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The role of mixotrophic protists in the biological carbon pump

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The oceans support ~46% of Earth's primary production, mainly through the phototrophic activities of protists and prokaryotic microbes that are at the base of the marine foodweb (Field et al., 1998). Classic texts relate physics and inorganic nutrients to the phototrophic producers (the phytoplankton), then to their consumers (as microand meso-zooplankton), and ultimately to fisheries (Cushing, 1975, 1995); within this framework the role of the microbial loop (Azam et al., 1983; Gifford, 1991; Sherr and Sherr, 2002) is typically embedded. The collective activity of marine microbes (i.e., bacteria, cyanobacteria, protists), together with their immediate trophic associates (zooplankton such as copepods and salps) drive the biological pump through production of particulate organic material (POM) that sinks from the upper mixed waters. While POM has a clear role within the biological pump, the role of long-lived forms of dissolved organic material (DOM) as a reservoir of C remains controversial (Azam and Worden, 2004; Hansell et al., 2009; Jiao et al., 2010).

Most of the energy and material in the marine foodweb flows through protists, organisms that are traditionally seen as (eukaryotic) phytoplankton and microzooplankton. In the evolution of these protists, phagotrophy is the ancestral state and phototrophy is the derived, more recent, state (Raven et al., 2009), with all protists engaging in osmotrophy (uptake of dissolved organic substrates, vitamins and others) to varying extents (Glibert and Legrand, 2006; Burkholder et al., 2008). This evolutionary pathway, which is neither fixed nor irreversible, has included a multitude of events associated with acquisition of structures and of symbionts that have led to the evolution of organelles including plastids and other characteristics (de Castro et al., 2009; Stoecker et al., 2009). "Strict" phototrophy and "strict" phagotrophy thus form the two extreme ends of a spectrum with most protist groups functionally occupying the intermediate niche zone as mixotrophs. In fact, evolution to the extent of rejecting phagotrophy completely appears to be restricted to a few, albeit important, groups of protists, most notably the diatoms.

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Here, we define mixotrophy in protists as the dual capability of engaging phototrophy and phagotrophy within a single cell. The varying proportions of phototrophic and phagotrophic activities in these mixotrophic protists (both for the individual cell, and among species and strains) depend on the availability of light, nutrient and/or prey or other particles upon which to feed. Mixotrophy is not displayed by any unique taxonomic group but, rather, occurs amongst different species ranging over a variety of groups (Stoecker et al., 2009; Flynn et al., 2013). It is a common phenomenon occurring widely in marine (e.g., eutrophic, mesotrophic and oligotrophic, coastal to open-ocean systems; Pitta and Giannakourou, 2000; Burkholder et al., 2008; Hartmann et al., 2012; Sanders and Gast, 2012) as well as freshwater systems (Sanders, 1991a). Indeed, there is increasing evidence that in most aquatic systems the majority of protists are mixotrophs engaging in varying proportions of phototrophy and phagotrophy (Sanders, 1991b; Burkholder et al., 2008; Raven et al., 2009; Stoecker et al., 2009; Jeong et al., 2010; Lindehoff et al., 2010; Hansen, 2011; Johnson, 2011). Often the marine ecosystems are heavily dependent upon the activity of these mixotrophic protists (Pitta and Giannakourou, 2000; Unrein et al., 2007; Zubkov and Tarran, 2008; Hartmann et al., 2012; Sanders and Gast, 2012).

In light of the above, Flynn et al. (2013) proposed a marked shift in the way that aquatic protists are popularly characterised and sub-divided. Instead of the traditional "black-and-white" view that characterizes typical marine microbial protists as being either phototrophic "phytoplankton" *or* phagotrophic "microzooplankton", they argued that a significant proportion of the protists in the photic zone are mixotrophic, and that aquatic scientists need to acknowledge this reality in empirical and theoretical studies. Stemming from this reappraisal, the revised marine foodweb paradigm conceptualizes the traditional phytoplankton–zooplankton structure as predominating only during short, though important, periods of the production cycle. These periods are characterised by pioneer species (akin to *r* selected species; Parry, 1981) growing in immature ecosystems exemplified by the spring bloom in temperate waters, or in upwelling waters (Fig. 1). They are typically associated with enhanced illumination and inorganic nutrient

