



Contribution to a bio-optical model for remote sensing of Lena River water

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Abstract. Bio-optical measurements and sampling were carried out in the delta of the Lena River (northern Siberia, Russia) between 26 June and 4 July 2011. The aim of this study was to determine the inherent optical properties of the Lena water, i.e., absorption, attenuation, and scattering coefficients, during the period of maximum runoff. This aimed to contribute to the development of a bio-optical model for use as the basis for optical remote sensing of coastal water of the Arctic. In this context the absorption by CDOM (colored dissolved organic matter) and particles, and the concentrations of total suspended matter, phytoplankton-pigments, and carbon were measured. CDOM was found to be the most dominant parameter affecting the optical properties of the river, with an absorption coefficient of $4.5\text{--}5\text{ m}^{-1}$ at 442 nm, which was almost four times higher than total particle absorption values at visible wavelength range. The wavelength-dependence of absorption of the different water constituents was characterized by determining the semi logarithmic spectral slope. Mean CDOM, and detritus slopes were 0.0149 nm^{-1} (standard deviation (stdev) = 0.0003, $n = 18$), and 0.0057 nm^{-1} (stdev = 0.0017, $n = 19$), respectively, values which are typical for water bodies with high concentrations of dissolved and particulate carbon. Mean chlorophyll *a* and total suspended matter were 1.8 mg m^{-3} (stdev = 0.734 $n = 18$) and 31.9 g m^{-3} (stdev = 19.94, $n = 27$), respectively. DOC (dissolved organic carbon) was in the range $8\text{--}10\text{ g m}^{-3}$ and the total particulate carbon (PC) in the range $0.25\text{--}1.5\text{ g m}^{-3}$. The light penetration depth (Secchi disc depth) was in the range 30–90 cm and was highly correlated with the suspended matter

concentration. The period of maximum river runoff in June was chosen to obtain bio-optical data when maximum water constituents are transported into the Laptev Sea. However, we are aware that more data from other seasons and other years need to be collected to establish a general bio-optical model of the Lena water and conclusively characterize the light climate with respect to primary production.

1 Introduction

The Lena River with its extensive delta is the pathway for transport of large amounts of organic and inorganic material, in diverse forms of carbon, from its huge catchment area of about 2 486 000 km² into the Arctic Ocean, particularly into the Laptev Sea (Kattner et al., 1999). It has an annual fresh water discharge of around 500 km³ (Dmitrenko et al., 2008; http://www.feow.org/ecoregion_details.php?eco=608). Of all major arctic rivers it dominates the supply of the Arctic Ocean with dissolved organic carbon, DOC, and colored dissolved organic matter, CDOM (Stedmon et al., 2011; <http://www.arcticgreativers.org/index.html>). Permafrost thawing and diminishing sea ice coverage are distinctive indicators of global warming in the Arctic (Walsh, 1991; Payette et al., 2004; Morison et al., 2012) and the input of carbon may increase with the thawing of Siberian permafrost due to global warming. Changes in hydrographic and meteorological conditions have been observed in the Lena Delta, which in turn influence the Arctic Ocean (Morison et al., 2012). Considerable water budget, geomorphological

and ecosystem changes are expected in the near future (McClelland et al., 2006; Dmitrenko et al., 2008; Morison et al., 2012). Currently, however, it is not so easy to predict the future of the Lena Delta and its output with the current lack of present, past and perhaps most importantly continuous measurement data. Gathering of relevant data is not always possible by in situ sampling in such inaccessible areas like the Lena Delta. Here remote sensing tools can be useful and contribute to the understanding of the current riverine conditions and facilitate a continuous monitoring program. Nevertheless, as the basis for optical remote sensing of water is the knowledge of its optical properties, it is imperative that the relationship between these and the concentrations of the water constituents is measured in situ. Such optical data, apart from underpinning remote sensing studies, are important for determining primary production.

Previous studies of optical properties in the Lena Delta region are mainly focused on the characterization and quantification of CDOM, POM (particulate organic matter), phytoplankton distribution, and primary production (Cauwet and Sidorov, 1996; Heiskanen and Keck, 1996; Sorokin and Sorokin, 1996; Kattner et al., 1999; Gueguen et al., 2005). Investigations in the Laptev Sea were mostly limited to the marine areas, and did not include the Lena River itself or its shallow coastal region of freshwater influence. Thus, the Lena River ecosystem is not well documented in terms of bio-optical properties.

According to these limited previous studies, the carbon budget of the Lena River varies a great deal. During the flood period TOC (total organic carbon) can be up to 14.4 g m^{-3} (June–July), while during the low water level period (November–April) this value is reported to be around 4 g m^{-3} , and the annual mean concentration is 10.2 g m^{-3} (Cauwet and Sidorov, 1996). Along the river the TOC concentration remains almost constant from Yakutsk (2000 km upstream) to the Laptev Sea (Lara et al., 1998). Turbidity of the water is higher during the flood period with concentrations of suspended matter dry weight of $50\text{--}70 \text{ g m}^{-3}$. The concentration of phosphate was found to be similar to other Arctic rivers but silicate and nitrate are three times higher (Cauwet and Sidorov, 1996).

The phytoplankton community of the Lena River is dominated by fresh water diatoms and picocyanobacteria (Heiskanen and Keck, 1996; Sorokin and Sorokin, 1996). Reported wet weights of the dominant groups and size-fractionated chlorophyll concentrations are given below (Table 1).

Bio-optical data of the Lena River are scarce. In some studies Secchi disk depths and light attenuation coefficients were measured or calculated (Sorokin and Sorokin, 1996). The DOM (dissolved organic matter) composition in the Lena is well documented (Lara et al., 1998; Kattner et al., 1999; Lobbes et al., 2000). Particulate matter of the Lena River is composed of a lithogenic fraction, which stems from erosion by the river, and of detritus, which is formed mainly by the debris of organisms. Phytoplankton contributes only a

Table 1. Summary of the phytoplankton composition, size fraction, and total production in the Lena River.

	Picocyanobacteria < 3 μm	Nanophytoplankton 3–10 μm	larger microalgae > 10 μm
¹ Fraction of Chl <i>a</i> in Lena	0.07	0.16	0.77
¹ Fraction of phae in Lena	0.16	0.21	0.63
² Wet weight mg m^{-3}	19	36	407
² Prim. prod. $\text{mg C m}^{-3} \text{ day}^{-1}$		50–126	
¹ Chl <i>a</i> concentration mg m^{-1}		1.5–4.5	

¹ Heiskanen and Keck (1996).

² Sorokin and Sorokin (1996).

few percent to the total particulate matter. The main sources of total suspended matter (TSM) are eroded land and the thawing permafrost (Rachold et al., 1996; Lobbes et al., 2000).

Similar studies in the Kara Sea about optical properties, remote sensing, and carbon content were carried out by Pozdnyakov et al. (2005), Hessen et al. (2010), and Korosov et al. (2011).

The goal of our study was to determine the optical properties of the Lena River water during the period of maximum river runoff, i.e., when most of the annual water load is transported into the Laptev Sea. The intention was to set up a preliminary bio-optical model, which describes the dominant components and which determines the optical properties, the ranges of their concentrations, their absorption and scattering spectra, and the relationships between concentrations and the optical coefficients. A bio-optical model is the basis for optical remote sensing and light climate related phytoplankton studies.

The selected investigation site was at the entrance of the Lena Delta, where the river water is divided up into different channels and flows via many arms into the coastal sea. The period and site of our investigation was selected to match the maximum concentrations of water constituents before they diffuse into the coastal sea. Unfortunately, the maximum water level was earlier in 2011 than predicted, so that we missed the extreme phase by 7–10 days, but which on the other hand allowed us to study both the end of the high-water period as well as the beginning of the summer period, thus yielding information for different situations.

