Disentangling residence time and temperature sensitivity

2 of microbial decomposition in a global soil carbon model

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Abstract

13 Recent studies have identified the first-order representation of microbial decomposition as a 14 major source of uncertainty in simulations and projections of the terrestrial carbon balance. 15 Here, we use a reduced complexity model representative of current state-of-the-art models of 16 soil organic carbon decomposition. We undertake a systematic sensitivity analysis to 17 disentangle the effect of the time-invariant baseline residence time (k) and the sensitivity of 18 microbial decomposition to temperature (Q_{10}) on soil carbon dynamics at regional and global 19 scales. Our simulations produce a range in total soil carbon at equilibrium of ~592 to 2745 Pg 20 C which is similar to the ~561 to 2938 Pg C range in pre-industrial soil carbon in models 21 used in the fifth phase of the Coupled Model Intercomparison Project. This range depends 22 primarily on the value of k, although the impact of Q_{10} is not trivial at regional scales. As climate changes through the historical period, and into the future, k is primarily responsible 23 24 for the magnitude of the response in soil carbon, whereas Q_{10} determines whether the soil 25 remains a sink, or becomes a source in the future mostly by its effect on mid-latitudes carbon 26 balance. If we restrict our simulations to those simulating total soil carbon stocks consistent 27 with observations of current stocks, the projected range in total soil carbon change is reduced 28 by 42% for the historical simulations and 45% for the future projections. However, while this 29 observation-based selection dismisses outliers it does not increase confidence in the future

30 sign of the soil carbon feedback. We conclude that despite this result, future estimates of soil

carbon, and how soil carbon responds to climate change should be constrained by available

32 observational data sets.

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1 Introduction

35 There is a 6-fold range in the amount of carbon stored in the soil in simulations conducted as

part of the fifth phase of the Coupled Model Intercomparison Project (CMIP5; Taylor et al.,

37 2012). This 6-fold range, identified by Todd-Brown et al. (2013), is consistent with results

38 from the recent model intercomparison projects such as the Coupled Climate-Carbon Cycle

39 Model Intercomparison Project (C⁴MIP; Friedlingstein et al., 2006). The analysis of carbon

40 stores in both C⁴MIP and CMIP5 have focused on the prediction of terrestrial and soil carbon

41 through time. In addition to demonstrating the large differences in carbon stocks (Todd-

42 Brown et al., 2013), they have also highlighted large inter-model differences in global and

regional land-atmosphere carbon (C) fluxes (e.g. Friedlingstein et al., 2006, 2014). This lack

of agreement between simulations exists in fully coupled models (e.g. C⁴MIP and CMIP-5)

45 but can also be found if sources of uncertainty are narrowed by relying on one weather

dataset to drive multiple land models (Friend et al., 2013; Nishina et al., 2014), or by using

47 one land model driven by multiple climate projections (Ahlström et al., 2013).

48 In these previous studies, critical uncertainties have been identified in the microbial

49 decomposition of soil organic C and the associated release of CO₂ via heterotrophic

50 respiration (R_h). This is despite all the current state-of-the-art global soil C models relying on

a similar representation of decomposition as a first-order process (see Exbrayat et al., 2013b;

52 Nishina et al., 2014; Todd-Brown et al., 2013). This conceptualization describes

53 decomposition and R_h as proportional to the availability of organic matter. The decay rate (or

54 R_h per unit of soil C) is modified based on an environmental scalar that intends to mimic the

55 dynamical response of microbial biomass to soil moisture and soil temperature.

56 This simple model structure has recently received some criticism because of its lack of

57 explicit representation of microbial physiology (Allison et al., 2010; Todd-Brown et al.,

58 2012; Wieder et al., 2013; Xenakis and Williams, 2014). However, it can successfully explain

59 some complex dynamic processes including the acclimation of decomposers to warming (Luo

et al., 2001) as a result of the quick depletion of labile pools by enhanced microbial biomass

61 (Kirschbaum, 2004; Knorr et al., 2005).

