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Probing the fate of soil-derived core and intact polar GDGTs in aquatic environments

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Abstract

We have performed incubation experiments in order to examine the fate of branched glycerol dialkyl glycerol tetraether (brGDGT) membrane lipids upon entering an aquatic environment and to assess the suitability of brGDGTs as tracers for fluvial land-sea transport of soil organic carbon. We incubated a soil from the Rakaia River catchment on the South Island of New Zealand using Rakaia River water and ocean water collected near the river mouth as inocula for a period of up to 152 days. The concentrations of brGDGTs derived from intact polar (“living”; IPL) lipids and core (“fossil”; CL) lipids remained stable over the course of the experiment, suggesting an absence of significant brGDGT production or degradation. Moreover, the lack of change in brGDGT distribution during the experiment implies that the initial soil signature remains unaltered during transport through the aquatic environment, at least over the time frame of the experiment. In contrast, the total pool of isoprenoid GDGTs (isoGDGTs), currently attributed to soil *Thaumarchaeota*, increased substantially (27–72 %) in all incubation setups by the end of the experiment. As a consequence, a decrease in Branched and Isoprenoid Tetraether (BIT) index values – a proxy for the relative input of fluvially discharged soil material into a marine system – became evident after an incubation period of 30 days, with a maximum final decrease of 0.88 to 0.74 in the experiment with river water. The relative distribution within the isoGDGT pool shows changes with time, suggesting different membrane adaptation rates to the aquatic environment, or a shift in source organism(s). While the stability of soil brGDGTs in aquatic environments reinforces their potential as tracers for land-sea transport of soil organic carbon and their use in paleoclimate reconstructions, the distributional differences between GDGTs in river water and nearby soil indicate that further research is needed to pinpoint the sources of GDGTs that are ultimately discharged to the oceans and are subsequently archived in continental margin sediments.

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1 Introduction

The global carbon cycle encompasses a myriad of biogeochemical processes that influence our climate and link all carbon reservoirs on the Earth surface. Soils are considered to play a very active and fundamental role in this cycle, as their mobilization from land and subsequent deposition in marine sediments serves as a long-term sink of atmospheric CO₂. However, the exact magnitude and mechanisms of soil organic carbon (OC) transfer from terrestrial source to marine sink remain elusive (e.g. Cole et al., 2007; Weyhenmeyer et al., 2012), which partly stems from the lack of a suitable tracer of the soil OC pool. Our current insights in land-sea soil OC transport dynamics are primarily based on bulk properties of river particulate suspended matter (SPM). However, next to soil OC, river SPM also comprises carbon derived from aquatic production and “fossil” carbon from rock erosion, each of which influences the bulk properties of SPM (Blair et al., 2010).

Analytical advances in the field of organic biogeochemistry have yielded a growing number of powerful tools with the potential to exclusively target components of specific pools, including soil OC. In this context, branched glycerol dialkyl glycerol tetraethers (brGDGTs; Fig. 1) have been put forward as tracers for soil-derived OC in carbon transport studies (Hopmans et al., 2004; Kim et al., 2006). Since their discovery (Sinninghe Damsté et al., 2000; Schouten et al., 2000), brGDGTs have been found in soils and peats worldwide (Weijers et al., 2007b; Peterse et al., 2012). The exact organism(s) that produce(s) these compounds have yet to be identified, but current evidence points towards an origin from heterotrophic soil bacteria (Pancost and Sinninghe Damsté, 2003; Oppermann et al., 2010; Weijers et al., 2010) from the phylum of *Acidobacteria* (Weijers et al., 2009; Peterse et al., 2010; Sinninghe Damsté et al., 2011). BrGDGTs have been used as a proxy for the relative input of fluvially transported soil OC in marine systems based on their abundance in coastal marine sediments relative to that of crenarchaeol (Fig. 1), an isoprenoid GDGT (isoGDGTs) produced by marine *Thaumarchaeota* (Sinninghe Damsté et al., 2002), parameterized as the Branched and Isoprenoid Tetraether

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from the river was sampled in jerry cans close to the town Rakaia (43°44′57.9″ S, 172°01′52.9″ E), about 20 km upstream from the river mouth and outside of the tidal influence. The ocean water used as inoculum was collected from the shoreline (43°54′15.1″ S, 172°11′41.9″ E), close to the river mouth. Based on pH measurements, this water (pH 7.6) was strongly fluviually influenced (pH 7.5). Surface soil (0–10 cm; pH 5.8) was collected on the Canterbury Plains within the Rakaia River catchment (43°52′00.8″ S, 172°09′10.8″ E). All soil material and incubation water was collected in January 2012 and directly transported to the University of Otago, where the incubation experiment was set up the next day. SPM in the river and ocean water was collected by filtration (10–100 L) onto pre-combusted 0.7 µm GF/F or polyethersulfone membranes in the field, after which they were stored frozen.

