

# Benthic foraminifera and gromiids from oxygen-depleted environments – survival strategies, biogeochemistry and trophic interactions

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Received: 3 March 2023 – Discussion started: 7 March 2023 Accepted: 30 June 2023 – Published: 17 August 2023

Abstract. The oceans are losing oxygen (O<sub>2</sub>), and oxygen minimum zones are expanding due to climate warming (lower O<sub>2</sub> solubility) and eutrophication related to agriculture. This trend is challenging for most marine taxa that are not well adapted to  $O_2$  depletion. For other taxa this trend might be advantageous because they can withstand low O<sub>2</sub> concentrations or thrive under O<sub>2</sub>-depleted or even anoxic conditions. Benthic foraminifera are a group of protists that include taxa with adaptations to partly extreme environmental conditions. Several species possess adaptations to O<sub>2</sub> depletion that are rare amongst eukaryotes, and these species might benefit from ongoing ocean deoxygenation. In addition, since some foraminifera can calcify even under anoxic conditions, they are important archives for paleoceanographic reconstruction in O2-depleted environments. This paper reviews the current state of knowledge about foraminifera from low-O2 environments. Recent advances in our understanding of specific survival strategies of foraminifera to withstand O2 depletion are summarized and discussed. These adaptations include an anaerobic metabolism, heterotrophic denitrification, symbiosis with bacteria, kleptoplasty and dormancy and have a strong impact on their preferred microhabitat in the sediments, especially the ability of some benthic foraminiferal species to denitrify. Benthic foraminifera also differ regarding their trophic strategies, which has an additional impact on the selection of their microhabitat. For example, some species are strict herbivores that feed exclusively on fresh phytodetritus and live close to the sediment surface, while some species are non-selective detrivores that occupy intermediate to deep infaunal habitats. There is evidence that foraminifers have the capacity to undergo phagocytosis, even under anoxia, and some foraminiferal species which can withstand low- $O_2$  conditions seem to prey on meiofauna. Also, due to their high abundances in  $O_2$ -depleted environments and their metabolic adaptations, benthic foraminifera are key players in marine nutrient cycling, especially within the marine N and P cycles. This review summarizes the denitrification rates for the species that are known to denitrify and the intracellular nitrate concentrations of the species that are known to intracellularly store nitrate. Finally, equations are provided that can be used to estimate the intracellular nutrient storage and denitrification rates of foraminifera and might be integrated into biogeochemical models.

# 1 Introduction

More than 2 decades have passed since Bernhard and Sen Gupta (1999) provided a comprehensive review about the history of research on foraminifera from  $O_2$ -depleted environments. About a decade later, Koho and Piña-Ochoa (2012) published another overview about benthic foraminifera as inhabitants of low- $O_2$  habitats, mainly focusing on the species distribution in different environments and the different depth layers in the sediment. They also summarized the early work on foraminiferal denitrification, kleptoplasty and evidence for bacterial symbiosis. Nevertheless, advances in methods to analyze the metabolic rates, intracellular nitrate storage and molecular genetics of foraminifera have changed our understanding of strategies such as an anaerobic metabolism that help them to withstand  $O_2$  depletion. This paper aims to summarize these developments, mainly focusing on benthic foraminifera. For the discussion about life in habitats where  $O_2$  is scarce or absent it is important to define the range of  $O_2$  concentration for terms such as anoxia, hypoxia, and suboxic and oxic conditions. The concentration range for these terms varies with the literature. To avoid confusion, this review only uses the following definitions from the literature:

- Anoxia usually indicates the complete absence of  $O_2$  ([ $O_2$ ] = 0  $\mu$ M; Diaz, 2016).
- Suboxic conditions indicate habitats where O<sub>2</sub> is low enough that denitrification and Mn and Fe reduction are present, but sulfide concentrations are still low due to the absence of sulfate reduction ( $[O_2] \sim 1-10 \,\mu\text{M}$ ; Oakley et al., 2007).
- Hypoxia in aquatic environments indicates habitats where  $O_2$  is present, but the  $O_2$  saturation is less than 30%, since most fish cannot survive below 30% saturation ( $[O_2] < 62.5 \mu M$  Levin et al., 2009).
- Low-O<sub>2</sub> or O<sub>2</sub>-depleted habitats summarize all environments that fulfill one of the above definitions (i.e., every environment where  $[O_2]$  is < 62.5 µM).

Knowledge about planktic foraminifera from O2-depleted habitats is scarce compared to the knowledge about benthic foraminifera. Nonetheless, at least two species (Globorotaloides hexagonus and Hastigerina parapelagica) are known to live in pelagic oxygen minimum zones (OMZs) (Davis et al., 2021). As a result, G. hexagonus has proven to be a valuable paleo-indicator for the presence of pelagic OMZs during the Pliocene (Davis et al., 2023). Benthic foraminifera from low-O2 environments have also been established as an invaluable archive for paleoceanography. However, this review summarizes redox proxies based on benthic foraminifera only briefly, since there is work in progress to give a comprehensive review about proxies for O<sub>2</sub> concentrations in paleoceanography (Hoogakker et al., 2023). Due to their ability to precipitate their calcitic tests even under anoxic conditions, fossil benthic foraminifera became routine tools in paleoceanography to reconstruct past redox conditions (Nardelli et al., 2014; Orsi et al., 2020). Some morphological adaptations are very common for benthic foraminifera that thrive in O<sub>2</sub>-depleted habitats. Small, more elongated and flattened morphologies are often characteristic for O<sub>2</sub> depletion, while more spherical forms can indicate oxygenated conditions (Bernhard, 1986; Bernhard et al., 1997). In addition, high porosity and thin test walls seem to be characteristic for foraminifera that live in low-O<sub>2</sub> environments (Kaiho, 1994). The porosity, including pore size and pore density, of foraminiferal tests recently received more attention as a possible paleoceanographic tool. Different foraminiferal species seem to adapt their pore characteristics in a different way to environmental conditions. Cibicides

spp. for example mainly thrive in well-oxygenated environments (Mackensen et al., 1995), and the porosity in epifaunal Cibicides spp. and Planulina spp. is significantly negatively correlated with the O<sub>2</sub> concentrations in the bottom water (Rathburn et al., 2018; Glock et al., 2022). If O2 is too depleted, these foraminifers increase their porosity to optimize the O<sub>2</sub> uptake. Furthermore, the mechanism of biomineralization in foraminifera can preserve the chemical signature of ambient seawater in their test calcite. These species precipitate their test calcite directly from vacuolized seawater (Erez, 2003; de Nooijer et al., 2014; Toyofuku et al., 2017), and thus the chemical composition of the test calcite reflects the chemical composition of the surrounding water in their habitats. Different element/Ca ratios are used as a proxy for various parameters. Over the past decades several redox-sensitive element/Ca ratios in foraminiferal calcite were identified as potential O2 proxies, where Mn/Ca (Reichart et al., 2003; Barras et al., 2018; Brinkmann et al., 2021) and I/Ca (e.g., Zhou et al., 2014, 2022; Lu et al., 2016; Glock et al., 2019d; Winkelbauer et al., 2021; Cook et al., 2022) are amongst the most prominent examples. The offset of the stable-carbon-isotope fractionation ( $\delta^{13}$ C) between the tests of epifaunal and deep-infaunal benthic foraminifera can also be used as a quantitative [O<sub>2</sub>]<sub>BW</sub> proxy (e.g., Mc-Corkle and Emerson, 1988; Schmiedl and Mackensen, 2006; Hoogakker et al., 2014, 2018). Finally, species compositions of benthic foraminifera assemblages are used to reconstruct past environmental conditions. Kaiho et al. (1994) developed the first benthic foraminifera O2 index (BFOI). Further development of this index is still ongoing, with recent developments by Tetard et al. (2021) and Kranner et al. (2022).

The first part of the present paper reviews recent advances in our understanding of the diverse strategies that foraminifera use to withstand  $O_2$  depletion, focusing mainly on denitrification, dormancy and kleptoplasty. The part about foraminiferal denitrification also incorporates denitrification into the conceptual TROX model of Jorissen et al. (1995). The TROX model explains the sediment microhabitats of benthic foraminifera in terms of an interplay between the supply of  $O_2$  and non-refractory organic matter that can be used as food. The next section briefly summarizes the knowledge about ecological and trophic interactions of foraminifera in marine biogeochemical cycling is discussed, with a focus on nitrogen and phosphorous cycling.

# 1.1 Survival strategies

Some benthic foraminiferal species have very specific adaptations that provide the opportunity either to thrive in anoxia or at least to survive periods of  $O_2$  depletion (see examples in Fig. 1).



Figure 1. Schematic representations for three survival strategy examples performed by benthic foraminifera under  $O_2$ -depleted conditions.

#### 1.1.1 Foraminiferal denitrification

More than a decade ago the first evidence emerged that some foraminifera from  $O_2$ -depleted environments are able to perform complete denitrification (Risgaard-Petersen et al., 2006). Heterotrophic denitrification describes the step-bystep reduction of nitrate ( $NO_3^-$ ) to inert  $N_2$  gas (Reaction R1 according to Jorgensen, 2006, and Fig. 2).

 $5[CH_2O] + 4NO_3^- \rightarrow 2N_2 + 4HCO_3^- + CO_2 + 3H_2O, \ \ (R1)$ 

where [CH<sub>2</sub>O] symbolizes organic matter of unspecified composition.

Heterotrophic denitrification provides energy to an organism for oxidative phosphorylation in a similar way to O<sub>2</sub> respiration. The  $\Delta G^0$  for heterotrophic denitrification per mole of carbon at a pH of 7 is  $-453 \text{ kJ mol}^{-1}$ , which is slightly less efficient O<sub>2</sub> respiration ( $\Delta G^0 = -479 \text{ kJ mol}^{-1}$  according to Jorgensen, 2006).

The discovery of foraminiferal denitrification by Risgaard-Petersen et al. (2006) was also the first evidence for complete denitrification in eukaryotic cells in general, and it showed that they likely take up  $NO_3^-$  from the surrounding pore water and store it within intracellular seawater vacuoles. Nevertheless, no later study could actually prove a bona fide "complete" denitrification pathway in foraminifera, and the eukaryotic foraminiferal denitrification pathway is considered today to be incomplete (Woehle et al., 2018; Orsi et al., 2020; Gomaa et al., 2021; see discussion below). Other eukaryotes that are known to perform incomplete denitrification are the

**Figure 2.** Schematic view of two alternative pathways suggested for foraminiferal denitrification. Abbreviations above the reaction arrows indicate the enzymes that are catalyzing the respective step (see legend). Enzymes in black have been found transcribed by eukaryotic (foraminiferal) RNA (Woehle and Roy et al., 2018). Enzymes in gray are missing in the foraminiferal denitrification pathway and are likely performed by bacterial symbionts (Woehle and Roy et al., 2022). The straight pathway above describes the normal heterotrophic denitrification pathway. The junction, catalyzed by the nitric oxide dismutase, which produces  $O_2$ , has been suggested as an alternative pathway for foraminiferal denitrification (Woehle and Roy et al., 2018).

primitive eukaryote *Loxodes* (Finlay et al., 1983) and two species of fungi (Usuda et al., 1995).

