



# Hydrogen dynamics in soil organic matter as determined by $^{13}\text{C}$ and $^2\text{H}$ labeling experiments

Alexia Paul<sup>1</sup>, Christine Hatté<sup>2</sup>, Lucie Pastor<sup>3</sup>, Yves Thiry<sup>4</sup>, Françoise Siclet<sup>5</sup>, and Jérôme Balesdent<sup>1</sup>

<sup>1</sup>Aix-Marseille Université, CNRS, Collège de France, IRD, INRA, CEREGE UM34, 13545 Aix-en-Provence, France

<sup>2</sup>Laboratoire des Sciences du Climat et de l'Environnement, LSCE/IPSL, UMR 8212 CEA-CNRS-UVSQ, Université Paris Saclay, 91198 Gif-sur-Yvette, France

<sup>3</sup>IFREMER/Centre de Brest, Département REM/EEP/LEP, CS 10070, 29280 Plouzané, France

<sup>4</sup>Andra, Research and Development Division, Parc de la Croix Blanche, 1/7 rue Jean Monnet, 92298 Châtenay-Malabry CEDEX, France

<sup>5</sup>EDF R&D, LNHE, 6 quai Watier, 78400 Chatou, France

Correspondence to: Alexia Paul (alexia.paul@inra.fr) and Jérôme Balesdent (jerome.balesdent@inra.fr)

Received: 27 July 2016 – Published in Biogeosciences Discuss.: 28 July 2016

Revised: 25 November 2016 – Accepted: 28 November 2016 – Published: 15 December 2016

**Abstract.** Understanding hydrogen dynamics in soil organic matter is important to predict the fate of  $^3\text{H}$  in terrestrial environments. One way to determine hydrogen fate and to point out processes is to examine the isotopic signature of the element in soil. However, the non-exchangeable hydrogen isotopic signal in soil is complex and depends on the fate of organic compounds and microbial biosyntheses that incorporate water-derived hydrogen. To decipher this complex system and to understand the close link between hydrogen and carbon cycles, we followed labeled hydrogen and labeled carbon throughout near-natural soil incubations. We performed incubation experiments with three labeling conditions: 1 –  $^{13}\text{C}^2\text{H}$  double-labeled molecules in the presence of  $^1\text{H}_2\text{O}$ ; 2 –  $^{13}\text{C}$ -labeled molecules in the presence of  $^2\text{H}_2\text{O}$ ; 3 – no molecule addition in the presence of  $^2\text{H}_2\text{O}$ . The preservation of substrate-derived hydrogen after 1 year of incubation (ca. 5% in most cases) was lower than the preservation of substrate-derived carbon (30% in average). We highlighted that 70% of the C–H bonds are broken during the degradation of the molecule, which permits the exchange with water hydrogen. Added molecules are used more for trophic resources. The isotopic composition of the non-exchangeable hydrogen was mainly driven by the incorporation of water hydrogen during microbial biosynthesis. It is linearly correlated with the amount of carbon that is degraded in the soil. The quantitative incorporation of water hydrogen in bulk material and lipids demonstrates that non-

exchangeable hydrogen exists in both organic and mineral-bound forms. The proportion of the latter depends on soil type and minerals. This experiment quantified the processes affecting the isotopic composition of non-exchangeable hydrogen, and the results can be used to predict the fate of tritium in the ecosystem or the water deuterium signature in organic matter.

## 1 Introduction

Our knowledge of the nature of soil organic matter (SOM) has made great progress in recent decades: it is now considered to be a continuum of progressively decomposing organic compounds (Lehmann and Kleber, 2015), composed of all the components of living material, such as glucides, peptides, lipids, organic acids and phenolic compounds (Keller and Simpson, 2006). Small molecules are associated with each other in supramolecular structures or with mineral particles by weak bonds, including H bonds (Sutton and Spósito, 2005; Lehmann and Kleber, 2015). Their lifetime in soils is controlled more by sorption or protection than by intrinsic chemical recalcitrance to biodegradation (Schmidt et al., 2011; Basile-Doelsch et al., 2015) with the exception of pyrolytic products. Highly degradable compounds, such as glucides and peptides, contribute to the oldest SOM components, and biomarkers tend to indicate that old SOM was

derived more from microbially derived products than from plant-derived molecules as a result of the mineral protection processes (Derrien et al., 2007; Bol et al., 2009). Carbon dynamics in this continuum have been widely studied using the natural  $^{14}\text{C}/^{12}\text{C}$  and  $^{13}\text{C}/^{12}\text{C}$  (Rafter and Stout, 1970; Balesdent et al., 1987) ratios and also through labeling experiments (Jenkinson, 1965; Murayama, 1988; Derrien et al., 2004). The results of these experiments have highlighted the different turnover of soil organic pools. Plant material is rapidly decomposed into microbial biomass, and a small portion of both can be protected from biodegradation for decades to centuries, representing the main part of SOM. One part of this organic matter remains stabilized for millennia, especially in deep soil horizons. It is therefore expected that the non-exchangeable hydrogen (NEH) dynamics, bound to carbon in soil, will be controlled by the same processes: organic carbon inherited from vegetation, biodegradation, microbial biosyntheses and stabilization.

Hydrogen has various molecular positions in soil. It can be organic or inorganic and non-exchangeable or exchangeable with available hydrogen. The abiotic exchange of organic hydrogen depends on the strength of the bond and the energy required for exchange (Belot, 1986; Schimmelmann, 1991; Ciffroy et al., 2006; Sauer et al., 2009). Bound to N, O and S, hydrogen is usually exchangeable with ambient water and water vapor (Schimmelmann et al., 1999; Wassenaar and Hobson, 2000, 2003). However, hydrogen bound to carbon is considered to be stable and non-exchangeable due to the strong covalent bonds (Baumgärtner and Donhaerl, 2004; Diabaté and Strack, 1997; Kim et al., 2013). On the ecosystem scale, H bound to C is not exchangeable (Sessions et al., 2004). Hydrogen can also interfere with clay minerals. Inter-layer water exchanges with free water within a few hours and is removed after drying (Savin and Epstein, 1970). However, the structural water and the hydroxyl hydrogen of clay are non-exchangeable at room temperature (Savin and Epstein, 1970).

The natural  $^2\text{H}/^1\text{H}$  ratio of plants and sediments has been used as a proxy to reconstruct past climate and paleoenvironmental conditions, such as temperature and water use efficiency (Epstein et al., 1976; Sessions et al., 2004; Zech et al., 2014; Tuthorn et al., 2015). The isotopic composition of the NEH preserves the initial composition of the plant and registers the rain isotopic composition (Sessions et al., 2004; Schimmelmann et al., 2006; Ruppenthal et al., 2010). The  $\delta^2\text{H}$  of water and exchangeable hydrogen is not stable. Whereas soil organic carbon and nitrogen cycles have been extensively studied, soil organic hydrogen and its recycling in the environment remain poorly understood due to hydrogen's complex behavior. The total bulk soil, composed of a mixture of non-exchangeable and exchangeable, organic and inorganic hydrogen, makes the hydrogen isotopic composition hard to determine.

The composition and exchanges between these pools can be of great importance when modeling, for instance, tritium

fate in the environment. Tritium is a radioactive isotope of hydrogen which was released in large amounts in the atmosphere by nuclear weapon tests in the 1960s. Since then, tritium levels have greatly declined because of its relatively short half-life (12.3 years). However the concentration of organically bound tritium can often be higher than the concentration of water tritium due its longer residence time (Gontier and Siclet, 2011; Kim et al., 2012; Eyrolle-Boyer et al., 2014; Thompson et al., 2015). Models used to predict the fate and behavior of tritium in the environment often simplify processes linked to the formation and degradation of organically bound tritium. Since tritium behaves as stable hydrogen in the environment, the assessment of the fate and residence time of organic tritium could be improved by quantifying the preservation of organic hydrogen from vegetation, the accumulation of hydrogen from water in the soil, and the processes involved in organic matter decomposition and mineralization.

To decipher and quantify the preservation of the organic material and the microbial biosyntheses incorporating water-derived hydrogen, we designed incubation experiments with labeled compounds by assuming that the non-exchangeable hydrogen dynamics are controlled by the carbon dynamics in the soil organic matter. Three scenarios were addressed:

1.  $^{13}\text{C}^2\text{H}$ -double labeled molecules in the presence of  $^1\text{H}_2\text{O}$
2.  $^{13}\text{C}$ -labeled molecules in the presence of  $^2\text{H}_2\text{O}$
3. no molecule addition in the presence of  $^2\text{H}_2\text{O}$ .

