

# On the relationship between ecosystem-scale hyperspectral reflectance and CO<sub>2</sub> exchange in European mountain grasslands

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## Abstract

In this paper we explore the skill of hyperspectral reflectance measurements and vegetation indices (VIs) derived therefrom in estimating carbon dioxide (CO<sub>2</sub>) fluxes (~~net ecosystem exchange—NEE; gross primary production—GPP~~), and some key ecophysiological variables related to NEE and GPP (light use efficiency— $\epsilon$ ; initial quantum yield— $\alpha$ ; and GPP at saturating light— $GPP_{max}$ ) of grasslands. Hyperspectral reflectance data (~~400–1000 nm~~), CO<sub>2</sub> fluxes and biophysical parameters were measured at three grassland sites located in European mountain regions using standardized protocols. The relationships between CO<sub>2</sub> fluxes, ecophysiological variables, traditional VIs and VIs derived using all two-band combinations of

1 wavelengths available from the whole hyperspectral data space were analysed. We found that  
2 ~~hyperspectral VIs VIs derived from hyperspectral data~~ generally explained a large fraction of the  
3 variability in the investigated dependent variables, ~~but differed in their ability to estimate midday~~  
4 ~~and daily average CO<sub>2</sub> fluxes and various derived ecophysiological parameters—and that they~~  
5 ~~generally exhibited more skill in estimating midday and daily average GPP and NEE, as well as~~  
6 ~~GPP<sub>max</sub>, than  $\alpha$  and  $\epsilon$ .~~ Relationships between VIs and CO<sub>2</sub> fluxes and ecophysiological  
7 parameters were site-specific, likely due to differences in soils, vegetation parameters and  
8 environmental conditions. Chlorophyll and water content related VIs ~~(e.g. CI, NPCI, WI),~~  
9 ~~reflecting seasonal changes in biophysical parameters controlling the photosynthetic process,~~  
10 explained the largest fraction of variability in most of the dependent variables. Band selection  
11 based on a combination of a genetic algorithm with random forests (GA-rF) confirmed that it is  
12 difficult to select a universal band region suitable ~~for describing ecophysiological parameters,~~  
13 ~~CO<sub>2</sub> fluxes and biophysical variables~~ across the investigated ecosystems. Our findings have  
14 major implications for up-scaling terrestrial CO<sub>2</sub> fluxes to larger regions and for remote and  
15 proximal sensing sampling and analysis strategies and call for more cross-site synthesis studies  
16 linking ground-based spectral reflectance with ecosystem-scale CO<sub>2</sub> fluxes.

17

## 18 **1 Introduction**

19 Understanding the mechanisms that drive the carbon dioxide (CO<sub>2</sub>) exchange of terrestrial  
20 ecosystems is one of the main challenges for ecologists working on climate change (Beer et al.,  
21 2010). Plant gross photosynthesis, also referred to as gross primary productivity (GPP), is one of  
22 the major components of the global carbon cycle. It interacts in complex ways with  
23 environmental factors such as radiation, nutrients, soil moisture, vapor pressure deficit, air  
24 temperature and soil temperature (Drolet et al. 2005). Plant biochemistry and structure determine  
25 many fundamental ecosystem patterns, processes and dynamics (Lambers et al. 1998; Waring  
26 and Running 1998). The canopy nitrogen content regulates the canopy photosynthetic capacity  
27 and the canopy light use efficiency ( $\epsilon$ ) (Ollinger et al., 2008). In addition, the canopy chlorophyll  
28 content plays an important role in controlling ecosystem photosynthesis and carbon gain (Peng et  
29 al., 2011; Gitelson et al., 2006).

1 Optical remote sensing can help ecologists in qualitatively and quantitatively assessing plant and  
2 canopy properties (e.g. biomass (Vescovo et al. 2012), water content (Clevers et al., 2010),  
3 nitrogen content (Ollinger et al., 2008; Knyazikhin et al., 2012), chlorophylls (Gitelson et al.,  
4 2006) and photosynthetic rate (Inoue et al., 2008) that drive ecosystem processes related to the  
5 carbon cycle).

6 Empirical and physical based methods have been proposed by several authors to interpret optical  
7 plant and canopy properties. Empirical methods consist of, for example, linear regression  
8 analysis between plant or canopy properties and optical data. The most used empirical methods  
9 are: hyperspectral index methods (Peñuelas et al., 1993; Sims and Gamon, 2002; Inoue et al.,  
10 2008) and multi-variable statistical methods, (e.g. stepwise linear regression, genetic algorithm,  
11 neural network (Grossman et al., 1996; Riaño et al., 2005a; Li et al., 2007). Physical methods are  
12 based on the use of radiative transfer models (RTMs) to simulate light absorption and scattering  
13 through the canopy as a function of canopy structure and leaf biochemical composition  
14 (Jacquemoud et al., 2000; Zarco-Tejada et al., 2003). Therefore, RTMs help in quantifying the  
15 contribution of canopy biophysical and biochemical variables to canopy reflectance. ~~One of the~~  
16 ~~most popular RTM is PROSAIL, based on the coupling of the SAIL canopy bidirectional~~  
17 ~~reflectance model (Verhoef et al., 1984) and the PROSPECT leaf optical properties model~~  
18 ~~(Jacquemoud and Baret, 1990).~~ Such models can be used for identifying regions of the light  
19 spectrum that are of particular importance for specific biophysical properties of vegetation. ~~The~~  
20 ~~sensitivity analysis of the PROSAIL model~~ For example, it was demonstrated that the red-edge  
21 region (between 680 nm to 730 nm) of the spectrum is sensitive to the leaf chlorophyll content  
22 and leaf area index (LAI) (Baret et al., 1992). It is also well accepted that an increase of LAI  
23 includes a decrease of reflectance in the red and an increase in the near-infrared (NIR) region  
24 (Jacquemoud, 1993). In the NIR region, effects of LAI and the leaf angle distribution equally  
25 contribute to the reflectance response (Bacour et al., 2002a). NIR reflectance between 800 nm  
26 and 850 nm is also related to canopy N content (Ollinger et al., 2008; Knyazikhin et al., 2012). In  
27 addition, the combination of the reflectance in NIR and in the short wave infrared region (SWIR)  
28 is correlated with canopy water content (Colombo et al., 2008), but the reflectance between 1000  
29 nm and 1400 nm is also highly sensitive to LAI. So, some attention is needed when these  
30 spectral regions are used to retrieve water content considering that the canopy properties in a  
31 given ecosystem often co-vary (Bacour et al., 2002b).

1 The drawback of such an approach consists in the fact that the process of building a model  
2 implies approximations and assumptions. For this reason we opted for a purely data based  
3 approach such as the hyperspectral index approach. This method consists of the use of spectral  
4 vegetation indices (VIs) defined as spectral band ratios, or normalized band ratios between the  
5 reflectance in the visible (VIS) vs. NIR region or VIS vs. VIS or NIR vs. NIR.

6 The typical optical sampling approach, which is linking spectral observations with CO<sub>2</sub> fluxes, is  
7 based on the Monteith equation (1972, 1977):

$$8 \quad GPP = \varepsilon * PAR * fAPAR \quad (1)$$

9 where  $\varepsilon$  is the light use efficiency and fAPAR is the fraction of absorbed photosynthetically  
10 active radiation); both  $\varepsilon$  and fAPAR can be retrieved by remote optical observations. A wide  
11 number of VIs that can potentially be used to model the productivity of terrestrial ecosystems (as  
12 a proxy of  $\varepsilon$  and fAPAR) has been suggested (Inoue et al., 2008; Coops et al., 2010; Peñuelas et  
13 al., 2011; Rossini et al. 2012). The various VIs differ in their sensitivity to changes in  
14 photosynthetic status. “Greenness indices” – such the widely used Normalized Difference  
15 Vegetation Index (NDVI) – demonstrated to be a good proxy for fAPAR, but are not sensitive to  
16 rapid changes in plant photosynthesis which are induced by common environmental and  
17 anthropogenic stressors (Gitelson et al., 2008; Hmimina et al., 2014; Soudani et al., 2014).  
18 However, in ecosystems characterized by strong dynamics (e.g. grasslands and crops with a  
19 strong green-up and senescence), other VIs are able to effectively monitor seasonal changes in  
20 biophysical parameters controlling canopy photosynthesis such as fAPAR and chlorophyll  
21 content and consequently, can be adopted to monitor the seasonal and spatial variability of  
22 carbon fluxes (Gitelson et al., 2012; Sakowska et al., 2014). Short-term changes in  $\varepsilon$  can be  
23 remotely detected through a spectral proxy of the xanthophyll cycle (Photochemical Reflectance  
24 Index, PRI; Gamon et al., 1992). The PRI is one of the most promising VIs for a direct  
25 estimation of photosynthetic light use efficiency and of its seasonal and diurnal variations  
26 (Nichol et al., 2002). Latest developments of the sun-induced fluorescence method may allow  
27 even more direct remote sensing of plant photosynthesis in the near future (Meroni et al., 2009;  
28 Rossini et al., 2010; Frankenberg et al., 2011). At canopy scale, the relationship between PRI and  
29  $\varepsilon$  was shown to be site-dependent (Garbulsky et al., 2011; Goerner et al., 2011) and strongly  
30 affected by environmental conditions (Soudani et al. 2014).

1 Whereas previous studies have demonstrated the ability of remote sensing data to allow  
2 modelling ecosystem GPP at ecosystem scale (e.g. Gianelle et al., 2009; Wohlfahrt et al., 2010;  
3 Rossini et al. 2012; Sakowska et al., 2014), a universal model for GPP estimation applicable  
4 across different ecosystems and a wide range of environmental conditions is still missing. ~~In~~  
5 ~~addition,~~ ~~p~~Previous studies often focussed on single sites with specific characteristics (e.g.  
6 climate, vegetation composition, soil type; see Wohlfahrt et al., 2010). ~~and~~In addition, different  
7 studies were often based on the use of used different sensors, platforms and protocols (Balzarolo  
8 et al., 2011), making generalisation difficult. Moreover, most of the studies have either relied on  
9 reflectance measurements in a few spectral wavebands (e.g. Wohlfahrt et al., 2010 and Sakowska  
10 et al, 2014) ~~or a minimum number of bands needed to calculate the most common VIs~~, missing  
11 potentially important information in under-sampled spectral regions ~~that could help explain~~  
12 ~~carbon fluxes and variability~~. In order to overcome such heterogeneity in spectrometry  
13 measurements, SpecNet (<http://specnet.info>; Gamon et al., 2006), the European COST Action  
14 ES0903 (EUROSPEC; <http://cost-es0903.fem-environment.eu/>) and the COST Action ES1309  
15 (OPTIMISE; [http://www.cost.eu/domains\\_actions/essem/Actions/ES1309](http://www.cost.eu/domains_actions/essem/Actions/ES1309)) focused on the  
16 definition of a standardized protocol for making optical measurements at the eddy covariance  
17 CO<sub>2</sub> flux towers (Gamon et al., 2010).

18 The overarching objective of the present paper is thus to develop a common framework for  
19 predicting grassland carbon fluxes and ecophysiological parameters based on optical remote  
20 sensing data across measurement sites exposed to diverse natural (climate) and anthropogenic  
21 (management) factors. To this end we combine eddy covariance CO<sub>2</sub> flux measurements with  
22 ground-based hyperspectral reflectance measurements for six different grasslands in Europe. In  
23 order to make the optical and fluxes measurements comparable, these were acquired at the six  
24 sites following a common protocol resulting in a unique standardized data set. We focused on  
25 European grasslands, ~~which since covering~~ roughly 22% (80 million ha) of the EU-25 land area,  
26 ~~grasslands and~~ are thus among the dominating ecosystem types in Europe (EEA, 2005).  
27 ~~Accordingly and~~ their role in the European carbon balance has received a lot of scientific interest  
28 (Soussana et al., 2007; Gilmanov et al., 2007; Wohlfahrt et al., 2008; Ciais et al. 2010). ~~While~~  
29 ~~d~~Direct measurements of the grassland carbon exchange have been carried out and are still  
30 ongoing at a number of different grassland sites in Europe ~~notably in the two EU projects~~  
31 ~~GreenGrass~~ (e.g. Soussana et al., 2007; ) ~~and CarboMont~~ (Cernusca et al., 2008). ~~s~~Scaling up

1 these plot-level measurements to the continental scale requires a modelling approach, typically  
2 based on or supported by remotely sensed data. Therefore, we believe that this study will  
3 improve the current knowledge on modelling the carbon dynamics of European grasslands.

4

## 5 **2 Materials and methods**

### 6 **2.1 Experimental site description**

7 This study was carried out at six experimental mountain grassland sites in Europe covering  
8 different climatic and grassland management conditions existing in the mountain regions of  
9 Europe, which were already part of the preceding study by Vescovo et al. (2012). This dataset  
10 combined *in-situ* hyperspectral, biophysical and flux measurements based on common protocol  
11 (for more details see sect. 2.2, sect. 2.3 and sect. 2.4). This dataset is unique since no common  
12 protocol for hyperspectral measurements exists in the various eddy covariance networks (e.g.  
13 FLUXNET). In this study, three of these sites (Amplero, Neustift and Monte Bondone, see  
14 Table 1 and S1 in the Supplemental section) composed the main dataset used in the analysis,  
15 while the other three sites (Table S2 in Supplemental section) were used to independently  
16 validate the models obtained with the main dataset.