In the laboratory at the University of Otago, the soil was sieved over a 2 mm mesh to remove roots and homogenize the sample, after which two 10 g subsamples of the soil were directly frozen at –20 °C for subsequent determination of initial GDGT composition. For the incubations, twenty-eight 250 mL bottles were filled with 10 g soil subsamples, after which 100 mL of river water or ocean water, untreated to preserve the natural microbial community, was added to twelve bottles each. Distilled water was added to the remaining four bottles to serve as control setup. The incubation bottles were loosely plugged with cotton wool, creating quasi-aerobic conditions, and placed on a shaker table from January to June 2012. Based on IPL degradation rates observed in previous studies (Harvey et al., 1986; Logemann et al., 2011), two bottles with river water and two bottles with ocean water were taken from the shaker table and immediately placed into a freezer (–20 °C) at 1, 7, 14, 30, 91, and 152 days after the start of the experiment. Duplicate control samples were frozen after 91 and 152 days. At the end of the experiment all bottles were shipped on ice to ETH Zürich, Switzerland, where the supernatant was pipetted off, the soils freeze dried, and stored frozen at –20 °C until subsequent sample work-up and analysis.

2.2 GDGT extraction and analysis

Freeze-dried soil samples (3–4 g) were solvent-extracted three times using a Bligh and Dyer technique modified from Sturt et al. (2004). In short, a solvent mixture of MeOH:dichloromethane (DCM):phosphate buffer at pH 7.4 (2:1:0.8, v/v/v) was added to the soils and ultrasonically extracted for 10 min. The extract was collected each time after centrifuging at 1000 rpm for 5 min. DCM and phosphate buffer were added to the combined extracts to a volume ratio of 1:1:0.9 to obtain phase separation. The DCM phase, containing the GDGTs, was collected after centrifuging, and the remaining solvent was rinsed twice with DCM. The combined DCM phases were dried under N₂ and passed over a silica column (deactivated with 1 weight % water) to separate CLs and IPLs according to (Pitcher et al., 2009), with the exception that hexane:ethyl acetate 1:1 (v/v) was used to elute the CLs. An aliquot of the IPL fraction was analyzed directly for CLs in order to assess potential carry over. The remainder of IPL fraction was dissolved in 6N HCl in MeOH and heated at 100 °C for at least 3 h to release IPL-bound CLs. A known amount of C₄₆ GDGT standard (Huguet et al., 2006) was added to all fractions prior to analysis.

The filters were freeze dried and extracted with a MARS Xpress microwave extraction system, using DCM:MeOH 9:1 (v/v). After centrifugation (5 min at 400 rpm), the total lipid extract (TLE) was pipetted off and the residues were rinsed twice with DCM:MeOH 9:1 (v/v). The combined extracts were dried under N₂ with a known amount of C₄₆ GDGT standard, after which the TLEs were separated into an apolar and a polar (GDGT) fraction by passing them over a silica (1 % water deactivated) column using hexane:DCM 9:1 (v/v) and DCM:MeOH 1:1 (v/v), respectively.

All GDGT-fractions were dissolved in hexane:isopropanol 99:1 (v/v), filtered over a 0.45 µm PTFE filter, and analyzed using high performance liquid chromatography/atmospheric pressure chemical ionization – mass spectrometry (HPLC/APCI-MS) with an Agilent 1260 Infinity series LC/MS at ETH Zürich according to Schouten et al. (2007a). Separation of the GDGTs was achieved with a Grace Prevail Cyano

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column (3 μm , 150 mm \times 2.1 mm) after passing through a guard column of the same material (5 μm , 7.5 mm \times 2.1 mm) with hexane : isopropanol (99 : 1, v/v) as an eluent at a flow rate of 0.2 mL min^{-1} . The GDGT-fractions eluted isocratically with 90 % A and 10 % B for 5 min, and then with a linear gradient to 18 % B for 34 min, where A = hexane and B = hexane : isopropanol 9 : 1 (v/v). Selective ion monitoring of the $[\text{M}+\text{H}]^+$ was used to detect the different GDGTs. Although brGDGTs appear to have a higher response factor than isoGDGTs (Schouten et al., 2013a), quantification was done assuming similar response factors for all GDGTs and the internal standard. Hence, reported brGDGT concentrations are likely overestimated compared to those of the isoGDGTs. The amounts of all GDGTs are given as the average and standard deviation of the duplicate incubation samples.