We previously identified (Exbrayat et al., 2013b, 2014) some further implications of the firstorder representation of microbial decomposition. First, in climate change experiments, model pools are usually initialised using a spin-up procedure with fixed pre-industrial atmospheric CO₂ concentrations until C pool trends are removed (Xia et al., 2012). Due to the interaction with substrate availability, the decay rate simulated by the model in response to steady boundary conditions determines the size of soil C pools reached at equilibrium. Because spinup is a long computational process, the magnitude of pool sizes is conserved during subsequent shorter simulations of climate change and, as a result, equilibrated stocks strongly explain final stocks (e.g. CMIP5 models as shown in supplementary Figure S1 after Exbrayat et al., 2014). Second, the microbial sensitivity to changing environmental conditions affects the response of the system under transient climate simulations (Falloon et al., 2011; Exbrayat et al., 2013a,b). However, because substrate availability also controls the amount of respired carbon, there is a "memory" control imposed by the initial conditions of this transient simulation (Exbrayat et al., 2013b and 2014) that also affects the response to perturbation in boundary conditions. The relative contribution of these two factors on soil C projections remains to be explored in detail especially since last generation models disagree on the carbon balance projected in the future (Friedlingstein et al., 2014; Nishina et al., 2014),

Here, we use a reduced complexity model representative of current state-of-the-art models of soil organic C decomposition. A systematic sensitivity analysis is performed to disentangle the effect of the time-invariant baseline residence time and the formulation of the dynamic response of microbial decomposition to climatic change on soil C dynamics at regional and global scale.

making it challenging to elaborate any land-based offsetting strategy.

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2 Materials and methods

2.1 Reduced complexity model

88 It is not possible to re-run each CMIP5 model or isolate the representation of soil carbon

89 processes from each model. We therefore use a reduced complexity model that simulates the

- 90 monthly evolution of a single soil organic carbon pool, C_s , in response to input derived from
- 91 Net Primary Productivity (NPP, g C m⁻² mth⁻¹) and output by heterotrophic respiration (R_h , g
- 92 C m⁻² mth⁻¹). For each monthly time step, the soil carbon balance can be described as:

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$$\frac{\partial C_s}{\partial t} = NPP - R_h \tag{1}$$

- 94 where NPP is a prescribed boundary condition in our model and R_h is simulated as a first-
- order process dependent on the availability of substrate C_s such as:

$$96 R_h = k^{-1} \times f_T \times f_W \times C_s (2)$$

- 97 where k is the baseline residence time at 15°C (Xia et al., 2013) adjusted at each time step by
- 98 f_T which is a function of soil temperature T_s (°C). The soil moisture (θ_s) modification
- 99 function, f_W , is usually expressed as a fraction of soil moisture saturation (Moyano et al.,
- 100 2012). We implement a classical formulation of the soil temperature sensitivity function f_T :

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$$f_T = Q_{10}^{\frac{(T_s - T_{ref})}{10}}$$
 (3)

- where Q_{10} is a constant factor that describes the relative increase in microbial activity for a
- warming of 10°C, and T_{ref} is the reference temperature (°C) for which $f_T(T_s) = 1$ (Lloyd and
- Taylor, 1994; Bauer et al., 2012). The chosen T_{ref} is the commonly used 15°C (Todd-Brown
- et al., 2013) so that the decomposition rate equals k^{-1} when moisture is non-limiting and
- temperature is approximately equal to the global average. We use the same formulation of f_W
- as in the CASA-CNP model (Wang et al., 2010):

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$$f_W(\theta_s) = \left(\frac{\theta_s - 1.70}{0.55 - 1.70}\right)^{6.6481} \times \left(\frac{\theta_s + 0.007}{0.55 + 0.007}\right)^{3.22}$$
 (4)

- which is a bell-shaped function that is equal to 1 for $\theta_s = 0.55$.
- 110 This first-order representation of microbial decomposition with a specified decay rate
- adjusted by environmental scalars is used in all 11 CMIP5 models that simulate soil carbon
- 112 (Todd-Brown et al., 2013) and all 7 Dynamic Global Vegetation Models used in the ISI-MIP
- project (Friend et al., 2013; Nishina et al., 2014). Typically, these models rely on a multi-pool
- architecture to represent the diversity in organic matter. Each pool has its own residence time
- that corresponds to a degree of resistance to decomposition (Davidson and Janssen, 2006).
- 116 Usually, part of the decomposition occurring in one pool is routed to one or several other

