



Effect of increased $p\text{CO}_2$ level on early shell development in great scallop (*Pecten maximus* Lamarck) larvae

S. Andersen¹, E. S. Grefsrud², and T. Harboe¹

¹Institute of Marine Research, Austevoll Research Station, 5392 Storebø, Norway

²Institute of Marine Research, Postbox 1870 Nordnes, 5817 Bergen, Norway

Correspondence to: S. Andersen (sissel.andersen@imr.no)

Received: 25 January 2013 – Published in Biogeosciences Discuss.: 22 February 2013

Revised: 16 August 2013 – Accepted: 22 August 2013 – Published: 1 October 2013

Abstract. As a result of high anthropogenic CO_2 emissions, the concentration of CO_2 in the oceans has increased, causing a decrease in pH, known as ocean acidification (OA). Numerous studies have shown negative effects on marine invertebrates, and also that the early life stages are the most sensitive to OA. We studied the effects of OA on embryos and unfed larvae of the great scallop (*Pecten maximus* Lamarck), at $p\text{CO}_2$ levels of 469 (ambient), 807, 1164, and 1599 μatm until seven days after fertilization. To our knowledge, this is the first study on OA effects on larvae of this species. A drop in $p\text{CO}_2$ level the first 12 h was observed in the elevated $p\text{CO}_2$ groups due to a discontinuation in water flow to avoid escape of embryos. When the flow was restarted, $p\text{CO}_2$ level stabilized and was significantly different between all groups. OA affected both survival and shell growth negatively after seven days. Survival was reduced from 45 % in the ambient group to 12 % in the highest $p\text{CO}_2$ group. Shell length and height were reduced by 8 and 15 %, respectively, when $p\text{CO}_2$ increased from ambient to 1599 μatm . Development of normal hinges was negatively affected by elevated $p\text{CO}_2$ levels in both trochophore larvae after two days and veliger larvae after seven days. After seven days, deformities in the shell hinge were more connected to elevated $p\text{CO}_2$ levels than deformities in the shell edge. Embryos stained with calcein showed fluorescence in the newly formed shell area, indicating calcification of the shell at the early trochophore stage between one and two days after fertilization. Our results show that *P. maximus* embryos and early larvae may be negatively affected by elevated $p\text{CO}_2$ levels within the range of what is projected towards year 2250, although the initial drop in $p\text{CO}_2$ level may have overestimated the effect of the highest $p\text{CO}_2$ levels. Future work should focus on long-term

effects on this species from hatching, throughout the larval stages, and further into the juvenile and adult stages.

1 Introduction

The increase of anthropogenic CO_2 emissions since the industrial revolution has led to an increase of carbon dioxide (CO_2) concentration and a decrease in pH in the oceans termed as ocean acidification (OA). Over the next 50–100 yr the CO_2 values are projected to reach 700 ppm (Zondervan et al., 2001; IPCC, 2001). Caldeira and Wickett (2003) showed through their geochemical model that with no reduction in fossil fuel emission of CO_2 into the atmosphere, there may be a rapid decline in pH in surface ocean waters over the next 100 yr of as much as 0.4 units (a doubling of H^+ -ion concentration). And within less than 250 yr the ocean surface pH may drop as much as 0.7 units, which may be the lowest pH value experienced during the last 300 million years (Caldeira and Wickett, 2003).

Depending on their distribution and habitats, marine organisms are exposed to various levels of pH fluctuations. In general, deep-sea species live in a more stable environment compared to shallow-living species that experience both seasonal and daily fluctuations in physio-chemical water parameters. In areas with natural high CO_2 supply (upwelling of deep-sea CO_2 -rich water and volcanic CO_2 vents) marine organisms are exposed to seawater CO_2 values as high as 2000 ppm, corresponding to a pH of 7.4–7.5 (Hall-Spencer et al., 2008; Thomsen et al., 2010). Some organisms can experience even lower pH values in nature, as shown for the oyster *Ostrea chilensis* (Chaparro et al., 2009). During valve closure

and isolation of the brood chamber, the brooded veligers were exposed to pH values as low as 7.0.

Although many marine organisms show a high tolerance to variations in pH, several experimental studies conducted over the last decade have shown negative effects of OA, especially on calcifying organisms (Kroeker et al., 2010). The effects of elevated CO_2 levels are shown on a range of marine invertebrates (embryonic and larval development in marine molluscs – Byrne, 2012; Ericson et al., 2010; Gaylord et al., 2011; Gazeau et al., 2011, Kurihara, 2008; Kurihara et al., 2007, 2008; Waldbusser et al., 2010; and effect on other invertebrates – Bechmann et al., 2011; Dupont et al., 2010; Jury et al., 2010; Comeau et al., 2009, 2010; Ries et al., 2009; Hoegh-Guldberg et al., 2007; Pane and Barry, 2007). Especially larval stages of bivalves seem to be extremely sensitive to enhanced levels of CO_2 (Fabry et al., 2008; Talmage and Gobler, 2009, 2010, 2011), and most studies have shown a decrease in survival, shell growth, and normal larval development. Some studies have also shown a negative effect on shell thickness and strength (Chaparro et al., 2009; Gaylord et al., 2011; Welladsen et al., 2010), calcification (Ries et al., 2009; Waldbusser et al., 2010; Range et al., 2011), lipid synthesis (Talmage and Gobler, 2011), and energy stores (Dickinson et al., 2012). The high sensitivity of bivalve larvae may be connected to the carbonate composition of the larval shell. A mollusc's larval shell contains mainly aragonite, the most soluble form of crystalline calcium carbonate (Weiss et al., 2002; Kudo et al., 2010; Barros et al., 2013). The predicted undersaturation of aragonite in the world's oceans may have a great negative impact on these calcifying organisms (Orr et al., 2005; Raven et al., 2005).

The numerous reports on effects of OA are slowly building up a picture of how marine coastal ecosystems may be affected (biodiversity, functioning, and service) in the future (Raven et al., 2005). It is crucial to gain more knowledge about the effects on different marine organisms in order to get realistic projections of future changes of the marine food web. A negative effect on early life stages may not only be detrimental to recruitment and endanger the species survival but also result in economic loss due to a collapse in global shellfish aquaculture production (Gazeau et al., 2010). Together with mussels and oysters, several scallop species are economically important in global aquaculture production (Bourne, 2000; FAO, 2010) in addition to being important calcifiers in the marine ecosystem.

The great scallop *Pecten maximus* Lamarck is distributed from the Iberian Peninsula in the south up to Lofoten in Norway in the north, and is found in highest abundances at 5–50 m depth in shell sand areas (Strand and Parsons, 2006). The larvae are pelagic for about 15–32 days at temperatures of 15–18 °C (Comely, 1972; Le Pennec et al., 2003). The great scallop is a highly valued sea food product, and has been cultured for more than 25 yr in Norway. The extended research efforts during the same period make it an excellent candidate for OA studies (Andersen et al., 2011). To our

knowledge, only the study of Schalkhausser et al. (2012) has shown the effect of OA on *P. maximus*. They found a negative effect of OA on adult *P. maximus* shell-clapping performance and aerobic scope. However, no studies have focused on the effect of OA on *P. maximus* larvae. Studies on other scallop species as *Argopecten irradians*, *Placopecten magellanicus* and *Chlamys nobilis* show a negative effect of increased pH in most life stages, including fertilization, embryos, larvae, and juveniles (Desrosiers et al., 1996; Talmage and Gobler, 2009, 2010, 2011; Liu and He, 2012).

Most OA studies on mollusc larvae focus on effects on shell formation (Kurihara et al., 2007, 2008; Kurihara, 2008; Ericson et al., 2010; Waldbusser et al., 2010; Byrne, 2012; Gaylord et al., 2011; Gazeau et al., 2011). The shell that starts to evolve at the trochophore stage consists of proteins (Bellioli et al., 1993; Casse et al., 1998). Onset of calcification in the initial organic larval shell was detected during the trochophore larval stage in *Mercenaria mercenaria* and *Crasostrea gigas* larvae (Weiss et al., 2002). Casse et al. (1998) did not detect calcification in *P. maximus* trochophores, and according to Cragg (2006) it is likely that there is initially little or no calcification in pectinid larvae, but that the calcification starts within less than a day after the initial shell formation. If calcification is the actual process being affected, effects on the shell formation should be detected only after the stage when the calcification starts. The onset of calcification can be determined by the use of an epifluorescent dye, calcein, which binds to calcium in metabolically active individuals (Körbl and Vydra, 1958; Day et al., 1995; Lucas et al., 2008; Chaparro et al., 2009).

The aim of the present study was to report the effect of elevated CO_2 levels on scallop *P. maximus* Lamarck larvae survival, shell size, and shell development from fertilized eggs to 7-day-old veligers. The exact initiation of calcification in *P. maximus* larvae is still unknown, and thus a pilot study using calcein staining as a method to elucidate this question was also conducted.

2 Material and methods

Local broodstock were collected in January 2012 from the outer Hardangerfjorden, located on the south-western coast of Norway. Ten scallops were conditioned for eight weeks at 12.0–12.5 °C and fed 13–15 cells μL^{-1} of a standard diet containing *Isochrysis galbana* (Tahitian), *Pavlova lutheri* and *Chaetoceros mulleri* (Andersen et al., 2011). Spawning was induced on 21 March by increasing the temperature to 16–17 °C. *Pecten maximus* is a simultaneous hermaphrodite, and some sperm will remain in the genital tract, resulting in self-fertilization (Mackie, 1984). Self-fertilization rate was checked before cross-fertilization, and egg batches with less than 10% self-fertilization were used in the experiment. Eggs from one individual were cross-fertilized with sperm from three other individuals. Egg:sperm ratio was

Table 1. Measured and calculated water parameters given as mean \pm sd for four different $p\text{CO}_2$ groups (μatm). Carbon chemistry values were computed based on daily measurements (0–6 days) of pH_{NBS} in all replicates ($n = 4$, $*n = 3$), means of hourly temperature measurements in three tanks ($n = 468$), salinity (daily), and total alkalinity (11 analyses from 3 dates) in seawater running into the lab. Coefficient of variation (CoV) range (sd as % of mean) is shown for variation in CO_2 (ppm) between days (d) for the replicates, and between replicates (r) for all days.

$p\text{CO}_2$ group	469 μatm	807 μatm	1164 μatm^*	1599 μatm
Measured parameters				
pH_{NBS}	7.98 \pm 0.01	7.78 \pm 0.05	7.64 \pm 0.07	7.51 \pm 0.10
Salinity	35.0	35.0	35.0	35.0
Temperature ($^{\circ}\text{C}$)	15.6 \pm 0.4	15.6 \pm 0.4	15.7 \pm 0.3	15.6 \pm 0.4
A_{T} (mmol kg^{-1} SW)	2341 \pm 11	2341 \pm 11	2341 \pm 11	2341 \pm 11
Calculated parameters				
$p\text{CO}_2$ (μatm)	469 \pm 9	807 \pm 106	1164 \pm 230	1599 \pm 421
HCO_3^- ($\mu\text{mol kg}^{-1}$ SW)	1973 \pm 5	2097 \pm 24	2159 \pm 25	2200 \pm 28
CO_3^{2-} ($\mu\text{mol kg}^{-1}$ SW)	149 \pm 2	99 \pm 10	74 \pm 10	57 \pm 11
CO_2 ($\mu\text{mol kg}^{-1}$ SW)	17.2 \pm 0.3	29.5 \pm 3.8	42.5 \pm 8.1	58.4 \pm 15.1
$\Omega_{\text{aragonite}}$	2.28 \pm 0.03	1.52 \pm 0.15	1.13 \pm 0.15	0.88 \pm 0.17
CO_2 (ppm)	477 \pm 9	821 \pm 108	1184 \pm 222	1627 \pm 429
CoV (d) CO_2	1.14–2.2	12.5–15.9	18.9–20.6	23.3–33.7
CoV (r) CO_2	0.0–2.2	1.3–11.4	1.4–11.4	1.2–13.8

approximately 1 : 100. Fertilized egg batches were mixed and incubated 1–2 h after fertilization. Egg number was determined by counts in 10 subsamples of 50 μL , and 13 fertilized eggs mL^{-1} were incubated in 38 L exposure tanks at an ambient pH_{NBS} of 7.98 (control) and mean pH_{NBS} levels of 7.71, 7.64, and 7.51, corresponding to a $p\text{CO}_2$ level of 469, 807, 1164, and 1599 μatm (Table 1). The pH levels were used based on the predicted drop of 0.5 units from present day to year 2250 (IPCC, 2001). Mean temperature based on hourly measurements in three replicates (\pm sd) was 15.6 \pm 0.3 $^{\circ}\text{C}$ ($n = 495$). Seawater with a salinity of 35 ppm was pumped from 160 m depth and filtered through a sand filter before temperature was adjusted in a heat pump. The water was aerated and finally filtered through a 50 μm filter. The experiment was conducted in a flow-to-waste system consisting of circular fiberglass tanks with slightly conical bottom, resulting in a depth difference of 22 mm between tank centre and edge. The inner tank diameter was 500 mm, and water depth was 226–257 mm at the tank edge. Four replicates were used per pH level, and the flow rate was set to 10 L h^{-1} per exposure tank. One replicate was lost after three days in the group exposed to a mean pH_{NBS} of 7.64 (1184 ppm) due to overflow. To prevent escape of embryos, water flow in the experimental tanks was turned off overnight at the incubation of fertilized eggs. A slow flow ($< 10 \text{L h}^{-1}$) was re-established after 10.5 h, and then increased to 15 L h^{-1} (flow meter readings) after another 3.3 h. Slow bubbling of pressurized air was introduced at the tank bottom through a perforated plastic tube. Larvae were not fed during the experimental period.