2.3 GDGT-based index calculations

The CBT and MBT' indices based on brGDGTs were calculated according to (Weijers et al., 2007b):

$$\text{CBT} = -\log((Ib + IIb)/(Ia + IIa)) \quad (1)$$

and (Peterse et al., 2012):

$$\text{MBT}' = (Ia + Ib + Ic)/(Ia + Ib + Ic + IIa + IIb + IIc + IIIa) \quad (2)$$

The BIT index was calculated following (Hopmans et al., 2004):

$$\text{BIT} = (Ia + IIa + IIIa)/(c_{\text{ren}} + Ia + IIa + IIIa) \quad (3)$$

and for the TEX_{86} index based on isoGDGTs the equation of (Schouten et al., 2002) was used:

$$\text{TEX}_{86} = (\text{GDGT-2} + \text{GDGT-3} + c_{\text{ren}}')/(\text{GDGT-1} + \text{GDGT-2} + \text{GDGT-3} + c_{\text{ren}}') \quad (4)$$

Roman numerals and GDGT names refer to the molecular structures in Fig. 1.

3 Results and discussion

3.1 BrGDGT concentrations over time

The total brGDGT pool (IPLs + CLs) initially present in the soil is $239 \pm 13 \text{ ng g}^{-1}$. In the different incubation experiments these concentrations varied between $187\text{--}325 \text{ ng g}^{-1}$ (river water), $186\text{--}322 \text{ ng g}^{-1}$ (ocean water), and $252\text{--}407 \text{ ng g}^{-1}$ (distilled water control) (Fig. 2). Although the abundance of CL-brGDGTs seems to increase with time in the control setup, the trend is only weak and not significant ($r^2 = 0.34$, $p = 0.122$) due to the large spread in concentration between the two replicate samples at $t = 152$ days (Fig. 2). Overall both IPL-derived and CL brGDGT concentrations in all experimental setups are within the same range and remain essentially constant through time. Since the soils were incubated under natural light conditions, this indicates that soil-derived brGDGTs are not sensitive to photodegradation during their presence in an aquatic environment.

The contribution of IPLs to the total amount of brGDGTs in the soil at $t = 0$ is $12.4 \pm 0.3\%$, which is in the same range as has previously been reported for temperate soils from the Netherlands, Scotland, and the UK (Peterse et al., 2010; Weijers et al., 2011). The contribution of IPL-derived brGDGTs to the total pool varies mostly within the same range in the river water (7–16%), ocean water (9–15%), and control (9–17%) setups, and like the total amount of CL-brGDGTs shows no significant trends or changes over time. This suggests that brGDGTs are either produced at the same rate that they are degraded, or that production and/or degradation of brGDGTs occurs at such low rates that these processes are not detected within the timeframe of this experiment. The apparent stability of the IPL-derived, or “living” brGDGT signature (Fig. 2) during the experiment is surprising, given the perceived lability of IPLs in general and their susceptibility to loss of their headgroups within a few days after cell lysis (Harvey et al., 1986; White et al., 1979). However, previous studies have indicated that the type of headgroup and the bond through which it is attached to a core lipid may influence the rate of degradation, offering a potential explanation for the absence of

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the latter phase of the warming experiment (Huguet et al., 2013). This finding is thus also consistent with the absence of any trends in our incubation setup, and suggests that longer incubation periods (> 1 year) with different temperature regimes may be necessary to reveal changes in brGDGT abundances.

While nutrient conditions in the river and ocean water added to the soil are presumably different, but this does not seem to have had any affect on brGDGT concentrations. Moreover, given the soil:water ratio (10 g in 100 mL) used in this experiment, the amounts of nutrients that are released from the soil after the addition of water is likely substantially higher than the concentrations initially present in the river and ocean. Nevertheless, the effect of nutrient availability on brGDGTs in soils has not yet received much attention, although recently was shown that variations in the amount and distribution of brGDGTs in lake sediments did not relate to water column nutrient concentrations (Loomis et al., 2014).