117 pools while the rest is emitted via R_h . At the ecosystem scale, however, the same 118 environmental scalar is applied despite the multi-pool architecture, and the heterotrophic 119 respiration flux is proportional to the amount of substrate available. Therefore, our simplified 120 model is broadly representative of the current paradigm and provides a useful framework to 121 undertake the sensitivity analysis described hereafter. 122 Soil moisture also has an influence on microbial decomposition (Falloon et al., 2011, Moyano 123 et al., 2012, 2013; Exbrayat et al., 2013a,b). However, Todd-Brown et al. (2013) recently 124 demonstrated that a one pool reduced complexity model could reproduce both total soil 125 carbon content and its spatial distribution for most of the CMIP5 models without considering 126 decomposition response to variations in soil moisture. We also recently showed that global features in the distribution and evolution of C_s were much more related to uncertainties in f_T 127 128 than uncertainties in the formulation of f_W (Exbrayat et al., 2013b). Therefore, in order to 129 keep the analyses as simple as possible and isolate the effect of f_T but still account for the 130 effect of soil moisture on R_h , we keep the formulation of f_W constant in the experiments that 131 follow. 132 We are aware of that our reduced complexity model relies on questionable assumptions such 133 as the use of a single soil carbon pool and global values of k, Q_{10} and T_{ref} . However, while we 134 agree that a multiple pool structure would provide diverging results, single pool soil carbon 135 carbon models similar to our design are used in 3 of the 11 CMIP5 models described by 136 Todd-Brown et al. (2013) and 2 of the 7 ISI-MIP models described by Nishina et al. (2014). 137 Further, using global parameter values of k, Q_{10} and T_{ref} is consistent with these state-of-theart models (Todd-Brown et al., 2013; Nishina et al., 2014). Of course, this does not allow 138 139 representing processes such as the remobilization of carbon in the active cycle following permafrost thaw (Koven et al., 2011) or the probably different behaviour of biological 140 141 systems in frozen conditions but these are not routinely implemented in the land component 142 of Earth system models and therefore fall beyond the scope of this paper. In summary, we 143 wish to reiterate that this study investigates the sensitivity of the first-order parameterization 144 of microbial decomposition and R_h processes used in current ecosystem models to its 145 uncertain parameters (Todd-Brown et al., 2013; Nishina et al., 2014). We do not intend to 146 provide improved results of the response of soil carbon to climate change but rather illustrate 147 and better understand the implications of the current ubiquitous approach to parameterization 148 and initial value prescription described in Section 2.2.

2.2 Model setup and experiments

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150 We configure the reduced complexity model in a spatially explicit way to represent global variations, implemented as a surrogate for the CASA-CNP biogeochemical module (Wang et 151 152 al., 2010) of the CABLE land surface model (Wang et al., 2011). A previous simulation by CABLE coupled to the coarse-resolution CSIRO Mk3L climate model (3.2 $^{\circ}$ latitude \times 5.6 $^{\circ}$ 153 154 longitude; Phipps et al., 2011) and driven by CMIP5 atmospheric CO₂ data provides monthly NPP, T_s and θ_s to the reduced complexity model. We use both historical simulations 155 (Exbrayat et al., 2013b) and 21st century projections using the Representative Concentration 156 157 Pathway 8.5 (RCP 8.5) atmospheric concentration scenario. 158 We perform a sensitivity analysis by running the simple model with various combinations of a Q_{10} value and a baseline residence time k. We use 11 equally-spaced values of Q_{10} ranging 159 160 from 1.5 to 2.5 (i.e. intervals of 0.1), and 31 equally-spaced values of k ranging from 120 161 months to 480 months (i.e. intervals of 12 months). These values are based on the range of 162 results previously obtained by Todd-Brown et al. (2013) with their own reduced complexity 163 model. Each value of Q_{10} is applied with each value of k for a total of 341 simulations. Model 164 versions are initialised via a classical spin-up procedure (Xia et al., 2012) using input data 165 from 1850 to 1859 for 10,000 years to ensure all soil carbon pools reach a steady-state. We 166 then continue simulations with NPP, T_s and θ_s data from 1850 to 2005, and continue with 167 RCP 8.5 projections to 2100. We note that these drivers do not include the representation of 168 land-use and land cover change and their effect on NPP, T_s and θ_s . Therefore, SOC input are likely to be higher than in reality. However, as stated earlier we are using the reduced 169 complexity framework to understand the behaviour of the SOC model in response to 170 171 variations in its parameters and we do not aim to provide improved estimates of global scale terrestrial carbon sinks. In each model version, both k and the sensitivity of R_h to temperature 172 173 (represented by Q_{10}) are constant globally, in accordance with observations (Mahecha et al., 174 2010) and state-of-the-art models (Todd-Brown et al., 2013; Nishina et al., 2014). However, 175 the actual value of the environmental scalar f_T will of course vary spatially and temporally as 176 a function of T_s . As we keep the same formulation of f_W between model versions, we can 177 attribute differences in results to the values of Q_{10} or k.