The  $^{13}\text{C}$  and  $^2\text{H}$  bulk soil isotopic compositions were analyzed at different times to quantify the processes involved. The isotopic composition of lipids was also analyzed as an indicator of organically bound NEH.

The medium-term  $^{13}\text{C}$  and  $^2\text{H}$  labeling experiments were conducted on different types of soil (clayey Leptosol, Cambisol and Podzol) from 0 to 1 year to highlight and quantify the processes affecting hydrogen based on the carbon dynamics in the soil organic matter and their dependence on the soil properties.

## 2 Materials and methods

### 2.1 Soil sampling

Three soils with contrasting physical and chemical properties were selected for this study (Table 1):

*Cambisol*: the 0–25 cm surface layer of a Cambisol was sampled from an INRA long-term field experiment in the Parc du Château de Versailles, France in March 2014. This soil is a neutral eutric Cambisol with a composition of 17 % clay, 33 % sand, 50 % silt and 0 % carbonate (Dignac et al., 2005). The plant cover is wheat. After each harvest, wheat

**Table 1.** Pedologic, geographic information and carbon and hydrogen isotopic composition of the three bulk soils.

Soil	Latitude, longitude	Sampling depth (cm)	Plant cover	% clay	% silt	% sand	C mg g <sup>-1</sup>	H mg g <sup>-1</sup>	N mg g <sup>-1</sup>	pH	Mean bulk $\delta^{13}\text{C}$ (‰)	Mean bulk $\delta^2\text{H}$ (‰)
Cambisol	48°48'02" N, 2°06'33" W	0–25	wheat	17	50	33	12	3.2	1.3	6.8	-26 ± 0	-53 ± 3
Podzol	44°44'33" N, 0°47'37" W	0–25	maize	5	3	92	20	2.3	0.8	5.5	-26 ± 0	-51 ± 3
Leptosol	43°56'06" N, 5°42'06" W	5–10	oak forest	54	37	9	39	13.7	2.6	7.5	-25 ± 0	-52 ± 2

residues were returned to the soil, and the first 25 cm were ploughed each fall.

*Podzol*: the 0–25 cm surface layer of a Podzol was sampled from an INRA field experiment in Pierroton (close to Bordeaux, SW France) in May 2014. The land cover of the Landes de Gascogne is a mixed forest dominated by *Pinus pinaster*. The sampling plot was converted into maize in 1992. The soil is a sandy hydromorphic Podzol (Jolivet et al., 2006) with a clay content of less than 5 % and a sand content higher than 90 %. The first 25 cm are ploughed and crop residues are returned to the soil after each harvest.

*Leptosol*: the 5–10 cm surface layer of a mollic Leptosol was sampled from the long-term ecosystem research experiment “Oak Observatory at Observatory of Haute-Provence” (O3HP), France, in July 2014. The vegetation is dominated by *Quercus pubescens*. This soil is derived from limestone, compact and iron-rich with a clay content (mainly smectite) of 54 %.

## 2.2 Soil incubation

### 2.2.1 Soil preparation

Soil samples were air-dried at 20 °C and sieved to 2 mm. Air-dry soil moisture was determined in parallel by oven-drying an aliquot at 105 °C during 24 h. The residual soil moisture was determined once before adding substrate and water in order to know the exact amount of water at the beginning of incubation.

The residual soil moisture was 5.9 % for the Cambisol, 0.8 % for the Podzol and 8.5 % for the Leptosol.

Thirty-five grams of Cambisol and Podzol and 30 g of mollic Leptosol were transferred into 210 mL incubation jars. Each incubator was then moistened with ultrapure water ( $\delta^2\text{H} = -63.8 \pm 0.5$  ‰) at 31 mg g<sup>-1</sup> of dry soil for Cambisol (6 mL of ultrapure water added) and Leptosol (4 mL of ultrapure water added) and 14 mg g<sup>-1</sup> for Podzol (3 mL of ultrapure water added) before preincubation at 28 °C in the dark for 10 days to reestablish the biological activity at the basal level and to avoid confusion between rewetting-induced and substrate-induced activity.

### 2.2.2 Substrate incubations

Glucose, palmitic acid, phenylalanine and isoleucine were introduced separately in different incubators. They represent the most common primary compounds of the glucide, lipid and protein families found in either plant or microbial matter and contain different functional groups.

<sup>13</sup>C-labeled and <sup>2</sup>H-labeled molecules and <sup>2</sup>H<sub>2</sub>O were provided by Euriso-Top (Cambridge Isotope Laboratories, Inc., Andover, England).

The isotopic abundance of each molecule was adjusted to the desired value by mixing labeled and unlabeled sources. We prepared “<sup>13</sup>C<sup>2</sup>H” (double-labeling) solutions

and “ $^{13}\text{C}^1\text{H}$ ” (mono-labeling) solutions for all molecules. The incubation characteristics are shown in Table S1 (Supplement). Mixing was performed gravimetrically.

For palmitic acid, the equivalent amount of unlabeled and labeled compound was added to 200 mg of soil and was melted at 70 °C. We finely ground the cooled mixture to obtain a homogenized powder that could be added to the incubators. Two powders were prepared: a  $^{13}\text{C}^2\text{H}$ -enriched powder and a  $^{13}\text{C}$ -enriched powder.

Three distinct labeling experiments were performed:

1. Experiment 1: “ $^{13}\text{C}^2\text{H} + \text{H}_2\text{O}$ ”: double-labeling molecule introduced to the soil with ultrapure water
2. Experiment 2: “ $^{13}\text{C}^1\text{H} + ^2\text{H}_2\text{O}$ ”: mono-labeling molecule introduced to the soil with deuterated water
3. Experiment 3: “no molecule +  $^2\text{H}_2\text{O}$ ”: only deuterated water introduced to the soil.

The final humidity of the soil was 30, 15 and 31 g g<sup>-1</sup> dry weight, respectively, for Cambisol, Podzol and Leptosol.

### 2.2.3 Incubations

The 300 incubators were incubated at 28 °C in the dark, and three were frozen at 0, 7, 14 and 28 days and 1 year for the Cambisol and at 0 and 7 days and 1 year for the two other soils. Jars were briefly opened (a few seconds for each sample) every 2 days during the first 3 weeks and then every week until the end of incubation to keep the system under aerobic conditions. The evaporation of water is very limited during this step. However, this leads to the renewal of water vapor in the jar head space by ambient atmosphere. The isotopic composition of  $^2\text{H}$  of the new atmosphere in the jar is depleted compared to the previous one. By taking the value of the saturation vapor pressure at 28 °C (28 g m<sup>-3</sup>), the amount of water contained in the headspace jar (0.17 dm<sup>3</sup>) was 4.8 mg. The proportion of the lost labeled water was estimated at 0.7 % in the first months and at 2 % for 1 year. The impact of the atmosphere renewal on the isotopic composition was therefore neglected.

Control incubators were prepared for each experiment at each time without any added substrate or deuterated water under the same incubation conditions.

To highlight the link between the NEH and carbon dynamics, we initially added three different amounts of labeled glucose to the Podzol and we analyzed the results after 7 days of incubation.

### 2.3 Lipid extraction

Lipids were extracted to isolate the organic non-exchangeable hydrogen. Lipids extractions were performed on samples that had received glucose and had been incubated for 1 year. Between 10 and 15 g of soil were subsampled and phosphate buffer, chloroform and methanol were

added (0.8:1:2;  $v:v:v$ ). After 2 min of ultrasound and 30 min of warming (37 °C), samples were centrifuged for 8 min at 2600 rpm. Supernatant was retrieved and stored at room temperature while chloroform and methanol (1:2;  $v:v$ ) were added to the remaining soil and centrifuged again. Supernatant was retrieved and added to the previous aliquot. Twenty milliliters of NaCl were then added to the supernatant to distinguish two phases. The denser part was collected and dried under nitrogen.

### 2.4 Isotopic measurements

Prior to analysis, incubated samples were freeze-dried during 28 h. Samples were then ground to a fine powder and kept in closed tubes under laboratory atmosphere.

Twenty to fifty-five milligrams of soil were then introduced into a 10 mm tin capsule.

Lipid samples were solubilized in dichloromethane before introducing them to the tin capsules. We let the solvent evaporate before the analysis. The mean isotopic signature of this bulk lipid fraction was measured using the same method as for the soil samples.