3.2 IsoGDGT concentrations over time

Crenarchaeol is the most abundant isoGDGT in the New Zealand soil at the start of the incubation, although its amount ($63 \pm 7 \text{ ng g}^{-1}$), as well as that of all other isoGDGTs ($50 \pm 5 \text{ ng g}^{-1}$) is low compared to the pool of brGDGTs ($239 \pm 13 \text{ ng g}^{-1}$). In contrast, the average fraction of isoGDGTs initially present in the soil as IPL is 54% for crenarchaeol to 76% for isoGDGT-2, substantially higher than proportion of brGDGTs occurring as IPLs (12%) at the commencement of the experiment. Similar proportions of IPL-derived crenarchaeol have been found in soils from the Amazon (~50%; Zell et al., 2013b). The total concentration (IPLs + CLs) of all isoGDGTs increases substantially during the experiment (Fig. 2), and that of isoGDGT-0 in the ocean water setup has even doubled by the end of the incubation.

The concentrations of the individual IPL-derived isoGDGTs vary between the different incubation setups, but are overall highest in ocean water (Fig. 2). The difference in isoGDGT concentration between incubations cannot be explained simply by the addition of aquatic isoGDGTs to the soil derived isoGDGT pool, as 100 mL of ocean water

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only contained ~ 0.5 ng isoGDGTs, which is $< 1\%$ of their initial concentration in the soil. Because *Thaumarchaeota* play an important role in the nitrogen cycle in soils and the marine environment (Leininger et al., 2006; Wuchter et al., 2006; Zhang et al., 2010), nutrient availability may explain the concentration differences between experiments. So far, crenarchaeol has been found in all (enrichment) cultures of ammonia-oxidizing *Thaumarchaeota*, and is considered as a biomarker lipid for this group (de la Torre et al., 2008; Pitcher et al., 2010, 2011; Sinninghe Damsté et al., 2012). Its high concentration in the incubation setup using ocean water suggests that more ammonium is released in this experiment, likely due to the higher abundance of exchangeable cations (e.g. Na^+ , Mg^+) in ocean water compared to fresh water. This is in agreement with the common method to determine available NH_4^+ in soils, which involves shaking soil in a KCl or Na_2SO_4 solution to extract ammonium from the soil matrix (e.g. Mehlich, 1953). The concentration profiles of the other isoGDGTs are comparable to that of crenarchaeol (Fig. 2), implying that they are also primarily derived from soil *Thaumarchaeota*. Although isoGDGTs are known to also commonly occur in the membranes of several other archaea (Schouten et al., 2013b, and references therein), the trace amounts of aquatic isoGDGTs added to the initial pool of soil isoGDGTs implies that production by soil archaea must be responsible for the isoGDGT increase during the experiment.

Despite the increase of the total isoGDGT pool, there is no statistically significant increase in the concentration of IPL-derived isoGDGTs during the course of the incubation experiment (e.g. max. $r^2 = 0.18$ for IPL-derived crenarchaeol in ocean water; Fig. 2). We assume therefore that the production rate of IPL-isoGDGTs is approximately constant. The proportional decrease of IPL-isoGDGTs with time (Fig. 3) subsequently indicates that the increase in the overall isoGDGT pool is a result of CL accumulation. This in turn implies that the turnover of the IPL pool, and thus the release of CL-isoGDGTs, is faster than the degradation of these CLs. The rapid IPL degradation rate may indicate that the isoGDGTs contain a phospho headgroup (cf. Harvey et al., 1986; Lengger et al., 2012), consistent with a recent survey of GDGTs in soil *Thaumarchaeota*.

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chaetoa enrichment cultures (Sinninghe Damsté et al., 2012) and with a molecular and genomic study of the archaeal community in the water column of Lake Challa in Africa (Buckles et al., 2013). Thus, in contrast to the apparent stability of IPL-brGDGTs, the likely predominance of isoGDGTs with phospho headgroups in our experiment supports the use of these specific IPLs as indicators for their living precursor organisms.

3.3 GDGT distribution changes

To evaluate if and how soil GDGT signatures are modified in aquatic systems, we calculated brGDGT and isoGDGT-based indices commonly used in paleoclimate and carbon cycle studies (i.e. the MBT', CBT, BIT, TEX₈₆, and isoGDGT-0/cren).