Harmonized World Soil Database

180 The Harmonized World Soil Database (HWSD; FAO, 2012) combines several national inventories and provides a number of chemical and physical soil properties at a 30 arc second 181 182 resolution globally. However, despite the availability of this dataset, CMIP5 models exhibit a six-fold range in their total soil carbon content (Todd-Brown et al., 2013) including values 183 184 well outside the uncertainty boundaries of observational data. We showed previously that 185 using this dataset to discriminate between acceptable and unacceptable simulations resulted 186 in a non-negligible reduction of the uncertainty in historical net carbon uptake (Exbrayat et 187 al., 2013b). While we do not aim to provide CMIP5-like projections of the soil carbon 188 balance with our reduced complexity model, we investigate the value of using the HWSD to 189 discriminate between plausible and implausible simulations. 190 We follow the method described by Todd-Brown et al. (2013) to derive an estimate of current 191 total soil carbon from the latest version of the Harmonized World Soil Database (HWSD). First, we re-grid the original 30 arc seconds raster to a $0.5^{\circ} \times 0.5^{\circ}$ resolution. Within each 192 half-degree cell we select the dominant soil type. For each soil type, the database provides 193 194 bulk density and organic carbon content for a top layer (0 - 30 cm depth) and a bottom layer

(30 – 100 cm depth). This allows us to calculate soil C density (in kg C m⁻²) in each cell. We

then multiply each grid cell by its area and sum to obtain a global estimate of ~1170 Pg C.

Similarly to Todd-Brown et al. (2013) we also consider the uncertainty associated to our re-198

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gridding process as well as analytical measurements of soil properties. We therefore obtain a 95% confidence interval (CI₉₅) of 29% below the mean to 32% above the mean, or ~830 – 1550 Pg C. We provide these gridded data as supplementary material.

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3 Results

Total soil carbon and global balance

204 Figure 1 presents snapshots of total soil carbon for all 341 model versions for three periods: at equilibrium (in 1850, Figure 1a), at the end of historical transient simulations (in 2005, 205 Figure 1b), and at the end of the projections with forcing corresponding to RCP 8.5 (in 2100, 206 207 Figure 1c). Figure 1a shows that the spin-up procedure causes different model versions to 208 equilibrate at widely varying levels of total soil carbon despite the use of the same boundary

