



Supplement of

Forest diversity and environmental factors shape contrasting soil-litter fluxes of biogenic volatile organic compounds and methane in three central Amazonian ecosystems

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1 Supplementary Material

2 S1. Soil Texture

3 **Table S1.** Soil particle size distribution in three forest types in central Amazonia. The data
4 represents the percentage composition of total sand, silt, and clay. Each sample is a composite
5 sample from two nearby points in the field. AR = Ancient River Terrace Forest, WS = White Sand
6 Forest, Up = Upland Forest.

7

		Soil Texture		
Forest type	samples	Sand (%)	Silt (%)	Clay (%)
AR	TR1.1	45.449	37.801	16.75
AR	TR1.2	45.449	37.801	16.75
AR	TR1.3	47.639	30.061	22.30
AR	TR1.4	47.639	30.061	22.30
AR	TR1.5	12.38	56.22	31.40
AR	TR1.6	12.38	56.22	31.40
AR	TR2.1	41.543	34.457	24.00
AR	TR2.2	41.543	34.457	24.00
AR	TR2.3	42.078	35.422	22.50
AR	TR2.4	42.078	35.422	22.50
AR	TR2.5	47.17	28.68	24.15
AR	TR2.6	47.17	28.68	24.15
WS	CP1.1	46.15	52.35	1.50
WS	CP1.2	46.15	52.35	1.50
WS	CP1.3	50.07	47.68	2.25
WS	CP1.4	50.07	47.68	2.25
WS	CP1.5	74.398	24.252	1.35
WS	CP1.6	74.398	24.252	1.35
WS	CP2.1	76.599	22.301	1.10
WS	CP2.2	76.599	22.301	1.10
WS	CP2.3	24.134	74.616	1.25
WS	CP2.4	24.134	74.616	1.25
WS	CP2.5	73.883	24.517	1.60
WS	CP2.6	73.883	24.517	1.60
Up	TF1.1	23.011	26.139	50.85
Up	TF1.2	23.011	26.139	50.85
Up	TF1.3	12.595	30.255	57.15
Up	TF1.4	12.595	30.255	57.15
Up	TF1.5	12.698	38.602	48.70
Up	TF1.6	12.698	38.602	48.70
Up	TF2.1	11.892	40.858	47.25

Up	TF2.2	11.892	40.858	47.25
Up	TF2.3	10.194	32.506	57.30
Up	TF2.4	10.194	32.506	57.30
Up	TF2.5	9.587	35.413	55.00
Up	TF2.6	9.587	35.413	55.00

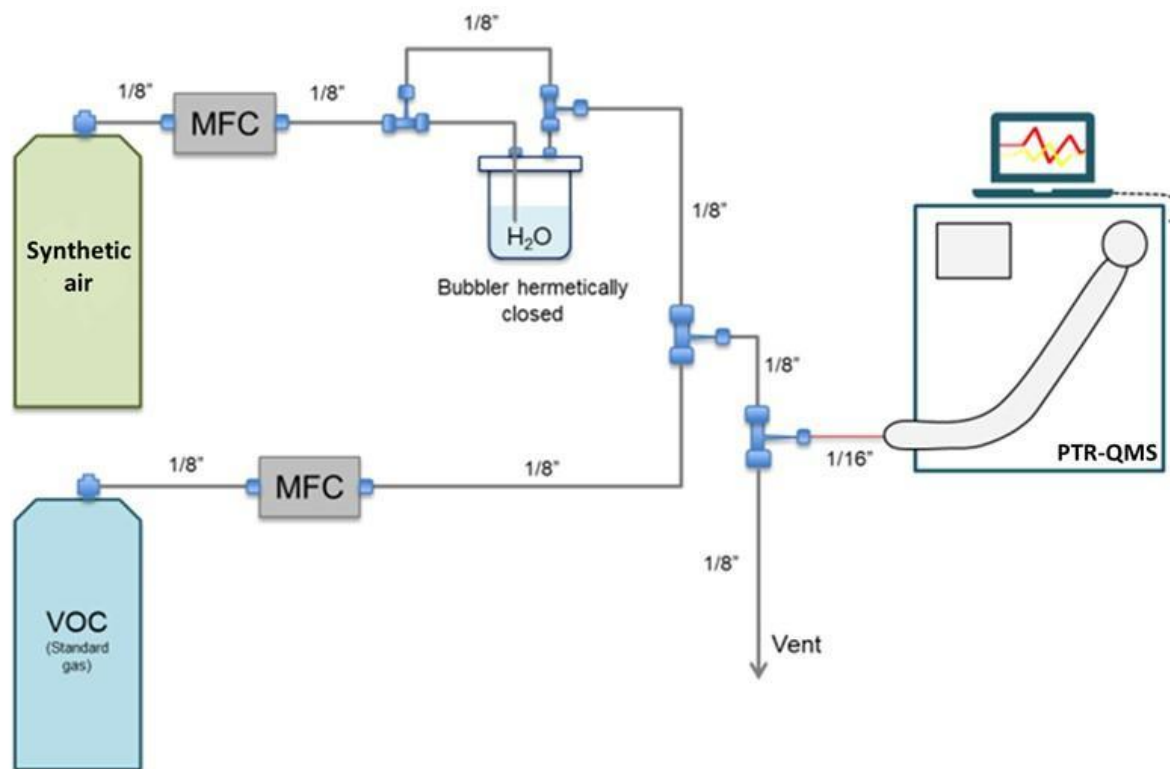
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9 **S2. Calibration gas composition**

10 **Table S2.** Multi-component calibration mixture in synthetic air, as used for the calibration of the PTR-
 11 QMS. Uncertainty is a conservative estimate of the combination of gravimetric preparation and analysis
 12 uncertainties.

Compound	CAS#	Concentration (ppb)	Uncertainty
Formaldehyde	50-0-0	992	5%
Acetaldehyde	75-07-0	426	5%
Methanol	67-56-1	521	5%
Ethanol	64-17-5	528	5%
Acetonitrile	75-05-8	523	5%
Acetone	67-64-1	497	5%
Isoprene	78-79-5	468	5%
DMS	75-18-3	487	5%
a-Pinene	80-56-8	488	5%
Camphene	79-92-5	504	5%
B-Pinene	18172-67-3	512	5%
a-Phellandrene	4221-98-1	432	5%
3-Carene	13466-78-9	458	5%
Limonene	5889-54-8	489	5%
o-Cymene	527-84-4	481	5%
y-Terpinene	99-85-4	476	5%
B-Caryophyllene	87-44-5	98.3	5%
a-Humulene	6753-98-6	95.2	5%

13



14 MFC= mass flow controller

15 **Figure S1.** PTR-QMS calibration setup scheme, the VOC standard gas is as described in Table
 16 S2, figure adapted from Caetano (2022).

17 S3. BVOC & GHG flux calculation

18 Fluxes were calculated using the closed chamber formulation, as given in equation 1 (manuscript,
 19 Section 2.5). It should be considered that the chamber was operated under a continuous flow (0.5 L
 20 min⁻¹), which introduces a constant small dilution of the accumulating BVOCs and GHG. Therefore, the
 21 system is formally a non-steady-state flow-through chamber rather than a strictly static chamber. Using
 22 the closed chamber formula instead of the non-steady-state flow-through chamber formula leads to an
 23 underestimation of 25%, as shown below.