2 Investigation site and methods

All measurements were carried out in the vicinity of Samoylov Island, which is located at the entrance of the delta (Fig. 1). Daily water samples were collected from the Lena River two times a day. The first sampling point was located at the Samoylov Island coast (Station 2 in Fig. 1 and Table 2), and the second sampling location (Station 1 in Fig. 1 and Table 2) was in the main stream of the Lena, close to the small rock island of Stolp. Samples were taken from a small boat.

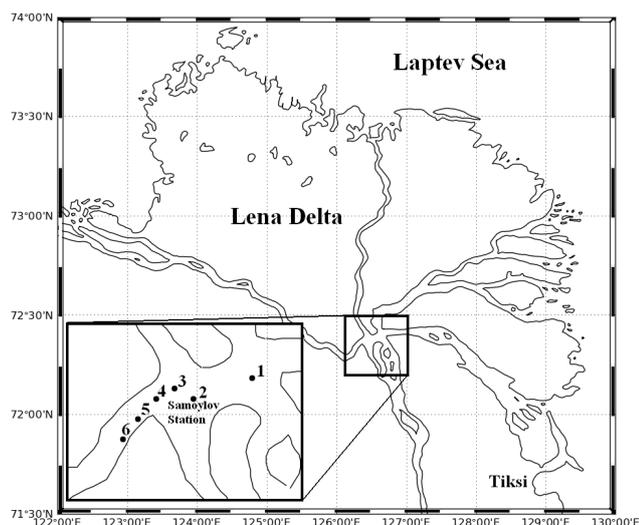


Fig. 1. Study area and location of sampling stations.

In addition to these positions another branch of the Lena River was sampled once (Stations 3–6 in Fig. 1 and Table 2). Samples were taken usually between 10:30 and 15:30 local time when the sun elevation was relatively high and overpasses of the ENVISAT satellite were expected. At the time of sampling we measured water and air temperatures, Secchi disk depths, reflectance spectra, phytoplankton fluorescence profiles, and GPS coordinates (Table 2).

Measurements of the in vivo fluorescence profiles were carried out with the Fluoroprobe Algae Analyser (BBE Moldaenke, Kiel, Germany, www.bbe-moldaenke.de). This instrument allows the differentiation of algal groups by using the different fluorescent spectra of them (Beutler et al., 2002).

All samples were processed at the Samoylov research station. Filters were prepared for measuring particle absorption, total suspended matter dry weight, and phytoplankton pigment concentrations by HPLC later in the home laboratory. Between 200 and 1500 mL of the water samples was filtered through pre-combusted GF/F filters and kept in a deep freezer (-18°C) until transfer in the freezer to Germany. These samples were analyzed as soon as received from Siberia 6–7 months later, to prevent or reduce further degradation of the biological material.

The absorption and scattering spectra of the fresh samples were measured before and after filtration at Samoylov using a Wetlabs hyperspectral absorption-attenuation meter (ACS). Using these measurements the absorption and scattering coefficients of particulate matter and the dissolved fraction (CDOM) could be determined.

3 Measurements and post processing

The ACS is designed for in situ measurements. It has two different cuvettes, one for measuring the beam attenuation coefficient c and the other for measuring the absorption coefficient a . To measure the samples at the station the instrument was installed in the lab and fixed upright to the bench leg. Silicon tubings were attached to the inlets of the cuvettes and the samples were poured from the lower inlet to avoid or minimize bubbles (schematic setup can be found in Sullivan et al., 2006). The absorption and beam attenuation spectra were recorded. Three replicates were measured for each sample together with a blank, which was recorded between the individual measurements to check the stability of the instrument. Temperature and scattering corrections were done by post processing of the raw data (Zaneveld et al., 1994; Sullivan et al., 2006; Boss et al., 2009). Blank values were not used for the calculations, because it was not possible to produce or store optically clean water at the station; instead previous blank values – measured before the campaign – were used for post processing of the raw data. However, this step was not critical, since blanks are less important for such highly absorbing water bodies.

The spectral shape of total absorption and that of the filtered water, which represents CDOM, were parameterized by fitting an exponential curve to the measured spectrum for the wavelength range 400–500 nm, which is the most important range for optical remote sensing of CDOM. This method was introduced by Jerlov in 1957. The spectral slope of the semi-logarithmic CDOM curve is variable and depends on the kind of matter being measured. Thus, the exponent s , which is the spectral slope on a semi-logarithmic plot, is an important parameter for the characterization of CDOM. The absorption at other wavelengths can then be computed using (Eq. 1):

$$a(\lambda) = a(\lambda_0) \exp\{-s(\lambda - \lambda_0)\}, \quad (1)$$

where a is absorption, λ is wavelength, λ_0 is the reference wavelength, and s is the spectral slope.

At each station the Secchi disk depth (SD) was determined. Due to the low transparency of Lena water and the high current speed a standard device with a marked cable could not be used. Instead a white disc was mounted at the tip of a pole of 2 m length, which was marked with a centimeter scale. This pole was deployed from the boat by hand.

The absorption of particulate matter was measured using the filter pad technique, which is a standard method to determine the absorption spectra of total suspended matter, detritus and phytoplankton (details in Mitchell et al., 2000). The main difficulty with the filter pad technique is the aggregating of particles in a small cross section, reasoning an artificial amplification of absorption. This amplification is corrected by an amplification factor. In this study we used the method of measurement and correction described in Röttgers and Gehnke (2012). Measurements were made in a dual-beam UV/VIS spectrophotometer (Lambda 950, Perkin Elmer) that

Table 2. Sampling protocol, complementary data, and filtered volume of the collected samples. (Station locations can be found in Fig. 1). n.a.: not available.

Location, Time and Station Information					Complementary Data			Filtered Volumes (mL)		
Date	Time (local)	Latitude	Longitude	Station Number	Secchi (cm)	Water Temp	Air Temp	TSM	HPLC	PABS
26 Jun 2011	10:00:00	72.37	126.48	2	70	n.a.	n.a.	500	750	550
26 Jun 2011	15:35:00	72.39	126.64	1	75	n.a.	n.a.	1000	n.a.	1000
27 Jun 2011	10:30:00	72.37	126.48	2	70	13		1000	1000	1000
27 Jun 2011	15:30:00	72.39	126.64	1	60	13.2	7.5	1000	1000	1000
28 Jun 2011	10:50:00	72.33	126.29	6	80	12.5	8.2	1000	1000	500
28 Jun 2011	11:20:00	72.35	126.33	5	85	12.3	9.5	1000	1000	500
28 Jun 2011	11:53:00	72.37	126.38	4	80	12.2	8.4	1000	1000	500
28 Jun 2011	12:10:00	72.38	126.43	3	75	12.2	8.6	1000	1000	500
29 Jun 2011	10:30:00	72.37	126.48	2	75	13.5	8.8	1500	n.a.	n.a.
29 Jun 2011	15:30:00	72.39	126.66	1	70	13.6	8.9	1500	1500	500
30 Jun 2011	10:30:00	72.37	126.48	2	n.a.	n.a.	n.a.	1500	1500	500
30 Jun 2011	15:30:00	72.39	126.64	1	n.a.	n.a.	n.a.	1500	1000	500
1 Jul 2011	10:30:00	72.37	126.48	2	70	13.4	7.3	1500	1000	500
1 Jul 2011	16:00:00	72.39	126.64	1	70	13	8.5	1500	1500	500
2 Jul 2011	10:30:00	72.37	126.48	2	70	13.3	8.4	1500	1000	500
2 Jul 2011	15:50:00	72.39	126.66	1	50	13.3	12	1000	1000	250
3 Jul 2011	19:30:00	72.37	126.46	2	35	13.7	17.8	700	700	200
3 Jul 2011	20:00:00	72.39	126.66	1	30	13.8	16	700	700	200
4 Jul 2011	09:00:00	72.37	126.47	2	40	14	18	1000	1000	250
4 Jul 2011	15:30:00	72.39	126.66	1	30	14.2	17.9	700	700	200

was equipped with a 150 mm integrating sphere (Labsphere Inc.). A piece of filter ($\sim 1 \times 2 \text{ cm}^2$) was cut out and placed in the middle of the integrating sphere. The filter was defrosted for 30–45 min before measurement. Measured filter pieces were then bleached for several minutes by sodium hypochlorite to destroy the phytoplankton pigments. Blanks were treated with the same procedure. The absorption by detritus, also called non-algal particulate matter, are determined from the bleached filters directly, while the difference between untreated and bleached filters provides the absorption spectra of phytoplankton pigments. The absorption spectra resembled the CDOM curves and were analyzed in the same way. The wavelength exponent of the absorption spectra of detritus was computed in the same way as CDOM by fitting an exponential curve to the measured spectrum with the same reference wavelength of 442 nm.