The apparent absence of brGDGT production/degradation already suggest that no distribution changes are expected. Indeed, the values for the MBT' and CBT indices for the ocean and river water experiments exhibit minimal variation over time, remaining within a range of 0.04 for the MBT' index, and 0.02 for the CBT (Fig. 4). Surprisingly, the largest changes are observed in the control experiment using distilled water. The MBT' index for the IPL-derived brGDGT fraction in the control setup increases from 0.23 ± 0.00 to 0.29 ± 0.04 ($r^2 = 0.57$, $p = 0.023$; Fig. 4), and is primarily caused by an increase in the concentration of brGDGT-Ia (from 23 ± 0.3 to 29 ± 3.9 %). A closer look at the relative distributions of each of the brGDGT types at $t = 0$ and $t = 152$ days reveals a subtle shift in the contribution of IPL-derived brGDGT-III (decreasing from 23 ± 1.3 to 19 ± 0.1 %) to brGDGT-II (increasing from 54 ± 0.4 to 58 ± 0.4 %) in river water (data not shown). Since both brGDGTs-II and III in the denominator of the MBT' index (Eq. 2), and the fraction of IPL-derived brGDGT-I remains stable, this change is not reflected by the index values. Given that the absolute amounts of IPL-derived brGDGTs do not significantly change with time, it is hard to determine the exact processes that cause the distributional changes, as in situ production or preferential degradation of specific brGDGTs could both influence the total concentration. Due to the absence of brGDGTs with cyclopentane moieties in the IPL-derived fraction it was not possible to calculate CBT index values for any of the incubation experiments, suggesting that if

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TEX₈₆ index values start to change between 30 and 91 days of incubation for the IPL-fraction in river water and the CL-fraction in ocean water experiments (Fig. 4). We find no clear explanation for the diverging trends for the different water types or IPL vs. CL fractions. Although the major part of the community producing the isoGDGTs seems to be soil-derived, the different trends may be caused by preferential growth of specific organisms. Alternatively, the relative distribution of isoGDGTs may differ per type of headgroup in both marine (Schouten et al., 2008; Pitcher et al., 2011; Lengger et al., 2012) and soil *Thaumarchaeota* (Sinninghe Damsté et al., 2012), so that the deviation in TEX₈₆ values may be explained by different turnover rates per headgroup. However, even though the total variation in TEX₈₆ corresponds with a 4 °C SST change in the experiment, given the low absolute amounts of terrestrial isoGDGTs that are finally discharged to the ocean in this case, this will have a minor impact on TEX₈₆ records from the marine environment where their abundances are much higher.

The ratio of isoGDGT-0/crenarchaeol is on average 0.26 ± 0.05 for the CL, and 0.51 ± 0.06 for the IPL-derived fraction (Fig. 4). Blaga et al. (2009) suggested that the ratio of isoGDGT-0/crenarchaeol could be used to detect methanogenic activity, as a high concentration of isoGDGT-0 is often associated with the presence of methanogens (Schouten et al., 2013b, and references therein). The ratio values below 2 indicate that the contribution of isoGDGT-0 produced by methanogenic archaea can be considered low, and that all isoGDGTs in our experiments are primarily produced by *Thaumarchaeota*.

3.4 Comparing brGDGT distributions in river SPM and catchment soil

Despite prior evidence for in situ production of brGDGTs in rivers (Kim et al., 2012; Zhang et al., 2012; Zell et al., 2013b; Yang et al., 2013; De Jonge et al., 2014), we have not observed this in our experiment. However, we can see that the distribution of brGDGTs in SPM of the Rakaia River does not precisely reflect that in the soil sample (Fig. 4), which does support the idea that brGDGTs in soils and river SPM have different sources. The absence of an in situ production signal in our data can be explained

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As opposed to brGDGTs, significant production of isoGDGT took place during the incubation experiments, altering their initial relative distribution. Although the changes in TEX_{86} index values correspond with a maximum change of about 4°C in reconstructed SSTs in the ocean water setup, the low abundance of isoGDGTs in soils relative to in marine settings suggests that their land-sea transport would have minimal impact on TEX_{86} recorded in the marine sedimentary record. Nevertheless, isoGDGT production after more than 30 days was sufficient to affect BIT index values, with potential consequences for the interpretation of TEX_{86} -based SST records with a BIT index close to the cutoff value of 0.3 (Weijers et al., 2006b), as the actual input of soil-derived GDGTs may be larger than anticipated