The  $^{13}\text{C}$  and  $^2\text{H}$  contents were analyzed simultaneously with a combustion module-cavity ring-down spectroscopy (CM-CRDS) isotope analyzer (Picarro, B2221-i). The organic standards polyethylene (IAEA CH7;  $\delta^{13}\text{C} = -32.15 \pm 0.05 \text{‰}$ ;  $\delta^2\text{H} = -100.3 \pm 2.0 \text{‰}$ ) and oil NBS-22 ( $\delta^{13}\text{C} = -30.03 \pm 0.05 \text{‰}$ ;  $\delta^2\text{H} = -119.6 \pm 0.6 \text{‰}$ ) were used to calibrate the measurements. A homemade standard (olive oil) was also used in each run ( $\delta^{13}\text{C} = -29.0 \pm 0.2 \text{‰}$ ;  $\delta^2\text{H} = -153 \pm 5 \text{‰}$ ). To validate the measurement of highly enriched samples by CM-CRDS, we compared the measurements at initial conditions (time 0 of the incubations, before the degradation of substrate occurred) to the theoretical isotopic composition at initial conditions obtained by calculation. Both the linearity of the  $\delta^2\text{H}$  measurement for enriched samples and the full recovery of labeled NEH during the drying process were confirmed by the measured vs. theoretical  $\delta^2\text{H}$  of initial labeled soil samples (mixtures of soil and labeled source before incubation), which yielded a slope of 1.02 and  $r^2 = 0.99$ .

To deal with the  $^2\text{H}$  memory effect often recorded with CM-CRDS (Koehler and Wassenaar, 2012), five repetitions were done for each sample; the last three were used for interpretation when standard deviation was less than 1.5 ‰ for natural samples and less than 10 ‰ for enriched samples. Moreover, we analyzed samples from the more depleted to the more enriched, and ashes were removed from the combustion tube each 45 samples to limit contamination.

The isotopic composition of  $^{13}\text{C}$  and  $^2\text{H}$  are expressed by abundance ( $A$ ) or as  $\delta$  (‰):

$$A^{13}\text{C} = \frac{^{13}\text{C}}{^{13}\text{C} + ^{12}\text{C}} \text{ and } A^2\text{H} = \frac{^2\text{H}}{^2\text{H} + ^1\text{H}}$$

$$\delta \text{‰} = [R_{\text{sample}}/R_{\text{standard}} - 1] \cdot 1000,$$

where  $R = {}^{13}\text{C}/{}^{12}\text{C}$  or  ${}^2\text{H}/{}^1\text{H}$ . The international standard was VPDB (Vienna Pee Dee Belemnite) for carbon and VS-MOW (Vienna Standard Mean Ocean Water) for hydrogen.

## 2.5 Quantification of NEH derived from the labeled source

### 2.5.1 Samples equilibration

We performed equilibration of the labeled and unlabeled samples with the same atmosphere to reach the same isotopic composition of exchangeable hydrogen in the dry soils. To do so, before analysis, unlabeled and labeled samples were equilibrated with the laboratory atmosphere for 2 h after soil grinding (exchanges also occur during the grinding  $\approx 20$  min and during evaporation by nitrogen flushing in the CM-CRDS introduction line). The differences in  $\delta^2\text{H}$  between unlabeled and labeled samples are a means to eliminate the contribution of labile hydrogen (short-time exchange) in the final isotopic calculation (see mass calculation below). Unlabeled and labeled samples received exactly the same treatment. Hydrogen that did not exchange during the length of equilibration is considered as non-exchangeable in this study.

### 2.5.2 Mass balance equations

Table 2 summarizes the different variables used in the mass balance equations and Supplement (S2) is provided for further understanding of calculations.

The carbon isotopic composition of the total bulk soil corresponds to the proportion of molecule-derived (labeled) and soil-derived (unlabeled) carbon (Eq. 1).

$$C_{\text{dfm}} = ({}^{13}A_{\text{tot}} - {}^{13}A_{\text{tot}_0}) / ({}^{13}A_{\text{m}} - {}^{13}A_{\text{tot}_0}) \cdot C_{\text{tot}} \quad (1)$$

Exchangeable hydrogen has the isotopic composition of the atmosphere when the sample is dry. Non-exchangeable hydrogen from the labeled source was estimated based on the simultaneous measurement of labeled and unlabeled samples equilibrated with the same atmosphere using Eqs. (2) and (3), which attributes all the excess deuterium (difference between the  ${}^2\text{H}$  abundances of the labeled sample and unlabeled control) to the NEH derived from the labeled source atoms (see Supplement S2 for the calculation).

In Experiment 1 (labeled molecule)

$$H_{\text{dfm}} = (A_{\text{tot}} - A_{\text{tot}_0}) / (A_{\text{m}} - A_{\text{tot}_0}) \cdot H_{\text{tot}} \quad (2)$$

In Experiment 2 (labeled water)

$$H_{\text{dfw}} = (A_{\text{tot}} - A_{\text{tot}_0}) / (A_{\text{w}} - A_{\text{tot}_0}) \cdot H_{\text{tot}} \quad (3)$$

The labeled source is highly enriched compared to natural soil or water:  ${}^{13}A_{\text{m}} = 6.08$  to  $16.08\%$ ,  $A_{\text{m}} = 2$  to  $3.5\%$  and  $A_{\text{w}} = 0.26\%$  (Supplement, Table S1).

The mean  ${}^2\text{H}$  abundance of unlabeled soil and water is approximately  $0.015\%$ , and the mean  ${}^{13}\text{C}$  abundance is approximately  $1.08\%$ .

Uncertainties in the element and isotope ratio measurements affect the estimate of the amount of labeled-source-derived carbon or hydrogen atoms. To assess the uncertainty in the calculated values  $H_{\text{dfm}}$  and  $C_{\text{dfm}}$ , we calculated the statistical error propagation of the uncertainties of the measured isotopic compositions and the element content of the replicated samples (Supplement S3).

## 3 Results

The raw isotopic composition of the highly enriched samples ( $\delta^{13}\text{C}$  and  $\delta^2\text{H}$  measurements) is presented over the incubation period in the Supplement (Sect. S4).

In unlabeled samples  $\delta^2\text{H}$  ranged from  $-43$  to  $+27\%$  for the Cambisol, from  $-58$  to  $-22\%$  for the Podzol and from  $-74$  to  $-43\%$  for the Leptosol with less than  $3\%$  variation.

In the first experiment (labeled molecule), measured  $\delta^2\text{H}$  values of labeled samples at initial conditions ranged from  $900$  to  $2500\%$  depending on the soil and the added molecule. They still reached  $85$  to  $576\%$  after 1 year of incubation.

$\delta^2\text{H}$  values of the second experiment (labeled water) ranged from  $0$  to  $260\%$  for the Cambisol, from  $-20$  to  $240\%$  for the Podzol and from  $40$  to  $1000\%$  for the Leptosol over the incubation period.

At each time step, as described in the methodology (see Sect. 2.5.2), the difference between the  ${}^2\text{H}$  isotopic composition of the labeled sample and the isotopic value of the unlabeled sample analyzed the same day is calculated. The differences between labeled and unlabeled samples are huge.

$\delta^{13}\text{C}$  values are between  $130$  and  $110\%$  at the beginning of the incubation and between  $0$  and  $15\%$  after 1 year, whereas unlabeled  $\delta^{13}\text{C}$  values were  $-26.2\%$  for the Cambisol,  $-25.4\%$  for the Podzol and  $-24.7\%$  for the Leptosol with less than  $0.3\%$  variation.

### 3.1 Comparison of the four substrates mineralization

The fates of labeled C or H atoms are presented as the mass of C or NEH derived from the labeled source, i.e., molecule or water ( $C_{\text{dfm}}$ ,  $H_{\text{dfm}}$ ,  $H_{\text{dfw}}$ ); see Eqs. (1), (2) and (3) (dfm and dfw stand for “derived from the molecule” and “derived from water”, respectively). We first tested the dependence on time (1 week and 1 year), molecule type and soil type on the basis of a three-way ANOVA of each explained variable. Both  $C_{\text{dfm}}$  and  $H_{\text{dfw}}$  were dependent on time ( $p < 0.001$ ) and soil ( $p < 0.001$ ) but not on molecule.  $H_{\text{dfm}}$  was dependent on time ( $p < 0.001$ ), soil type ( $p < 0.001$ ) and molecule ( $p < 0.001$ ). The differences in results for  $H_{\text{dfm}}$  can be explained by the uncertainty in experiments and measurements. Because we found no significant differences between the molecules for  $C_{\text{dfm}}$  and  $H_{\text{dfw}}$ , we considered the different molecule incubations as replicates to simplify the presentation of results, with only the mean values shown in graphs.