Four years after the study by Risgaard-Petersen et al. (2006), Pina-Ochoa et al. (2010b) documented that intracellular NO<sub>2</sub><sup>-</sup> storage and denitrification are not an exception, limited to a few specialized foraminiferal species, but actually a widespread phenomenon. Within a couple of years more studies either quantified denitrification rates or the intracellular NO<sub>3</sub><sup>-</sup> storage of various foraminifera and gromiid species (Høgslund et al., 2008; Glud et al., 2009; Piña-Ochoa et al., 2010b, a; Koho et al., 2011; Bernhard et al., 2012b). The intracellular  $NO_3^-$  storage can reach concentrations up to 567 mM in gromiids (Piña-Ochoa et al., 2010b), and experiments with isotopically labeled NO<sub>3</sub><sup>-</sup> showed that Globobulimina turgida takes up NO<sub>3</sub><sup>-</sup> at a similar rate, independently of the presence or magnitude of the intracellular  $NO_3^$ pool (Koho et al., 2011). It has been hypothesized that at least some denitrifying foraminifera seem to take up  $NO_3^$ through the pores in their tests, and the pore density (number of pores per area) of some denitrifying species, such as Boliv*ina spissa*, turned out to be a promising proxy for quantitative NO<sub>3</sub><sup>-</sup> reconstructions (Glock et al., 2011, 2018). However, not all benthic foraminifera are able to denitrify, even if they live in environments that are periodically exposed to anoxia such as representatives of the intertidal species morphogroup Ammonia tepida (either Ammonia veneta, Ammonia aberdoveyensis or Ammonia confertitesta according to Hayward et al., 2021), which neither store  $NO_3^-$  nor show any denitrification activity (Piña-Ochoa et al., 2010b). Some foraminifera from the Bering Sea have been shown to store  $NO_3^-$  but did not denitrify in incubation experiments (Langlet et al., 2020). These species include Nonionella pulchella, Uvigerina peregrina and Bolivinellina pseudopunctata. Also, the  $NO_3^-$  storage in U. peregrina shows a high variability, depending on the environment. Individuals of U. peregrina from the Bay of Biscay lack significant  $NO_3^-$  storage, while U. peregrina

from the North Sea and the Bering Sea both show intracellular  $NO_3^-$  enrichment (Piña-Ochoa et al., 2010b; Langlet et al., 2021). Other *Uvigerina* and *Nonionella* species have been shown to denitrify (Risgaard-Petersen et al., 2006; Høgslund et al., 2008; Piña-Ochoa et al., 2010b; Glock et al., 2019b; Gomaa et al., 2021). Many miliolids and allogromiids, several intertidal rotaliid species, and also some other rotaliids and textulariids completely lack intracellular  $NO_3^-$  storage (Piña-Ochoa et al., 2010b).

The observations that some species store NO<sub>3</sub><sup>-</sup> and denitrify in some environments and not in others might have two reasons. One reason could be that these species belong to an opportunistic group of foraminifera that can adapt well to both oxygenated environments where they respire  $O_2$  and do not denitrify and O<sub>2</sub>-depleted environments where they switch to denitrification. The other reason could be that some of these foraminifera belong to morphogroups that are identified as a single species but indeed are a mixture of cryptic and pseudocryptic species that include denitrifying and nondenitrifying species. An example for such a morphogroup that has recently had a revision is A. tepida. This morphogroup includes three species (A. veneta, A. aberdoveyensis and A. confertitesta) that can now be morphologically distinguished (Richirt et al., 2019; Hayward et al., 2021). A similar case concerns the morphogroup Nonionella stella, where representatives have been found to denitrify (Høgslund et al., 2008; Choquel et al., 2021). The morphogroup N. stella also consists of several cryptic to pseudocryptic species (Deldicq et al., 2019). The situation might be similar with other Nonionella species and the widespread species U. peregrina.

There is strong evidence for symbiosis between foraminifera and prokaryotes in many hosts from O2depleted environments, which is most likely an adaptation to survive within the steep geochemical gradients close to the oxic-anoxic boundary (Bernhard et al., 2000, 2006, 2018; Bernhard, 2003; Nomaki et al., 2014). Most of the observed prokaryotic associates are endobionts within the foraminiferal cytoplasm, but some are ectobionts that are often observed close to the pores in the foraminiferal shell (Bernhard et al., 2001, 2010a, 2018). For about a decade after the first discovery of foraminiferal denitrification it remained unclear if foraminifera indeed denitrify themselves or if the bacterial symbionts are responsible for the denitrification. Evidence came up for both hypotheses. Bernhard et al. (2012b) showed that Bolivina argentea consumed its intracellular NO<sub>3</sub><sup>-</sup> storage in O<sub>2</sub>-free incubations even after a very harsh treatment with antibiotics, which indicates that this species can denitrify even when the activity of potential bacterial symbionts would be inhibited. Other studies showed that bacterial endobionts likely perform denitrification in some allogromiid foraminifera and gromiid species (Bernhard et al., 2012a; Høgslund et al., 2017). Gromiida are a separate group of protists within the Rhizaria and are closely related to foraminifera. With the recent advances in molecular biology, however, it became possible to analyze the transcriptome of denitrifying foraminifera, and Woehle and Roy et al. (2018) showed that the enzymes responsible for denitrification in *Globobulimina* spp. from an O<sub>2</sub>-depleted Swedish fjord basin are indeed transcribed by eukaryotic RNA. These enzymes are homologues of enzymes that are also used by bacteria for denitrification, which indicates an ancient prokaryotic origin of denitrification in foraminifera. Nevertheless, the homologues of the enzymes that catalyze the first and the last step of foraminiferal denitrification (reduction of NO<sub>3</sub><sup>-</sup> to nitrite (NO<sub>2</sub><sup>-</sup>) and reduction of nitrous oxide (N<sub>2</sub>O) to N<sub>2</sub> gas; Fig. 2) have not been identified yet. This indicates that foraminifera use other enzymes to catalyze these steps, that they rely on bacterial symbionts for these steps or that they use an alternative denitrification pathway in general.

One hypothesis, brought up by Woehle and Roy et al. (2018), is that the homologue of the nitric oxide reductase (Nor) is indeed a nitric oxide dismutase that has been proposed to catalyze the enzymatic reaction  $2NO \rightarrow N_2 + O_2$ (alternative pathway in Fig. 2) (Ettwig et al., 2012). The presence of the eukaryotic denitrification pathway found in foraminifera (Woehle and Roy et al., 2018) has been confirmed through other analyses of foraminiferal transcriptomes (Orsi et al., 2020; Gomaa et al., 2021). Gomaa et al. (2021) also identified an enzyme of yet unknown functionality that might be responsible for the first step in the foraminiferal denitrification pathway. Recent metagenomics and transcriptomic results of denitrifying foraminifera indicate that bacterial symbionts might perform the missing steps in the foraminiferal denitrification pathway or that they at least partly contribute to the amount of  $NO_3^-$  that is denitrified within foraminiferal cells (Woehle and Roy et al., 2022). It has already been hypothesized before that the ectobionts, found on Bolivina pacifica from the Santa Barbara Basin, are either sulfate-reducing or sulfur-oxidizing bacteria (Bernhard et al., 2010a). The possible complementation of the foraminiferal denitrification with bacterial symbionts appears to be contradictory to the results of Bernhard et al. (2012b), who showed that B. argentea consumed its intracellular  $NO_2^-$  storage (likely for denitrification) even after the antibiotic treatment. Gomaa et al. (2021) confirmed that B. argentea also lacks the first and last denitrification step in its transcriptome, although it lacks intracellular bacterial symbionts (Bernhard et al., 2012b). Future studies might decipher if indeed bacteria are responsible for the missing denitrification step and are immune to such antibiotic treatment; if an oxygenic nitric oxide dismutase skips the last denitrification step, as discussed by Woehle and Roy et al. (2018); and/or if foraminifera have unknown enzymes that catalyze the missing steps, as suggested by Gomaa et al. (2021). The study by Woehle and Roy et al. (2022) also showed that the last common ancestor of denitrifying foraminifera likely has its origin during the Cretaceous, possibly related to the occurrence of the Cretaceous anoxic events. Since the foraminiferal denitrification pathway is incomplete, and the first and last steps might be performed by Desulfobacteraceae in their microbiome, the authors suggested that the acquisition of denitrification ability in foraminifera occurred in multiple stages (starting during the Cretaceous) but is not yet complete (Woehle and Roy et al., 2022).

It is noteworthy that denitrifying foraminifera from the Peruvian OMZ show a metabolic preference for  $NO_3^-$  over O<sub>2</sub> as an electron acceptor (Glock et al., 2019b). These foraminifera show an increasing cell volume with increasing ambient  $NO_3^-$  and decreasing  $O_2$  concentrations. Similar observations have been made at the California Borderland, where some benthic foraminifera also increase their cell volume with decreasing ambient O2 concentrations (Keating-Bitonti and Payne, 2017). Additional evidence for the metabolic  $NO_3^-$  preference came from comparing denitrification and O<sub>2</sub> respiration rates and scaling them to their cell volume (Glock et al., 2019b). The scaling is lower for O<sub>2</sub> respiration than for denitrification, indicating that the  $NO_3^$ metabolism during denitrification is more efficient than the O<sub>2</sub> metabolism during aerobic respiration in foraminifera from the Peruvian OMZ. This might explain why some infaunal denitrifying foraminifera follow the oxycline within sediments (Linke and Lutze, 1993; Duijnstee et al., 2003). We have to keep in mind that  $O_2$  can be quite harmful for organisms that are not adapted to higher O<sub>2</sub> concentrations due to its strong reactivity. Even trace amounts of O<sub>2</sub> can inhibit denitrification, and O<sub>2</sub> can repress the denitrifying enzyme synthesis (Smith and Tiedje, 1979; Knowles, 1981; Tiedje, 1988; Mckenney et al., 1994). Thus, if denitrifying foraminifera are exposed to small amounts of O2 they cannot denitrify but also do not have enough O<sub>2</sub> to supply their demands for electron acceptors. Larger amounts of O2 might supply this demand but also harm the cell. For example, O<sub>2</sub> can inhibit the growth of some obligate anaerobes' poison enzymes that are important for their metabolism (Lu and Imlay, 2021). Also, for aerobes, O2 can be harmful. "Hyperoxia", an excess supply of O<sub>2</sub>, leads to damaging effects by highly reactive metabolic products of O<sub>2</sub> (free O<sub>2</sub> radicals) that inactivate enzymes in the cell, damage DNA and destroy lipid membranes (Frank and Massaro, 1980). Furthermore, for aminifera are able to store  $NO_3^-$  within vacuoles due to its lower reactivity and still have an electron acceptor reservoir if  $NO_3^-$  is depleted in their microhabitat. This is not possible for O<sub>2</sub> due to its high reactivity (Auten and Davis, 2009). Finally, a review by Zimorski et al. (2019) addresses the common misconception that the presence of  $O_2$  improves the overall energetic state of the cell. It is a fact that the energy yield from remineralizing glucose or amino acids is higher in the presence of  $O_2$  (" $O_2$  respiration"), but it is also a fact that the synthesis of biomass consumes 13 times more energy per cell if O<sub>2</sub> is present compared to anoxic conditions. This is related to the chemical equilibrium between organic matter and  $CO_2$ , which strongly shifts to the side of  $CO_2$  in the presence of O<sub>2</sub> (Zimorski et al., 2019). All this might explain why the metabolism of at least some foraminifera is better adapted to denitrification than to  $O_2$  respiration.