17

18 Main study sites (Table 1 and S1 in the Supplemental section):

#### 19 *Amplero*

20 The Amplero site is situated in the Mediterranean Appennine mountain region of Italy (41.90409  
21 N, 13.60516 E) at 884 m a.s.l.. This site is characterized by mild, rainy winters and by an intense  
22 drought in summer. Amplero is managed as a hay meadow with one cut in late June and  
23 extensive grazing during summer and autumn.

#### 24 *Monte Bondone*

25 The Monte Bondone site is situated in the Italian Alps (46.01468 N, 11.04583 E) at 1550 m  
26 a.s.l.. This site is characterized by a typical sub-continental climate with mild summers and

1 precipitation peaks in spring and autumn. Monte Bondone is managed as an extensive meadow  
2 with one cut in mid-July.

### 3 *Neustift*

4 The Neustift grassland site is located in the Austrian Alps (47.11620 N, 11.32034 E) at 970 m  
5 a.s.l.. The climate of this area is continental/Alpine, with precipitation peaks during the summer  
6 (July). This site is intensively managed as a hay meadow with three cuts in mid-June, beginning  
7 of August and at the end of September.

8

9 Validation sites:

### 10 *Längenfeld*

11 The site Längenfeld is located in the Austrian Alps (47.0612 N, 10.9634 E) at 1180 m a.s.l.. The  
12 climate of this area is continental/Alpine, however compared to the other Alpine sites in this  
13 study, the site receives comparably less precipitation due to rain shadowing effects from both the  
14 North and South. The site is intensively managed as a hay meadow with three cuts in mid-June,  
15 mid-August and mid-October.

16

### 17 *Leutasch*

18 The site Leutasch is located in the Austrian Alps (47.3780 N, 11.1627 E) at 1115 m a.s.l.. The  
19 climate of this area is Alpine with substantial precipitation due to its position on the north range  
20 of the Alps. The site is extensively managed as a hay meadow with two cuts at the end of June  
21 and beginning of September.

22

### 23 *Scharnitz*

24 The site Scharnitz is located very close to Leutasch (47.3873 N, 11.2479 E) at 964 m a.s.l. and  
25 the climate is thus very similar to Leutasch. The site is extensively managed as a hay meadow  
26 with two cuts at the beginning of July and beginning of September.

27

## 1 **2.2 Hyperspectral reflectance measurements**

2 The canopy hyperspectral reflectance measurements were collected at each site under clear sky  
3 conditions close to solar noon (between 11:00 to 14:00 Central European time) using the same  
4 model of a portable spectroradiometer (ASD FieldSpec HandHeld, Inc., Boulder, CO, USA;  
5 ~~serial numbers: 1275 for Amplero, 6354 for Monte Bondone in 2006/2013 and 1191 for Neustift,~~  
6 ~~Längenfeld, Leutasch, Scharnitz and Monte Bondone in 2005~~) at all sites. The spectroradiometer  
7 acquires reflectance values between 350 and 1075 nm with a Full Width Half Maximum  
8 (FWHM) of 3.5 nm and a spectral resolution of 1 nm. In order to achieve a better match between  
9 the eddy covariance flux footprint and optical measurements, a cosine diffuser foreoptic (ASD  
10 Remote Cosine Receptor, Inc., Boulder, CO, USA), calibrated by the manufacture, was used for  
11 nadir/zenith measurements (Gianelle et al., 2009; Fava et al., 2009; Meroni et al. 2011). The  
12 ASD's cosine receptor is designed with a geometry and material that provides a hemispherical  
13 field of view (FOV) of 180° and optimizes the cosine response. To reduce the nadir FOV  
14 contamination (i.e. sky irradiance and for canopy irradiance) due to the hemispherical view of  
15 the sensor the instrument was placed on a 1.5 m long horizontal arm at a height of 1.5 m above  
16 the ground. To avoid the zenithal FOV contamination, the measurements were taken at least at a  
17 15 m distance from the eddy covariance tower (maximum height of the tower was 6 m). The  
18 vegetation irradiance (sensor pointing nadir) and sky irradiance (sensor pointing zenith) were  
19 measured by rotating the spectroradiometer alternately to acquire spectra from the vegetation and  
20 from the sky. Hemispherical reflectance was derived as the ratio of reflected to incident radiance.  
21 Each reflectance spectrum was automatically calculated and stored by the spectroradiometer as  
22 an average of 20 readings. Before starting each spectral sampling, a dark current measurement  
23 was done. For more details on experimental set-up see Vescovo et al. (2012). Spectral  
24 measurements were collected from spring until the cutting date at Amplero and Monte Bondone,  
25 while at the site in Neustift, which is cut three times during the season, spectral measurements  
26 were taken about once per week throughout the growing season of 2006.

27



## 1 **2.3 Biophysical and biochemical canopy properties**

2 Samples for dry phytomass, nitrogen and water content measurements were collected at the time  
3 of the hyperspectral measurements in the field of view of the hyperspectral sensor (see Vescovo  
4 et al. 2012 for more details). A similar dataset was collected in 2013 at Monte Bondone by  
5 combining hyperspectral data with chlorophylls measurements. Chlorophylls samples were  
6 collected in the field of view of the hyperspectral sensor and chlorophylls content was detected  
7 by UV-VI spectroscopy. First, the samples were grinded in presence of liquid nitrogen and then  
8 immersed in 80% acetone solution (0.1 g per 10 ml), shaken for 10 min in an automatic shaker at  
9 250 rpm (Universal Table Shaker 709), and centrifuged at 4000 rpm for 10 min (Eppendorf 5810  
10 R) in order to remove particles from the solution. The absorbance of extracted solutions was  
11 measured at 470, 646.8 and 663.2 nm by a UV/VIS spectrophotometer (Shimadzu UV-1601),  
12 and the concentrations of chlorophyll a ( $C_a$ ), chlorophyll b ( $C_b$ ) and carotenoids ( $C_{x+c}$ ) were  
13 calculated as proposed by Lichtenthaler (1987). The weight of sampled sediment was used to  
14 calculate pigments concentrations per unit leaf mass ( $\text{mg g}^{-1}$ ) and the weight of green biomass  
15 per ground area was used to obtain the total chlorophylls content ( $\text{mg m}^{-2}$ ).

## 16 **2.4 CO<sub>2</sub> flux measurements**

17 Continuous measurements of the net ecosystem CO<sub>2</sub> exchange (NEE) were made by the eddy  
18 covariance (EC) technique (Baldocchi et al., 1996; Aubinet et al., 2012) at the six study sites  
19 using identical instrumentation. The three wind components and the speed of sound were  
20 measured using ultra-sonic anemometers, and CO<sub>2</sub> molar densities using open-path infrared gas  
21 analyzers (IRGAs), as detailed in Tables S1 and S2 in the Supplement section. Raw data were  
22 acquired at 20 Hz and averaged over 30 min time windows in post-processing. Turbulent fluxes  
23 were obtained from raw data by applying block averaging (Monte Bondone, Neustift, validation  
24 sites) or linear de-trending (Amplero) methods with a time window of 30 minutes. A 3D  
25 coordinate correction was performed according to Wilczak et al. (2001). The CO<sub>2</sub> fluxes were  
26 corrected for the effect of air density fluctuations as proposed by Webb et al. (1980). Low- and  
27 high-pass filtering was corrected for following Aubinet et al. (2000) (Amplero, Monte Bondone)  
28 or Moore (1986) (Neustift, validation sites). Data gaps due to sensors malfunctioning or violation  
29 of the assumptions underlying the EC method were removed and filled using the gap-filling and

1 flux-partitioning techniques as proposed in Wohlfahrt et al. (2008). Ecosystem respiration (Reco)  
2 was calculated from the y-intercept of the light response model (see eq. 4). Gross primary  
3 productivity (GPP) was calculated as the difference between NEE and Reco. Half-hourly NEE  
4 and GPP values were averaged between 11:00 to 14:00 solar local time (at the time window of  
5 optical measurements) to allow for direct comparison with the hyperspectral data, and daily sums  
6 were also computed. At each site the following supporting environmental measurements were  
7 acquired: photosynthetically active radiation (PAR; quantum sensors), air temperature (Ta;  
8 PT100, thermistor and thermoelement sensors), and humidity (RH; capacitance sensors) at some  
9 reference height above the canopy, and soil temperature (Ts; PT100, thermistor and  
10 thermoelement sensors) and volumetric water content (SWC; dielectric and time-domain  
11 reflectometry sensors) in the main rooting zone. In this study we used CO<sub>2</sub> flux and  
12 meteorological data of the years 2005 and 2006 for Monte Bondone and of 2006 for the other  
13 sites.

## 14 **2.5 Estimation of grassland ecophysiological parameters**

15 Canopy light use efficiency ( $\epsilon$ ) was derived from photosynthetically active radiation (PAR)  
16 absorbed by the canopy (APAR) as:

$$17 \quad \epsilon = \frac{GPP}{APAR} = \frac{GPP}{PAR * fAPAR} \quad (2)$$

18 and was estimated both at midday and daily time resolution. We estimated the fraction of PAR  
19 absorbed by the canopy (fAPAR) from measured values of the leaf area index (LAI) using the  
20 Lambert-Beer law:

$$21 \quad fAPAR = 0.95 \left( 1 - e^{(-k LAI)} \right) \quad (3)$$

22 where k is the canopy extinction coefficient (fixed at k=0.4 as defined for southern mixed-grass  
23 prairie in Texas; Kiniry et al., 2007) and 0.95 is the proportion of intercepted PAR that is  
24 absorbed by plants (Schwalm et al., 2006). LAI was quantified non-destructively by an indirect  
25 method based on canopy PAR transmission using line PAR sensors (SunScan, Delta-T, UK) and  
26 inversion of a RTM (Wohlfahrt et al., 2001). These measurements were done within the footprint  
27 area of the spectroradiometer simultaneously with the hyperspectral measurements.

1 Three additional key parameters of the response of NEE to PAR were extracted by fitting  
2 measured NEE and PAR to a simple Michaelis-Menten-type model:

$$3 \quad NEE = \frac{-\alpha \text{ PAR} F_{\text{sat}}}{\alpha \text{ PAR} + F_{\text{sat}}} + R_{\text{eco}} \quad (4)$$

4 where  $\alpha$  represents the apparent quantum yield ( $\mu\text{mol CO}_2 \mu\text{mol photons}^{-1}$ ),  $F_{\text{sat}}$  the asymptotic  
5 value of GPP ( $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ ), PAR the photosynthetically active radiation ( $\mu\text{mol photons}$   
6  $\text{m}^{-2} \text{ s}^{-1}$ ) and  $R_{\text{eco}}$  the ecosystem respiration ( $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ ). For all sites, using the  
7 Levenberg-Marquardt (1963) algorithm the parameters of Eq. (4) were estimated by fitting Eq.  
8 (4) to both day and nighttime data, which were pooled into 3-day blocks centered on the date of  
9 the hyperspectral data acquisition. For each acquisition date, we then used Eq. (3) to derive GPP  
10 at an incident PAR of  $1500 \mu\text{mol m}^{-2} \text{ s}^{-1}$ , referred to as  $\text{GPP}_{\text{max}}$  in the following.

## 11 **2.6 Hyperspectral data analysis**

12 In order to explore the information content of the hyperspectral data for estimating  $\text{CO}_2$  fluxes  
13 (i.e. midday/daily average of NEE and GPP) and ecophysiological parameters (i.e.  $\alpha$ ,  $\varepsilon$  and  
14  $\text{GPP}_{\text{max}}$ ), we performed a correlation analysis between spectral reflectance indices (independent  
15 variables) and these (dependent) variables. To this end, we derived spectral ratio (SR; Eq. (5)),  
16 spectral difference (SD; Eq. (6)) and normalized spectral difference (NSD; Eq. (6)) indices using  
17 all possible two-band (i and j) reflectance ( $\rho$ ) combinations between 400 and 1000 nm (180600  
18 combinations). These three formulations were selected since they represent the most common  
19 equations used to compute vegetation indices (see Table 2).

$$20 \quad SR_{i,j} = \frac{\rho_i}{\rho_j} \quad (5)$$

$$21 \quad SD_{i,j} = \rho_i - \rho_j \quad (6)$$

$$22 \quad NSD_{i,j} = \frac{\rho_i - \rho_j}{\rho_i + \rho_j} \quad (7)$$

23 Linear regression analysis was performed among all possible wavelength-combinations for all  
24 three index-types (SR, NSD and SD) and the investigated dependent variables.

25 The performance of linear models in predicting dependent variables (i.e. carbon fluxes and  
26 ecophysiological parameters) was evaluated by the coefficient of determination ( $R^2$ ) and root

1 mean square error (RMSE). The coefficients of determination ( $R^2$ ) resulting from the linear  
2 models were visualized in correlograms as depicted in an exemplary fashion in Figure 1.

3 We also calculated four SR- and seven NSD-indices which are commonly used in relation to  
4 vegetation activity and  $CO_2$  fluxes (Table 2). Figure 1 shows the location of these indices in the  
5 waveband space of the correlograms. In this analysis, we also considered the Enhanced  
6 Vegetation Index (EVI), which is one of the most frequently used vegetation index to predict  
7  $CO_2$  fluxes. In the Fig. 1 the location of EVI is not shown since this index is computed by the  
8 combination of three spectral bands as shown in Table 2.

9 The robustness of the model selected on the basis of the best band combinations for all  
10 ecophysiological parameters for each site and all sites pooled was tested by the leave-one-out  
11 cross-validation technique. The predictive performance was expressed as the cross-validated  
12 coefficient of determination ( $R^2_{CV}$ ) and the cross-validated root mean square error ( $RMSE_{CV}$ ). In  
13 addition, the capability of the selected models in predicting different ecophysiological  
14 parameters was tested by applying the selected models to the validation dataset (Table S2)  
15 composed by three different grasslands not used in the previous analysis. This dataset was  
16 selected because the hyperspectral and flux data were collected by using exactly the same  
17 protocol applied for the main dataset (see sect.2.1).