Samples for total particulate carbon (PC), which is the sum of inorganic (PIC) and organic (POC) particulate carbon, were taken daily by filtering 500–1000 mL of water onto pre-combusted and washed Whatman GF/F filters. The carbon content of the samples was measured with an Elementar Vario MICRO cube CHN analyser (Elementar Analysensysteme, www.elementar.de).

DOC was analyzed in the filtrates of the samples. The filtrate was acidified by the addition of 100 μL of HCl (25 %) before being sealed and stored in a dark fridge at 4 °C until analysis (Elementar Analysensysteme Vario TOC cube; www.elementar.de). All glassware used in the preparation of the DOC sample were previously washed in 10 % HCl and

rinsed with ultrapure water (Milli-Q Integral Water Purification System, Millipore).

The wavelength dependent total scattering coefficients b were obtained by subtracting total absorption (a of CDOM + water + particles) from the beam attenuation c with $b = c - a$ where c is the beam attenuation coefficient, a is the absorption and b is the scattering coefficient. By using b of particles only, the single scatter albedo $\omega_0 = b/c$ can be calculated, which provides information about the nature of particles. A high ω_0 indicates inorganic suspended matter with a relatively low absorption coefficient.

Phytoplankton pigments were determined using HPLC (high performance liquid chromatography) and the method developed by Zapata et al. (2000). A total of 29 different pigments, including chlorophyll and degradation products, could be separated from the Lena samples. Pigment concentrations were computed from the area of the concentration peaks relative to pigment standards (DHI and Sigma). Filters were extracted for 24 h at -40 °C with 100 % acetone (chromatography grade), then filtered through Whatman Spartan filters (13 mm \varnothing , 0.2 μm pore size) and transferred to the vials for chromatography.

TSM dry weight was determined by filtering the sampled water through pre-combusted (450 °C) and pre-weighted GF/F filters (Whatman). The nominal pore size of 0.7 μm of these filters defines the separation of dissolved and particulate matter for this study. Before weighting the filters were dried at 60 °C for 24 h and cooled to room temperature in a desiccator to prevent any change in weight due to humidity. The difference between the pre- and post-filter weights

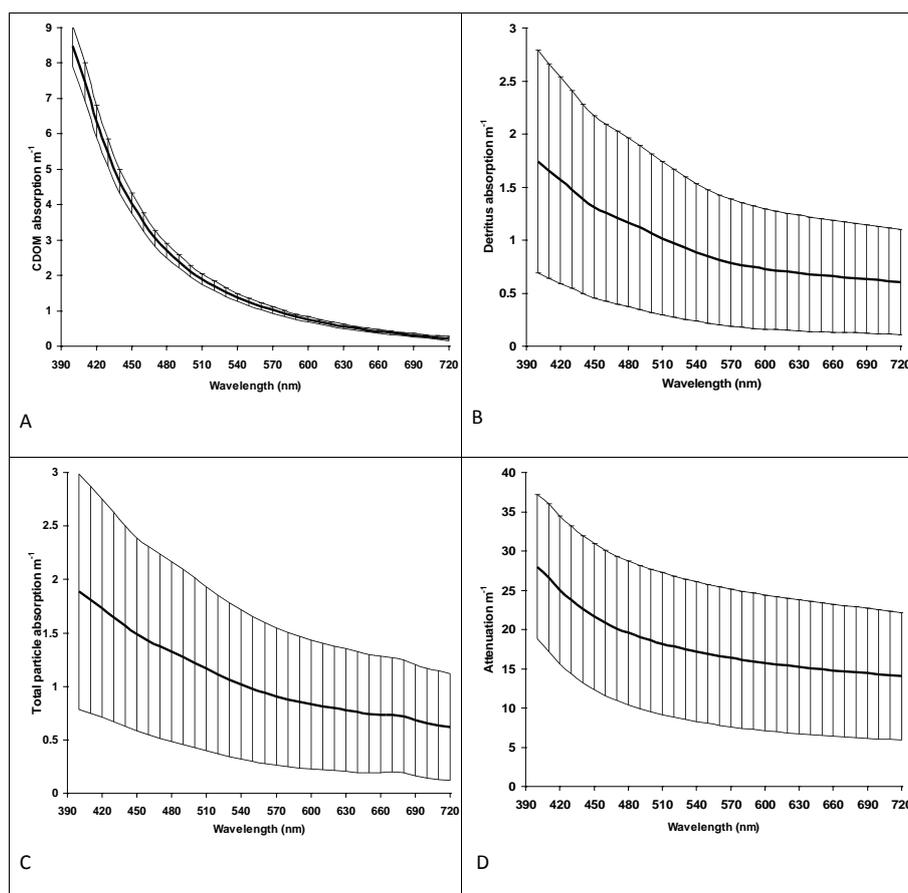


Fig. 2. Mean and standard deviations of the absorption coefficients of (A) CDOM, (B) particulate detritus, and (C) total particulate matter, and (D) of the beam attenuation of the samples.

divided by filtered sample volume is the dry weight of suspended matter per unit water volume.

4 Results

4.1 Absorption

Of all water constituents CDOM is the main source of absorption in the visible wavelength range in the Lena River. During the first week of our campaign CDOM absorption was almost uniform within the sampling area (Fig. 2a). At 442 nm it was in the range 4.25–5.25 m^{-1} , with a standard deviation of $\pm 0.36 \text{ m}^{-1}$. The decrease over time (Fig. 3) is not significant. Contrary to this slight decrease, an increase of the spectral slope was found (Fig. 4, Table 3). It varied between 0.0142 and 0.0158 nm^{-1} with a standard deviation of $\pm 0.0003 \text{ nm}^{-1}$.

The absorption coefficients of TSM and of the detrital fraction (after bleaching) are very close to each other (Fig. 2c and b, respectively). Thus, the difference between both, which is assumed to be the absorption by phytoplankton pigments, is low and noisy and thus not represented here. This is ex-

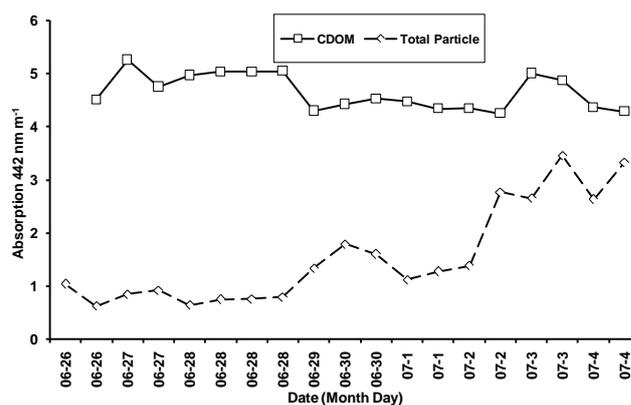


Fig. 3. CDOM and total particle absorption variability at 442 nm during the sampling period.

acerbated by the fact that the bleaching process does not only eliminate the absorption by pigments, but also partly of other organic matter. Another way to determine the absorption by chlorophyll is to calculate the difference between the absorption maximum of chlorophyll at 676 and 710 nm to

Table 3. Slopes of CDOM and detritus absorption and concentrations of total suspended matter mg L^{-1} , and chlorophyll a $\mu\text{g L}^{-1}$ (including all degradation products). Given below are the overall means and standard deviations.

