

1 Differences in microbial community composition between injection and
2 production water samples of water-flooding petroleum reservoirs

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17 **Conflict of interest**

18 The authors declare that there is no conflict of interest regarding the publication of

19 this article.

20

21 **Abstract.** Microbial communities in injected water are expected to have significant
22 influence on those of reservoir strata in long-term water-flooding petroleum reservoirs.
23 To investigate the similarities and differences in microbial communities in injected
24 water and reservoir strata, high-throughput sequencing of microbial partial 16S rRNA
25 of the water samples collected from the wellhead and downhole of injection wells,
26 and from production wells in a homogeneous sandstone reservoir and a heterogeneous
27 conglomerate reservoir were performed. The results indicate that a small number of
28 microbial populations are shared between the injected and produced water samples in
29 the sandstone reservoir, whereas a large number of microbial populations are shared
30 in the conglomerate reservoir. The bacterial and archaeal communities in the reservoir
31 strata have high concentrations, which are similar with those in the injected water.
32 However, microbial population abundance exhibited large differences between the
33 injected and produced water samples. The number of shared populations reflects the
34 influence of microbial communities in injected water on those in reservoir strata to
35 some extent, and show strong association with the unique variation of reservoir
36 environments.

37

38 **Keywords** 16S rRNA • Microbial community • Pyrosequencing • Miseq • MEOR

39

40 1 Introduction

41 Water-flooding is an efficient and least expensive oil recovery process that is
42 employed worldwide. Water-flooding is believed to be a continued reinoculation of
43 reservoir with surface microorganisms. After long-term water-flooding, microbial
44 populations possessing exceptional survival abilities in injected water are expected to
45 flow into oil-bearing strata, in where, exogenous and indigenous microbial
46 populations form a new complex ecosystem (Zhang et al., 2012a). When injecting
47 nutrients and air into reservoir strata, these microbial populations can be stimulated,
48 and produce metabolites, such as polysaccharide, surfactants, acid, alcohol, and
49 biogas. Because these metabolites can improve reservoir properties by blocking
50 preferred water flow paths, lowering interfacial tension between brines and the oil
51 phase, and decreasing oil viscosity (Youssef et al., 2009), microbial enhanced oil
52 recovery (MEOR) have been applied to petroleum industry, and is currently studied
53 extensively (Abdel-Waly, 1999;Zhang et al., 2012b;Bao et al., 2009;Gao et al.,
54 2013;Li et al., 2014).

55

56 Microbial populations inhabiting petroleum reservoirs play critical roles in the
57 microbial enhancing of the oil recovery process. As a result, an increasing number of
58 studies, especially those based on culture-independent methods, have been performed,
59 and broaden our knowledge of microbial diversity in oil reservoirs (Al-Bahry et al.,
60 2013;Kumaraswamy et al., 2011;Lenchi et al., 2013;Okoro et al., 2014;Wang et al.,
61 2012). However, to our best knowledge, detailed studies about the effects of the
62 microbial communities in injected water on those in reservoir strata remains poorly
63 understood. Based on the 16S rRNA gene clone library method, several studies to date
64 have suggested that despite being flooded by the same injected water, there is a

65 significant difference in the communities between each production well (Tang et al.,
66 2012;Ren et al., 2011). Zhang et al. compared microbial communities in samples of
67 injection and production wells from reservoirs with different in situ temperatures, and
68 pointed that the effects of microorganisms in the injected waters on microbial
69 community compositions in produced waters are strong associated with reservoir
70 temperature (Zhang et al., 2012a). However, because of the low throughput of the
71 clone library method, many infrequent microbial taxa may not be detected, making it
72 difficult to compare microbial communities in detail.

73

74 If microbial populations in injected water can flow into reservoir strata and reach
75 production wells, is the microbial community in the injected water expected to have a
76 similar community composition to those in the production wells? If there is a large
77 difference in community composition, what is the difference and how many microbial
78 populations are shared? To explore these issues, we investigated the microbial
79 populations and their abundance in injection and production wells in a homogeneous
80 sandstone petroleum reservoir with a permeability of $522 \times 10^{-3} \mu\text{m}^2$ and inter-well
81 spacing of 300–425 m using 16S rRNA pyrosequencing and real-time fluorescent
82 quantitative PCR (qPCR). At the same time, we analyzed microbial communities in
83 water samples collected from the wellhead and downhole of injection wells, and from
84 production wells in a heterogeneous conglomerate water-flooding petroleum reservoir
85 with a permeability of $362 \times 10^{-3} \mu\text{m}^2$ and inter-well spacing of 100–150 m.
86 High-throughput sequencing provides the opportunity to compare microbial
87 populations with unprecedented levels of coverage and detail. The variation
88 in permeability, interwell spacing, and heterogeneity of the reservoirs is benefit for
89 exploring the influence of reservoir physical properties on microbial distribution in

90 injected water and reservoir strata. The similarity among microbial communities was
91 investigated using hierarchical clustering and Principal Coordinate Analysis.
92 Microbial populations were also clustered according to injection and production wells
93 to highlight the populations that showed the highest variability.

94

95 **2 Materials and methods**

96 **2.1 Sampling locations**

97 The Lu and Liu field block reservoirs are located in the Xinjiang Oil Field, in the
98 Junggar Basin of Xinjiang Uygur Autonomous Region, Northwest China. The
99 Lu field block is a homogeneous sandstone reservoir that has been water-flooded
100 since 2001. The depth of the sampling horizon is approximately 1200 m with
101 a temperature of 37°C. The porosity of the reservoir is 29.9%, with an
102 average permeability of $522 \times 10^{-3} \mu\text{m}^2$. The density of the crude oil is 0.846 g/cm^3 ,
103 with an oil viscosity of 18 mPa•s. In the investigated well group (an injection well and
104 four production wells), injection well Lu3084, located in the center of the production
105 wells, has a direct influence on the neighboring producers, with inter-well distances of
106 300–425 m. The Liu field block is a heterogeneous conglomerate reservoir that has
107 been water-flooded for approximately 30 years. The depth of the block horizon is
108 approximately 1088 m, with a temperature of 22.6°C. The porosity of the reservoir is
109 18.96 %, with an average permeability of $362 \times 10^{-3} \mu\text{m}^2$. The oil density is
110 0.912 g/cm^3 , with an oil viscosity of 80 mPa•s. The selected well group includes two
111 injection and three production wells, with an inter-well spacing of 100–150 m. The
112 production well T90 is located at the center of injection wells T86 and T93, while
113 production wells T95 and T96 are located at the edge of the field block and are mainly
114 flooded by injection well T93 (Fig. 1).