18 In order to explore the basis of the correlation between the selected band combinations and  
19 ecophysiological variables (e.g.  $\alpha$ ,  $GPP_{max}$ ,  $GPP$ ,  $\varepsilon$ ) the relationship between the selected bands  
20 and biophysical parameters such as dry phytomass, nitrogen and water content collected during  
21 the field campaign in the same footprint of the hyperspectral measurements was examined.

## 22 **2.7 Band selection based on the combination of random forests and genetic** 23 **algorithm (GA-rF)**

24 In order to complement the more conventional analysis described in the previous section, we also  
25 explored the use of a hybrid feature selection strategy based on a genetic algorithm and random  
26 forests (GA-rF). The first method was used for the feature selection and the second one as  
27 regression for predicting the target variables. First of all, the original dataset was aggregated to  
28 10 nm bands in order to reduce the effects of autocorrelation in frequency space. The algorithm  
29 generates a number of possible model solutions (chromosomes) and uses these to evolve towards

1 an approximation of the best solution of the model. In our case the genes of each chromosome  
2 correspond to the wavebands. We made use of 5 genes for each chromosome in order to  
3 overcome overfitting. Each population of 1000 chromosomes evolved for 200 generations. The  
4 mutation chance was set to the inverse of population size increased by one. The fitness of each  
5 chromosome was measured by applying the random forest algorithm (Breiman, 2001). This was  
6 used as an ensemble method for regression that is based on the uncontrolled development of  
7 decision trees (n=100). We opted for this method because of its demonstrated efficiency with  
8 large datasets. In combining the two methods we choose the mean squared error as the target  
9 variable to be minimized.

10

### 11 **3 Results**

#### 12 **3.1 Seasonal variation of meteorological variables, LAI and CO<sub>2</sub> fluxes**

13 Environmental conditions and the seasonal development of LAI, NEE, GPP  $\alpha$ ,  $\varepsilon$  and GPP<sub>max</sub>  
14 during the study period are shown in Figure 2. A strong influence of the typical climatic  
15 conditions at the three study sites ~~is~~ was evident: Amplero was characterized by a Mediterranean  
16 climate, with highest incoming radiation and temperatures, and the lowest amount of  
17 precipitation which translated into a substantial seasonal drawdown of soil moisture~~;~~. Monte  
18 Bondone and Neustift, more influenced by continental Alpine climate, experienced comparably  
19 lower temperatures with higher precipitation and soil moisture with respect to Amplero (Fig. 2).

20 Maximum LAI values were similar at Monte Bondone and Amplero (2.8-3.4 m<sup>2</sup> m<sup>-2</sup>), while,  
21 twice as much leaf area developed at the more intensively managed study site Neustift~~,-~~. ~~which~~  
22 ~~The latter~~ was also characterized by higher NEE and GPP (i.e. more photosynthesis and net  
23 uptake of CO<sub>2</sub>). The reductions in leaf area associated with the cuts of the grasslands were  
24 associated as expected with marked increases and reductions in NEE and GPP, respectively. The  
25 canopy light use efficiency,  $\varepsilon$ , was inversely related to GPP and LAI, peaking at the beginning of  
26 the season at Amplero and Monte Bondone (0.01-0.10  $\mu\text{mol photons } \mu\text{mol CO}_2^{-1}$ ), while for  
27 Neustift  $\varepsilon$  showed the highest values after the cuts (0.01-0.20  $\mu\text{mol photons } \mu\text{mol CO}_2^{-1}$ ). At  
28 Amplero,  $\alpha$  and GPP<sub>max</sub> peaked in spring and then decreased during the summer drought period,

1 while at Neustift and Monte Bondone, temporal patterns of  $\alpha$  and  $GPP_{max}$  were more strongly  
2 affected by management.

### 3 **3.2 Hyperspectral data and their relation to CO<sub>2</sub> fluxes and ecophysiological** 4 **parameters**

5 ~~Figure 3 reports key spectral signatures of the grasslands collected during the study period. The~~  
6 ~~reflectances in the NIR region decreased (NIR; 700–1000 nm) and increased in the blue region~~  
7 ~~(420–540 nm) from early to late spring until the harvest for the Mediterranean grassland of~~  
8 ~~Amplero (Fig. 3a) (Balzarolo, 2008). This is a typical trend for Mediterranean grasslands~~  
9 ~~characterized by leaf senescence due to drought conditions (Fava et al., 2007; Vescovo et al.~~  
10 ~~2012). For Monte Bondone in 2006 and Neustift (Fig. 3b, d) the reflectance in the green (540–~~  
11 ~~580 nm) and NIR region increased and decreased in the visible region with increasing LAI and~~  
12 ~~phytomass.~~

13 Figures 3-5 show correlograms between NSD-, SR- and SD-type indices, respectively, and the  
14 investigated dependent midday ecophysiological parameters and fluxes. The correlograms for  
15 daily data can be found in the Supplement (Figures S2-S4). ~~Selected examples of key spectral~~  
16 ~~signatures of the investigated grasslands are shown in Figure S1 in the Supplementary material.~~

17  
18 A number of interesting insights may be gained from Figures 3-5 and Figures S2-S4, which we  
19 summarize in the following:

- 20 (i) The correlograms exhibited quite different patterns – some correlograms showed that a  
21 wide range of band combinations was able to explain the simulated quantities (e.g. GPP at  
22 Amplero; Fig. 3; Fig. S2), while some correlograms exhibited very pronounced patterns,  
23 with the  $R^2$  value changing greatly with subtle changes in band combinations (e.g.  $\epsilon$  at  
24 Neustift; Fig. 3; Fig. S2).
- 25 (ii) Maximum  $R^2$  values were often clearly higher than the surrounding areas of high  
26 predictive power (e.g.  $\epsilon$  at Amplero; Fig. 3).
- 27 (iii) The different types of indices (compare Figs. 3-5) yielded similarly high correlations with  
28 the same dependent variable at the same site in similar spectral regions, ~~–. This indicatesing~~  
29 that band selection is more important for explanatory power than the mathematical

1 formulation of the VI (i.e. ratio vs. difference, with/without normalization). SR and NSD  
2 indices (Figs. 3 and 4) yielded similar results compared to SD indices (Fig. 5).

3 (iv) The highest correlations for all dependent variables were found either for indices  
4 combining bands in the visible range (VIS: <700 nm) or the red edge and NIR (NIR: >700  
5 nm), corresponding to spectral regions used by indices such as the SRPI, NPCI, PRI and  
6 NPQI, and the CI and WI, respectively. Spectral regions of well-known indices, such as  
7 NDVI, SR, SIPI or GRI, which exploit the contrasting reflectance magnitudes in the visible  
8 and NIR (Fig. S1), resulted in comparably lower correlations.

9 (v) For midday and daily time resolutions different band combinations were selected (e.g.  
10 NEE at Amplero; compare Figs. 4 and S3). For similar selected regions, daily averages  
11 were characterized by higher explanatory power compared to midday averages (e.g.  $\epsilon$  at  
12 Neustift).

13  
14 Figure 6 shows the performance of linear regression models for the best NSD-type indices for  
15 midday ecophysiological parameters for each site and all sites pooled (Figure S7 in the  
16 Supplement shows the results of the same analysis for daily averages). Large differences existed  
17 between the study sites in the explanatory power of the same index for the same dependent  
18 variable. The highest  $R^2_{cv}$  and values were generally obtained for Amplero, followed by Neustift  
19 and then Monte Bondone and the lowest  $R^2_{cv}$  values resulted when data from all three sites were  
20 pooled, confirming the difficulties in finding a general relation valid among sites.

21 For Amplero and Neustift the NIR vs. NIR combinations showed a positive linear regression  
22 model with  $\alpha$ ,  $GPP_{max}$  and GPP, while for Monte Bondone a negative linear correlation was  
23 observed. For Amplero the VIS vs. VIS combination showed a good performance in predicting  
24  $\epsilon$ ; the NIR vs. NIR combinations showed good performance for Neustift and VIS vs. NIR  
25 combination for Monte Bondone. The linear models for NEE were site-specific. In fact, Amplero  
26 and Monte Bondone showed a positive linear regression model for NEE but the VIS vs. VIS  
27 band combination was selected for Amplero and NIR vs. NIR combination for Monte Bondone.  
28 Neustift performed well with NEE for NIR vs. NIR combinations, but with an inverse  
29 relationship.

1 The different type of indices (compare Figs. 6, S5 and S6) resulted in similar models. The  
2 different time resolutions gave different models (e.g. GPP,  $\epsilon$  and NEE at Monte Bondone,  
3 compare Figs. 6 and S7 or Figs. S5 and S8 or Figs. S6 and S9).

### 4 **3.3 Correlation between conventional VIs, ecophysiological variables and CO<sub>2</sub>** 5 **fluxes**

6 The correlation analysis between the conventional VIs, the midday CO<sub>2</sub> fluxes (Table 3) and  
7 ecophysiological parameters (Table 4), generally confirmed the results obtained with the  
8 hyperspectral data.

9 For the same dependent variable ( $\alpha$ , GPP<sub>max</sub>, GPP,  $\epsilon$  and NEE), the performance of the various  
10 VIs showed large differences between sites. For example, for GPP<sub>max</sub> all of the investigated  
11 indices except NPQI resulted in significant linear correlations at Amplero, explaining 41-89% of  
12 the variability in GPP<sub>max</sub>. In contrast, only NDVI, PRI, NPCI and SRPI showed a slightly  
13 significant linear performance (17-26%) for GPP<sub>max</sub> at Neustift.

14 The different VIs performed differently in predicting the same dependent variable at the different  
15 study sites. For all dependent variables (Tables 3, 4 and S3), the VI resulting in the highest R<sup>2</sup>  
16 values was never the same at all sites. Often the best fitting VI at one site resulted in a non-  
17 significant correlation at another site. Therefore, none of the dependent variables clearly emerged  
18 as the one best predicted (Tables 3, 4 and S3).

19 When data from all sites were pooled, models showed the same performance for the same VI and  
20 dependent variable except for GPP and NEE. The best performing VI for GPP and NEE was  
21 SIPI, NPCI performed best for  $\alpha$ , GRI for  $\epsilon$ , SIPI for GPP<sub>max</sub>.

22 The choice of the averaging period (midday *vs.* daily) applied to  $\epsilon$ , NEE and GPP did generally  
23 not modify the ranking of the VIs, but the R<sup>2</sup> values tended to be similar or somewhat higher at  
24 the daily time scale (compare Tables 3 and 4 with Table S3).

### 25 **3.4 Evaluation of the model performance**

26 [Figure 7](#) shows the results of the validation for each ecophysiological parameter and midday  
27 averaged fluxes and NSD-, SR- and SD-type indices against data from the validation sites. The



1 models used in the validation are based on the best models determined for each site (i.e.  
2 Amplero, Neustift and Monte Bondone) and by pooling together the two alpine grasslands of  
3 Monte Bondone and Neustift (referred to as M.Bondone&Neustift).

4 Overall, the results of the validation were mixed. Good performance was observed mainly for the  
5 Neustift, Monte Bondone and pooled M.Bondone&Neustift models (see Table S4 in the  
6 Supplement section). In particular, the best performance values were obtained for: (i)  $\alpha$  for the  
7 pooled Monte Bondone and Neustift model for SR-type indices; (ii)  $GPP_{max}$  for NSD-, SR- and  
8 SD-type indices and models except for Amplero; and (iii) for midday GPP for all NSD-, SR-,  
9 and SD-type indices and models. It is interesting to note that lower performances were generally  
10 found for the models based on the Amplero parameterization. This is understandable as Monte  
11 Bondone and in particular Neustift were structurally and functionally much more similar to the  
12 validation sites compared to Amplero (Tables 1 and S2). Considerably poorer performance was  
13 observed for  $\epsilon$  and NEE across all model-index type combinations (Table S4). The validation at  
14 daily time scale always resulted in a poorer performance compared to the midday average time  
15 scale (Fig. S10 and Table S5 in the Supplement section).

### 16 **3.5 Effects of canopy structure on selected band combinations**

17 Tables 5 and S6 show the results of the correlation analysis between the selected models for  
18 ecophysiological variables and fluxes and biophysical properties of vegetation such as dry  
19 phytomass, nitrogen and water content. Overall, the spectral response in the selected band  
20 combinations for NSD, SR and SD-type indices was strongly related to vegetation properties of  
21 the three grasslands (e.g. nitrogen and dry phytomass) which impacted on their spectral response  
22 in the NIR and VIS regions. For the Mediterranean site (Amplero) and for all eco-physiological  
23 parameters (i.e.  $\alpha$ ,  $GPP_{max}$ , GPP,  $\epsilon$ ), dry phytomass was the main driving factor of the spectral  
24 response in the selected bands, while nitrogen content drove the spectral response in the NIR  
25 region for Neustift. For Monte Bondone, both dry phytomass and nitrogen content affected the  
26 spectral response of the grassland. Similar results were obtained for SR- and SD-type indices.

27 Figs. 8 and S11 in the Supplement show the correlation analysis between the selected bands for  
28 NSD-, SR- and SD-type indices and chlorophylls content for Monte Bondone in 2013. The  
29 chlorophylls content showed a very good correlation for all selected models and for all indices.