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availability coupled with an absence of effective grazing control (Irigoien et al., 2005), favouring rapid proliferation of "strict" phototrophs (e.g., diatoms, and non-phagotrophic stages of the coccolithophorids; Rokitta et al., 2011). Such conditions lead to emergence of the "strict" phagotrophs as the dominant protist predators. However, as these waters and their ecosystems mature, their nutrient conditions and particulate organic loading change, giving rise to conditions that favour mixotrophy (Fig. 1). Therefore, during much of the planktonic production cycle these mixotrophs are abundant or dominant; they support important events such as early-stage life cycles of many finfish and shellfish, and development of the autumn and overwintering zooplankton populations which then impact upon the biogeochemistry of the following spring bloom (Cushing, 1995; Montagnes et al., 2010).

Most plankton ecologists and those working in allied fields of research (e.g., biological oceanographers, modellers) emphasize, mostly or entirely, the organisms representing the two extremes in the evolution of protists: the strictly heterotrophic microzooplankton and the strictly phototrophic phytoplankton. Mixotrophy is also ignored or marginalised in theoretical as well as modelling studies of aquatic ecosystems (Table 2 in Mitra et al., 2013). Thus, the nutrient-phytoplankton-zooplankton-bacteria (NPZB) foodweb models, which have been a feature of marine research for decades (Cushing, 1975, 1995; Fasham et al., 1990; Totterdell et al., 1993; Rose et al., 2010), do not include mixotroph functional types at all. Even recent reviews focussing on the importance of microbial activity in global productivity and functioning of the biological pump make scant or no mention of the existence of mixotrophs (Azam and Worden, 2004; Jiao et al., 2010; Steele et al., 2010).

Here, we explore the ramifications of the proposed new paradigm for trophic dynamics within the plankton foodweb, demonstrating the need to consider the activities of the mixotrophic protists. We also show how exclusion of these organisms from conceptual and mathematical models misrepresent the functioning of the biological pump with important implications for our understanding of the influence of climate change on aquatic ecosystems. For this work, we have considered an oligotrophic plankton food**BGD**

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web framework because, while providing a relatively simple structure, the oligotrophic gyres are recognized to be the largest oceanic ecosystems (ca. 40% by area) and are spatially expanding at substantial annual and seasonal rates (Polovina et al., 2008; Hartmann et al., 2012).

5 2 Methods

2.1 Foodweb framework

In order to explore the role of mixotrophic protists, we compare the outputs from two contrasting in silico plankton foodweb structures operating in an oligotrophic setting, as detailed below.

- (i) A traditional foodweb structure without an explicit description of mixotrophs (Fig. 2a), henceforth termed the "classic paradigm". This framework includes non-motile microalgae (NMA) and photo-autotrophic nanoflagellates (ANF) as primary producers (both being non-mixotrophic), bacteria as decomposers, heterotrophic nanoflagellates (HNFs) that ingest the bacteria, and microzooplankton (μ Z) which graze on the NMA, ANFs and HNFs. The μ Z are in turn consumed by higher trophic levels (simulated here through a closure function; Mitra, 2009).
- (ii) An alternative foodweb framework incorporating mixotrophs (Fig. 2b), henceforth termed the "revised paradigm". This revised paradigm includes the same components as the classic paradigm, excepting one difference. The photo-autotrophic nanoflagellates (ANF) are now replaced with nanoflagellates that engage in phagotrophy; that is, in keeping with our revised understanding, they are mixotrophic. These mixotrophic nanoflagellates (MNF) in the revised paradigm are de facto representatives of the mixotrophic flagellates observed within oligotrophic systems (Unrein et al., 2007; Zubkov and Tarran, 2008; Stukel et al., 2011; Hartmann et al., 2012). They photosynthesize using their constitutive chloroplasts and attain additional nutrition through the ingestion of bacteria, thence competing with the HNFs for bacterial prey. While

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mixotrophic algal protists have been shown to feed on a wide array of types and sizes of prey (Jeong et al., 2010), in the system that we consider, bacteria are the primary prey item (Hartmann et al., 2012).