115

116 The concentrations of potential nutrient factors, including crude oil properties, total
117 nitrogen (TN), total phosphorus (TP), and ion concentration of formation brines, are
118 listed in Table 2. The differences in geochemical parameters between crude oil
119 samples from the two blocks are not obvious, indicating similar oil formation
120 characteristics and maturity. The crude oil in both blocks had a higher content of
121 saturates and aromatics, which favor the growth of hydrocarbon-degrading bacteria
122 (HDB), and some anaerobes, such as sulfate-reducing bacteria. The cations and anions
123 among the water samples in the two blocks were similar, with lower nitrogen and
124 phosphorus content, which are essential for the survival and growth of
125 microorganisms. The lack of nitrogen and phosphorus implies a low metabolism level
126 of microorganisms.

127

128 **2.2 Water samples collection and DNA extraction**

129 Based on tracer techniques, the time interval for injected water to flow from an
130 injection well into neighboring production wells was approximately 30–45 days in the
131 sandstone reservoir, and 7–10 days in the conglomerate reservoir (data provided by
132 the Xinjiang Oil Field Company). Injected water from the sandstone reservoir was
133 collected on three occasions every 15 days between October 2012 and November
134 2012, and the produced water samples (from the reservoir strata) were collected along
135 with the second injected water sample on three occasions at a 30-day interval. All the
136 injected and produced water samples were collected randomly from sampling valves
137 located on the wellhead. In the conglomerate reservoir, the injected water samples
138 were collected in November 2011 from the wellhead and the zone close to downhole
139 (obtained by backflow, that is, the injected water flowed upward through the injection

140 well) of the injection wells. Seven days later, the produced water samples were
141 collected from neighboring production wells on three occasions at a 7-day
142 interval. The collected water samples were completely filled into 15 L sterilized
143 plastic bottles, which were immediately capped and sealed to avoid contamination and
144 oxygen intrusion.

145

146 For DNA extraction, the residual oil was first removed by heating the sample to 60°C
147 for 30 min and by phase separation in sterilized separatory funnels. Microbial cells
148 were then collected from 5 L of each water sample by centrifugation at 4°C for 15
149 min at 10,000 × g in a high-speed centrifuge (Beckman, CA 92821, USA). The cell
150 deposits collected from the same sampling location were mixed and resuspended with
151 TE buffer (Tris 80 mM, EDTA 40 mM, pH 8.0), and then lysed using a mini
152 bead-beater (BioSpec, Bartlesville, OK 74005, USA) at 200 rpm for 1 min at room
153 temperature with 0.1 mm glass beads. After bead beating, lysozyme was added (final
154 concentration of 1 mg/ml), and the samples were incubated at 37°C for 1 h. Following
155 the lysozyme treatment, 120 µL sodium-dodecyl sulphate (20% SDS, W/V) was
156 added and the samples were incubated at 65°C for 60 min. Total genomic DNA was
157 then extracted from the suspension solution using an AxyPrep™ Genomic DNA
158 miniprep kit (Axygen Biosciences, Tewksbury, MA 01876, USA) according to the
159 manufacturer's instructions and stored at -80°C for subsequent study.

160

161 **2.3 Pyrosequencing of partial 16S rRNA genes and sequence analysis**

162 Broadly conserved primers 27F (5'-AGA GTT TGA TCC TGG CTC AG-3') and
163 533R (5'-TTA CCG CGG CTG CTG GCA C-3') were used to amplify the bacterial
164 16S rRNA gene, while primers 344F (5'-ACG GGG YGC AGC AGG CGC GA-3')

165 and 915R (5'-GTG CTC CCC CGC CAA TTC CT-3') were used to amplify the
166 archaeal 16S rRNA gene. PCR reactions were performed following the protocol
167 described in the Supporting Information (SI). Replicate PCR products of the same
168 sample were mixed in a PCR tube. Then, the amplicons from each reaction mixture
169 were pooled in equimolar ratios based on concentration and subjected to emulsion
170 PCR to generate amplicon libraries. Amplicon pyrosequencing was performed on a
171 Roche Genome Sequencer GS FLX+ platform at Majorbio Bio-Pharm Technology,
172 Shanghai, China.

173

174 Sequences generated from pyrosequencing were analyzed using default settings in the
175 open source software package mothur (Schloss et al., 2009). The detailed process is
176 described in the SI. Alpha diversity analyses, including rarefaction and computation
177 of the Shannon, Simpson, Chao1 metric, and phylogenetic diversity (PD), were used
178 to assess biodiversity. The similarity among microbial communities was determined
179 using UniFrac analysis in which weighted and unweighted Principal Coordinate
180 Analysis (PCoA) were performed based on OTUs abundance or phylogenetic
181 relationships. Specific differences in community composition of samples were
182 visualized using heatmaps, ggplot, and Venn diagrams using the R software package.

183

184 **2.4 Miseq-sequencing of partial 16S rRNA genes and sequence analysis**

185 The bacterial and archaeal 16S rRNA gene V4 region (300–350 bp) were amplified
186 using primer set 515f (GTG CCA GCM GCC GCG GTAA) and 806r (GGA CTA
187 CHV GGG TWT CTA AT) with the protocol described by Caporaso et al. (Caporaso
188 et al., 2011; Caporaso et al., 2012). A composite sample for sequencing was created by
189 combining equimolar ratios of amplicons from the individual samples, followed by

190 gel purification and ethanol precipitation to remove any remaining contaminants and
191 PCR artifacts. Amplicon sequencing was conducted on an Illumina MiSeq platform at
192 Novogene Co., Beijing, China.

193

194 Pairs of reads from the original DNA fragments were merged using FLASH (Magoc
195 and Salzberg, 2011). Sequences were then analyzed using the Quantitative Insights
196 Into Microbial Ecology (QIIME) software package (Caporaso et al., 2010), and the
197 UPARSE pipeline (Edgar, 2013). The detailed process is described in the SI. The
198 similarity among microbial communities was determined using UniFrac analysis in
199 which weighted PCoA was performed based on OTUs composition and phylogenetic
200 relationships. Specific differences in community composition of samples were
201 visualized using heatmaps, ggplot, and Venn diagrams using the R package.

202

203 **2.5 Quantification of community abundance**

204 Evaluation of community abundance by real-time fluorescent qPCR was performed
205 using the 16S rRNA gene as a molecular marker. Reactions were performed using the
206 FastStart Universal SYBR Green Master PCR mix in a Bio-Rad iQ5 Sequence
207 detection system. The primer set 8F (5'-AGA GTT TGA T(CT)(AC) TGG
208 CTC-3')/338R (5'-GCT GCC TCC CGT AGG AGT-3') were used to quantify
209 bacterial community, while 806F (5'-ATT AGA TAC CCS BGT AGT CC-33')/958R
210 (5'-YCC GGC GTT GAM TCC AAT T-3') were used to quantify archaeal community
211 (Gittel et al., 2009). Ten-fold serial dilutions of a known copy number of plasmid
212 DNA containing the target gene were subjected to real-time PCR in triplicate to
213 generate an external standard curve. The PCR efficiency and correlation coefficients
214 for the standard curves were higher than 95%, and R^2 values were greater than 0.99

215 for the curves. The specificity of the PCR amplification was determined by the
216 melting curve. Gene copy numbers in unknown samples were determined based on
217 standard curves.