The circumstance that some foraminifera have a metabolic preference for  $NO_3^-$  over  $O_2$  as an electron acceptor (Glock et al., 2019b) and that other species like U. peregrina denitrify in some environments but completely lack intracellular  $NO_{3}^{-}$ storage in others (Piña-Ochoa et al., 2010b) might partly explain the microhabitat selectivity of benthic foraminifera in the sediment. According to the conceptual TROX model, benthic foraminifera can be divided into groups due to their microhabitat preference: epifauna, shallow infauna, intermediate infauna and deep infauna (Jorissen et al., 1995). The presence of this species-specific microhabitat structure was first documented by Corliss (1985). These microhabitats are mainly controlled by bottom-water O<sub>2</sub> concentrations and the supply of non-refractory organic matter (i.e., food; Jorissen et al., 1995). Due to our increasing understanding about the anaerobic metabolism of foraminifera we can now assume that  $NO_3^-$  availability is another controlling factor (Fig. 3). This is also indicated by a study that coupled early diagenetic modeling with foraminiferal ecology to model the microhabitats of benthic foraminifera (Jorissen et al., 2022). According to their metabolic preference for  $NO_3^-$  or  $O_2$  as electron acceptors, many benthic foraminifera species that typically occupy a certain microhabitat (epifauna, shallow infauna and deep infauna) might partly be assigned to three different attributes (aerobe, facultative anaerobe and facultative aerobe). There are most likely exceptions to these classifications, which are discussed below. Another controlling factor of the microhabitat is the specific trophic strategy of the foraminiferal species, which is further discussed in Sect. 3.

Deep-infaunal species can most likely be considered to be facultative aerobes that have a metabolic preference for  $NO_3^$ over O<sub>2</sub> (Glock et al., 2019b) and try to avoid trace amounts of O<sub>2</sub>. They cannot be counted as obligate anaerobes though since they can withstand periods of oxygenation. Many experiments show that denitrifying foraminifera can switch to O<sub>2</sub> respiration if they are exposed to O<sub>2</sub> (i.e., Piña-Ochoa et al., 2010b). Still, they follow the oxycline in the sediments to avoid the inhibition of denitrification by trace amounts of O<sub>2</sub>. The  $\delta^{13}$ C signature of shells of deep-infaunal globobuliminids indicates that they calcify in sediment depth where the pore water O<sub>2</sub> level reaches zero or even deeper in the sediments. The offset between  $\delta^{13}$ C of *Globobulimina* spp. tests and  $\delta^{13}$ C of epifaunal foraminifera or of bottom-water dissolved inorganic carbon (DIC) is nearly equal to the offset between DIC at the zero-O<sub>2</sub> layer and the bottom water (Schmiedl and Mackensen, 2006) and often can be even higher (Costa et al., 2023), indicating that many globobuliminids live even below the oxycline. Even though they can switch to O<sub>2</sub> respiration (Piña-Ochoa et al., 2010b), these species would most likely try to avoid crossing the oxycline since denitrification would already be inhibited by O<sub>2</sub> concentrations within the nanomolar range (Dalsgaard et al., 2014), and the  $O_2$  concentration slightly above the oxycline



**Figure 3.** TROX model modified after Jorissen et al. (1995) and Xu et al. (2021). The supply of organic matter and bottom-water  $O_2$  and  $NO_3^-$  concentrations in different environments control the penetration depth of  $O_2$  and  $NO_3^-$  into the sediment. Benthic foraminifera choose their microhabitat according to their metabolic preferences for  $O_2$  or  $NO_3^-$  as an electron acceptor and the availability of food. Intermediate infauna is not specifically schematized in the figure but peaks between the shallow and deep infauna with an overlap in both directions. Note that denitrifying foraminifera can actively transport intracellular  $NO_3^-$  below the  $NO_3^-$  penetration depth in the sediments. The deeper regions where production of free sulfide occurs will mainly be avoided. For further details see the text.

is not high enough to fulfill their metabolic demands. Indeed, the model by Jorissen et al. (2022) describes the distribution of deep infauna very well by using the presence of  $O_2$  as an inhibiting factor, which also suggests that they can be considered facultative aerobes instead of facultative anaerobes. Taxa belonging to the deep-infaunal group that might be considered to be facultative aerobes that prefer  $NO_3^-$  over  $O_2$  include, for example, *Valvulineria inflata* and *bradyana*, *Bolivina seminuda*, *Globobulimina pyrula*, *Globobulimina affinis*, and *Cancris carmenensis* (Jorissen et al., 1995; Schmiedl and Mackensen, 2006; Mojtahid et al., 2010; Glock et al., 2019b).

Shallow infauna can in many cases be considered to be facultative anaerobes that are well adapted to the presence of low  $O_2$  concentrations but can switch to denitrification if they are exposed to anoxic conditions or need to enter the deeper sediment parts to find food or avoid competitive stress. These species have the advantage that they can utilize both fresh phytodetritus from the top of the sediments and organic matter of lower quality from the deeper parts of the sediments. A good example for a shallow-infaunal-facultative anaerobe species is U. peregrina, which is well known for its shallowinfaunal lifestyle (Schmiedl and Mackensen, 2006) and has been found with and without intracellular NO<sub>3</sub><sup>-</sup> storage in different environments (Piña-Ochoa et al., 2010b; Langlet et al., 2020). Of course, it cannot be generalized that all foraminifera from a shallow infaunal habitat are indeed facultative anaerobes. At least some species that can be considered shallow-infaunal have been shown neither to be able to store  $NO_3^-$  nor to denitrify. As mentioned above, all specimens from the Ammonia tepida morphogroup that have been analyzed so far lack intracellular NO<sub>3</sub><sup>-</sup> storage and cannot denitrify (Piña-Ochoa et al., 2010b). Nevertheless, these taxa are often exposed to anoxia and can sometimes even be found alive in 4 to 26 cm sediment depth (Alve and Murray, 2001; Thibault de Chanvalon et al., 2015). It is possible that these foraminifera indeed only have an aerobe metabolism and just become dormant under exposure to anoxia (dormancy is discussed in another section). However, another possibility is that intertidal species such as A. veneta, A. aberdoveyensis and A. confertitesta have other adaptations to anoxia than denitrification. Recent studies revealed other possible anaerobic metabolic pathways in foraminifera such as fermentation and dephosphorylation of creatine phosphate, which are discussed in Sect. 2.1.3 (Orsi et al., 2020; Gomaa et al., 2021). Also, an Ammonia sp. has been shown to take up nitrogen from  ${}^{15}N$ -labeled  $NO_3^-$  under  $O_2$  depletion and to assimilate it within cell organelles known as electron dense bodies (Nomaki et al., 2016). Eventually, studies on the transcriptome of non-denitrifying species from infaunal environments might be able to show if some of these species can switch to an alternative anaerobe metabolism under exposure to anoxia.

Many epifaunal species can most likely be considered to be aerobes that typically occur at the sediment-water interface or on elevated surfaces. Typical epifaunal-aerobe taxa include Cibicides spp. and Planulina spp. (Corliss and Chen, 1988; Lutze and Thiel, 1989). These species have the advantage that they are well adapted to collect fresh food supply from above (Wollenburg et al., 2021) but usually cannot withstand longer O2-depleted periods (Mackensen et al., 1995). Nevertheless, recent genetic data indicate that Cibicidoides wuellerstorfi clusters very close to known denitrifying species in the phylogenetic tree, so it cannot be excluded that some Cibicides spp. may denitrify under certain circumstances (Woehle and Roy et al., 2022). In the same way as for the other microhabitats, not all species with an epifaunal lifestyle should be automatically considered to be aerobes. There are examples of epifaunal benthic foraminifera that have not been found in well-oxygenated environments but reach high abundances in O2-depleted environments. One example is Epistominella smithi, which has been described in low-O2 environments, such as the Santa Barbara Basin (Harman, 1964) or the Peruvian OMZ (Erdem and Schönfeld, 2017). Nevertheless, the morphology of *E. smithi* strongly suggests an epifaunal lifestyle. Another example is the epifaunal species Planulina limbata. This species is abundant only in O<sub>2</sub>-depleted environments on continental margins within the eastern Pacific (Natland, 1938; Erdem and Schönfeld, 2017; Glock et al., 2022). Recent P. limbata specimens are present in severely O2-depleted water masses within the Peruvian OMZ ( $[O_2] = 3-12 \,\mu\text{mol kg}^{-1}$ ; Glock et al., 2022). Nevertheless, P. limbata also adapts its pore density to the availability of  $O_2$  (Glock et al., 2022), which might indicate that it has an aerobic metabolism, despite the fact that its presence appears to be limited to low-O<sub>2</sub> environments. Another possibility is that species such as E. smithi or P. limbata may denitrify under certain circumstances and therefore can also be considered to be facultative anaerobes. Hopefully, measurements of metabolic rates, intracellular nutrient content and enzymatic activity might bring further evidence in the future if at least some epifaunal species can switch to an anaerobe metabolism when  $O_2$  is too depleted.

The intermediate infauna is somehow an exceptional case. Common representatives of intermediate-infaunal taxa are Melonis barleeanus and Pullenia spp. (Corliss, 1991). The typical example for intermediate-infaunal species, M. barleeanus, is interesting since it stores either no or only very small amounts of  $NO_3^-$  (see Tables 2 and 3). Still, several studies indicate that *M. barleeanus* lives deeper in the sediments than some Uvigerina spp. (Corliss, 1991; Ní Fhlaithearta et al., 2018), although many Uvigerina species have been shown to store  $NO_3^-$  and denitrify (Tables 1 and 2). This might give room to speculate if *M. barleeanus* has other metabolic adaptations to O<sub>2</sub> depletion than denitrification or if it simply does not store large amounts of  $NO_3^-$  but denitrifies  $NO_3^-$  directly after the uptake from the seawater. Indeed, a recent study predicted the microhabitats of infaunal benthic foraminifera using an early diagenetic model and showed that the intermediate-infaunal clusters around the NO<sub>3</sub><sup>-</sup> maximum in the pore water (Jorissen et al., 2022). Future perspectives on understanding the biology of intermediate infauna might include transcriptome analyses to decipher other anaerobe metabolic pathways and testing the denitrification capacity after incubation in NO<sub>3</sub><sup>-</sup>-free and NO<sub>3</sub><sup>-</sup>-containing seawater.

Note that the deep infauna can even migrate deeper into the sediments below the depth of  $NO_3^-$  penetration if they must due to their ability to intracellularly store  $NO_3^-$  as a reservoir (Fig. 3). The deeper boundary of the deep infauna might be controlled by the zone of sulfate reduction, where free sulfide is produced, which could be toxic for the foraminifers. Research to measure denitrification rates in different benthic foraminiferal species continues (Langlet et al., 2020; Choquel et al., 2021). This will add to the scarce available data and contribute to estimates of the role of foraminifera in oceanic N cycling. This topic is discussed a bit further in Sect. 4.

