| 2 | The effects of intermittent exposure to low pH and oxygen conditions on survival and | | | |
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| 3 | growth of juvenile red abalone | | | |
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| 5 | Running head: Low O ₂ and pH effects on abalone | | | |
| 6 | Authors: Tae Won Kim ^{1,2} , James P. Barry ¹ , Fiorenza Micheli ² | | | |
| 7 | | | | |
| 8 | Affiliation | | | |
| 9 | ¹ Monterey Bay Aquarium Research Institute, 7700 Sandholdt Road, Moss Landing, CA95039 | | | |
| 10 | ² Hopkins Marine Station, Stanford University, 120 Oceanview Blvd., Pacific Grove, | | | |
| 11 | CA93950 | | | |
| 12 | | | | |
| 13 | Correspondence: Tae Won Kim (e-mail: <u>ktwon@mbari.org</u>) | | | |

14 TEL: 831-775-1903

15 ABSTRACT

Exposure of nearshore animals to hypoxic, low pH waters upwelled from below the 16 17 continental shelf and advected near the coast may be stressful to marine organisms and lead to impaired physiological performance. We mimicked upwelling conditions in the laboratory 18 and tested the effect of fluctuating exposure to water with low pH and/or low oxygen levels 19 on the mortality and growth of juvenile red abalone (Haliotis rufescens, shell length 5-10mm). 20 Mortality rates of juvenile abalone exposed to low pH (7.5, total scale) and low O₂ (40% 21 saturation, 5 mg l⁻¹) conditions for periods of 3 to 6 hours every 3-5 days over 2 weeks did 22 not differ from those exposed to control conditions (O₂: 100% saturation, 12 mg l^{-1} ; pH 8.0). 23 However, when exposure was extended to 24 h, twice over a 15 day period, juveniles 2425 experienced 5-20 % higher mortality in the low oxygen treatments compared to control conditions. Growth rates were reduced significantly when juveniles were exposed to low 26 oxygen and low pH treatments. Furthermore, individual variation of growth rate increased 27 when juveniles were exposed simultaneously to low pH and low O₂ conditions. These results 28 indicate that prolonged exposure to low oxygen levels is detrimental for the survival of red 29 30 abalone, whereas pH is a crucial factor for their growth. However, the high individual variation in growth rate under low levels of both pH and oxygen suggests that cryptic 31 phenotypic plasticity may promote resistance to prolonged upwelling conditions by a portion 32 33 of the population.

Keywords: abalone, hypoxia, growth, mortality, ocean acidification, upwelling 34

35

1. INTRODUCTION

Marine ecosystems are under threat from ocean acidification as the excess burden of fossil 36 fuel CO₂ dissolves into the ocean (Orr et al., 2005;Hofmann et al., 2010). Many studies of the 37 biological effects of ocean acidification have focused on the predicted change of mean pH of 38 ocean surface waters derived from IPCC climate scenarios (e.g., Bibby et al., 2007;Orr et al., 39 2005; Dupont et al., 2010; Byrne et al., 2011). However, nearshore environments are also 40 influenced by low pH, low oxygen waters upwelled from below the continental shelf and 41 42 advected to shallow nearshore environments (Grantham et al., 2004;Feely et al., 2008;Hauri et al., 2009). Even though the low-pH, low-oxygen conditions generally persist in the 43 nearshore environment for only a few hours, this can happen routinely (50- 200 times / yr) in 4445 upwelling-driven ecosystems such as the California Current and Humboldt Current large marine ecosystems (Garcia-Reyes and Largier, 2010;Booth, 2011;Micheli et al., 2012). 46

It is assumed that many benthic animals living in nearshore marine environments are either 47 adapted to the local natural fluctuations of pH and dissolved oxygen (DO), or tolerate 48occasional short-lived exposure to potentially stressful conditions (Alenius and Munguia, 49 50 2012; Vaquer-Sunyer and Duarte, 2008). Nevertheless, more frequent and prolonged 51 upwelling or hypoxic events induced by climate change (Garcia-Reyes and Largier, 2010) 52 may impact marine animals. Local extremes in several factors, associated with upwelling (e.g. 53 oxygen, pH, and temperature), could be stressful for the population. In particular, juveniles 54 and other sensitive life stages may be highly vulnerable to these extremes. To date, however, 55 few experimental studies have evaluated the influence of upwelling-related exposure to low 56 pH and O₂ events on the mortality or growth of nearshore marine animals.