Date	Time (local)	CDOM	Detritus	TSM	Chl_ a^*
26 Jun 2011	10:00:00	No Data	0.00622	15.06	0.920
26 Jun 2011	15:35:00	0.014323	0.006039	21.66	No Data
26 Jun 2011	10:30:00	0.014676	0.00527	19.51	2.184
27 Jun 2011	15:30:00	0.014781	0.004893	30.42	2.100
28 Jun 2011	10:00:00	No Data	No Data	15.43	No Data
28 Jun 2011	10:50:00	0.014859	0.008914	7.92	0.939
28 Jun 2011	11:20:00	0.014801	0.008649	9.36	1.610
28 Jun 2011	11:53:00	0.014711	0.00866	9.08	0.315
28 Jun 2011	12:10:00	0.014816	0.008331	8.45	2.306
29 Jun 2011	10:30:00	No Data	No Data	24.27	No Data
29 Jun 2011	15:30:00	0.014849	0.005714	27.25	2.043
30 Jun 2011	10:30:00	0.015026	0.004277	52.94	2.904
30 Jun 2011	15:30:00	0.014944	0.004771	37.34	2.995
1 Jul 2011	10:30:00	0.014958	0.005967	25.87	2.902
1 Jul 2011	16:00:00	0.015219	0.005387	28.63	1.930
2 Jul 2011	10:30:00	0.015159	0.005364	25.95	2.215
2 Jul 2011	15:50:00	0.015241	0.004244	54.77	1.888
3 Jul 2011	19:30:00	0.015369	0.004785	55.58	1.336
3 Jul 2011	20:00:00	0.01559	0.003884	58.81	1.380
4 Jul 2011	09:00:00	0.014992	0.003885	48.77	1.613
4 Jul 2011	15:30:00	0.015324	0.004056	67.59	1.192
Mean		0.01498	0.005753	31.89	1.820627
Stdev		0.0003	0.001688	19.94	0.733779

compensate for non-pigmented detritus absorption. For the Lena samples this reduces the uncertainty in the mass specific absorption coefficient from 60 to 34 %.

The absorption by TSM increased by a factor of three during the sampling period (Fig. 3, Table 3). The spectral slopes of detritus absorption are listed in Table 3. A decreasing trend of these slopes was observed over the sampling period. Exceptions were seen for stations 3–6, which were located in a tributary branching off from the main Lena stream, the “Kurungnakh arm”. All other samples were collected from the main stream and off Samoylov Island (Fig. 1). Slopes of the four Kurungnakh stations are almost twice those of the other samples (Fig. 5). Slopes in the main stream varied from 0.0038 to 0.0062 nm^{-1} , while Kurungnakh station slopes varied between 0.0083 and 0.0089 nm^{-1} .

4.2 Phytoplankton pigments

The major pigments, which we found in phytoplankton of the Lena River during our campaign, were chlorophyll a , fucoxanthin (fuco), and chlorophyll b (Chl b). Diadinoxanthin (diadino), alloxanthin (alloxan), zeaxanthin (zeaxan), and lutein were found in lower concentrations at all stations. Some other pigments such as chlorophyll $c1$, $c2$ (Chl $c1$, $c2$), neoxanthin (neoxan), and violoxanthin (viola) were not always found and when they were, only at lower concentrations (Fig. 6). Total chlorophyll was calculated by summing

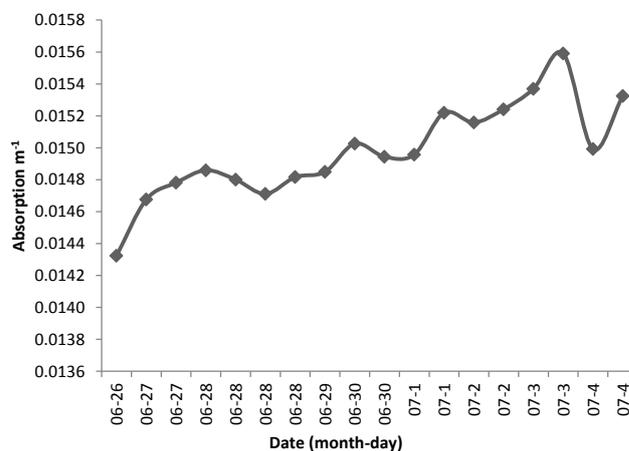


Fig. 4. CDOM semi-logarithmic slopes calculated between 400 and 500 nm with 442 nm as reference wavelength using a nonlinear fit.

up chlorophyll a and all of its degradation products. On average the degradation products added 33 % (min 15 %, max 50 %) to the total chlorophyll concentration. Relatively high concentrations of fucoxanthin and chlorophyll b indicate that the majority of the species in the Lena River were composed of green algae (Chlorophyta) and diatoms (Bacillariophyta). This was confirmed by the measurements in situ using the BBE multialgal-fluorescence probe.

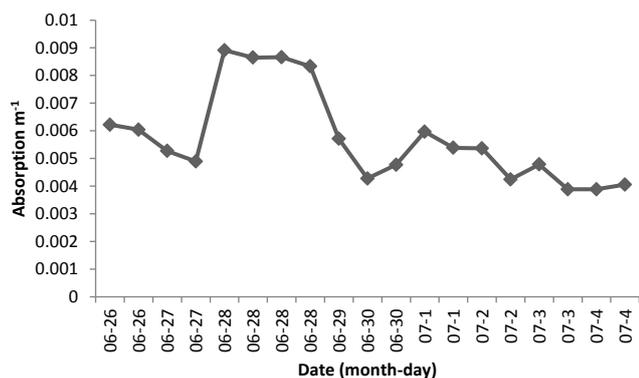


Fig. 5. Detritus semi-logarithmic slopes calculated between 400 and 500 nm with 442 nm as reference wavelength using a nonlinear fit.

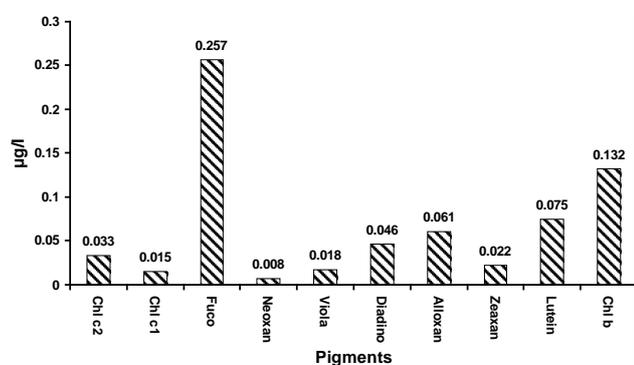


Fig. 6. Mean pigment concentrations, chlorophyll *a* not included here.

4.3 Total suspended matter

Total suspended matter concentration was spatially and temporally variable during the sampling period. Concentrations in samples, which were taken from the Kurungnakh arm (stations 3–6, Fig. 1), were lower than those of the samples of the main channel (Fig. 7). During the observation period concentrations of TSM increased by a factor of four and reached its maximum at the last days of the campaign.

4.4 Carbon concentrations

The concentration of DOC was constant over the observation period in June/July at all stations with values of around $9.4 \pm 0.45 \text{ g m}^{-3}$ when omitting one value at 16 g m^{-3} . These values were double those from samples taken in the second half of August 2011 at the same stations, with values of around $4.6 \pm 0.46 \text{ g m}^{-3}$ ($n = 38$), also when omitting one potential erroneous value at 16 g m^{-3} (data not included in the publication).

In contrast, the values of the sum of organic and inorganic particulate carbon (PC) were much more variable. During the campaign in June/July PC values were in the range $0.3\text{--}1.9 \text{ g m}^{-3}$ with a mean of $0.85 \pm 0.44 \text{ g m}^{-3}$,

$n = 26$. In the second half of August these values were more variable with a range of $0.2\text{--}6.0 \text{ g m}^{-3}$ and a mean of $1.07 \pm 1.08 \text{ g m}^{-3}$, $n = 36$.

