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Microbial strong organic-ligand production is tightly coupled to iron in hydrothermal plumes

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Abstract. Hydrothermal vents have emerged as important sources of iron to seawater, yet only a subset of this iron is soluble and persists long enough to impact the deep-ocean iron inventory. The longevity and solubility of iron in seawater is in part governed by strong organic ligands that are produced by microorganisms and are a part of the ocean's dissolved organic iron-binding ligand pool. Organic ligands have long been recognized to support elevated dissolved iron in hydrothermal vent plumes. Siderophores are one group of microbially produced organic ligands that have especially high binding affinities for iron. Here we present the first direct measurements of siderophore concentrations in hydrothermal vents, which we compare to bulk strong ironbinding ligand concentrations, along a 1700 km section of the Mid-Atlantic Ridge. Siderophores were found in hydrothermal plumes at all sites, with proximity to the vent playing an important role in dictating siderophore type and diversity. The notable presence of amphiphilic siderophores may point

to microbial utilization of siderophores to access particulate hydrothermal iron and the exchange of dissolved and particulate iron. The tight coupling between strong ligands and dissolved iron within neutrally buoyant plumes across distinct hydrothermal environments, as well as the presence of dissolved siderophores with siderophore-producing microbial genera, suggests that biological production of ligands influences iron chemistry in hydrothermal systems.

1 Introduction

Over the last few decades, observations and modeling efforts have increased our understanding about the critical role organic ligands play in the cycling, transport, and utilization of trace metals (Tagliabue et al., 2017; Buck et al., 2018; Bundy et al., 2018; Moore et al., 2021; Hawkes et al., 2013b; Kleint et al., 2016). Iron (Fe)-binding organic ligands in seawater have a wide range of sources, which are only just beginning to be understood. Recent observations suggest that microbial production of siderophores, humic-like substances, and exopolysaccharides are some of the major contributors of marine organic ligands (Hassler et al., 2017), and microbial production and alteration of ligands influence Fe cycling in environments ranging from hydrothermal plumes (Cowen and Bruland, 1985; Cowen et al., 1990) to the open ocean (Lauderdale et al., 2020; Whitby et al., 2024, 2020; Misumi et al., 2013). Strong Fe-binding organic ligands (defined as L_1 ligands) are a heterogeneous mixture of microbially produced compounds that are operationally classified based on their binding strength with Fe (defined as $\log K_{\text{Fe}',\text{Fe}}^{\text{cond}} > 12$). They are thermodynamically favored to complex and stabilize external sources of Fe to prevent its scavenging and removal (Fishwick et al., 2014; Aguilar-Islas et al., 2010).

Siderophores are the strongest known Fe-binding organic ligands. They are produced by bacteria and fungi to facilitate Fe uptake and solubilize otherwise inaccessible phases in the marine environment (Butler, 2005; Manck et al., 2022). They have primarily been considered an important microbial strategy for Fe acquisition in the low dissolved Fe (dFe) surface ocean (Vraspir and Butler, 2009; Butler, 2005). However, siderophore uptake and biosynthesis genes were observed in > 70 % of Fe-related bacterial transcripts in a hydrothermal environment in the Guaymas Basin (Li et al., 2014), have been identified in oxygen-deficient zones (Moore et al., 2021), and are a common Fe acquisition strategy within terrestrial and pathogenic ecosystems (Sandy and Butler, 2009), all of which are environments where Fe concentrations are orders of magnitude higher than surface seawater.

Previous studies have examined total concentrations of Fe-binding ligands in hydrothermal plumes and throughout the deep ocean (Sander and Koschinsky, 2011; Hawkes et al., 2013b; Mahieu et al., 2024; Buck et al., 2018; Kleint et al., 2016), as well as siderophores observed below the euphotic zone (Park et al., 2023; Boiteau et al., 2019; Bundy et al., 2018; Moore et al., 2021; Li et al., 2024). A "stabilizing agent" has been proposed for the long-range transport of hydrothermal dFe into the ocean interior, which has been hypothesized to be inorganic colloids (Fitzsimmons et al., 2017; Fitzsimmons and Boyle, 2014; Yücel et al., 2011; Lough et al., 2019), organic ligands including strong ligands and weaker ligands (Hawkes et al., 2013b; Mahieu et al., 2024; Kleint et al., 2016; Hassler et al., 2020; Slagter et al., 2019), or a combination of the two. The role of strong Fe-binding ligands in hydrothermal dFe transport represents an important knowledge gap in how hydrothermal vents may impact the ocean dFe inventory (Resing et al., 2015) and how siderophores may influence Fe transformations in hydrothermal plumes. While genetic evidence suggests that siderophore cycling may occur in hydrothermal systems (Li et al., 2014), no previous studies have ever directly measured siderophores in hydrothermal systems due to the high-sample-volume requirements, difficulty in obtaining deep-ocean trace metal samples, and the time-intensive nature of the analyses. Here, for the first time, we identified siderophores and siderophore-producing microbes in 11 geochemically distinct hydrothermal plume environments along the slow-spreading (20–50 mm yr−¹) Mid-Atlantic Ridge (MAR). Four black smokers (high temperature, high Fe), four off-axis sites, one diffuse vent (low temperature, low Fe), one alkaline vent (pH 9–11, very low Fe), and one non-vent fracture zone were investigated using both competitive ligand exchange–adsorptive cathodic stripping voltammetry and state-of-the-art liquid chromatography coupled to electrospray ionization mass spectroscopy (Boiteau et al., 2016) in a targeted approach to search for known siderophores and possible compounds present in the L_1 ligand pool in hydrothermal plumes. Microbial community analysis was also compared at three sites to understand whether siderophore production impacts Fe transformation in hydrothermal plumes.