209 conditions of NPP and T_s . Differences in residence time k contribute most of the ~592 to 2745 Pg C range, with larger values of k resulting in larger pools (Figure 1a). Variations in 210 211 the Q_{10} parameter of f_T have a smaller influence on total soil carbon but lower values do result in lower total soil carbon. For the same value of k, simulations with $Q_{10} = 1.5$ equilibrate with 212 213 total soil carbon equal to $86\% \pm 0.005\%$ (mean ± 1 standard deviation) of the amount with 214 $Q_{10} = 2.5$. Figure 1b shows that the distribution of total soil carbon between model versions 215 does not vary much during historical simulations (1850-2005). Models with large total soil 216 carbon pools over this period remain versions with long residence time k and higher values of 217 Q_{10} . Note, however, that the range of total soil carbon in 2005 grows to ~709 to 2943 Pg C. 218 Dashed contours on Figure 1b indicate the limits of the CI₉₅ of the HWSD for current total 219 soil carbon. Here, 115 simulations with values of k ranging approximately from 150 to 250 220 months all fall within this range for 2005, regardless of the Q_{10} value used. Finally, Figure 1c 221 continues to indicate a strong control of k on the total soil carbon in 2100. The projected range narrows to ~684 to 2825 Pg C throughout the 21st century. However, we note there is 222 223 an inversion in the influence of Q_{10} on simulated total soil carbon with lower values of Q_{10} 224 resulting in larger pools especially for longer baseline residence times k. Nevertheless, this is 225 still minor compared to the influence of k on C_s . 226 Although the range in simulated soil carbon remains similar through time, non-negligible 227 changes occur. This is highlighted in Figure 2 which shows ΔC_s , the change in total soil 228 carbon as a function of model parameters k and Q_{10} for the historical simulations (1850 – 229 2005, Figure 2a) and RCP 8.5 projections (2006 – 2100, Figure 2b). First, Figure 2a clearly 230 shows that all model versions act as a net carbon sink during historical simulations, 231 accumulating between 81 and 283 Pg C. Model versions with longer residence time k tend to 232 accumulate more carbon through time. However, models with the largest value of Q_{10} tend to 233 accumulate only 69% \pm 0.4% (mean \pm 1 standard deviation) of the amount that the lowest Q_{10} 234 models do. By analysing Figure 2b, we see that the influence of Q_{10} on the total soil carbon balance grows during RCP 8.5 projections where Q_{10} now determines whether the soil 235 236 remains a sink or becomes a source. This change between a source or a sink for different Q_{10} 237 values follows a near linear relationship with k (solid line on Figure 2b). Interestingly, the -238 179 to 168 Pg C range in the change in total soil carbon during RCP 8.5 is mostly a function 239 of Q_{10} as both extremes are achieved with the longest residence time used here. In other

words, while Q_{10} decides of the sign of the change, k, and hence the initial stocks of SOC 240 241 after spin-up, drives the magnitude of the response. 242 If we consider only models that fall within the CI₉₅ of the HWSD for current total soil carbon 243 (dashed contours on Figure 2a and 2b) the spread in simulated total soil carbon balance is 244 largely reduced. During the historical simulations, the range of this subset of models shrinks 245 by 84 Pg C to between 87 and 205 Pg C. It corresponds to a reduction of about 42% of the 246 initial uncertainty. Similarly, the range in projected soil carbon balance is reduced by 157 Pg 247 C to -129 to 61 Pg C, a reduction of about 45% of the initial uncertainty. We note, however, 248 that this restriction does not necessarily increase confidence in sign of the future soil carbon 249 change under RCP8.5. 250 Differences in the behaviour between the full set of models and this subset of observationally 251 constrained models can be seen in the time series and probability density functions (PDFs) 252 for the historical period, shown in Figure 3. First, the time series from 1850 shows there is no 253 noticeable difference between the full set of simulations (in grey) and the subset of 254 simulations with acceptable current soil carbon (in green) until 1900. During the first half of the 20th century, stronger sinks are excluded as they lie outside the CI₉₅ range, which 255 correspond to the upper tail of the distribution of ΔC_s (see PDF inset for 1950). However, the 256 257 kurtosis of the distribution, or most probable change from our simulations, changes 258 negligibly. After ~1960, we observe a step-change in cumulative ΔC_s that follows a strong 259 response in NPP to the rapid increase in atmospheric CO₂ (please refer to Exbrayat et al., 260 2013b for a more detailed account of this behaviour). The spread between simulations grows 261 and most of the excluded simulations based on the CI₉₅ range are the strongest sinks (as in 262 Figure 2a) while a few of the least accumulating simulations are also excluded. This does have a large impact on the most probable change in storage, reducing it from ~200 PgC to 263

We now examine future simulations and present time series and PDFs of change in total soil carbon during RCP 8.5 projections in Figure 4. All simulations continue to accumulate carbon at the beginning of the 21^{st} century and remain net carbon sinks until about 2060. At the end of the century, some model versions have simulated positive ΔC_s corresponding to a net carbon sink over the 21^{st} century, while other ends their projections with negative ΔC_s , or a net carbon loss. However, all simulations show the same overall behaviour with first an increase in C_s that peaks, and then a decrease in C_s . The timing of the peak, i.e. when soil

~140 PgC.