24

25 For a flow-through chamber with volume V , surface area A , flow rate Q , and enclosure time t , the
 26 transient mass balance gives:

27
$$C(t) - C_{in} = \frac{FA}{Q} (1 - e^{-Qt/V}) \tag{S1}$$

28 Solving for flux yields the non-steady-state formulation:

29

$$30 \quad F_{open} = \frac{Q}{A} \frac{\Delta C(t)}{1 - e^{-Qt/V}} \frac{P}{RT} \quad (S2)$$

31

32 where $\Delta C(t) = C(t) - C_{in}$, P is pressure, T is temperature, and R is the gas constant.

33

34 In contrast, the closed chamber formulation assumes linear accumulation:

35

$$36 \quad F_{closed} = \frac{V}{A} \frac{\Delta C(t)}{t} \frac{P}{RT} \quad (S3)$$

37

38 The relative bias introduced by applying the closed formulation to a flow-through system depends only
39 on the dimensionless term Qt/V :

40

$$41 \quad \frac{F_{closed}}{F_{open}} = \frac{1 - e^{-Qt/V}}{Qt/V} \quad (S4)$$

42

43 For the present study:

$$44 \quad t = 25 \text{ min}$$

$$45 \quad Q = 0.5 \text{ L min}^{-1}$$

$$46 \quad V = 21 \text{ L}$$

47 This gives:

$$48 \quad \frac{Qt}{V} = 0.595$$

49

50 and therefore:

$$51 \quad \frac{F_{closed}}{F_{open}} = 0.754$$

52

53 indicating that the closed chamber approach underestimates the flux by approximately:

$$54 \quad 1 - 0.754 = 0.246 \approx 25\%$$

55

56 Thus, for the chamber configuration used here ($A = 0.0855 \text{ m}^2$, $V = 21\text{L}$, $Q = 0.5 \text{ L min}^{-1}$, enclosure time
57 25 min), the expected underestimation due to assuming closed chamber conditions is on the order of
58 25%. Importantly, this correction depends only on the Qt/V ratio and is independent of the absolute
59 magnitude of the flux. The reported fluxes should therefore be considered conservative estimates, and
60 this potential bias does not affect the main conclusions of the study.

61

62

63 **S4. Gas Chromatography - Time of Flight Mass Spectrometer (GC-TOF-MS) analysis**

64 **S4.1 GC-TOF-MS: Material & Methods**

65 After determining gas mixing ratios on the PTR-QMS and the Los Gatos analyzer, samples were
66 collected onto cartridges (stainless steel tubes filled with Tenax TA and Carbograph 5 TD
67 adsorbents, Markes International, UK) at a rate of 200 sccm for 10 min, which resulted in 2L air
68 samples for compound identification and quantification analyses with the GC-TOFMS. The
69 cartridges were stored in an air-conditioned room for up to six months until they were analyzed.

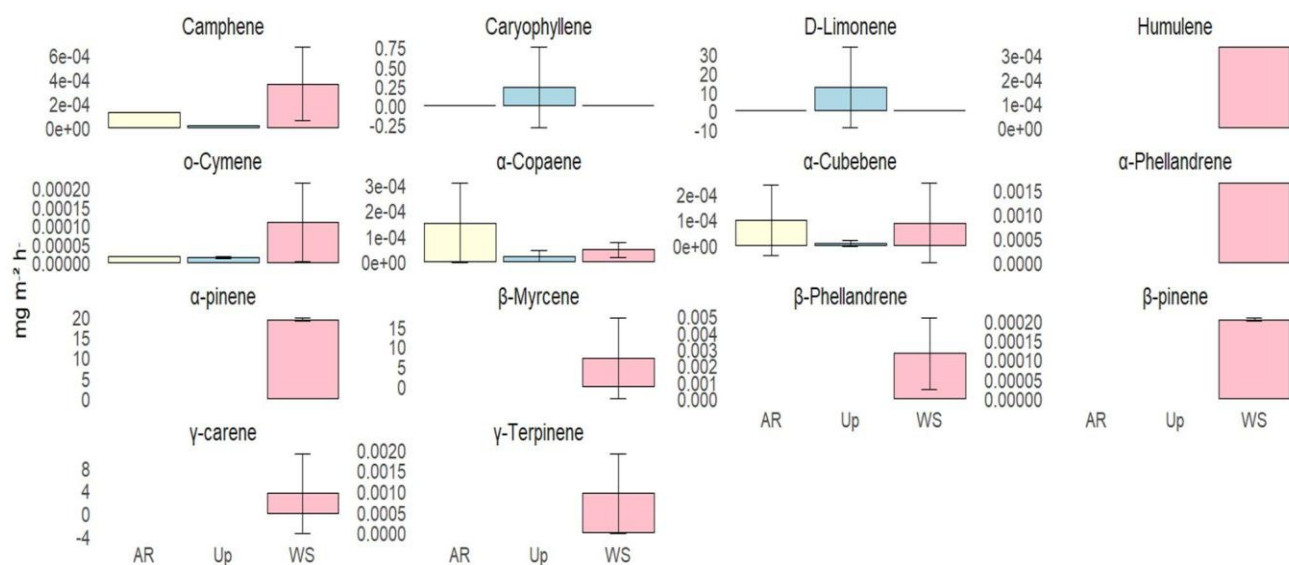
70 Samples were examined with a thermal-desorption gas chromatograph coupled to a time-of-flight
71 mass spectrometer (TD-GC-TOF-MS; Bench TOF Tandem Ionisation, Markes International, UK).
72 A complete description of the instrument, analytical procedure, and performance is available in
73 Zannoni et al. (2020a, 2020b). In the TD-GC-TOF-MS, the analytes are desorbed from the
74 sampling tubes using two sequential stages, performed at 250°C for 10 minutes, then carried by a
75 flow of He into the separating column in the gas chromatograph. The column is a dimethyl TBS
76 β -cyclodextrin column (0.25 μm , 0.25mm ID, 30 mL, from MEGA, Italy), which separates the
77 analytes according to volatility and enantiomeric configuration. The separation method was
78 specifically designed for the separation of chiral monoterpenes ($\text{C}_{10}\text{H}_{16}$) and sesquiterpenes
79 ($\text{C}_{15}\text{H}_{24}$) and consisted of an initial 5 min when the oven temperature was held at 40°C, after which
80 it was increased at a rate of 1.5 °C/min from 40°C to 150°C. Finally, the temperature increased
81 further at a rate of 30°C/min from 150°C to 200°C. Detection was performed by a Time-of-Flight
82 Mass Spectrometer, which fragments the analytes through electron impact ionization at -70 eV for
83 quantification and identification of the chemical species. Identification of the main chemical

84 compounds was obtained by comparing the MS spectra with the MS library for the same ionization
85 energy (NIST library), by injection of a calibration gas standard mixture (19 biogenic volatile
86 organic compounds provided by Apel Riemer, USA), and by use of liquid standards.
87 Chromatogram peak areas were integrated through TOF-DS software provided with the instrument
88 (Markes International, UK). The standard gas used is a certified concentrated mixture (~100 ppbv)
89 of 19 BVOCs, including, among the other species, four pairs of chiral terpenoids (α -pinene, β -
90 pinene, limonene, linalool), isoprene, and one sesquiterpene (β -caryophyllene). The GC-TOF
91 sensitivity was quantified by measuring the concentration of BVOCs in standard sorbent tubes
92 filled with 2 L of gas standard diluted in a pure synthetic air flow with four dilution steps to achieve
93 concentrations in the range of 0.05-1 ppb. Such calibrations were performed before or at the end
94 of each sample batch. The obtained calibration factors were confirmed using standard cartridges
95 containing a known BVOC concentration (~0.150 ppb) in a workflow, with a standard cartridge
96 measured for every 10 samples analyzed. The standard cartridges were analyzed in this way to
97 determine the analysis precision (including any mass spectrometer drift), which was quantified at
98 22% (Zannoni *et al.* 2020a and Zannoni *et al.* 2020b).