218

219 **2.6 Sequence accession numbers**

220 The raw reads were deposited in the National Center for Biotechnology Information
221 (BioProject ID: PRJNA246768, <http://www.ncbi.nlm.nih.gov/bioproject/246768>).

222

223 **3 Results**

224 **3.1 Microbial community composition in the sandstone reservoir**

225 Up to 4016–5060 bacterial and 2688–2857 archaeal sequences were obtained by
226 pyrosequencing in the sandstone reservoir. These sequences were assigned into
227 249–538 bacterial and 45–130 archaeal OTUs at a 3% cutoff. The individual
228 rarefaction, Shannon, and Phylogenetic diversity curves tended to approach the
229 saturation plateau (Fig. S1). The results of qPCR indicated that the copy number of
230 bacterial 16S rRNA in the injected water was 8.25×10^6 copies ml^{-1} , while 1.5×10^6 to
231 2.75×10^6 copies ml^{-1} in the produced water samples. Comparing with the bacteria, the
232 number of archaea was about one percent of the bacterial number, with 3.75×10^4 16S
233 rRNA copies ml^{-1} in injected water and 8.5×10^3 to 5.75×10^4 copies ml^{-1} in the
234 produced water samples.

235

236 Phylogenetic analysis indicated that the injected water (Lu3084) was dominated by
237 *Proteobacteria* (50.43%), *Cyanobacteria* (15.51%), and *Chloroflexi* (9.12%). Among
238 the *Proteobacteria*, *Betaproteobacteria* (20.42%) and *Alphaproteobacteria* (19.63%)
239 were numerically dominant, while a small quantity of *Deltaproteobacteria* (5.49%),

240 *Gamaproteobacteria* (4.44%), and *Epsilonproteobacteria* (0.32%) were detected (Fig.
241 2a I). The produced water from Lu3073 was dominated by *Proteobacteria* (65.35%)
242 *Spirochaetes* (13.38%), and *Bacteroidetes* (12.38%). *Gammaproteobacteria* (23.96%),
243 *Deltaproteobacteria* (22.16%), *Alphaproteobacteria* (13.47%), and *Spirochaetes*
244 (13.38%) dominated at class level (Fig. 2a I). In the produced water from Lu3095,
245 Lu1039, and Lu2180, *Proteobacteria* composed 78.58%–95.75% of the bacterial
246 communities. *Alphaproteobacteria* (15.43%, 26.77%, 53.54%), *Betaproteobacteria*
247 (23.48%, 50.57%, 12.94%), and *Epsilonproteobacteria* (2.79%, 4.38%, 25.54%) were
248 dominant (Fig. 2a I).

249

250 To archaeal communities, more than 95% of the archaeal sequences were assigned to
251 *Methanobacteria*, *Methanococci*, and *Methanomicrobia* (Fig. 2a II). In the injected
252 water, 87% sequences were classed into *Methanomicrobia*, and the dominant genera
253 were *Methanosaeta* (42.39%), *Methanomethylovorans* (25.57%), and *Methanolobus*
254 (10.96%). Among them, *Methanosaeta* uses only acetate to produce CH₄.
255 *Methanomicrobia* accounted for 84.03% in the produced water of Lu1039, and
256 *Methanolobus* (83.46%) and *Methanococcus* (11.23%) were the dominant genera. The
257 archaeal communities were much more conserved in the produced water at Lu2180,
258 Lu3073, and Lu3095, with *Methanococcus* accounting for 95.34%, 90.79%, and
259 86.79%, respectively. The *Methanolobus* and *Methanococcus* are methylotrophic and
260 hydrogenotrophic methanogens.

261

262 **3.2 Microbial community composition in the conglomerate reservoir**

263 Between 52719 to 129106 16S rRNA gene sequences were analyzed and assigned to
264 2623 to 3414 genus-level OTUs. In combination with the relative abundance, the

265 number of bacterial and archaeal sequences was calculated, with the number of
266 sequences per sample ranging in size from 51273 to 128980 and 85 to 1445,
267 respectively (Fig. S2). Based on the results of qPCR, the copy number of bacterial
268 bacterial 16S rRNA in the water samples ranged from 1.5×10^7 to 6.5×10^7 copies ml⁻¹,
269 while archaeal 16S rRNA ranged from 4.5×10^5 to 8.5×10^5 copies ml⁻¹.

270

271 In contrast to the sandstone reservoir, *Proteobacteria*, *Bacteroidetes*, *Firmicutes*,
272 *Spirochaetes*, and *Synergistetes* were simultaneously detected in both the injected and
273 produced water, composing 85.7%–94.1% of all bacterial communities. Similar to the
274 sandstone reservoir, more *Proteobacteria* were detected in the produced water
275 samples. At the class level, *Gammaproteobacteria*, *Epsilonproteobacteria*,
276 *Alphaproteobacteria*, *Betaproteobacteria*, *Deltaproteobacteria*, *Bacteroidia*, *Bacilli*,
277 and *Clostridia* composed 74.5%–83.7% of the bacterial communities in both the
278 injected and produced water samples (Fig. 2b I).

279

280 The archaea were mainly assigned to *Methanomicrobia*, *Methanococci*,
281 *Methanobacteria*, *Thaumarchaeota*, *Parvarchaea*, and *Thermoplasmata* (Fig. 2b II).
282 Among them, *Methanobacteria*, *Methanococci*, and *Methanomicrobia* were
283 simultaneously detected in both the injected and produced water, and composed
284 64.3%–94.6% of the archaeal communities. Compared with the injected water
285 collected from the wellhead of the injection wells (T86-0 and T93-0), more
286 *Methanomicrobia* were detected in the downhole of injection wells (T86-8 and T93-7)
287 and production well T90. At genus level, *Methanocorpusculum*, *Methanococcus*, and
288 *Methanocalculus* were dominant, accounting for 60.3–88.5% of the archaeal
289 communities in the injection wells and production well T90. The three taxa can use

290 H₂ and formate as carbon sources to produce CH₄.

291

292 **3.3 Shared microbial populations between injected water and reservoir strata**

293 The shared microbial OTUs and genera between communities in the injected and
294 produced water samples were investigated using Venn diagrams, histograms, and
295 heatmap. Based on the Venn diagrams, 16.3%–32.81% of bacterial OTUs and
296 13.73%–51.61% of archaeal OTUs were shared between the injected water and each
297 produced water sample in the sandstone reservoir (Fig. 3a). These shared bacterial
298 OTUs accounted for 4.6–24.71% of the total bacterial community in the injected
299 water, and 43.23–76.18% in each produced water sample (Fig. 4a I). Furthermore,
300 only 13 bacterial and 3 archaeal OTUs were shared by both the injected and produced
301 water samples (Fig. 3a). Accordingly, the shared genera only accounted for 2.26% of
302 the bacterial community in the injected water, but dominated each production well
303 (12.02%–36.5%; Fig. 4a II). Similar to the bacteria, three archaeal genera belonging
304 to *Methanobacterium*, *Methanococcus*, and *Methanolobus* were detected in the
305 injected and produced water samples, comprising 13.58% of the total archaea in the
306 injected water, and 90.4%–96.89% in each of the produced waters (Fig. 4a III).