Seawater at different pH levels was produced by mixing seawater with an acid stock solution of pH_{NBS} 5.80, made from mixing CO_2 gas and seawater with an ambient pH_{NBS} of 7.95. The pH_{NBS} in each mixing tank was continuously adjusted to preset levels by addition of stock solution with dosage pumps (IWAKI) controlled by feedback from pH electrodes to pH transmitters (Endress & Hauser). The different water qualities were distributed by gravity to the exposure tanks via four high-mounted header tanks. The water level in header and mixing tanks was kept constant by flotation valves, and water was pumped in a loop between the mixing tanks and header tanks. Water flow in the loop was much higher than the flow from the header tanks to the exposure tanks in order to keep the pH level constant in the header tanks.

Exposure tanks were illuminated continuously by fluorescent tubes (Philips TL-D 18w/830) 110 cm above the tanks. Light was dispersed through an opaque sheet of polycarbonate 70 cm above the exposure tanks.

The pH level in each exposure tank was measured daily in a 100 mL sample using a Mettler Toledo equipped with a Mettler Toledo InLab[®] ExpertPro pH-probe, calibrated with 4.00 and 7.00 buffers (Certipur[®] buffer solutions, Merck KGaA, 64271 Darmstadt, Germany) traceable to standard reference material from NIST (NBS). Total alkalinity (A_{T}) was analysed in four replicate samples on 9, 15, and 26 March 2012 by a TitraLab radiometer. Salinity was checked daily using a WTW LF330 conductivity meter, and temperature recorded every hour using a four-detector (one in air and three in exposure tanks) EBI – 1 Ebro 4 temperature

logger. The $p\text{CO}_2$ values (μatm) corresponding to the pH_{NBS} values were calculated from the daily average temperature ($^{\circ}\text{C}$) based on hourly recordings, salinity, A_{T} and pH using the macros in Ernie Lewis' "CO2SYS.BAS" basic program (Pierrot et al., 2006).

Larvae samples were taken after carefully stirring from bottom to top using a plunger. A 50 mm diameter pipe was lowered to the bottom, and both ends were plugged before the pipe was pulled out of the water, collecting a volume of 400 mL. On day 2 a volume of 800 mL from each replicate was withdrawn to measure shell size and classify larval deformities. When terminating the experiment after seven days, all tanks were drained and a total sample from each replicate was collected and preserved to determine final shell size, survival, and classify deformities. To determine timing of shell calcification, larvae from ambient pH in an additional tank were stained for 18–20 h from day 1 to day 2.

When collected for preservation, embryos and larvae were retained on a 35 μm mesh screen and preserved in 4 % formalin (diluted from 36.5 % formaldehyde solution, Sigma-Aldrich, cat. no. 33220) in PSB for two hours, rinsed twice in 70 % ethanol, and then kept in 70 % ethanol until the ratio of developmental stages was investigated. The formalin was made from mixing 8.14 g of disodium hydrogen phosphate dihydrate ($\text{Na}_2\text{HPO}_4 \cdot 2\text{H}_2\text{O}$, Merck 1.06580.1000) and 4.00 g of sodium dihydrogen phosphate monohydrate ($\text{NaH}_2\text{PO}_4 \cdot \text{H}_2\text{O}$, Merck 1.06346.1000) with 100 mL of 37 % formalin in 900 mL of distilled water.

A calcein stock solution was made from mixing 1 g of calcein (Sigma C0875-5G) in 1 L of distilled water adjusted to pH_{NBS} 7.0 using a sodium bicarbonate (NaHCO_3 , Merck 1.06329.1000) solution of 2.4 g in 400 mL of distilled water (Moran, 2000; Moran and Marko, 2005; Fitzpatrick et al., 2010). For staining of calcium, 20 mL of calcein stock solution was added to 400 mL of seawater with larvae, giving a final calcein concentration of 22 mg L^{-1} . Only larvae at ambient pH_{NBS} were stained, as the aim of this pilot study was to elucidate when the calcification of the larvae shell is normally initiated. We assumed that the pH was not affected during staining since the calcein stock solution was less than 5 % of the total water volume; however the pH was not measured. Calcein has little or no effect on mollusc larvae survival or growth (Moran, 2000; Moran and Marko, 2005; Chick, 2010; Fitzpatrick et al., 2010).

Live (calcein stained) and preserved larvae were photographed using a Canon EOS 5D Mark II camera attached to an Olympus BX60 microscope. Final shell length (parallel to the hinge) and height (perpendicular to the hinge) were measured in 21–60 individuals to the nearest 4.8 μm from printed photographs. Survival was based on the initial number of fertilized eggs, and was estimated by concentrating all day-7 larvae in 25–110 mL and counting the number in 10 subsamples of 50 μL . Preserved larvae photographed at 100 magnifications using bright field were classified according to shell shape and counted. The number of trochophores classi-

fied on day 2 was 76–150, and 46–178 veligers were examined on day 7. Since we were unable to find any published study on deformities in scallop larvae, deformities were categorized according to deformities as described in larvae of mussel (*Mytilus* sp) and the Pacific oyster (*C. gigas*) (His et al., 1997). We observed the same "convex hinge" and "indented shell margins" in our larvae, but called the categories "hinge deformity" and "edge deformity", respectively. We classified preserved larvae into four categories according to shell shape, and counted (1) normal (Fig. 1a), (2) hinge deformity (Fig. 1b–c), (3) edge deformity (Fig. 1d–e), and (4) both edge and hinge deformity (Fig. 1f). Trochophore larvae on day 2 were only classified into category 1 and 2, since shell edge was not visible until the shell valves covered the whole larva. In veligers on day 7, all four categories were used.

Live, calcein-stained larvae were photographed at 10×40 magnification using an U-MSWB2 cube filter (excitation 420–440 nm, emission 475 nm) and UV light to show yellow-green fluorescence when calcein bonded with calcium (Moran, 2000; Moran and Marko, 2005; Fitzpatrick et al., 2010).

Statistica version 11 (Statsoft Inc.) was used to run statistical tests. Arcsine-transformed survival data and shell dimension data were not confirmed to normality (Shapiro–Wilk's W test), and significant effects were tested using Kruskal–Wallis ANOVA by ranks. Differences between groups were tested using multiple comparison p values (two-tailed). The rest of the data were tested using one-way ANOVA, followed by unequal N HSD post hoc test to find differences between groups. The significance level used in all tests was set to 0.05.

3 Results

3.1 Initial variation in $p\text{CO}_2$ (μatm)

Measured and calculated water parameters for the four $p\text{CO}_2$ groups are given in Table 1. Salinity and total alkalinity (A_{T}) showed little or no variation, including in periods before and after the experiment (February–May). The $p\text{CO}_2$ values (μatm) for the elevated groups showed a sharp decrease during the first night after incubation due to a discontinuation in seawater flow in the exposure tanks (Fig. 2). This drop in $p\text{CO}_2$ (delta, $\Delta p\text{CO}_2$) increased with increasing $p\text{CO}_2$ levels, and fell from a mean of 2489 to 1203 μatm ($\Delta p\text{CO}_2$ of 1286 μatm) in the highest $p\text{CO}_2$ group. The ambient group showed only an insignificant $\Delta p\text{CO}_2$ of 21 μatm . Naturally, the variation in $p\text{CO}_2$ values given as the coefficient of variation (CoV), increased with increasing $p\text{CO}_2$ levels, and was higher between days for replicates than between replicates on any day (Table 1).

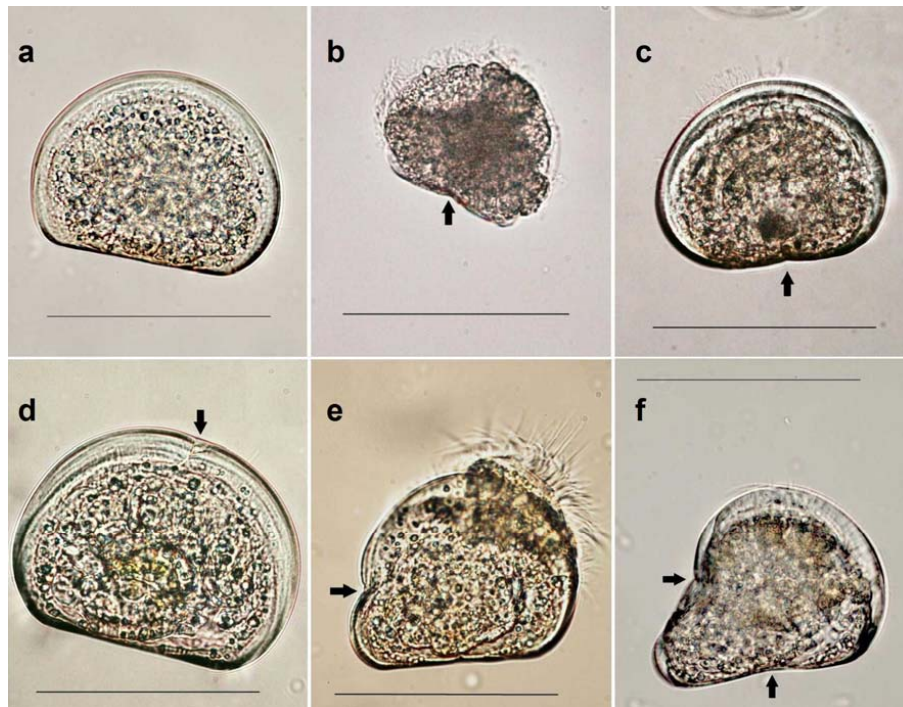


Fig. 1. Shell shape categories observed in *Pecten maximus* larvae. (a) Normal shell shape in formalin-preserved day-7 veliger larva; (b) hinge deformity in formalin-preserved day-2 trochophore larva and (c) in day-3 veliger larva; (d) edge deformity in shell edge centre of a live day-7 veliger larva and (e) in shell edge side of a live day-6 veliger larva; (f) both hinge and edge deformity in a live day-7 veliger larva. The scale bar is 100 μm .

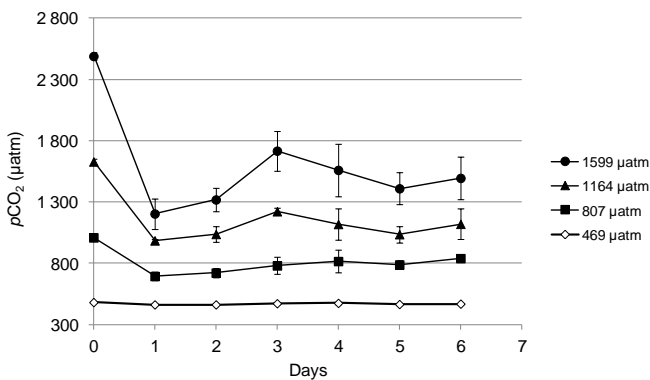


Fig. 2. The daily means of $p\text{CO}_2$ (μatm) \pm sd for replicates in the treatments and ambient group during the experimental period of seven days ($n = 4$, $n = 3$ for 1164 μatm).