2 Results and discussion

2.1 The role of iron-binding ligands in hydrothermal plumes

Strong Fe-binding ligands (L_1) have previously been found in neutrally buoyant hydrothermal plumes across a variety of systems (Wang et al., 2022; Bennett et al., 2008; Tagliabue et al., 2017; Hawkes et al., 2013b; Resing et al., 2015; Buck et al., 2018). However, the relationship between organic ligands and dFe has never been investigated together systematically across a wide variety of vents in the same study. In this work, the average binding strength and concentration of organic Fe-binding ligands were quantified in 11 vent systems that spanned a wide range in dFe concentrations (0.41– 90 nM) and underlying vent geology. Over 99 % of dFe in the neutrally buoyant plume samples was complexed by L_1 ligands, and the ligands were almost always completely saturated with dFe, meaning Fe-free "excess" L_1 ligands capable of binding additional Fe were present in low concentrations $(< 1 \text{ nM}$; Fig. S1 in the Supplement). As a result, dFe concentrations were tightly coupled to L_1 ligands in a nearly 1 : 1 ratio (Fig. [1d](#page-2-0)), similar to previous studies in other neutrally buoyant plumes (Fig. [1e](#page-2-0)) (Buck et al., 2015, 2018).

The strong coupling between dFe and ligands was only observed at sites where L_1 ligands were detected. Some samples that were closer to the buoyant plume and vent source contained high concentrations of weaker ligands (log $K_{\text{Fe}',\text{FeL}}^{\text{cond}}$ < 12; Tables S2 and S3 in the Supplement) whose concentrations had no correlation with dFe. This is consistent with these environments likely being dominated by complex Fe phases, which could include various inorganic forms (e.g., nanopyrite, Fe oxyhydroxide) as well as mixed organic phases of Fe as hydrothermal fluids initially mix with

Figure 1. Dissolved iron is strongly correlated with L_1 iron-binding ligands in diverse hydrothermal systems. (a) Station map showing the 11 sites investigated along the MAR. Known hydrothermal vents are marked as red triangles (Beaulieu and Szafranski, 2020). Two expanded ´ inset maps for (b) Rainbow and (c) TAG hydrothermal vent fields. For additional information about vent site characteristics, refer to Table [1.](#page-5-0) (d) dFe versus L₁ iron-binding ligands at each vent site in this study showing a ~ 1 : 1 correlation ($m = 0.88$, $R^2 = 0.96$) with dFe in neutrally buoyant plumes along the MAR. (e) dFe versus L¹ ligands from previous studies over the ridge axis and ∼ 80 km from the ridge axis in the southern East Pacific Rise hydrothermal plume (Buck et al., 2018) and over TAG hydrothermal vent field (Buck et al., 2015). The solid black lines in (d) and (e) are the 1 : 1 ratio line between dFe and ligand concentrations, and dashed lines show the linear regression for the corresponding data. Square symbols refer to spreading centers, triangles refer to fracture zones, and circles refer to alkaline vents. Error bars represent the 95 % confidence interval of the data fit as calculated by ProMCC (Omanovic et al., 2015). The map was created using GeoMapApp version 3.6.14.

oxygenated seawater. High concentrations of weaker ligands have also been observed in samples near the vent orifice in previous studies (Hawkes et al., 2013). These ligands can include humic-like substances, exopolysaccharides, or other organic degradation products (Slagter et al., 2019; Hassler et al., 2020; Mahieu et al., 2024; Hawkes et al., 2013b). In this study, we were not able to discern the exact chemical

composition of the ligands we detect via voltammetric methods, and thus the weaker and some portion of the stronger ligands we observe likely represent a mix of different inorganic and organic ligands. Similar to what was described in Hawkes et al. (2013b), the ligands we measure could represent multiple layers of coordination bonds, forming complex Fe phases, similar to the "onion" concept (Mackey and Zirino, 1994). For example, colloidal Fe phases are common in hydrothermal plumes and can form aggregates that bind Fe but not in traditional organic coordination bonds (Fitzsimmons et al., 2017; Honeyman and Santschi, 1989). There are also likely processes occurring near the vent source in such a complex environment that causes some Fe phases to be in various stages of disequilibria that we also measure as ligands via our voltammetric methods.

The sources of weaker Fe-binding ligands (log $K_{Fe',FeL}^{cond} <$ 12) that have been observed in other hydrothermal plumes are not well understood, and their impact on Fe cycling over the lifetime of neutrally buoyant plume is unclear. Recent studies have shown microbes may use siderophores or siderophore-like (strong binding ligands) ligands to access Fe associated with weaker ligands – such as humic substances and thiols – to enhance the bioavailability of Fe (Kuhn et al., 2014; Muller, 2018). However, to date, the few studies that have explored ligand concentrations and binding strengths within hydrothermal plumes (Buck et al., 2015, 2018; Hawkes et al., 2013b; Kleint et al., 2016; Sander and Koschinsky, 2011; Mahieu et al., 2024) have mixed hypotheses as to the role and sources of weaker-type ligands within plumes. Additional studies are needed to investigate the sources and mechanisms of weaker-type ligands in hydrothermal plumes and understand their impact on the Fe cycle in hydrothermal plumes.