**Table 2.** Definition of the variable used in calculations.

	Variables	Definition
Quantity (mg g <sup>-1</sup> of dry soil)	H <sub>tot</sub>	Total amount of hydrogen in the soil
	C <sub>tot</sub>	Total amount of carbon in the soil
	H <sub>dfw</sub>	Amount of non-exchangeable hydrogen derived from water
	H <sub>dfm</sub>	Amount of non-exchangeable hydrogen derived from molecule
	C <sub>dfm</sub>	Amount of carbon derived from molecule
	H <sub>m</sub>	Initial amount of non-exchangeable hydrogen in the added molecule
	C <sub>m</sub>	Initial amount of carbon in the added molecule
Abundance	<sup>13</sup> A <sub>tot_0</sub>	<sup>13</sup> C abundance of the unlabeled experiment (control)
	A <sub>tot_0</sub>	<sup>2</sup> H abundance of the unlabeled experiment (control)
	<sup>13</sup> A <sub>tot</sub>	<sup>13</sup> C abundance of the total bulk soil
	A <sub>tot</sub>	<sup>2</sup> H abundance of the total bulk soil
	<sup>13</sup> A <sub>m</sub>	Initial <sup>13</sup> C abundance of the labeled molecule
	A <sub>m</sub>	Initial <sup>2</sup> H abundance of the labeled molecule
	A <sub>w</sub>	Initial <sup>2</sup> H abundance of the labeled water

The difference in degradation between molecules is therefore contained in the error bars in Figs. 1 and 2.

### 3.2 Carbon mineralization

Figure 1 shows the amount of carbon and hydrogen derived from the initial molecule in the soil: the amount of molecule-derived carbon (C<sub>dfm</sub>) and molecule-derived hydrogen (H<sub>dfm</sub>) as calculated in Eqs. (1) and (2) is relative to the amount of C<sub>m</sub> and H<sub>m</sub> in the added molecule. The degradation of the added molecule was very fast. After 7 days, 42, 31 and 53 % of molecule-derived carbon remained in the Cambisol, Podzol and Leptosol, respectively (Fig. 1). This trend is in agreement with previous studies (Murayama, 1988; Derrien et al., 2007) and illustrates the almost complete consumption of the substrate in a few days. Approximately 30 to 50 % of the consumed material was converted into microbial products, and the remaining part was used for heterotrophic respiration. During the following months, the mineralization of organic carbon continued due to the partial consumption of the newly formed microbial carbon by the soil food web.

During the incubation, non-labeled carbon (soil-derived carbon) also decreased by 1.8, 2.4 and 4.5 mg g<sup>-1</sup> within 1 year in the Cambisol, Podzol and Leptosol, respectively.

### 3.3 Molecule-derived non-exchangeable hydrogen

After incubation, molecule-derived NEH (H<sub>dfm</sub>) was considerably lower than molecule-derived carbon for the Cambisol and for the Podzol (Fig. 1): expressed in percentage of the initial labeled NEH present in the molecule, it was 12 and 5 % compared to the 42 and 31 % recorded for carbon. It is important to note the different fate of the Leptosol, where the yield of transfer of NEH reached 55 % after 7 days of incubation (Fig. 1). During the following months, H<sub>dfm</sub> slightly

decreased in the three soils (by approximately 6 % ± 5 after 1 year).

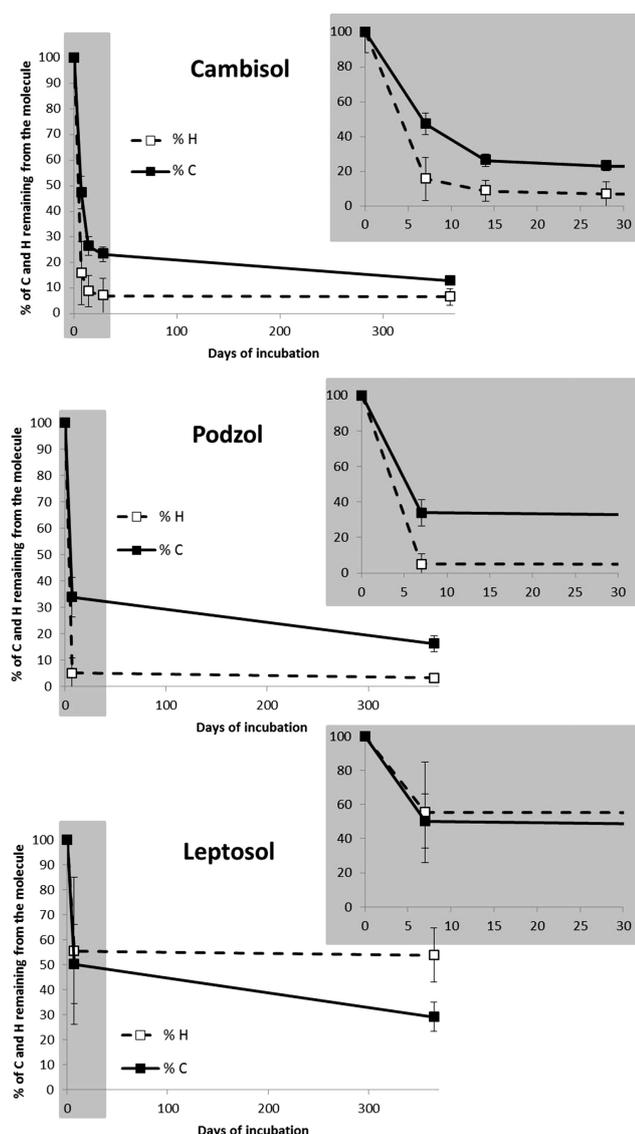
### 3.4 Incorporation of water hydrogen

Experiment 2 (molecule <sup>13</sup>C<sup>1</sup>H + <sup>2</sup>H<sub>2</sub>O) highlights the incorporation of water hydrogen in the non-exchangeable pool of soil calculated in Eq. (3), called water-derived hydrogen. For the three soils, the incorporation of hydrogen from water increased during the first 7 days and continued to slowly increase during the incubation year (Fig. 2). Respectively, 0.013 ± 0.001, 0.008 ± 0.002 and 0.33 ± 0.07 mg g<sup>-1</sup> of hydrogen derived from water was found after 7 days of incubation, and 0.06 ± 0.03, 0.023 ± 0.004 and 0.845 ± 0.003 mg g<sup>-1</sup> was found after 1 year of incubation for Cambisol, Podzol and Leptosol (Fig. 2).

Figure 3 shows the difference in the incorporation of water hydrogen with and without added substrate. The incorporation of water-derived hydrogen was higher when associated with substrate addition. It was twice as high for the Podzol after 7 days of incubation. Figure 4 illustrates that this enhancement of the incorporation of water hydrogen was linearly dependent on the amount of the substrate added to the soil.

### 3.5 Isotopic composition of lipids

Carbon and hydrogen isotopic compositions of bulk lipids at 365 days for the control soil are presented in Table 3. The proportions of labeled carbon and hydrogen were calculated as the proportion of the total lipid carbon and hydrogen ((C<sub>dfm</sub>/C)<sub>lipids</sub>, (H<sub>dfm</sub>/H)<sub>lipids</sub> and (H<sub>dfw</sub>/H)<sub>lipids</sub>) and were compared to the proportion in the bulk soil ((C<sub>dfm</sub>/C)<sub>bulk</sub>, (H<sub>dfm</sub>/H)<sub>bulk</sub> and (H<sub>dfw</sub>/H)<sub>bulk</sub>). The δ<sup>13</sup>C and δ<sup>2</sup>H of the lipids in the control samples were lower than that of the