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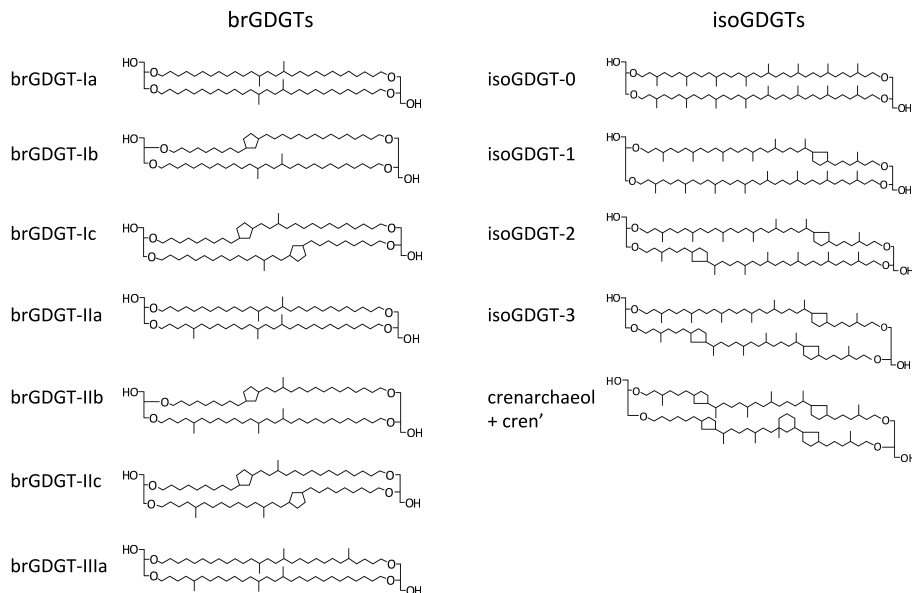


Figure 1. Molecular structures of the brGDGTs and isoGDGTs monitored during the incubation experiment.

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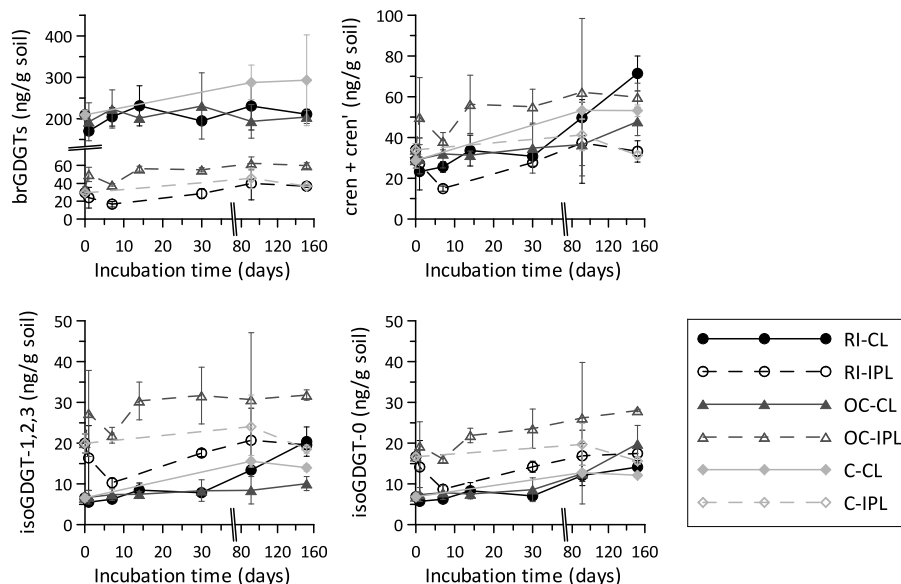


Figure 2. Average concentrations and standard deviation of core lipid (CL) and intact polar lipid (IPL)-derived (a) brGDGTs, (b) crenarchaeol + cren', (c) isoGDGT-1,2,3, and (d) isoGDGT-0 in a sandy loam soil from the Raiaka River catchment, New Zealand, during incubation ($t = 152$ days) in river (RI), ocean (OC), and distilled (C) water under quasi-aerobic conditions.