307

308 In the conglomerate reservoir, most of the OTUs and genera were simultaneously
309 detected in the injected and produced water samples (Fig. 3b and 4b). Similar with the
310 sandstone reservoir, these shared populations accounted for a minor proportion of the
311 communities in the water samples collected from the wellhead of injection wells, but
312 dominated the water samples obtained from the downhole of injection wells and each
313 production well (Fig.3b).

314

315 **3.4 Microbial population distribution in injected water and reservoir strata**

316 Microbial populations were clustered according to injection and production wells to
317 highlight the populations that showed the most variability (Fig. 5). In the sandstone
318 reservoir, more *Sphingomonas* and *Azospirillum* were detected in the injected water,
319 while *Arcobacter*, *Marinobacterium*, *Pseudomonas*, *Hyphomonas*, *Novispirillum*,
320 *Proteiniphilum*, *Spirochaeta*, and *Rhizobium* were highly abundant in the produced
321 water. In the conglomerate reservoir, higher amounts of *Paracoccus*, *Bacillus*,
322 *Ochrobactrum*, *Parabacteroides*, *Sphaerochaeta*, *Thauera*, *Halomonas*, and
323 *Alcanivorax* were detected in the injected water, while *Arcobacter*, *Marinobacterium*,
324 *Pseudomonas*, *Bacteroides*, *Oleibacter*, *Marinobacter*, and *Shewanella* were
325 dominant in the downhole of the injection and production wells. Among them,
326 *Marinobacterium*, *Paracoccus*, *Ochrobactrum*, *Sphingomonas*, *Alcanivorax*, and
327 *Azospirillum* are aerobic bacteria, while *Pseudomonas*, *Rhizobium*, *Arcobacter*,
328 *Halomonas*, *Spirochaeta*, *Bacillus*, *Thauera*, *Halomonas*, and *Bacteroides* are
329 microaerophilic bacteria, facultative anaerobes, or anaerobes.

330

331 To further investigate the microbial distribution in injected water and reservoir strata,
332 hierarchical clustering and Unifrac PCoA were performed based on microbial OTUs
333 abundance and phylogenetic relationships. In the sandstone reservoir, hierarchical
334 clustering showed that the community in the injected water was distinct from that of
335 the produced water (Fig. S3). Weighted PCoA distinguished the bacterial community
336 of the injected water from that of the production wells, while communities of the
337 production wells were placed at a comparatively decentralized position (Fig. 6a I).
338 Similar to the bacterial communities, hierarchical clustering and PCoA distinguished
339 the archaeal community of the injected water from those of the production wells,

340 whereas production wells were placed at a close proximity (Fig. 6a II). In the
341 conglomerate reservoir, communities of water samples collected from the wellhead of
342 injection wells clustered into a group in the PCoA plot, indicating that communities
343 remained unchanged before injected water flowed into the injection wells (Fig. 6b).
344 Communities in the water samples collected from the downhole of injection wells and
345 neighboring production well T90 clustered into one group, while production well T95
346 and T96 clustered into another (Fig. 6b). This shows that the microbial community
347 reassembled during the process of the injected water flowing into the reservoir strata
348 and each production well.

349

350 **4 Discussion**

351 MEOR technique is generally classified into exogenous MEOR and indigenous
352 MEOR (Youssef et al., 2009). The former includes injection of exogenous
353 microorganisms and injection of ex-situ produced products into reservoirs to enhance
354 oil recovery (Zobell, 1947). This is an effective way to quickly improve oil recovery.
355 However, because of the sieve effect of strata on microbial cells, the injected
356 microorganisms are generally difficult to migrate into reservoir strata (Youssef et al.,
357 2009;Brown, 2010). Diverse microbial populations inhabit petroleum reservoirs.
358 Among them, hydrocarbon-degrading bacteria (HDB), nitrate-reducing bacteria
359 (NRB), sulfate-reducing bacteria (SRB), and methanogens are the most important
360 functional populations in reservoir ecosystems (Youssef et al., 2009). When injecting
361 nutrients into reservoir, these microbial populations can be stimulated and produce
362 biosurfactants, organic acids, and gas, which improve oil recovery (Belyaev SS et al.,

363 1998). Comparing with exogenous MEOR, indigenous microorganisms are more
364 adapt to the environmental conditions present in reservoirs. Additionally, because
365 nutrients can be easier to migrate into reservoir strata, indigenous MEOR has higher
366 oil displacement efficiency. Despite the validity in field trial, indigenous MEOR also
367 has some limitations, in particular, the uneven oil displacement efficiency in different
368 production wells in the same reservoir block. The community composition and
369 diversification have been found to have a significant influence on oil displacement
370 efficiency (Li et al., 2014). Therefore, it is needed to investigate the relationship
371 between microbial communities in injected water and reservoir strata, because
372 microbial communities in injected water are expected to flow into oil-bearing strata,
373 and produce a significant influence on those of reservoir strata in long-term
374 water-flooding petroleum reservoirs (Youssef et al., 2009;Dahle et al., 2008).

375

376 Molecular methods have been widely used to assess the microbial diversity of
377 petroleum reservoirs. Compared to the traditional 16S rRNA gene clone library and
378 sequencing, high-throughput sequencing has generated hundreds of thousands of short
379 sequences, and significantly improved our ability to compare microbial populations
380 with unprecedented levels of coverage and detail (Caporaso et al., 2012). In the
381 conglomerate reservoir, Miseq-sequencing produced approximately 52719 to 129106
382 16S rRNA gene sequences. The sequencing reads was approximately 10–20 folds of
383 those (obtained by pyrosequencing) of the sandstone reservoir, and 50–400 folds of
384 the 16S rRNA gene clone library (assuming 300 clones per library). We
385 simultaneously sequenced the bacterial and archaeal V4 region of 16S rRNA gene,
386 obtaining a total of 51273–128980 bacterial sequences per sample, but only 85–1445

387 archaeal sequences. This is consistent with the count result for archaea, which are
388 about one percent of bacterial number. The result implies that deeper sequencing is
389 needed for detecting rare archaeal populations using Miseq-sequencing based on
390 bacterial and archaeal universal primer set 515f and 806r. In contrast, the bacterial and
391 archaeal communities were sequenced independently using pyrosequencing in the
392 sandstone and we obtained 4016–5060 bacterial and 2688–2857 archaeal sequences.
393 The rarefaction curves, Shannon diversity estimates, and observed species, suggesting
394 that this sequencing depth was enough for the investigation of the bacterial and
395 archaeal communities.