Configuration of the model

Five plankton functional types (bacteria, autotrophic non-motile microalgae – NMA, autotrophic nanoflagellates - ANF, mixotrophic nanoflagellates - MNF, heterotrophic nanoflagellates – HNF, and microzooplankton – μZ) were constructed using mechanistic adaptive plankton models, which have all been previously described (see below for references). The models were C-biomass based, with explicit inclusion of N and P. Variable C: N: P stoichiometry was simulated in those capable of phototrophy (NMA, ANF and MNF) and in the DOM. The bacteria, HNFs and µZ were assigned a fixed C: N: P stoichiometry of 50: 10: 1 (Mitra, 2006; Flynn and Mitra, 2009; Mitra and Flynn, 2010). Growth of those capable of phototrophy (NMA, ANF and MNF) contained components linking growth rates to their variable C: N: P physiology and to light through photoacclimation (variable Chl: C, Flynn, 2001). Osmotrophy (i.e., nutrition using dissolved organics) was not factored into the models of any of the protists (NMA, ANF, MNF, HNF, μZ), although it is common among many aquatic protists (e.g., Glibert and Legrand, 2005; Burkholder et al., 2008).

The microalgal assemblage (NMA) consuming inorganic substrates as nitrate, ammonium and phosphate, and releasing dissolved organic carbon (DOC) was configured using the model of Flynn (2001). The bacterial assemblage consuming inorganic forms of N and P, and labile and semi-labile dissolved organics was as described by Flynn (2005). Labile DOM was considered to have been generated directly as a function of C-fixation (i.e., low-molecular-weight primary metabolites; Hansell et al., 2009), while all other forms of DOM were semi-labile (sDOM); labile forms were used by preference and allocated as described in Flynn (2005). The HNFs and µZ were described using the zooplankton model of Mitra (2006).

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The ANF and MNF models were described using the model of Flynn and Mitra (2009), configured to consume nitrate, ammonium, phosphate and to release DOC and semi-labile organics. For the classic paradigm, phagotrophy in this submodel was not enabled; in this form the model described ANFs. With phagotrophy upon bacteria enabled, allowing mixotrophy, this submodel described MNFs.

The HNFs and MNFs had only one prey option (bacteria; cell diameter, $0.45\,\mu m$, Andersson et al., 1986). However, the μZ could graze on NMA, ANFs, MNFs and HNFs. Prey selection by the μZ was a function of prey availability as related to prey numeric abundance and prey encounter, assuming cell diameters as follows: NMA, ANF/MNF, and HNFs, $3\,\mu m$, $3\,\mu m$, and $2.9\,\mu m$, respectively (Hartmann et al., 2012), and for μZ of $20\,\mu m$ (Pérez et al., 1997). Prey selectivity was implemented through a modified version of the ingestion-based selectivity function of Mitra and Flynn (2006a).

The interactions between the different plankton communities were modelled within a physical description of a mixed water column of 150 m depth. A low level of mixing (equivalent to a dilution rate of $0.01\,d^{-1}$; see Fasham et al., 1990) between the mixed and lower water masses, removed organisms and residual nutrients and introduced fresh nutrients from the sub-mixed layer waters. The initial (and sub-mixed layer) dissolved inorganic nitrogen (DIN, as nitrate and ammonium at a ratio of 9 : 1) was assumed to be 1 μ M. Phosphate was supplied at three different ratios relative to DIN: – Redfield ratio (molar N : P = 16, Redfield, 1958), elevated (molar N : P = 64, imparting P-stress) or depressed (molar N : P = 4, for N-stress). To provide a source of organics to support bacterial growth, the initial (and sub-mixed layer) semi-labile dissolved organic nitrogen (sDOM-N) was assumed to be 0.5 μ M (i.e., 50 % of initial DIN). The initial (and sub-mixed layer) semi-labile dissolved organics in the form of carbon (sDOM-C) and phosphorus (sDOM-P) were calculated from sDOM-N using the Redfield ratio.