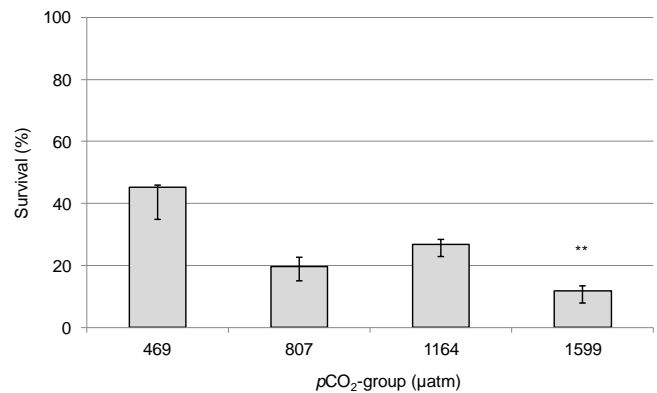


Fig. 3. Survival of day-7 scallop larvae given as median \pm quartile in the four $p\text{CO}_2$ groups as percentage of eggs incubated. ** Significantly different from the other $p\text{CO}_2$ groups.

3.2 Survival

The median survival of day-7 veligers based on the initial number of incubated eggs reared at ambient $p\text{CO}_2$ was 45 % (range 25–46) and decreased with elevated $p\text{CO}_2$ levels (Fig. 3). At the highest $p\text{CO}_2$ level the median survival was 12 % (range 6–13 %). There was a significant effect of $p\text{CO}_2$ on survival ($p = 0.013$), but only the group at the highest $p\text{CO}_2$ level was significantly different from the control.

3.3 Shell size

Shell length and height on day 7 were both negatively affected by an increase in $p\text{CO}_2$ level ($p < 0.001$ for both) showing a decrease in median values of 8 and 15 %, respectively, when $p\text{CO}_2$ increased from 469 μatm (ambient) to the highest $p\text{CO}_2$ of 1599 μatm (Fig. 4). The median shell length and height were 115 and 91 μm , respectively, in the ambient

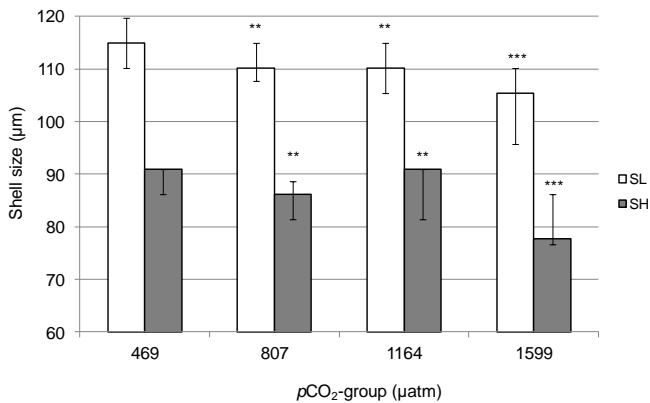


Fig. 4. Scallop larvae shell length (SL) and height (SH) on day 7 given as median \pm quartile for four $p\text{CO}_2$ groups. Each median represents measurements of 21–60 individuals in 3–4 replicates. Unmarked, **, and ***-marked CO_2 groups are significantly different within the same variable.

group. All the $p\text{CO}_2$ groups were significantly different for both parameters, except for the two $p\text{CO}_2$ groups 807 and 1164 μatm .

3.4 Shell deformities

On day 2 an average of 85 % (82–88 %) of examined individuals had developed a shell, and only 1 % had not developed further from egg stage. Around 28–68 % of all individuals in the different $p\text{CO}_2$ groups on day 2 had developed a normal-looking straight hinge (Fig. 5). On day 7 the percentage of normal hinges was higher: 69–97 % for all $p\text{CO}_2$ groups. The percentage of larvae with a normal hinge was negatively affected by $p\text{CO}_2$ on both days, ($p = 0.004$ and $p = 0.021$ for d2 and d7, respectively), and the lowest percentages were found at the highest $p\text{CO}_2$ level (Fig. 5). The percentages of all four shell shape categories on day 7 were affected by increasing $p\text{CO}_2$ level ($p < 0.000$ – 0.005) (Fig. 6). The percentage of normally shaped larval shells decreased from 64 to 28 % with an increase in $p\text{CO}_2$ level from ambient to 1599 μatm . The effect of elevated $p\text{CO}_2$ level seemed to follow no clear pattern as the ambient group was significantly different only from the least (807 μatm) and the most elevated $p\text{CO}_2$ level (1599 μatm). Shell edge deformities accounted for 30–57 % in the different $p\text{CO}_2$ groups (Fig. 6), and the control was significantly different only from the least elevated group. The percentage of larvae with hinge deformities, and both hinge and edge deformities, increased when the $p\text{CO}_2$ level increased. Only the most elevated $p\text{CO}_2$ group was significantly different from the ambient group for these two shell deformity categories. Variation within replicates was high, and the coefficient of variation (CoV) ranged between 5 and 70 % for all $p\text{CO}_2$ groups and categories.

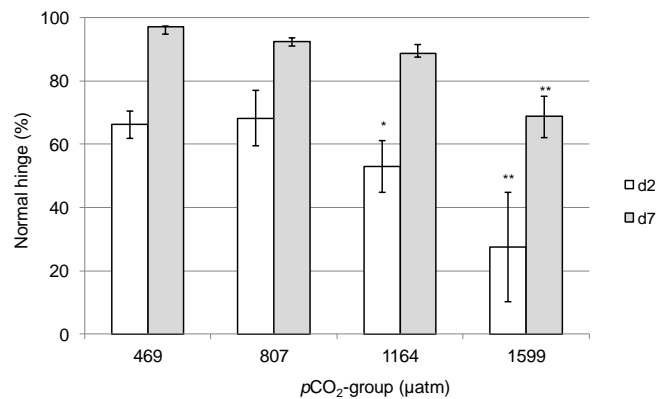


Fig. 5. Percentage of larvae showing a normal, straight hinge on day 2 (d2) and day 7 (d7) based on the total larval number. Values are means \pm sd for d2, and medians \pm quartiles for d7 ($n = 4$, and $n = 3$ for 1164 μatm). * denotes not significantly different from any group within same variable; **, significantly different from unmarked groups within same variable.

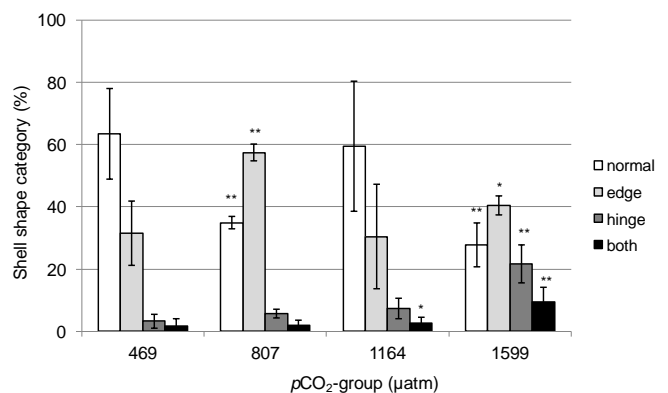


Fig. 6. Percentage of the four scallop larvae shell shape categories on day 7 for different $p\text{CO}_2$ groups. Values are mean ($n = 4$, $n = 3$ for 1164 μatm), and error bars are sd. * denotes not significantly different from any group within same category; **, significantly different from unmarked groups within same category.

3.5 Calcification

Trochophore larvae in ambient seawater (Fig. 7a, b(i)) stained with calcein overnight (day 1–2) showed green fluorescence when exposed to UV light (Fig. 7a, b(ii), (iii)). The fluorescence was highly concentrated in a small area on each valve, closer to the hinge than the shell margin centre. Also, there was a weak fluorescence from the total shell area, indicating that calcification of the organic shell started between day 1 and day 2.

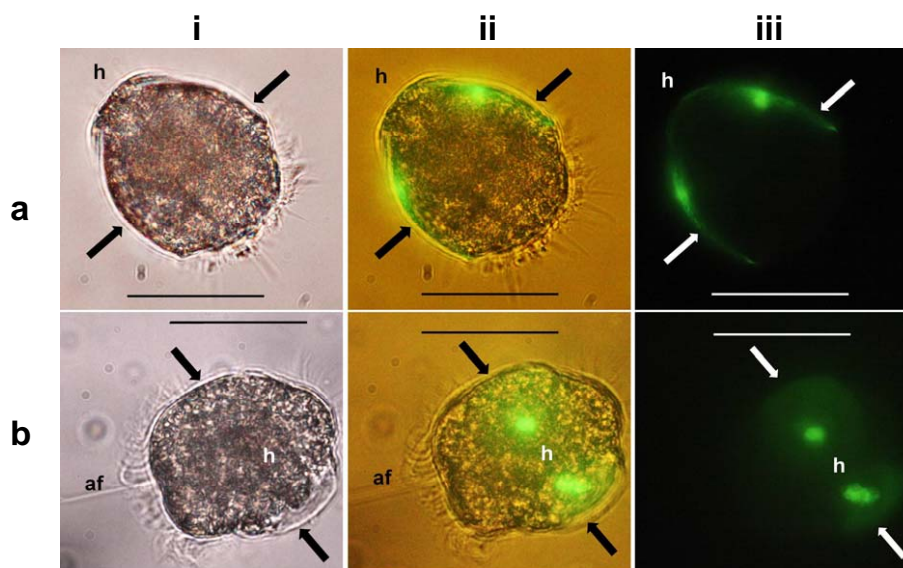


Fig. 7. Stereomicroscopic images of live trochophore scallop larvae stained with calcein from day 1 and overnight (18–20 h). **(a)** A side view with the hinge area (h) oriented towards the upper-left corner and a newly formed shell valve on each side (arrows). **(b)** The hinge area (h) and early shell formation (arrow) using (i) bright field, (ii) bright field and UV light, and (iii) UV light. Green colouring in UV light (fluorescence) shows areas containing calcium. The scale bar is 50 μm , and magnification is the same for all photographs.

4 Discussion

4.1 Non-feeding regime

In the present study, scallop larvae were not fed during the 7-day experimental period. Lack of food is known to affect the energy status of scallop larvae after only three days (Andersen et al., 2012); however, the survival and growth in the control group indicate that the unfed larvae were healthy and viable. Previous OA experiments on bivalve larvae support that the decreased survival, smaller size, and increased percentage of deformed *P. maximus* larvae found in the highest $p\text{CO}_2$ group was caused by the change in water chemistry rather than the feeding regime (feeding or non-feeding) (Table 2).

Thomsen et al. (2010) showed that calcifying keystone species may be able to handle the change in seawater chemistry projected for the end of the century as long as food supply is sufficient. The lack of food can be regarded as a stress factor that may push the larval energy status towards a point where the effects will be visible as slower growth, increased mortality, and increased shell deformity. We believe that in nature, bivalve larvae may experience shorter or longer periods of food scarcity; thus unfed larvae may not be uncommon, and most species may be quite well adapted to such conditions. More studies are needed to elucidate how well *P. maximus* larvae are able to handle the future $p\text{CO}_2$ levels at different feeding regimes.

4.2 Initial drop in $p\text{CO}_2$ level

The initial drop in $p\text{CO}_2$ level was most likely due to both the discontinuation in seawater flow and gas exchange at the water surface due to slow aeration during the first 12 h. This drop may have added additional physiological stress to the scallop larvae, especially in the higher $p\text{CO}_2$ groups. Both the delta $p\text{CO}_2$ and the initial maximum values may have contributed to an overestimation of the effect on all parameters (survival, growth, and deformities) that were given for the average $p\text{CO}_2$ levels, especially for the two groups exposed to the highest levels. Surface waters (upper 100 m) of the open ocean vary between pH 7.9 and 8.3 depending on season and geographic region, and coastal areas may even show higher fluctuations and relatively rapid changes in $p\text{CO}_2$ levels may be expected (Raven et al., 2005; Blackford and Gilbert, 2007; Kerrison et al., 2011, and references therein). *P. maximus* is most abundant at 5–50 m depth in Norwegian coastal waters, and may be exposed to such variable levels in pH in nature. However, short-term fluctuations within hours have not yet been quantified to our knowledge. Comparing our results with OA studies using more or less the same $p\text{CO}_2$ or pH levels indicates that the effects we measured were most likely due to elevated $p\text{CO}_2$ over time rather than a result of the sudden drop in the first hours of the experiment (Talmage and Gobler, 2010, 2009; Watson et al., 2009). Experiments with rapid changes in pH can elucidate how variation experienced under natural conditions may affect marine organisms. However, such experiments should be conducted under controlled conditions, not as a result of methodological constraints.