396

397 If the microbial populations in the injected water could flow into the reservoir strata
398 and reach the production wells along with the injected water, is the microbial
399 community in the injected water expected to have a similar community composition
400 with those in the production wells? In the homogeneous sandstone reservoir, we found
401 the number of shared bacterial and archaeal populations between the injected water
402 and each production well was different. As shown in Fig. 3a, 16.3%–32.81% of
403 bacterial OTUs and 13.73%–51.61% of archaeal OTUs were shared between the
404 injected water and each produced water sample. It is reasonable to speculate that
405 microbial populations in the injected water produce different levels of impact on those
406 in production wells. Based on the previous research, the main reason may be the sieve
407 effect that can be enhanced by the long inter-well spacing (Ren et al., 2011). Because
408 of this effect on microbial cells when injected fluid passes through a subsurface
409 formation, it is more difficult for microbial cells to migrate in reservoir strata. In
410 contrast, almost all OTUs and genera detected in the injected water were also
411 observed in downhole of the injection and neighboring production wells in the

412 heterogeneous reservoir. Compared with the sandstone reservoir, this reservoir has a
413 similar permeability, but shorter inter-well spacing. It appears that most microbial
414 populations in the injected water flowed into the oil-bearing strata and reached the
415 production wells during water-flooding process. Additionally, despite lacking for
416 sufficient nutrients, bacterial and archaeal communities in the reservoir strata have
417 high concentrations, which are similar with those in the injected water. This
418 phenomenon implies some correlations of microbial communities in injected water
419 and reservoir strata. However, we appreciate that it is less rigorous to delineate the
420 transport of microbial populations in the reservoir strata simply by detecting the
421 shared microbial populations in the injection and production wells using 16S rRNA
422 sequencing, because this method is not able to demonstrate whether the species
423 detected in the produced water are the same ones as in the injected water. Therefore,
424 labelled strains, such as ones containing green fluorescent protein, may be a suitable
425 way to investigate microbial migration in petroleum reservoirs.

426

427 The number of shared microbial populations reflects the influence of microbial
428 communities in injected water on those in reservoir strata to some extent, and show
429 strong association with the unique variation of reservoir environments. Differ with the
430 sandstone reservoir, a large number of microbial populations were shared by the
431 injected water and produced water samples in the conglomerate reservoir. However,
432 the community structure, in particular, the abundance of the shared OTUs and genera,
433 exhibited a large difference between the injected water and reservoir strata. The
434 environmental variables, such as salinity, pH, and nutrients, have been supposed to be
435 the primary drivers for the community diversification (Kuang et al., 2013). However,
436 few differences in cations and anions among the injected and produced water samples

437 were observed. Petroleum reservoir represents an anaerobic environment with
438 multiphase fluids of oil, gas and water. Therefore, except for the sieve effect of
439 reservoir strata on microbial migration, dissolved oxygen, which is known to be
440 strongly related to reservoir microbial growth and metabolism (Gao et al., 2013), may
441 be the main factor influencing the community structures. In both the reservoirs,
442 aerobic bacteria, including *Sphingomonas*, *Azospirillum*, *Paracoccus*, *Ochrobactrum*,
443 *Alcanivorax*, and *Hydrogenophilaceae* were more frequently detected in the injected
444 water, while microaerophilic bacteria, facultative anaerobes, and anaerobes, including
445 *Pseudomonas*, *Rhizobium*, *Arcobacter*, *Halomonas*, *Spirochaeta*, and *Bacteroides*,
446 were found to have higher relative abundance in reservoir strata (Fig. 5).

447

448 Apart from the dissolved oxygen, crude oil, in particular, the saturated and aromatic
449 hydrocarbon, may also strongly influence the microbial distribution in injected water
450 and reservoir strata. Petroleum reservoirs represent oligotrophic environments.
451 Although diverse microbial populations inhabit the reservoirs, only
452 hydrocarbon-degrading bacteria and some anaerobes, such as sulfate-reducing
453 bacteria, could grow with crude oil as carbon source. This is consistent with the
454 observed results that more hydrocarbon-degrading bacteria, including
455 *Marinobacterium*, *Pseudomonas*, *Rhizobium*, *Halomonas*, and *Oleibacter*, were
456 detected downhole of injection and production wells.

457

458 This study compared the differences in microbial community composition between
459 injected water and reservoir strata using microbial genomes obtained from the
460 aqueous phase. In fact, each component of the reservoir multiphase fluid, including
461 crude oil, gases, and insoluble particles, may act as an important habitat for microbial

462 growth in addition to the water phase within the petroleum reservoir (Kryachko et al.,
463 2012; Kobayashi et al., 2012). Recent research has also compared microbial
464 communities in aqueous and oil phases of water-flooded petroleum reservoirs, and
465 found that the oil phase also harbored a large number of microorganisms, with large
466 differences in the bacterial community between the aqueous and oil phases of the
467 reservoir fluid (Wang et al., 2014). Therefore, simultaneous analysis of DNA
468 extracted from both aqueous and oil phases may provide a better understanding of the
469 microbial communities in injection and production water samples.

470

471 In summary, this study investigated the relationship shared by microbial communities
472 in injected water and reservoir strata in two long-term water-flooding reservoirs.
473 However, the results cannot provide any reliable information on the indigenous
474 microbial community. The indigenous microbial populations may be those in newly
475 drilled wells without water-flooding in the same oil-bearing block. However, the two
476 reservoirs have been water-flooded for decades. Due to the introduction of exogenous
477 microorganisms in injected water and other sources of contaminations by enhanced
478 oil recovery processes, determining whether a microorganism is indigenous to a
479 petroleum reservoir become increasingly difficult. This study implies that the number
480 of shared populations reflects the influence of microbial communities in injected
481 water on those in reservoir strata to some extent, and show strong association with the
482 unique variation of reservoir environments. However, it cannot make a conclusion on
483 the transport of microbial populations in the reservoir strata by detecting the shared
484 microbial populations in injected water and produced water samples using 16S rRNA
485 sequencing. To further investigate the relationship shared by microbial communities
486 in injection and production water samples, injecting labelled strains containing

487 marked gene (e.g., green fluorescent protein coded gene) into reservoirs may bring
488 novel insight and greater predictive power. Therefore, further research on microbial
489 diversification and transferability as injected water flows into reservoir is needed.
490 Solving these problems is significant to guide the application of MEOR approaches
491 based on injecting nutrients or microbial populations into reservoirs.

492

493 **5 Conclusions**

494 Using high-throughput sequencing, this study revealed the similarities and differences
495 in microbial communities in the injected water and reservoir strata in a homogeneous
496 sandstone reservoir and a heterogeneous conglomerate reservoir. The bacterial and
497 archaeal communities in the reservoir strata have high concentrations, which are
498 similar with those in the injected water. However, microbial community compositions
499 exhibited large differences between the injected produced water samples. The number
500 of shared populations reflects the influence of microbial communities in injected
501 water on those in reservoir strata to some extent, and show strong association with the
502 unique variation of reservoir environments. Additionally, aerobic bacterial populations
503 were more frequently detected in injected water, while microaerophilic bacteria,
504 facultative anaerobes, and anaerobes dominated the reservoir strata.

