

1 **Ocean acidification accelerates dissolution of experimental coral reef communities**

2 Comeau S., Carpenter R. C., Lantz C. A., Edmunds P. J.

3 Department of Biology, California State University, 18111 Nordhoff Street, Northridge,

4 CA 91330-8303, USA.

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6 **Corresponding author:** Steeve Comeau, Email: steve.comeau@csun.edu

7 **Abstract**

8 Ocean acidification (OA) poses a severe threat to tropical coral reefs, yet much of what is
9 know about these effects comes from individual corals and algae incubated in isolation
10 under high pCO₂. Studies of similar effects on coral reef communities are scarce. To
11 investigate the response of coral reef communities to OA, we used large outdoor flumes
12 in which communities composed of calcified algae, corals, and sediment were combined
13 to match the percentage cover of benthic communities in the shallow back reef of
14 Moorea, French Polynesia. Reef communities in the flumes were exposed to ambient (~
15 400 μatm) and high pCO₂ (~ 1300 μatm) for 8 weeks, and calcification rates measured for
16 the constructed communities including the sediments. Community calcification was
17 reduced by 59% under high pCO₂, with sediment dissolution explaining ~ 50% of this
18 decrease; net calcification of corals and calcified algae remained positive, but was
19 reduced by 29% under elevated pCO₂. These results show that despite the capacity of
20 coral reef calcifiers to maintain positive net accretion of calcium carbonate under OA
21 conditions, reef communities might transition to net dissolution as pCO₂ increases,
22 particularly at night, due to enhanced sediment dissolution.

23

24 **1 Introduction**

25 The calcium carbonate framework produced by coral reefs hosts the highest
26 known marine biodiversity, and protects tropical shores from wave erosion (Ferrario et
27 al., 2014). However, in recent decades coral reefs have been impacted by a diversity of
28 disturbances, and now are threatened by an increase in seawater temperature and ocean
29 acidification (OA) (Hoegh-Guldberg et al., 2007; Kleypas and Yates, 2009). OA is
30 caused by the dissolution of atmospheric CO₂ in seawater, which reduces pH, depresses
31 carbonate ion concentration, and increases bicarbonate ion concentration with no change
32 in total alkalinity (Feely et al., 2004). The net effects of OA on coral reefs remain unclear
33 as most studies show a decrease in organismic calcification under OA (Erez et al. 2011;
34 Chan and Connolly, 2012), while recent work describes species-specific responses with
35 some corals and calcifying algae resistant to decreasing pH (Comeau et al., 2013;
36 Takahashi and Kurihara, 2013). Critically, most of these studies have been performed on
37 individuals maintained in isolation in laboratory conditions.

38 The results from laboratory studies are valuable, but to examine the potential for
39 emergent properties of coral reefs exposed to OA effects, now it is necessary to scale up
40 from individual- to community-level experiments (Leclercq et al., 2002; Jokiel et al.,
41 2008; Andersson et al., 2009; Dove et al., 2013; Edmunds et al., 2013). Generally there
42 are three complementary approaches to studying the responses of coral reef communities
43 to OA. First, *in situ* observations of communities living in naturally acidified water
44 (Fabricius et al., 2011) due to volcanic activities or local conditions (Shamberger et al.,
45 2014). Second, carbonate chemistry can be manipulated directly *in situ* (Kline et al.,
46 2012), although this approach is challenging technically and has yet to be used to study

47 intact communities. Third, reef communities can be created *ex situ* (Andersson et al.,
48 2009; Dove et al., 2013) to allow precise control of the physical parameters predicted
49 under future OA conditions. For our experiment, we chose to construct *ex situ*
50 communities and used, for the first time, large outdoor flumes (after Atkinson and Bilger,
51 1992) to investigate the effects of OA on coral reef communities.

52 In addition to corals and macroalgae, it is important to incorporate sediments in
53 OA experiments, as this component of reef ecosystems may be sensitive to decreasing pH
54 (Cyronak et al. 2013a, b; Andersson et al., 2009). Dissolution occurs on coral reefs in
55 sediment pore-waters, or in particular microenvironments where pCO₂ is elevated due to
56 biological activity (Andersson and Gledhill, 2013). Observations in Bermuda have shown
57 that the dissolution of Mg-calcite sediments occurs in a location with seawater pCO₂
58 naturally elevated to values expected by the end of the century (Andersson et al., 2007).
59 Further, *in situ* manipulations show that elevated pCO₂ (~ 800 µatm) can transition the
60 calcification budget of coral reef sediments from net precipitation to net dissolution
61 (Cyronak et al., 2013a). Increasing pCO₂ likely will lead to increasing dissolution and
62 decreased precipitation of calcium carbonate, resulting in coral reef community
63 calcification changing from net precipitation to net dissolution (Yates and Halley, 2006;
64 Silvermann et al., 2009; Andersson et al., 2009). Given the aforementioned results that
65 highlight the importance of sediments in the community calcification of entire coral reefs,
66 we included sediment chambers in our flumes (Fig. 1) to integrate reef carbonate
67 sediments into the analysis of OA effects on communities under ecologically relevant
68 conditions.

69 Our experiment investigated the response to OA of constructed reef communities
70 representative of present day back reef communities of Moorea. Communities were
71 incubated for 8 weeks in two flumes at ambient seawater pCO₂ (~ 400 μatm) and two
72 flumes at an elevated pCO₂ (~1300 μatm) under natural lighting, controlled temperature,
73 and water flow similar to those experienced on a shallow reef flat (Atkinson and Bilger,
74 1992; Carpenter and Williams, 2007). Calcification was measured at three levels of
75 biological function using the alkalinity anomaly technique: whole community, sediments,
76 and macro-calcifiers (i.e., corals and calcified algae as determined by subtraction).