4.5 Beam attenuation and scattering

The variability of the beam attenuation coefficient c of the Lena River water was determined mainly by the scattering and to less extent by the absorption of TSM, while the CDOM absorption was relatively constant (Fig. 8a, b). Similar to TSM dry weight the beam attenuation coefficient increased by a factor of four during the observation period (Fig. 8b).

The scattering spectrum is given in Fig. 9a. The relationships between the concentrations of TSM and the beam attenuation and absorption coefficients, respectively, show similar scatter.

4.6 Mass specific inherent optical properties

The mass specific absorption and scattering coefficients were computed from the slope of the linear regressions between concentrations of chlorophyll, TSM, DOC, and PC and the corresponding absorption and scattering coefficients at 442 nm. The relative uncertainty is calculated from the regression on the logarithm scale to get a mean relative error over the entire range. Due to the limited number of data and the narrow concentration ranges these coefficients have to be regarded as preliminary. For chlorophyll the problem in deriving the absorption coefficients from the minimal difference between the total and the absorption after bleaching, led to a high uncertainty in the concentration specific absorption coefficient.

Pigment absorption:

$$a_{\text{pig}_442} [\text{m}^{-1}] = 0.128 [\text{m}^2 \text{mg}^{-1}] \cdot \langle \text{Chl}_a \rangle > [\text{mg m}^{-3}], \pm 0.08 [\text{m}^{-1}], \pm 60 \% \text{ relative uncertainty.}$$

$$a_{\text{pig}_676} [\text{m}^{-1}] = 0.05 [\text{m}^2 \text{mg}^{-1}] \cdot \langle \text{Chl}_a \rangle > [\text{mg m}^{-3}] \pm 0.025 [\text{m}^{-1}], \pm 34 \% \text{ relative uncertainty.}$$

CDOM absorption:

$$a_{\text{cdom}_442} [\text{m}^{-1}] = 0.56 [\text{m}^2 \text{g}^{-1}] \cdot \langle \text{DOC} \rangle > [\text{g m}^{-3}] \pm 0.2 [\text{m}^{-1}], \pm 5 \% \text{ relative uncertainty.}$$

When converting $\mu\text{mol DOC}$ into a_{cdom_442} the conversion factor is $0.00672 [\text{m}^2 \mu\text{mol}^{-1}]$.

Absorption by bleached TSM (detritus):

$$a_{\text{det}_442} [\text{m}^{-1}] = 0.07 [\text{m}^2 \text{g}^{-1}] \cdot \langle \text{TSM} \rangle > [\text{g m}^{-3}] \pm 0.4 [\text{m}^{-1}], \pm 42 \% \text{ relative uncertainty.}$$

Scattering by particles at 442 nm derived from a linear regression (Fig. 9c), which has a bias of 1.07 m^{-1} :

$$b_{\text{tsm}_442 \text{ nm}} [\text{m}^{-1}] = 0.45 [\text{m}^2 \text{g}^{-1}] \cdot \langle \text{TSM} \rangle > [\text{g m}^{-3}] \pm 5.8 [\text{m}^{-1}], \pm 35 \% \text{ relative uncertainty.}$$

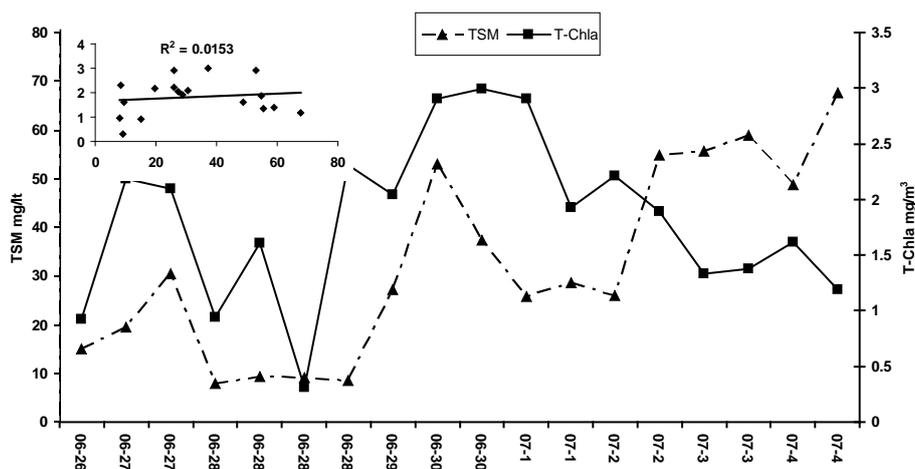


Fig. 7. Fluctuations of chlorophyll *a* and total suspended matter concentrations. The insert shows the relationship between suspended matter (x axis, g m^{-3}) and chlorophyll (y axis, mg m^{-3}).

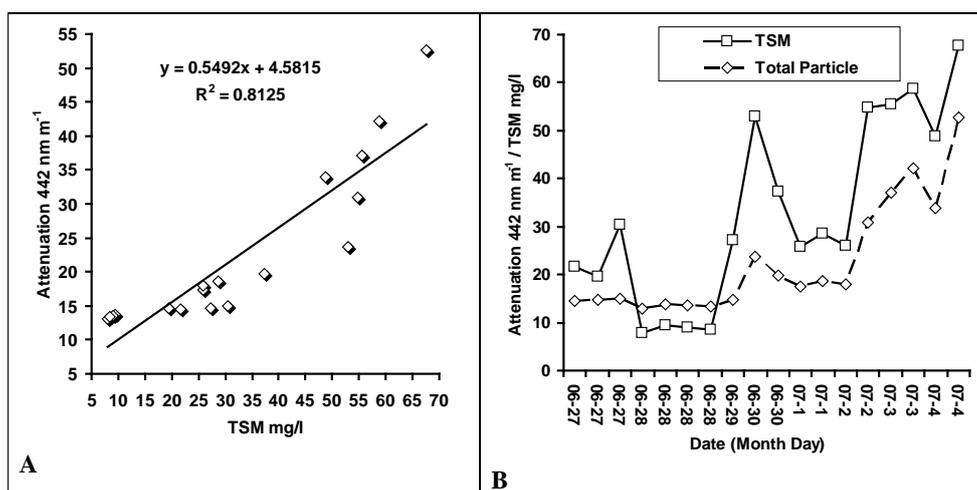


Fig. 8. Co-variation (A) and temporal variations (B) of TSM and the beam attenuation at 442 nm.

Absorption by all particles (TSM) at 442 nm derived from a linear regression (Fig. 9d), which has a bias of 0.23 m^{-1} :

$$a_{\text{tsm}_{442 \text{ nm}}} \quad [\text{m}^{-1}] = 0.040[\text{m}^2 \text{ g}^{-1}] \cdot \langle \text{TSM} \rangle \\ > [\text{g m}^{-3}] \pm 0.43 [\text{m}^{-1}], \pm 31 \% \text{ relative uncertainty.}$$

Particulate carbon [g m^{-3}] related absorption coefficient of particulate matter:

$$a_{\text{p}_{442 \text{ nm}}} \quad [\text{m}^{-1}] = 0.83[\text{m}^2 \text{ g}^{-1}] \cdot \langle \text{PC} \rangle \\ [\text{g m}^{-3}] \pm 0.43 [\text{m}^{-1}], \pm 37 \% \text{ relative uncertainty.}$$

5 Discussion

A major question underpinning our investigations on the bio-optical properties of the Lena water is as to how the results compare to other Arctic rivers and estuaries. This leads on to whether a generic bio-optical model can be derived for

optical remote sensing of the coastal waters of the Arctic Sea with an eye on other rivers besides the Lena, namely the Yenisei, Ob, and Mackenzie, which also determine the fresh water flux into the Arctic Sea.