1 The values of  $R^2$  were always higher than the values of  $R^2$  obtained for the other biophysical  
2 variables (Tables 5 and S6). In Figure 9, it is possible to see that, NSD- and SR-type indices for  
3 the selected bands for estimating GPP (i.e. 996 nm and 710 nm) are strongly correlated with  
4 canopy total chlorophyll content ( $R^2 > 0.80$ ).

### 5 **3.6 Band selection using GA-rF method**

6 Figure 9 shows the results of the band selection based on GA-rF method. In particular, each plot  
7 represents the frequency of the occurrence of each band in the genetic algorithm.

8 Overall, using the GA-rF method it was possible to identify portions of the spectrum that were of  
9 particular significance for estimating specific properties of the different ecosystems. For  
10 example, for predicting midday GPP (Fig. 9b) for all sites pooled together, the bands at 430 nm,  
11 630 nm, 660 nm and 710 nm showed the best results. The bands at 505 nm played an important  
12 role in predicting midday GPP for Amplero; bands at 660 nm for Neustift and bands 710 nm for  
13 Monte Bondone. Some differences were found for the different time resolutions (compare Figs.  
14 9b and 9c). For example, the bands at 580 nm and 800 nm showed the best results for Amplero  
15 and bands at 530 nm for Neustift.

16 Figure 10 shows the results for the band selection by GA-rF methods for biophysical variables  
17 (i.e. dry phytomass, nitrogen and water content). For the variables related to slow processes the  
18 GA-rF method highlighted different bands for different sites; a much higher between site  
19 variability for the variables related to ecophysiological processes (e.g.  $\epsilon$ ,  $\alpha$  and  $GPP_{max}$ ) was  
20 detected and we weren't able to identify common "hot spots".

21

## 22 **4 Discussion**

23 This study aimed at evaluating the potential of hyperspectral reflectance measurements to  
24 simulate  $CO_2$  fluxes and ecophysiological variables of European mountain grasslands over a  
25 range of climatic conditions and management practices (grazing, harvest). To this end, we  
26 combined eddy covariance  $CO_2$  flux measurements with ground-based hyperspectral  
27 measurements at six mountain grassland sites in Europe.

28

1 *Up-scaling of in-situ relationships between VI indices and CO<sub>2</sub> fluxes and ecophysiological*  
2 *parameters*

3 Despite the fact that we focused on a single type of ecosystem, our results showed that large  
4 differences existed among the investigated sites in the relationships between hyperspectral  
5 reflectance data and CO<sub>2</sub> fluxes and ecophysiological parameters. For all study sites pooled,  
6 hyperspectral reflectance data explained 40-68% of the variability in the dependent variables  
7 (Figs. 3-5). The conventional VIs yielded a maximum of 47% of explained variability in the data  
8 (Tables 3-4).

9 This is the first study comparing different grasslands characterized by different plant species and  
10 environmental conditions. The use of simple models based on a linear relationship between GPP  
11 and VIs, related to canopy greenness, has proven to be a good proxy for GPP of ecosystems with  
12 strong green-up and senescence (Peng et al., 2011; Rossini et al., 2012). The loss of this  
13 relationship may be related to low  $\epsilon$  variability due to abiotic and biotic stressors, the  
14 dependency of PRI on LAI, leaf and canopy biochemical structure (e.g. leaf orientation), and  
15 xanthophyll cycle inhibition or saturation and zeaxanthin-independent quenching (Gamon et al.,  
16 2001; Filella et al., 2004; Rahimzadeh-Bajgiran et al. 2012; Hmimina et al., 2014). For alpine  
17 grasslands, a key meteorological variable that played a relevant role in stimulating  $\epsilon$  was high  
18 soil water content associated with low temperatures (Polley et al. 2011). Low soil water contents  
19 triggered a decrease in leaf conductance as well as in  $\epsilon$  and in  $\alpha$  also for two oak and beech  
20 ecosystems (Hmimina et al., 2014). However, no significant differences in leaf biochemical and  
21 structural properties of the canopy at lowest and highest water content were found. In addition, in  
22 this special issue, Sakowska et al. (2014) showed that  $\epsilon$  is also strongly affected by the  
23 directional distribution of incident PAR, i.e. the ratio of direct to diffuse PAR.

24 Considering all sites pooled together (Figs. 3 and S2), NSD-type indices showed a very poor  
25 correlation in the VIS vs. NIR band combinations (i.e. traditional "greenness" indices, see Table  
26 2) with GPP. It is well-known in the literature (Rossini et al., 2010, 2012; Peng et al., 2010;  
27 Sakowska et al., 2014) that "greenness" indices, for grasslands and crops, are often good proxies  
28 of fAPARgreen (and thus carbon fluxes). Interestingly, in our study their performance was  
29 considerably poorer than expected. The NSD-type index showed a better performance in VIS vs.  
30 VIS band combinations than VIS vs. NIR ones. VIS vs. VIS band combination for NSD-type

1 indices (e.g. green vs. blue or red, green vs. green wavelengths; see e.g. Inoue et al, 2008) are  
2 defined as “greenness” indices (Fig. 1), although their performance is generally much poorer  
3 than NSD VIS vs. NIR indices. These results are likely due to the confounding effects of the  
4 different canopy structures, and consequently of the different NIR response of the investigated  
5 grasslands (see Fig. S1). In fact, the different grassland structures (spatial distribution of  
6 photosynthetic, and also non photosynthetic material, leaf angles, etc.) is affecting our ability to  
7 use traditional indices to estimate fAPARgreen (and fluxes) when we consider different  
8 grasslands together because the structural effects on scattering are very complex in the NIR  
9 (Jacquemoud et al., 2009; Knyazikhin et al., 2012). These results are of importance for the  
10 community, which still relies a lot on these relationships, also favoured by the availability of  
11 affordable narrow-band sensors that allow continuous monitoring of e.g. NDVI. These results  
12 suggest that waveband combinations not exploited by presently used (conventional) VIs may  
13 offer considerable potential for predicting grassland CO<sub>2</sub> fluxes, which has implications for the  
14 design and capabilities of future space/airborne or ground-based low cost sensors. In particular,  
15 these results also have a strong impact on our ability to up-scale grassland fAPARgreen and  
16 carbon fluxes using upcoming sensors (e.g. Sentinel 2).

17 The evaluation of the models found for the main dataset against three new sites ~~confirm~~ showed  
18 that ~~at least some of~~ these models can be transferred to predict carbon fluxes and  
19 ecophysiological parameters for similar grasslands (Fig. 7). However, ~~for some parameters (e.g.~~  
20 ~~ε), the independent validation indicated a poor performance, it ~~these findings also~~ challenges~~ the  
21 current practice in up-scaling to larger regions by grouping all grasslands into a single plant  
22 functional type (PFT). We advocate more studies to be conducted merging CO<sub>2</sub> flux with  
23 hyperspectral data by means of models which use a more process-oriented and coupled approach  
24 to simulating canopy CO<sub>2</sub> exchange and reflectance in order to explore the causes underlying the  
25 observed differences between seemingly closely related study sites.

26

### 27 *Grassland structural characteristic and their spectral response*

28 Although we considered similar ecosystems (belonging to the same vegetation type) the  
29 investigated canopies were very different and included Mediterranean, extensive alpine and  
30 intensive alpine grasslands with very different canopy structures in terms of leaf orientation,

1 amount and spatial distribution of green and non-photosynthetic components, leaf nitrogen and  
2 water content as detailed in Vescovo et al. (2012).

3 For Amplero and Neustift NSD-type indices performed well for NIR vs. NIR band combinations  
4 for all investigated parameters, while Monte Bondone showed best performances in the VIS vs.  
5 NIR band combinations for  $GPP_{max}$  and  $\epsilon$  (Fig. 6). The dry phytomass was the main driving  
6 factor of the spectral response in NIR vs. NIR band combinations for Amplero, while nitrogen  
7 content drove the spectral response in NIR vs. NIR band combinations of Neustift for all  
8 parameter except for  $\alpha$  (Tab. 5 and S6). Interestingly, for Monte Bondone both dry phytomass  
9 and nitrogen content explained the spectral response of the grassland in VIS vs. NIR band  
10 combinations for  $GPP_{max}$  and  $\epsilon$  while no significant relationships with biophysical variables were  
11 found for  $\alpha$ , GPP, NEE. These results partially confirm the findings of Vescovo et al. (2012),  
12 who highlighted a strong relationship, for several grassland types, between an NSD-type index  
13 and phytomass.

14 For Monte Bondone, NSD- and SR-type indices for the selected bands for estimating all  
15 variables except  $\alpha$  were strongly correlated with canopy total chlorophyll content ( $R^2 > 0.85$ ).

16 The chlorophyll indices (e.g. RedEdge NDVI and CI; see Tables 3 and 4) – which are considered  
17 the best indices for estimating carbon fluxes on grasslands and crops) – showed in our dataset a  
18 good performance for Amplero and Monte Bondone, but performed poorly for Neustift.

19 It was demonstrated by many authors that the red edge domain, where reflectance changes from  
20 very low in the absorption region to high in the NIR, is one of the best descriptors of chlorophyll  
21 concentration. On the other hand, it is well known that the canopy structure can be a very strong  
22 confounding factor. Our results confirm that this topic needs to be further investigated, as this  
23 finding has a relevant impact concerning the use of Sentinel 2 to upscale fAPAR and carbon flux  
24 observations.

25 It is interesting to see that the NSD-type indices in the NIR vs. NIR band combinations appeared  
26 to be the best proxy for GPP fluxes when all the grasslands were pooled together. These results  
27 can be linked to the controversial paper focused on the strong impact of structure on the ability to  
28 estimate canopy nitrogen content (Knyazikhin et al., 2012) and confirm the need for more  
29 studies in this direction. Good relationships were found between the NIR vs. NIR band

1 combinations (>750nm wavelengths) and fluxes; the physical basis of these relationships needs  
2 to be further investigated. In fact, it is important to highlight that the literature indicates that the  
3 wavelengths in the NIR (>750nm) are not sensitive to chlorophyll content, but they are related to  
4 leaf, canopy structure, and -around the 970nm area- to water.

5 As confirmed by comparing the correlation matrix approach with the GA-rF approach we  
6 couldn't find a universal relationship between reflectance in specific wavelengths of the light  
7 spectrum to biophysical properties of vegetation. We think that this is strongly linked to  
8 vegetation structure effects. For this reason we believe that further research for disentangling the  
9 impact of factors like bidirectional reflectance distribution function and scaling effects is  
10 necessary.

11

## 12 **5 Conclusions**

13 The present study focused on understanding the potential of hyperspectral VIs in predicting  
14 grassland CO<sub>2</sub> exchange and ecophysiological parameters ( $\alpha$ ,  $\varepsilon$  and GPP<sub>max</sub>) for different  
15 European mountain grasslands.

16 The major finding of this study is that the relationship between ground-based hyperspectral  
17 reflectance and the ecosystem-scale CO<sub>2</sub> exchange of mountain grasslands is much more variable  
18 than what might be supposed given this closely related group of structurally and functionally  
19 similar ecosystems. As a consequence, the unique models of mountain grassland CO<sub>2</sub> exchange,  
20 i.e. the best fitting models for all sites pooled, explained 47% and 68% of the variability in the  
21 independent variables when established VIs and optimized hyperspectral VIs, respectively, were  
22 used. Interestingly, VIs based on reflectance either in the visible or NIR part of the  
23 electromagnetic spectrum were superior in predicting mountain grassland CO<sub>2</sub> exchange and  
24 ecophysiological parameters compared to commonly used VIs which are based on a combination  
25 of these two wavebands. The band selection based on GA-rF algorithm confirmed that is difficult  
26 to define a universal band range able to describe ecophysiological parameters, carbon fluxes and  
27 biophysical variables even for a closely related group of ecosystems.

28 The take-home message from this study thus is that continuing efforts are required to better  
29 understand differences in the relationship between ecosystem-scale reflectance and CO<sub>2</sub>

1 exchange and to improve models of this relationship which can be employed to up-scale the land  
2 CO<sub>2</sub> exchange to larger spatial scales based on optical remote sensing data. Initiatives such as  
3 SpecNet (<http://specnet.info>; Gamon et al., 2006), the COST Action ES0903 (EUROSPEC;  
4 <http://cost-es0903.fem-environment.eu/>) and the COST Action ES1309 (OPTIMISE;  
5 [http://www.cost.eu/domains\\_actions/essem/Actions/ES1309](http://www.cost.eu/domains_actions/essem/Actions/ES1309)) are instrumental to this end as they  
6 provided the scale-consistent combination of hyperspectral reflectance and CO<sub>2</sub> exchange data.