77

78 **2 Materials and methods**

79 **2.1 Collection and sample preparation**

80 This study was carried out in August-October 2013 in Moorea, French Polynesia,
81 using organisms collected from the back reef of the north shore at ~ 1–2 m depth. The
82 organisms used to construct communities in outdoor flumes were assembled to match the
83 contemporary (in 2013) mean cover of a back reef in Moorea (Carpenter, 2014;
84 Edmunds, 2014). Coral communities were built from the four dominant coral taxa found
85 on the back reefs of Moorea: massive *Porites* spp. (11% cover in 2013), *Porites rus* (6%),
86 *Montipora* spp. (3%), and *Pocillopora* spp. (2%), that together accounted for 98% of the
87 coral cover in this habitat in 2013. In addition to corals, 6% of the surface was covered by
88 crustose coralline algae (66% *Porolithon onkodes* and 33% *Lithophyllum flavescens*), and
89 5% by rubble (dead coral skeletons). After collection of corals and algae (all ~ 10 × 10
90 cm), they were returned to the Richard B. Gump South Pacific Research and attached to

91 plastic supports using epoxy glue. Following preparation, samples were left to recover in
92 a seawater table for 3 d.

93 Sediments were collected from the lagoon on the north shore, ~ 200 m from the
94 reef crest, at 2-m depth using 24 custom made boxes ($0.4 \times 0.3 \times 0.3$ m). Sediment boxes
95 were inserted into the sediment and left *in situ* for 4 d to allow chemical stratification in
96 the sediment to re-establish before transferring the boxes to the flumes. It was not
97 possible to subsample these boxes to quantify the stratification of the sediment, and
98 therefore we assume that 4 d was adequate for stratification to be re-established.

99 The four outdoor flumes consisted of a working section measuring $5.0 \times 0.3 \times 0.3$
100 m. Water was re-circulated using water pumps (W. Lim Wave II 373 J s^{-1}) to obtain a 10
101 cm s^{-1} flow. Flow was measured across the working section of the flume using a Nortek
102 Vectrino Acoustic Doppler Velocimeter. At each end of the flume seawater passed
103 through an 88-cm long transition section (circular to rectangular) that housed 20-cm
104 (length) flow straighteners made of stacked, 3-cm diameter PVC pipe, and then into a
105 12.5-cm (diameter) return section. Fresh sand-filtered seawater, pumped from Cook's Bay
106 at 12-m depth, was dispensed continuously into the flume at 5 L min^{-1} . Flumes
107 experienced natural sunlight that was attenuated using fiberglass screens to maintain
108 irradiances similar to ambient irradiances in the back reefs of Moorea (daily maximum of
109 $\sim 1500 \mu\text{mol photons m}^{-2} \text{ cm}^{-1}$ over the incubation period determined with a 4π quantum
110 sensor LI-193 and a LiCor LI-1400 meter). Temperature in the flumes was maintained at
111 $\sim 27 \text{ }^\circ\text{C}$ to match the ambient temperature in the back reef of Moorea in September-
112 October.

113

114 2. 2 Carbonate chemistry control and measurements

115

116 As the pCO₂ level in the back reef of Moorea is close to open-ocean and current
117 atmospheric values (e.g., Comeau et al. 2014a), pCO₂ levels for the incubations were
118 chosen to match ambient pCO₂ (~ 400 μatm) and the pCO₂ expected in the atmosphere by
119 the end of the present century under a pessimistic scenario (Representative Concentration
120 Pathway 8.5, ~1300 μatm, Moss et al., 2010). pCO₂ in the flumes was controlled using a
121 pH-stat (Aquacontroller, Neptune systems, USA) that actuated the bubbling of either pure
122 CO₂ or CO₂-free air into the seawater. To match the natural diel variation in pH in the
123 back reef of Moorea (Hofmann et al., 2011; Comeau et al., 2014a) pH was maintained 0.1
124 unit lower at night (from 18:00 to 6:00) than during the day.

125 pH was measured daily using a portable pH meter (Orion 3-stars, Thermo-
126 Scientific, USA) fitted with a DG 115-SC pH probe (Mettler Toledo, Switzerland)
127 calibrated every other day with Tris/HCl buffers (Dickson et al., 2007). pH also was
128 measured spectrophotometrically using m-cresol dye (Dickson et al., 2007) at regular
129 intervals. pH measured spectrophotometrically or using a pH electrode provided similar
130 results with means differing < 0.01 pH unit. Measurement of total alkalinity (A_T) was
131 made using open-cell potentiometric titrations (Dickson et al., 2007) using 50-mL
132 samples of seawater collected every 2-3 d. Titrations of certified reference materials
133 provided by A. G. Dickson (batch 122) yielded A_T values within 3.5 μmol kg⁻¹ of the
134 nominal value (SE = 3.1 μmol kg⁻¹; n = 14). Parameters of the carbonate system in
135 seawater were calculated using the R package seacarb (Lavigne and Gattuso, 2013).