505

506 **Acknowledgments**

507 This study was supported by the National High Technology Research and
508 Development Program of China (Grant No. 2013AA064402), the National Natural
509 Science Foundation of China (Grant No. 41373074), the Chinese Postdoctoral Science
510 Foundation (Grant No. 2014M561175), and the Key Technology Research and
511 Development Program of Tianjin (Grant No. 14ZCZDSY00016).

512

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599

600 **Figure captions**

601 **Fig. 1.** Schematic diagram showing (a) the distribution of the injection and production
602 wells of the Liu and Lu field blocks, and (b) the wellhead and downhole of injection
603 and production wells, and the location where the water samples were collected. The
604 injected water samples were collected from the wellhead and the zone close to
605 downhole (obtained by backflow, that is, the injected water flowed upward through
606 the injection well) of the injection wells. The water samples in reservoir strata were
607 collected from the wellhead of production wells. T86, T93 and Lu3084 are the
608 selected injection wells, and T90, T95, T96, Lu1039, Lu2180, Lu3073 and Lu3095
609 are the production wells.

610 **Fig. 2.** Relative proportion of microbial taxa at class level in the injected and
611 produced water samples. (a) Sandstone and (b) conglomerate reservoirs. I: bacterial
612 taxa at class level; II: archaeal taxa at class level; Lu3084, T86-0, and T93-0: water
613 samples from the well head of the injection wells; T86-8 and T93-7: water samples
614 from downhole of the injection wells; T90, T95, T96, Lu1039, Lu2180, Lu3073, and
615 Lu3095: water samples from the well head of the production wells.

616 **Fig. 3.** Venn diagrams of the bacterial and archaeal OTUs in the injection and
617 production wells. (a) Sandstone and (b) conglomerate reservoirs. I: bacterial OTUs; II:
618 archaeal OTUs. Venn diagrams indicate the shared microbial OTUs between
619 communities in the injected and produced water samples.

620 **Fig. 4.** Comparison of shared microbial genera between the injection and production
621 wells. a-I: pairwise comparison between injection and production wells in the
622 sandstone reservoir; a-II: shared bacterial genera in the injection and production wells;
623 a-III: shared archaeal genera in the injection and production wells; b-I and

624 b-II: comparison between injection and production wells on the conglomerate
625 reservoir; and b-III: dominant shared bacterial genera in the conglomerate reservoir.

626 **Fig. 5.** Genera showing the most variability in the injected water and production wells.
627 (a) Sandstone and (b) conglomerate reservoirs. The black-bordered box indicates the
628 genera most detected in the production wells.

629 **Fig. 6.** Principal coordinate analysis of microbial communities used to investigate the
630 microbial distribution in injected water and reservoir strata. (a) Sandstone and (b)
631 conglomerate reservoirs. I: bacterial community distribution; II: archaeal community
632 distribution. Sample points that are close together are more similar in community
633 composition than those that are far apart. The arrows in panel (b) indicate the
634 community succession during the process of the injected water flowing into the
635 injection wells and the neighbouring production wells.

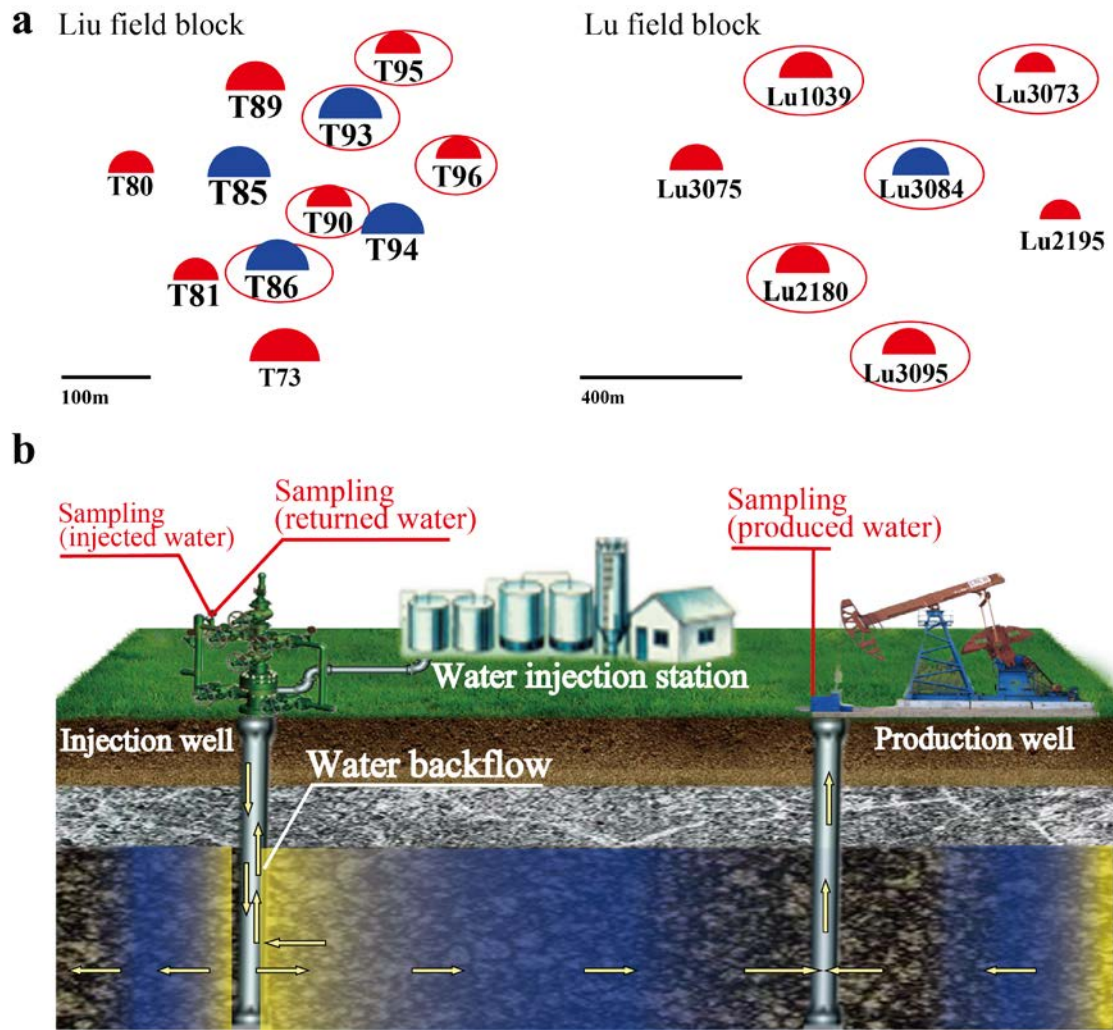
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Table 1 Reservoir characteristics of Lu and Liu field block