Table 2. Overview of ocean acidification (OA) effect studies. The focus is on bivalves, but some studies on other calcifying invertebrates are included. “Volume” is the experimental unit volume, “n” is the number of replicates, “water exchange” is the flow rate in the flow-through system, and if the system is stagnant “water exchange” gives the exchange rate of the total volume. NG – information not given in the original publication.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L ⁻¹)	Flow or stagnant (water exchange)	Temp. °C (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects: Negative (–) None (0) Positive (+)	Reference
Bivalves										
<i>Argopecten irradians</i> bay scallop	Embryos (0 d)	1 L (3, 4)	350 (350)	Stagnant (2 times a week, + antibi-otics)	24, 28	3 weeks	Yes (1)	pHNBS 8.2, 8.1, 7.8 CO ₂ 220–247, 374–387, 756–795 ppm	Development (–) Survival (–) Growth (–) Lipid synthesis (–) for both OA and temperature, additive effects	Talmage and Gobler (2011)
	Larvae (1 d)	1 L (4)	200 (200)	Stagnant (2 times a week, + antibi-otics)	24 (28)	52 days	Yes (1)	CO ₂ 244, 387, 739, 1529 ppm, pHr 8.17, 8.041, 7.801, 7.530	at pH < ambient Growth (–) Metamorphosis (–) Survival (–) Lipid accumulation rates (–) Normal shell development (–) Shell thickness (–) at pH > ambient Growth (+) Metamorphosis (+) Survival (+) Lipid accumulation rates (+) Shell thickness (+)	Talmage and Gobler (2010)
<i>A. irradians</i>	Larvae (1 d)	1 L (4)	100 (100)	Stagnant (2 times a week, + antibi-otics)	24 (28)	18–20 days	Yes (1)	pHr 8.1, 7.8, 7.5 pCO ₂ 39, 66, 152 Pa	Survival (–) Metamorphosis (–) Size (–)	Talmage and Gobler (2009)
	Juveniles (15,9 mm)	10 L (3)	10 (1)	Stagnant (every days, + antibi-otics)	24, 28	45 days	Yes (1)	pHNBS 8.1, 7.6 CO ₂ 400–400, 1665–1737 ppm	OA Shell growth (–) Tissue growth (0) Temperature Shell growth (–) Tissue growth (–) no additive effects	Talmage and Gobler (2011)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L^{-1})	Flow or stagnant (water exchange)	Temp. $^{\circ}\text{C}$ (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects: Negative (-) None (0) Positive (+)	Reference
<i>Pecten maximus</i> (king scallop)	Adults (108.5 mm shell height)	200 L (8)	10 (0.05)	Recirculation (exchanged twice a week)	4, 10	33–60 days	Yes (3)	$p\text{CO}_2$ at 4 $^{\circ}\text{C}$: 0.040 and 0.110 kPa $p\text{CO}_2$ at 10 $^{\circ}\text{C}$: 0.040 and 0.115 kPa pHNBS at 4 $^{\circ}\text{C}$: 8.19 and 7.76 pHNBS at 10 $^{\circ}\text{C}$: 8.25 and 7.81 Temperature	Temperature Survival (-) OA Clapping performance (-) Metabolic rate (-)	Schalkhauser et al. (2012)
<i>Placopecten magellanicus</i> (sea scallop)	Gametes, oocytes, embryos	NG	NG	NG	10 (25)	180 h	No	pH 7.0, 7.5, 8.0, 8.5	Extrusion of polar bodies (0) Duration of first mitotic division shorter at pH 8–8.5 Completion of first embryonic cleavage (-) at pH 7.0 and 7.5 Levels of polyspermy (+) at pH 7.0 and 7.5	Desrosiers et al. (1996)
<i>Chlamys nobilis</i> noble scallop	Juveniles (40–44 mm)	10 L (5)	4 (0.4)	Stagnant (daily)	25 (32)	5 days	Yes (1)	pHNBS 8.1, 7.7, 7.4	Clearance (-) Respiration rate (0), (-) at pH 7.4 Excretion (-) at pH 7.4	Liu and He (2012)
<i>Crassostrea gigas</i> Pacific oyster	Gametes	100 mL (3)	500 (5000)	Stagnant	18, 22, 26, 30 (35)	48 h	No	$p\text{CO}_2$ 375, 600, 750, 1000 μatm	D-veliger % (-) Abnormality % (+) Size (-) All less at optimum temperatures, more when fertilized at elevated $p\text{CO}_2$	Parker et al. (2010)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (indL ⁻¹)	Flow or stagnant (water exchange)	Temp. °C (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects: Negative (-) None (0) Positive (+)	Reference
<i>C. gigas</i>	Embryo (0d)	4.5 L (3)	67 500 (15 000)	Stagnant	18.9 (34.0)	3 days	No	pH _T 8.03, 7.72, 7.41, 7.67, 7.62 pCO ₂ 449, 1020, 2171, 494, 3730 A _T 2453, 2446, 2443, 1094, 6727 S _{car} 2.8, 1.5, 0.8, 1.6, 3.5	Above CO ₃ ²⁻ saturation level: Developmental success (0) Shell length (0) Shell area (0) Incorp. calcium (0) Below CO ₃ ²⁻ saturation level: Developmental success (-) Shell length (-) Shell area (-) Incorp. calcium (-)	Garzeau et al. (2011)
	Embryo (0d)	50 mL (5)	< 500 (< 10 000)	Stagnant (sealed)	23 (33.7)	2 days	No	pHNBS 8.21, 7.42	Normal development of gastrula (0) Shell size (-) Normal shell shape (-) Shell mineralization (-)	Kurihara et al. (2007)
<i>C. gigas</i>	Larvae (0d)	20 L (3)	60 000	Flow	17.9 (30.6)	148 h	Yes (1)	pH _T s 8.09, 7.76, 7.37 pCO ₂ 580, 1386, 3573 µatm	Fertilization (-) Sperm motility and velocity (-) Survival (-) Normal shell development (-) Growth (-)	Barros et al. (2013)
	Umbonate (137 µm)	100 mL (2500)	250 (2500)	Stagnant (daily)	18, 22, 26, 30 (35)	4 days	Yes (3)	pCO ₂ 375, 600, 750, 1000 µatm Temperature	Size (-), not at 18 °C	Parker et al. (2010)
<i>C. gigas</i>	Pediveliger (246 µm)	100 mL (2000)	200 (2000)	Stagnant (daily)	18, 22, 26, 30 (35)	4 days	Yes (3)	pCO ₂ 375, 600, 750, 1000 µatm Temperature	Size (-), not at 30 °C	Parker et al. (2010)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L^{-1})	Flow or stagnant (water exchange)	Temp. $^{\circ}\text{C}$ (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects: Negative (-) None (0) Positive (+)	Reference
<i>C. gigas</i>	Juveniles (5.5 mm)	100 mL	15 (150)	Stagnant (daily)	18, 22, 26, 30 (35)	4 days	Yes (3)	$p\text{CO}_2$ 375, 600, 750, 1000 μatm Temperature	Size (-), more so at 30 $^{\circ}\text{C}$	Parker et al. (2010)
<i>C. virginica</i> eastern oyster	Embryo	5 L (2)	5 (1)	Stagnant (every other day)	21 (30)	2 weeks	Yes	380, 3500 μatm	Standard metabolic rate (+)	Beniash et al. (2010)
<i>C. virginica</i>	Embryos (0 d)	1 L (3, 4)	350 (350)	Stagnant (2 times a week, + antibiotics)	24, 28	3 weeks	Yes (1)	pHNBS 8.2, 8.1, 7.8 CO_2 220–247, 374–387, 756–795 ppm	Development (-) Survival (-) Growth (-) Lipid synthesis (-) for both OA and temperature, additive effects	Talmage and Gobler (2011)
<i>C. virginica</i>	Larvae (1 d)	1 L (4)	100 (100)	Stagnant (2 times a week, + antibiotics)	24 (28)	18–20 days	Yes (1)	pH _T 8.1, 7.8, 7.5 $p\text{CO}_2$ 39, 66, 152 Pa	Survival (-) Metamorphosis (-) Size (-) only at 152 Pa	Talmage and Gobler (2009)
<i>C. virginica</i>	Juveniles (< 1 mm)	5 L (2)	50 (10)	Stagnant (every other day)	21 (30)	20 weeks	Yes (3)	$p\text{CO}_2$ 380, 3500 μatm	Survival (-) Dry shell mass (-) Soft-tissue mass Shell fracture toughness (-) Shell growth (-) Shell area (0) Standard metabolic rate (+)	Beniash et al. (2010)
<i>C. virginica</i>	Juveniles (7 weeks)	30 L (6 _{control} , 11–12)	NG	Stagnant (every other day)	21 (30, 15 psu)	11 weeks	Yes (3)	400, 700–800 μatm Salinity	Survival (-) Tissue energy stores (-) Soft tissue growth (-) salinity more than high CO_2	Dickinson et al. (2012)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (indL^{-1})	Flow or stagnant (water exchange)	Temp. ($^{\circ}\text{C}$ (Salinity))	Duration	Food supply (spp)	OA Treatment	Effects: Negative (-) None (0) Positive (+)	Reference
<i>C. virginica</i>	Juveniles (11.5 mm)	10L (3)	10 (1)	Stagnant (every 3 days, + antibiotics)	24, 28	45 days	Yes (1)	pHNBS 8.1, 7.6 CO_2 400–400, 1665–1737 ppm Temperature	OA Shell growth (-) Tissue growth (0) Temperature Shell growth (-) (0) at pH 7.6 Tissue growth (0) <i>C. virginica</i> No additive effects	Talmage and Gobler (2011)
	Embryos (0 d)	1L (3, 4)	350 (350)	Stagnant (2 times a week, + antibiotics)	24, 28	3 weeks	Yes (1)	pHNBS 8.2, 8.1, 7.8 CO_2 220–247, 374–387, 756–795 ppm Temperature	Both OA and temperature Development (-) Survival (-) Growth (-) Lipid synthesis (-) additive effects	Talmage and Gobler (2011)
<i>M. mercenaria</i> (hard clam)	Larvae (1 d)	1L (4)	100 (100)	Stagnant (2 times a week, + antibiotics)	24 (28)	18–20 days	Yes (1)	pHr 8.1, 7.8, 7.5 $p\text{CO}_2$ 39, 66, 152 Pa	Survival (-) Metamorphosis (-) Size (-)	Talmage and Gobler (2009)
	Larvae (1 d)	1L (4)	100 (100)	Stagnant (2 times a week, + antibiotics)	24 (28)	18–20 days	Yes (1)	pHr 8.02, 7.97, 7.88, 7.79 $p\text{CO}_2$ 41, 49, 60, 73 Pa	Survival (-) Metamorphosis (-)	Talmage and Gobler (2009)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L ⁻¹)	Flow or stagnant (water exchange)	Temp. °C (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects:	Reference
<i>M. mercenaria</i>	Larvae (1 d)	1 L (4)	200 (200)	Stagnant (2 times a week, + antibiotics)	24 (28)	36 days	Yes (1)	CO ₂ 247, 380, 742, 1529 ppm, pH _T 8.171, 8.052, 7.801, 7.532	Negative (-) None (0) Positive (+) pH < ambient Growth (-) Metamorphosis (-) Survival (-) Lipid accumulation rates (-) Normal shell development (-) Shell thickness (-) pH > ambient Growth (+) Metamorphosis (+) Survival (+) Lipid accumulation rates (+) Shell thickness (+)	Talmage and Gobler (2010)
<i>M. mercenaria</i>	Juveniles (6.1 mm)	10 L (3)	10 (1)	Stagnant (every days, + antibiotics)	24, 28	45 days	Yes (1)	pHNBS 8.1, 7.6 CO ₂ 400–400, 1665–1737 ppm	OA Shell growth (0) Tissue growth (0) Temperature Shell growth (-) Tissue growth (0) No additive effects	Talmage and Gobler (2011)
<i>Mercenaria</i> spp.	Post-larvae – juveniles (0.39–2.9 mm)	50 mL (1, 3, 4)	1–2 g (20–40 g)	Stagnant (daily)	20	8 h	Yes (1+ Shell-fish diet)	pHNBS 8.02, 7.64, 7.41 pCO ₂ 424, 1120, 1950 µatm	Calcification rate (-), all sizes, smaller sizes more affected, and at higher pH Dependent on hatchery populations (hybridized and pure <i>Mercenaria</i>)	Waldbusser et al. (2010)
<i>Mytilus californianus</i> California mussel	Larvae	2 L (6)	1000 (500)	Stagnant (every days)	15	9 days	Yes (1)	380, 540, 970 ppm	Shell length (-) Shell thickness (-) Shell size (-) Tissue mass (-)	Gaylord et al. (2011)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (indL ⁻¹)	Flow or stagnant (water change)	Temp. °C (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects:	Reference
<i>Mytilus edulis</i> blue mussel	Embryo (0 d)	130 L (3)	(10 000)	Stagnant	16.2–16.7 (31.9–32.1)	2 days	No	pHNBS 8.1 (460–640 µatm), pH 7.8 (1100–1200 µatm), pH 7.6 (460–640 µatm)	Shell length (–) Shell thickness (–) Hatching rate (0)	Gazeau et al (2010)
<i>M. edulis</i>	Larvae (2 d)	130 L (3)	(10 000)	Flow	19.2, 19.5 (31.4, 31.5)	13 days	Yes (2–3)	pHNBS 8.1, 7.8	Shell length and thickness (–) Survival (0)	Gazeau et al (2010)
<i>M. edulis</i> (20% hybrid with <i>Mytilus trossulus</i>)	Post-larvae	500 mL (7)	100 (200)	Stagnant (400 mL daily)	17	7 weeks	Yes (1)	$p\text{CO}_2$ 470, 1020, 2110, 3350 µatm 3 food levels	OA Shell size (–) Survival (0) Calcification (–) Organic growth (0) Food Shell size (+) Survival (0) Calcification (+) Organic growth (+) Explained most of the variation	Thomsen et al (2013)
<i>M. edulis</i>	Juveniles – Adults (5.5, 13.3 mm)	16 L (4)		Flow (100 mL min ⁻¹)	13.8 (15)	8 weeks	Yes (1)	pHNBS 8.13, 7.72, 7.26 $p\text{CO}_2$ 493, 1332, 3898 µatm	Shell length (–) at 3898 µatm Individual aragonite platelet thickness (–) at 3898 µatm Aragonite layer thickness (0) Calcite layer thickness (0) Shell dissolution Index (+) Shell dissolution area (+)	Thomsen et al (2010)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L ⁻¹)	Flow or stagnant (water exchange)	Temp. °C (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects:	Reference
<i>M. edulis</i>	Juveniles (16.6–17.6 mm)	18 L (4)	4 (0.22)	Flow (50 mL min ⁻¹)	4.7–5.4 (16)	7 weeks	Yes (1)	$p\text{CO}_2$ mean 54, 117, 248, 385 Pa pH _{NBS} mean 8.05, 7.73, 7.43, 7.19 Food 0.6 and 5.7 cells μL^{-1}	OA Shell growth (–) at 385 Pa Internal shell dissolution (–) no interactions Food Shell growth (+) Internal shell dissolution (+) interactions?	Melzner et al. (2011)
	Juveniles (45–55 mm)	50 L (2)	32 (0.64)	Flow (60 mL min ⁻¹)	16.03–16.11 (35.13–35.37)	60 days	Yes (1)	pH _{NBS} 8.0, 7.8, 7.6, 6.5	Health (NRR assay) (–)	Beesley et al. (2008)
<i>M. edulis</i>	Adults (76 mm)	16 L (2?)	6 (0.4)	Flow	12.5 (11.8)	14 days	Yes (3)	pH _{NBS} 8.05, 7.89, 7.81, 7.70, 7.56, 7.08 $p\text{CO}_2$ 464, 661, 789, 1046, 1480, 4254 μatm	pHe (–) Haemolymph Mg^{2+} (0) Haemolymph Ca^{2+} (0)	Thomsen et al. (2010)
<i>M. galloprovincialis</i> Mediterranean mussel	Embryos	50 mL (6)	400 (<10 000)	Stagnant (55, 96 and 121 h)	13.0 (35.5)	6 days	No	pH _{NBS} 8.13, 7.42	Development rate from trochophores (–) Shell size (–) Normal shell shape (–)	Kurihara et al. (2008)
<i>Mytilus trossulus</i>	Larvae	22 mL (240)	72.6 (3300)	Stagnant	12 (34.3)	60 h	No	$p\text{CO}_2$ 253, 571 pH _{NBS} 8.3, 7.9	Shell size (–)	Sunday et al. (2011)
<i>Perma viridis</i> green-lipped mussel	Juveniles (40–44 mm)	10 L (5)	4 (0.4)	Stagnant (daily)	25 (32)	5 days	Yes (1)	pH _{NBS} 8.1, 7.7, 7.4	Clearance (–) Respiration rate (0) Excretion (–) at pH 7.4	Liu and He (2012)
<i>Pinctada fucata</i> pearl oyster	Juveniles (40–44 mm)	10 L (5)	4 (0.4)	Stagnant (daily)	25 (32)	5 days	Yes (1)	pH _{NBS} 8.1, 7.7, 7.4	Clearance (–) at pH 7.4 Respiration rate (0) Excretion (–)	Liu and He (2012)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L^{-1})	Flow or stagnant (water exchange)	Temp. ($^{\circ}\text{C}$ (Salinity))	Duration	Food supply (spp)	OA Treatment	Effects:	Reference
<i>Ruditapes decussatus</i>	Juveniles (10.2 mm)	15 L (3)	100 (6.7)	Flow (3–4 times day^{-1})	22–23 (31–32)	75 days	Yes (2)	pH 8.25, 7.85, 7.67	Net calcification (0) Size (0) Weight (0) may be due to natural elevated total alkalinity Survival (+) may be due to spawning at pH 8.25, 7.85	Range et al. (2011)
<i>Saccostrea glomerata</i>	Gametes	100 mL (3)	500 (5000)	Stagnant	18, 22, 26, 30 (35)	48 h	No	$p\text{CO}_2$ 375, 600, 750, 1000 μatm Temperature Fertilization	D-veliger % (–) Abnormality % (+) Size (–) All less at optimum temperatures, more when fertilized at elevated $p\text{CO}_2$	Parker et al. (2010)
Sydney rock oyster										
<i>S. glomerata</i>	Gametes	200 L (3)	3 million (15 000) After 12 h: 1 million (5000)	Stagnant (every 2 days)	24 (34, 6)	19 days	Yes (3)	pHNSB 8.2, 7.9 $p\text{CO}_2$ 380, 856 μatm Parental OA exposure, and origin (wild and selected)	Survival (–), also (–) for wild vs. breed parental origin, no interaction with OA, (0) for parental exposure Development rate to veliger (–), (+) for parental exposure, and (+) for bred origin Development rate to eyed (–), interaction parental exposure – origin – OA Shell length (–), (0) for bred and exposed parents, (+) for exposed parents	Parker et al. (2012)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L^{-1})	Flow or stagnant (water exchange)	Temp. °C (Salinity)	Duration	Food supply (spp)	OA Treatment	Effects:	Reference
<i>S. glomerata</i>	D-veligers (24 h)	60 L (1)	90 000 (1500)	Flow	26	8 days	Yes (2)	pH: 8.1, 7.8, 7.6 $p\text{CO}_2$ 580, 1386, 3573	Negative (-) None (0) Positive (+) Survival (-) Growth (-) Shell morphology (-) Amount of empty shells (-) Shell surface characteristics (-) Larval dry mass (0)	Watson et al. (2009)
<i>S. glomerata</i>	Umbonate (142 μm)	100 mL (2500)	250 (2500)	Stagnant (daily)	18, 22, 26, 30 (35)	4 days	Yes (3)	$p\text{CO}_2$ 375, 600, 750, 1000 μatm Temperature	Size (-), not at 18 °C	Parker et al. (2010)
<i>S. glomerata</i>	Pediveliger (264 μm)	100 mL (2000)	200 (2000)	Stagnant (daily)	18, 22, 26, 30 (35)	4 days	Yes (3)	$p\text{CO}_2$ 375, 600, 750, 1000 μatm Temperature	Size (-), not at 18 and 26 °C	Parker et al. (2010)
<i>S. glomerata</i>	Juveniles (2.0 mm)	100 mL (150)	15 (150)	Stagnant (daily)	18, 22, 26, 30 (35)	4 days	Yes (3)	$p\text{CO}_2$ 375, 600, 750, 1000 μatm Temperature	Shell growth (-), more so at 18 and 22 °C	Parker et al. (2010)
<i>S. glomerata</i>	Post-larvae	120 mL (3)	30 (250)	Stagnant (daily)	25 (35)	4 days	Yes (3)	$p\text{CO}_2$ 375, 1000 ppm* Wild and mass-selected populations (<i>msp</i>), and families within <i>msp</i>	Shell growth (-), <i>msp</i> were less affected, large variations between families	Parker et al. (2011)
spp from bivalvia	NG	38 L (6)	NG	Stagnant (filtration 600 L h^{-1} , 75 % every 14 days)	25	60 days	Yes (several)	$p\text{CO}_2$ 409, 606, 903, 2856 pHNBS 8.03, 7.85, 7.72, 7.31	Net calcification (-) for scallops, oysters, hard clams, soft clams	Ries et al. (2009)