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carbon starts to deplete, varies between ~2035 and 2075 and is explained by the value of Q_{10} ($R^2 = 0.74$, data not shown) with higher values leading to an earlier peak. This indicates that, in all simulations, soil has become a net source of carbon by the end of the 21^{st} century, regardless how much carbon was accumulated since 2005, and hence since 1850. The PDFs in 2050 show that selecting only observationally consistent models results in the most heavily accumulating simulations, i.e. those that would peak later, to be dismissed. However, by 2100, both the lower and upper tails of the initial distribution are clipped, reducing the simulated range from -178 to 168 Pg C (all simulations) to -129 to 61 Pg C. In both cases, differences in the kurtosis of both distributions remains very small which indicates that our selection scheme dismisses outliers. We note that the lower bound of ΔC_s for both sets of models is the same until late in the projections (~2085).

3.2 Regional differences

Although Figure 1 indicates that the range in k can explain most of the variability in total soil carbon content at equilibrium and hence through transient simulations, Q_{10} is likely to influence the local response of f_T . Figure 5 shows the relative value of f_T for different temperatures and values of Q_{10} . Since the chosen T_{ref} =15°C, all Q_{10} values lead f_T to be equal at this particular temperature. However, the more difference there is between the actual temperature and T_{ref} , the more sensitive f_T becomes to values of Q_{10} . As our simulations are spatially-explicit, this may introduce non-negligible regional differences in C pools at equilibrium and their response to transient changes in T_s and NPP.

To investigate this more in detail, we present the zonal averages of soil C density for different values of Q_{10} with k set to 180 months (Figure 6). We choose this particular residence time as example because all corresponding simulations are within the CI_{95} of the HWSD for 2005 regardless the value of Q_{10} . Figure 6a shows that Q_{10} values do introduce non-negligible differences in local equilibrated soil C density. Steady-state pools at low latitudes (30°S to 30°N) are larger with low values of Q_{10} (blue in Figure 6). Conversely, high latitude pools are larger with high values of Q_{10} (red in Figure 6). Overall, the range in the value of zonally averaged soil C density at equilibrium is up to three-fold depending on the chosen value of Q_{10} . This is particularly obvious in regions with high NPP including low-latitude tropical rainforests or northern taigas. As was the case with total C_8 , the zonal distribution soil C

303 density and the relative position of simulations with different Q_{10} do not vary much between 304 1850 and 2005 (Figure 6b) although there is a slight shift towards uniformly higher densities 305 as all model versions are net global carbon sinks (Figure 2a and 3). The pattern of zonal soil 306 carbon remains essentially the same at the end of RCP 8.5 projections. However, models with 307 lower values of Q_{10} now have more carbon than those with high values of Q_{10} over a broader zone $(40^{\circ}S - 50^{\circ}N)$. 308 309 Figure 7 shows the zonal change in soil C density for the same simulations as in Figure 6. 310 Figure 7a indicates that all simulations simulate a net sink almost everywhere during historical simulations, except at latitudes > 70°N. However, the strength of this sink is 311 312 strongly dependent upon the value of Q_{10} , especially in low latitudes. There is an approximately two-fold difference between the high accumulation of low Q_{10} models, and the 313 314 low accumulation of high Q_{10} models. Differences between Q_{10} values are negligible at 315 higher latitudes. Figure 7b shows the same information for RCP 8.5 projections. Simulations 316 with lower values of Q_{10} almost always accumulate more C (except between 0° and 10° N). 317 While all model versions with k = 180 months lose carbon at low latitudes $(20^{\circ}S - 20^{\circ}N)$, 318 and gain carbon at high latitudes in the northern hemisphere (> 50°N), the value of Q_{10} , and 319 hence the environmental scalar f_T , decides of the sign of the local soil C balance in the 21st century at mid-latitudes. Within the mid-latitudes, high values of Q_{10} are more likely to 320 321 simulate a net loss of soil carbon. We can therefore narrow down the dependence of the 322 global ΔC_s on Q_{10} to its affect at mid-latitudes.