57 Individual variation in response to elevated environmental CO_2 and decreased O_2 is another 58 concern when investigating the response of organisms to upwelling events. Even when 59 environmental change causes significant negative impacts on most individuals' performance, tolerance by a subset of the population may promote adaptation for population persistence (Sih et al., 2012;Charmantier et al., 2008). High variation among individuals in response to elevated CO_2 has been shown to represent genetic diversity in populations of some marine taxa (Sunday et al., 2011;Langer et al., 2006;Pistevos et al., 2011) and this high individual variation could benefit the population and species as a whole.

Here we examine the effect of fluctuating exposure to low pH and low oxygen water on 65 juvenile stages of the red abalone, Haliotis rufescens. It is a large gastropod mollusk 66 67 inhabiting lower intertidal to shallow subtidal (to 30 m depth) environments from Oregon to Baja California (Micheli et al., 2008;Rogers-Bennett et al., 2007). As a key recreational 68 fishery and aquaculture resource in California, it is legally harvested north of San Francisco 69 70 and also cultured in abalone farms. To assess the sustainability of red abalone populations, their sensitivity to current environmental stressors should be identified. We investigated how 71fluctuating exposure to low pH and low oxygen waters affects the mortality and growth of 72 juvenile abalone using a pulse exposure experiment mimicking upwelling conditions. Booth 73 et al. (2012) reported that the mean duration of upwelling nearshore water (17 m in depth) 7475 with pH 7.6 and DO 4.6 mg/L off Monterey CA is 2.4 hrs, with a maximum duration of 40 hrs. Moreover, pH can drop dramatically by ca. 0.4 units within an hour. First, we studied the 76 effects of short-term (3 – 6 h) exposure to low pH (pH 7.5) and low oxygen (5 mg/ L) 77 78conditions that occur typically during coastal upwelling events (Booth et al., 2012). Second, to determine how abalone respond to prolonged low pH and hypoxic events, we extended the 79 exposure to "upwelling" conditions to 24 h twice within a 2 week period – an uncommon, but 80 81 occasional event in upwelling regions (Booth et al., 2012; Micheli et al., 2012).

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83 2. MATERIALS AND METHODS

84 2.1. Study species

Juvenile abalone (Haliotis rufescens, shell length 4-12 mm, 4-5 months of age) were 85 acquired from the Abalone Farm Co. located in Cayucos, California, where many individuals 86 collected over several years from Southern California abalone populations are reared at 87 ambient temperature (Mean \pm SD: 11.9 \pm 0.76). Juveniles in this study were maintained at 88 Hopkins Marine Station of Stanford University (HMS) in aquaria with replenishing water 89 flow at ambient temperature (Mean \pm SD: 12.6 \pm 3.4, 11-13.8 °C) and pH (Mean \pm SD: 7.79 90 \pm 0.16, Ranges: 7.61-7.98). They were fed *ad libitum* with small portions of ground Hikari ® 91 92 algae wafers every other day. The opening of each jar was covered by a 12×12 cm nylon mesh (1×1 mm mesh size), secured with a rubber band. For each experiment, jars with 93 juvenile abalone were maintained at HMS for 3 days. Jars were then transferred to the 94 95 Monterey Bay Aquarium Research Institute (MBARI), Moss Landing, California, and immersed in a transparent plastic holding tank ($60 \times 30 \times 35$ cm) with flow-through circulation 96 in a chilled environmental chamber for at least 3 days for acclimation. The water temperature 97 of the holding tank was maintained at approx. 9 °C using two heaters and water pH_T (total 98 scale) was maintained at pH 8.0 (normal). 99