7

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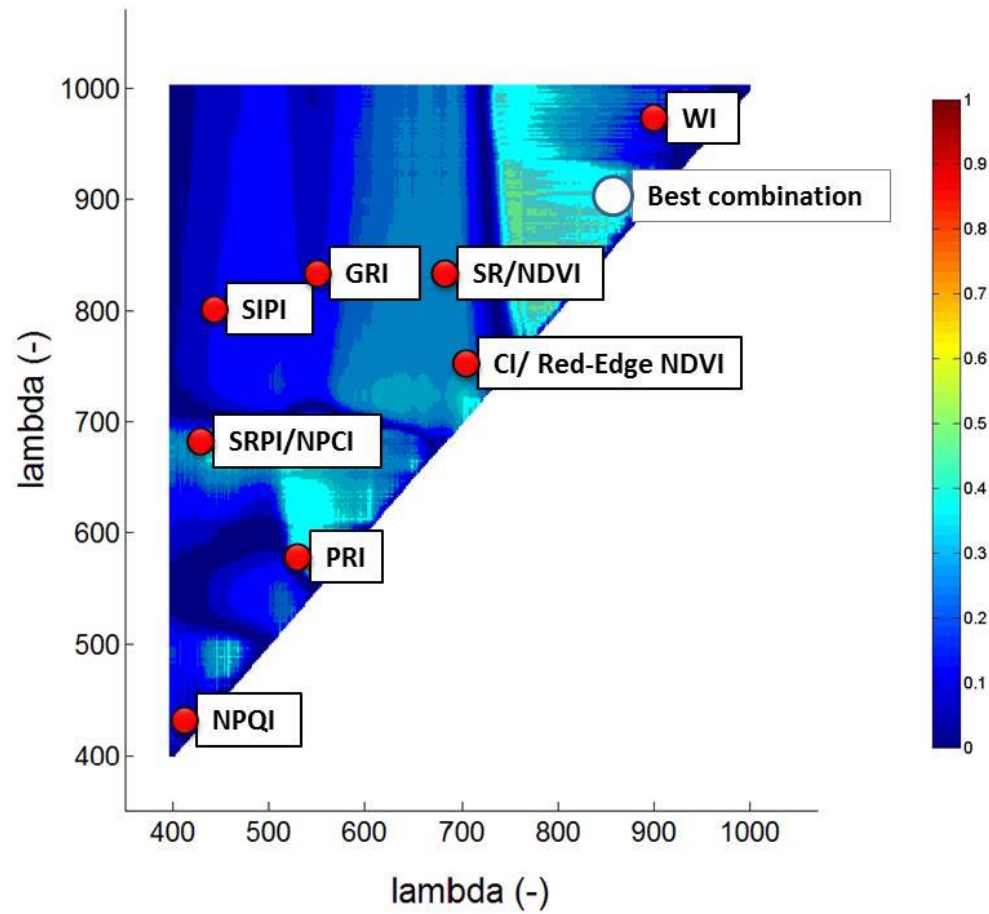


Figure 1. A selected example of a correlogram between NSD-type indices and midday average GPP for all sites pooled. The correlogram shows all  $R^2$  values, the white dot indicates the two-band combination with the highest  $R^2$  value and the red dots indicate the location of the reference VIs reported in Table 2 (SR: Simple ratio; GRI: Green Ratio Index; WI: Water Index; SRPI: Simple Ratio Pigment Index; NDVI: Normalized Difference Vegetation Index; NPQI: Normalized Phaeophytinization Index; NPCI: Normalized Pigment Chlorophyll Index; CI: Chlorophyll Index; Red Edge NDVI; SIPI: Structural Independent Pigment Index; PRI: Photochemical Reflectance Index).

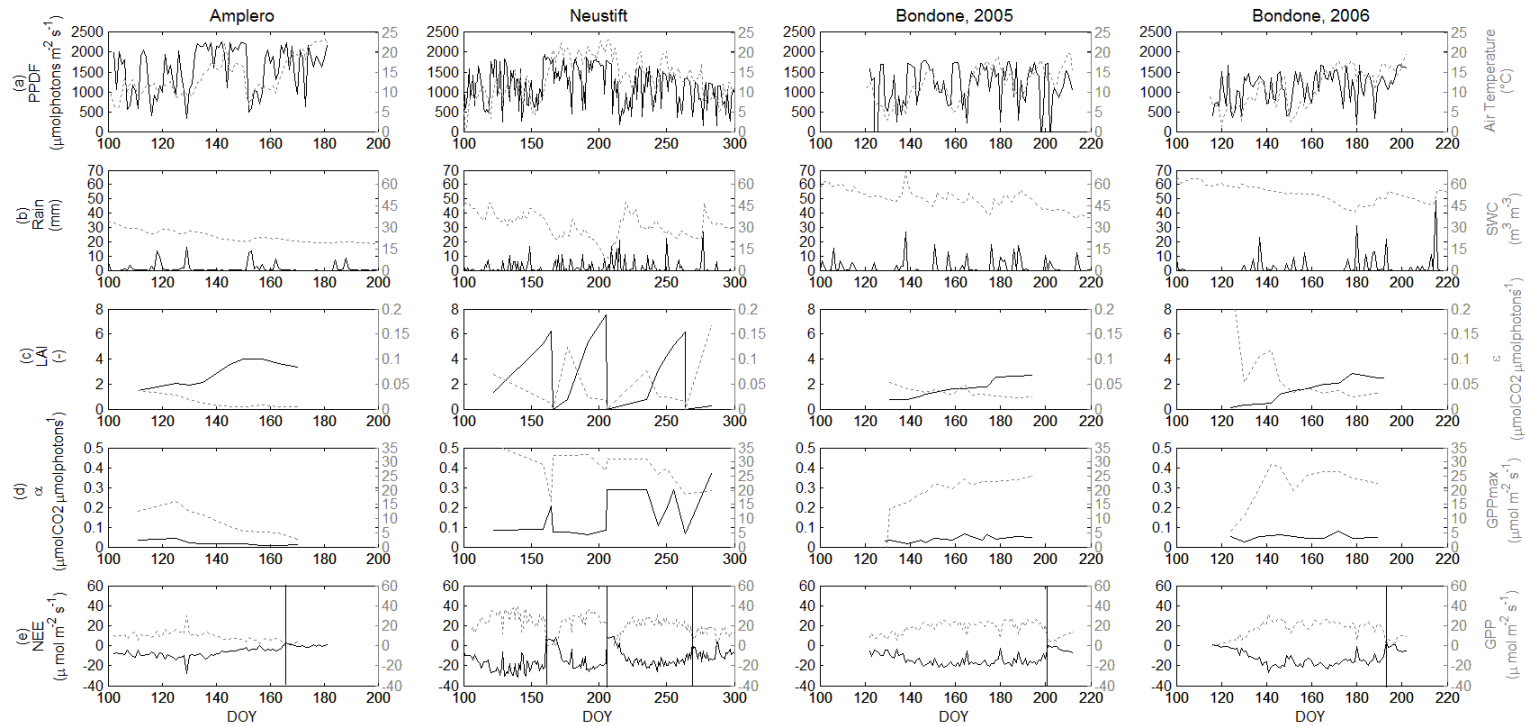


Figure 2. Seasonal variation of meteorological variables, LAI, CO<sub>2</sub> fluxes and ecophysiological parameters for the period of the hyperspectral measurements at the three investigated grasslands. (a) midday average photosynthetically active radiation (PAR;  $\mu\text{mol m}^{-2} \text{s}^{-1}$ ; solid black line) and daily average air temperature ( $^{\circ}\text{C}$ ; dotted grey line); (b) daily precipitation (Rain; mm; solid black line) and daily average soil water content (SWC;  $\text{m}^3 \text{m}^{-3}$ ; dotted grey line); (c) Leaf Area Index (LAI;  $\text{m}^2 \text{m}^{-2}$ ; solid black line) and light use efficiency ( $\epsilon$ ;  $\mu\text{mol photons } \mu\text{mol CO}_2^{-1}$ ; dotted grey line); (d) apparent quantum yield ( $\alpha$ ;  $\mu\text{mol CO}_2 \mu\text{mol photons}^{-1}$ ; solid black line) and gross primary production at saturating light ( $\text{GPP}_{\text{max}}$ ;  $\mu\text{mol m}^{-2} \text{s}^{-1}$ ; dotted grey line); (e) midday average net ecosystem CO<sub>2</sub> exchange (NEE;  $\mu\text{mol m}^{-2} \text{s}^{-1}$ ; solid black line) and gross primary production (GPP;  $\mu\text{mol m}^{-2} \text{s}^{-1}$ ; grey dotted line); vertical lines in the lowermost panels indicate the dates of mowing.



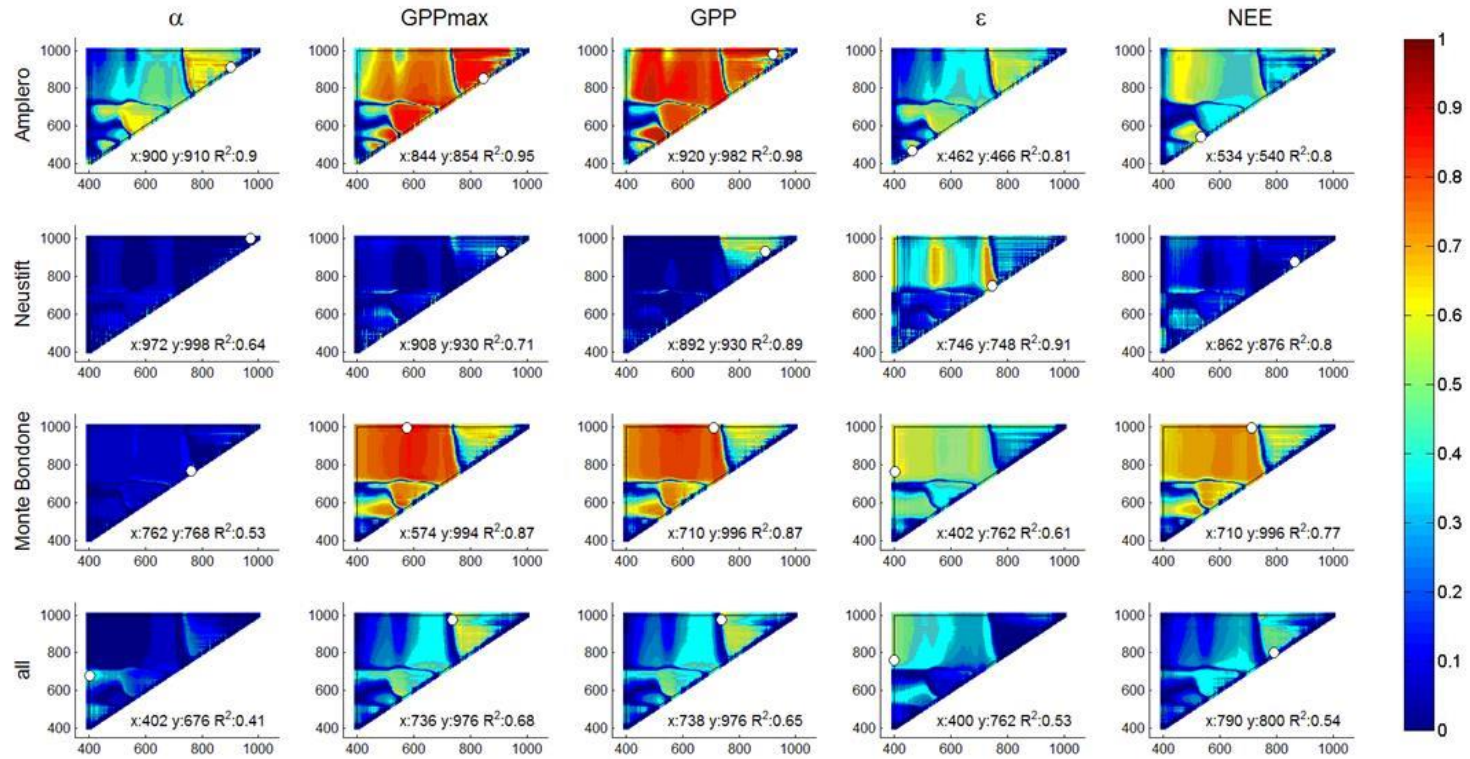


Figure 3. Correlograms of  $R^2$  values for  $\alpha$ , GPP<sub>max</sub> and midday averaged GPP,  $\epsilon$  and NEE and NSD-type indices for Amplero, Neustift, Monte Bondone (both study years pooled) and all sites pooled. The white dots indicate the position of paired band combinations corresponding to the maximum  $R^2$ .

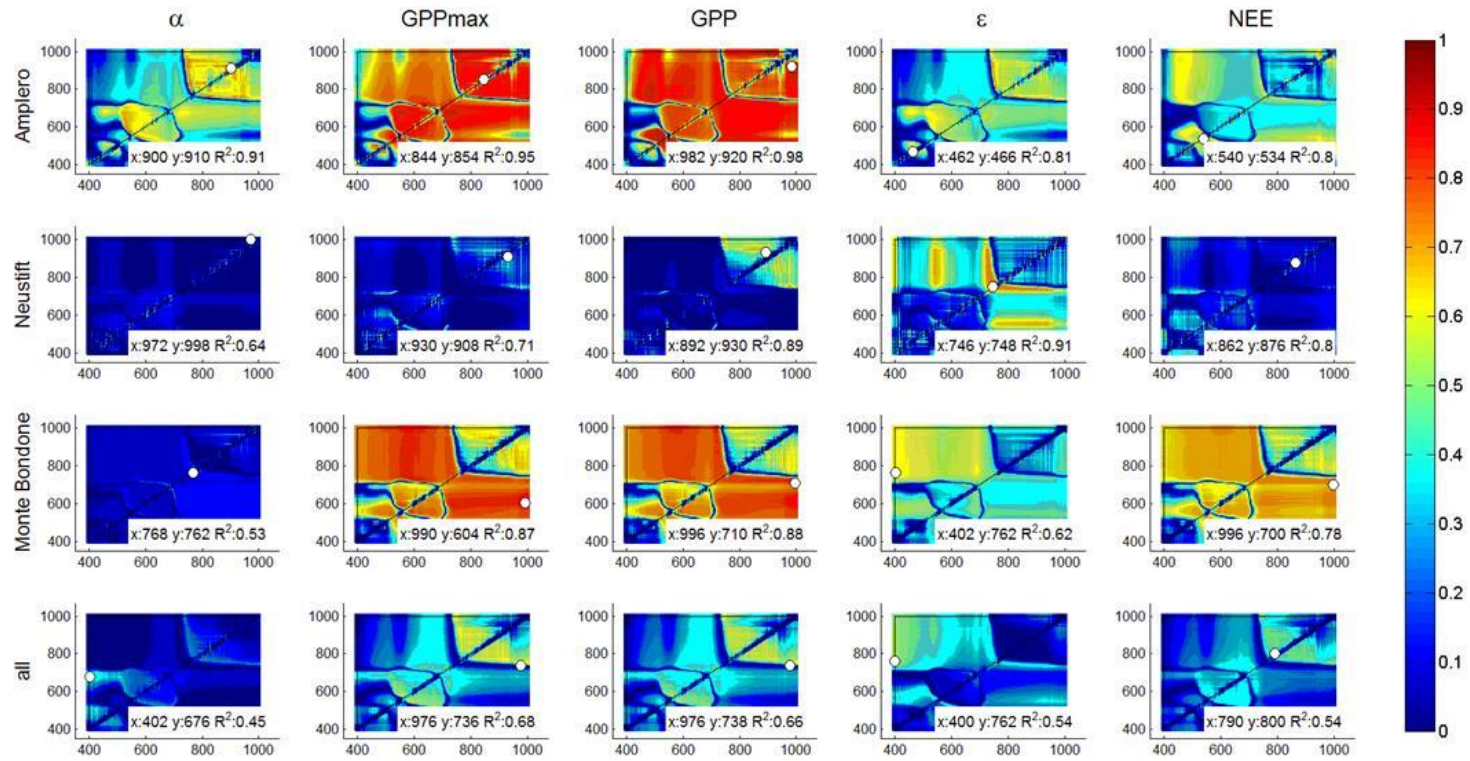


Figure 4. Correlograms of  $R^2$  values for  $\alpha$ ,  $GPP_{max}$  and midday averaged GPP,  $\epsilon$  and NEE and SR-type indices for Amplero, Neustift, Monte Bondone (both study years pooled) and all sites pooled. The white dots indicate the position of paired band combinations corresponding to the maximum  $R^2$ .