Table 2. Continued.

Species	Life stage (age/size)	Volume (n)	Stocking density (ind L ⁻¹)	Flow or stagnant (water ex-change)	Temp. (°C (Salinity))	Duration	Food supply (spp)	OA Treatment	Effects: Negative (-) None (0) Positive (+)	Reference
Others										
<i>Dendaxster ex-centricus</i> sand dollars (sea urchin)	Gastrulae	3.7 L (3)	7400 (2000)	Stagnant (every other day)	20		Yes (2)	CO ₂ 380, 1000 ppm (pHNBS 8.06, 7.75)	Stomach size (-) Body size (-) Swimming speed (0) between-family variation	Chan et al. (2011)
<i>Echthometra viridis</i> tropical urchin	Juveniles (dw 1.2–6.1 g)	34 L	6–8	Stagnant (filtered at 757 L h ⁻¹)	20.4, 29.9 (32)	60 days	Yes (NG)	CO ₂ 524, 827 ppm at 20.4 °C CO ₂ 448, 783 ppm at 30.0 °C	Survival (0) Calcification rate (-), (+) by temperature	Courtney et al. (2013)
<i>Madracis aur-retentra</i> coral	Branches	250 mL (3)	4*	Stagnant	28 (36)	2 h	No	pH _T 8.06, 7.78, 7.60, CO ₃ ²⁻ 105, 150, 260 μmol kg ⁻¹	Calcification (-) at pH 8.06 if HCO ₃ ⁻ < 1800 1M; (0) or (+) at pH 7.6–7.8 if HCO ₃ ⁻ > 1800 1M	Jury et al. (2010)
<i>Triplonastes garutilla</i> sea urchin	Embryo (0 d)	100 mL (3)	100–200 (2000)	Flow (7.8 mL min ⁻¹)	24, 27, 30	5 days	No	pHNBS 8.15, 7.85, 7.65 Temperature	Growth (-) Calcification (-)	Brennand et al. (2010)
spp from: Crustacea Chitardia Echinoidea Rhodophyta Chlorophyta Gastropoda Amelida	NG	38 L (6)	NG	Stagnant (filtration 600 L h ⁻¹ , 75 % every 14 days)	25	60 days	Yes (several)	pCO ₂ 409, 606, 903, 2856 pHNBS 8.03, 7.85, 7.72, 7.31	Net calcification (-) for: corals, urchins, whelks, serpulid worms, periwinkles (+) for: limpets, purple urchins, coralline red algae, calcareous green algae	Ries et al. (2009)
<i>Strongylocentrotus franciscanus</i> Larvae	NG	22 mL (1000)	39.6 (1800)	Stagnant	12 (34.3)	7 days	No	pCO ₂ 253, 571 pHNBS 8.3, 7.9	Larvae size (-)	Sunday et al. (2011)

4.3 Inter- and intra-species response and synergistic factors

In the present study scallop larvae showed a reduction in shell size, survival, and percentage of normally developed larvae from the ambient group to the highest $p\text{CO}_2$ group (discussed in further detail in the sections below). These responses to OA seem to vary little between bivalve species (Table 2). However, the magnitude of the response vary, although the pH window, temperature, and rearing volume used during experiments vary, and one should be careful in drawing conclusions about species-specific effects between studies. Comparative studies of two (Sunday et al., 2011; Parker et al., 2010; Talmage and Gobler, 2010) or several species (Liu and He, 2012; Ries et al., 2009; Talmage and Gobler, 2009, 2011) mostly conclude that species have different sensitivity to OA. Within-species and between-population variation was reported by Parker et al. (2011) when they investigated differences between a wild population and selectively bred populations in sensitivity to OA exposure. Their results indicated that sensitivity to rearing conditions may have been different between the populations, as growth rate in the ambient group was higher for the selected group. This may have overestimated the differences in sensitivity to OA exposure alone.