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101 **2.2.** Treatment protocol

102 Three treatments with two different levels of pH and oxygen were used: 1. Low pH (7.5) & 103 Low O₂ (40% saturation, 5 mg l^{-1}) = ' upwelling"; 2. High pH (8.0) & Low O₂ = ' low 104 oxygen' and 3. High pH (8.0) & High O₂ (100% saturation, 12 mg l^{-1}) as a 'control' (Table 1). 105 A low pH and high DO treatment was not included in this experiment due to space, and 106 because the primary focus of the experiment concerned the effects of upwelling conditions 107 rather than acidification alone.

108 Seawater pH, oxygen, and temperature levels were controlled using a gas-regulated 109 aquarium system (Barry et al., 2008). This system is capable of modifying ambient seawater from MBARI's ocean intake to produce water with specified pH, pO₂, and temperature. For each treatment, water produced by the control system was delivered at approx. 30 ml/min to each jar through a 3 cm diameter PVC manifold, and a 10 mm diameter hose. Consequently, when experimental animals in the holding tanks were transitioned to treatment conditions (Δ pH, Δ pO₂), inflowing treatment waters would mix with normal water in each jar causing the pH and pO₂ to approach specified treatment levels over ca. 15-20 minutes as treatment waters were flushed from the jar.

117 Seawater temperature for all treatments was maintained at 6 °C. Though this low 118 temperature may be near the minimum level tolerable level for red abalone, the two to three 119 degree difference between the holding tank and treatment water effectively mimicked the 120 temperature reduction associated with upwelling in the central and northern Pacific coast of 121 USA (e.g. Point St. George, CA, and Oregon).

122 The pH and DO of seawater of treatment waters produced by the aquarium control system 123 were monitored continuously. In addition, conditions were measured in one randomly-124 selected jar at least twice for each exposure period, using a portable pH, DO meter (Thermo Scientific®: Orion 5 Star Series) (Table 1). Continuous monitoring of pH in control waters in 125the aquarium system was not available for the experimental period. Measurements of the 126 127 ambient pH water (nominal pH 8.0) at MBARI performed before (Oct 26, Nov 22, Dec 16 2010) and after the experiment (Mar 30, 2011) using the spectrophotometric method of 128 Dickson et al. (2007) were pH 7.99 \pm 0.04 (SD), and were very similar to values measured 129 for the high pH & low O_2 (pH 8.01 ± 0.18) treatment and control (high pH & high O_2) 130 treatment (pH 8.09 \pm 0.18) during the experiment using a portable pH meter. Therefore, we 131 132 assumed that there was little deviation from pH 8.0 both in the high pH treatment water and in normal water in the holding tank at MBARI. However, aquaria at HMS during the 133 recovery period for experiment II experienced rather low pH (See the supplementary Figure 134

135 1a). pH meters were regularly calibrated with pH values of Fisher Scientific® buffer salt
136 solutions (pH 6.86 and pH 9.18, total scale) measured by a UV spectrophotometer (UV-1601
137 Shimadzu®). To determine the calcite and aragonite saturation states of treatment waters,
138 samples were collected from all treatments three times after the experiment. pH of the
139 samples was measured using the spectrophotometric method (Dickson et al., 2007) and DIC
140 (Dissolved Inorganic Carbon) was measured by non-dispersive infrared analysis (LI-COR
141 model 6262), as detailed by Friederich et al. (2002)

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143 2.3. Experiment I

We tested the effects of repeated short-term exposure on the mortality of juvenile abalone in a 144145 controlled laboratory experiment. Twenty juvenile abalone individuals (Mean ± SD shell length: 7.36 ± 0.76 mm) were assigned to each of twelve 500ml transparent glass jars with 146 seawater. After 3 days of acclimation to natural (high O_2 (100% sat.), high pH_T (8.0)) 147 conditions in the holding tank, 4 randomly-selected jars were assigned to each of 3 treatments 148 (control, upwelling, and low oxygen treatments). At the completion of the acclimation period, 149 150 abalone were exposed to their assigned treatment for 3 h, the returned to the holding tank for 3 days. This 3h treatment: 3d control cycle was then repeated once. Abalone were then 151 exposed to treatment conditions for 6h, followed by a 2 week recovery period (See SM Fig. 152153 1).