Reservoir characteristics	Lu field block	Liu field block
Oil Reservoir		
Formation lithology	Sandstone	Conglomerate
Average depth (m)	1200	1088
Pressure (MPa)	10.2	7.2
Stratal temperature (°C)	37	22.6
Average water content, %	80.8%	86.8%
Interwell distances, m	300-425	100-150
Average permeability, μm^2	522×10^{-3}	362×10^{-3}
Effective porosity, %	29.9	18.96
Water flooding (yr)	13	30
Crude oil properties		
Density (g/cm^3)	0.846	0.912
Viscosity in situ ($\text{mPa}\cdot\text{s}$)	18	80.0
Saturates (%)	71.29	61.94
Aromatic (%)	14.85	11.24
Resin (%)	5.94	18.85
Asphalte (%)	5.94	7.97

639 **Table 2** Chemical properties of the water samples obtained from Lu and Liu field
 640 block (Unit: mg/L)

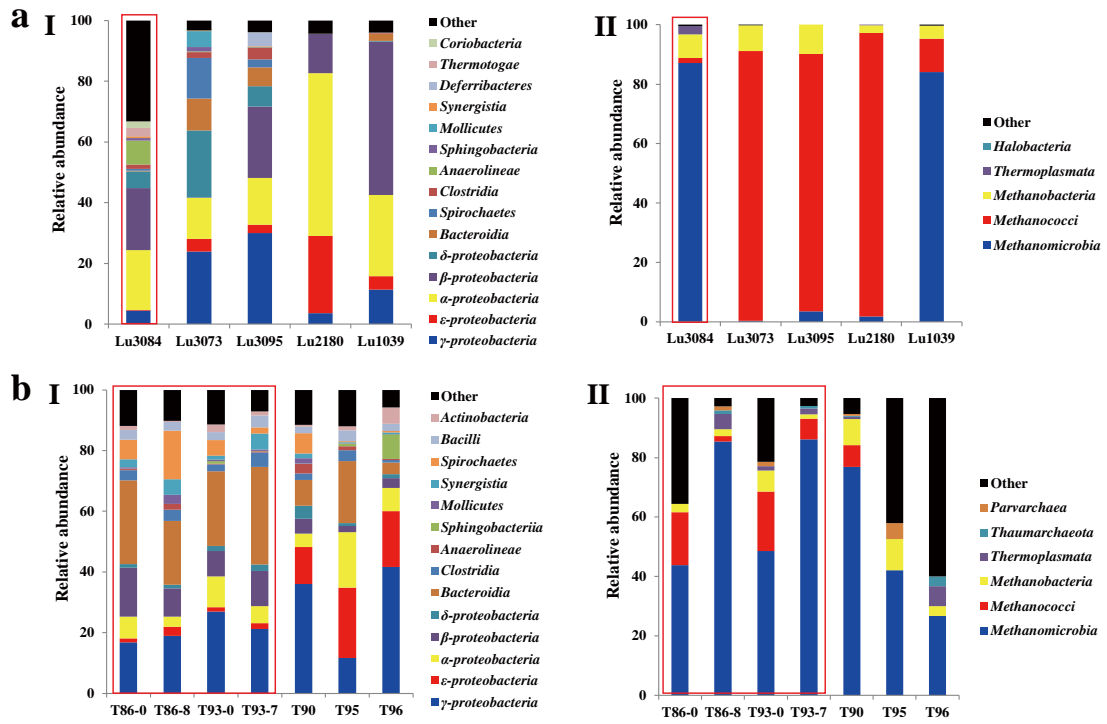
Samples	Lu field block					Liu field block						
	3084	1039	2180	3073	3095	T86-0	T86-8	T93-0	T93-7	T90	T95	T96
Salinity	10850	11690	11170	10545	11102	10101	11313	11399	13991	13203	8997	9710
Nitrogen	15.1	11.5	10.6	12.7	11.6	7.6	6.8	10.2	11.5	5.7	6.3	8.5
Phosphorus	20.2	19.1	17.5	18.8	19.5	16.8	15.1	22.6	21.2	18.5	16.2	12.1
Na ⁺ K ⁺	4525	4803	4565	4309	4487	3364	3630	3802	4349	4014	3097	3139
Mg ²⁺	21.7	32.1	31.6	26.0	28.8	33.1	63.1	28.4	63.1	68.1	17.5	50.2
Ca ²⁺	191.3	281.9	284.7	181.6	216.4	70.2	77.9	72.8	78.0	96.2	86.4	108.7
Cl ⁻	5640	6125	5820	5160	5850	3010	3630	2922	3453	3099	3816	3406
SO ₄ ²⁻	116.2	14.04	4.86	23.13	8.86	483.4	89.4	523	362.4	89.9	65.6	165.1
HCO ₃ ⁻	356	434	464	846	511	3140	3823	4052	5687	5837	1915	2841



641

642 **Fig. 1**

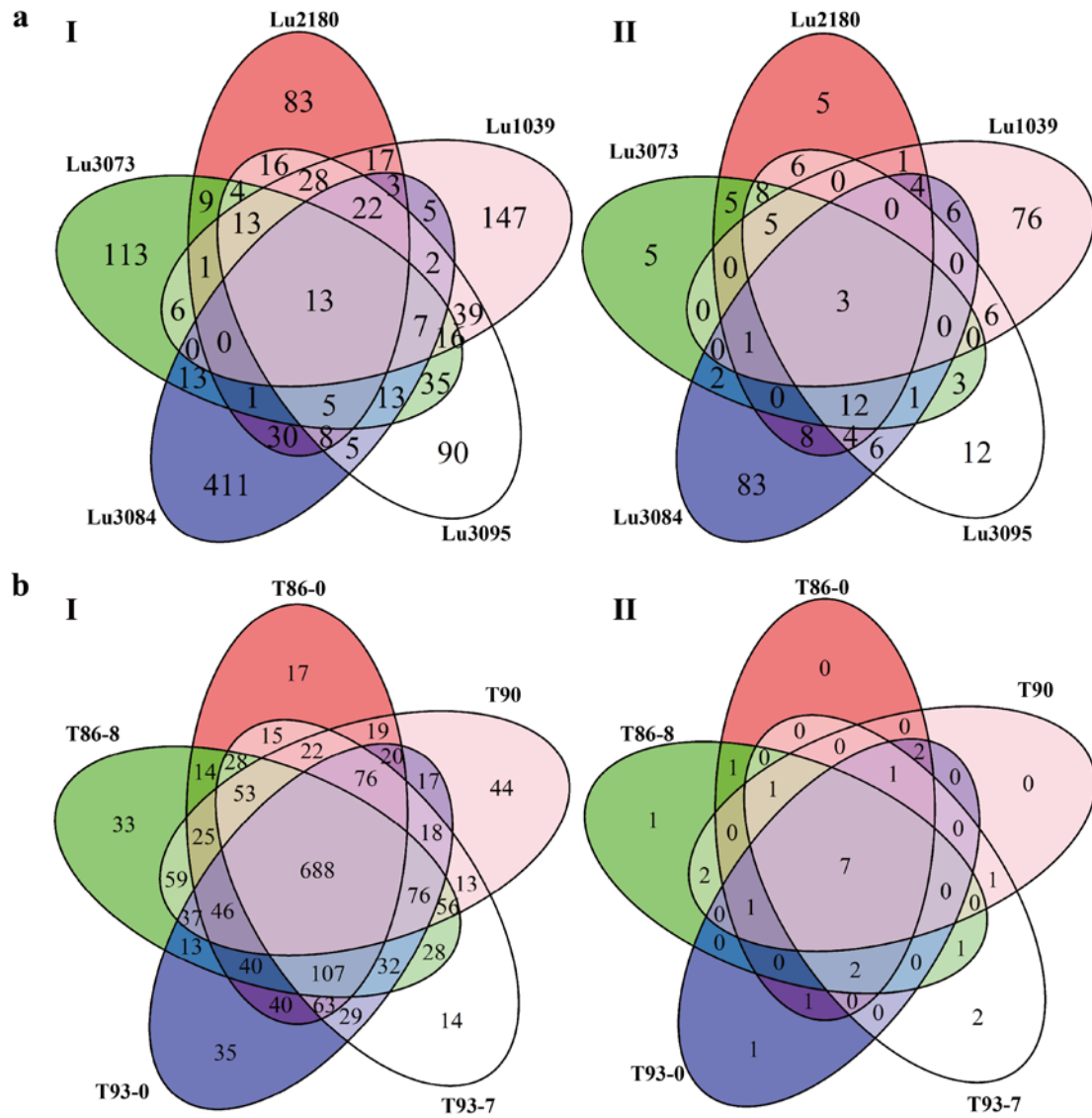
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645 **Fig. 2**