99

100 **S4.2 GC-TOF-MS: Results**

101 For all BVOC fluxes studied, emissions were highest in the White Sand Forest. Notably, some
102 BVOCs were detected exclusively in the White Sand Forest and were completely absent from the
103 other forest types (Figure S2).



█ AR (Ancient River Terrace Forest)
 █ WS (White Sand Forest)
 █ Up (Upland Forest)

104

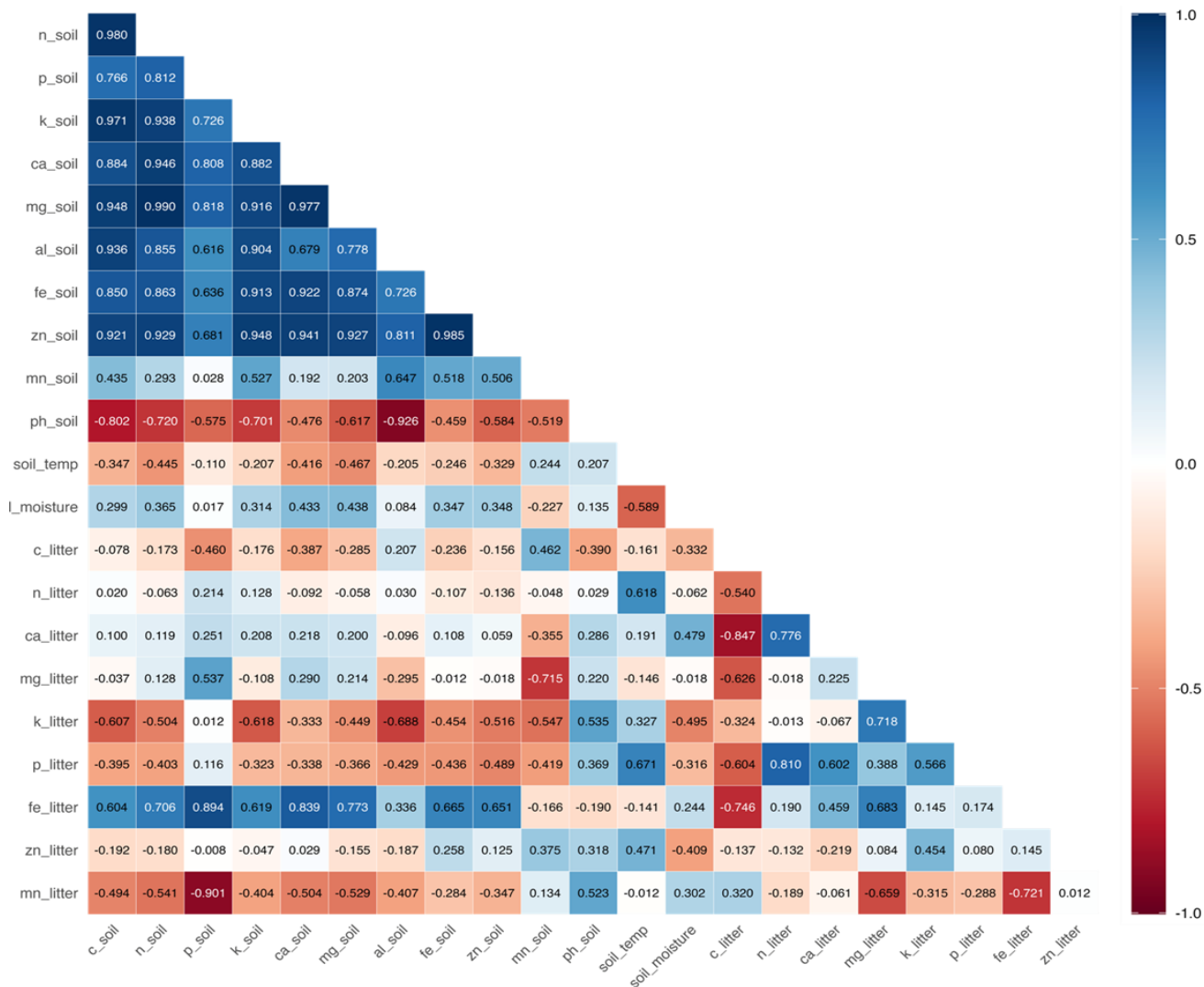
105 **Figure S2.** BVOC Fluxes from the soil and litter across the three forest types. Bar plot of the 12 samples
 106 from each environment.

107

108 **S5. Linear Models (LM)**

109 **S5.1 White sand forest fluxes and their potential predictors**

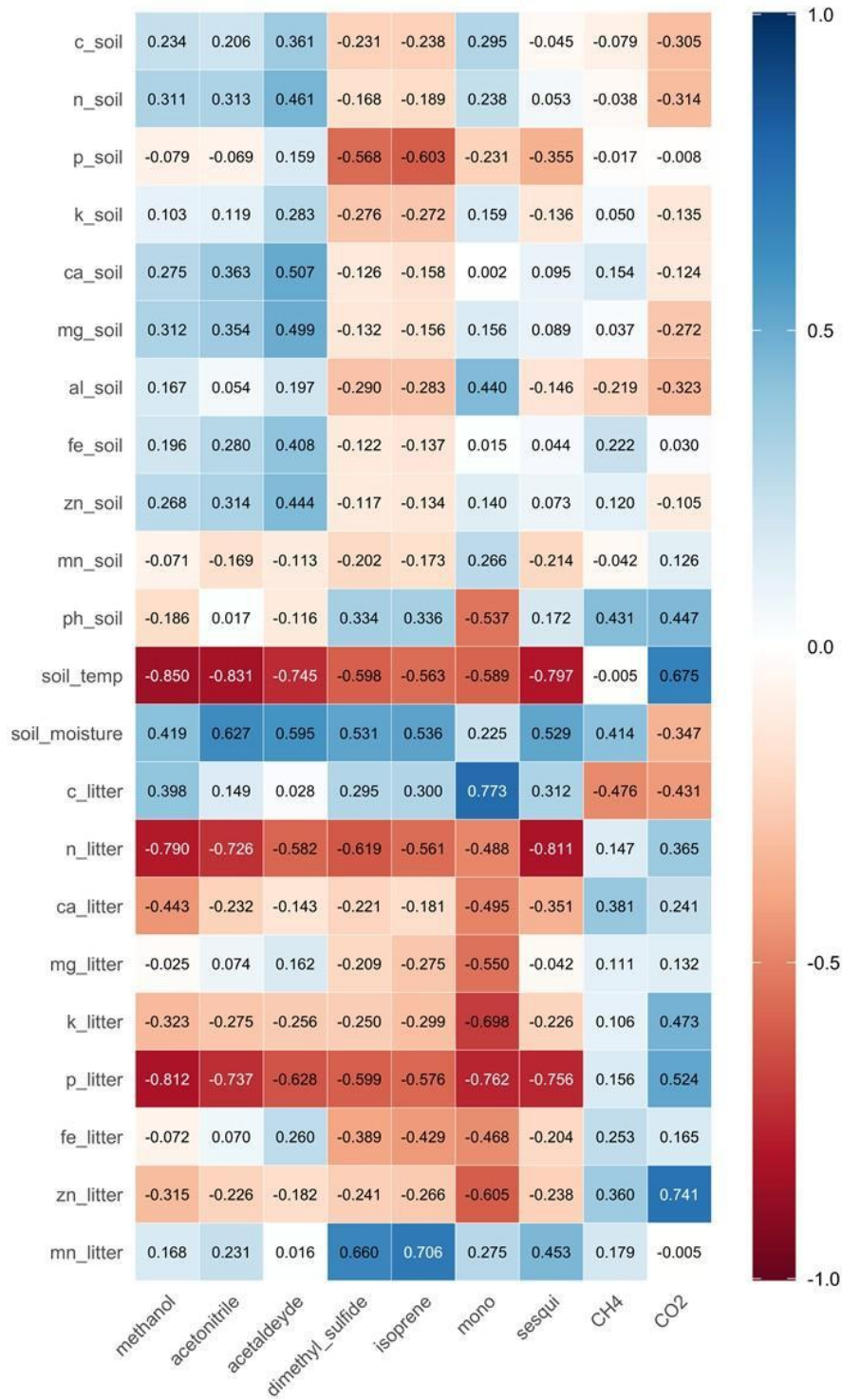
110 Firstly, we investigated the relationships among the potential predictor variables in the white sand
 111 forest ecosystem, independent of their relation to gas fluxes. These predictor variables include soil
 112 and litter characteristics, microbial biomass, soil moisture, and temperature. The Pearson
 113 correlations for these variables are presented in Figure S3a, which highlights the interactions and
 114 dependencies between environmental drivers only.