Most striking in Lena water is the high concentration of all kinds of humic substances (see also Kattner et al., 1999), defined here as organic degradation products and whereby dissolved and particulate matter are merely defined by the pore size of the filter (see data at http://www.feow.org/ecoregion_details.php?eco=608). These dominate the optical properties of all major Arctic estuaries in the form of CDOM and detritus. The DOC concentrations of Ob and Yenisei, close to their mouths, where river water is not diluted by marine water, is in the range of $5\text{--}10 \text{ g C m}^{-3}$ with extreme values of 15 g C m^{-3} . The concentrations in the Kara Sea, where salinity exceeds 30 ppt, is $< 1 \text{ g C m}^{-3}$ (Korosov et al., 2011), also reported by Hessen et al. (2010). The concentrations, we

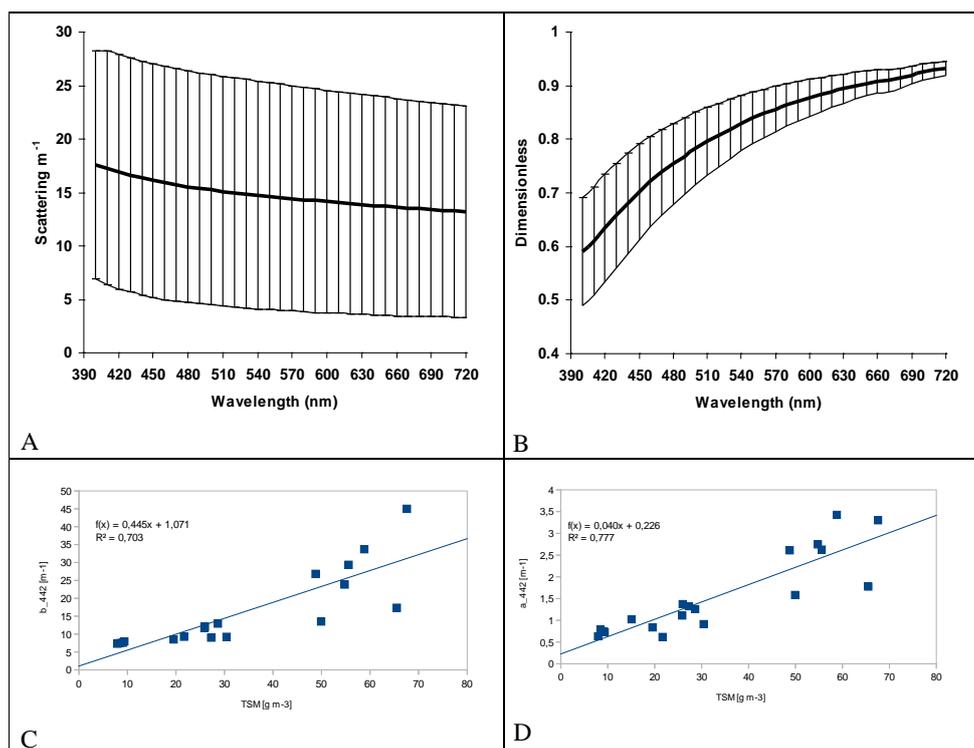


Fig. 9. Means and standard deviations of (A) scattering and (B) the single scatter albedo, as well as the relationship between the total suspended matter concentrations and (C) scattering and (D) absorption coefficients.

found during our campaign in the fresh water region of the Lena were rather constant with a range of 8.5–10.5 g C m^{-3} . When including the measurements of August 2011 at the same stations with a mean of $4.6 \pm 0.46 \text{ g m}^{-3}$, we found almost the same range of DOC as reported for the Ob and Yenisei whereby even our extreme values of 16 g C m^{-3} agree. In contrast, the values of the sum of organic and inorganic PC were much more variable. Here one has to consider that also the absorption by minerals may contribute to the overall absorption coefficient of TSM (Estapa et al., 2012). During the campaign in June/July PC values were in the range $0.3\text{--}1.9 \text{ g m}^{-3}$ with a mean of $0.85 \pm 0.44 \text{ g m}^{-3}$, $n = 26$. In the second half of August these values were even more variable with a range of $0.2\text{--}6.0 \text{ g m}^{-3}$ and a mean of $1.07 \pm 1.08 \text{ g m}^{-3}$, $n = 36$.

The absorption coefficients of CDOM are easier to compare. In the paper of Hessen et al. (2010) the downwelling irradiance attenuation coefficient k_d is included for some wavelengths in the ultraviolet (UV) spectral range, which are around 10 m^{-1} for 380 nm for the Ob and Yenisei. Since for highly absorbing water, k_d is mainly determined by the absorption coefficient, it can be transferred to an absorption coefficient of 4 m^{-1} at 442 nm assuming a spectral slope of 0.015. For the Lena we found the corresponding coefficient in the range $4\text{--}5 \text{ m}^{-1}$.

Absorption coefficients for the Mackenzie Delta (Canada) are reported in Bélanger et al. (2006). Values close to the coast are in the range of $4.5\text{--}5.4 \text{ m}^{-1}$ at 330 nm with a spectral slope of around 0.02. The corresponding values for 442 nm using this slope would be $0.45\text{--}0.55 \text{ m}^{-1}$.

Mean spectral slopes of the absorption coefficients are available for the Great Whale River and Mackenzie River in the Canadian Arctic (Retamal et al., 2007) at 0.016 and 0.0185 m^{-1} , respectively. However, they were derived from the wavelength range 300–650 nm with a reference wavelength of 320 nm. Slopes derived from the UV spectral range are normally higher (Diehl & Haardt, 1980), since they are determined by other components of CDOM. Also the spectral slopes in Bélanger et al. (2006) with values in the range of $0.019\text{--}0.02 \text{ m}^{-1}$ have obviously been derived from the UV part of the spectrum (but not explicitly mentioned in the paper). Similar CDOM spectral slope values can be found for different coastal and river waters worldwide. Helms et al. (2008) reported values between 0.0158 and 0.0186 nm^{-1} from the Great Dismal Swamp and the Suwannee River. The mean CDOM spectral slope of Elbe river water is reported to be 0.0188 (Helmholtz Center Geesthacht, unpublished data) or 0.014 nm^{-1} (Doerffer, 1979). More reports of this parameter can be found in Wozniak and Dera (2007). Slopes of CDOM, which we found for the Lena, are in the range $0.0145\text{--}0.0155 \text{ nm}^{-1}$. They were derived from the

spectral range of 400–500 nm, which is the most relevant part for optical remote sensing.

Variability of the slopes of detritus absorption (range 0.004–0.009 nm⁻¹) can be attributed to variations in particle size and/or composition including inorganic particles and content of organic carbon (detritus). Compared to spectra of Estapa et al. (2012) the shape of the spectra of Lena water does not indicate a significant contribution by iron. The spectral slopes are lower compared to mean Baltic and North Sea detritus absorption slopes, which are 0.013 and 0.0116 nm⁻¹, respectively (Babin et al., 2003a). Ferrari et al. (2003) reported values between 0.0095 and 0.0125 nm⁻¹ for the Baltic Sea. These are similar to values found in the NOMAD database (Werdell and Bailey, 2005) which are in the range 0.007–0.015 m⁻¹ (5–95 % percentile).

Concentrations of phytoplankton pigments found are similar to previous studies carried out in the Lena River (Heiskanen and Keck, 1996; Sorokin and Sorokin, 1996). Total chlorophyll *a* concentration was variable within our sampling period with values between 0.5 and 3 mg m⁻³. The time series does not show any significant trend during our campaign. The samples taken from the Kurungnakh river arm were not significantly different from the major river. These values match the data of Hessen et al. (2010) for the Ob (0.4–3.4 mg m⁻³) and Yenisei (0.6–2.9 mg m⁻³). Chlorophyll values derived from satellite data are in most cases overestimated due to the difficult separation between CDOM and phytoplankton pigment absorption spectra (Heim et al., this volume). In Hessen et al. (2010) chlorophyll values derived from MODIS satellite data are up to 15 mg m⁻³ for the Yenisei and up to 8 mg m⁻³ for the Ob. The chlorophyll values derived from MERIS satellite data (Korosov et al., 2011) are even higher with values up to 30 mg m⁻³ for the Ob and Yenisei.