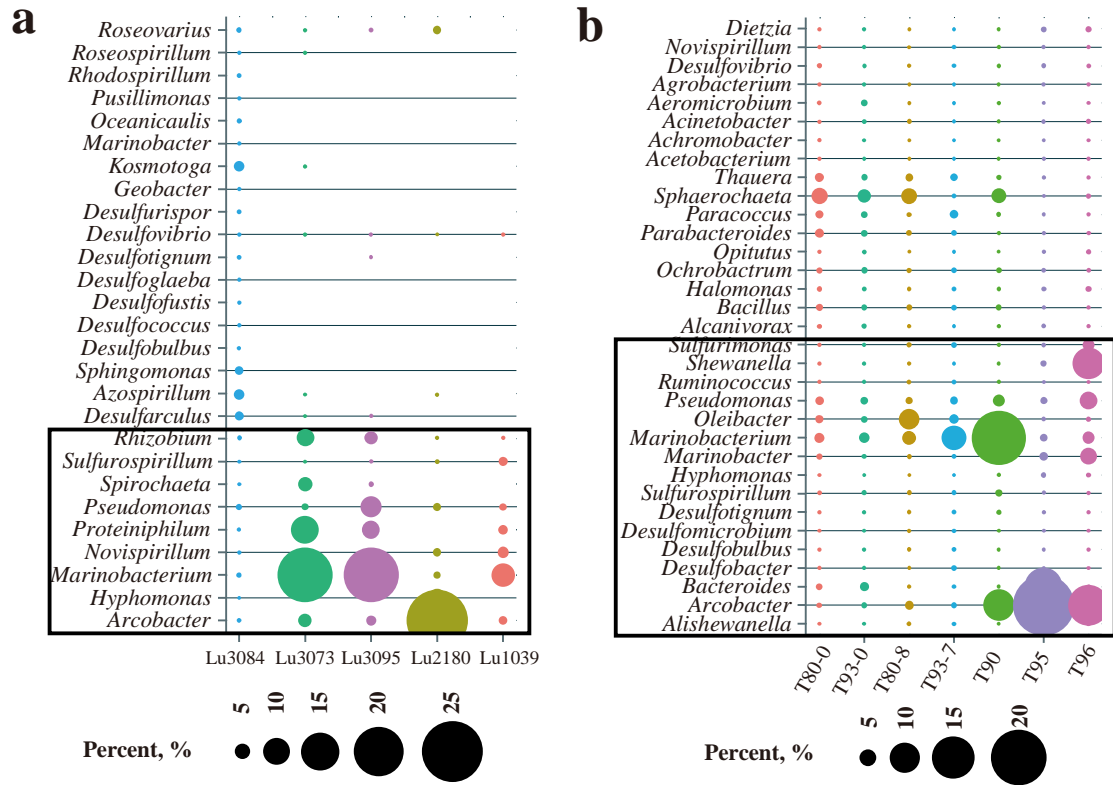
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648 **Fig. 3**

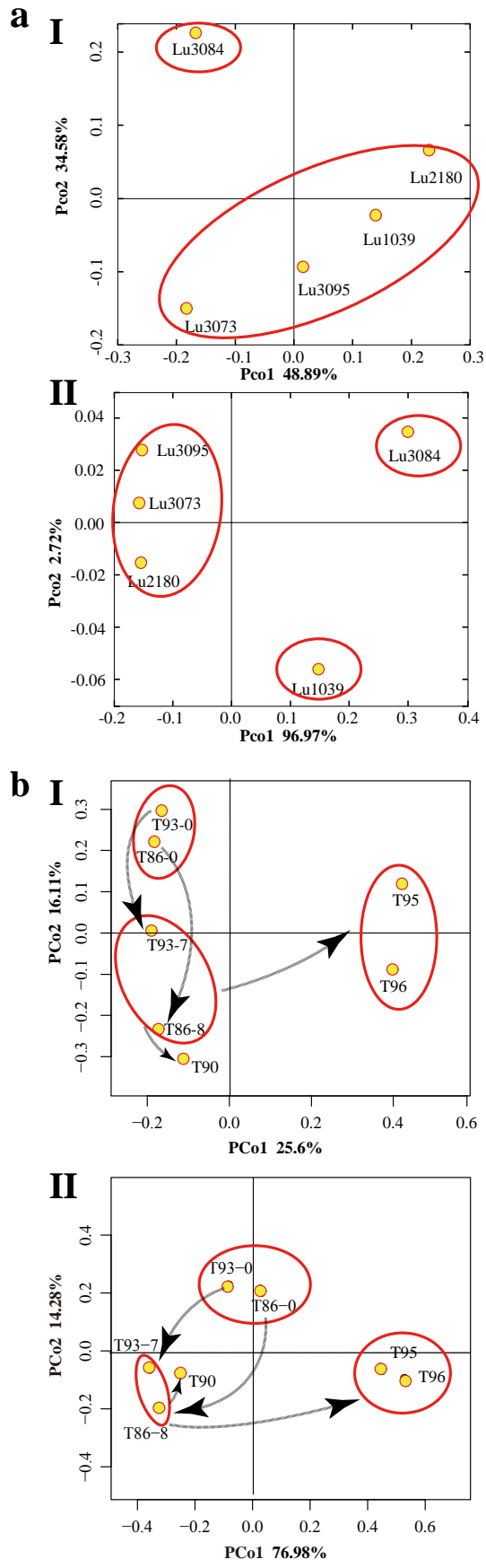
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654 **Fig. 5**

655



656

657 **Fig. 6**