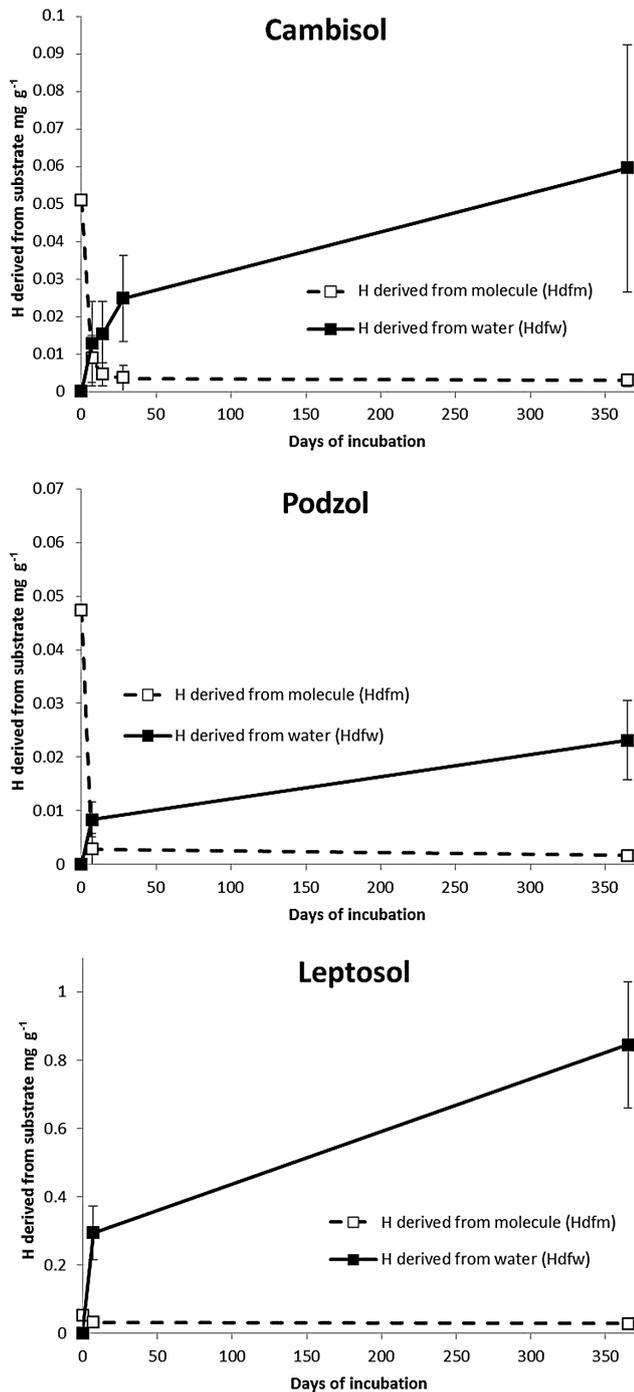
**Figure 1.** Percentage of non-exchangeable hydrogen ( $H_{dfm}/H_m \cdot 100$ ) and carbon ( $C_{dfm}/C_m \cdot 100$ ) remaining from the added molecule in the total soil during 1 year of incubation for Cambisol, Podzol and mollic Leptosol. The grey part corresponds to the results from 0 to 28 days. The line corresponds to the mean value calculated at each time for all molecule experiments.

bulk soil, in agreement with previous work, where the lipid  $\delta^{13}C$  was 2–3‰ lower than the bulk  $\delta^{13}C$  (Chikaraishi and Naraoka, 2001; Hayes, 2001), and the lipid  $\delta^2H$  was 150‰ lower than the bulk  $\delta^2H$  (Sessions et al., 1999; Chikaraishi and Naraoka, 2001). The average measured H / C ratio of the lipids of the three soils was 2.1 (molar ratio).

The proportion of molecule-derived carbon in the lipids was 1.0, 0.4 and 0.8% for the Cambisol, Podzol and Leptosol, respectively, compared to the corresponding values of 0.6, 0.4 and 0.4% in the bulk organic carbon. The proportion of molecule-derived hydrogen was, respectively, 0.10, 0.02

**Table 3.**  $\delta^{13}C$ ,  $\delta^2H$  of bulk soil and lipids at 365 days of incubation for Cambisol, Podzol and Leptosol and the proportion of carbon and hydrogen derived from the labeled source. In brackets the concentration in  $mg\ g^{-1}$  of carbon and hydrogen derived from the labeled source.

	$\delta^{13}C\text{‰}$		$\delta^2H\text{‰}$		Percentage and concentration ( $mg\ g^{-1}$ ) of the labeled source in lipids			
	Bulk	Lipids	Bulk	Lipids	% $C_{dfm}$ ( $mg\ g^{-1}$ )	% $H_{dfw}$ ( $mg\ g^{-1}$ )	% $C_{dfm}$ ( $mg\ g^{-1}$ )	% $H_{dfw}$ ( $mg\ g^{-1}$ )
<b>Cambisol</b>								
Experiment 1	5.3 ± 0.9	19 ± 3	179 ± 2	133 ± 43	0.6% (0.054 ± 0.009)	0.082% (0.0031 ± 0.0003)	1% (3.3 ± 0.6)	0.1% (0.13 ± 0.01)
Experiment 2	5 ± 1	17 ± 11	422 ± 15	36 ± 6	0.6% (0.055 ± 0.009)	–	1% (4.2 ± 0.8)	1% (0.68 ± 0.09)
<b>Podzol</b>								
Experiment 1	13.2 ± 0.7	7 ± 5	94 ± 39	–106 ± 42	0.4% (0.07 ± 0.01)	0.07% (0.0017 ± 0.0007)	0.4% (1.8 ± 0.9)	0.02% (0.05 ± 0.02)
Experiment 2	7.0 ± 0.7	6 ± 7	259 ± 25	–38 ± 55	0.4% (0.07 ± 0.01)	–	0.4% (1.4 ± 0.1)	–
<b>Leptosol</b>								
Experiment 1	22 ± 3	97 ± 2	723 ± 10	458 ± 65	0.4% (0.15 ± 0.03)	0.24% (0.029 ± 0.003)	0.8% (5.9 ± 0.2)	0.19% (0.18 ± 0.01)
Experiment 2	25 ± 1	76 ± 20	798 ± 58	23 ± 5	0.4% (0.14 ± 0.03)	–	0.8% (4.8 ± 0.8)	1.1% (0.95 ± 0.03)



**Figure 2.** Concentration of water-derived ( $H_{dfw}$ ) and molecule-derived ( $H_{dfm}$ ) non-exchangeable hydrogen in the total soil for Cambisol, Podzol and mollic Leptosol from 0 to 365 days. The line corresponds to the mean value calculated at each time for all molecule experiments. Note that the scale of y axis is different for the three soils.

and 0.19 % for the Cambisol, Podzol and Leptosol (Table 3). These values were on the same order of magnitude as the molecule-derived hydrogen in the bulk soil. The proportion

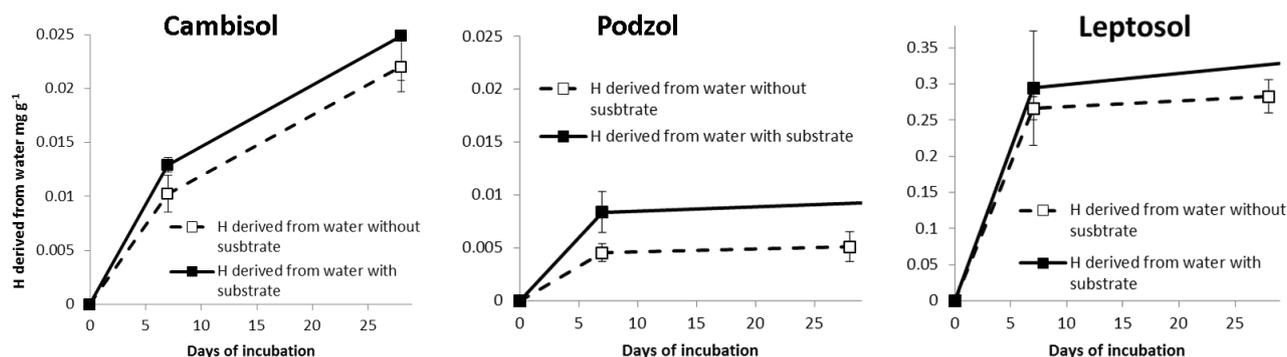
of labeled water-derived hydrogen was 1.0, 0.4 and 1.1 % of the total hydrogen content in the lipids and 1.5, 0.9 and 6.8 % in the total bulk soil for the Cambisol, Podzol and Leptosol, respectively.

## 4 Discussion

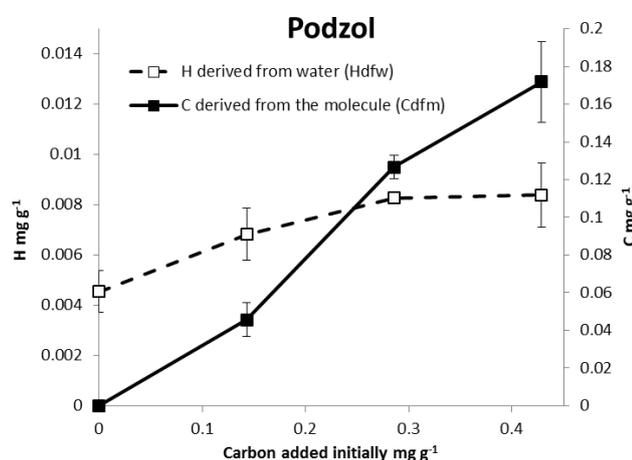
### 4.1 Preservation of the organic substrate hydrogen in biosyntheses

The microbial activity is initiated during the first days after the addition of the substrate. The added molecule regardless of its quality is quickly metabolized (Fig. 1).