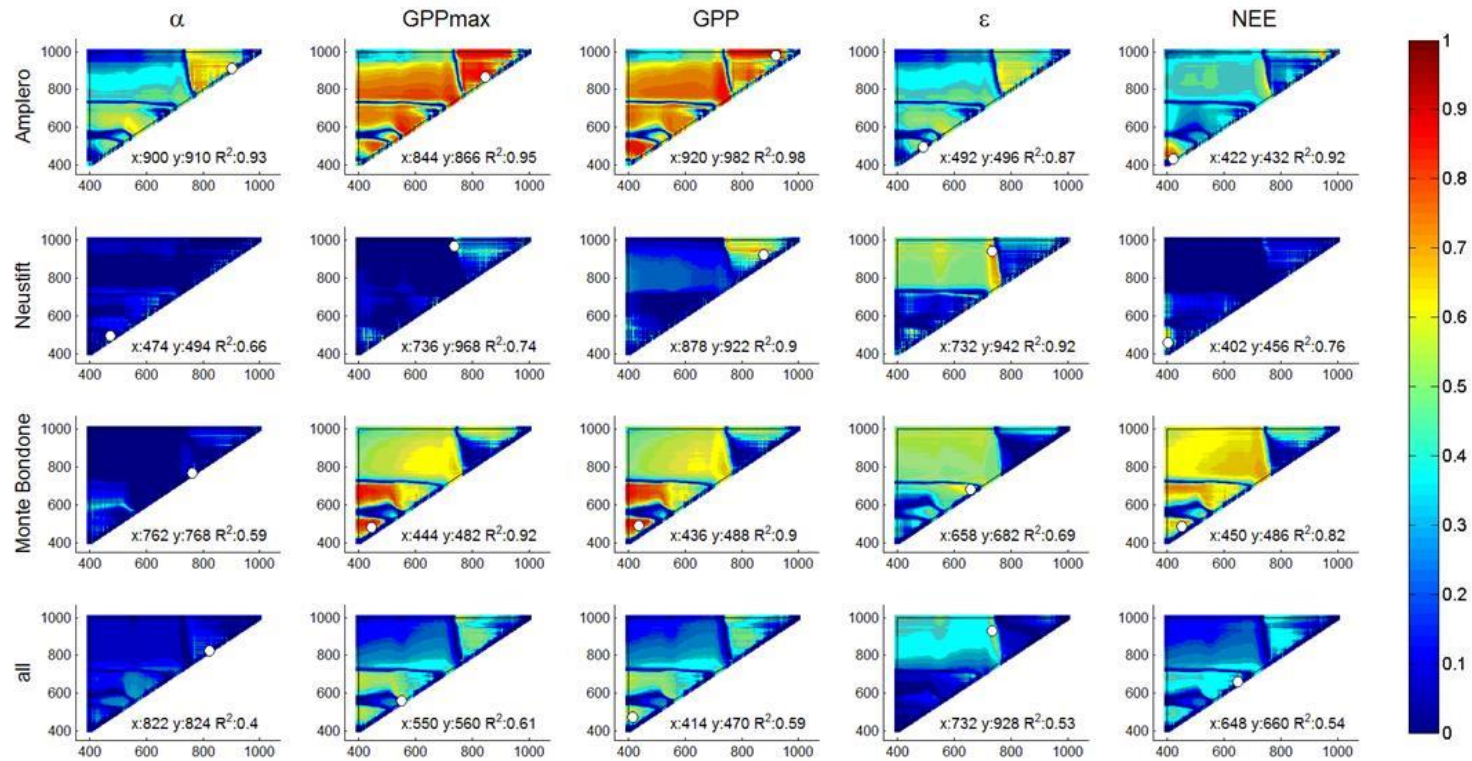


Figure 5. Correlograms of  $R^2$  values for  $\alpha$ ,  $GPP_{max}$  and midday averaged GPP,  $\epsilon$  and NEE and SD-type indices for Amplero, Neustift, Monte Bondone (both study years pooled) and all sites pooled. The white dots indicate the position of paired band combinations corresponding to the maximum  $R^2$ .

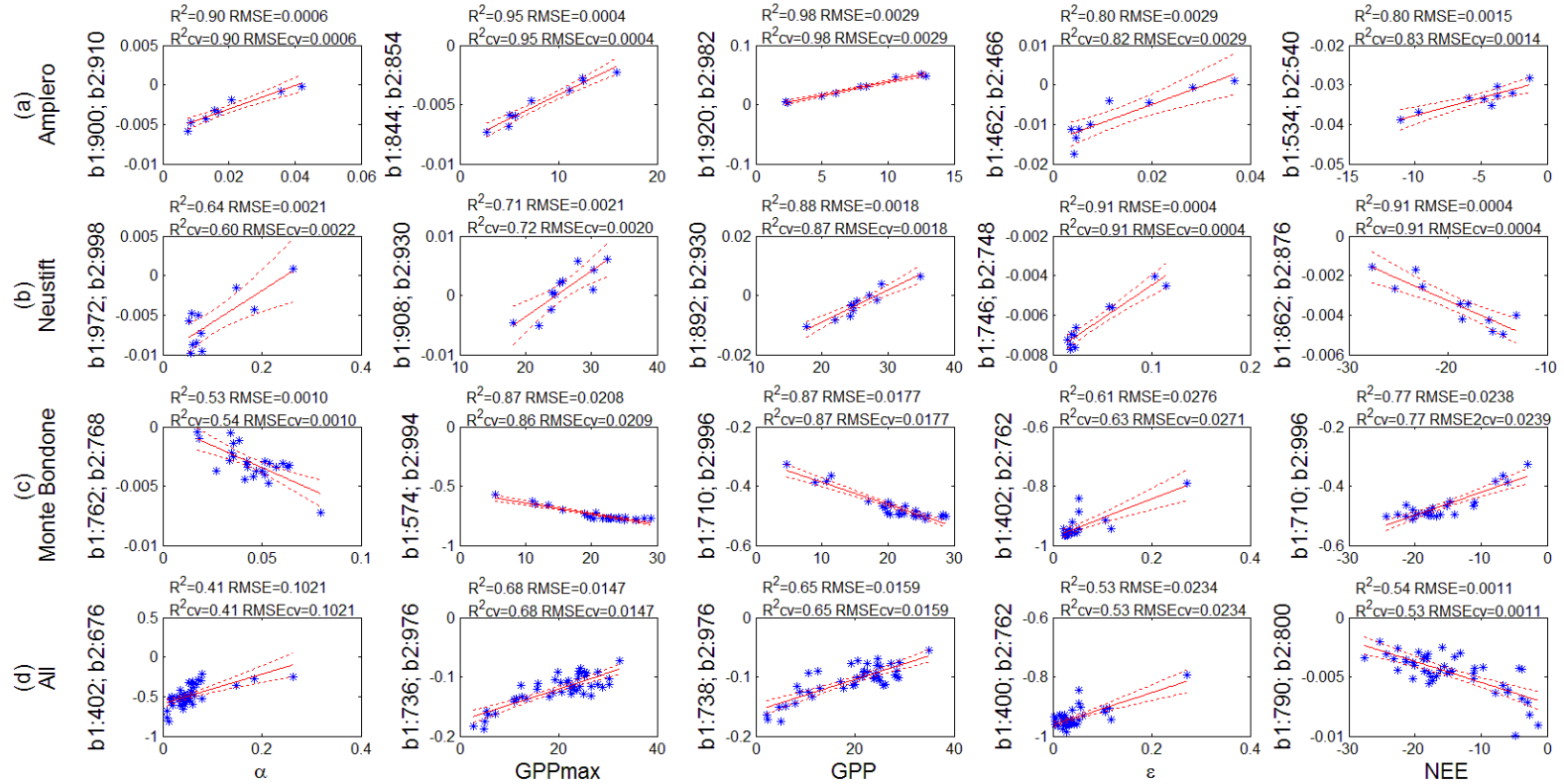


Figure 6. Results of linear correlation analysis for  $\alpha$ ,  $GPP_{max}$  and midday averaged GPP,  $\epsilon$  and NEE and selected best NSD-type indices for (a) Amplero, (b) Neustift, (c) Monte Bondone (both study years pooled) and (d) all sites pooled.  $R^2$ —Coefficient of determination; RMSE—Root Mean Square Error;  $R^2_{cv}$ —Cross-validated coefficient of determination;  $RMSE_{cv}$ — Cross-validated root Mean Square Error. The solid red lines indicate the fitted models and the dotted red lines represent the 95% upper and lower confidence bounds.

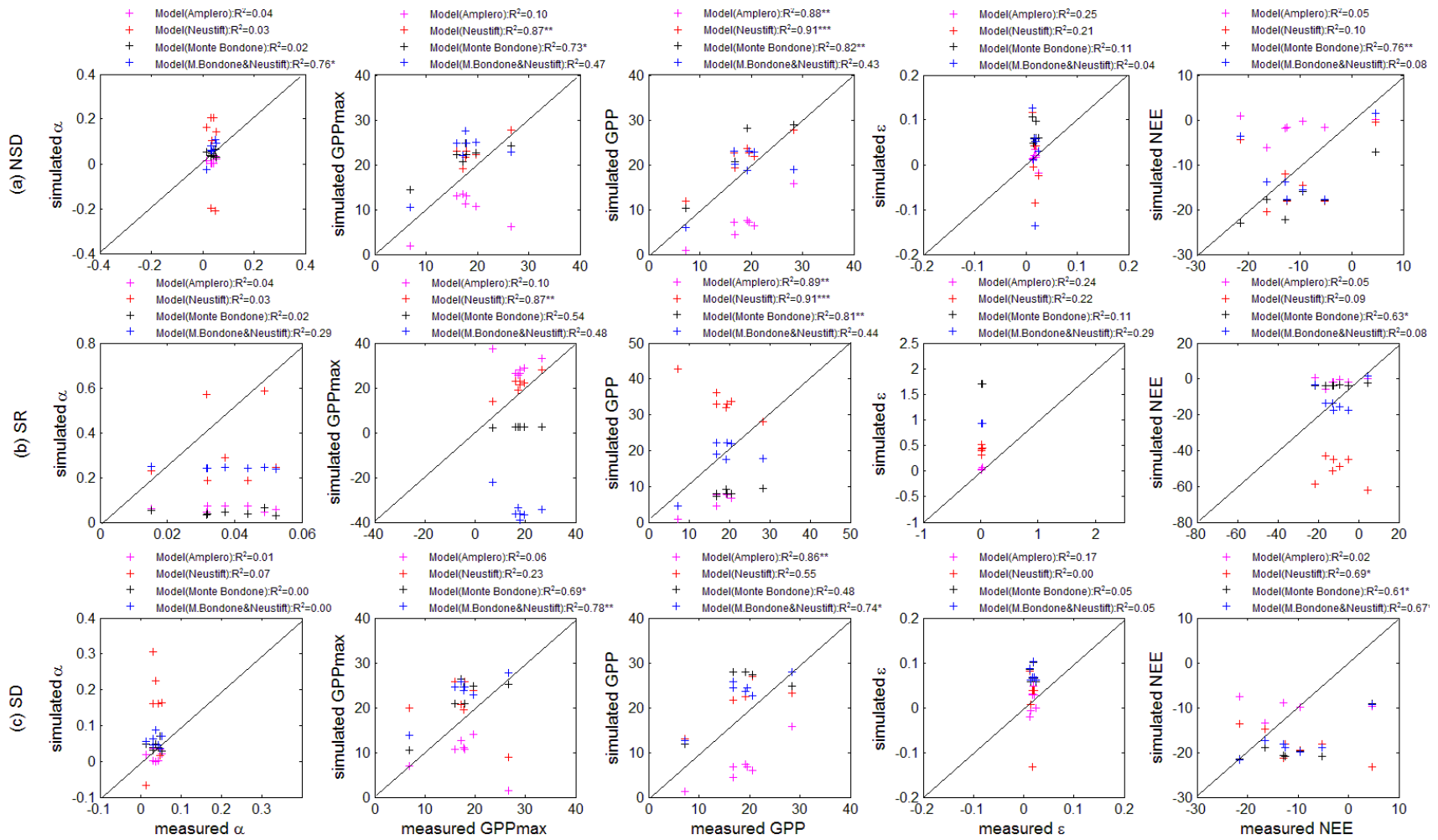


Figure 7. Results of validation of linear regression models between VIs ((a) NSD-type; (b) SR-type; (c) SD-type) and ecophysiological parameters:  $\alpha$ ,  $\epsilon$  (midday average),  $GPP_{max}$  and midday average  $CO_2$  fluxes ( $NEE$  and  $GPP$ ).  $R^2$  – coefficients of determination. Different colours represent results of the validation performed applying to the three new sites the model for Amplerio (in magenta), Neustift (in red) and Monte Bondone (in blue) and a model parameterized grouping Monte Bondone & Neustift

(M.Bondone&Neustift; in black). Statistical significance is indicated as \* ( $p < 0.05$ ), \*\* ( $p < 0.01$ ), and \*\*\* ( $p < 0.001$ ). The black lines are 1:1 lines.