115

116 **Figure S3a.** Heatmap for the Pearson correlation coefficients for the potential predictors of fluxes
 117 in the white sand forest. N = 12.

118 Secondly, we analyzed the correlations between the potential predictor variables and the response
 119 variables, which are the BVOC and GHG fluxes measured in this study. The Pearson correlations
 120 for these relationships are shown in Figure S3b, which specifically highlights how the fluxes of
 121 BVOCs and GHGs are associated with environmental drivers. This distinction between Figure S3a
 122 (predictor-predictor correlations) and Figure S3b (predictor-response correlations) underscores the
 123 separate analyses of interactions among environmental variables and their direct associations with
 124 gas fluxes.

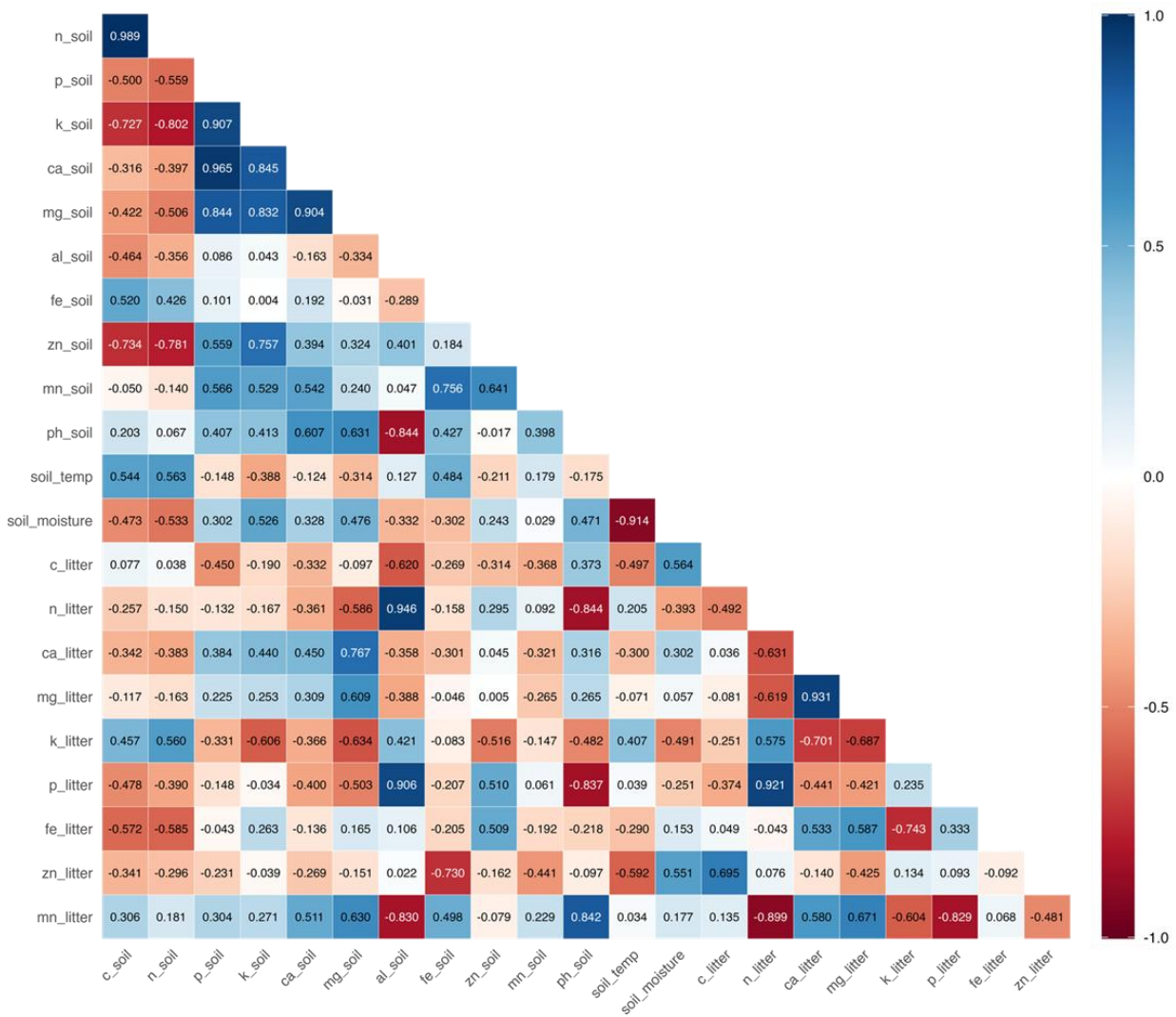


125

126 **Figure S3b.** Heatmap for the Pearson correlation coefficients for the correlation between gas
 127 fluxes and their potential predictors in the white sand forest. N = 12.