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All simulations of the classic paradigm demonstrated a significant period of oscillation before entering steady state (Fig. 3; similar results were found using different N:P nutrient ratios, not shown). On entering steady state, the autotrophic nanoflagellates (ANFs) were slowly eliminated due to their slightly slower emergent growth rate (lower by ca. 3%) compared to that of the non-motile microalgae (NMA).

In the simulations of the revised paradigm, the predator-prey oscillations were muted in comparison with those seen in the classic foodweb (Fig. 3b cf. Fig. 3a), and the system entered steady-state much more rapidly. The NMA were eliminated as a consequence of their inability to compete with the mixotrophic nanoflagellates (MNFs). In addition, the heterotrophic nanoflagellates (HNFs) were eliminated because they could not compete with the MNFs for bacterial prey. The advantage of the MNFs over both the NMA (for phototrophy) and the HNFs (for bacterivory) was attained through a synergistic co-operation between phototrophy *and* phagotrophy within the same organism. The results shown in Fig. 3 are consistent with those of Hartmann et al. (2012), in that the dominance of bacterivory by HNFs expected under the classic paradigm is replaced by a dominance of MNFs.

A comparison between steady-state standing stock C-biomass levels for the different simulations is shown in Fig. 4. For Redfield N : P and low N : P nutrient configurations, the total plankton biomass was either slightly higher, or essentially comparable in the classic vs. revised paradigms. However, the revised paradigm, with its lower bacterial biomass and higher biomass of larger celled μZ , contained a higher biomass contribution by larger organisms (average ESD, 9 μ m for revised paradigm simulations vs. 7 μ m for classic paradigm). In contrast, in the P-limited (high N : P) scenario, where again the MNFs dominated, the total plankton biomass and μZ biomass was lower for revised vs. classic paradigm. Bacterial biomass was also much lower in the revised paradigm because of the higher abundance of their grazers (MNFs vs. HNFs). The explanation for the decreased μZ in the revised paradigm, despite the similar C-biomass of their col-

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lective prey between paradigms, is the lower food quality; the HNFs, with their more closely matched lower C:N and C:P, were scarce when MNFs were present. Other simulations with different nutrient concentrations and/or different light showed similar trends.

Comparisons of the primary, bacterial and DOC production in the different scenarios showed that, in every instance, primary production (C-fix, Fig. 5) was enhanced in the revised paradigm, wherein C-fixation is partly supported directly (through phagotrophy) by nutrients originally acquired by bacteria (Fig. 5). Production of DOC originating directly from primary production was similarly enhanced, but total DOC production (which includes DOC produced as sDOM-C through grazing activity) was lower in the high N:P (P-limited) scenario of the revised paradigm (Fig. 5), because of the lower μZ activity (Fig. 4). In all instances, planktonic net DOC production was always higher in the presence of mixotrophy (revised paradigm); although under high N:P, net DOC production was positive only with mixotrophy (Fig. 5c). In all the other scenarios, bacterial production was augmented through the use of dissolved organics mixed up into the upper layer. Although bacterial production in the revised paradigm under high N:P was lower than in the classic paradigm, it was higher than one may expect from the standing stock (Fig. 4); this is because the lower bacterial biomass was more active (higher growth rate) in the revised paradigm containing the mixotrophs.

4 Discussion

4.1 Interpreting the simulations

The results from the simulations recreate the pattern of a typical plankton community observed in oligotrophic waters (Zubkov and Tarran, 2008; Hartmann et al., 2012) of a plankton community comprised primarily of bacteria and mixotrophic flagellates. The heterotrophic nanoflagellates (HNFs) and non-motile microalgae (NMA), which separately were better competitors for bacteria or nutrients, respectively, were ultimately

excluded by the mixotrophic nanoflagellates (MNFs). The expected boom-and-bust predator-prey dynamics of the classic paradigm did not occur in the revised paradigm (Fig. 3a vs. Fig. 3b), indicating the stabilising nature of mixotrophic nutrition. Algivory, rather than bacterivory, as the heterotrophic component in mixotrophs has been previously noted to enhance foodweb stability (Jost et al., 2004). In a maturing ecosystem such stability is important because instability would result in nutrients being lost from the system due to predator-prey mismatch (Flynn, 1989). The net result is the appearance of a stable plankton community, which displays different trophic dynamics than in the classic paradigm.