# 1.1.2 Kleptoplasty

Kleptoplasty describes a symbiosis between algal chloroplasts and a host organism that sequesters the chloroplasts from algae (Clark et al., 1990). The word originates from the Greek word *Kleptes*, which means "thief". Kleptoplasty in foraminifera is most extensively studied for shallow *Elphidium* and *Haynesina* species that often thrive within the photic zone, and this research originated in the 1970s (Lopez, 1979; Lee et al., 1988; Correia and Lee, 2000, 2002b, a; Goldstein et al., 2004; Pillet et al., 2011, 2013; Cevasco et al., 2015; Jauffrais et al., 2016, 2017, 2018; Cesbron et al., 2017; Goldstein and Richardson, 2018; Jesus et al., 2021). Several studies showed that the sequestered chloroplasts in the intertidal species *Haynesina germanica* are still capable of photosynthesis under light exposure (Lopez, 1979; Ces**Table 1.** Summary of foraminiferal denitrification rates (individual and volume-specific), where Ind. refers to the number of individuals used for one incubation. Individual denitrification rates refer to average rates per individual, while specific denitrification rates refer to rates normalized to the biovolume of the foraminifers. Specimens earlier identified as *A. tepida* are likely either *A. aberdoveyensis* or *A. confertitesta* according to Hayward et al. (2021).

Species	Location	Ind.	Denitrification (pmol N	Specific denitrification (pmol N
			per individual per day)	$\mu M^{-3} d^{-1}$ )
Ammonia tepida <sup>*ª</sup>	Aiguillon Bay	2	0 (n = 1)	0 (n = 1)
Bolivina argentea <sup>b</sup>	Santa Barbara Basin	10	$1976 \pm 1103 \ (n = 8)$	NA
Bolivina costata <sup>c</sup>	OMZ, Peru	13-14	$21 \pm 8 \ (n = 3)$	$3.42\text{E-}5 \pm 1.53\text{E-}5 \ (n=3)$
Bolivina plicata <sup>c</sup>	OMZ, Peru	5-8	$105 \pm 33 \ (n=2)$	$2.49\text{E-5} \pm 3.27\text{E-6} (n = 2)$
Bolivina plicata <sup>a</sup>	OMZ, Peru	3	79 ( $n = 1$ )	1.05E-5 (n = 1)
Bolivina seminuda <sup>c</sup>	OMZ, Peru	6–13	$86 \pm 57 \ (n = 11)$	$5.73E-5 \pm 2.53E-5 \ (n = 10)$
Bolivina seminuda <sup>a</sup>	OMZ, Peru	3	216 (n = 1)	4.15E-5 (n = 1)
Bolivina spathulata <sup>d</sup>	Bering Sea	19	11 (n = 1)	9.17E-7 $(n = 1)$
Bolivina spissa <sup>c</sup>	OMZ, Peru	4–7	$373 \pm 205 \ (n = 5)$	$9.12\text{E-5} \pm 3.66\text{E-5} (n = 5)$
Bolivina subaenariensis <sup>a</sup>	Bay of Biscay	10-12	$78 \pm 2 \ (n = 2)$	$3.12\text{E-}6 \pm 5.43\text{E-}7 \ (n=2)$
Cancris carmenensis <sup>c</sup>	OMZ, Peru	3–4	$765 \pm 306 \ (n = 3)$	$1.86E-5 \pm 4.25E-6 (n = 3)$
Cassidulina limbata <sup>c</sup>	OMZ, Peru	4–6	$45 \pm 16 \ (n = 4)$	$7.62\text{E-}6 \pm 9.25\text{E-}6 \ (n = 3)$
Fursenkoina cornuta <sup>b</sup>	Santa Barbara Basin	10	$1386 \pm 320 \ (n=2)$	NA
Globobulimina auriculata <sup>e</sup>	Gullmar Fjord	4–5	$75 \pm 44 \ (n = 4)$	$2.39\text{E-}6 \pm 1.50\text{E-}6 (n = 4)$
Globobulimina pacifica <sup>d</sup>	Bering Sea	4–5	$378 \pm 471 \ (n=2)$	$1.63E-5 \pm 2.07E-5 (n = 2)$
Globobulimina turgida <sup>a</sup>	Gullmar Fjord	2-3	$358 \pm 134 \ (n = 2)$	7.16E-7 + 5.16E-6 (n = 2)
Globobulimina turgida <sup>f</sup>	Gullmar Fjord	3	$565 \pm 339 \ (n = 10)$	1.13E-6 (n = 1)
Globobulimina turgida <sup>e</sup>	Gullmar Fjord	3–5	$310 \pm 573 \ (n = 8)$	$9.34\text{E-}6 \pm 1.34\text{E-}5 \ (n = 8)$
Nonionella auris <sup>c</sup>	OMZ, Peru	10	$7 \pm 1 \ (n = 1)$	2.70E-6 (n = 1)
Nonionella cf. stella <sup>f, g</sup>	OMZ, Chile	3–5	$84 \pm 33 \ (n = 3)$	$1.62\text{E-5} \pm 6.72\text{E-6} (n = 3)$
Nonionella sp. (T1) <sup>g</sup>	Gullmar Fjord	5	38 (n = 1)	NA
Stainforthia sp. <sup>a</sup>	OMZ, Peru	4	70 (n = 1)	NA
Uvigerina phlegeri <sup>a</sup>	Rhône	10	$46 \pm 2 \ (n = 1)$	5.48E-6 (n = 1)
Uvigerina striata <sup>c</sup>	OMZ, Peru	6-13	$244 \pm 35 \ (n = 3)$	$9.26E-6 \pm 1.50E-6 (n = 3)$
Valvulineria bradyana <sup>a</sup>	Rhône	10	$183 \pm 10 \ (n=2)$	$1.22E-5 \pm 1.32E-6 (n = 2)$
<i>Valvulineria</i> cf. <i>laevigata</i> <sup>a</sup>	OMZ, Peru	10	$248 \pm 180 \ (n=2)$	1.31E-5 + 9.81E-6 (n = 2)
Valvulineria inflata <sup>c</sup>	OMZ, Peru	2–3	$2241 \pm 1825 \ (n=2)$	$3.50\text{E-5} \pm 2.49\text{E-5}$ ( <i>n</i> = 2)

Errors are given as standard deviations (1 SD). <sup>a</sup> Data from Piña-Ochoa et al. (2010b), <sup>b</sup> data from Bernhard et al. (2012b), <sup>c</sup> data from Glock et al. (2019b), <sup>d</sup> data from Langlet et al. (2020), <sup>e</sup> data from Woehle and Roy et al. (2018), <sup>f</sup> data from Risgaard-Petersen et al. (2006), <sup>g</sup> data from Choquel et al. (2021), <sup>h</sup> data from Høgslund et al. (2008). \* *Ammonia tepida* is a morphogroup of pseudocryptic species that recently had a revision.

bron et al., 2017). *H. germanica* often shares a habitat with species from the *Ammonia tepida* morphogroup (*Ammonia aberdoveyensis* or *Ammonia confertitesta* according to Hayward et al., 2021), which also tend to ingest chloroplasts, but these chloroplasts do not show any photosynthetic activity anymore (Jauffrais et al., 2016).

The kleptoplasts in foraminifera originate from diatoms, which has been confirmed on the basis of the chloroplast shape in transmission electron microscopy (TEM) observations and by sequencing the chloroplasts using molecularbiological methods (Lopez, 1979; Lee et al., 1988; Cedhagen, 1991; Lee and Anderson, 1991; Bernhard and Bowser, 1999; Grzymski et al., 2002; Goldstein et al., 2004). Austin et al. (2005) hypothesized that the tooth plates in *H. germanica* are morphological adaptations to crack diatom frustules for access to their chloroplasts. Recently, LeKieffre et al. (2018) showed in (aerated) incubation experiments with  $H^{13}CO_3^-$  and  $^{15}NH_4^+$  during a light-dark cycle that Haynesina germanica is indeed able to fix inorganic carbon and nitrogen under light exposure. Intertidal foraminifera are often exposed to  $O_2$ -depleted or even anoxic conditions when water stagnates during low tide or if they are transported to deeper anoxic sediment layers by bioturbation (Rybarczyk et al., 1996; Cesbron et al., 2017). Oxygen penetration depths in tidal flats can vary from a few millimeters during low tide to several centimeters during high tide (Jansen et al., 2009). Thus, intertidal foraminifera are often exposed to anoxia, even within the first centimeter of the sediment column. H. germanica is also supposed to occur in black sediments of the British salt marsh tide pools (Bernhard and Bowser, 1999), which likely become anoxic during tidal cycles (Rybarczyk et al., 1996), and it was among the first recolonizers of a fjord suffering from organic pollution (Cato et al., 1980; Bernhard and Bowser, 1999). Kleptoplasty might thus be an additional