The concentration of TSM was variable in the Lena River and its side arms. We assume that the variability is mainly due to differences in water turbulence, which, however, could only be rated visually. TSM ranged from 10–60 g m⁻³ during the observation period and did not co-vary with the chlorophyll concentration (Fig. 7). Due to the high variability, which depends on local currents, water depth, and river runoff, a direct comparison with corresponding snapshot observations of other Arctic rivers does not help in assessment of these values.

The calculated single scattering albedo of the Lena River was lower in the short and higher in the long wavelength range. This effect is caused by the strong absorption by CDOM at shorter wavelengths (Fig. 9b), which reduces the number of photons to be scattered at these wavelengths. The spectrum of scattering coefficient shows an increase in the short wavelength range and remains flat at longer wavelengths (Fig. 9a). This type of spectral shape is determined by the particle composition and concentration. Especially small lithogenic particles can increase the scattering coefficients

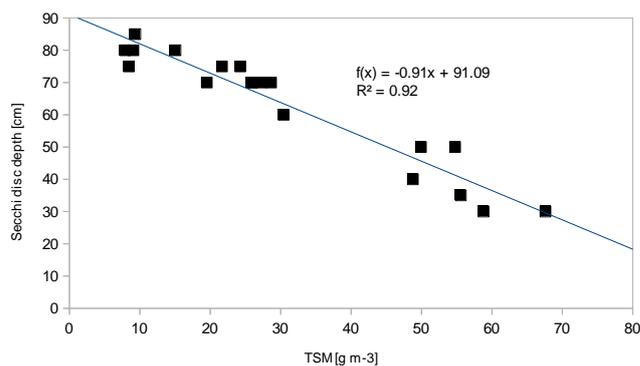


Fig. 10. Relationship between TSM dry weight and the Secchi disc depth for all Lena stations.

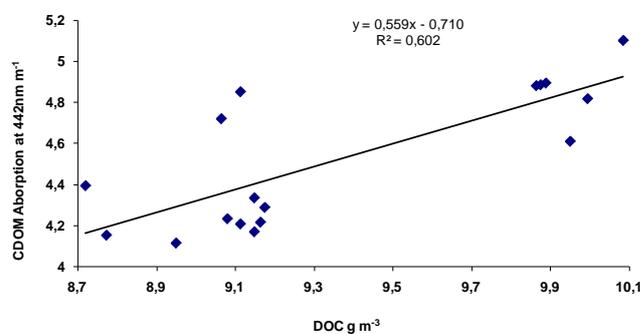


Fig. 11. Relationship between DOC and the absorption coefficient of the filtrate at 442 nm against pure water as blank.

(Whitmire et al., 2007) and the backscattering ratio (Boss et al., 2004).

The relationship between the beam attenuation and the TSM concentration could be compared with results of Hill et al. (2011), who analyzed the effect of different particle sizes. They found a mean conversion factor of 0.22 g m⁻² with a range between 0.2 and 0.4 for a wavelength of 670 nm. For the Lena water we found a corresponding factor (for 670 nm) of 0.37 g m⁻² (see Figs. 8a and 2d).

The relationship between the TSM dry weight and the scattering coefficient of TSM shows a clear linear relationship (Fig. 9c) when we omit the measurements of the first day. The mass specific scattering coefficient is 0.45 m² g⁻¹ ± 35 %. Babin et al. (2003b) found average values of 1.0 and 0.5 m² g⁻¹ in optical case 1 and case 2 waters, respectively, of the North Sea and Baltic Sea. The scatter diagram of TSM concentration and the detritus absorption at 442 nm (Fig. 12) shows a logarithmic relationship, which might be due to changes in the composition of TSM at different concentration levels.

For the mass specific absorption coefficient of TSM of samples from coastal Louisiana and the Mississippi and Atchafalaya Rivers, Estapa et al. (2012) found values between 0.01 and 0.05 m² g⁻¹ at 550 nm. For the

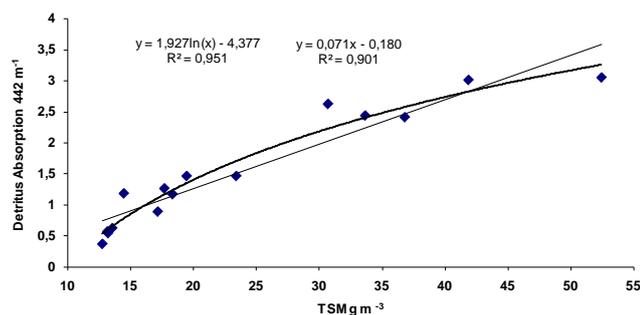


Fig. 12. Relationship between total suspended matter dry weight (TSM) and the absorption coefficient of the particulate matter at 442 nm, logarithmic and linear fit.

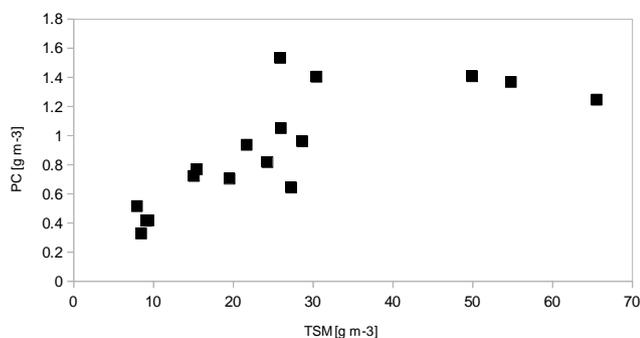


Fig. 13. Relationship between particulate carbon and total suspended matter concentration.

Lena the corresponding specific absorption coefficient is $0.027 \text{ m}^2 \text{ g}^{-1} \pm 31\%$ at 550 nm.

A high correlation was also found for the relationship between TSM dry weight and the Secchi disc depth (Fig. 10). CDOM contributes also to the light penetration, but since its concentration was rather constant during the observation period, the variability in the Secchi disc depth was mainly determined by suspended particles.

Of further interest was the relationship between particulate organic carbon and TSM, which is shown in Figs. 13 and 14. These figures show that the three samples with very high concentrations of TSM have lower amount of PC, since they presumably consist mainly of mineral suspended matter.

Due to the dominating effect of absorption mainly by organic particulate matter and the low chlorophyll concentration, it was difficult to derive the specific phytoplankton pigment absorption coefficient using the filter pad and bleaching method. The relationship between chlorophyll *a* and the pigment absorption coefficient at 442 nm was: $\langle \text{Chl}_a \rangle [\text{mg m}^{-3}] = 7.8 [\text{m}^2 \text{ mg}^{-1}] \cdot \text{apig}_{442 \text{ nm}} [\text{m}^{-1}]$, which is a lower conversion factor than normally found for open and coastal water phytoplankton.

The relationship between DOC and the absorption by CDOM at 442 nm (Fig. 11) were compared to measurements in the Baltic Sea. Ferrari et al. (1996) found mean values be-

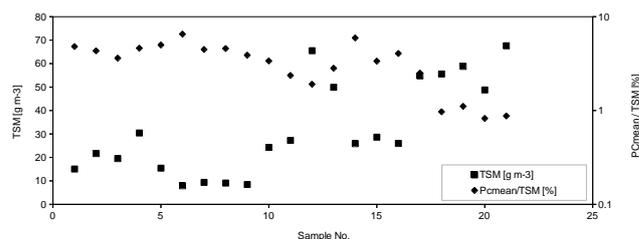


Fig. 14. Total suspended matter dry weight (TSM) in g m^{-3} and the PC content as % of TSM, in sequence of sample number.