The mortality of juvenile abalone in each jar was checked immediately after exposure to treatment conditions and daily during the 3 day and 2 week recovery periods. To determine the live / dead status of juvenile abalone, they were overturned and then pricked using a dissecting pin. Individuals that did not respond to the pin were considered dead, and were transferred to vials containing 70% ethanol solution. Due to the short duration of this experiment, growth rates of abalone were not measured. 160

161 2.4. Experiment II

162 To test the effects of more prolonged periods of exposure to upwelling waters, we used an experimental protocol similar to the Experiment I, but with extended periods of exposure to 163 treatment waters. Twelve abalone juveniles (Mean \pm SD, shell length: 7.42 \pm 1.23 mm) were 164 allotted to each of 18 jars. Each individual was marked with a combination of color dots on 165 the shell using paint markers, which were then coated with instant glue to preserve the color 166 167coding. The shell length and width of each individual was measured using a digital caliper to the nearest 0.01mm. The control and treatment waters were the same used in the previous 168 experiment. After an acclimation period of 3 days under control conditions, abalone were 169 170 exposed to their assigned treatment conditions for 24h, followed by immersion in water in the holding tank for 6 days. Abalone were then returned for 24 h to treatment waters, and then 171immersed in the holding tank for 6 days again and then returned to the aquarium at HMS for 172 recovery and immersed for 11 days (SM Fig. 1). The survival of each abalone was checked 173 on the 1st, 2nd, 3rd, 5th, and 6th day after each exposure. The shell length and width of each 174 175 individual were measured after its death or at the end of the experiment, for both exposures.

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177 2.5. Statistical analyses

Mortality was calculated by dividing the number of dead abalone by the number of abalone initially placed in each jar. We calculated the cumulative mortality of all abalone in each jar after each exposure time point and used repeated measures, one-way ANOVA for each exposure, including post-exposure periods. Repeated-measures ANOVA was then applied to the entire experimental period with the same abalone group. In all repeated-measures analyses, the assumption of equal between-group correlations and group variances ("sphericity") was not violated (Mauchly's test, all p > 0.05). When significant differences 185 were detected among treatments, Tukey tests were applied for post-hoc comparisons.

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$$\frac{L_f - L_i}{L_i} \times \frac{1}{D_{if}}$$

187

 L_{f} is the final shell length at the time of death or in the end of the experiment, L_{i} is the initial shell length measured before the exposure experiment, D_{if} is the interval (days) between measurements of initial shell length and final shell length.

The daily proportional rate of change in shell length (growth) was calculated as:

We assumed linear growth for juvenile abalone because their shell lengths were similar at the beginning of the experiments (for both experiments) and the experimental period was short. To evaluate the relationship between size and growth, we plotted daily proportional growth rate versus initial shell size and found no significant slope ($F_{1,149} = 1.80$, p = 0.1818), indicating that, at least in the range of sizes used here, growth rate is unrelated to initial size.

To verify if both treatment level and live/dead status at the end of the experiment influence the proportional shell growth, we applied a two-way ANOVA. All growth data were transformed by arcsine square root prior to analysis. Fisher's PLSD tests were applied for post-hoc comparisons. An *F*-Test of equality for variances was used to determine differences in the variance of growth rate between treatments.