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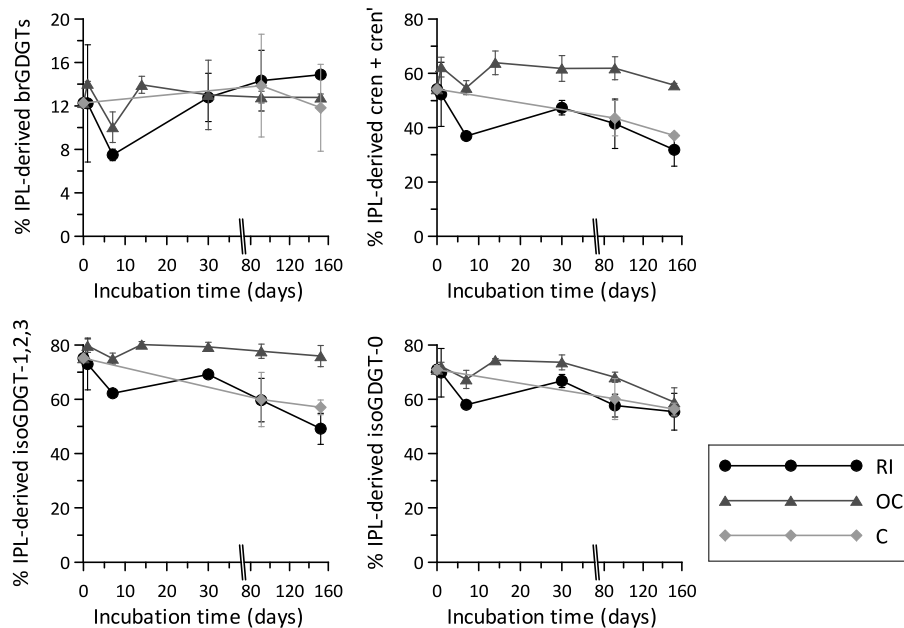


Figure 3. Average percentage of the total pool of (a) brGDGTs, (b) crenarchaeol + cren', (c) isoGDGTs-1,2,3, and (d) isoGDGT-0 present in “living”, or intact polar lipid (IPL)-derived form.

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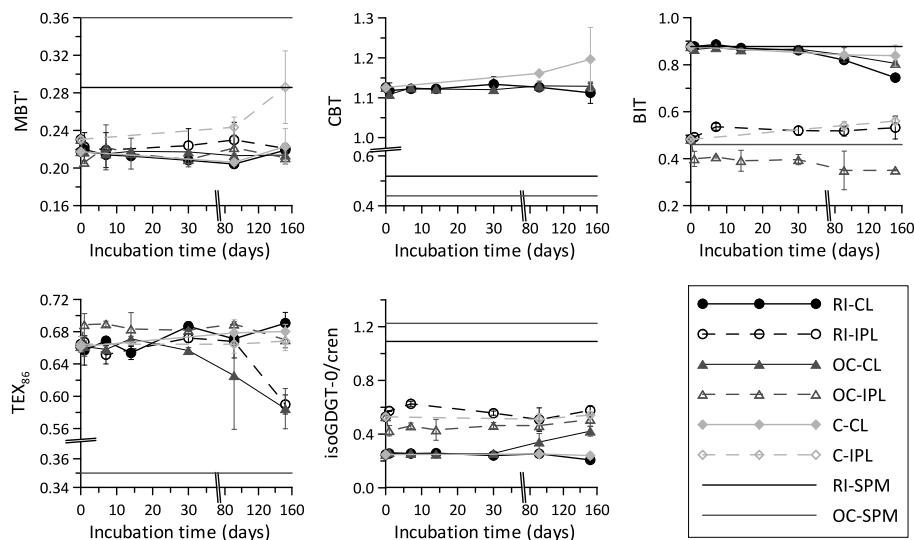


Figure 4. Average distribution and deviation of core lipid (CL) and intact polar lipid (IPL)-derived GDGTs in a sandy loam soil from the Rakaia River catchment, New Zealand, as reflected by the (a) MBT' index, (b) CBT index, (c) BIT index, (d) TEX₈₆ index, and (e) isoGDGT-0/crenarchaeol ratio during incubation ($t = 152$ days) in river (RI), ocean (OC), and distilled (C) water under quasi-aerobic conditions. Horizontal straight lines represent GDGT composition in river (black) and ocean water (grey) suspended particulate material (SPM).

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