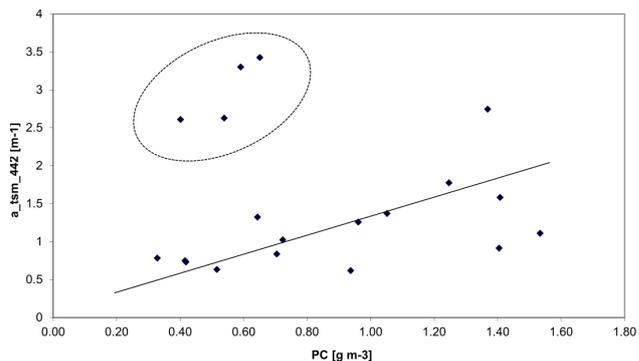


Fig. 15. Relationship between PC and the absorption coefficient of particulate matter at 442 nm, data points within the circle are only from samples of 3 and 4 July 2011, when omitting these points the regression is $a_{\text{TSM}_{442} < \text{m}^{-1} > = 0.83 < \text{m}^2 \text{ g}^{-1} > \cdot \text{PC} < \text{g m}^{-3} > + 0.26 < \text{m}^{-1} >$.

tween 49 and $73 \text{ m}^2 \mu\text{mol}^{-1}$ for 355 nm, which depend on the DOC analysis technique (see their Table 6). When we convert our values to a wavelength of 355 nm using a spectral slope of 0.014 m^{-1} , we get a conversion factor of $44 \text{ m}^2 \mu\text{mol}^{-1}$. Fichot and Benner (2011) measured the absorption at 275 and 295 nm in the northern Gulf of Mexico and Beaufort Sea. For their conversion from CDOM absorption to DOC on the log scale they used the absorption at both wavelengths: $\log(\text{DOC}) = a + b \cdot \log(\text{ag}_{275}) + c \cdot \log(\text{ag}_{295})$. The conversion factors for the concentration range $\text{ag}_{275} > 3.5 \text{ m}^{-1}$ were $a = 2.9$, $b = 2.77$, and $c = -2.04$ for the northern Gulf of Mexico and $a = 2.36$, $b = 3.04$, and $c = -2.13$ for the Beaufort Sea. When we transfer our values to a wavelength of 285 nm using again a spectral slope of 0.014 m^{-1} we get corresponding conversion factors for the log scale $\log(\text{DOC}) = a + b \cdot \log(\text{ag}_{285})$ of $a = 4.73$ and $b = 0.52$ with an uncertainty of $\pm 3\%$ on the logarithmic scale.

6 Summary and conclusions

The primary goal of this investigation was to determine the bio-optical properties of Lena River water during the phase of maximum runoff in order to obtain a basis for optical

remote sensing and primary production studies. The investigation site, where we expected maximum concentrations, was the entrance of the Lena Delta before the water enters the coast of the Laptev Sea. The measured variables include absorption and scattering coefficients of water samples before and after filtration to separate the particulate from the dissolved fraction of water constituents. The absorption by particulate matter was further split into the absorption coefficients of phytoplankton pigments and detritus. Also the concentrations of total suspended matter dry weight and of phytoplankton pigments were determined.

Although the short investigation period could provide us with only a snapshot of Lena River optical properties, it was the first time that crucial bio-optical properties of the Lena River were studied during the most dynamic period when the flow and material transport are at a maximum. Furthermore, the data were already important for the evaluation of satellite data (Heim et al., this volume).

Most remarkable was the extremely high absorption of light by the dissolved and particulate fraction of humic substances. Secchi disc depths (SD) showed a maximum of about 90 cm. According to our visual inspection of the water and the regression (Fig. 10) we have to attribute this maximum SD to the dissolved organic matter fraction. This depth was then reduced to 30 cm by variable concentrations of TSM in the range of 8–70 g m⁻³ with a linear relationship between TSM and SD. However, for this phase of maximum river runoff we expected higher TSM concentrations compared to late summer conditions with low water. Visual turbidity observations in August 2008 in the same area and also satellite data of summer periods indicate much higher TSM concentrations than during this campaign.

The linear relationship between DOC and the CDOM absorption at 442 nm allows the determination of DOC also from optical remote sensing data, from which CDOM absorption can be determined (see Heim et al., this volume). More difficult is the determination of PC from optical data, since the scattering as well as the absorption coefficient is also determined by particulate matter without carbon.

With respect to the light climate, which is determined by the extremely high absorption in the blue spectral range, the chlorophyll concentration of 1–4 mg m⁻³ was relatively high. Since the water depth of the Lena at the sampling stations was around 6 m with a range of 4–21 m, the question that arises is how phytoplankton can develop under these poor light conditions.

Of interest for remote sensing applications are the spectral slopes of the absorption coefficients of organic matter. The mean value of about 0.015 m⁻¹ corresponds to what was found in many other estuaries, e.g., the river Elbe. The slight but clear trend over our observation period is presumably due to changes in the composition of organic matter with the river runoff. This corresponds to an increase of the absorption by TSM and a decrease of the absorption by dissolved organic matter. Thus, for remote sensing algorithms it might

be important to introduce not only one spectral slope but two, which bracket the mean minimum and maximum values.

Since the organic matter dominates the absorption of Lena water, it will be difficult to identify phytoplankton from reflectance measurements, since its absorption contributes less than 10% to the overall absorption. Due to the high absorption by CDOM in the blue part of the spectrum the sunlight-induced fluorescence of chlorophyll will also be low and thus difficult to determine from reflectance spectra.

Although we have sampled at only a few locations and for a short period we expect no significant differences at other locations in the Lena Delta during this time, since the environmental conditions elsewhere were very similar, as indicated also by satellite data (Heim et al., this volume).

It was possible to establish a preliminary bio-optical model of the Lena water, which can be used for optical remote sensing. This model links the concentration of different water constituents with their optical properties. Where a comparison with other areas was possible, in particular with the Ob and Yenisei, the values of Lena were found to be similar. The concentration range of phytoplankton and CDOM was small, which limits the general application of the results. However, since our observation period was during the phase with the maximum runoff, the data can be used for estimating the input of substances from the Lena into the coastal water of the Laptev Sea and related to remote sensing data.

Appendix A

Our present bio-optical model for the Lena water consists of four components: (1) absorption by phytoplankton pigments, (2) absorption by CDOM, (3) absorption by bleached total suspended matter (detritus), and (4) scattering by TSM, all defined for 442 nm.

The mean ranges and conversion factors are the following.

Range of chlorophyll: 0.0–10 mg m⁻³

Conversion to absorption by pigments at 442 nm:

$$a_{\text{pig}}_{442} [\text{m}^{-1}] = 0.128 [\text{mg}^{-1} \text{m}^2] \cdot \langle \text{Chl}_a \rangle [\text{mg m}^{-3}].$$

Absorption by CDOM at 442 nm range: up to 5 m⁻¹

Spectral slope of CDOM: 0.015 m⁻¹

Relationship between DOC [g m⁻³] and absorption coefficient of CDOM:

$$a_{\text{cdom}}_{442 \text{ nm}} [\text{m}^{-1}] = 0.56 [\text{g}^{-1} \text{m}^2] \cdot \langle \text{DOC} \rangle [\text{g m}^{-3}].$$

Absorption by bleached TSM range: up to 5 m⁻¹

Spectral slope of bleached TSM absorption spectrum: 0.006 m⁻¹

Conversion: absorption by bleached TSM as

$$a_{\text{det}}_{442} [\text{m}^{-1}] = 0.07 [\text{g}^{-1} \text{m}^2] \cdot \langle \text{TSM} \rangle [\text{g m}^{-3}].$$

Scattering by particles b_{tsm}_{442} range: 5–50 m⁻¹

Conversion: scattering by particles as

$$b_{\text{tsm}}_{442 \text{ nm}} [\text{m}^{-1}] = 0.45[\text{g}^{-1} \text{ m}^2] \cdot \langle \text{TSM} \rangle [\text{g m}^{-3}].$$

Conversion: particulate carbon [g m^{-3}] and absorption coefficient of bleached TSM (detritus) as

$$a_{\text{det}}_{442 \text{ nm}} [\text{m}^{-1}] = 0.83[\text{g}^{-1} \text{ m}^2] \cdot \langle \text{PC} \rangle [\text{g m}^{-3}].$$

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