201

3. RESULTS

203 3.1. Experiment I

Abalone mortality did not differ among treatments during both 3 h exposure periods (1st exposure: $F_{2,9} = 1.286$, p = 0.322; 2nd exposure: no mortality in any treatments) or during the 3 days after those exposures (1st exposure: $F_{2,9} = 0.300$, p = 0.7479, 2nd exposure: $F_{2,9} =$ 1.544, p = 0.265, Fig. 1a). There was also no significant difference in mortality between treatments during the 6 h-exposure ($F_{2,9} = 1.000$, p = 0.405) or post-exposure periods ($F_{2,9} =$ 209 0.459, p = 0.645, Fig. 1a). Overall, there was no significant difference in cumulative mortality 210 among treatments during Experiment I ($F_{2,9} = 0.175$, p = 0.841).

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3.2. Experiment II

Repeated measures ANOVA indicated significant variation in mortality among treatments following 24 h exposures to treatment levels, for up to 3 days after exposure (2 days: $F_{2,15}$ = 4.059, p = 0.039, 3 days: $F_{2,15} = 3.966$, p = 0.041, Fig. 1b). Mortality was significantly higher in the low oxygen treatment than the control (Tukey post-hoc test, p = 0.035) but all other comparisons among treatments were not significant. Cumulative mortality over 4 days after exposure to control or treatment conditions did not vary significantly, indicating that the effects of exposure may not persist beyond 4 days.

After the first 24 h exposure, the daily proportional growth rate did not differ among 220 treatments (ANOVA: $F_{2.27} = 3.168$, p = 0.0586). By the end of the second 24 h exposure to 221 treatment conditions, growth rates differed significantly between treatments ($F_{2, 149} = 16.509$, 222 p < 0.0001, Fig. 2). In fact, shells in the upwelling treatment (low pH and low O₂) showed 223 224 negative growth, significantly lower than for animals from control or low-oxygen conditions (p < 0.0001 for both). Variation in growth rate was significantly different between treatments 225 (*F*-test of variance: $F_{2,149} = 16.509$, p < 0.0001). In particular, it was higher in the upwelling 226 227 treatment than in control (F-test of variance: $F_{53, 50} = 0.411$, p < 0.0001) and low-oxygen 228treatments ($F_{49,52} = 0.411$).

Whether abalone individuals died during the experiment was not related to differences in growth. Comparison of "alive" or "dead" status or its interaction with treatment did not influence the growth rate (Two-way ANOVA: Alive/dead status: $F_{1, 146} = 0.116$, p = 0.7343, Alive/dead status × treatment : $F_{1, 146} = 1.011$, p = 0.3665).

4. DISCUSSION

These experiments indicate that repeated short-term (3-6 h) exposure of abalone juveniles to 235 236 conditions documented during upwelling (Booth, 2011) has no detectable immediate effect on their survival. In contrast, periodic, prolonged exposure (24 h twice) to low oxygen 237 significantly decreased survival and exposure to upwelling treatment significantly reduced 238 the shell growth. These results suggest that abalone populations may be adapted to 239 fluctuations in pH and O₂ over time scales typically occurring along the central California 240 241 coast, but not to more prolonged duration of low pH and O₂, similar to those documented off 242 the coast of Mexico (Micheli et al., 2012). Low O₂ may always be stressful, but tolerable for short periods. Extended periods of hypoxia and/or hypercapnia may lead to accumulation of 243 244 impacts and eventually reduced growth or death.