128 **S5.2 Ancient River Terrace Forest fluxes and their potential predictors**

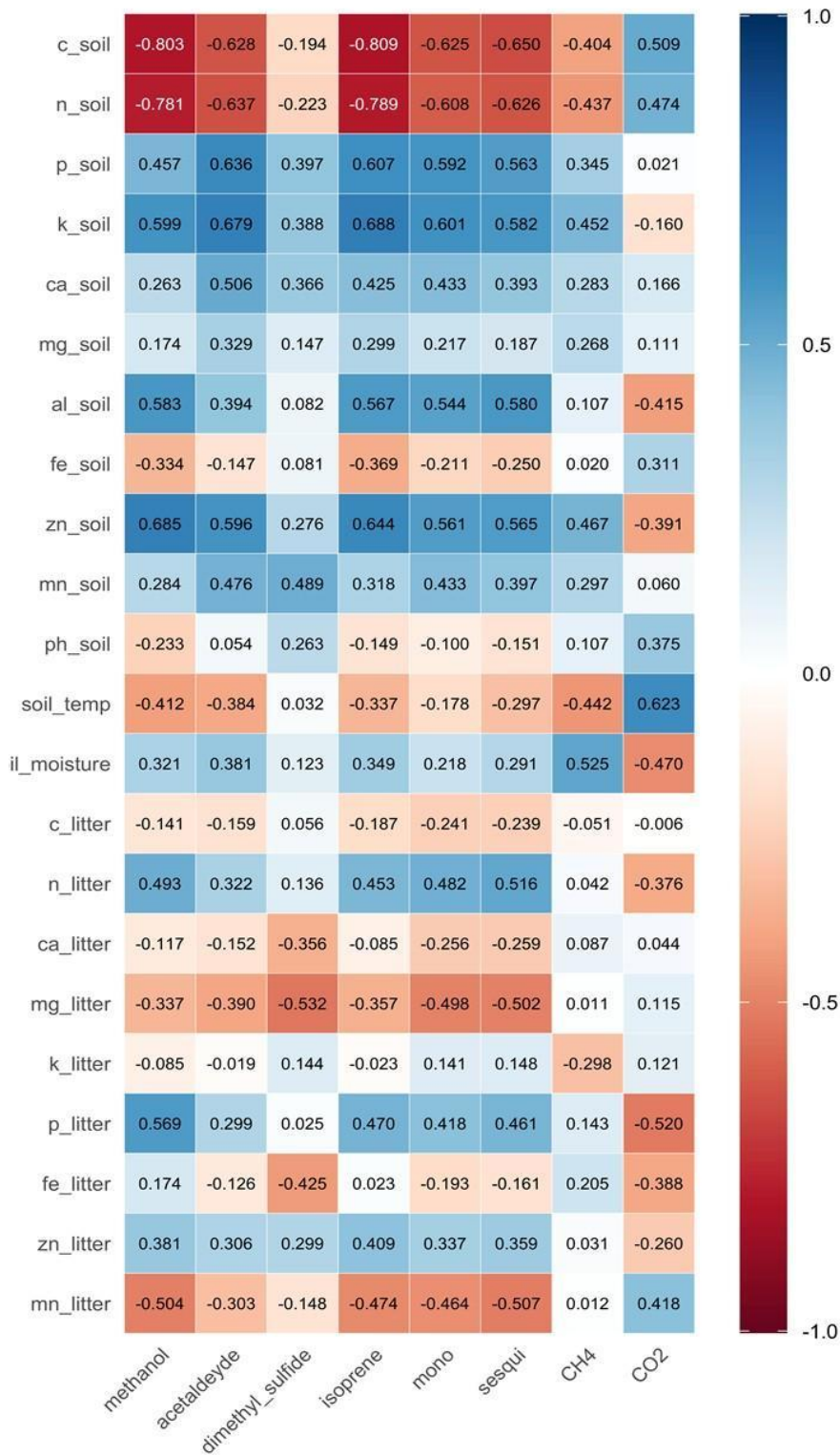
129 Initially, our investigation into the ancient river terrace forest ecosystem centered on delineating
 130 the intricate relationships among its key predictor variables. These variables encompass a range of
 131 soil and litter characteristics, microbial biomass metrics, and essential environmental parameters
 132 such as soil moisture and temperature. The outcomes of these Pearson correlation analyses are
 133 visually encapsulated in Figure S4a, which specifically illuminates the interdependence and
 134 patterns observed exclusively among these potential environmental drivers.



135
 136 **Figure S4a.** Heatmap for the Pearson correlation coefficients for the potential predictors of gas
 137 fluxes in the ancient river terrace forest. N = 12.

138 Following this, direct correlations were examined between these same predictor variables and the
139 measured response variables: BVOC and GHG fluxes. Figure S4b illustrates these Pearson
140 correlations, highlighting specific associations between environmental factors and observed
141 magnitudes of BVOC and GHG exchange. By presenting Figure S4a for predictor-predictor
142 correlations and Figure S4b for predictor-response correlations, the goal was to provide a clear,
143 two-level understanding of the underlying factors influencing gas fluxes in this type of forest.

144



145

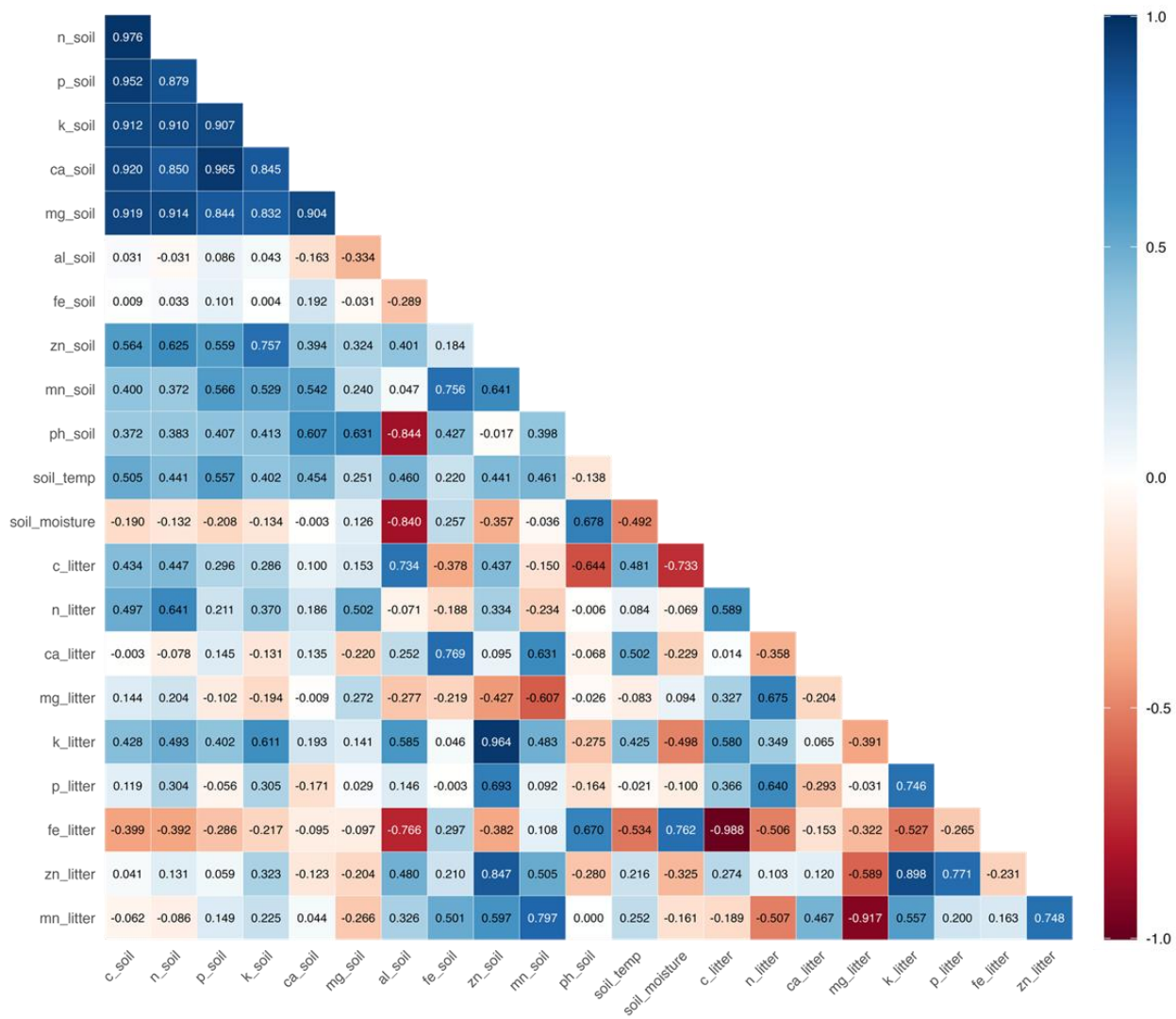
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147

Figure S4b. Heatmap for the Pearson correlation coefficients for the correlation between gas fluxes and their potential predictors in the ancient river terrace forest. N = 12.