We independently traced the preservation of organic hydrogen (Experiment 1) and the incorporation of water-derived hydrogen (Experiment 2) during decomposition and biosynthesis. The conservation of organic hydrogen from the initial substrate is very low in both the total and lipid NEH. The carbon–hydrogen bonds are broken during decomposition, and exchange with exchangeable hydrogen can occur. The difference between carbon and hydrogen isotopic fates during the first 7 days (Fig. 1) reflects the exchange of hydrogen with water during the early stage of degradation. Subsequently, new organic exchangeable hydrogen derived from water can be incorporated into the non-exchangeable pool of organic matter by biological processes. Furthermore, because mineralization of a substrate also results in  $^2\text{H}_2\text{O}$  release, one part of the soil organic non-exchangeable  $^2\text{H}$  may originate from incorporation of this substrate-derived deuterated water into the non-exchangeable pool. Using the assumption that water is a well-mixed isotopic compartment, this amount is between 3 and 7 % of the residual hydrogen from the organic substrate at 365 days (see Supplement for calculation). The isotopic composition of the non-exchangeable organic hydrogen is mainly determined by the water isotopic composition (Fig. 2). It is in accordance with the work of Baillif and colleagues that have grown fungi with labeled glucose, water and acetate to trace the incorporation of  $^2\text{H}$  during fatty acid biosynthesis. They have demonstrated that water is the main donor of hydrogen atoms in the non-exchangeable pool within the biosynthesis cycle (Baillif et al., 2009). Ruppenthal and colleagues have shown as well that precipitation contributes up to 80 % of the isotopic composition of non-exchangeable hydrogen (Ruppenthal et al., 2010). Moreover, the incorporation of water hydrogen is favored by the strength of the C–H bond breakage. It can be weak or strong, depending on the enzyme activated for the degradation of the molecule and the position of the bond (Augusti et al., 2006). When C–H breakdown is favored, the surrounding water imprints its hydrogen isotopic signature on the former bounded H (Augusti et al., 2006). In the present experiment, we show that more than 70 % of the H–C bonds of the initial molecule are broken in the first biodegradation steps (7 days, Fig. 1); therefore, the added molecules are used more for energetic



**Figure 3.** Concentration of non-exchangeable hydrogen derived from water with and without addition of substrate in the total soil for the Cambisol, Podzol and Leptosol from 0 to 28 days of incubation. Note that the scale of y axis is different for the three soils.



**Figure 4.** Amount of substrate-derived H and C at 7 days of incubation vs. the concentration of substrate as carbon (0.14, 0.29 and 0.43 mg g<sup>-1</sup>). The concentration C in Experiment 1 is equal to 0.43 mg g<sup>-1</sup>.

and trophic resources than as building blocks in the biosynthesis.

Deuterium can also accumulate in hydration shells, which have stronger hydrogen bridges than the biomolecule (Baumgärtner and Donhaerl, 2004). The accumulation of deuterium from water occurs in the biomatter during biological processes but also during the hydration of molecules (Baumgärtner and Donhaerl, 2004; Turner et al., 2009).

#### 4.2 Carbon-driven acquisition of the non-exchangeable hydrogen isotope signature

The rapid mineralization of hydrogen from the molecule is due to biodegradation, whereas the rapid incorporation of water during the first 7 days of incubation is associated with biosynthesis (Figs. 1 and 2). Results obtained on the lipid fraction provide evidence of the formation of organic non-exchangeable hydrogen from H<sub>dfw</sub> (Table 3). Carbon min-

eralization fosters the formation of non-exchangeable hydrogen from water (Figs. 3 and 4). There is also an incorporation of hydrogen from water in the soil in the experiments without substrate. This could be due to the microbial transformation of carbon already present in soil (Fig. 3) but also to the inorganic fraction that could incorporate hydrogen from water. In Fig. 2, exchange of hydrogen with water seems to be continuous: H<sub>dfw</sub> continues increasing, following carbon mineralization. The incorporation of water hydrogen in the organic non-exchangeable pool but also in the inorganic fraction occurs throughout the incubation.

Results of lipids isotopic compositions show that the amount of newly formed NEH (% of H<sub>dfw</sub> + % of H<sub>dfm</sub> in lipids; Table 3) is slightly higher than expected from the theoretical organic C–H bond (% of C<sub>dfm</sub> in lipids; Table 3). This could be due to complete, stoichiometric labeling of newly biosynthesized lipids, i.e., lipids formed on the labeled organic carbon plus a smaller amount of newly synthesized lipids from unlabeled organic matter. The proportion of molecule-derived carbon is higher in the lipid than in the bulk soil, and the lipids are derived mainly from microbial biosynthesis. However, the proportions of H<sub>dfm</sub> in the lipids and in the bulk soil are of the same magnitude for the three soils (Table 3): hydrogen is derived evenly from the labeled molecule and from the unlabeled soil during lipid biosynthesis.

The proportion of hydrogen derived from the water in lipids is lower than the respective proportion in the bulk soil, which means the proportion of H<sub>dfw</sub> is not necessarily organic. It could be inorganic hydroxyl, hydrated ions or water in different states. The difference is even higher in the clayey soil (Leptosol; Table 3) that contains the highest proportion of hydrogen (Table 1).

In lipids, hydrogen corresponds to the organic, non-exchangeable hydrogen. The hydrogen in lipids formed from water (H<sub>dfw</sub>) is only organic, whereas in the bulk soil, the H<sub>dfw</sub> is organic and inorganic. To estimate the proportion of organic non-exchangeable hydrogen, we assume that H<sub>dfw</sub>/C<sub>dfm</sub> in lipids (0.20 on average for the three soils based

on mass ratio; Table 3) is approximately the same as the  $H_{dfw}/C_{dfm}$  in the organic fraction of the bulk soil. Using the measured  $C_{dfm}$  in bulk soil, we can then estimate the total organic  $H_{dfw}$  as 0.014, 0.015 and 0.028 mg g<sup>-1</sup>, respectively, for the Cambisol, Podzol and Leptosol. The proportion of inorganic  $H_{dfw}$  is therefore 0.046, 0.008 and 0.82 mg g<sup>-1</sup> for the three soils. The NEH isotopic composition is mainly controlled by the incorporation of water through biosynthesis, but the inorganic  $H_{dfw}$  is not negligible, especially in the clayey soil.

### 4.3 Hydrogen dynamics in different soil types

The association of organic matter with minerals is known to decrease the decomposition rate of the former (Feng et al., 2013; Jenkinson and Coleman, 2008; Vogel et al., 2014). This result is observed in our experiment by comparing the three soils with increasing clay content and is applicable to both H and C in both bulk soil (Fig. 1) and lipids (Table 3). However, the clay content has an important role in the incorporation of water-derived hydrogen beyond this organic matter stabilization effect. In clayey Leptosol, the amount of labeled NEH from the molecule ( $H_{dfm}$ ; Fig. 1) is much higher than in the other soils, which may be explained by the preferential use of hydrogen locally near biological reactions. Hydrogen derived from the mineralization of the substrate does not directly exchange with the total pool of water but with a smaller pool. The resulting local water pool has a less negative isotopic signature than the remaining water pool. Water incorporation through biosynthesis could then occur with this smaller pool of <sup>2</sup>H-enriched water. Moreover, hydrogen exchange within the whole water pool is slowed by the presence of clay, which accumulates molecule-derived hydrogen in hydroxyl sites. For this reason,  $H_{dfw}$  is also much higher in Leptosol than in the other soils. The  $H_{dfw}$  pool in Leptosol is bigger because of the presence of inorganic  $H_{dfw}$ . The non-exchangeable hydrogen pool considered in the present equilibration method contains H in various positions, and it may include water molecules acting as non-exchangeable hydrogen. The inorganic  $H_{dfw}$  pool accounts for less than 1/40 of the amount of water H in the Leptosol. This pool appears as very slowly exchangeable with “free water”. In situ, on a short-term dynamics scale, this pool acts as non-exchangeable and is mostly at the hydroxyl position (López-Galindo et al., 2008).