136

137

138 **2.3 Calcification measurements and sediment analysis**

139 Calcification rates were measured using the total alkalinity anomaly method
140 (Chisholm and Gattuso, 1991). Calcification measurements were made every 7 d on the
141 constructed community, and in the analysis of sediments alone, after 7, 30, and 56 d of
142 treatments. During incubations, the addition of seawater was stopped so that each flume
143 operated in a closed loop; seawater samples for A_T then were taken every 3 h during the
144 day and every 6 h at night. To maintain A_T and nutrients close to ambient levels, water in
145 the flumes was refreshed every 3-6 h for 30 min. Regular refreshing limited changes in
146 alkalinity during incubations to $< 50\text{-}100 \mu\text{mol kg}^{-1}$, which corresponded to variations in
147 aragonite saturation state (Ω) of $< 0.1\text{-}0.2$. Nutrient changes in the flumes were monitored
148 during four incubations and the changes in nitrate and ammonium during incubations
149 were $< 2 \mu\text{mol L}^{-1}$. To conduct incubations with sediments alone, corals and coralline
150 algae were removed from the flumes for 24 h and held in a separate tank where
151 conditions were identical to those in the flumes. Corals and coralline algal calcification
152 was calculated by subtracting the mean light and dark net calcification of the sediments
153 from the community calcification. For both corals and algae, buoyant weight (Davies,
154 1989) was recorded before and after the 8-week treatments and converted to dry weight
155 to quantify the contribution of each functional group to the calcification budget. Sediment
156 grain size of each flume was analyzed in triplicate using sediment sieves. Three vertical
157 cross sections of sand ($\sim 600\text{g}$) were collected from each flume sediment chamber and
158 dried at 60°C to remove moisture. Sand then was sieved through five separate sediment
159 sieves ($149 \mu\text{m}$, $420 \mu\text{m}$, $840 \mu\text{m}$, $3360 \mu\text{m}$) yielding six size class fractions for each
160 flume ($n = 3$).

161 2.4 Statistical analysis

162 All analyses were performed using R software (R Foundation for Statistical
163 Computing), and assumptions of normality and equality of variance were evaluated
164 through graphical analyses of residuals. Calcification rates were analyzed using a
165 repeated measure ANOVA in which the within subject factor was time (week), pCO₂ was
166 a fixed effect, and duplicate flumes was a nested effect.

167

168 3 Results

169 3.1 Carbonate chemistry and organism condition

170 Mean pCO₂ in the four flumes during the 8-week incubation was 456 ± 21 μatm
171 and 451 ± 21 μatm in the ambient treatments, and 1329 ± 28 μatm and 1306 ± 41 μatm in
172 the high pCO₂ treatments (± SE, n = 42). pCO₂ differed between treatments (repeated
173 measure ANOVA, $F_{1,232} = 734.38$, $p < 0.001$), but there was no difference within
174 treatments ($F_{2,232} = 0.16$, $p = 0.852$). Communities were maintained in conditions within
175 the flumes that were super-saturated with respect to aragonite, as $\Omega_{\text{arag}} \sim 3.5$ under
176 ambient conditions, and ~ 1.6 in the high pCO₂ treatment.

177 No *Pocillopora* spp. and *Montipora* spp. colonies died during the 8-week
178 treatments, but 10% of the *Porites* pooled across flumes died by the end of the
179 experiment, regardless of treatment, because of an outbreak of corallivorous nudibranchs
180 feeding on this taxon (*Phestilla* spp.). Coralline algae ($\sim 70\%$) died at the end of the
181 incubation, which was likely due to sediment abrasion. No difference in mortality or
182 bleaching was observed between treatments for corals and calcified algae.

183

184 3.2 Community

185 Net calcification was higher at ambient versus high pCO₂ (Fig. 2A), both during
186 the day and night (repeated measures ANOVA, $F_{1,2} = 84.9$, $p = 0.012$ and $F_{1,2} = 44.9$,
187 0.022 , respectively); there were no differences between flumes within each treatment so
188 the nested factor was removed from the final analysis. At night, treatment effects were
189 more striking than during the day, as calcium carbonate dissolution exceeded
190 precipitation at high pCO₂ (-1.6 ± 0.9 gCaCO₃ m⁻² d⁻¹), whereas net calcification
191 remained positive at ambient pCO₂ (2.7 ± 0.6 gCaCO₃ m⁻² d⁻¹) (both means \pm SE, n =
192 16). Calcification integrated over 24 h highlighted the difference between treatments ($F_{1,2}$
193 = 869.2 , $p = 0.001$), with calcification 59% lower at high pCO₂ than at ambient pCO₂.

194

195 3.3 Sediments

196 Sediment grain sizes in the flumes were similar between flumes and fractionated
197 (by weight) to $5.3 \pm 0.5\%$ < 149 μ m grain size, $56.5 \pm 1.4\%$ > 149 μ m and < 420 μ m,
198 $25.9 \pm 0.4\%$ > 420 μ m < 840 μ m, $10.1 \pm 0.5\%$ > 840 μ m and < 3360 μ m, and $2.2 \pm 0.9\%$
199 > 3360 μ m. Net calcification of the sediments alone differed between treatments, during
200 the day and night ($F_{1,2} = 344.2$, $p = 0.003$ and $F_{1,2} = 282.6$, $p = 0.003$, respectively) (Fig.
201 2B), but there were no differences between flumes within each treatment, hence the
202 nested factor was removed from the final analysis. Net calcification pooled among
203 treatments was negative during the day (-1.0 ± 1.3 gCaCO₃ m⁻² d⁻¹) and night (-3.72 ± 0.6
204 gCaCO₃ m⁻² d⁻¹) at high pCO₂, whereas net calcification was positive during the day (1.4
205 ± 1.1 gCaCO₃ m⁻² d⁻¹) and negative at night (-1.0 ± 0.8 gCaCO₃ m⁻² d⁻¹) in the ambient
206 treatment. When calcification was integrated over 24 h, pCO₂ effects were significant ($F_{1,$

207 $2 = 886.5, p = 0.001$), with dissolution exceeding precipitation at high $p\text{CO}_2$ (-2.3 ± 1.1
208 $\text{gCaCO}_3 \text{ m}^{-2} \text{ d}^{-1}$), and a nearly balanced calcification budget under ambient $p\text{CO}_2$ ($0.2 \pm$
209 $0.9 \text{ gCaCO}_3 \text{ m}^{-2} \text{ d}^{-1}$).