In the neutrally buoyant plume samples, stronger L_1 ligands were present and were correlated with the dFe concentrations (Fig. [1\)](#page-2-0), and weaker ligands were no longer dominant. In other systems with a high dFe and ligand endmember such as estuaries, a decrease in weaker ligands along with dFe concentrations has also been observed (Buck et al., 2007; Bundy et al., 2015). This has been interpreted as scavenging of weaker Fe-binding ligand complexes, while the dFe that remains in solution is that which is bound to stronger ligands (Bundy et al., 2014). A similar control on dFe concentrations by L_1 ligands has also been previously observed in aerosol solubility experiments (Fishwick et al., 2014). There are a few possible explanations for the correlation of dFe and L_1 ligands in the neutrally buoyant plume. One possible explanation is that both the dFe and L_1 ligands originate from the vent fluids themselves, yielding a tightly coupled hydrothermal end-member. However, the concentration of L¹ ligands did not correlate with excess mantle helium-3 $(^{3}He_{xs};$ Fig. S2, Tables S2 and S3) (Lough et al., 2023), a nearly conservative tracer of the mixing of hydrothermal fluids with seawater (Buck et al., 2018). Moreover, our samples closer to the vent source were dominated by weaker organic ligands showing no correlation to dFe. This suggests the L_1 ligands were not directly sourced from the vent fluids along with dFe. Biological sources represent another likely explanation for the coupling of L_1 ligands and dFe if the ligands observed in the neutrally buoyant plume are from bacteria that produced them in surrounding deep-ocean seawater that was then entrained, local production from vent biota and/or microbial mats, diffusion from microbial production in sediments, or in situ production by bacteria within the neutrally buoyant plume (Dick et al., 2013; Li et al., 2014; Sheik et al., 2015).

2.2 The presence of siderophores in hydrothermal systems

Siderophores were measured in a subset of the samples to further explore the source of the L_1 ligands coupled to dFe in the neutrally buoyant plume. Marine organic-ligand composition changes with environmental gradients (Gledhill and Buck, 2012; Boiteau et al., 2016), making the structure and functional groups of siderophores identified in hydrothermal samples of particular interest. Somewhat surprisingly, siderophores were found in all samples, and we observed a large diversity of siderophores with high confidence using mass-to-charge ratio (m/z) , MS/MS spectra, and specific chromatographic characteristics (Fig. [2a](#page-4-0)). On-axis spreading centers contained the highest dFe concentrations $(> 20 \text{ nM})$ and a wider variety of siderophores than samples from fracture zones, diffuse, and off-axis sites (dFe ≤ 1 nM). The greatest number of distinct siderophores was identified at Lucky Strike, Broken Spur, Rainbow, and TAG (Fig. [2\)](#page-4-0). On average, 13 compounds were identified with high confidence per on-axis spreading center sample, compared with 5 per diffuse–fracture zone sample, and 2.5 per off-axis sample (Figs. [2b](#page-4-0) and S4). Mixed-type siderophores – containing different moieties that bind to Fe(III) – were common at all sites. Hydroxamates were identified at and around spreading centers, yet none of these were detected with high confidence in samples from diffuse–fracture zones (Fig. S4). Summed siderophore abundance in neutrally buoyant plumes above spreading centers was similarly more than twice that of samples from fracture zones or off-axis (Fig. [2c](#page-4-0)). Thus, vent type and proximity played a role in the diversity and abundance of siderophore types observed, likely related to the diversity of the microbial community and/or unique Fe acquisition strategies across sites.

Siderophores are putatively part of the operational L_1 ligand pool based on their binding strength (Gledhill and Buck, 2012), and patterns in their distributions were similar to those of the strong ligands. The peak areas of each putative siderophore we identified were used as a proxy for concentrations (Sect. [3.3\)](#page-8-0), and these concentrations significantly correlated with dFe, as observed with dFe and L_1 ligands (Fig. [2b](#page-4-0)). Siderophores were present in concentrations similar to the surface ocean (Park et al., 2022; Boiteau et al., 2016; Moore et al., 2021; Bundy et al., 2018) and were equivalent to concentrations representing 0.01 %–0.4 % of the total L_1 ligands (Table [1\)](#page-5-0). This is a substantial underestimate of siderophore contributions to the L_1 ligand pool due to analytical constraints in identifying unknown siderophores. Recent work on siderophore biosynthesis pathways and advances in genome mining suggest that known siderophores represent a

C. L. Hoffman, P. J. Monreal, et al.: Microbial strong organic-ligand production 5237

small fraction of what is expected to be produced in nature (Hider and Kong, 2010; Reitz et al., 2022), and our analyses in this study were limited to only known siderophores. We also restricted our reporting to compounds identified with very high confidence (Figs. [2a](#page-4-0) and S3). In addition, most siderophores are not commercially available to use as standards, and individual siderophores have different ionization

or extraction efficiencies. The extraction efficiency for the solid phase extraction technique is approximately 5 %–10 % for bulk Fe-binding organics (Bundy et al., 2018) and 40 % for a siderophore standard (Waska et al., 2015). Employing both corrections yields siderophore contributions to the total L₁ pool of 0.1 %–4 % and 0.025 %–1 %, respectively. We are inevitably missing many naturally occurring unknown com-

in Table [1.](#page-5-0) Symbols follow Fig. [1.](#page-2-0)