A problem in modelling mixotrophy is that it is all too easy to configure an organism that is completely dominating – the "perfect beast" (Flynn and Mitra, 2009). In the past such an in silico outcome was considered to be at odds with reality. Now, however, it appears that mixotrophic protists are indeed nearly ubiquitous in mature ecosystems (Sanders, 1991a; Stoecker et al., 2009; Sanders and Gast, 2012; Flynn et al., 2013). This reality has not been widely appreciated and embraced because traditional field and laboratory studies typically concentrate on strict phototrophs and strict phagotrophs; in addition, experimental protocols are typically suboptimal for elucidating the mixotrophic activities. Development of a sensitive, radiotracer-based experimental protocol (Zubkov and Tarran, 2008) enabled quantification of the major contribution of the phytoflagellates (formally considered to be "strict" phototrophs) to bacterivory in both mesotrophic and oligotrophic oceanic ecosystems (Zubkov and Tarran, 2008; Hartmann et al., 2012).

It appears that bacterivory, as well as predation upon larger organisms, is a potential nutritional acquisition route for many flagellates (Unrein et al., 2007; Burkholder et al., 2008; Jeong et al., 2010). Importantly though, this form of mixotrophy differs from the traditional expectation of two contributing nutritional pathways, both phototrophy and phagotrophy, contributing C, N, and P. Rather, mixotrophy in these protists appears synergistic and cooperative in its action: phagotrophy provides primarily N and P (and other non-carbon elements), whereas C acquisition happens mainly through photoau-

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- (i) Measurements using C as the tracer probably will not indicate the true importance of mixotrophy. In consequence, field measurements of mixotrophy need to deploy mixed tracers (i.e., not solely C-tracers) and use a variety of approaches (Zubkov and Tarran, 2008; Calbet et al., 2012). Furthermore, significant changes in photosynthesis may or may not be expected (depending on the fate of the prey-C).
 - (ii) Phago-mixotrophy by protists with a constitutive ability to photosynthesize (as simulated here) probably provides nutrients that are ultimately handled by the cells' photo-dominated physiology in a similar fashion to inorganic nutrients. In consequence, there is scope for enhanced DOM-C release as a function of phagotrophy, both through non-assimilation of some portion of the prey-C, and through DOM-production as some portion of the continuing C-fixation process (Flynn et al., 2008).
 - (iii) Models of mixotrophy and of its role in the biological pump need to be multi-element based (C, N, P, etc.), else they cannot capture the synergistic/cooperative nature of the interactions. In addition, variability in stoichiometry is an important feature in simulating the dynamics of predation (Grover, 2003; Mitra and Flynn, 2005; Glibert et al., 2011). In the simulations presented here, the MNFs did not win by eating their competitors (eating the HNFs or non-phagotrophic microalgae; cf. Thingstad et al., 1996). Instead, they succeeded through a combination of co-operative nutrition and, especially in the low-P simulations, through having a disadvantageous stoichiometric content (i.e., poor quality food for μ Z). Even though the nutritional routes (phototrophy vs. phagotrophy) are not fully substitutable, an additional factor in favour of mixotrophy is that (similar to that mentioned in Tittel et al., 2003), the combination of nutritional routes enables the mixotrophs to decrease resources to levels below the critical thresholds required to support effective growth by either of the non-mixotrophic NMA and HNFs.

Oligotrophic systems are low in phosphorus (high N:P), although as nutrient concentrations become vanishingly low, the critical N:P ratio for equal N and P sufficiency