210

211 **3.4 Corals and calcifying algae**

212 The total net calcification of corals and calcifying algae was estimated by
213 subtracting the mean sediment calcification rates from the total community calcification
214 in each flume. Net calcification of the corals and calcifying algae differed between
215 treatments during the day ($F_{1,2} = 32.3, p = 0.030$) and night ($F_{1,2} = 22.9, p = 0.041$) (Fig.
216 2C). In contrast to the whole community and the sediments alone, net calcification of
217 corals and calcifying algae was positive at night at high $p\text{CO}_2$ ($1.1 \pm 0.5 \text{ gCaCO}_3 \text{ m}^{-2}$ in
218 12 h), but was 24% and 44% lower at high $p\text{CO}_2$ compared to ambient $p\text{CO}_2$ during the
219 day and night, respectively. Net calcification integrated over 24 h also differed between
220 treatments ($F_{1,2} = 2569, p < 0.001$) with calcification at ambient $p\text{CO}_2$ 29% higher than at
221 high $p\text{CO}_2$.

222 Calcification of the constructed reef communities was driven principally by
223 corals, since their contribution to the calcification budget, based on dry weight calculated
224 from the changes in buoyant weight, was $\sim 98\%$ of the total (Fig. 3). Massive *Porites* spp.
225 was the main contributor among the corals, with an increased contribution to the
226 calcification budget at high $p\text{CO}_2$ (40% at ambient $p\text{CO}_2$, and 48.5% at high $p\text{CO}_2$, Fig.
227 3). In contrast, the importance of *P. rus*, *Montipora* spp., and *Pocillopora* spp. was
228 reduced at high $p\text{CO}_2$. The small contribution of coralline algae to the calcification
229 budget was due to high mortality perhaps leading to potential dissolution during the last

230 weeks of the incubation. Furthermore, while the ratio of planar area to surface area for
231 crustose coralline algae is close to one, corals have a disproportionately large surface area
232 to planar area ratio due to their three-dimensional structure. With such a large actual
233 surface area, the corals made a large contribution to the calcification budget of the
234 communities assembled in the flumes.

235

236 **4 Discussion**

237 Using outdoor flumes, we show that the effects of OA on coral reef communities
238 are greater than estimates obtained by summing results obtained by incubating organisms
239 in isolation under similar conditions and assuming their contribution to community
240 calcification is proportional to their planar cover. Indeed, at the community level, the
241 reduction in net calcification attributed to high pCO₂ was greater than the mean reduction
242 of 26% calculated in a recent meta-analysis of the effects of future conditions (~1300
243 µatm pCO₂) based on the consequences of high pCO₂ on organismic calcification (Chan
244 and Connolly, 2013). This discrepancy likely is not caused by experimental bias, as rates
245 of net community calcification in the flumes in the ambient treatment were similar to
246 rates measured for back reef communities on the north shore of Moorea. For instance, in
247 2012 and 2013 we measured calcification rates during the day that ranged from 5 to 25
248 gCaCO₃ m⁻² d⁻¹ (R.C. Carpenter, unpublished data), which spans the rates measured in
249 flumes during the present study (i.e., 13.9 gCaCO₃ m⁻² d⁻¹ in the light, Fig. 2A). Net
250 community calcification for the back reef of Moorea in 1991 (~ 19–25 gCaCO₃ m⁻² d⁻¹;
251 Gattuso et al., 1996) also was similar to the rates measured in the flumes (this study) and
252 in the field as described above. Rates of calcification in the present study under ambient

253 conditions also are similar to the $7.9 \text{ gCaCO}_3 \text{ m}^{-2} \text{ d}^{-1}$ reported by Andersson et al. (2009)
254 for a reef community from Kaneohe Bay (Hawaii) that was assembled and incubated in
255 mesocosms. However, while community calcification was still positive under high pCO_2
256 in the present study, Andersson et al. (2009) measured negative calcification (i.e., net
257 dissolution) in their coral reef communities incubated at a pCO_2 twice that of current
258 ambient values. The differences between the present study and that of Andersson et al.
259 (2009) may be due to methodological effects. Andersson et al. (2009) manipulated pH
260 through acid additions (we used CO_2 bubbling), and also used a different assemblage of
261 species and sediments in dissimilar proportions compared to the present study.