The zonal distribution of organic compounds associated with minerals (Kleber et al., 2007; Vogel et al., 2014) may control the exchange between soil solution and organic compounds at kinetics that differ according to the layer within the organo-mineral interaction zone. The non-exchangeable hydrogen dynamics in soil organic matter are not independent of the mineral structure. The type of clay plays a role in carbon sequestration, depending on the specific surface area of the mineral or aggregate (Vogel et al., 2014; López-Galindo et al., 2008). In Leptosol, clays are mainly smectite and have

a high specific surface area. The high content of iron and hydroxide present in the Leptosol also increases the specific surface area of the aggregates, which increases the organo-mineral association (Baldock and Skjemstad, 2000). Organic carbon cycling itself may be associated with mineral transformation (Basile-Doelsch et al., 2015), which may involve the newly formed hydroxyl.

The short-term dynamics of hydrogen are driven by the incorporation of hydrogen from water by isotopic exchange and by microbial biosynthesis. However, the increase in the incorporation of water hydrogen with the soil clay content suggests that part of the hydrogen is bound to clay or organo-mineral complexes. The production of NEH from water occurs mainly during the first weeks, but slow exchange of water hydrogen continues during the following year. López-Galindo et al. (2008) observed the same trend, and they related the accumulation rate to the clay mineral properties.

### 4.4 Ecosystem-scale production and fate of non-exchangeable hydrogen

In the present experiment, the preservation of non-exchangeable hydrogen from an organic substrate is less than 5 % after 1 year in soil with a low clay content. Water is the main donor of hydrogen during the biosynthesis cycle favored by the breakage of the C–H bonds of the initial substrate. In this work, we showed that 70 % of the C–H bonds of the initial substrate were broken during biosynthesis. Concerning the fate of tritium in terrestrial ecosystems, the isotopic composition of the organic plant material is a minor determinant of the bulk soil organic matter composition. However, a better preservation of hydrogen from vegetation could arise from a higher soil clay content and the subsequent organo-mineral and zonal interactions. Water will be the main donor of organically bound tritium in the soil and the incorporation will be dependent on the carbon mineralization. Lipid isotopic composition highlighted that (i) water derived hydrogen is involved in biosynthesis and (ii) the newly formed non-exchangeable hydrogen is not necessarily organic. Therefore, the incorporation of tritium from water in the NEH pool is dependent on the clay content and on the soil hydrodynamics.

In our work, the isotopic composition of the NEH pool is determined by comparing labeled samples with unlabeled samples equilibrated under the same atmosphere. This method includes inorganic NEH. A proportion of the inorganic, non-exchangeable hydrogen should be taken into account in the prediction of the dynamics of hydrogen and tritium.

The main finding of the work is that the long-term fate of hydrogen in terrestrial environments and by extension the fate of tritium will depend on the status of soil carbon dynamics.

The proportion of NEH associated with minerals is itself partially related to the carbon dynamics.

Both the carbon dynamics and the incorporation of inorganic hydrogen in soils should therefore be taken into account in a conceptual model for the prediction of the long-term fate of hydrogen, and thereafter of tritium, in soil organic matter. The results of the present study can be used for the parameterization of the carbon–hydrogen coupling in such prediction models.

## 5 Data availability

The raw data are available in the Supplement as figures. More details can be obtained by contacting the author (alexia.paul@inra.fr).

## Information about the Supplement

The Supplement includes the summary of incubation characteristics (Table S1), mass balance calculation (S2), propagation error calculation (S3), and  $\delta^{13}\text{C}$  and  $\delta^2\text{H}$  results of the incubation samples (S4).

**The Supplement related to this article is available online at doi:10.5194/bg-13-6587-2016-supplement.**

*Acknowledgements.* This study was funded by EDF and Andra. We would like to thank Sabine Maestri for her great help with the isotopic measurements. We thank the following research teams for having kindly put at our disposition their long-term experiments and associated data: INRA-Agro-Paristech ECOSYS and INRA-UVSQ GCVG (Les Closeaux SolFit experiment in Versailles); INRA Unité Expérimentale Forêt Pierroton and INRA ISPA (Pierroton forest reserve and fertilization experiment); OSU-Pytheas (O3HP LTER in Saint-Michel l'Observatoire). We sincerely thank the anonymous referees for their relevant suggestions to improve the manuscript. This is a LSCE contribution no. 5959.

Edited by: S. Bouillon

Reviewed by: two anonymous referees

## References

Augusti, A., Betson, T. R., and Scleucher, J.: Hydrogen exchange during cellulose synthesis distinguishes climatic and biochemical isotope fractionations in tree rings, *New Phytol.*, 172, 490–499, 2006.

Baillif, V., Robins, R. J., Le Feunteun, S., Lesot, P., and Billault, I.: Investigation of Fatty Acid Elongation and Desaturation Steps in *Fusarium lateritium* by Quantitative Two-dimensional Deuterium NMR Spectroscopy in Chiral Oriented Media, *J. Biol. Chem.*, 284, 10783–10792, doi:10.1074/jbc.M807826200, 2009.

Baldock, J. A. and Skjemstad, J. O.: Role of the soil matrix and minerals in protecting natural organic materials against biological attack, *Org. Geochem.*, 31, 697–710, doi:10.1016/S0146-6380(00)00049-8, 2000.

Balesdent, J., Mariotti, A., and Guillet, B.: Natural  $^{13}\text{C}$  abundance as a tracer for studies of soil organic matter dynamics, *Soil Biol. Biochem.*, 19, 25–30, 1987.

Basile-Doelsch, I., Balesdent, J., and Rose, J.: Are Interactions between Organic Compounds and Nanoscale Weathering Minerals the Key Drivers of Carbon Storage in Soils?, *Environ. Sci. Technol.*, 49, 3997–3998, doi:10.1021/acs.est.5b00650, 2015.

Baumgärtner, F. and Donhaerl, W.: Non-exchangeable organically bound tritium (OBT): its real nature, *Anal. Bioanal. Chem.*, 379, 204–209, 2004.

Belot, Y.: Tritium in plants: A review, *Radiat. Prot. Dosimetry*, 16, 101–105, 1986.

Bol, R., Poirier, N., Balesdent, J., and Gleixner, G.: Molecular turnover time of soil organic matter in particle-size fractions of an arable soil, *Rapid Commun. Mass Spectrom.*, 23, 2551–2558, 2009.

Chikaraishi, Y. and Naraoka, H.: Organic hydrogen-carbon isotope signatures of terrestrial higher plants during biosynthesis for distinctive photosynthetic pathways, *Geochem. J.*, 35, 451–458, doi:10.2343/geochemj.35.451, 2001.

Ciffroy, P., Siclet, F., Damois, C., and Luck, M.: A dynamic model for assessing radiological consequences of tritium routinely released in rivers. Application to the Loire River, *J. Environ. Radioact.*, 90, 110–139, 2006.

Derrien, D., Marol, C., and Balesdent, J.: The dynamics of neutral sugars in the rhizosphere of wheat. An approach by  $^{13}\text{C}$  pulse-labelling and GC/C/IRMS, *Plant Soil*, 267, 243–253, doi:10.1007/s11104-005-5348-8, 2004.

Derrien, D., Marol, C., and Balesdent, J.: Microbial biosyntheses of individual neutral sugars among sets substrates and soils, *Geoderma*, 139, 190–198, 2007.

Diabaté, S. and Strack, S.: Organically Bound Tritium in Wheat after Short-Term Exposure to Atmospheric Tritium under Laboratory Conditions, *J. Environ. Radioact.*, 36, 157–175, 1997.

Dignac, M., Bahri, H., Rumpel, C., Rasse, D., and Bardoux, G.: Carbon-13 natural abundance as a tool to study the dynamics of lignin monomers in soil: an appraisal at the Closeaux experimental field (France), *Geoderma*, 128, 3–17, 2005.

Epstein, S., Yapp, C. J., and Hall, J. H.: The determination of the D/H ratio of non-exchangeable hydrogen in cellulose extracted from aquatic and land plants, *Earth Planet. Sc. Lett.*, 30, 241–251, 1976.

Eyrolle-Boyer, F., Boyer, P., Claval, D., Charmasson, S., and Cossonnet, C.: Apparent enrichment of organically bound tritium in rivers explained by the heritage of our past, *J. Environ. Radioact.*, 136, 162–168, doi:10.1016/j.jenvrad.2014.05.019, 2014.