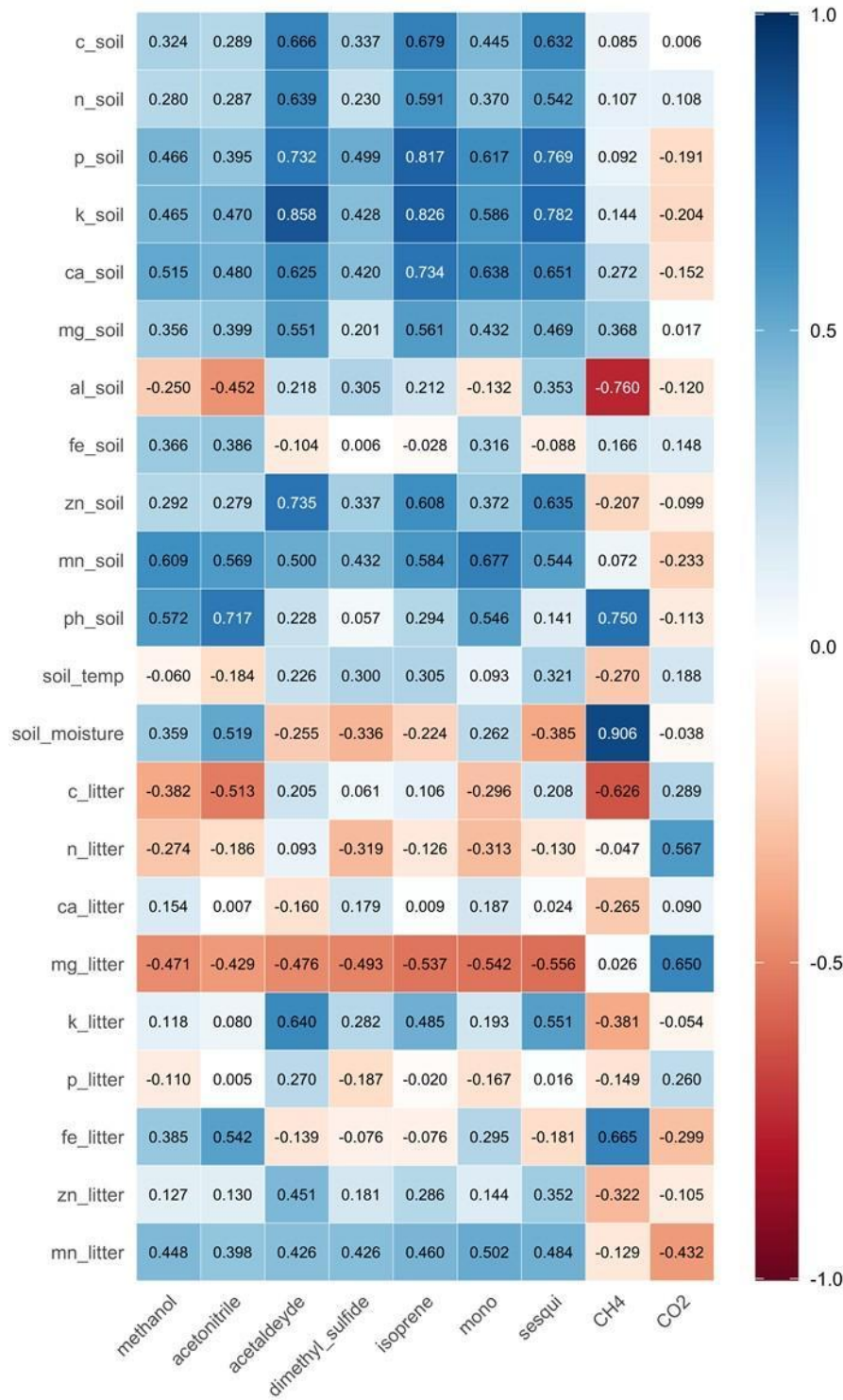
148 **S5.3 Upland forest fluxes and their potential predictors**

149 For the upland forest, we started by exploring the interrelationships existing between its various
 150 potential predictor variables. These predictors comprise elements such as soil and litter properties,
 151 indicators of microbial activity, and critical environmental conditions like soil moisture and
 152 temperature. The Pearson correlation coefficients that define these connections are systematically
 153 displayed in Figure S5a, providing a comprehensive overview of the intrinsic linkages among these
 154 environmental factors.



155
 156 **Figure S5a.** Heatmap for the Pearson correlation coefficients for the potential predictors of gas
 157 fluxes and their potential predictors in the upland forest. N = 12.

158 Subsequently, the focus shifted to quantifying the correlations between these identified predictor
159 variables and the study's response variables, which are BVOC and GHG fluxes. Figure S5b
160 graphically represents these Pearson correlations, offering specific insights into how variations in
161 environmental factors are associated with patterns of BVOC and GHG fluxes in the upland forest.
162 This dual visualization, with Figure S5a detailing the links between predictors and Figure S5b
163 focusing on the interactions between predictors and responses, allows for a nuanced interpretation
164 of the factors driving gas fluxes.



165

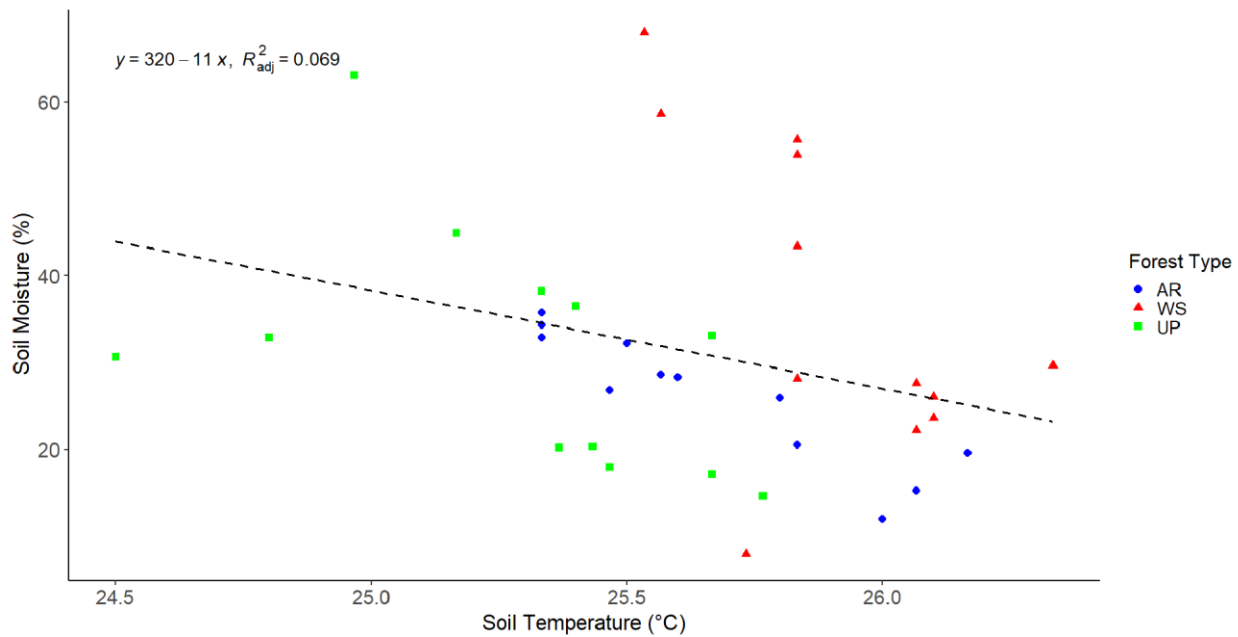
166 **Figure S5b.** Heatmap for the Pearson correlation coefficients for the correlation between gas
 167 fluxes and their potential predictors in the upland forest. N = 12.