highest confidence). Gray boxes indicate a detection with lower confidence (Sect. 3.4), and white boxes indicate no detection at those sites. The y axis is ordered from top to bottom in terms of descending mass of the apo (without Fe) form of the siderophore. (b) Model II ordinary least squares regression on dFe versus summed siderophore concentrations (of detections in Fig. [2b](#page-4-0)), calculated from peak areas, at each site. Since the siderophore analysis was performed on pooled samples, the dFe values in the regression are weighted values based on measured dFe and volume of each constituent of the pooled sample. The vertical error bars represent the standard deviation of dFe of the constituents. TAG (Station 35) – denoted by the asterisk – was not included in the regression due to its large range of dFe values and outlier behavior. (c, d) Fe-bound versus total summed concentration of (c) all types of siderophores and (d) amphiphilic siderophores at each station. The vertical green lines separate fracture and diffuse sites from off-axis sites, and vertical red lines separate off-axis from on-axis sites as defined

| Vent name | Abbreviation | Station | Geology | Host rock | Vent type | Spreading rate $(mmyr^{-1})$ | Summed putative siderophore concentration (pM) | Summed siderophore concentration/ L_1 ligand $(\%)^*$ |
|-----------------------------------|--------------|---------|---------------------|------------------------|--------------------------|------------------------------------|--|---|
| Lucky Strike | LS | 7/8 | spreading center | gabbro | black smoker | 20.2 | 4.38 | $0.034 - 0.19$ |
| 33 km east of Rainbow | CER | 12 | spreading center | $\qquad \qquad -$ | | | 0.537 | $0.013 - 0.017$ |
| Rainbow | \mathbb{R} | 38 | spreading center | ultramafic | black smoker | 20.6 | 12.1 | n.a. |
| Hayes Frac- ture Zone | HFZ | 21 | fracture zone | peridotites/ gabbro | \equiv | 21.2 | 2.74 | $0.20 - 0.39$ |
| Lost City | LC | 23 | fracture zone | ultramafic/ gabbro | alkaline | 22.6 | 3.47 | $0.27 - 0.35$ |
| Broken Spur | BS | 24 | spreading center | gabbro | black smoker/ diffuse | 22.9 | 6.30 | $0.07 - 0.29$ |
| 29 km north of TAG | CNT | 26 | spreading center | $\qquad \qquad -$ | | | 0.968 | $0.014 - 0.079$ |
| 30 km west of TAG | CWT | 30 | spreading center | $\qquad \qquad -$ | $\overline{}$ | $\overline{}$ | 2.91 | 0.15 |
| 30 km east of TAG | CET | 31 | spreading center | $\qquad \qquad -$ | | $\overline{}$ | 2.19 | 0.31 |
| Trans- Atlantic Geotraverse | TAG | 35 | spreading center | gabbro | black smoker | 23.6 | 6.61 | 0.18 |
| Low Temp Slope | LTS | 37 | | | diffuse fluids | | 1.13 | $0.079 - 0.087$ |

Table 1. Characteristics of sample locations along the Mid-Atlantic Ridge.

Spreading rates along the Mid-Atlantic Ridge were gathered from the InterRidge Vents Database v3.4. Host rock groups were determined from previously discussed classifications (Bazylev, 1997; Kelley and Shank, 2010). Off-axis sites – 33 km east of Rainbow, 29 km north of TAG, 30 km east of TAG, and 30 km west of TAG – were far-field locations of their respective vent field. Low Temp Slope was a diffuse-dominated site that was sampled for the first time as a part of this study. Summed putative siderophore concentrations and the
percent of L₁ ligand are reported for com ENV columns for Fe-binding organics. Average L₁ ligand and siderophore concentrations can be viewed in Table S3, and concentrations for individual siderophores can be observed in Table S5. [∗] The siderophore sample at each site was pooled from ligand samples, so the percentage of siderophores in the ^L¹ pool is presented as a range based on the range of L_1 concentrations at each site. "n.a." denotes unable to be determined. – denotes unknown.

pounds, and thus we consider this a lower bound. Regardless of the small percentage contribution to total L_1 ligands, it is evident that microbially produced siderophores were ubiquitous across all vent sites and had similar distributional patterns as L_1 ligands. There are also likely other compounds such as some strong binding humics that are also contributing to the L_1 ligand pool (Laglera and van den Berg, 2009). Future work with much larger water volumes will be able to reduce uncertainty and identify a greater number of compounds. Still, the identification of siderophores here – as well as their relationship with dFe – provides compelling evidence that microbial production of ligands is responsible for at least some portion of the tight coupling between L_1 and dFe in hydrothermal systems along the MAR.

The presence and diversity of siderophores identified in this system was surprising given the relatively high Fe concentrations of hydrothermal environments, but some compelling patterns were observed. Amphiphilic siderophores comprised 57 % of the siderophores in our samples (Fig. S5), supporting the ubiquity of amphiphilic siderophores in marine environments (Butler and Theisen, 2010). Amphiphilic siderophores were found in concentrations between 0.3– 4.7 pM, with the highest concentrations found at Rainbow (Fig. [2d](#page-4-0), Table S6). These concentrations were similar to those observed in the upper ocean (Boiteau et al., 2019, 2016; Bundy et al., 2018). Amphiphilic siderophores have long hydrocarbon tails that can be embedded into the lipid bilayer of the bacterial cell membrane, providing a mechanism to shuttle Fe into the cell and prevent diffusive loss (Martinez et al., 2003). Marine bacteria produce suites of amphiphilic siderophores as a way to adapt to the change in hydrophilicity in the surrounding environment (Homann et al., 2009; Sandy and Butler, 2009). Amphiphilic siderophores in plumes could be a way for bacteria to access Fe as they are physically transported and cope with strong chemical gradients, similar to the production of multiple siderophores in terrestrial and pathogenetic systems as a means to access inorganic particulate Fe for cellular uptake and storage (Hider and Kong, 2010).