262 The discrepancy in the evaluation of the effects of high pCO_2 at the community
263 level (the present study) versus organismic level (previous studies) was the result of
264 dissolution of sediments that represented up to 50% of the decrease in calcification at
265 high pCO_2 . Increased dissolution of sediments at high pCO_2 likely was caused by the
266 reduction of the seawater saturation state in the flumes, as we did not detect any
267 difference in respiration and photosynthesis under elevated pCO_2 (results not shown) that
268 could also affect sediment dissolution (Andersson and Gledhill, 2013). Our results reveal
269 the sensitivity of carbonate sediments to dissolution at elevated pCO_2 , and they are in
270 agreement with a recent manipulative experiment conducted on Heron Island, where
271 dissolution of in situ areas of sand (1.7 m depth) exceeded precipitation at $\text{pCO}_2 > 500$
272 μatm (Cyronak et al., 2013a). During a mesocosm experiment, Dove et al. (2013) also
273 demonstrated that a pH of 7.7 caused a change in sediment granularity to favor small-
274 grained (i.e., $\leq 1 \text{ mm}$) sediments as a result of dissolution or increased bioerosion of
275 larger grains. In this case, bioerosion was more likely than dissolution, as dissolution

276 would favor a loss of the smallest grains as a result of their higher surface area to volume
277 ratio. Size-frequency distribution of sediment grain was not different between treatments
278 at the end of our incubations and therefore is unlikely to have affected the treatment
279 effects we detected. Sensitivity of coral reef communities to dissolution has been shown
280 previously for communities constructed in mesocosms in Hawaii, where dissolution (-3.6
281 $\text{mmol CaCO}_3 \text{ m}^{-2} \text{ h}^{-1}$) was detected at night under conditions of double ambient pCO_2
282 (Andersson et al., 2009). In this case, dissolution was attributed to the thin layer of
283 sediment that accumulated at the bottom of the mesocosms (Andersson et al., 2009).

284 In addition to chemical dissolution occurring in the communities constructed in
285 the present study, we cannot exclude the possibility that at least some of the apparent
286 community dissolution was caused by enhanced bioerosion, which for example
287 previously has been shown to occur when blocks of *Porites lobata* are incubated under 750
288 $\mu\text{atm pCO}_2$ for 3-month (Tribollet et al. 2009). In future work it will be important to
289 census the fragments of coral and rock to quantify the presence of bioeroders and their
290 relative contribution to dissolution under ambient and OA conditions.

291 When the effect of sediment dissolution was subtracted from the overall net
292 calcification rate for the communities assembled in our flumes, corals and coralline algae
293 alone exhibited a decrease in net calcification of 29% over 24 h at elevated pCO_2 versus
294 ambient pCO_2 . Such a decrease falls within the range of values we have previously
295 reported for organismic effects of high pCO_2 , in which the calcification rates of 16
296 calcifiers in Moorea declined 0–40% at $1300 \mu\text{atm pCO}_2$ compared to ambient pCO_2
297 (Comeau et al., 2013; Comeau et al., 2014b). It is also within the range of the predicted
298 changes for calcification of corals under a tripling of pCO_2 (relative to present values)

299 estimated by meta-analysis (i.e., a ~ 26% reduction; Chan and Connolly, 2012). As the
300 decrease in calcification recorded in the present study for corals and coralline algae alone
301 was within the range of previous studies, this supports our assumptions that calcification
302 of macro-calcifiers is equal to the difference between net sediment calcification and net
303 community calcification. This “subtraction method” for calculating the calcification rate
304 of corals and coralline algae included in community experiments has some limitations, as
305 it assumes that the calcification of the sediments and the macro-calcifiers are
306 independent. Such interactions might occur, for example, if dissolution of the sediment
307 would locally enhance total alkalinity that would, in turn, favor calcification by macro-
308 calcifiers. Testing for such feedback mechanisms among the different compartments of
309 the communities we built was beyond the scope of the present study, but it will be
310 important to consider such effects in future experiments.

311 Our results demonstrate the utility of large outdoor flumes for investigating the
312 responses of coral reef communities to OA. Similar rates of calcification in the field and
313 in the flumes suggest that the communities assembled in the flumes effectively mimicked
314 both the biological communities and the physical and chemical conditions characterizing
315 the back reef of Moorea. The ability to create ecologically relevant flow conditions in the
316 flumes is likely to be especially important for establishing ecological relevance, as flow is
317 critical in modulating mass transfer and metabolism of coral reef organisms (Atkinson
318 and Gilmer, 1992; Carpenter and Williams, 2007, Comeau et al. 2014c). In the case of
319 stony corals, for example, high flow speeds are suspected to enhance coral calcification
320 by favoring proton export from coral tissue through boundary layers (Jokiel, 2011; Jokiel
321 et al., 2014), and for coralline algae, might increase sensitivity to OA by reducing the

322 capacity to maintain high pH in the diffusion boundary layer adjacent to the algal thallus
323 (Cornwall et al., 2013, 2014).

324

325 **5 Conclusion**

326 The present results suggest that, despite a reduction in calcification, calcifying reef
327 organisms may maintain net positive calcification under $p\text{CO}_2$ as high as $1300 \mu\text{atm}$.
328 However, at the scale of coral reef communities in back reef habitats, community net
329 calcification will be affected strongly and negatively, at least for reefs similar in
330 community structure to those in Moorea in 2013. The present experiments demonstrate
331 the importance of living organisms on benthic surfaces in maintaining a positive balance
332 between precipitation and dissolution of calcium carbonate. Whereas several reefs around
333 the world are already at the threshold between precipitation and dissolution of calcium
334 carbonate (Silverman et al., 2009, 2014), the susceptibility of coral reefs to net
335 dissolution in the future likely will be linked directly to the proportion of the reef covered
336 by macro-calcifiers and sediments. In addition to dissolution, it also is possible that coral
337 reefs will be exposed to increased bioerosion at high $p\text{CO}_2$ (Wisshak et al., 2012; Crook
338 et al., 2013) that will decrease the integrity of the carbonate framework. Our results
339 suggest that under OA conditions anticipated by the end of the current century, at least
340 some tropical corals and calcifying algae will persist, but the function of the coral reef
341 community as a net precipitator of calcium carbonate and as a physical structure to
342 protect coasts against erosion (Ferrario et al., 2014) will be challenged.