168

169 **S6. Tangled Influences of Soil Moisture and Temperature on BVOC and CH₄ Fluxes**

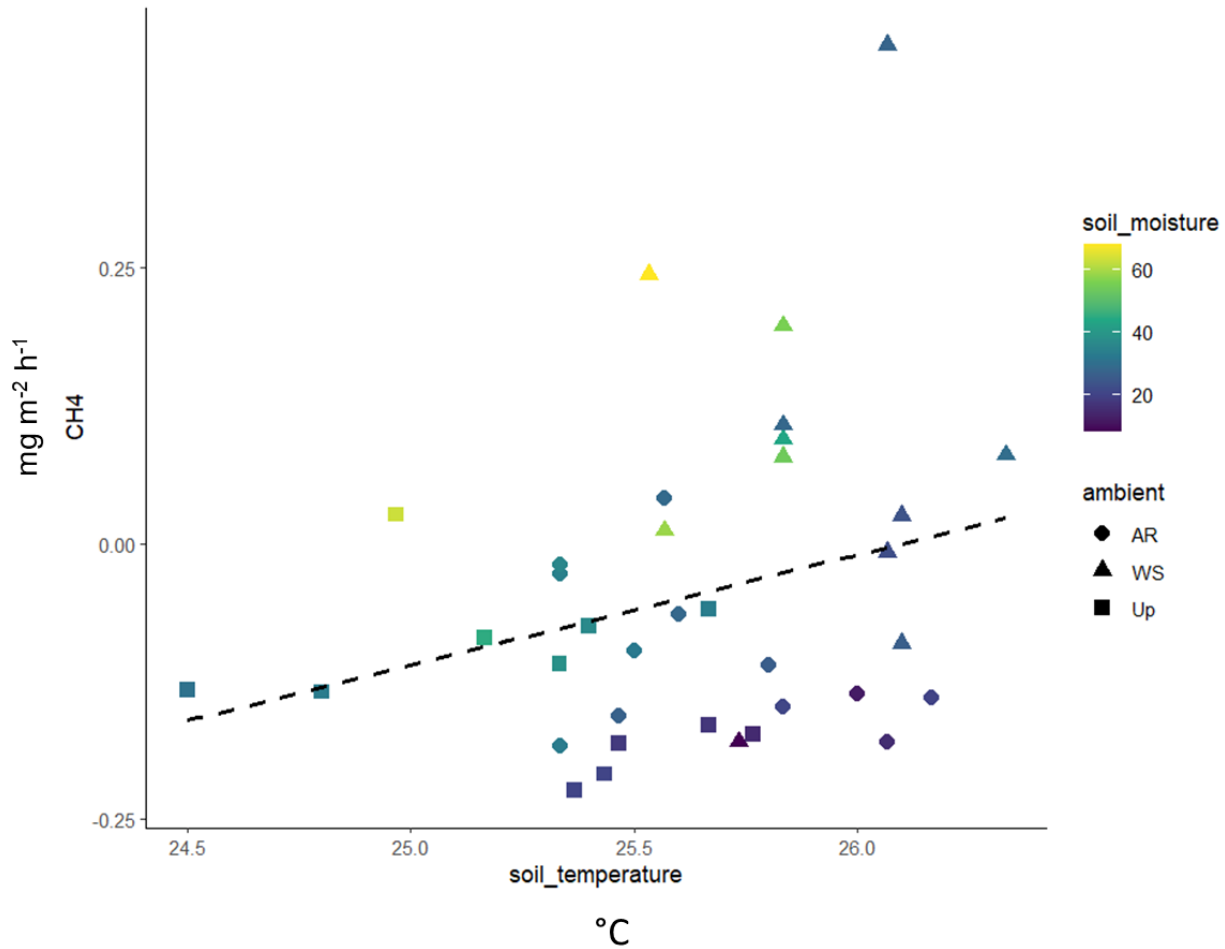
170 **S6.1 Relation between soil moisture and soil temperature (all forest types)**

171 Generally, drier soil conditions were associated with warmer soil temperatures (Figure S6), with
172 coefficients of determination (R²) of 0.82 (AR), 0.28 (WS), and 0.17 (UP). This interdependent
173 behavior complicates the interpretation of direct relationships between soil moisture, soil
174 temperature, and soil fluxes (Figure S7).



175
176 **Figure S6.** Relation between soil moisture and soil temperature in all forest types (AR, WS and
177 UP), average of five measurements taken in the topsoil. The dashed line represents the linear
178 regression fitted to the data, indicating the general trend of the relationship between soil moisture
179 and soil temperature. This regression model was built based on 36 observations.

180 **S6.2 Relation between soil temperature, soil moisture, and CH₄ fluxes (all forest types)**

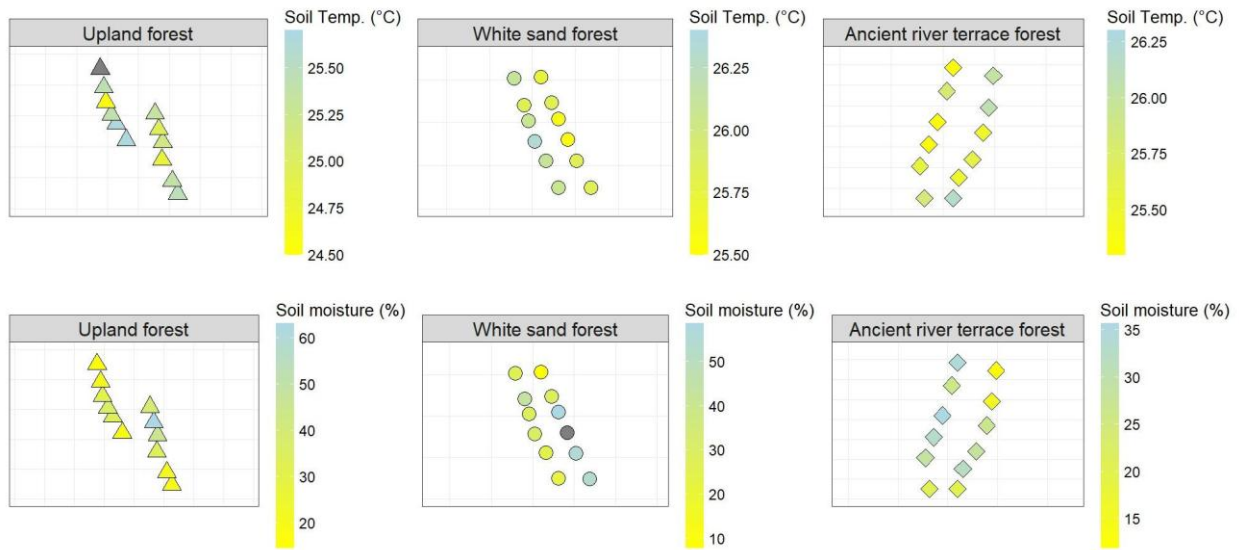


181
182 **Figure S7.** CH₄ flux from the soil and litter in the different forest types. The figure shows the
183 relationship between soil-litter CH₄ flux (mg m⁻² h⁻¹, y-axis), soil temperature (x-axis), and soil
184 moisture (color bar), where soil moisture and soil temperature values are an average of five
185 measurements taken in the topsoil. The dashed line represents the linear regression fitted to the
186 data, indicating the general trend of the relationship between soil temperature, moisture, and CH₄
187 fluxes. This regression model was built based on 36 observations.