2.3 Microbial sources of siderophores in hydrothermal plumes

The high diversity of siderophores across a huge range of hydrothermal vent systems revealed several surprising aspects of Fe cycling. The biosynthesis of a siderophore is energy-intensive and is regulated by Fe concentration in the surrounding environment (Rizzi et al., 2019). Siderophore presence suggests that bacteria are producing these compounds despite the overall higher Fe concentrations in the deep ocean and within hydrothermal plumes. Consistent with siderophore utilization in terrestrial ecosystems (Hider and Kong, 2010; Sandy and Butler, 2009), one hypothesis is that siderophore production is beneficial to bacteria in the plumes for transforming Fe from otherwise inaccessible forms, such as particulate nanopyrites or Fe oxyhydroxides that are present close to the vent source. To explore the potential for microbial production of siderophores, we examined microbial community composition around Rainbow (Stations 11 and 17) and Lucky Strike (Station 7; Tables [1](#page-5-0) and S1) using 16S rRNA gene-based amplicon sequencing to detect bacteria with the metabolic potential to synthesize siderophores (Figs. [3](#page-7-0) and S11). The presence of taxa encoding siderophore biosynthetic gene clusters indicates whether the microbial community has the genetic potential to produce the compounds we observed. Bacterial genera containing known siderophore producers were found at all three MAR sites examined, and putative siderophore producers represented 3 %–20 % of the relative abundance of the community (Fig. [3\)](#page-7-0). Putative siderophore producers were more abundant in the $3 \mu m$ (particle-attached) size fraction than in the $0.2 \mu m$ (free-living) fraction, suggesting siderophore production is more common in particle-associated bacteria in hydrothermal environments.

We found microbial genera in our samples that can produce a subset of the siderophores identified here, including ferrioxamines, vibrioferrin, and acinetoferrin (Vraspir and Butler, 2009; Butler, 2005; Moore et al., 2021; Bundy et al., 2018; Boiteau et al., 2016). Genera with the genetic potential to produce ferrioxamines were present at all three sites, while those known to produce vibrioferrin were present at Lucky Strike and Rainbow, and those producing acinetoferrin were also present at Rainbow (Tables S1 and S7). Mycobactins were detected with high confidence in every sample of this study, and genes encoding mycobactin have been detected in a cultured organism from a hydrothermal system (Gu et al., 2019), but no mycobactin producers were identified in this study. We detected woodybactin D with high confidence in 5 out of 11 sites analyzed and compared it to the known siderophore library (Fig. [2\)](#page-4-0). Woodybactin-D biosynthetic genes were not identified in any of the genera from the 16S rRNA gene amplicon sequences; however, woodybactin D is a carboxylate siderophore isolated from *Shewanella* (Carmichael et al., 2019), and groups of deepsea *Shewanella* (Kato and Nogi, 2001) were found in the dataset (Fig. S11). The biosynthesis genes for many of the siderophores identified are unknown. Thus, finding genera capable of producing only a subset of the siderophores characterized is not surprising. The observation that a portion of the in situ microbial community is capable of synthesizing siderophores (Fig. [3\)](#page-7-0) suggests that siderophore production is more widespread in the deep ocean than previously believed and could contribute to the "microbial iron pump" in hydrothermal plumes (Li et al., 2014)

2.4 The impact of strong ligands and siderophores on dissolved iron in neutrally buoyant plumes

Evidence that siderophores are ubiquitous in marine environments – including higher Fe environments – has been increasing (Park et al., 2022). The high dFe associated with hydrothermal plumes may still not be high enough to suppress siderophore production due to the elevated Fe requirements of heterotrophic bacteria (Tortell et al., 1996). It is also likely that not all of the Fe is bio-accessible in hydrothermal plumes. Soil microbes secrete siderophores to solubilize particulate Fe (Crowley et al., 1991), and similar processes could be occurring in hydrothermal plumes, where Fe mineral phases associated with organic compounds are common (Hoffman et al., 2020; Toner et al., 2009; Hoffman et al., 2018; German and Seyfried, 2014; Holden et al., 2012; Fitzsimmons et al., 2017). Although our measurements suggest that dFe in the neutrally buoyant plume is likely dominated by organic complexation, the L_1 measurements alone cannot distinguish between purely organic phases or a mixture of inorganic and organic ligands in complex aggregations or small colloids, as discussed above (Sect. 2.1). Given the evidence from particulate Fe studies in neutrally buoyant plumes (Yücel et al., 2011; Fitzsimmons et al., 2014; Hoffman et al., 2020; Toner et al., 2009; Fitzsimmons et al., 2017; Hoffman et al., 2018), it is highly likely that some portion of what is detected in the L_1 pool is a mixture of organic and inorganic Fe in small colloids, which are operationally in the dFe pool (Fitzsimmons et al., 2017). It is also telling that $4-5\times$ more siderophore-producing genera were found to be particle-associated (Fig. [3\)](#page-7-0), providing additional evidence that siderophores might be produced to solubilize particulate Fe or access other colloidal phases. Further work that assesses why bacteria are producing siderophores in neutrally buoyant plumes will be important for understanding microbial metabolism in these systems and the impact of siderophore production on Fe dispersal.