Species	Location	NO <sub>3</sub> <sup>-</sup> (pmol per cell)	1 SEM	Volume $(\mu m^3 \times 10^{-6})$	1 SEM	[NO <sub>3</sub> <sup>-</sup> ] (mM)	1 SEM
Foraminifera							
Allogromia sp. <sup>i</sup>	Santa Barbara Basin	570	354	NA	NA	70.0	49.0
Ammonia sp. <sup>m</sup>	Sagami Bay	80	4	NA	NA	NA	NA
Bolivina alata <sup>a</sup>	Bay of Biscay	615	154	17.0	1.1	37.0	12.0
Bolivina argentea <sup>b</sup>	Santa Barbara Basin	NA	NA	NA	NA	195.1	160.3
Bolivina cf. abbreviata <sup>a</sup>	OMZ, Peru	1081	368	12.0	2.7	153.0	49.0
Bolivina cf. skagerrakensis <sup>a</sup>	North Sea	83	NA	17.0	0.0	5.0	NA
Bolivina costata <sup>k</sup>	OMZ, Peru	34	4	0.8	0.0	43.1	4.3
Bolivina interjuncta <sup>k</sup>	OMZ, Peru	1239	267	15.6	0.5	80.2	18.9
Bolivina plicata <sup>a</sup>	OMZ, Peru	478	72	7.5	1.0	79.0	15.0
Bolivina robusta <sup>j</sup>	Yellow Sea	212	46	6.1	0.4	35.0	6.0
Bolivina seminuda <sup>k</sup>	OMZ. Peru	140	45	1.6	0.1	88.6	29.8
Bolivina seminuda <sup>a</sup>	OMZ. Peru	564	135	5.2	1.8	118.0	18.0
Bolivina spathulata <sup>d</sup>	Bering Sea	154	NA	10.3	NA	14.9	NA
Bolivina spissa <sup>m</sup>	Sagami Bay	190	72	NA	NA	NA	NA
Bolivina subaenariensis <sup>a</sup>	Bay of Biscay	285	46	25.0	4.3	44.0	9.0
Bolivinellina pseudopunctata <sup>d</sup>	Bering Sea	133	NA	0.9	NA	148.1	NA
Bulimina aculeata <sup>a</sup>	Bay of Biscay	19	12	7.4	0.4	3.0	2.0
Bulimina cf. elongata <sup>a</sup>	OMZ, Peru	817	287	7.9	1.2	116.0	43.0
Bulimina marginata <sup>1</sup>	Yellow Sea	70	11	2.7	0.3	26.0	1.0
Bulimina marginata <sup>a</sup>	Skagerrak	5	NA	1.1	11.0	0.5	0.2
Bulimina marginata <sup>a</sup>	Bay of Biscay	40	4	32.0	1.1	4.0	1.0
Bulimina subula <sup>1</sup>	Yellow Sea	79	8	1.7	0.3	51.0	5.0
Buliminella tenuata <sup>b</sup>	Santa Barbara Basin	NA	NA	NA	NA	217.4	150.5
Cancris auriculus <sup>j</sup>	East China Sea	3211	1046	28.0	5.1	114.0	23.0
Cancris inflatus <sup>a</sup>	OMZ, Peru	263 877	4253	120.0	24.0	262.0	37.0
Cassidulina carinata <sup>a</sup>	Rhône delta	3	1	4.1	0.2	1.0	0.5
Cassidulina cf. laevigata <sup>a</sup>	North Sea	21	NA	4.1	0.0	5.0	5.0
Cassidulina cf. laevigata <sup>a</sup>	OMZ, Peru	523	289	12.0	3.6	41.0	12.0
Cassidulina limbata <sup>k</sup>	OMZ, Peru	1408	710	16.8	2.9	72.9	37.8
Chilostomella oolina <sup>a</sup>	Bay of Biscay	1124	520	20.0	2.0	65.0	36.0
Chilostomella ovoidea <sup>m</sup>	Sagami Bay	50	13	NA	NA	NA	NA
Clavulina cylindrica <sup>a</sup>	Rhône delta	2202	480	35.0	1.0	48.0	13.0
Clavulina cylindrica <sup>a</sup>	Bay of Biscay	1941	314	37.0	5.8	61.0	12.0
Cyclammina cancellata <sup>a</sup>	OMZ, Peru	45 563	45 563	380.0	3.1	119.0	118.0
Fursenkoina cornuta <sup>b</sup>	Santa Barbara Basin	NA	NA	NA	NA	125.2	68.9
Globobulimina affinis <sup>m</sup>	Sagami Bay	480	116	NA	NA	NA	NA

**Table 2.** Summary of intracellular nitrate (NO<sub>3</sub><sup>-</sup>) storage in benthic foraminifera and gromiids from different environments. Only species where intracellular [NO<sub>3</sub><sup>-</sup>] was at least 0.1 mM are listed. Species with intracellular [NO<sub>3</sub><sup>-</sup>] < 0.1 mM are listed in Table 3.

### Table 2. Continued.

Species	Location	NO <sub>3</sub>	1 SEM	Volume	1 SEM	[NO <sub>3</sub> <sup>-</sup> ]	1 SEM
		(pmol per cell)		$(\mu m^3 \times 10^{-6})$		(mM)	
Globobulimina auriculata cf. arctica <sup>a</sup>	Greenland	10 624	3555	100.0	17.0	113.0	43.0
Globobulimina cf. ovula <sup>a</sup>	OMZ, Peru	3369	1602	1.0	2.3	375.0	174.0
Globobulimina pacifica <sup>j</sup>	East China Sea	1167	455	75.0	7.0	16.0	5.0
Globobulimina pacifica <sup>d</sup>	Bering Sea	6530	5563	34.2	8.9	243.9	203.6
Globobulimina turgida <sup>f</sup>	Gullmar Fjord	18 000	4852	500.0	360.0	10.0	2.0
Globobulimina turgida <sup>a</sup>	Skagerrak	8192	1497	100.0	17.0	71.0	13.0
Goesella flintii <sup>a</sup>	OMZ, Peru	459	424	100.0	27.0	24.0	23.0
Gyroidina neosoldanii <sup>a</sup>	OMZ, Peru	13 190	480	27.0	12.0	241.0	46.0
Hanzawaia nipponica <sup>j</sup>	Yellow Sea	316	73	30.0	0.5	11.0	3.0
Hanzawaia nipponica <sup>l</sup>	Yellow Sea	296	49	16.2	4.9	25.0	9.0
Hyalinea balthica <sup>a</sup>	North Sea	8	2	8.0	120.0	1.0	0.3
Labrospira cf. kosterensis <sup>a</sup>	OMZ, Peru	3139	845	51.0	12.0	57.0	12.0
Melonis barleeanus <sup>a</sup>	North Sea	9	3	14.0	20.0	0.6	0.2
<i>Nonionella</i> cf. <i>stella</i> <sup>h</sup>	OMZ, Chile	186	24	5.2	0.7	35.0	5.0
Nonionella pulchella <sup>d</sup>	Bering Sea	31	7	6.7	2.0	7.6	2.2
Nonionella stella <sup>j</sup>	Yellow Sea	162	27	53.0	3.9	3.0	0.6
Nonionella stella <sup>l</sup>	Yellow Sea	178	28	5.5	0.9	34.0	3.0
Nonionella stella <sup>b</sup>	Santa Barbara Basin	NA	NA	NA	NA	11.6	15.7
Protelphidium tuberculatum <sup>1</sup>	Yellow Sea	232	26	3.7	0.5	68.0	9.0
Pyrgo elongata <sup>a</sup>	Rhône delta	43	14	47.0	5.8	0.8	0.2
Pyrgo williamsoni <sup>a</sup>	North Sea	5	NA	47.0	0.0	0.1	NA
Pyrgoella sphaera <sup>a</sup>	North Sea	6	1	47.0	5.8	0.1	0.0
Stainforthia sp. var. I <sup>a</sup>	OMZ, Chile	60	46	0.3	0.0	180.0	29.0
Textularia cf. tenuissima <sup>a</sup>	OMZ, Peru	450	432	11.0	2.9	43.0	7.0
Uvigerina akitaensis <sup>m</sup>	Sagami Bay	210	73	NA	NA	NA	NA
Uvigerina elongatastriata <sup>a</sup>	Bay of Biscay	274	244	5.1	0.6	60.0	55.0
Uvigerina mediterranea <sup>a</sup>	Bay of Biscay	101	66	20.0	6.6	6.0	4.0
Uvigerina peregrina <sup>d</sup>	Bering Sea	74	20	9.9	4.1	10.0	4.7
Uvigerina peregrina <sup>a</sup>	North Sea	332	184	20.0	6.6	16.0	9.0
Uvigerina phlegeri <sup>a</sup>	Rhône delta	444	44	8.4	0.2	209.0	48.0
Valvulineria bradyana <sup>a</sup>	Rhône delta	1268	164	15.0	1.4	95.0	15.0
Valvulineria cf. laevigata <sup>a</sup>	OMZ, Peru	865	640	19.0	3.7	25.0	12.0
Valvulineria inflata <sup>k</sup>	OMZ, Peru	17 666	5319	135.4	16.4	120.1	34.1
Verneuilinulla advena <sup>l</sup>	Yellow Sea	86	15	2.5	0.3	34.0	3.0
Gromiids							
Gromia sp. <sup>a</sup>	Bay of Biscay	2846	1275	93.0	20.0	35.0	21.0
Gromia sp. <sup>a</sup>	Skagerrak	35 277	16546	510.0	110.0	53.0	19.0
Gromia sp. <sup>a</sup>	Rhône delta	3889	1024	160.0	110.0	91.0	26.0
Gromia sp. <sup>a</sup>	North Sea	14 682	4649	160.0	3500.0	140.0	46.0
Gromia sp. <sup>a</sup>	Greenland	12 997	2954	80.0	23.0	163.0	NA
Gromia spp. <sup>d</sup>	Bering Sea	367	85	11.3	6.1	40.2	14.1

Errors are given as standard error of the mean (SEM). <sup>a</sup> Data from Piña-Ochoa et al. (2010b), <sup>b</sup> data from Bernhard et al. (2012b), <sup>d</sup> data from Langlet et al. (2020), <sup>f</sup> data from Risgaard-Petersen et al. (2006), <sup>h</sup> data from Høgslund et al. (2008), <sup>i</sup> data from Bernhard et al. (2012a), <sup>j</sup> data from Xu et al. (2017), <sup>k</sup> data from Glock et al. (2020), <sup>1</sup> data from Xu et al. (2021), <sup>m</sup> data from Nomaki et al. (2015). NA: not available.

#### N. Glock: Benthic foraminifera and gromiids from oxygen-depleted environments

Table 3. Summary of benthic foraminifera from different environments that lack intracellular nitrate (NO <sub>3</sub> <sup>-</sup> ) storage. Only species with
intracellular $[NO_3^-] < 0.1 \text{ mM}$ are listed. Species in bold letters have been found to store $NO_3^-$ in other environments (see Table 2). Specimens
earlier identified as A. tepida are likely either A. aberdoveyensis or A. confertitesta according to Hayward et al. (2021).

Species	Location	Species	Location
Agglutinated sp. <sup>a</sup>	Rhône delta	Hippocrepinella alba <sup>a</sup>	Skagerrak
Ammonia beccarii <sup>a</sup>	Rhône delta	Hyalinea balthica <sup>a</sup>	North Sea
Ammonia beccarii <sup>a</sup>	Bay of Biscay	Komokiacea <sup>a</sup>	OMZ, Peru
Ammonia sp. <sup>a</sup>	Limfjorden	Labrospira cf. L. subglobosa <sup>a</sup>	OMZ, Peru
Ammonia sp. <sup>m</sup>	Sagami Bay	Melonis barleeanus <sup>a</sup>	Rhône delta
Ammonia tepida <sup>*a</sup>	Aiguillon Bay	Nonion scaphum <sup>a</sup>	Rhône delta
Arenoparella asiatica <sup>j</sup>	Yellow Sea	Nonion scaphum <sup>a</sup>	Bay of Biscay
Bathysiphon cf. argenteus <sup>a</sup>	OMZ, Peru	Nouria polymorphinoides <sup>a</sup>	Bay of Biscay
Bathysiphon minutus <sup>a</sup>	Skagerrak	Pelosina variabilis <sup>a</sup>	Skagerrak
Biloculinella depressa <sup>a</sup>	North Sea	Pseudoeponides falsobeccarii <sup>a</sup>	Rhône delta
Bolivinita quadrilatera <sup>a</sup>	Bay of Biscay	Quinqueloculina seminulum <sup>a</sup>	Skagerrak
Bulimina aculeata <sup>a</sup>	Rhône delta	Quinqueloculina seminulum <sup>a</sup>	Bay of Biscay
Bulimina marginata <sup>a</sup>	Rhône delta	Quinqueloculina seminulum <sup>a</sup>	Rhône delta
Cibicidoides pachyderma <sup>a</sup>	Bay of Biscay	Quinqueloculina sp. <sup>a</sup>	OMZ, Peru
Crithionina hispida <sup>a</sup>	OMZ, Peru	Reophax micaceus <sup>a</sup>	Bay of Biscay
Cyclammina cancellata <sup>a</sup>	Bay of Biscay	<i>Reophax</i> sp. <sup>a</sup>	OMZ, Peru
Cypris subglobosa <sup>a</sup>	Bay of Biscay	Rhabdammina inaequalis <sup>a</sup>	North Sea
Dentalina sp. <sup>a</sup>	Rhône delta	Saccammina sp. <sup>a</sup>	Bay of Biscay
Epistominella exigua <sup>a</sup>	OMZ, Peru	Technitella legumen <sup>a</sup>	Skagerrak
Gyroidina altiformis <sup>a</sup>	Bay of Biscay	Triloculina tricarinata <sup>a</sup>	North Sea
Haynesina germanica <sup>a</sup>	Aiguillon Bay	Uvigerina peregrina <sup>a</sup>	Bay of Biscay

<sup>a</sup> Data from Piña-Ochoa et al. (2010b), <sup>j</sup> data from Xu et al. (2017), <sup>m</sup> data from Nomaki et al. (2015). \* *Ammonia tepida* is a morphogroup of pseudocryptic species that recently had a revision.

adaptation of foraminifera from photic environments to stay active during periods of  $O_2$  depletion, which has already been hypothesized by Cesbron et al. (2017).