Many factors may contribute to a synergistic response of a species to OA exposure (Table 2). Some have been investigated for early life stages of bivalves: food availability (Thomsen et al., 2010, 2013; Melzner et al., 2011), temperature (Brennand et al., 2010; Courtney et al., 2013; Parker et al., 2010; Talmage and Gobler, 2011), salinity (Dickinson et al., 2012), parental exposure to OA (Parker et al., 2012), fertilization under OA conditions (Parker et al., 2010), and genetic selection (populations or families, Parker et al., 2011, 2012). Additionally, there are other external factors that have not yet been investigated that may impact OA effects through reducing or improving the energy level: rearing volume, diet, day length, UV radiation, etc.

A large part of the work undertaken to investigate effects of OA on marine mollusc larvae has been carried out in volumes of 50 mL–200 L (Table 2). Wernberg et al. (2012) points out that it may be problematic that so many experimental studies of ocean acidification effects are undertaken in relatively small volumes. Although the larvae are small in size, small rearing volumes may not give a sufficiently stable rearing environment, and result in negative environmental stress. The volume of exposure tanks in the present study was 38 L, and we cannot rule out that these tanks are too small to keep variations in the rearing environment at a low level. However, the low variation in survival in the control group indicates that the experimental conditions were within acceptable limits.

None of the studies conducted on bivalve larvae are directly comparable with the present study (Table 2). When $p\text{CO}_2$ was relatively similar, the experiments were con-

ducted in stagnant systems with different rearing volumes and densities, or the larvae were fed. As discussed above, all these factors may be synergistic. Larvae in our experiment were reared from eggs of one individual and sperm from three different individuals (i.e. three half-sibling groups) after being well mixed and distributed in equal densities in the experimental tanks. We assume that the half-sibling groups were distributed more or less equally in the tanks, and the variation we found between tanks is most likely not a result of family differences (intra-species differences).

4.4 Larvae survival

The survival of day-7 veligers for the ambient group replicates in this study was 25–46%. In comparison, the average survival of day-3 veligers (referred to as hatching) in a Norwegian hatchery was 36.7% of 12 broodstock groups (Magnesen et al., 2006). This indicates that the larval group used in the present study was viable and healthy. In the treatment groups, median survival was reduced by 41–74%, which was much higher compared to larvae of hard clam (*M. mercenaria*), bay scallop (*A. irradians*), and eastern oyster (*C. virginica*) (Talmage and Gobler, 2009). The highest reduction in survival on day 6 for these larvae was 20% for the oyster larvae at 64 and 150 Pa CO_2 (pH 7.83–7.85 and pH 7.48–7.50, respectively) when ambient conditions were 36 Pa (pH 8.02–8.08). A study on abalone larvae (*Haliotis kamtschatkana*) showed that both survival and growth were reduced by 40% on day 8, when CO_2 was elevated from an ambient 400 ppm up to 800 ppm (Crim et al., 2011). These variations in survival may indicate inter-species differences, but we cannot rule out that the variation at least partly may be a result of the differences in experimental design (culture volume, water regime, feeding regime).

In an OA study on the Pacific oyster, *C. gigas*, Gazeau et al. (2011) manipulated the A_T levels in one of their treatment groups and showed that the saturation level of CO_3^{2-} had a more severe effect on development and shell growth in oyster embryos than did the pH or $p\text{CO}_2$ level. However, none of the other OA experiments on bivalve larvae have manipulated the A_T level, and thus elevated $p\text{CO}_2$ levels seem to be the most plausible cause of reduced survival in these studies. Although survival of bivalve larvae in OA studies shows a significant variation, a clear trend of a negative effect due to elevated $p\text{CO}_2$ levels or lower pH is found (Table 2). Of six studies calculating survival on larvae, all reported a negative effect on survival (Table 2).

4.5 Larvae shell size

On day 7 larvae size (shell length and shell height) was smaller in the treatment groups compared to the group held in ambient water, but no difference was found between the two mid-treatment groups (807 and 1164 μatm). Reduced shell growth is in accordance with previous stud-

ies (Table 2). Gaylord et al. (2011) found that when CO_2 concentration increased from ambient 380 to 970 ppm, the shell area of mussel larvae (*Mytilus californianus*) on day 5 and day 7 was reduced by 7 and 5 %, respectively. Kurihara et al. (2007) showed the shell length of oyster (*C. gigas*) larvae after 48 h to be ca. 20 % smaller at a $p\text{CO}_2$ of 2268 μatm (pH 7.42) than at the ambient 348 μatm (pH 8.21). Kurihara et al. (2008) also found that the shell length of *Mytilus galloprovincialis* larvae was reduced by about 20 % after six days when pH was reduced from the ambient 8.05 (14.3 $\mu\text{mol CO}_2 \text{ kg}^{-1}$) to 7.43 (83.3 $\mu\text{mol CO}_2 \text{ kg}^{-1}$). Parker et al. (2010) found a synergistic effect of elevated $p\text{CO}_2$ and suboptimal temperatures on larvae size of the two oyster species *Saccostrea glomerata* and *C. gigas*, the first being the most sensitive to the exposure.

The reduction in shell growth found in the present study was at the lower end of the scale compared to other studies. This may reflect an inter-species difference, and indicates that *P. maximus* shell growth is not the most sensitive to OA compared to other bivalves. However, the high reduction in survival may indicate that other physiological processes are highly affected by OA. The duration of the study was only seven days, and keeping the scallop larvae in the same conditions throughout the larval cycle will allow for elucidation of how OA affects growth, survival, and the critical metamorphosis over a longer term.

4.6 Shell deformities

In the present study the percentage of normally developed larvae on day 7 was 64 % in the control group. The common ratio of normal shell development in *P. maximus* larvae has been poorly described and is not known. In a control group of oyster (*C. gigas*) larvae, Kurihara et al. (2007) described that 68 % developed into normal D-shaped veligers, and 72 % of these larvae had fully mineralized shells after 48 h. This was explained as rather low values for the control group, and was supposed to be a result of using gonadal stripping to produce the embryos. In a study on *M. galloprovincialis*, Kurihara et al. (2008) reported abnormal morphology in < 1 % of control larvae, showing that the variation in percentage of normally developed larvae may vary between species, although other conditions may play a significant role.

The percentage of deformed larvae varied highly between the treatments in our study, and the results showed a tendency of more shell deformities with increasing $p\text{CO}_2$ levels. The negative effects observed have been reported also in other OA studies on bivalves (Table 2). Watson et al. (2009) found that in 8-day-old *S. glomerata* larvae, both shell morphology and shell surface characteristics were negatively affected by reduced pH (7.8 and 7.6). Talmage and Gobler (2010) found that the level of CO_2 strongly affected the development of the hinge structure in larvae of *M. mercenaria* and *A. irradians* using CO_2 levels of 250 (pre-industrial), 390 (ambient), 750, and 1500 ppm. They found a decline in size, integrity,

and connectedness of the hinge with increasing CO_2 levels. As the authors pointed out, the hinge facilitates opening and closing of the shell, which again is crucial for both feeding and excretion. Losing the ability to perform these basic functions most certainly has a negative effect on larval survival. In the present study, the higher percentage of normally developed hinges on day 7 compared to day 2 in all larvae groups was most likely a result of increased mortality of abnormal larvae from day 2 to 7. The final sample at the end of the experiment did not show a high amount of deformed empty shells to confirm this assumption. However, empty larvae shells are extremely fragile and dissolve fast (and even faster in low-pH water), which may explain why we did not find high numbers of them in the final samples. In future experiments, more frequent sampling may reveal whether the deformed larvae die off throughout the experimental period.

It is known that other factors like self-fertilization, high egg : sperm ratio, temperature, and pollutants may cause an increase in larval deformities (Conroy et al., 1996; Spangenberg and Cherr, 1996; His et al., 1997; Krassoi et al., 1997, Concha et al., 2011; Kobayashi and Kijima, 2010). It has also been reported that factors in the rearing environment such as small rearing volumes, stagnant systems, addition of antibiotics, or pulse feeding may cause larvae deformities in control groups (Kurihara et al., 2007, 2008; Talmage and Gobler, 2009, 2010, 2011; Waldbusser et al., 2010; Gaylord et al., 2011; Gazeau et al., 2011). Although we cannot rule out the effect of small rearing volumes, other factors should play a minor role in the present study as we cross-fertilized the eggs to minimize the amount of self-fertilization, egg : sperm ratio was within the recommended numbers (Martinez et al., 2007; S. Andersen, personal communication, 2012), temperature was within the optimum range (Davenport et al., 1975), we used a flow-to-waste system, and no antibiotics or feed were added.

The ambient water used for the control group in the present study had a pH of 7.98, which was 0.19 units lower than the pH values Talmage and Gobler (2010) used when mimicking the pre-industrial level to investigate past, present and future ocean CO_2 concentrations. They found improved results at pre-industrial pH levels compared to ambient pH levels of 8.05 and 8.04. The shell deformities we found in our control group may therefore at least partly be a result of an already elevated CO_2 level; thus future studies on *P. maximus* should include experimental groups kept in pre-industrial CO_2 conditions to elucidate whether the ongoing OA already may have an impact on scallop larvae.

For the seven OA studies looking at the normal development of bivalve embryos and larvae (a total of seven different species), all reported a negative effect of elevated CO_2 levels (Table 2). Although the results in our study showed a high variation in shell edge deformities between treatment groups, the percentage of hinge deformities increased with increasing $p\text{CO}_2$ levels. As the determination of shell edge deformities can be a bit challenging due to damage caused

by handling the samples, calculating the amount of hinge deformity is probably a better estimate of abnormalities in our study.

4.7 Calcification

In the present study we confirmed that calcein bonded to the calcium in 1–2-day-old trochophore larvae, confirming similar timing for calcification in *P. maximus* larvae as in larvae of *M. mercenaria* and *C. gigas* (Weiss et al., 2002), but at an earlier stage than was suggested by Cass et al. (1998) and Cragg (2006). We were unable to determine the exact time of initiation of calcification in the larvae since the staining period was 18–20 h. Staining embryos and larvae frequently (e.g. every 5 h) for a shorter period (e.g. 3 h) from day 0 to day 2 may reveal more details on the exact onset of the calcification process. Comeau et al. (2009) used calcein staining of the pteropod *Limacina helicina* to compare shell extension in animals held at $p\text{CO}_2$ levels of 350 and 760 μatm over a period of five days. Although too few organisms were used to get statistically robust quantitative estimations, the results showed that shell extension in pteropods held at the highest $p\text{CO}_2$ levels were lower compared to those held at 350 μatm . Future OA experiments on *P. maximus* larvae should elucidate whether calcein staining can be used to calculate calcification either as shell extension or size of shell area in larvae held at different $p\text{CO}_2$ levels.

4.8 Concluding remarks and future work

The rather severe effects of elevated future $p\text{CO}_2$ levels on a number of bivalve larvae seem undisputable (Table 2). Talmage and Gobler (2010) showed that the CO_2 levels of today may already affect bivalve larvae negatively compared to pre-industrial levels, and hence, negative effects of OA are happening today rather than in the future. Also, OA should be seen in combination with the increase in global ocean temperature (Blunden and Arndt, 2013). Effects of a temperature increase may amplify the OA effects or reduce them. Talmage and Gobler (2011) showed that the negative effects of elevated CO_2 and temperature on larvae were additive.