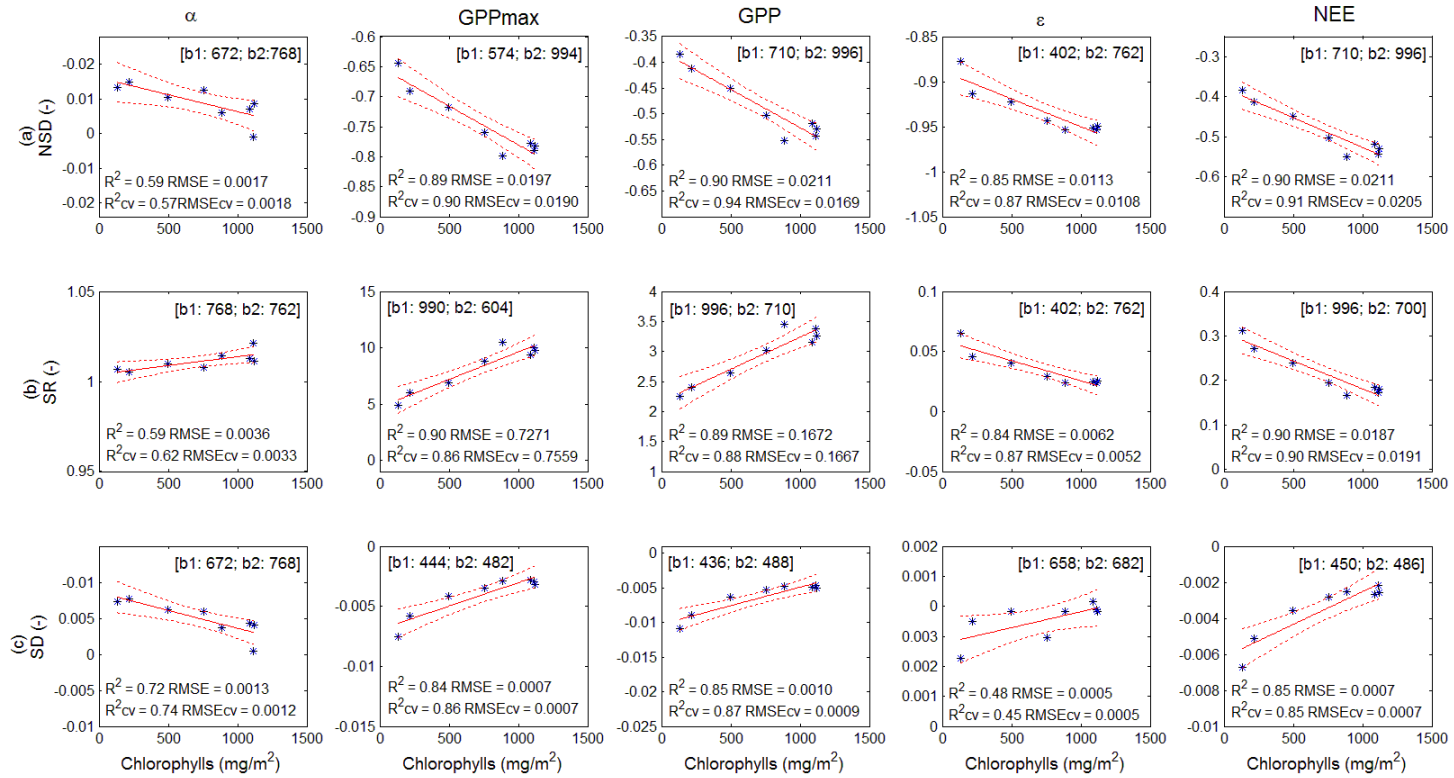
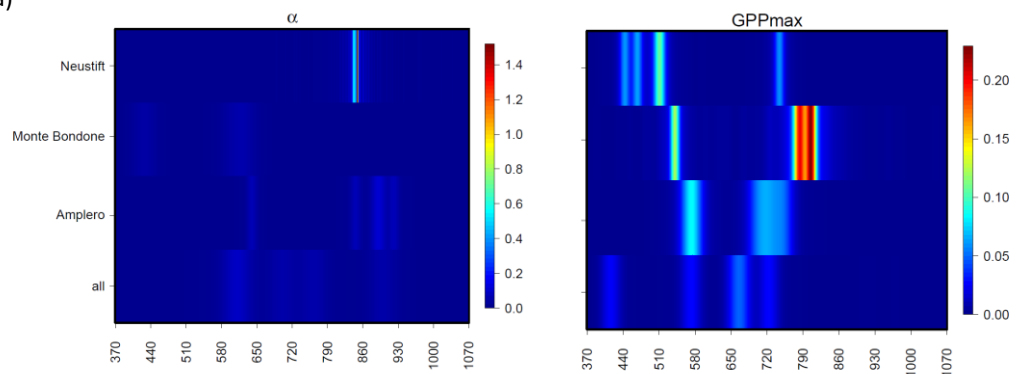


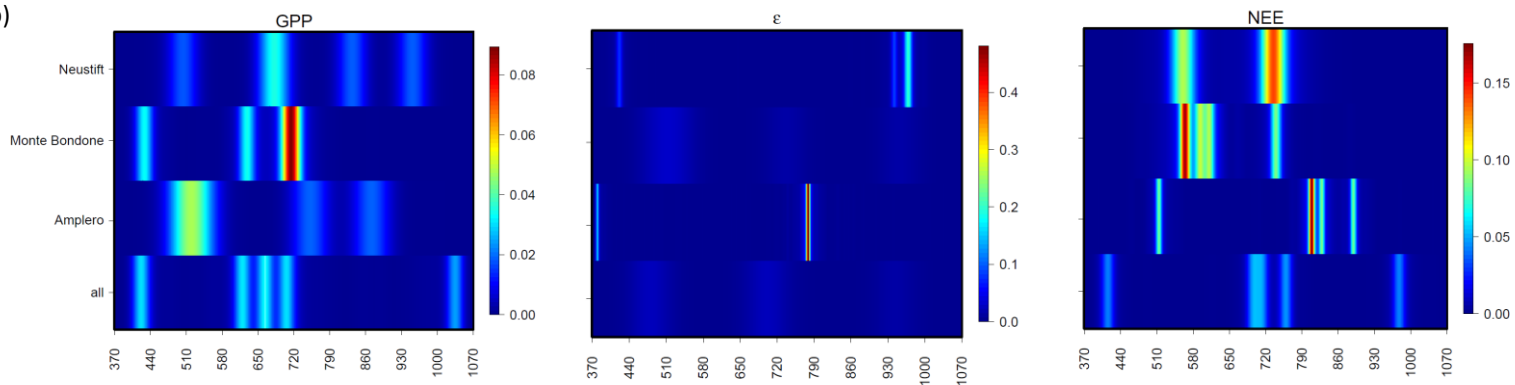
Figure 8. Correlation between selected (a) NSD-, (b) SR- and (c) SD-type indices for the  $\alpha$ , GPP<sub>max</sub>, midday GPP, midday  $\epsilon$  and midday NEE (plots in the columns) and the total chlorophyll content for Monte Bondone in 2013.  $R^2$ — coefficient of correlation; RMSE—root mean square error;  $R^2_{cv}$ — cross-validated coefficient of correlation; RMSE<sub>cv</sub>— cross-validated root mean square error.

The solid red lines indicate the fitted models and the dotted red lines represent the 95% upper and lower confidence bounds. The selected bands to compute NSD-, SR- and SD-type indices are reported in brackets.

(a)



(b)



(c)

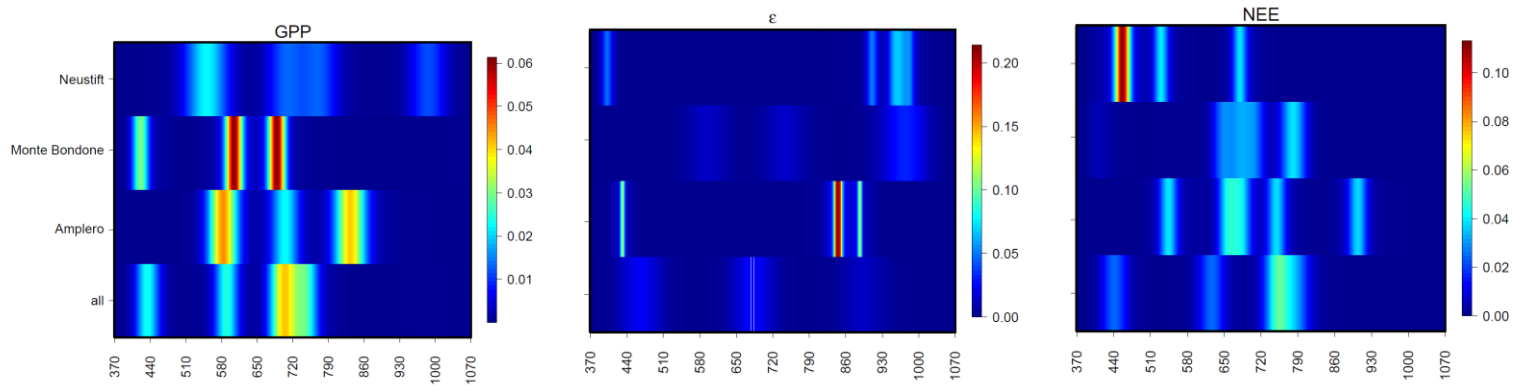




Figure 9. Results of the GA-rF method for band selection for Amplero, Neustift, Monte Bondone and all sites pooled for (a)  $\alpha$  and  $GPP_{max}$ , (b) midday average  $\varepsilon$ ,  $CO_2$  fluxes (NEE and GPP); (b) daily average  $\varepsilon$  and  $CO_2$  fluxes (NEE and GPP).

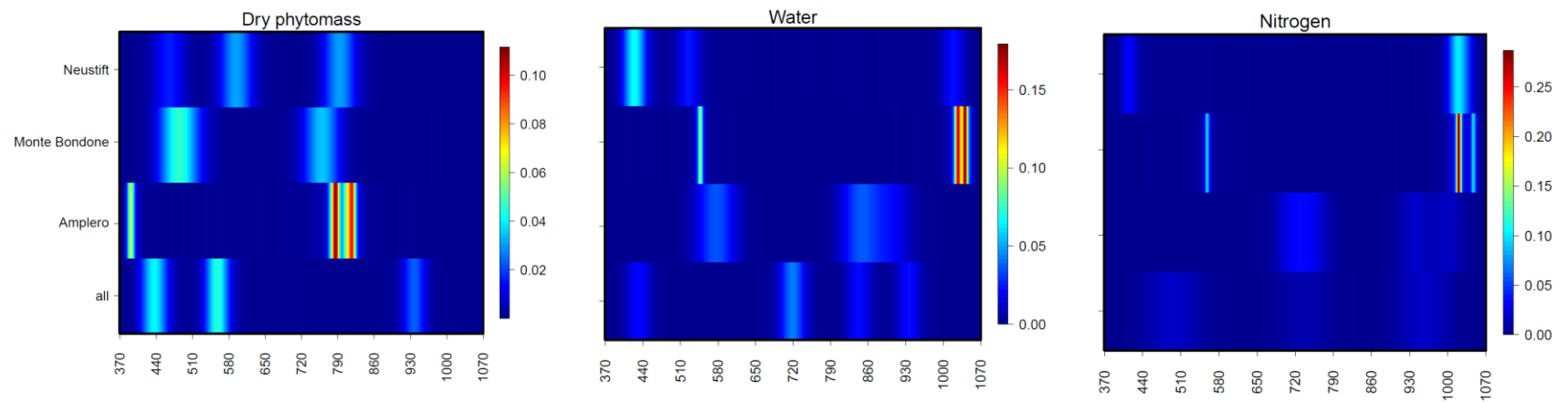


Figure 10. Results of the GA-rF method for band selection for Amplero, Neustift, Monte Bondone and all sites pooled for dry phytomass, water and nitrogen content.