343 **Authors contributions:** S.C. designed and performed experiments, analyzed data and
344 wrote the paper; C.L. performed experiments and wrote the paper; B.C. and P.E.
345 designed experiments, analyzed data and wrote the paper.

346

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353 **References**

- 354 Andersson, A. J., Bates, N. R., and Mackenzie, F. T.: Dissolution of carbonate sediments
355 under rising pCO₂ and ocean acidification: observations from Devil's Hole,
356 Bermuda, *Aquatic. Geochem.*, 13, 237–264, 2007.
- 357 Andersson, A. J., Kuffner, I. B., Mackenzie, F. T., Jokiel, P. L., Rodgers, K. S., and Tan,
358 A.: Net Loss of CaCO₃ from a subtropical calcifying community due to seawater
359 acidification: mesocosm-scale experimental evidence, *Biogeosciences*, 6, 1811–
360 1823, 2009.
- 361 Andersson, A. J., and Gledhill, D.: Ocean acidification and coral reefs: effects on
362 breakdown, dissolution, and net ecosystem calcification, *Ann. Rev. Mar. Sci.*, 5,
363 321–348, 2013.
- 364 Atkinson, M. J. and Bilger, R. W.: Effects of water velocity on phosphate uptake in coral
365 reef-flat communities, *Limnol. Oceanogr.*, 37, 273-279, 1992.
- 366 Carpenter, R. C. and Williams, S. L.: Mass transfer limitation of photosynthesis of coral
367 reef algal turfs, *Mar. Biol.*, 151, 435–450, 2007.
- 368 Carpenter, R. C. of Moorea Coral Reef LTER.: MCR LTER: Coral Reef: Long-term
369 population and community dynamics: benthic algae and other community
370 components, ongoing since 2005. knb-lter-mcr.8.27.
371 <http://metacat.lternet.edu/knb/metacat/knb-lter-mcr.8.27/lter>, 2014.
- 372 Chan, N. C. S. and Connolly, S. R.: Sensitivity of coral calcification to ocean
373 acidification: a meta-analysis, *Glob. Change Biol.*, 19, 282–290, 2013.

374 Chisholm, J. R. M. and Gattuso, J.-P.: Validation of the alkalinity anomaly technique for
375 investigating calcification and photosynthesis in coral reef communities, *Limnol.*
376 *Oceanogr.*, 36, 1232–1239, 1991.

377 Comeau, S., Edmunds, P. J., Spindel, N. B., and Carpenter, R. C.: The responses of eight
378 coral reef calcifiers to increasing partial pressure of CO₂ do not exhibit a tipping
379 point, *Limnol. Oceanogr.*, 58, 388–398, 2013.

380 Comeau, S., Edmunds, P. J., Spindel, N. B., and Carpenter, R. C.: Diel pCO₂ oscillations
381 modulate the response of the coral *Acropora hyacinthus* to ocean acidification, *Mar.*
382 *Ecol. Prog. Ser.*, 501, 99–111, 2014a.

383 Comeau, S., Edmunds, P. J., Spindel, N. B., and Carpenter, R. C.: Fast coral reef
384 calcifiers are more sensitive to ocean acidification in short-term laboratory
385 incubations, *Limnol. Oceanogr.*, 59, 1081–1091, 2014b.

386 Comeau, S., Edmunds, P. J., Lantz, C. A., and Carpenter, R. C.: Water flow modulates
387 the response of coral reef communities to ocean acidification, *Scientific Reports*, 4,
388 doi:10.1038/srep06681, 2014c.

389 Cornwall, C. E., Hepburn, C. D., Pilditch, C. A., and Hurd, C. L.: Concentration
390 boundary layers around complex assemblages of macroalgae: Implications for the
391 effects of ocean acidification on understory coralline algae, *Limnol. Oceanogr.*, 58,
392 121–130, 2013.

393 Cornwall, C. E., Boyd, P. W., McGraw, C. M., Hepburn, C. D., Pilditch, C. A., Morris, J.
394 N., Smith, A. M., and Hurd, C. L.: Diffusion boundary layers ameliorate the
395 negative effects of ocean acidification on the temperate coralline macroalga

396 *Arthrocardia corymbosa*, PLoS One, 9, e97235. doi:10.1371/journal.pone.0097235,
397 2014.

398 Crook, E. D., Cohen, A. L., Rebolledo-Vieyra, M., Hernandez, L., and Paytan, A.:
399 Reduced calcification and lack of acclimatization by coral colonies growing in areas
400 of persistent natural acidification, Proc. Natl. Acad. Sci., 110, 11044–11049, 2013.

401 Cyronak, T., Santos, I. R., and Eyre, B. D.: Permeable coral reef sediment dissolution
402 driven by elevated pCO₂ and pore water advection, Geophys. Res. Lett., 40, 4876–
403 4881, 2013a.

404 Cyronak, T., Santos, I. R., McMahon, A., and Eyre, B. D.: Carbon cycling hysteresis in
405 permeable carbonate sands over a diurnal cycle: implications for ocean acidification,
406 Limnol. Oceanogr., 58, 131–143, 2013b.

407 Davies, P. S.: Short-term growth measurements of corals using an accurate buoyant
408 weighing technique, Mar. Biol., 101, 389–95, 1989.

409 Dickson, A. G., Sabine, C. L., and Christian, J. R. (Eds.): Guide to best practices for CO₂
410 measurements, PICES Special Publication, 3, 191 pp., 2007.