Figure 3. Relative abundance of putative siderophore-producing taxa. Bar height indicates the proportion of 16S rRNA genes recovered in each sample, separated by depth from water surface, filter size fraction, and site location. Colors correspond to taxonomy. Genera found in MAR vent microbial communities with members in the antiSMASH database predicted to produce siderophores are depicted at the family level.

Organic Fe-binding ligands have been implicated in playing a critical role in the preservation and transport of hydrothermal dFe into the ocean interior (Bennett et al., 2011; Hoffman et al., 2018; Fitzsimmons et al., 2017; Toner et al., 2009; Bennett et al., 2008; Resing et al., 2015; Buck et al., 2018; Sander and Koschinsky, 2011). In this work, L_1 ligands were tightly coupled to dFe in neutrally buoyant plumes along the MAR, and the presence of siderophores in these samples provided evidence for the first time that at least some of these ligands are microbially produced. How these complexes may facilitate the exchange of Fe between dissolved and particulate phases (Fitzsimmons et al., 2017) and whether siderophores are present across additional hydrothermal vent systems will aid in understanding how microorganisms might play a role in shaping the hydrothermal dFe supply to the deep ocean.

3 Appendix: Materials and methods

3.1 Sampling and cruise transect

Samples were collected as part of the 2017–2018 UK GEO-TRACES GA13 section cruise along the Mid-Atlantic Ridge (FRidge GA13). Water samples from 11 venting and nearventing locations were collected using a Sea-Bird 911 conductivity, temperature, and depth (CTD) titanium rosette using conducting Kevlar wire with an oxidation–reduction potential (ORP) sensor to detect plumes. Teflon-coated OTE (Ocean Test Equipment) bottles were pressurized to approximately 7 psi with 0.2 µm filtered air using an oil free compressor. A Sartobran 300 (Sartorius) filter capsule $(0.2 \,\mu\text{m})$ was used to collect filtered seawater samples into clean 250 mL LDPE sample bottles. Bottles and caps were rinsed three times with the filtered sample before being filled. Samples were stored frozen at −20 °C for Fe-binding organic ligand characterization by voltammetry and mass spectrometry.

3.2 Fe-binding ligand concentration and binding strengths competitive ligand exchange–adsorptive cathodic stripping voltammetry

Fe-binding ligand concentrations and binding strengths (defined as conditional binding constants; log $K_{Fe',FeL}^{cond} >$ 12) were determined by competitive ligand exchange– adsorptive cathodic stripping voltammetry (CLE–ACSV) with a BAS*i* controlled growth mercury electrode (CGME) with an Ag/AgCl[−] reference electrode and platinum auxiliary electrode (Bioanalytical Systems Incorporated). Using previously established methods (Abualhaija and van den Berg, 2014; Buck et al., 2015; Hawkes et al., 2013b; Buck et al., 2018; Bundy et al., 2018), 40 frozen filtrate $(< 0.2 \,\mu m)$ samples with dFe concentrations between 0.41–11.67 nM (Tables S1 and S2) were thawed in a 4 °C fridge prior to analysis. A 15-point titration curve was analyzed for each sample. Briefly, within each titration, every point sequentially received 10 mL of sample, 7.5 mM of borate–ammonium buffer, 10μ M salicylaldoxime (SA) added ligand, and a dFe addition (see Supplemental Methods 1.1. for additional details). Samples were then equilibrated overnight before being measured on the BASi. Data were collected using the Epsilon Eclipse Electrochemical Analyzer (v.213) with a deposition time of 120 s and analyzed using ElectroChemical Data Software (v2001-2014) and ProMCC (v2008-2018) to determine peak areas and Fe-binding ligand parameters, respectively. All results were confirmed to fall within the analytical window of the method by comparing the side reaction coefficient of the added ligand α_{SA} to the side reaction coefficient of the natural ligands detected (α _L). When the α _L was within an order of magnitude of α_{SA} , then the results were deemed to fall within the analytical window.

3.3 Reverse titration–CLE–ACSV

Reverse titration–CLE–ACSV (RT–CLE–ACSV) (Hawkes et al., 2013a) was completed on 10 samples from Broken Spur and TAG hydrothermal vent fields with dFe concentrations between 19.01–90.25 nM (Table S3). Briefly, a 10 point titration curve was analyzed for each sample with each titration point consisting of 10 mL of sample buffered with 7.5 mM boric acid and the competitive ligand 1-nitroso-2 napthol (NN) additions. All samples were analyzed on a BASi controlled growth mercury electrode (CGME) with the Epsilon Eclipse Electrochemical Analyzer (v.213) and deposition time of 120 s. For each sample, competitive ligand NN additions were 0.5, 1, 2, 3, 4, 6, 9, 15, 20, and 40 µM. Samples were equilibrated overnight and purged with N_2 (99.99%) for 5 min before analysis. At the end of each titration, three Fe additions (3–15 nM) were made to the final titration point to obtain the total concentration of Fe in equilibrium with ligands. Data were analyzed using ElectroChemical Data Software (version 2001–2014) to acquire peak areas and a package in R using the model parameters of $\beta_{\text{FeNN3}} = 5.12 \times 10^{16}$, $\chi_{\text{min}} = 0.8$, $\chi_{\text{max}} = 0.9$, and $\text{c1high} = 0.75$ to determine the Fe-binding ligand parameters (Hawkes et al., 2013a). These parameters were chosen based on the recommendations for undersaturated samples and titration curves where ip_{max} was not reached (Hawkes et al., 2013a). All other parameters within the model were kept at the default values.