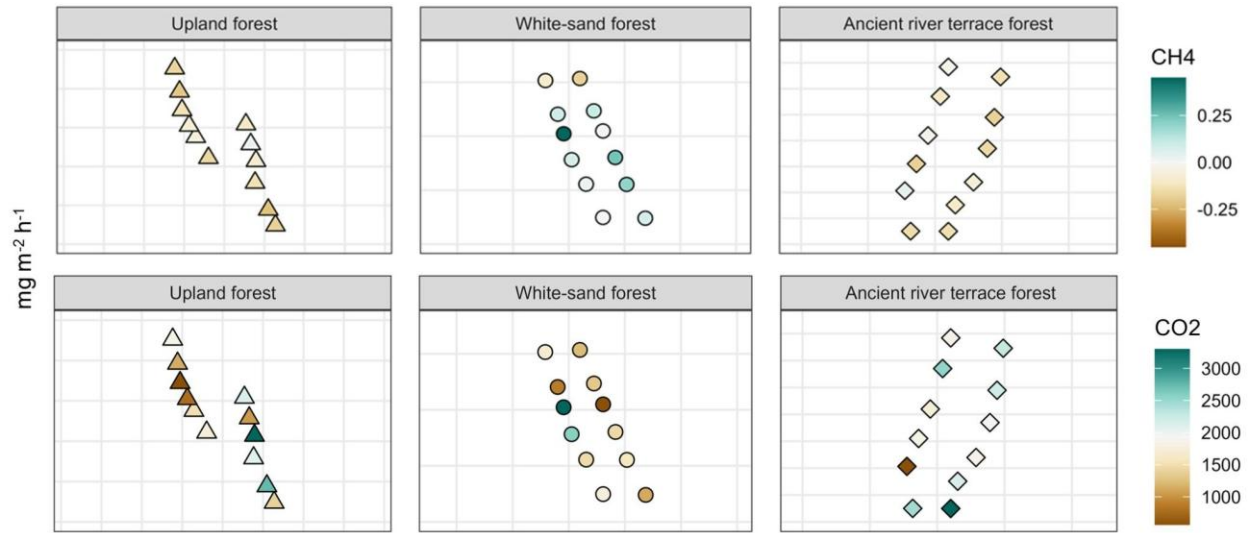
188 **S7. Spatial variability**

189 Although we attempted to select homogeneous and representative transects within each forest type,
190 we still observed differences in soil temperature and soil moisture between transects of the same
191 forest type, even when measurements were taken at the same time of day and only one day apart

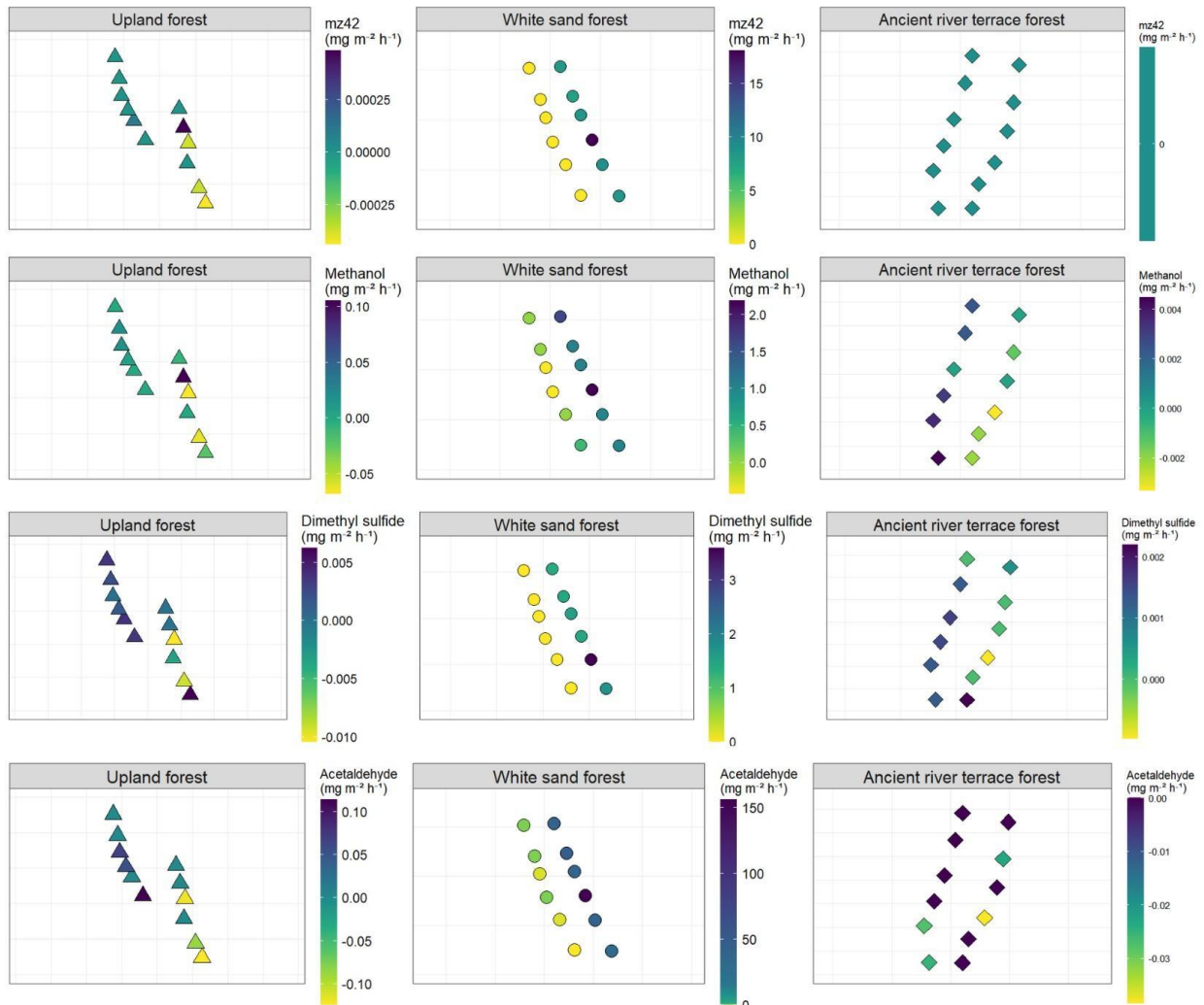
192 (Figure S8). For instance, in the upland forest, transect 1 was drier and warmer than transect 2. In
193 the white sand forest, transect 2 was colder and wetter than transect 1, likely partly due to a rain
194 event shortly before the measurements. CO₂ and CH₄ fluxes also exhibited considerable spatial
195 variability between and within transects (Figure S9). For BVOC fluxes, transects within the same
196 forest type even showed opposite flux directions, for example, with mostly uptake and mostly
197 emission in respectively transect 2 and transect 1 of the ancient river terrace forest (Figure S10).



198
199 **Figure S8.** Map showing the soil temperature (°C) (upper row) and soil moisture (%) (lower row)
200 for the two transects in each of the three forest types: Up, WS, and AR. Each transect contained
201 six sampling points, totaling 12 measurement points per forest type. Left - Transect 1; Right –
202 Transect 2.



203
 204 **Figure S9.** Map of the points collected showing the flux of GHG - methane (CH₄) (upper row) and
 205 CO₂ (lower row) in each of the three forest types: Up, WS, and AR. Each transect contained six
 206 sampling points, totaling 12 measurement points per forest type. Left - Transect 1; Right - Transect
 207 2. The gas flux data are expressed in mg m⁻² h⁻¹.



208
 209 **Figure S10.** Map of the points collected showing the BVOC fluxes (mg m⁻² h⁻¹) in each of the
 210 three forest types: Up, WS, and AR. Each transect contained six sampling points, totaling 12
 211 measurement points per forest type. Left - Transect 1; Right - Transect 2.
 212