Feng, W., Plante, A. F., and Six, J.: Improving estimates of maximal organic carbon stabilization by fine soil particles, *Biogeochemistry*, 112, 81–93, doi:10.1007/s10533-011-9679-7, 2013.

Gontier, G. and Siclet, F.: Le tritium organique dans les écosystèmes d'eau douce: évolution à long terme dans l'environnement des centres nucléaires de production d'électricité français, *Radioprotection*, 46, 457–491, doi:10.1051/radiopro/2011139, 2011.

Hayes, J. M.: Fractionation of Carbon and Hydrogen Isotopes in Biosynthetic Processes, *Rev. Mineral. Geochem.*, 43, 225–277, doi:10.2138/gsrmg.43.1.225, 2001.

Jenkinson, D. S.: Studies on the decomposition of plant material in soil. I. Losses of carbon from  $^{14}\text{C}$  labeled ryegrass incubated with soil in the field, *J. Soil Sci.*, 16, 104–115, 1965.

- Jenkinson, D. S. and Coleman, K.: The turnover of organic carbon in subsoils. Part 2. Modelling carbon turnover, *Eur. J. Soil Sci.*, 59, 400–413, 2008.
- Jolivet, C., Angers, D. A., Chantigny, M. H., Andreux, F. and Arrouays, D.: Carbohydrate dynamics in particle-size fractions of sandy spodosols following forest conversion to maize cropping, *Soil Biol. Biochem.*, 38, 2834–2842, doi:10.1016/j.soilbio.2006.04.039, 2006.
- Kelleher, B. P. and Simpson, A. J.: Humic Substances in Soils: Are They Really Chemically Distinct?, *Environ. Sci. Technol.*, 40, 4605–4611, doi:10.1021/es0608085, 2006.
- Kim, S. B., Bredlaw, M., and Korolevych, V. Y.: HTO and OBT activity concentrations in soil at the historical atmospheric HT release site (Chalk River Laboratories), *J. Environ. Radioact.*, 103, 34–40, doi:10.1016/j.jenvrad.2011.08.013, 2012.
- Kim, S. B., Baglan, N., and Davis, P.: Current understanding of organically bound tritium (OBT) in the environment, *J. Environ. Radioact.*, 126, 83–91, 2013.
- Kleber, M., Sollins, P., and Sutton, R.: A conceptual model of organo-mineral interactions in soils: self-assembly of organic molecular fragments into zonal structures on mineral surfaces, *Biogeochemistry*, 85, 9–24, doi:10.1007/s10533-007-9103-5, 2007.
- Lehmann, J. and Kleber, M.: The contentious nature of soil organic matter, *Nature*, 528, 60–68, doi:10.1038/nature16069, 2015.
- López-Galindo, A., Hach-Alí, P. F., Pushkarev, A. V., Lytovchenko, A. S., Baker, J. H., and Pushkarova, R. A.: Tritium redistribution between water and clay minerals, *Appl. Clay Sci.*, 39, 151–159, doi:10.1016/j.clay.2007.06.005, 2008.
- Murayama, S.: Microbial synthesis of saccharides in soils incubated with  $^{13}\text{C}$ -labelled glucose, *Soil Biol. Biochem.*, 20, 193–199, 1988.
- Rafter, T. A. and Stout, J. D.: Radiocarbon measurements as an index of the rate of turnover of organic matter in forest and grassland ecosystems in New Zealand, in: *Radiocarbon Variations and Absolute Chronology*, edited by: Olsson, I. U., Nobel symposium, 7th Proc., 401–418, Stockholm, 1970.
- Ruppenthal, M., Oelmann, Y. and Wilcke, W.: Isotope ratios of nonexchangeable hydrogen in soils from different climate zones, *Geoderma*, 155, 231–241, 2010.
- Sauer, P. E., Schimmelmann, A., Sessions, A. L., and Topalov, K.: Simplified batch equilibration for D/H determination of non-exchangeable hydrogen in solid organic material, *Rapid Commun. Mass Sp.*, 23, 949–956, doi:10.1002/rcm.3954, 2009.
- Savin, S. and Epstein, S.: The oxygen and hydrogen isotope geochemistry of clay minerals, *Geochim. Cosmochim. Ac.*, 94, 25–42, 1970.
- Schimmelmann, A.: Determination of the concentration and stable isotopic composition of nonexchangeable hydrogen in organic matter, *Anal. Chem.*, 63, 2456–2459, 1991.
- Schimmelmann, A., Lewan, M. D., and Wintsch, R. P.: D/H isotope ratios of kerogen, bitumen, oil, and water in hydrous pyrolysis of source rocks containing kerogen types I, II, IIS, and III, *Geochim. Cosmochim. Ac.*, 63, 3751–3766, doi:10.1016/S0016-7037(99)00221-5, 1999.
- Schimmelmann, A., Sessions, A. L., and Mastalerz, M.: Hydrogen Isotopic ( $\text{d}/\text{H}$ ) Composition of Organic Matter During Diagenesis and Thermal Maturation, *Annu. Rev. Earth Planet. Sc.*, 34, 501–533, doi:10.1146/annurev.earth.34.031405.125011, 2006.
- Schmidt, M. W. I., Torn, M. S., Abiven, S., Dittmar, T., Guggenberger, G., Janssens, I. A., Kleber, M., Kögel-Knabner, I., Lehmann, J., Manning, D. A. C., Nannipieri, P., Rasse, D. P., Weiner, S., and Trumbore, S. E.: Persistence of soil organic matter as an ecosystem property, *Nature*, 478, 49–56, doi:10.1038/nature10386, 2011.
- Sessions, A. L., Burgoyne, T. W., Schimmelmann, A., and Hayes, J. M.: Fractionation of hydrogen isotopes in lipid biosynthesis, *Org. Geochem.*, 30, 1193–1200, doi:10.1016/S0146-6380(99)00094-7, 1999.
- Sessions, A. L., Sylva, S. P., Summons, R. E., and Hayes, J. M.: Isotopic exchange of carbon-bound hydrogen over geologic timescales, *Geochim. Cosmochim.*, 68, 1545–1559, 2004.
- Sutton, R. and Sposito, G.: Molecular structure in soil humic substances: the new view, *Environ. Sci. Technol.*, 39, 9009–9015, 2005.
- Thompson, P. A., Kwamena, N.-O. A., Ilin, M., Wilk, M., and Clark, I. D.: Levels of tritium in soils and vegetation near Canadian nuclear facilities releasing tritium to the atmosphere: implications for environmental models, *J. Environ. Radioact.*, 140, 105–113, doi:10.1016/j.jenvrad.2014.11.009, 2015.
- Turner, A., Millward, G. E., and Stemp, M.: Distribution of tritium in estuarine waters: the role of organic matter, *J. Environ. Radioact.*, 100, 890–895, doi:10.1016/j.jenvrad.2009.06.018, 2009.
- Tuthorn, M., Zech, R., Ruppenthal, M., Oelmann, Y., Kahmen, A., del Valle, H. F., Eglinton, T., Rozanski, K., and Zech, M.: Coupling  $\delta^2\text{H}$  and  $\delta^{18}\text{O}$  biomarker results yields information on relative humidity and isotopic composition of precipitation – a climate transect validation study, *Biogeosciences*, 12, 3913–3924, doi:10.5194/bg-12-3913-2015, 2015.
- Vogel, C., Mueller, C. W., Höschel, C., Buegger, F., Heister, K., Schulz, S., Schloter, M., and Kögel-Knabner, I.: Submicron structures provide preferential spots for carbon and nitrogen sequestration in soils, *Nat. Commun.*, 5, 2947, doi:10.1038/ncomms3947, 2014.
- Wassenaar, L. I. and Hobson, K. A.: Improved Method for Determining the Stable-Hydrogen Isotopic Composition ( $\text{dD}$ ) of Complex Organic Materials of Environmental Interest, *Environ. Sci. Technol.*, 34, 2354–2360, 2000.
- Wassenaar, L. I. and Hobson, K. A.: Comparative equilibration and online technique for determination of non-exchangeable hydrogen of keratins for use in animal migration studies, *Isotopes Environ. Health Stud.*, 39, 211–217, doi:10.1080/1025601031000096781, 2003.
- Zech, M., Zech, R., Rozanski, K., Hemp, A., Gleixner, G., and Zech, W.: Revisiting Mt. Kilimanjaro: Do n-alkane biomarkers in soils reflect the  $\delta^2\text{H}$  isotopic composition of precipitation?, *Biogeosciences Discuss.*, 11, 7823–7852, doi:10.5194/bgd-11-7823-2014, 2014.