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decreases (Flynn, 2010). It is interesting to note that the modelled high N: P scenario behaved differently than the others and was the only scenario that resulted in net production of DOM (Fig. 5). Taking everything into account, the events seen for the high N: P scenario arise because: (1) the mixotrophs release DOC that supports bacterial acquisition of the nutrients, (2) the mixotrophs then acquire those nutrients directly through consumption of the bacteria, (3) this acquisition is insufficient to match the total nutritional demand by the mixotrophs (due to P deficiency in the environment), and therefore (4), the C: N: P of the mixotrophs is/remains disadvantageous (of poor quality) to the µZ predators. In essence, the dominance of the mixotrophs in the P-limited high N: P system developed through a process not dissimilar to that proposed by Mitra and Flynn (2006b) for the formation of an ecosystem disruptive harmful algal bloom (cf. EDAB; Sunda et al., 2006). This development of a bloom could be considered to represent a product of a dysfunctional microbial loop (Thingstad et al., 1997), in that the transmission of C and energy up through μZ to higher trophic levels is restricted. Either way, a system is created that generates DOM via the "microbial carbon pump" that some (Azam and Worden, 2004; Hansell et al., 2009; Jiao et al., 2010) have suggested to represent an additional contributor to the biological pump.

The revised paradigm for marine plankton foodweb and implications for the biological pump

An appreciation of the existence of protist mixotrophs is not new. What is new is the realisation that these organisms are major players in the planktonic foodweb, contributing substantially to the flow of carbon and other nutrients in aquatic ecosystems. It is thus appropriate to ask whether this realisation warrants a revision of our understanding and simulation of foodweb dynamics and allied biogeochemistry coupled to the biological pump.

Figure 6 presents, in simplified form, the bacteria-centric parts of the foodweb (Fig. 2), as we have explored through simulations. In the classic paradigm (Fig. 6a), DOM release from phototrophy supports growth of bacteria, which then enables them,

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on an individual cell basis (due to their smaller size), to outcompete phototrophs (microalgae, Fig. 6) for nutrients. Grazing on bacteria by microzooplankton (specifically HNFs, and then by larger μZ) provides the main route for nutrient regeneration in the classic paradigm (yellow arrow in Fig. 6a); this is due to stoichiometric constraints linked to respiration and an inability of assimilation efficiencies to approach 100 % (Mitra and Flynn, 2005). In the revised paradigm (Fig. 6b), DOM release from phototrophy again supports growth of bacteria, enabling them to acquire inorganic nutrients unavailable to the phototrophic protists. However, grazing on bacteria by these protists now acts through mixotrophy as a direct conduit for the support of primary production (yellow arrow in Fig. 6b). This could be considered akin to a symbiotic or mutualistic relationship between phototrophy and heterotrophy (as discussed in the context of primary production by Flynn, 1988), or a relationship wherein the bacteria are being farmed by the mixotrophs. The consequence is that primary production can now be supported by nutrients that would otherwise be unavailable (present at very low concentrations, and/or in combined forms of DOM that are not chemically suitable for transport into the protist cell).

Both the bacteria (with extracellular digestion) and the mixotrophs (through voiding of un-required complex-compounds of bacterial origin) will generate DOM that will over time become increasingly refractory and hence accumulate. Indeed, in the simulations, especially in the high N:P (P-limiting) scenario, the net DOM-C increased (Fig. 5c). The extent to which DOM-C contributes to a biological pump (Hansell et al., 2009), depends on the removal of this material into deeper water, away from abiotic processes (notably sunlight and oxygen) that would promote its degradation. Presumably any contribution of DOM-C to the biological pump is (similar to the abiotic removal of CO₂ into cold waters) only of transient importance in mitigating the increasing atmospheric CO₂ (otherwise, concentrations of DOM must be increasing over the millennia). Nonetheless, it could be important and merits consideration in long-term forecast models.

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Climate and anthropogenic changes to the marine ecosystem include an increase in water column stability (Doney et al., 2009, but cf. Lozier et al., 2011), and changes if not increases to coastal eutrophication (Burkholder et al., 2008; Burkholder and Glibert, 2013). Both of these events are likely to favour the growth of mixotrophic plankton, including potentially HAB species (noting that HABs are dominated by mixotrophic forms; Burkholder et al., 2008). Along with this we may expect to see changes in plankton trophic dynamics, and in the functioning of the biological pump.