1 Table 1. Description of the study sites and period.

<i>Site characteristics</i>	Amplero (IT-Amp)	Neustift (AT-Neu)	Monte Bondone (IT-MBo)
Latitude	41.9041	47.1162	46.0296
Longitude	13.6052	11.3204	11.0829
Elevation (m)	884	970	1550
Mean annual temperature (°C)	10.0	6.5	5.5
Mean annual precipitation (mm)	1365	852	1189
Vegetation type	Seslerietum apenninae	Pastinaco- Arrhenatheretum	Nardetum Alpigenum
Study period <sup>1</sup>	111-170, 2006 (9)	122-303, 2006 (16)	129-201, 2005 (13) 124-192, 2006 (12)

2 <sup>1</sup> from-to DOY, year (number of hyperspectral measurement dates)

3

4

Table 2. Summary of the vegetation indices characteristics used in this study.

Index name and acronym	Formula	Use	Reference
<b><i>Simple Spectral Ratio Indices</i></b>			
Simple Ratio (SR or RVI)	$SR = R_{830}/R_{660}$	Greenness	Jordan (1969)
Green Ratio Index (GRI)	$GRI = R_{830}/R_{550}$	Greenness	Peñuelas and Filella (1998)
Water Index (WI)	$WI = R_{900}/R_{970}$	Water content, leaf water potential, canopy water content	Peñuelas et al. (1993)
Simple Ratio Pigment Index (SRPI)	$SRPI = (R_{430})/(R_{680})$		Peñuelas et al. (1995)
Chlorophyll Index (CI)	$CI = (R_{750}/R_{720}) - 1$	Chlorophyll content	Gitelson et al. (2005)
<b><i>Normalized Spectral Difference Vegetation Indices</i></b>			
Normalized Difference Vegetation Index (NDVI)	$NDVI = (R_{830} - R_{660}) / (R_{830} + R_{660})$	Greenness	Rouse et al. (1973)
Normalized Phaeophytinization Index (NPQI)	$NPQI = (R_{415} - R_{435}) / (R_{415} + R_{435})$	Carotenoid /Chlorophyll ratio	Barnes et al. (1992)
Normalized Pigment Chlorophyll Index (NPCI)	$NPCI = (R_{680} - R_{430}) / (R_{680} + R_{430})$	Chlorophyll ratio	Peñuelas et al. (1994)
Red-edge NDVI (Red-edge NDVI)	$Red-edge NDVI = (R_{750} - R_{720}) / (R_{750} + R_{720})$	Chlorophyll content	Gitelson and Merzlyak (1994)
Structural Independent Pigment Index (SIPI)	$SIPI = (R_{800} - R_{445}) / (R_{800} + R_{445})$	Chlorophyll content	Peñuelas et al. (1995)

Table 3. Results of statistic of linear regression models between VIs and ecophysiological parameters:  $\alpha$ ,  $\varepsilon$  (midday average) and  $GPP_{max}$ .  $R^2$ —Coefficient of determination; and RMSE—Root Mean Square Error. Bold letters indicate the best fitting model.

VI	$\alpha$								$\varepsilon$								GPPmax							
	Amplero		Neustift		Monte Bondone		All		Amplero		Neustift		Monte Bondone		All		Amplero		Neustift		Monte Bondone		All	
	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE	$R^2$	RMSE
	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$	-	$\frac{\mu mol_{CO_2}}{\mu mol_{phat}}$
SR	0.57	0.01	0.04	0.07	<b>0.13</b>	0.01	0.06	0.04	0.50	0.01	0.33	0.03	0.35	0.04	0.18	0.04	<b>0.89</b>	1.58	0.01	4.31	0.78	2.76	0.28	6.71
GRI	0.29	0.01	0.00	0.07	0.13	0.01	0.00	0.05	0.26	0.01	<b>0.67</b>	0.02	0.44	0.04	<b>0.47</b>	0.03	0.69	2.66	0.00	4.35	0.81	2.53	0.09	7.51
WI	0.50	0.01	0.01	0.07	0.08	0.01	0.03	0.04	0.41	0.01	0.22	0.03	0.36	0.04	0.25	0.04	0.86	1.82	0.16	3.99	0.54	3.95	0.24	6.87
NDVI	0.44	0.01	0.04	0.07	0.06	0.01	0.06	0.04	0.40	0.01	0.30	0.03	0.53	0.04	0.43	0.03	0.79	2.21	0.03	4.28	<b>0.82</b>	2.50	0.37	6.27
SIPI	0.37	0.01	<b>0.07</b>	0.07	0.02	0.01	0.18	0.04	0.35	0.01	0.29	0.03	<b>0.64</b>	0.03	0.44	0.03	0.66	2.80	0.06	4.21	0.74	2.96	<b>0.47</b>	5.74
CI	0.49	0.01	0.00	0.07	0.09	0.01	0.01	0.05	0.41	0.01	0.65	0.02	0.43	0.04	0.34	0.04	0.81	2.08	0.01	4.34	0.80	2.62	0.16	7.24
PRI	<b>0.71</b>	0.01	0.02	0.07	0.02	0.01	0.14	0.04	0.50	0.01	0.19	0.03	0.28	0.05	0.40	0.04	0.41	3.68	<b>0.26</b>	3.75	0.11	5.50	0.14	7.33
EVI	0.47	0.01	0.03	0.07	0.03	0.01	0.14	0.04	0.46	0.01	0.43	0.03	0.53	0.04	0.38	0.04	0.78	2.25	0.01	4.33	0.70	3.21	0.32	6.50
NPQI	0.06	0.01	0.06	0.07	0.05	0.01	0.31	0.04	0.04	0.01	0.30	0.03	0.17	0.05	0.11	0.04	0.00	4.78	0.07	4.20	0.21	5.17	0.14	7.31
NPCI	0.50	0.01	0.07	0.07	0.03	0.01	<b>0.37</b>	0.04	0.51	0.01	0.17	0.03	0.00	0.05	0.00	0.05	0.53	3.28	0.17	3.97	0.17	5.33	0.32	6.52
SRPI	0.51	0.01	0.06	0.07	0.03	0.01	0.36	0.04	<b>0.56</b>	0.01	0.15	0.04	0.00	0.05	0.00	0.05	0.50	3.38	0.17	3.97	0.17	5.31	0.28	6.69
RedEdgeNDVI	0.48	0.01	0.00	0.07	0.07	0.01	0.01	0.05	0.40	0.01	0.65	0.02	0.47	0.04	0.40	0.04	0.79	2.16	0.00	4.34	0.80	2.58	0.19	7.09

Table 4. Results of statistic of linear regression models between VIs and midday average CO<sub>2</sub> fluxes: NEE and GPP. R<sup>2</sup>—Coefficient of determination; and RMSE—Root Mean Square Error. Bold letters indicate the best fitting model.

	GPP								NEE							
	Amplero		Neustift		Monte Bondone		All		Amplero		Neustift		Monte Bondone		All	
	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE	R <sup>2</sup>	RMSE
	- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$		- $\frac{\mu\text{mol}_{\text{CO}_2}}{\text{m}^2\text{s}}$	
<b>SR</b>	0.86	1.59	0.08	4.56	0.75	3.12	0.27	7.09	0.36	2.76	0.08	4.77	0.68	3.19	0.18	6.35
<b>GRI</b>	0.85	1.67	0.01	4.44	0.80	2.78	0.10	7.85	<b>0.54</b>	2.32	0.01	4.96	0.68	3.21	0.08	6.73
<b>WI</b>	<b>0.92</b>	1.23	0.05	3.25	0.50	4.41	0.24	7.20	0.44	2.57	0.05	4.87	0.43	4.28	0.17	6.42
<b>NDVI</b>	0.82	1.79	<b>0.14</b>	4.58	0.80	2.82	0.36	6.60	0.42	2.63	<b>0.14</b>	4.63	0.72	3.01	0.29	5.94
<b>SIPI</b>	0.65	2.50	0.08	4.57	0.72	3.32	<b>0.46</b>	6.08	0.33	2.82	0.08	4.79	0.65	3.34	<b>0.39</b>	5.51
<b>CI</b>	0.88	1.44	0.00	4.31	0.81	2.69	0.17	7.56	0.43	2.59	0.00	4.98	<b>0.75</b>	2.82	0.12	6.59
<b>PRI</b>	0.25	3.69	0.05	4.34	0.14	5.79	0.10	7.84	0.00	3.44	0.05	4.87	0.15	5.20	0.05	6.84
<b>EVI</b>	0.75	2.11	0.01	4.31	0.68	3.51	0.33	6.79	0.36	2.74	0.01	4.97	0.71	3.03	0.26	6.05
<b>NPQI</b>	0.04	4.17	0.08	4.27	0.14	5.78	0.16	7.57	0.24	2.99	0.08	4.78	0.19	5.08	0.12	6.60
<b>NPCI</b>	0.40	3.29	0.01	4.45	0.14	5.76	0.30	6.92	0.11	3.25	0.01	4.95	0.21	5.03	0.25	6.09
<b>SRPI</b>	0.35	3.42	0.01	4.44	0.15	5.74	0.27	7.08	0.08	3.30	0.01	4.95	0.22	5.01	0.22	6.19
<b>RedEdgeNDVI</b>	0.87	1.51	0.00	4.35	<b>0.81</b>	2.68	0.20	7.40	0.43	2.60	0.00	4.98	0.75	2.84	0.15	6.47

Table 5. Results of the correlation (R<sup>2</sup>—Coefficient of determination) between the best NDS, SR and SD-type indices selected for the  $\alpha$ , GPP<sub>max</sub>, midday GPP, midday  $\epsilon$  and midday NEE and dry phytomass, nitrogen and water content for Amplero, Neustift, Monte Bondone and all sites pooled. The selected bands to compute NSD-, SR- and SD-type indices are reported in brackets. Statistical significance is indicated as \* (p < 0.05), \*\* (p < 0.01), and \*\*\* (p < 0.001).

Index	Site	Parameter	$\alpha$	GPPmax	GPP	$\varepsilon$	NEE
			R <sup>2</sup>	R <sup>2</sup>	R <sup>2</sup>	R <sup>2</sup>	R <sup>2</sup>
			(-)	(-)	(-)	(-)	(-)
NSD-type	Amplero	Dry phytomass (g m <sup>-2</sup> )	0.66**	0.72**	0.58*	0.76**	0.36
	Amplero	Nitrogen content (%)	0.30	0.32	0.19	0.49*	0.15
	Amplero	Water content (%)	0.28	0.53*	0.56*	0.44	0.55*
	Neustift	Dry phytomass (g m <sup>-2</sup> )	0.00	0.26	0.35	0.44*	0.02
	Neustift	Nitrogen content (%)	0.16	0.15	0.21	0.77**	0.03
	Neustift	Water content (%)	0.00	0.00	0.03	0.59*	0.10
	Monte Bondone	Dry phytomass (g m <sup>-2</sup> )	0.02	0.59***	0.49***	0.55***	0.49***
	Monte Bondone	Nitrogen content (%)	0.08	0.52***	0.38**	0.48***	0.38**
	Monte Bondone	Water content (%)	0.09	0.48***	0.35**	0.42***	0.35**
	All	Dry phytomass (g m <sup>-2</sup> )	0.05	0.02	0.02	0.05	0.00
	All	Nitrogen content (%)	0.26***	0.04	0.02	0.41***	0.09
	All	Water content (%)	0.00	0.01	0.01	0.10*	0.00
SR-type	Amplero	Dry phytomass (g m <sup>-2</sup> )	0.66**	0.72**	0.58*	0.76*	0.36
	Amplero	Nitrogen content (%)	0.30	0.32	0.20	0.49*	0.15
	Amplero	Water content (%)	0.28	0.53*	0.56*	0.44	0.55*
	Neustift	Dry phytomass (g m <sup>-2</sup> )	0.00	0.26	0.35	0.44*	0.02
	Neustift	Nitrogen content (%)	0.16	0.15	0.21	0.77**	0.03
	Neustift	Water content (%)	0.01	0.00	0.03	0.59*	0.10
	Monte Bondone	Dry phytomass (g m <sup>-2</sup> )	0.02	0.50***	0.50***	0.55***	0.45*
	Monte Bondone	Nitrogen content (%)	0.08	0.48***	0.38**	0.48***	0.35
	Monte Bondone	Water content (%)	0.09	0.44***	0.34**	0.41***	0.30*
	All	Dry phytomass (g m <sup>-2</sup> )	0.08	0.01	0.01	0.05	0.00
	All	Nitrogen content (%)	0.26***	0.03	0.02	0.40***	0.09
	All	Water content (%)	0.00	0.01	0.01	0.11*	0.00
SD-type	Amplero	Dry phytomass (g m <sup>-2</sup> )	0.64***	0.81**	0.59*	0.58*	0.25
	Amplero	Nitrogen content (%)	0.22	0.30	0.22	0.30	0.03

Amplero	Water content (%)	0.16	0.45*	0.59*	0.19	0.49*
Neustift	Dry phytomass (g m <sup>-2</sup> )	0.21	0.04	0.37	0.20	0.00
Neustift	Nitrogen content (%)	0.11	0.01	0.12	0.81**	0.08
Neustift	Water content (%)	0.02	0.26	0.00	0.64*	0.52*
Monte Bondone	Dry phytomass (g m <sup>-2</sup> )	0.15	0.42***	0.36**	0.45***	0.36**
Monte Bondone	Nitrogen content (%)	0.28***	0.34**	0.34**	0.38**	0.35**
Monte Bondone	Water content (%)	0.27***	0.34**	0.34	0.31**	0.30**
All	Dry phytomass (g m <sup>-2</sup> )	0.20***	0.01	0.00	0.30***	0.01
All	Nitrogen content (%)	0.01	0.01	0.02	0.02	0.11*
All	Water content (%)	0.01	0.03	0.04	0.28***	0.06

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