Because extended exposure to low oxygen per se has lethal effects on juvenile abalone, 245 oxygen concentration is likely to be a crucial factor influencing the persistence of abalone 246 populations. However, seawater pH should be also considered as an important factor 247influencing abalone populations because low pH has a deleterious effect on growth. Growth 248 249 can influence reproductive success and susceptibility to predation (Rossetto et al., 2012). Shell length of individuals exposed to the low pH, low O₂ treatment simulating upwelled 250 waters decreased, suggesting that shells were dissolving in the low pH water. Given that 251 252 ocean acidification did not change the expression of shell growth genes in the larval stage in a previous experiment (Zippay and Hofmann, 2010), epigenetic processes or differential 253 physiological investment might have influenced the shell growth. The aragonite and calcite 254 255saturation states of the upwelling treatment waters were less than 1 (Table 1), which should increase the energetic cost of shell formation for abalone (Portner, 2008;Byrne et al., 2011). 256

The effects of low pH and low oxygen may have different effects on the growth of different life stages and different species. Interestingly, mortality did not differ between the control and upwelling treatment with low pH and low O_2 . Our expectation was that exposure to low pH waters during upwelling, coupled with low oxygen levels, would be the most stressful, leading to reduced survival or growth or both. Our results suggest otherwise, and further information is required to understand the combined effects of pH and O2 on abalone performance.

Although very few data are available concerning variation in the carbonate chemistry of 264 nearshore waters in Monterey Bay, time-series measurements of conditions at the intake pipe 265 266 for the Monterey Bay Aquarium indicate that the average duration of upwelling conditions (~4.6 mg .L $^{-1}$ O2, pH ~7.6) in nearshore Monterey Bay environments is 2.4 \pm 2.6 h (Booth, 267 2011). Juvenile abalone appear to have the capacity to cope with short exposure to these 268269 conditions. Prolonged upwelling conditions stressful to juvenile abalone may occur only 270 rarely. Over the past 12 years, the maximum observed duration of low oxygen and low pH events nearshore Monterey Bay was 40 h (Booth, 2011), similar to recent measurements off 271 Baja California (Micheli et al., 2012). A trend towards prolonged upwelling conditions 272 associated with recent climatic changes (Garcia-Reyes and Largier, 2010) is expected to 273 274 increase the frequency and duration of low-oxygen, low-pH events (Feely et al., 2008;Nam et al., 2011;Melzner et al., 2012), possibly leading to more frequent prolonged exposure. 275

Prolonged upwelling conditions can have a significant impact on the growth of juvenile abalone, and possibly other calcifying, marine invertebrates (e.g., Vaquer-Sunyer and Duarte, 2008;Kroeker et al., 2010). Changes in individual performance (growth, survival, reproduction) of abalone in response to future ocean conditions will likely have significant, but poorly known effects on the demographic dynamics of California abalone populations and the ecology of nearshore ecosystems.

Our observation of increased individual variation in growth rate after exposure to upwelling conditions is intriguing. Such variation among individuals suggests that phenotypic 284 plasticity or genetic variation or both promote a range of performance within the population in response to shifts in pH and oxygen. While a large portion of the population may 285286 experience reduced growth or survival, some individuals may acclimate to extended upwelling events by maintaining their calcification, thereby promoting adaptation through 287 enhanced survival and fitness by this subpopulation. This is in agreement with results of other 288 studies that have shown high individual variation of marine animals in response to low pH 289 conditions (Sunday et al., 2011;Schlegel et al., 2012;Pistevos et al., 2011). Phenotypic 290 291 plasticity of abalone populations may explain higher individual variation in growth rate in 292 response to low pH and low oxygen. Cryptic genetic variation might also be expressed when juveniles are exposed to the novel environments (Ghalambor et al., 2007). Some part of the 293 294 reaction norm can be adaptive to new conditions. If such adaptive variation is heritable, red abalone may be able to adapt to extended exposure to upwelling conditions. Indeed, recent 295 genetic information on red abalone suggests that genes for biomineralization and resistance to 296 hypoxia are under selection in upwelling regions (de Wit and Palumbi, 2013). Gene 297 expression studies of abalone responses to upwelling conditions are likely to shed light on 298 this issue. 299