Temporal and spatial events that see mixotrophs as important members of the plankton include survival and growth of larval fish in temperate waters (de Figueiredo et al., 2007; Montagnes et al., 2010), and production in the oligotrophic systems that cover most of the oceans (Pitta and Giannakourou, 2000; Unrein et al., 2007; Zubkov and Tarran, 2008; Stukel et al., 2011; Hartmann et al., 2012). The interplay between nutrient stoichiometry, ecosystem maturity and success of mixotrophy also helps to explain why eutrophication is often associated with mixotrophic HABs and EDABs (Burkholder et al., 2008; Glibert and Burkholder, 2011; Burkholder and Glibert, 2013). It also explains why many offshore mixotroph-dominated bloom events are difficult to relate to land-based nutrient sources (Anderson et al., 2011; Harrison et al., 2011). Offshore blooms of mixotrophic species may, in fact, be the successional endpoint of near-shore nutrient loading and nutrient stoichiometric shifts. For example, some offshore blooms of the green mixotrophic *Noctiluca scintillans* are now recognized to be the displaced result of near-shore eutrophication (Harrison et al., 2011). The future would likely see an expansion of such events unless steps are taken to control eutrophication.

Acknowledging the importance of mixotrophy is not simply a case of recognising mixotrophs as significant; it is of recognising the importance of mixotrophy in organisms that we already know to be significant. While many aspects of mixotroph physiology remain enigmatic, they offer fertile ground for investigation by aquatic ecologists and thus improve understanding of aquatic foodwebs and elemental cycles in theoreti-

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cal and computational ecological modelling. The work reported in this paper, together with that of Mitra and Flynn (2010), have demonstrated the importance of providing an explicit multi-nutrient description of mixotrophy in models if we are to simulate these events adequately. Only through such efforts will we better understand the future contribution of mixotrophy to the biological pump as well as to food security issues (as feed for fisheries, Montagnes et al., 2010 vs. HABs, Burkholder et al., 2008) and interactions with climate change in shaping marine plankton assemblages.

Acknowledgements. Construction, validation and application of models were funded through grants awarded by the Natural Environment Research Council (UK) to AM and KJF. All authors acknowledge the Leverhulme Trust for their funding of an International Network Grant. AC was partially supported by the Ministry of Science and Innovation of Spain through project PROTOS (CTM2009-08783). DKS was partially supported by NSF award 1031344. J. Hawkey assisted with graphic preparation. AM thanks Rohan Mitra-Flynn for his continuing support. This is contribution number xxxx from UMCES. The University of Dundee is a registered charity, No SC010596.

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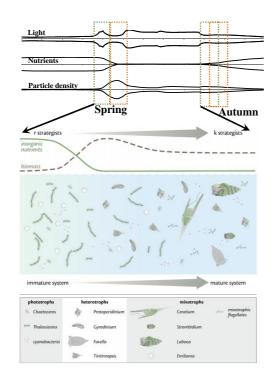


Fig. 1. Diagrammatic portrayal of the changes to the planktonic foodweb over a year, with transitions between immature and mature ecosystem states. The upper panels show changing patterns of light, inorganic nutrients and particle density (i.e., total plankton biomass) over the temperate year. Transitions from mature to immature (spring or autumn "blooms") to mature again, are as indicated; green dashed line indicating conditions optimal for phototrophy, orange dashed lines for phagotrophy. Other periods are sub-optimal for strict phototrophs and/or strict phagotrophs, and preferable for mixotrophs. The lower panel shows in detail the transition from immature to mature, with changes in selection priorities from so-called "r select" phototrophs and phagotrophs in immature ecosystems, to a mature ecosystem more optimal for "K select" mixotrophs. (See Parry, 1981 for discussion concerning r vs. K selection.)

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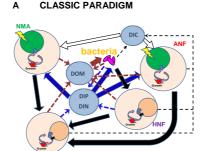
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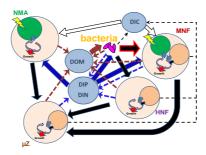


Fig. 2. Schematic representing the two alternate modelled foodweb structures. In the classic paradigm **(A)** the physiology of the photoautotrophic flagellates (ANF) is similar to that of the non motile microalgae (NMA); neither are mixotrophic. Inorganic nutrients for the support of primary production are regenerated via bacteria, heterotrophic nanoflagellates (HNF) and phagotrophic microzooplankton (μ Z). In the revised paradigm **(B)**, the ANF are replaced with mixotrophic nanoflagellates (MNF), conforming to our revised understanding of protist physiology. The MNFs are capable of eating bacteria (red-lined black arrow), and hence deriving nutrients for the support of their growth, and of their primary production, directly rather than (for ANF in panel **A**) via the activities of HNF + μ Z. Dashed arrows indicate functions contributing to nutrient pools (blue for inorganic, brown for organic). Heavy black arrows indicate predatory links.