411 Dove, S. G., Kline, D. I., Santos, S., Angly, F. E., Tyson, G. W., and Hoegh-Guldberg,
412 O.: Future reef decalcification under a business-as-usual CO₂ emission scenario,
413 Proc. Natl. Acad. Sci., 110, 15342–15347, 2013.

414 Edmunds, P. J., Carpenter, R. C., and Comeau, S.: Understanding the threats of ocean
415 acidification to coral reefs, Oceanogr., 26, 149–152, 2013.

416 Edmunds, P. J. of Moorea Coral Reef LTER.: MCR LTER: Coral Reef: Long-term
417 population and community dynamics: corals. knb-lter-mcr.4.31
418 <http://metacat.lternet.edu/knb/metacat/knb-lter-mcr.4.31/lter>, 2014.

419 Erez, J., Reynaud, S., Silverman, J., Schneider, K., and Allemand, D.: Coral
420 calcification under ocean acidification and global change, in: Coral reefs: An
421 ecosystem in transition, Dubinsky, Z. and Stambler, N. (Eds.), Springer, Germany,
422 151–176, 2011.

423 Fabricius, K. E., Langdon, C., Uthicke, S., Humphrey, C., Noonan, S., De'ath, G.,
424 Okazaki, R., Muehllehner, N., Glas, M. S., and Lough, J.M.: Losers and winners in
425 coral reefs acclimatized to elevated carbon dioxide concentrations, *Nature Clim.*
426 *Change*, 1, 165–169, 2011.

427 Feely, R. A., Sabine, C. L., Lee, K., Berelson, W., Kleypas, J., Fabry, V. J., and Millero,
428 F.J.: Impact of anthropogenic CO₂ on the CaCO₃ system in the oceans, *Science*, 305,
429 362–366, 2004.

430 Ferrario, F., Beck, M. W., Storlazzi, C. D., Micheli, F., Shepard, C. C., and Airoidi, L.:
431 The effectiveness of coral reefs for coastal hazard risk reduction and adaptation,
432 *Nature Comm.*, 5, doi:10.1038/ncomms4794, 2014.

433 Gattuso, J.-P., Pichon, M., Delesalle, B., Canon, C., and Frankignoulle, M.: Carbon
434 fluxes in coral reefs. I. Lagrangian measurement of community metabolism and
435 resulting air-sea CO₂ disequilibrium, *Mar. Ecol. Progr. Ser.*, 145, 109–121, 1996.

436 Hoegh-Guldberg, O., Mumby, P. J., Hooten, A. J., Steneck, R. S., Greenfield, P., Gomez,
437 E., Harvell, C. D., Sale, P. F., Edwards, A. J., Caldeira, K., Knowlton, N., Eakin, C.
438 M., Iglesias-Prieto, R., Muthiga, N., Bradbury, R. H., Dubi, A., and Hatzioios M. E.:
439 Coral reefs under rapid climate change and ocean acidification, *Science*, 318, 1737–
440 1742, 2007.

441 Hofmann, G. E., Smith, J. E., Johnson, K. S., Send, U., Levin, L. A., Micheli, F., Paytan,
442 A., Price, N. N., Peterson, B., Takeshita, Y., Matson, P. G., Derse Crook, E.,
443 Kroeker, K. J., Gambi, M. C., Rivest, E. B., Frieder, C. A., Yu, P. C., and Martz, T.
444 R.: High-frequency dynamics of ocean pH: a multi-ecosystem comparison. PLoS
445 ONE, 6, e28983. doi:10.1371/journal.pone.0028983, 2011.

446 Jokieli, P. L., Rodgers, K. S., Kuffner, I. B., Andersson, A. J., Cox, E. F., and Mackenzie,
447 F. T.: Ocean acidification and calcifying reef organisms: a mesocosm investigation.
448 Coral Reefs, 27, 473–483, 2008.

449 Jokieli, P.L.: The reef coral two compartment proton flux model: A new approach relating
450 tissue-level physiological processes to gross corallum morphology, J. Exp. Mar.
451 Biol. Ecol., 409, 1–12, 2011.

452 Jokieli, P. L., Jury, C. P., and Rodgers, K. S.: Coral-algae metabolism and diurnal changes
453 in the CO₂ -carbonate system of bulk sea water, PeerJ, 2, e378,
454 doi:10.7717/peerj.378, 2014.

455 Kleypas, J. and Yates, K.: Coral reefs and ocean acidification, Oceanogr., 22, 108–117,
456 2009.

457 Kline, D. I., Teneva, L., Schneider, K., Miard, T., Chai, A., Marker, M., Headley, K.,
458 Opdyke, B., Nash, M., Valetich, M., Caves, J. K., Russell, B. D., Connell, S. D.,
459 Kirkwood, B. J., Brewer, P., Peltzer, E., Silverman, J., Caldeira, K., Dunbar, R. B.,
460 Koseff, J. R., Monismith, S. G., Mitchell, B. G., Dove, S., Hoegh-Guldberg, O.: A
461 short-term in situ CO₂ enrichment experiment on Heron Island (GBR), Scientific
462 Reports, 2, 413, doi:10.1038/srep00413, 2012.

463 Lavigne, H. and Gattuso, J.-P.: seacarb, seawater carbonate chemistry with R. R package
464 version 2. 4. 10. <http://CRAN.R-project.org/package=seacarb>, 2013.

465 Leclercq, N., Gattuso, J.-P., and Jaubert, J.: Primary production, respiration, and
466 calcification of a coral reef mesocosm under increased CO₂ partial pressure, *Limnol.*
467 *Oceanogr.*, 47, 558–564, 2002.