Long-term studies on larvae of the great scallop *P. maximus* are needed to elucidate the effect of OA throughout the larvae stages, metamorphosis, and settlement, and also the potential of an evolutionary response in larval development rate. Such studies should involve assessment of the energy budget, take place in facilities that allow a normal development, and include studies on temperature effects. Effects on important processes for survival, such as immune defence functions, feeding rate, and shell mass and strength should also be investigated. More studies using calcein at the earliest larvae stages (day 0–2) are needed to increase knowledge about the onset of the larval shell calcification process, which is crucial to the existence of calcifying organisms.

Supplementary material related to this article is available online at <http://www.biogeosciences.net/10/6161/2013/bg-10-6161-2013-supplement.pdf>.

Acknowledgements. We thank Ørjan Karlsen for helpful assistance with statistics, and the staff at the Institute of Marine Research – Austevoll Research Station for rearing broodstock, producing algal feed, mixing chemical solutions, and assisting in photographic work. This study was supported by the Institute of Marine Research through project number 13192-04, Ocean Acidification – Scallops.

Edited by: D. Schmidt

References

- Andersen, S., Christophersen, G., and Magnesen, T.: Spat production of the great scallop (*Pecten maximus*): a roller coaster, *Can. J. Zool.*, 89, 585–604, doi:10.1139/z11-035, 2011.
- Andersen, S., Christophersen, G., and Magnesen, T.: Implications of larval diet concentration on post-larval yield in a production scale flow through system for scallops (*Pecten maximus* Lamarck) in Norway, *Aquacult. Int.*, 15, 1–18, doi:10.1007/s10499-012-9570-0, 2012.
- Barros, P., Sobral, P., Range, P., Chícharo, L., and Matias, D.: Effects of sea-water acidification on fertilization and larval development of the oyster *Crassostrea gigas*, *J. Exp. Mar. Biol. Ecol.*, 440, 200–206, doi:10.1016/j.jembe.2012.12.014, 2013.
- Bechmann, R. K., Taban, I. C., Westerlund, S., Godal B. F., Arnberg, M., Vingen, S., Ingvarsdottir, A., and Baussant, T.: Effects of ocean acidification on early life stages of shrimp (*Pandalus borealis*) and mussel (*Mytilus edulis*), *J. Toxicol. Env. Heal. A*, 74424–74438, doi:10.1080/15287394.2011.550460, 2011.
- Beesley, A., Lowe, D. M., Pascoe, C. K., and Widdicombe, S.: Effects of CO_2 -induced seawater acidification on the health of *Mytilus edulis*, *Clim. Res.*, 37, 215–225, doi:10.3354/cr00765, 2008.
- Bellioli, G., Lohrman, K., and Dupré, E.: Larval morphology of the scallop *Argopecten purpuratus* as revealed by scanning electron microscopy, *The Veliger*, 36, 332–342, 1993.
- Beniash, E., Ivanina, A., Lieb, N. S., Kurochkin, I., and Sokolova, I. M.: Elevated level of carbon dioxide affects metabolism and shell formation in oysters *Crassostrea virginica*, *Mar. Ecol.-Prog. Ser.*, 419, 95–108, doi:10.3354/meps08841, 2010.
- Blackford, J. C. and Gilbert, F. J.: pH variability and CO_2 induced acidification in the North Sea, *J. Marine Syst.*, 64, 229–241, doi:10.1016/j.jmarsys.2006.03.016, 2007.
- Blunden, J. and Arndt, D. S.: State of the climate in 2012, *B. Am. Meteorol. Soc.*, 94, S1–S258, 2013.
- Bourne, N. F.: The potential for scallop culture – the next millennium, *Aquacult. Int.*, 8, 113–122, 2000.
- Brennand H. S., Soars, N., Dworjanyn, S. A., Davis, A. R., and Byrne, M.: Impact of Ocean Warming and Ocean Acidification on Larval Development and Calcification in the Sea Urchin *Triploneustes gratilla*, *PLoS ONE* 5, e11372, doi:10.1371/journal.pone.0011372, 2010.

- Byrne, M.: Global change ecotoxicology: Identification of early life history bottlenecks in marine invertebrates, variable species responses and variable experimental approaches, *Mar. Environ. Res.*, 76, 3–15, doi:10.1016/j.marenvres.2011.10.004, 2012.
- Caldeira, K. and Wickett, M. E.: Anthropogenic carbon and ocean pH, *Nature*, 425, 365, 2003.
- Casse, N., Devauchelle, N., and Le Pennec, M.: Embryonic shell formation in the scallop *Pecten maximus* (Linnaeus), *Veliger*, 41, 133–141, 1998.
- Chan, K. Y., Grünbaum, K. D., and O'Donnell, M.: Effects of ocean-acidification-induced morphological changes on larval swimming and feeding, *J. Exp. Biol.*, 214, 3857–3867, doi:10.1242/jeb.054809, 2011.
- Chaparro, O. R., Montory, J. A., Segura, C. J., and Pechenik, J. A.: Effect of reduced pH on shells of brooded veligers in the estuarine bivalve *Ostrea chilensis* (Philippi, 1845), *J. Exp. Mar. Biol. Ecol.*, 377, 107–112, doi:10.1016/j.jembe, 2009.
- Chick, R. C.: Batch tagging blacklip abalone (*Haliotis rubra*) for identification of hatchery-reared individuals on natural coastal reefs in New South Wales, Australia, *J. Shellfish Res.*, 29, 209–215, 2010.
- Comeau, S., Gorsky, G., Jeffree, R., Teyssié, J.-L., and Gattuso, J.-P.: Impact of ocean acidification on a key Arctic pelagic mollusc (*Limacina helicina*), *Biogeosciences*, 6, 1877–1882, doi:10.5194/bg-6-1877-2009, 2009.
- Comeau, S., Gorsky, G., and Gattuso, J.-P.: Larvae of the pteropod *Cavolinia inflexa* exposed to aragonite undersaturation are viable but shell-less, *Mar. Biol.*, 157, 2341–2345, 2010.
- Comely, C. A.: Larval culture of the scallop, *Pecten maximus*, *J. Cons. Int. Explor. Mer.*, 34, 365–378, 1972.
- Concha, C., Figueroa, E., and Winkle, F. M.: Association between self-fertilization rates and the frequency of malformed larvae in farmed populations of the northern scallop *Argopecten purpuratus* (Lamarck, 1819), *Lat. Am. J. Aquat. Res.*, 39, 327–337, doi:10.3856/vol39-issue2-fulltext-14, 2011.
- Conroy, P. T., Hunt, J. W., and Anderson, B. S.: Validation of a short-term toxicity test endpoint by comparison with longer-term effects on larval red abalone *Haliotis rufescens*, *Env. Toxicol. Chem.*, 15, 1245–1250, doi:10.1897/1551-5028(1996)015<1245:voastt>2.3.co;2, 1996.
- Courtney, T., Westfield, I., and Ries, J. B.: CO_2 -induced ocean acidification impairs calcification in the tropical urchin *Echinometra viridis*, *J. Exp. Mar. Biol. Ecol.*, 440, 169–175, doi:10.1016/j.jembe.2012.11.013, 2013.
- Cragg, S. M.: Development, physiology, behavior, and ecology of scallop larvae, 45–122. In *Scallops: Biology, ecology, and aquaculture*, edited by: Shumway, S. E. and Parsons, G. J., Elsevier, 2006.
- Crim, R. N., Sunday, J. M., and Harley, C. D. G.: Elevated seawater CO_2 concentrations impair larval development and reduce larval survival in endangered northern abalone (*Haliotis kamtschatkana*), *J. Exp. Mar. Biol. Ecol.*, 400, 272–277, 2011.
- Day, R. W., Williams, M. C., and Hawkes, G. P.: A comparison of fluorochromes for marking abalone shells, *Mar. Freshwater Res.*, 46, 599–605, 1995.
- Davenport, J., Gruffydd, L. D., and Beaumont, A. R.: An apparatus to supply water of fluctuating salinity and its use in a study of the salinity tolerances of larvae of the scallop *Pecten maximus* L., *J. Mar. Biol. Assoc. UK*, 55, 391–409, 1975.
- Desrosiers, R. R., Désilets, J., and Dubé, F.: Early developmental events following fertilization in the giant scallop *Placopecten magellanicus*, *Can. J. Fish. Aquat. Sci.*, 53, 1382–1392, 1996.
- Dickinson, G. H., Ivanina, A. V., Matoo, O. B., Pörtner, H. O., Lannig, G., Bock, C., Beniash, E., and Sokolova, I. M.: Interactive effects of salinity and elevated CO_2 levels on juvenile eastern oysters, *Crassostrea virginica*, *J. Exp. Biol.*, 215, 29–43, doi:10.1242/jeb.061481, 2012.
- Dupont, S., Ortega-Martínez, O., and Thorndyke, M.: Impact of near-future ocean acidification on echinoderms, *Ecotoxicology*, 19, 449–462, doi:10.1007/s10646-010-0463-6, 2010.
- Ericson, J. A., Lamare, M. D., Morley, S. A., and Barker, M. F.: The response of two ecologically important Antarctic invertebrates (*Sterechnus neumayeri* and *Parborlasia corrugates*) to reduced seawater pH: effects on fertilization and embryonic development, *Mar. Biol.*, 157, 2689–2702, 2010.
- Fabry, V. J., Seibel, B. A., Feely, R. A., and Orr, J. C.: Impacts of ocean acidification on marine fauna and ecosystem processes, *ICES J. Mar. Sci.*, 65, 414–432, 2008.
- FAO: Food and Agriculture Organization of the United Nations, Fisheries and Aquaculture Information and Statistics Service, Rome, Italy, 2010.
- Fitzpatrick, M. P., Jeffs, A. G., and Dunphy, B. J.: Identification of the optimal fluorochrome for marking larvae of the pulmonate limpet *Siphonaria australis*, *J. Shellfish Res.*, 29, 941–944, doi:10.2983/035.029.0427, 2010.
- Gaylord, B., Hill, T. M., Sanford, E., Lenz, E. A., Jacobs, L. A., Sato, K. N., Russell, A. D., and Hettinger, A.: Functional impacts of ocean acidification in an ecologically critical foundation species, *J. Exp. Biol.*, 214, 2586–2594, doi:10.1242/jeb.055939, 2011.
- Gazeau, F., Gattuso, J.-P., Dawber, C., Pronker, A. E., Peene, F., Peene, J., Heip, C. H. R., and Middelburg, J. J.: Effect of ocean acidification on the early life stages of the blue mussel *Mytilus edulis*, *Biogeosciences*, 7, 2051–2060, doi:10.5194/bg-7-2051-2010, 2010.
- Gazeau, F., Gattuso, J.-P., Greaves, M., Elderfield, H., and Peene, J.: Effect of Carbonate Chemistry Alteration on the Early Embryonic Development of the Pacific Oyster (*Crassostrea gigas*), *PLoS ONE*, 6, e23010, doi:10.1371/journal.pone.0023010, 2011.
- Hall-Spencer, J. M., Rodolfo-Metalpa, R., Martin, S., Ransome, E., Fine, M., Turner, S. M., Rowley, S. J., Tedesco, D., and Buia, M.-C.: Volcanic carbon dioxide vents show ecosystem effects of ocean acidification, *Nature*, 454, 96–99, doi:10.1038/nature07051, 2008.
- His, E., Seaman, M. N. L., and Beiras, R.: A simplification the bivalve embryogenesis and larval development bioassay method for water quality assessment, *Water Res.*, 31, 351–355, 1997.
- Hoegh-Guldberg, O., Mumby, P. J., Hooten, A. J., Steneck, R. S., Greenfield, P., Gomez, E., Harvell, C. D., Sale, P. F., Edwards, A. J., Caldeira, K., Knowlton, N., Eakin, C. M., Iglesias-Prieto, R., Muthiga, N., Bradbury, R. H., Dubi, A., and Hatziolos, M. E.: Coral Reefs Under Rapid Climate Change and Ocean Acidification, *Science*, 318, 1737, doi:10.1126/science.1152509, 2007.
- IPCC: Climate Change 2001: The Scientific Basis, in: Contribution of Working Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change, edited by: Houghton, J. T., Ding, Y., Griggs, D. J., Noguera, M., van der Linden, P. J.,