3.4 Siderophore quantification and characterization

In addition to measuring Fe-binding ligands by voltammetry, we also identified and quantified siderophores. Between 0.65–1.5 L of 0.2 µm filtered seawater pooled from ligand samples at each site (described above) was pumped slowly (15–20 mLmin−¹) onto a polystyrene–divinylbenzene (Bond

Elut ENV) solid phase extraction (SPE) column (Bundy et al., 2018; Boiteau et al., 2016). SPE columns were rinsed with Milli-Q water and stored at −20 °C until analysis. For the analytical measurements, samples were thawed in the dark, eluted in 12 mL of distilled methanol, and dried down to between 0.2–0.5 mL of sample eluent (Table S1). Aliquots were analyzed by reverse-phase liquid chromatography (LC) on a trace-metal-clean bio-inert LC (Thermo Scientific Dionex 3000 NCS). The LC was interfaced with an electrospray ionization–mass spectrometer (ESI-MS; Thermo Scientific Q Exactive HF) to identify and quantify the compounds based on accurate mass $(MS¹)$ and the fragmentation $(MS²)$ data (Bundy et al., 2018; Boiteau et al., 2016). MSconvert (ProteoWizard) was used to convert MS data to an open-source mzXML format, and two stages of data processing were conducted using modified versions of previously reported R scripts (Bundy et al., 2018; Boiteau et al., 2016). In the first stage, mzXML files were read into R using new package "RaMS" (Kumler and Ingalls, 2022) , and extracted ion chromatograms (EICs) were generated for each targeted m/z of interest from an in-house database of siderophores. The m/z targets were the ionized apo, ⁵⁴Febound, and ⁵⁶Fe-bound version of each siderophore, with a tolerance of 7.5 ppm. Putative siderophore candidates were filtered through a series of hard thresholds such that $MS¹$ spectra were quality-controlled to contain a minimum of 25 data points, and the maximum intensity of each EIC was greater than 10^4 counts. Spectra meeting these criteria and containing either 54 Fe-bound and 56 Fe-bound m/z peaks within 30 s of each other or an apo peak were displayed for the user to further inspect peak quality and make the final decision of whether to move on to Stage 2 of processing with a given siderophore candidate.

Stage 2 of processing extracted $MS²$ spectra of the apo and Fe-bound forms of candidate siderophores to compare with the predicted $MS²$ generated by in silico fragmenter MetFrag (Ruttkies et al., 2016). The in silico fragmenter feature was run with a tolerance of 10 ppm in " $[M + H]$ ^{+"} and " $[M + Na]$ ⁺" modes. A confidence level of 1–4, from highest to lowest confidence, was then assigned to putative siderophores based on the following criteria: (1) peaks were present in $MS¹$ and $MS²$ spectra, and at least one of the three most-intense $MS²$ fragments matched in silico fragmentation; (2) peaks were present in $MS¹$ and $MS²$ spectra, and smaller-intensity fragments matched in silico fragmentation; (3) peaks were present in $MS¹$ and $MS²$ spectra, but little to no fragments matched in silico fragmentation; and (4) nicely shaped peaks were identified in $MS¹$ spectra, but no $MS²$ spectra were collected (outlined in Table S5; example spectra in Figs. S6–S9). The confidence levels were modeled after reporting standards for metabolite identification (Sumner et al., 2007). MetFrag pulls chemical structures from publicly available databases like PubChem or COCONUT (Sorokina et al., 2021), which contain most but not all variations of siderophores. As such, Fe-bound candidates were

usually run against the apo form available in the database, and for siderophores with similar structures but variations in fatty chain length or double-bond placement, sometimes only one parent structure was available.

A five-point standard curve with known concentrations of siderophore ferrioxamine E was used for quantification of putative siderophores, with a limit of detection of 0.257 nM in the eluent (Fig. S10) or 0.07–0.21 pM in the sample depending on sample-to-eluent volume ratio at each site (Table S1). $MS¹$ peaks were integrated for all putatively identified siderophores, and peak areas were converted to concentration using the standard curve and the concentration factor of sample volume to eluent volume (Fig. S10). Commercial standards are not available for most siderophores, and different compounds have distinct ionization efficiencies in ESI-MS. Thus, the siderophore concentrations reported here are estimates of siderophore concentrations in these environments based on ferrioxamine E, chosen for its commercial availability and its use in prior studies (e.g., Boiteau et al., 2016). Additionally, 1 µM of cyanocobalamin was added as an internal standard to each sample aliquot to address any changes in sensitivity during LC-ESI-MS runs. All putative siderophores that were identified with peak areas less than the detection limit were discarded, and all remaining putative compounds with at least confidence levels 1 and 2 at one site were included in the article and are referred to as siderophores throughout. Siderophore identification remains putative due to inherent uncertainty with assignments by mass, but the confidence levels were designed such that highconfidence candidates contain siderophore-like moieties in their fragments. Limited sample volumes prevented analysis via liquid chromatography–inductively coupled plasma mass spectrometry like previous studies, which, in addition to greater availability of commercial standards and more analytical comparisons between ferrioxamine E with other siderophore types, would allow definitive characterization in future studies. Confidence level 3 and 4 putative siderophores are only included in the Supplement (Table S6). In a final step of quality control, EICs for 13 C isotopologues of candidates were inspected to verify matching peak structure.