Less well understood is the phenomenon of kleptoplasty, observed in the benthic foraminifers Nonionella stella, Virgulina fragilis and Nonionellina labradorica, which can thrive below the photic zone and often inhabit O<sub>2</sub>-depleted sediments (Cedhagen, 1991; Bernhard and Bowser, 1999; Grzymski et al., 2002; Bernhard, 2003; Tsuchiya et al., 2015; Jauffrais et al., 2019; Gomaa et al., 2021; Powers et al., 2022). Experiments to test if N. labradorica is able to photosynthesize with its sequestered chloroplasts have been inconclusive. While Cedhagen (1991) found active photosynthesis in *N. labradorica* specimens incubated with <sup>14</sup>C, Jauffrais et al. (2019) showed an increased O<sub>2</sub> respiration rate instead of O2 production and chloroplast degradation in specimens exposed to light. Recently, Gomaa et al. (2021) found chloroplast encoded in transcripts of N. stella, indicating that the kleptoplasts in this species are still active. Genetic analyses revealed that the kleptoplasts in N. stella and N. labradorica are also mainly sequestered from diatoms, most likely after ingestion and selective digestion of phytodetritus (Grzymski et al., 2002; Jauffrais et al., 2019; Gomaa et al., 2021). Grymzki et al. (2002) calculated that the required amount of light for N. stella specimens collected from aphotic depths at the Santa Barbara Basin is too low to sustain active photosynthesis. Instead, they suggested that the kleptoplasts in foraminifera from aphotic environments provide the ability to fix inorganic nitrogen via the glutamine synthetase and glutamate 2-oxoglutarate amidotransferase (GOGAT) pathway. Indeed, Jauffrais et al. (2019) showed that kleptoplastic N. labradorica are able to fix inorganic nitrogen, but coupled TEM-nanoscale secondary ion mass spectrometry (nanoSIMS) revealed that the assimilated nitrogen is associated with electron opaque bodies instead of sequestered chloroplasts. Analyses of the transcriptome of N. stella by Gomaa et al. (2021) support the observations by Grimzki et al. (2002), since N. stella appears to be able to fix ammonia by itself. They also found that the fucoxanthin-chlorophyll binding protein (FCP) was expressed in the transcriptome of N. stella and speculated that the ability to synthesize FCP was derived from the kleptoplasts by horizontal gene transfer. FCP is a pigment, commonly found in chloroplasts of brown algae, and allows a more efficient photosynthesis with a light absorption bandwidth, which is especially useful in aquatic environments (Papagiannakis et al., 2005; Premvardhan et al., 2008). The true function of the kleptoplasts in deep-sea benthic foraminifera from aphotic, often O2-depleted environments still remains enigmatic, though.



**Figure 4.** Examples for molecules and processes that are relevant in the anaerobic metabolism of foraminifera. (a) Structural formula of creatine phosphate. (b) The role of creatine kinase (Ck) and creatine phosphate in the anaerobic metabolism. High-energy creatine phosphate is produced by phosphorylation of creatine. Creatine phosphate can rapidly recycle adenosine diphosphate (ADP) to adenosine triphosphate (ATP) to provide resources for rapid energy bursts. This pathway has been described by Orsi et al. (2020). (c) Fermentation has been found to be relevant in the anaerobic metabolism of foraminifera by both Orsi et al. (2020) and Gomaa et al. (2021). The possibility of a H<sub>2</sub>-producing fermentation pathway, catalyzed by Fe-hydrogenase, has been described by Gomaa et al. (2021).

# 1.1.3 Other strategies: fermentation, utilization of high-energy phosphates and peroxisome proliferation

Several recent publications based on advances in molecularbiological methods (e.g., next generation sequencing) have revealed some other metabolic adaptations of foraminifera that thrive under  $O_2$  depletion (see examples in Fig. 4) (Woehle and Roy et al., 2018, 2022; Orsi et al., 2020; Gomaa et al., 2021). In N. stella and Bolivina argentea, Gomaa et al. (2021) found evidence for the expression of proteins, including pyruvate:ferredoxin oxidoreductase (PFOR) and [FeFe]-hydrogenase, which are characteristic of anaerobic metabolism. These PFOR sequences were indeed eukaryotic and closely related to those of the facultative anaerobe polychaete *Capitella teleta* and the anaerobic protistan parasite Blastocystis. The [FeFe]-hydrogenase is very similar to those in the amoeba/flagellate Naegleria gruberi, which has experimentally been shown to be active and to produce molecular hydrogen even under aerobic conditions (Tsaousis et al., 2014). Due to these observations, Gomaa et al. (2021) suggested that N. stella and B. argentea might be able to produce H<sub>2</sub> gas and have the capacity for an anaerobic energy metabolism.

Another important observation was made by Orsi et al. (2020). They used metatranscriptomics on sediments from the Namibian shelf, where the foraminiferal community is dominated by *Bolivina* and *Stainforthia* species. Presumably living foraminifera were present in the sediment column up to 28 cm depth in an anoxic habitat with high sulfide concentrations. The gene expression of those foraminifers increased under sulfidic conditions, which indicates that they not only survive but thrive under anoxic conditions. The anaerobic energy metabolism of these foraminifers seems to be sufficient enough to support calcification and phagocytosis even under anoxic conditions. Evidence for foraminiferal calcification under anoxia already came up in a study by Nardelli et al. (2014). Orsi et al. (2020) suggested that the Namibian foraminifera use phagocytosis (vacuolic ingestion of food particles) to ingest prey cells even under anoxic conditions. These processes (calcification and the ingestion of prey cells by phagocytosis) require bursts of high energy, which the authors suggest is generated by dephosphorylation of intracellular creatine phosphate storage to regenerate ATP

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from ADP. Evidence for the capacity for the dephosphorylation of creatine phosphate under anoxia was indicated by the metatranscriptomes. In addition, high intracellular dissolved inorganic phosphate storage has been found in benthic foraminifera from the Peruvian OMZ, which might serve as a reservoir to synthesize creatine phosphate and/or to synthesize polyphosphates that might be broken down to harvest energy (Glock et al., 2020). Orsi et al. (2020) and Gomaa et al. (2021) also found evidence for another anaerobic metabolism. Their data indicate that the foraminifers metabolize hydrolyzed organics to produce ATP using fermentation and fumarate reduction.

Most foraminifera species from O<sub>2</sub>-depleted habitats possess numerous peroxisomes that are usually associated with mitochondria and the endoplasmatic reticulum (Bernhard and Bowser, 2008). Bernhard and Bowser (2008) hypothesized that these peroxisome proliferations might be used to either metabolize H<sub>2</sub>O<sub>2</sub> and other highly reactive oxygen species that are produced within the chemocline close to the oxic-anoxic boundary or to reduce the oxidative stress by these compounds. Indeed, they showed in an experiment that ATP concentrations in foraminifera increased proportionally to ambient H2O2 concentrations. A recent study on the transcriptome and metatranscriptome of N. stella and B. argentea from the Santa Barbara Basin revealed that these species utilize an adaptable mitochondrial and peroxisomal metabolism, depending on the chemical treatment in the experiment (Powers et al., 2022). The high plasticity of their peroxisomal and mitochondrial metabolism might be substantial for survival under the highly variable conditions at the chemocline in the sediments. The results of Powers et al. (2022) indicate that at least some processes that are involved in foraminiferal denitrification are associated with mitochondria. Interestingly, the expression of denitrificationrelated genes in both species was upregulated after incubation with elevated  $H_2O_2$  but without  $NO_3^-$  and downregulated if they were incubated without  $H_2O_2$  but with  $NO_3^-$ , compared to a control treatment with both  $H_2O_2$  and  $NO_3^-$ . In the same way several peroxisomal processes were upregulated in the  $H_2O_2$ -only treatment. In addition, despite the fact that both species are able to denitrify, Powers et al. (2022) found distinct metabolic adaptations to anoxia in both species. For example, a quinol:fumarate oxidoreductase, which is considered to be an adaptive mechanism for anaerobic respiration in eukaryotic organisms, was present in N. stella but not in B. argentea. On the other hand, B. argen*tea* has the capacity to digest food vacuole contents under O<sub>2</sub> depletion, while N. stella was lacking food vacuoles (Powers et al., 2022).

## 1.1.4 Dormancy

Dormancy is another strategy to survive anoxia or extreme  $O_2$  depletion for some benthic foraminifera that cannot denitrify. Dormancy is defined as the reduced or suspended metabolic activity in response to exogenous factors (Ross and Hallock, 2016). Observations that indicate the potential of dormancy in foraminifera have been documented since the 1950s and are extensively reviewed by Ross and Hallock (2016). Nevertheless, many aspects of foraminiferal dormancy, such as its role in the foraminiferal life cycle or its role in structuring foraminiferal assemblages, remain unexplored (Ross and Hallock, 2016).

In the 1990s some studies suggested that some foraminifera may become dormant when exposed to anoxia. Bernhard and Alve (1996) observed that the ATP concentration ([ATP]) of the benthic foraminiferal species Bulimina marginata, Stainforthia fusiformis and Adercotryma glomerata flushed with N<sub>2</sub> gas to drive out O<sub>2</sub> was significantly lower than in specimens from well-aerated conditions. They interpreted this observation as an indication that dormancy is a survival strategy for some foraminiferal species when they are exposed to periods of anoxia. Linke and Lutze (1993) observed cysts of *Elphidium incertum* from putative anoxic habitats that might be interpreted as a sign for dormancy, and Hannah and Rogerson (1997) hypothesized that foraminifera transported to an anoxic sediment layer might become dormant until they return to aerated conditions by transport through bioturbation.