- Dai, X., Maskell, K., and Johnson, C. A., Cambridge University Press, Cambridge, UK and New York, NY, USA, 881 pp., 2001.
- Jury, C. P., Whitehead, R. F., and Szmant, A. M.: Effects of variations in carbonate chemistry on the calcification rates of *Madracis auretenra* (= *Madracis mirabilis* sensu Wells, 1973): bicarbonate concentrations best predict calcification rates, *Glob. Change Biol.* 16, 1632–1644, doi:10.1111/j.1365-2486.2009.02057.x, 2010.
- Kerrison, P., Hall-Spencer, J. M., Suggett, D. J., Hepburn, L. J., and Steinke, M.: Assessment of pH variability at a coastal CO_2 vent for ocean acidification studies, *Estuar. Coast. Shelf S.*, 94, 129–137, doi:10.1016/j.eccs.2011.05.025, 2011.
- Kobayashi, T. and Kijima, A.: Effects of inbreeding depression in Pacific abalone *Haliotis discus hannai*, *J. Shellfish Res.*, 29, 643–649, 2010.
- Kudo, M., Kameda, J., Saruwatari, K., Ozaki, N., Okano, K., Nagasawa, H., and Kogure, T.: Microtexture of larval shell of oyster, *Crassostrea nippona*: A FIB-TEM study, *J. Struct. Biol.*, 169, 1–5, doi:10.1016/j.jsb.2009.07.014, 2010.
- Kurihara, H.: Effects of CO_2 -driven ocean acidification on the early developmental stages of invertebrates, *Mar. Ecol.-Prog. Ser.*, 373, 275–284, doi:10.3354/meps07802, 2008.
- Kurihara, H., Kato, S., and Ishimatsu, A.: Effects of increased seawater $p\text{CO}_2$ on early development of the oyster *Crassostrea gigas*, *Aquat. Biol.*, 1, 91–98, doi:10.3354/ab00009, 2007.
- Kurihara, H., Asai, T., Kato, S., and Ishimatsu, A.: Effects of elevated $p\text{CO}_2$ on early development in the mussel *Mytilus galloprovincialis*, *Aquat. Biol.*, 4, 225–233, doi:10.3354/ab00109, 2008.
- Körbl, J. and Vydra, F.: Metallochromic indicators. IV. A note on the preparation and properties of “calcein”, *Collect. Czech Chem. C.*, 23, 622–627, 1958.
- Krassoi, R., Anderson, I., and Everett, D.: Larval abnormalities in doughboy scallops *Chlamys (mimachlamys) asperrima* L. In response to test conditions and six reference toxicants, *Australasian Journal of Ecotoxicology*, 3, 65–74, 1997.
- Kroeker, K. J., Kordas, R. L., Crim, R. N. and Singht, G. G.: Meta-analysis reveals negative yet variable effects of ocean acidifications on marine organisms, *Ecol. Lett.*, 13, 1419–1434, 2010.
- Le Pennec, M., Paugam, A., and Le Pennec, G.: The pelagic life of pectinid *Pecten maximus* – a review, *ICES J. Mar. Sci.*, 60, 211–223, 2003.
- Liu, W. and He, M.: Effects of ocean acidification on the metabolic rates of three species of bivalve from southern coast of China, *Chin. J. Oceanol. Limn.*, 30, 206–211, doi:10.1007/s00343-012-1067-1, 2012.
- Lucas, T., Palmer, P. J., Wang, S., Scoones, R., and O'Brien, E.: marking the shell of the saucer scallop *Amusium balloti* for sea ranching using oxytetracycline, calcein and Alizarin red, *J. Shellfish Res.*, 27, 1183–1188, 2008.
- Mackie, G. L.: Bivalves, in: *The Mollusca*, 7, Reproduction, edited by: Trompa, A. S., Verdonk, N. H., and Van Den Biggelaar, J. A. M., Academic Press, San Diego, 351–403, 1984.
- Magnesen, T., Bergh, Ø., and Christophersen, G.: Yields of great scallop, *Pecten maximus*, larvae in a commercial flow-through system in Norway, *Aquacult. Int.*, 14, 377–394, 2006.
- Martinez, G., Mettifogo, L., Perez, M. A., and Callejas, C.: A method to eliminate self-fertilization in a simultaneous hermaphrodite scallop. 1. Effects on growth and survival of larvae and juveniles, *Aquaculture*, 273, 459–469, doi:10.1016/j.aquaculture.2007.10.033, 2007.
- Melzner, F., Stange, P., Trübenbach, K., Thomsen, J., Casties, I., Panknin, U., Gorb, S. N., and Gutowska, M. A.: Food Supply and Seawater $p\text{CO}_2$ Impact Calcification and Internal Shell Dissolution in the Blue Mussel *Mytilus edulis*, *PLoS ONE*, 6, e24223, doi:10.1371/journal.pone.0024223, 2011.
- Moran, A. L.: Calcein as a marker in experimental studies newly-hatched gastropods, *Mar. Biol.*, 137, 893–898, doi:10.1007/s002270000390, 2000.
- Moran, A. L. and Marko, P. B.: A simple technique for physical marking of larvae of marine bivalves, *J. Shellfish Res.*, 24, 567–571, 2005.
- Orr, J. C., Fabry, V. J., Aumont, O., Bopp, L., Doney, S. C., Feely, R. A., Gnanadesikan, A., Gruber, N., Ishida, A., Joos, F., Key, R. M., Lindsay, K., Maier-Reimer, E., Matear, R., Monfray, P., Mouchet, A., Najjar, R. G., Plattner, G.-K., Rodgers, K. B., Sabine, C. L., Sarmiento, J. L., Schlitzer, R., Slater, R. D., Totterdell, I. J., Weirig, M.-F., Yamanaka, Y., and Yool, A.: Anthropogenic ocean acidification over the twenty-first century and its impact on calcifying organisms, *Nature*, 437, 681–686, doi:10.1038/nature04095, 2005.
- Pane, E. F. and Barry, J. P.: Extracellular acid-base regulation during short-term hypercapnia is effective in a shallow-water crab, but ineffective in a deep-sea crab, *Mar. Ecol.-Prog. Ser.*, 334, 1–9, 2007.
- Parker, L. M., Ross, P. M., and O'Connor, W. A.: Comparing the effect of elevated $p\text{CO}_2$ and temperature on the fertilization and early development of two species of oysters, *Mar. Biol.*, 157, 2435–2452, doi:10.1007/s00227-010-1508-3, 2010.
- Parker, L. M., Ross, P. M., and O'Connor, W. A.: Populations of the Sydney rock oyster, *Saccostrea glomerata*, vary in response to ocean acidification, *Mar. Biol.*, 158, 689–697, doi:10.1007/s00227-010-1592-4, 2011.
- Parker, L. M., Ross, P. M., O'Connor, W. A., Borysko, L., Raftos, D. A., and Pörtner, H.-O.: Adult exposure influences offspring response to ocean acidification in oysters, *Glob. Change Biol.*, 18, 82–92, doi:10.1111/j.1365-2486.2011.02520.x, 2012.
- Pierrot, D. E., Lewis, D. W., and Wallace, R.: MS Excel Program Developed for CO_2 System Calculations, ORNL/CDIAC-105a. Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, US Department of Energy, Oak Ridge, Tennessee, doi:10.3334/CDIAC/otg.CO2SYS_XLS_CDIAC105a, 2006.
- Range, P., Chícharo, M. A., Ben-Hamadou, R., Piló, D., Matias, D., Joaquim, S., Oliveira, A. P., and Chíchar, L.: Calcification, growth and mortality of juvenile clams *Ruditapes decussatus* under increased $p\text{CO}_2$ and reduced pH: Variable responses to ocean acidification at local scales?, *J. Exp. Mar. Biol. Ecol.*, 396, 177–184, doi:10.1016/j.jembe.2010.10.020, 2011.
- Raven, J., Caldeira, K., Elderfield, H., Hoegh-Guldberg, O., Liss, P., Riebesell, U., Sheperd, J., Turley C., and Watson, A.: Ocean acidification due to increasing atmospheric carbon dioxide, *The Royal Society*. 2005.
- Ries, J. B., Cohen, A. L., and McCorkle, D. C.: Marine calcifiers exhibit mixed responses to CO_2 -induced ocean acidification, *Geology*, 37, 1131–1134, doi:10.1130/G30210A.1, 2009.
- Schalkhauser, B., Bock, C., Stemmer, K., Brey, T., Pörtner, H.-O., and Lanning, G.: Impact of ocean acidification on escape perfor-

- mance of the king scallop, *Pecten maximus*, from Norway, *Mar. Biol.*, doi:10.1007/s00227-012-2057-8, 2012.
- Spangenberg, J. V. and Cherr, G. N.: Developmental effects of barium exposure in a marine bivalve (*Mytilus californianus*), *Environ. Toxicol. Chem.*, 15, 1769–1774, 1996.
- Strand, Ø. and Parson, G. J.: Scandinavia, 1067–1091, in: *Scallops: Biology, ecology, and aquaculture*, edited by: Shumway, S. E. and Parsons, G. J., Elsevier, 2006.
- Sunday, J. M., Crim, R. N., Harley, C. D. G., and Hart, M. W.: Quantifying rates of evolutionary adaptation in response to ocean acidification, *PLoS ONE*, 6, e22881, doi:10.1371/journal.pone.0022881, 2011.
- Talmage, S. C. and Gobler, C. J.: The effects of elevated carbon dioxide concentrations on the metamorphosis, size, and survival of larval hard clams (*Mercenaria mercenaria*), bay scallops (*Argopecten irradians*), and Eastern oysters (*Crassostrea virginica*), *Limnol. Oceanogr.*, 54, 2072–2080, 2009.
- Talmage, S. C. and Gobler, C. J.: Effects of past, present, and future ocean carbon dioxide concentrations on the growth and survival of larval shellfish, *P. Natl. Acad. Sci. USA*, 107, 17246–17251, 2010.
- Talmage, S. C. and Gobler, C. J.: Effects of elevated temperature and carbon dioxide on the growth and survival of larvae and juveniles of three species of Northwest Atlantic bivalves, *PLoS ONE*, 6, e26941, doi:10.1371/journal.pone.0026941, 2011.
- Thomsen, J., Gutowska, M. A., Saphörster, J., Heinemann, A., Trübenbach, K., Fietzke, J., Hiebenthal, C., Eisenhauer, A., Körtzinger, A., Wahl, M., and Melzner, F.: Calcifying invertebrates succeed in a naturally CO_2 -rich coastal habitat but are threatened by high levels of future acidification, *Biogeosciences*, 7, 3879–3891, doi:10.5194/bg-7-3879-2010, 2010.
- Thomsen, J., Casties, I., Pansch, C., Körtzinger, A., and Melzner, F.: Food availability outweighs ocean acidification effects in juvenile *Mytilus edulis*: laboratory and field experiments, *Glob. Change Biol.*, 19, 1017–1027, doi:10.1111/gcb.12109, 2013.
- Waldbusser, G. G., Bergschneider, H., and Green, M. A.: Size-dependent pH effect on calcification in post-larval hard clam *Mercenaria* spp., *Mar. Ecol.-Prog. Ser.*, 417, 171–18, doi:10.3354/meps08809, 2010.
- Watson, S. A., Southgate, P. C., Tyler, P. A., and Lloyd, P. S.: Early larval development of the Sydney rock oyster *Saccostrea glomerata* under near-future predictions of CO_2 -driven ocean acidification, *J. Shellfish Res.*, 28, 431–437, doi:10.2983/035.028.0302, 2009.
- Weiss, I. M., Tuross, N., Addadi, L., and Weiner, S.: Mollusc larval shell formation: Amorphous calcium carbonate is a precursor phase for Aragonite, *J. Exp. Zool.*, 293, 478–491, 2002.
- Welladsen, H. M., Southgate, P. C., and Heimann, K.: The effects of exposure to near-future levels of ocean acidification on shell characteristics of *Pinctada fucata* (Bivalvia: Pteriidae), *Molluscan Res.*, 30, 125–130, 2010.
- Wernberg, T. D., Smale, A., and Thomsen, M. S.: A decade of climate change experiments on marine organisms: procedures, patterns and problems, *Glob. Change Biol.*, 18, 1491–1498, doi:10.1111/j.1365-2486.2012.02656.x, 2012.
- Zondervan, I., Zeebe, R. E., Rost, B., and Riebesell, U.: Decreasing marine biogenic calcification: A negative feedback on rising atmospheric $p\text{CO}_2$, *Global Biogeochem. Cy.*, 15, 507–516, 2001.