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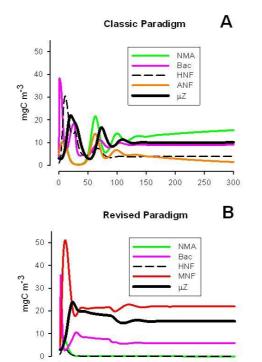


Fig. 3. Temporal pattern of the development of biomass in the simulated communities. In the classic paradigm (A), the autotrophic nanoflagellates (ANFs) have no mixotrophic potential, while in the revised paradigm (B), these nanoflagellates (as MNF) exhibit mixotrophy (see also Fig. 2). The inorganic nutrient regimes used for this simulation (with an inorganic N input of 1 μM) was Redfield N: P (molar ratio 16).

Time (d)

100

200

250

300

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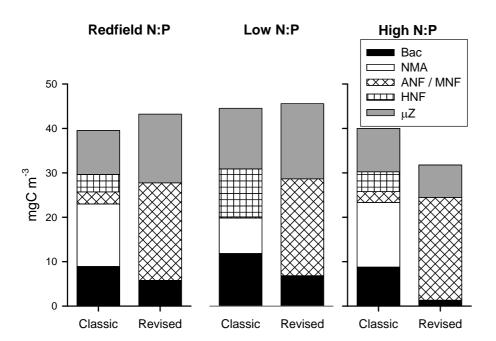


Fig. 4. Model output at steady-state, showing the standing stock contributions to biomass of bacteria (Bac), non-motile microalgae (NMA), heterotrophic nanoflagellates (HNF), and the microzooplankton (μZ). In the classic paradigm, the flagellates capable of phototrophy (ANF) have no mixotrophic potential, while in the revised paradigm, they (MNF) exhibit mixotrophy (see also Figs. 2 and 3). The inorganic nutrient regimes (all with 1 μM inorganic N) are in Redfield N: P (molar ratio 16), low N: P (molar ratio 4), or high N: P (molar ratio 64).

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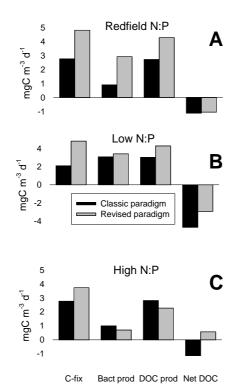


Fig. 5. Model output showing rates of primary production (C-fix), bacterial production (Bact prod), production of DOC (from all sources, including voiding of material by grazers and primary production leakage, DOC prod), and net DOC (i.e., biological production of DOC - bacterial uptake of DOC; a negative value indicates that bacteria are reliant on DOC in part from outside of the mixed layer). In the classic paradigm, the flagellates capable of phototrophy (ANF) have no mixotrophic potential, while in the revised paradigm, they (MNF) exhibit mixotrophy (see also Figs. 2, 3 and 4). The inorganic nutrient regimes (all with an inorganic N input of 1 µM) are in Redfield N: P (molar ratio 16; A), low N: P (molar ratio 4; B), or high N: P (molar ratio 64; C).

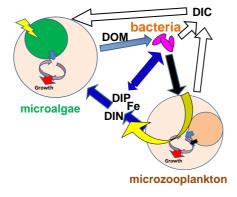
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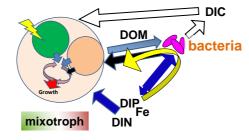


Fig. 6. Schematic showing the detailed involvement of bacteria and DOM for the supply of nutrients to support primary production (yellow arrows) in the classic paradigm (A) vs. the revised paradigm (B). See Sect. 4.2. Black arrows indicate predatory links.

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