Recently, dormancy of foraminifera exposed to anoxia has gained more attention again. LeKieffre et al. (2017) did a feeding experiment with specimens from the Ammonia tepida morphogroup (A. confertitesta according to Koho et al., 2018, and Hayward et al., 2021) using a <sup>13</sup>C-labeled diatom film as a food source. They compared the metabolic differences in Ammonia sp. between oxic and anoxic conditions by mapping the distribution of <sup>13</sup>C within the cells using coupled TEM-nanoSIMS and by analyzing the carbon concentration and stable-carbon-isotope composition of the total organic matter and individual fatty acids in the foraminifer. Nearly the complete diatom biofilm was consumed, and the foraminiferal cytoplasm was strongly enriched in <sup>13</sup>C under oxic conditions. Specimens from the anoxic incubation ingested only few of the diatoms, and those were neither assimilated nor metabolized further. In addition, the specimens from the oxic incubation produced a significant quantity of specific polyunsaturated fatty acids, which was not the case under anoxic conditions. A. confertitesta reacted to the induced anoxia with a severely reduced metabolic rate within less than 24 h. All these observations provide solid evidence that dormancy is a survival strategy of A. confertitesta under anoxia.

Koho et al. (2018) further analyzed cell structural changes in *Ammonia* spp. under exposure to anoxia collected from the field as well as from incubations. The specimens from anoxia showed an increase in lipid droplets and electron dense bodies within their cytoplasm. The cytoplasm itself was thinned out, which was interpreted as metabolization of the cytosol. In addition, while absent within the specimens from oxic environments, various bacteria were present within the cytoplasm of the specimens from anoxia. These were interpreted as endobionts but might also be parasites that could not be fended off due to the drastically reduced metabolism during dormancy under anoxia. A continuum of intracellular bacteria including prey in food vacuoles, endobionts, parasites and necrophages has been documented before in benthic foraminifera from cold seeps (Bernhard et al., 2010b). It has already been hypothesized by the authors that bacteria can switch their function from endobionts to predators, depending on the vitality of the host cell. Considering all the studies about dormancy, it is likely that dormancy is a common survival strategy for foraminiferal species whose supply of suitable electron acceptors is exhausted (i.e.,  $O_2$  or  $NO_3^-$ ) or who are exposed to periods of extreme environmental conditions. Since there is evidence for dormancy in both S. fusiformis and B. marginata (Bernhard and Alve, 1996), it is likely that even denitrifying species can become dormant under unfavorable conditions. Another Stainforthia sp. has been shown to denitrify, and *B. marginata* stores  $NO_3^-$  in some environments (Piña-Ochoa et al., 2010b).

# 2 Trophic interactions in O<sub>2</sub>-depleted environments

In general, benthic foraminifera show a wide range of trophic strategies. Gooday et al. (2008) suggested that they can be separated according to their main trophic types (see examples in Fig. 5): A – selective herbivores, which include phytophagous species that consume only phytodetritus; B - seasonal herbivores, which feed on fresh phytodetritus when available and consume sedimentary organic matter at other times; C - detrivores that non-selectively ingest sediment and consume the present degraded organic matter, bacteria and/or other organisms; D - selective bacterivores that consume only bacteria; and E - suspension feeders that either arise from the sediments or occur on elevated substrates. The latter two are not discussed in detail, since they mainly apply to abyssal species that inhabit more oxygenated environments. Nevertheless, some Cibicides and Planulina species can also inhabit environments with relatively low O2 concentrations (Erdem and Schönfeld, 2017; Rathburn et al., 2018; Hoogakker et al., 2018b; Glock et al., 2022), and at least some of these Cibicides species are certainly suspension feeders (Wollenburg et al., 2018, 2021). The trophic types that have been introduced above suggest that foraminifera mainly feed on a low trophic level, and it has been suggested that they constitute a trophic link to higher levels in the food chain (Lipps and Valentine, 1970; Gooday et al., 1992; Nomaki et al., 2008).

There are a few studies that specifically focused on trophic interactions of foraminifera in environments where  $O_2$  is scarce or absent. Early observations have been documented by Nomaki et al. (2006), who conducted an in situ feeding experiment at central Sagami Bay (1450 m), Japan, using <sup>13</sup>C-labeled algae and bacteria. Bottom-water  $O_2$  con-

centration at this location is usually less than 60 µM, and O<sub>2</sub> penetration depth into sediments varies between 3 and 10 mm, indicating that infaunal foraminifera in this habitat are regularly exposed to hypoxia and anoxia (Glud et al., 2005). Nomaki et al. (2006) described three different feeding strategies by benthic foraminifera in this environment. Since the bottom-water O<sub>2</sub> concentrations at central Sagami Bay fluctuate and are not strictly hypoxic, these observations likely apply to more oxygenated environments as well, especially for the shallow-infaunal species. Uvigerina akitaensis, Bolivina spissa and Bolivina pacifica selectively ingest fresh phytodetritus and thus can be described as phytophagous species (selective herbivores). Bulimina aculeata, Textularia kattegatensis and Globobulimina affinis ingest fresh phytodetritus selectively but feed on sedimentary organic matter instead when fresh phytodetritus is unavailable (seasonal herbivores). The species Cyclammina cancellata and Chilostomella ovoidea ingest sedimentary organic matter at random and can thus be described as detrivores. A later study confirmed these trophic types for most of the species at Sagami Bay by measuring the nitrogen isotope fractionation ( $\delta^{15}$ N) of their amino acids, which is commonly used to trace the trophic position of an organism in the food chain (Nomaki et al., 2015). Another feeding experiment at Sagami Bay by Nomaki et al. (2011) revealed that all of the analyzed benthic species assimilated carbon from <sup>13</sup>C-labeled glucose and thus can effectively also utilize dissolved organic carbon. The same study indicated that even the deep-infaunal detrivores can be selective regarding their food source. Four of the five analyzed species, excluding C. cancellata, incorporated proportionally more <sup>13</sup>Clabeled organic matter from the green algae Dunaliella sp. than from other carbon sources, while C. cancellata preferentially incorporated carbon from Chlorella sp. (Nomaki et al., 2005, 2006, 2011). Additional feeding experiments have been conducted at the Arabian Sea OMZ, where benthic foraminifera from locations with different bottom-water O<sub>2</sub> concentrations have been supplied with <sup>13</sup>C- and <sup>15</sup>Nlabeled algae (Enge et al., 2014, 2016). Nine out of nine analyzed species took up labeled phytodetritus during the 4 d experimental phase (Enge et al., 2014). The foraminifera took up the highest amount of labeled carbon in the OMZ center, and the uptake decreased with distance from the OMZ (Enge et al., 2016). The authors hypothesized that either the foraminifera from the core OMZ have a higher carbon demand or that there was less food competition with macrofauna at the O<sub>2</sub>-depleted locations. Similar to the studies by Nomaki et al. (2005, 2006, 2011) at Sagami Bay, the experiments by Enge et al. (2014, 2016) showed a more or less selective ingestion at the Arabian Sea OMZ depending on the foraminiferal species. For example, several Uvigerina species took up large amounts of carbon from the labeled algae and are thus either selective or seasonal herbivores, while Globobulimina spp. took up either no or only small amounts of the labeled carbon, indicating their detritivore



**Figure 5. (a)** Schematic representation of a *bolivinid* ingesting bacterial cells. Recent studies have shown that benthic foraminifera from  $O_2$ -depleted habitats have the capacity to undergo phagocytosis even under anoxia (Orsi et al., 2020). **(b)** Schematic representation of *Ammonia* sp. preying on a nematode. Some benthic foraminifera are known to prey on meiofauna (Dupuy et al., 2010), and there is evidence that even some *globobuliminids* that usually thrive under  $O_2$ -depleted conditions might prey on nematodes (Glock et al., 2019a).

behavior (Enge et al., 2016). Further examples for selective herbivores, opportunistic omnivores (which include seasonal herbivores) and sediment detrivores are discussed by Gooday et al. (2008). It appears that many of the species that are considered to be selective herbivores (e.g., B. spissa, U. akitaensis, Eponides pusillus and Cassidulina carinata) have epifaunal or shallow-infaunal lifestyles, although the selective herbivore B. pacifica can also be considered to be intermediate infauna (Gooday et al., 2008). The seasonal herbivores (or opportunistic omnivores, e.g., U. peregrina, G. affinis and G. pacifica) can be found in a relatively wide range of microhabitats, from shallow to deep infauna (Gooday et al., 2008). Species that are considered to be sediment deposit feeders (or detrivores, e.g., C. ovoidea and M. barleeanum) are usually found in the deeper habitats and belong to intermediate to deep infauna (Gooday et al., 2008). This indicates that the selective herbivores must live closer to the source of fresh food supply, while the less selective species can also feed on degraded organic matter or bacteria deeper in the sediments. Thus, the specific trophic type is another control on the microhabitat of benthic foraminifera in addition to the availability of  $O_2$  and  $NO_3^-$  and the metabolic adaptations discussed in Sect. 2. Indeed, the coupled diagenetic and ecologic model of Jorissen et al. (2022) successfully uses different types of food particles as a controlling factor to simulate the microhabitats of benthic foraminifera.

Although benthic foraminifera feed mainly on detritus and minute organisms, there is also (less common) evidence for carnivorous behavior when foraminifera prey on meiofauna (e.g., Lee, 1980; Bowser et al., 1986, 1992; Hallock and Talge, 1994). These observations have mainly been made for species that usually live in oxygenated environments. Dupuy et al. (2010) also documented carnivorous behavior in a laboratory experiment for the Ammonia tepida morphogroup (A. aberdoveyensis or A. confertitesta according to Hayward et al., 2021), which is not uncommon in anoxic layers of tidal mudflats. A study on the trophic behavior of intertidal foraminifera using metabarcoding brought up evidence that A. confertitesta actively preys on small eukaryotes (e.g., nematodes), even in their natural environment (Panagiota-Chronopoulou et al., 2019). The intracellular eukaryotic community in A. confertitesta varies with sediment depth, but even up to 10 cm depth the metabarcoding indicates freshly ingested eukaryotic prey in this species (Panagiota-Chronopoulou et al., 2019). Still, the main eukaryotic prey of A. confertitesta appears to be diatoms (Panagiota-Chronopoulou et al., 2019). Similar results have been documented by Schweizer et al. (2022). Recently, new evidence came up indicating ingestion of nematodes by Globobulimina auriculata from the O<sub>2</sub>-depleted Alsbäck Deep in Gullmar Fjord, Sweden (Glock et al., 2019a). The species G. auriculata denitrifies and lives under O<sub>2</sub>-depleted conditions (Woehle and Roy et al., 2018). It is inconclusive though if the foraminifer preys on the nematode or vice versa, but the nematodes have most likely been ingested in the natural O<sub>2</sub>-depleted habitat (Glock et al., 2019a). Although predation is the main type of interaction in aerobic communities, it usually plays a much smaller role in anoxic communities (Fenchel and Finlay, 1995). This is related to the low growth yields associated with the anaerobic metabolism, which results in very short food chains. Thus, the decrease in energy flow along the anaerobic food chains is higher than along the aerobic food chain (Fenchel and Finlay, 1995). The predatory isopod Saduria entomon for example strongly reduces its predatory activity under hypoxia in comparison to aerobic conditions (Sandberg, 1994), and the predator / prey biomass ratio has been shown to be 4 times lower in anoxic environments compared to oxic environments (Fenchel and Finlay, 1995). There is evidence that foraminifera from the Namibian shelf can perform phagocytosis (vacuolic ingestion of food particles) even under anoxic conditions, which usually requires bursts of energy (Orsi et al., 2020). This study provides further evidence that the Namibian foraminifera express enzymes for lysing digested prey cells inside food vacuoles after phagocytosis (schematic representations for phagocytosis and predation on meiofauna shown in Fig. 5). The evidence for phagotrophy and predation on or by benthic foraminifera under O2-depleted conditions, although it is rare, is thought-provoking, and future studies might shed more light on predator-prey interactions of benthic foraminifera in O2-depleted environments. In general, future metabarcoding studies to identify food sources of deep infauna or foraminifera that inhabit anoxic environments might shed more light on trophic strategies in O<sub>2</sub>depleted environments.