3.5 Microbial community analysis

Microbial community composition was assessed in neutrally buoyant plumes and near venting sites at three sites: Lucky Strike (Station 7; 1670 m), 10 km south of Rainbow (Station 17; 2000 m), and 200 km east of Rainbow (Station 11; 600, 1600 and 2250 m). A range of 1–2 L of seawater was filtered by pressure filtration through sequential 25 mm membrane filters housed in polypropylene filter holders (Whatman SwinLok, GE Healthcare, Pittsburgh, Pennsylvania) using a peristaltic pump and silicone tubing. Samples first passed through a 3 µm pore size polyester membrane filter (Sterlitech, Auburn, Washington) and then onto a 0.2 µm pore size polyethersulfone membrane filter (Supor 200, Pall Corporation, Port Washington, New York). Pump tubing was acid-washed with 10 % hydrochloric acid and flushed with ultrapure water between each sample. The filters were flash frozen in liquid nitrogen in 2 mL gasketed bead beating tubes (Fisher Scientific) at sea.

DNA was extracted using modified previously described methods (Santoro et al., 2010). Briefly, cells on the filters were lysed directly in the bead beating tubes with a sucrose– ethylenediaminetetraacetic acid (EDTA) lysis buffer (0.75 M sucrose, 20 mM EDTA, 400 mM NaCl, 50 mM Tris) and 1 % sodium dodecyl sulfate. Tubes were then agitated in a bead beating machine (BioSpec Products) for 1 min and subsequently heated for 2 min at 99 °C in a heat block. Proteinase K (New England Biolabs) was added to a final concentration of $0.5 \text{ mg} \text{ mL}^1$. Filters were incubated at 55 °C for approximately 4 h, and the resulting lysates were purified with the DNeasy kit (Qiagen) using a slightly modified protocol (Santoro et al., 2010). The purified nucleic acids were eluted in 200 µL of DNase-free and RNase-free water and quantified using a fluorometer (Qubit and Quanti-T HS reagent, Invitrogen Molecular Probes).

The 16S rRNA gene was amplified in all samples using V4 primers (Apprill et al., 2015; Parada et al., 2016) (515F-Y and 806RB) following a previously established protocol (Stephens et al., 2020). Amplicons were sequenced using a paired-end 250 bp run, on an Illumina MiSeq 500 and demultiplexed by the UC Davis Genome Center. The resulting 16S rRNA amplicon sequences were filtered and trimmed using the DADA2 pipeline in R (Callahan et al., 2016). Taxonomic assignments were made with version 138.1 of the SILVA SSU database (Quast et al., 2013) (silva_nr99_v138.1_wSpecies_train_set.fa.gz; https://doi.org[/10.5281/zenodo.4587955,](https://doi.org/10.5281/zenodo.4587955) McLaren and Callahan, 2021). Chloroplast and mitochondrial sequences were filtered out of the dataset using the "phyloseq" R package (v. 1.38.0), after which samples had read depths ranging from $9375-65486$ reads (average $28425 + 20014$ reads) and represented 1010 unique amplicon sequence variants (ASVs). Read counts were transformed from absolute to relative abundance, and taxa were aggregated to the family level. The 10 most abundant families present in each sample were visualized using the "ggplot2" package (v. 3.3.5).

In order to assess the potential of the observed prokaryotic taxa to produce siderophores, we downloaded all siderophore biosynthetic gene clusters (BGCs) in the version-3 anti-SMASH secondary metabolite database $(n = 7909)$ and used text-string matching to compare genera containing these BGCs to the genera found in our 16S rRNA gene dataset (Blin et al., 2021). We cross-referenced the nomenclature of antiSMASH-predicted siderophores with that of the siderophores identified by LC-ESI-MS in this study, accounting for minor differences in naming convention between the two databases, to determine whether microbial community members present at each site were predicted to make any of the siderophores that were measured at that site. Stations 38

C. L. Hoffman, P. J. Monreal, et al.: Microbial strong organic-ligand production 5243

and 12 were the closest sites with siderophore measurements for comparison against the taxonomic samples taken at 200 km east of Rainbow and 10 km south of Rainbow, respectively. Samples for microbial taxonomy and siderophore identification were taken from the same location at Lucky Strike and thus directly compared.

Data availability. The CSV data reported in this study were deposited in Zenodo at https://doi.org[/10.5281/zenodo.7325154](https://doi.org/10.5281/zenodo.7325154) (Hoffman and Bundy, 2022). The LC-ESI-MS data were deposited in MassIVE at https://doi.org[/10.25345/C5V97ZW7N](https://doi.org/10.25345/C5V97ZW7N) (Monreal, 2024). Microbial 16S rRNA data were deposited in GenBank under BioProject accession no. PRJNA865382. All data are freely available in each of these data repositories.

Supplement. The supplement related to this article is available online at: [https://doi.org/10.5194/bg-21-5233-2024-supplement.](https://doi.org/10.5194/bg-21-5233-2024-supplement)

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Competing interests. The contact author has declared that none of the authors has any competing interests.

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