468 Moss, R. H., Edmonds, J. A., Hibbard, K. A., Manning, M. R., Rose, S. K., VanVuuren,
469 D. P., Carter, T. R., Emori, S., Kainuma, M., Kram, T., Meehl, G. A., Mitchell, J. F.,
470 Nakicenovic, N., Riahi, K., Smith, S. J., Stouffer, R. J., Thomson, A. M., Weyant, J.
471 P., and Wilbanks, T. J.: The next generation of scenarios for climate change research
472 and assessment, *Nature*, 463, 747–756, 2010.

473 Shamberger, K. E. F., Cohen, A. L., Golbuu, Y., McCorkle, D. C., Lentz, S. J., and
474 Barkley, H. C.: Diverse coral communities in naturally acidified waters of a Western
475 Pacific reef, *Geophys. Res. Lett.*, 41, 499–504, 2014.

476 Silverman, J., Lazar, B., Cao, L., Caldeira, K., and Erez, J.: Coral reefs may start
477 dissolving when atmospheric CO₂ doubles, *Geophys. Res. Lett.*, 36, L05606,
478 doi:200910.1029/2008GL036282, 2009.

479 Silverman, J., Schneider, K., Kline, D. I., Rivlin, T., Rivlin, A., Hamylton, S., Lazar, B.,
480 Erez, J., Caldeira, K.: Community calcification in Lizard Island, Great Barrier Reef:
481 A 33 year perspective, *Geochim. Cosmochim. Acta*, 144, 72–81, 2014.

482 Takahashi, A., and Kurihara, H.: Ocean acidification does not affect the physiology of the
483 tropical coral *Acropora digitifera* during a 5-week experiment, *Coral Reefs*, 32,
484 305–314, 2013.

485 Tribollet, A., Godinot, C., Atkinson, M., and Langdon, C.: Effects of elevated pCO₂ on
486 dissolution of coral carbonates by microbial euendoliths, *Global Biogeochem. Cycl.*,
487 23, GB3008. doi:10.1029/2008GB003286, 2009.

488 Wisshak, M., Schönberg, C. H. L., Form, A., and Freiwald, A.: Ocean acidification
489 accelerates reef bioerosion, *PLoS One*, 7, e45124.
490 doi:10.1371/journal.pone.0045124, 2012.

491 Yates, K. K., and Halley, R. B.: CO₃²⁻ concentration and pCO₂ thresholds for
492 calcification and dissolution on the Molokai reef flat, Hawaii, *Biogeosciences*, 3,
493 357–369, 2006.

494 **Table 1.** Mean carbonate chemistry in the four flumes (F1-4) during the 8-week incubation. The partial pressure of CO₂ (pCO₂), the
 495 aragonite saturation state (Ω_{arag}) and the calcite saturation state (Ω_{calc}) were calculated from pH_T, total alkalinity (A_T), temperature and
 496 salinity. The values presented are mean \pm SE (n = 56). SE for salinity was < 0.1.

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498

Flume	Treatment	pH _T	A_T ($\mu\text{mol. kg}^{-1}$)	pCO ₂ (μatm)	Ω_{arag}	Ω_{calc}	Temperature (°C)	Salinity
F1	High pCO ₂	7.603 \pm 0.008	2343 \pm 1	1329 \pm 28	1.60 \pm 0.03	2.41 \pm 0.04	27.0 \pm 0.1	35.9
F2	Ambient	8.010 \pm 0.012	2339 \pm 1	456 \pm 19	3.49 \pm 0.07	5.26 \pm 0.11	26.8 \pm 0.1	35.9
F3	High pCO ₂	7.617 \pm 0.014	2345 \pm 1	1306 \pm 42	1.68 \pm 0.05	2.53 \pm 0.08	27.1 \pm 0.1	35.9
F4	Ambient	8.015 \pm 0.013	2339 \pm 1	451 \pm 18	3.53 \pm 0.07	5.32 \pm 0.11	26.9 \pm 0.1	35.9

501 **Figure legends**
502

503 **Figure 1.** Photographs of the outdoor flumes. a) The flumes consisted of a $5.00 \times 0.30 \times$
504 0.30 m working section, and a lower sediment chamber ($2.50 \times 0.30 \times 0.25$ m) in which
505 sediments were maintained, and together contained ~ 600 L of seawater. b) Communities
506 matching the average composition (in 2013) of the back reef in Moorea were constructed
507 in the flumes.

508

509 **Figure 2.** Calcification in the light, dark, and integrated over 24 h for intact communities
510 (A), sediment (B), and corals and coralline algae (C) maintained under ambient and high
511 $p\text{CO}_2$ ($\sim 1300 \mu\text{atm}$). The grey bars represent the calcification measured in the ambient
512 conditions and the black bars are calcification in the elevated $p\text{CO}_2$ treatment.

513

514 **Figure 3.** Relative contribution of each functional group of corals and calcifying algae to
515 the calcification budget of communities as a function of their contribution to the planar
516 surface area of calcifiers in the flumes. Contribution to the calcification budget was
517 derived from the buoyant weight measurements made on each individual at the beginning
518 and end of the 8-week incubation. The grey (ambient condition) and black (high $p\text{CO}_2$)
519 squares correspond to the mean \pm SD specific contributions of massive *Porites* (mP),
520 *Porites rus* (Pr), *Pocillopora* spp. (Po), *Montipora* spp. (Mo), *Porolithon onkodes* (Ph),
521 and *Lithophyllum flavescens* (Lf). The dashed line corresponds to a contribution to the
522 calcification budget equivalent to the planar surface areas of calcifier in the flumes.

A**B**