# **3** The role of foraminifera in benthic nutrient cycling and biogeochemistry

Pina-Ochoa et al. (2010b) also suggested the possible importance of denitrifying foraminifera for the benthic N cycle due partly to their high abundances in O<sub>2</sub>-depleted environments. In some environments, such as certain habitats in the Peruvian OMZ, foraminifera even seem to be the key players in benthic denitrification (Glud et al., 2009; Glock et al., 2013, 2019b; Choquel et al., 2021). Complete heterotrophic denitrification produces non-reactive (i.e., not bioavailable) N<sub>2</sub> gas. Denitrifying benthic foraminifera can thus be considered a sink for bioavailable N. Recent genetic studies on denitrifying benthic foraminifera did not find transcripts for homologues of enzymes that catalyze the last step of denitrification – the reduction of N<sub>2</sub>O to N<sub>2</sub> (Woehle and Roy et al., 2018, 2022; Orsi et al., 2020; Gomaa et al., 2021). Some

globobuliminids from the O<sub>2</sub>-depleted Alsbäck Deep in the Swedish Gullmar Fjord have been shown to produce N<sub>2</sub>O gas as a product of denitrification, although the rates were lower than their rates for complete denitrification (Piña-Ochoa et al., 2010a). The NO<sub>3</sub><sup>-</sup> storage in denitrifying foraminifera, but also in some sulfur bacteria, such as *Beggiatoa*, is of greater importance for benthic biogeochemical cycling due to the potential of biological transport of these intracellular reservoirs (Dale et al., 2016). Most of the other diagenetic models that describe and calculate benthic N cycling are based on (and limited to) diffusive transport of the different N species in bottom and pore water. Active biological transport of different N species can thus efficiently influence the benthic fluxes of different N species (Dale et al., 2016).

The estimates of total benthic foraminiferal denitrification rates are mainly based on upscaling individual speciesspecific denitrification rates by the living abundances of benthic foraminifera in different environments (Piña-Ochoa et al., 2010b; Glock et al., 2013, 2019b). This approach is limited by the availability of species-specific denitrification rates, although various approximations can be used to calculate estimated denitrification rates for species with unknown denitrification rates (Glock et al., 2013). A summary of all published benthic foraminiferal denitrification rates can be found in Table 1. Further data on species-specific foraminiferal denitrification rates will improve our estimates regarding the role of foraminifera in benthic N cycling and thus also models for benthic biogeochemical cycling.

Recently, it has been found that some benthic foraminifera not only store  $NO_3^-$  for denitrification but also larger amounts phosphate (Glock et al., 2020). The intracellular phosphate concentration can exceed the concentration in the surrounding pore waters by a factor of 10 to 100. The use of this intracellular phosphate storage is still under debate. Hypotheses include the synthesis of polyphosphates or a reservoir for the synthesis of phospholipids for the cell membranes (Glock et al., 2020). In addition, there is evidence that the intracellular phosphate storage in foraminifera facilitates phosphogenesis in some environments, similar to the intracellular polyphosphate enrichment in some sulfur bacteria (Schulz and Schulz, 2005). The release of phosphate after breakdown of these polyphosphates to harvest energy in times of electron acceptor depletion results in apatite supersaturation and initiates phosphogenesis (Schulz and Schulz, 2005). Sediments at the lower boundary of the Peruvian OMZ contain many small phosphorite grains with foraminifera of similar size and shape (Manheim et al., 1975; Glock et al., 2020). The sand fraction of the surface sediments in this region is a mixture of pristine living foraminifer shells with dead tests that show a transition from shells that are filled with phosphorites to small phosphorite grains that only retain the size and coarse shape of a foraminifer. It is likely that a postmortem release of the intracellular phosphate storage results in a supersaturated microenvironment within the shells that initiates apatite formation (Glock et al., 2020) in a similar way



**Figure 6.** Log–log plot and power regression of intracellular NO<sub>3</sub><sup>-</sup> content (NO<sub>3</sub><sup>-</sup>) against the biovolume (V<sub>cell</sub>) of benthic foraminiferation from diverse environments (Table 2). Only species with an intracellular [NO<sub>3</sub><sup>-</sup>]  $\geq$  1 mM where both NO<sub>3</sub><sup>-</sup> and V<sub>cell</sub> were published were considered for the power regression.

to that suggested for other organisms (Kulakovskaya, 2014). The recent evidence for the potential of benthic foraminifera to use dephosphorylation of intracellular creatine phosphate storage to regenerate ATP under anoxic conditions might be another explanation for the high intracellular phosphate storage (Orsi et al., 2020). It might be that this is an adaptation of foraminifera to enable phagocytosis even under anoxic conditions.

# **3.1** Estimating the contribution of foraminifera to benthic nutrient budgets and fluxes

The intracellular  $NO_3^-$  storage in benthic foraminifera from different environments shows a relatively wide concentration range (Table 2). In addition, species that lack intracellular  $NO_3^-$  storage are relatively widespread, and there are species that, depending on the environment, either have or lack intracellular  $NO_3^-$  (Tables 2 and 3). Most of the species that have been found both with and without intracellular  $NO_3^-$  in different environments (bold species in Table 3) are species that are typically shallow-infaunal. They belong to the group of foraminifera that might partly be considered facultative anaerobe and are likely opportunistic species that are well adapted to transitional environments with periodic  $O_2$  depletion, since they apparently can handle oxygenated and anoxic environments (see Sect. 2.1.1). In addition, the  $NO_3^-$  is most likely stored in seawater vacuoles, and the vacuole volume of foraminifera can have a large variability (LeKieffre et al., 2018).

Given this variation in  $NO_3^-$  storage capability, the reliability of estimates for the foraminiferal contribution to  $NO_3^-$  budgets depends crucially on the availability of data. The more data there are, the better we are able to calculate foraminiferal NO<sub>3</sub><sup>-</sup> budgets. Nevertheless, there are thousands of benthic foraminiferal species, and a considerable number of these species inhabit O<sub>2</sub>-depleted environments and potentially store  $NO_3^-$  and denitrify. It will be unrealistic to measure the intracellular nutrient content and metabolic rates for all foraminifera. Thus, functions to estimate the contribution of species with unknown denitrification rates or intracellular  $NO_3^-$  will provide more data for better estimates of total foraminiferal budgets within the nitrogen cycle. Of course, it is not possible to strictly define which for a species are able to denitrify or to store  $NO_3^$ without real measurements. If a foraminiferal species inhabits O<sub>2</sub>-depleted environments and belongs to a genus of the species listed in Tables 1 and 2, as a rule of thumb, they are good candidates for potential denitrifiers. In addition, if a species is known to inhabit well-oxygenated environments and/or belongs to a genus of the species shown in Table 3, use of the equations presented below to estimate NO<sub>3</sub><sup>-</sup> storage or denitrification rates should be avoided. Considering this, an analysis of published data on intracellular NO<sub>3</sub><sup>-</sup> content reveals a highly significant correlation between the intracellular  $NO_3^-$  and the cell volume of denitrifying benthic

for aminifera (Fig. 6; power regression,  $R^2 = 0.59$ , F = 86,  $P = 3 \times 10^{-13}$ ).

Thus, the intracellular  $NO_3^-$  content of a potentially denitrifying foraminifer can be estimated from its biovolume according to the following equation:

$$\ln(\text{NO}_{3^-}) = 1.07(\pm 0.11) \times \ln(\text{V}_{\text{cell}}) - 11.5(\pm 1.9), \quad (1)$$

where  $NO_{3_i^-}$  is the intracellular  $NO_3^-$  content in picomoles per individual and  $V_{cell}$  is the cell volume in cubic micrometers. Note that only species from Table 2 with an intracellular  $[NO_3^-] \ge 1$  mM were considered for the power regression. In addition, two extreme data points were discarded as outliers (see Supplement). Similar equations have been published to estimate foraminiferal denitrification rates (Glock et al., 2019b; Eq. 2 herein) and intracellular dissolved inorganic phosphorous content (Glock et al., 2020; Eq. 3 herein).

$$\ln(R_{\rm den(ind)}) = 0.68(\pm 0.12) \times \ln(V_{\rm cell}) - 5.57(\pm 1.9)$$
(2)

$$\ln(\text{DIP}_i) = 0.82(\pm 0.03) \times \ln(\text{V}_{\text{cell}}) - 7.65(\pm 0.52), \quad (3)$$

where  $R_{\text{den(ind)}}$  is the individual denitrification rate in picomoles per individual per day, and  $\text{DIP}_i$  is the intracellular dissolved inorganic phosphorous content in picomoles per individual.

Further equations and principles for upscaling foraminiferal nitrogen and phosphorous budgets from abundances of living foraminifera can be found in Glock et al. (2013, 2019b and 2020) and Xu et al. (2021). Formulae to estimate the biovolume of many different common shapes of foraminifera have recently been published (de Freitas et al., 2021). Due to the high uncertainties related to the natural variability in metabolic rates and nutrient storage, a thorough error estimation is recommended (see Appendix B in Glock et al., 2020). With an increasing amount of data on metabolic rates and intracellular nutrient storage, more accurate models and equations might become available in the future that describe the role of benthic foramifera within marine biogeochemistry. Similar models and equations might also be very helpful for exploring the role of planktonic foraminifera in pelagic biogeochemistry.

*Data availability.* No new data have been used for this paper. All original data can be found in the cited articles, and all data I compiled are available in the tables of the paper.

*Supplement.* The supplement related to this article is available online at: https://doi.org/10.5194/bg-20-3423-2023-supplement.

*Competing interests.* The contact author has declared that neither of the authors has any competing interests.

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*Special issue statement.* This article is part of the special issue "Low-oxygen environments and deoxygenation in open and coastal marine waters". It is not associated with a conference.

Acknowledgements. I would like to thank Gerhard Schmiedl for providing constructive feedback on an early draft of this paper. In addition, I acknowledge the extensive and constructive feedback of Andrew Gooday, Frans Jorissen, another anonymous reviewer and the editor Lisa Levin, which significantly improved this paper. Funding was provided by the Deutsche Forschungsgemeinschaft (DFG) through Heisenberg Grant GL 999/3-1 to Nicolaas Glock. Finally, I would like to thank all the authors and co-authors that are cited in this review because of their pioneering research on benthic foraminifera from O<sub>2</sub>-depleted environments.

*Financial support.* This research has been supported by the Deutsche Forschungsgemeinschaft (grant no. GL 999/3-1).

*Review statement.* This paper was edited by Lisa Levin and reviewed by Andy Gooday and one anonymous referee.

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