Low O₂ and low pH events during upwelling periods are largely coupled because these 300 parameters are regulated by respiratory oxygen consumption and carbon dioxide release by 301 302 the deep water biological community. Advection of deeper, low O2, low pH waters toward the surface and inshore leads to exposure of coastal taxa to potentially stressful 303 environmental conditions (Bianucci et al., 2011; Nam et al., 2011). Future increases in the 304 305 influx of CO₂ from the atmosphere to the ocean will further reduce the pH of upwelled waters (Feely et al., 2008;Feely et al., 2010), and will decouple, to some extent, the linkage between 306 oxygen and carbon dioxide in these waters. This study did not assess the influence of 307 increasing temperature, which is also changing with upwelling, pH, and oxygen. Further 308

309 study of the effects of projected changes in ocean conditions in upwelling regimes on diverse 310 nearshore species are needed to understand and predict future shifts in the structure and 311 function of coastal ecosystems.

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Table 1. Carbonate system and other physical parameters for experimental treatments measuring the response of juvenile abalone to upwelling conditions (Mean \pm SD).

436

| Variable | Treatments | | | |
|---------------------------------------|------------------------------|-----------------------------|-------------------------------|--|
| | High pH & Low O ₂ | Low pH & Low O ₂ | High pH & High O ₂ | |
| | (low-oxygen) | (upwelling) | (control) | |
| Measured pH | 8.00 ± 0.05 | 7.51 ± 0.08 | 7.93 ± 0.03 | |
| (total scale) | | | | |
| $DO (mg l^{-1})$ | 5.99 ±0.91 | 6.38 ±1.53 | 11.68 ±2.86 | |
| $TCO_2 (\mu mol kg^{-1})$ | 2116.10±56.74 | 2233.33±54.29 | 2196.72 ± 141.23 | |
| Salinity (ppt) | 33.0±0.1 | 33.0 ±0.1 | 33.0 ±0.1 | |
| Temperature (°C) | 6.0 ± 0.0 | 6.0 ± 0.0 | 6.0 ± 0.0 | |
| PCO ₂ (µatm) | 434.36±47.33 | 1424.36±224.91 | 535.48 ± 23.65 | |
| Alkalinity (µEq kg ⁻¹) | 2249.75 ± 80.59 | 2217.70 ± 73.34 | 2310.09 ± 152.68 | |
| Calcite saturation | 2.43 ± 0.38 | 0.85 ± 0.19 | 2.23 ± 0.25 | |
| Aragonite | 1.53 ± 0.23 | 0.54 ± 0.12 | 1.41 ± 0.16 | |
| saturation | | | | |
| HCO_3^- (µmole ⁻ | 1993.14 ± 43.57 | 2125.44 ± 56.19 | 2076.88 ± 130.79 | |
| kg ⁻¹) | | | | |
| CO_3^2 (µmole kg | 100.86 ± 15.67 | 35.40 ± 7.80 | 92.58 ± 10.55 | |
| ¹) | | | | |

The parameters were calculated with CO2sys (Pierrot et al., 2006) using the pH and TCO₂
values with dissociation constants from Dickson & Millero (1987) and KSO₄ using Dickson

(1990).

- 441 Figure 1.
- **(a)**



(b)







449 SM Figure 1

450 (a)



(b)



454 Figure legends

455

Fig. 1. *Haliotis rufescens*. Cumulative mortality (Mean \pm s.e.) of juveniles abalone after each exposure to control, upwelling, and low-oxygen treatments. (a) Experiment I. (b) Experiment II. Shaded days represent exposure of abalones to different treatments. "Day number" represents the days after the exposure treatment.

460

Fig. 2. *Haliotis rufescens*. Daily proportional change in shell length of juveniles among control, upwelling, or low-oxygen treatments. Square dot and line represent mean \pm s.e.

463

SM Fig. 1. Diagram describing treatment exposure and recovery schedule and pH values changing for (a) Experiment I and (b) Experiment II. Red triangle (high pH) and blue square (low pH) dots represent the pH values measured using a pH meter and dashed line represent the assumed pH value throughout the experiment.

468

469 Short tile: Low pH and O₂ effects on abalone