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4 Discussion

4.1 Effect of k and Q_{10} on soil carbon

326 In our simulations, the range in total soil carbon at equilibrium (~592 to 2745 Pg C) depends 327 on which value of Q_{10} and especially k is used (Figure 1a). This range captures the ~561 to 328 2938 Pg C range in soil carbon in CMIP-5 in 1860 (see Supplementary Figure S1). We note 329 of course that CMIP5 models not only vary in their soil C component, but simulate different 330 NPP and T_s and also integrate a range of soil moisture limitations (Todd-Brown et al., 2013). The range achieved here at the end of the historical simulations (~709 to 2943 Pg C) is, for 331 example, larger than the 1090 to 2646 Pg C range in 2000 from 7 DGVMs in the ISI-MIP 332 333 project (Nishina et al., 2014) which were driven by a harmonised weather dataset.

We can attribute this range to the first-order representation of decomposition and its response to the initialisation procedure used in most CMIP-5 simulations. By spinning-up the model, the goal is to stabilise pools so that total NPP is exactly compensated by total R_h over the selected period of time (here 10 years). In Equation (2), a longer residence time k results in a lower decay rate (i.e. R_h per unit of C_s). Therefore, model versions that have a slower turnover will require more substrate to simulate the same R_h needed to compensate NPP. As the baseline residence time k is applied globally, it drives the global pool size (Figure 1) much more than changing Q_{10} affects f_T . However, as seen in Figure 6, when considered regionally, Q_{10} plays a non-negligible role for the local response of decomposition and the definition of equilibrium soil C density. High values of Q_{10} lead f_T to trigger strong decay rates in warm regions (Figure 5) that require less substrate (see low latitudes in Figure 6a) to compensate the same NPP. Conversely, high Q_{10} lead to low values of f_T in cold regions. Therefore, more substrate is required to bring the pool to equilibrium as seen in high latitudes in Figure 6a. Low values of Q_{10} show an opposite regional behaviour. Regional differences compensate each other and therefore f_T with different Q_{10} values can only explain a small fraction of the range in equilibrated total soil carbon. Of course, if another T_{ref} was used, the relative differences between f_T with different Q_{10} would be altered and the influence of Q_{10} and its effect on f_T on total and local C_s would vary. Furthermore, the difference between f_T with different Q_{10} grows with the absolute value of the difference T_s - T_{ref} . Therefore, using a value of T_{ref} that is outside the range of actual temperatures would lead f_T with different Q_{10} to keep the same relative position globally. It would introduce larger relative differences between these functions. Comparing Figures 1a, 1b and 1c suggests that the range in total C_s at equilibrium is a good predictor of the current and future range in total soil carbon. Despite differences in the magnitude of the change in C_s through time (Friedlingstein et al., 2014), equilibrium conditions achieved under pre-industrial conditions largely define current and future pool sizes. Examining Figure 6 confirms that this global effect can also be seen regionally, especially in low (20°S to 20°N) and high (>50°N) latitudes, where carbon pools are largest. This is of concern as substrate availability also influences R_h and hence its response to changes.

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Changes in C_s through time are nevertheless non-negligible, and it is important to quantify the response of the system to perturbations. Our results show increasing atmospheric CO_2

concentrations enhances NPP more than the simultaneous warming enhances R_h during historical simulations. This historical net carbon sink that is driven by the response of vegetation to increasing atmospheric CO₂ (and hence SOC_{in}) is in accordance with previous studies (Friedlingstein et al., 2006; Sarmiento et al., 2010; Zhang et al., 2011; Wania et al., 2012; Anav et al., 2013; Exbrayat et al., 2013b). Therefore, all model versions with longer residence time accumulate more C_s over the same time period as a result of a slower turnover of carbon in soils, and this mirrors the state of the equilibrium stores. However, despite the dominance of the increased NPP on ΔC_s , the historical warming signal is influential. Specifically, those model versions more sensitive to changes in temperature (i.e. with high values of Q_{10}) accumulate less soil carbon during the 20^{th} century even though they initially equilibrated with larger global pools. This is also true of local soil C density where high Q_{10} values are less accumulating regardless of the initial soil C density. We however note that the value T_{ref} used in our experiments is well within the range of actual temperatures. Therefore, the historical warming does not induce large changes in the values of f_T with different Q_{10} . Projections under the strong-forcing RCP 8.5 scenario also see an increase in the influence of the value of Q_{10} on ΔC_s . Figure 2b clearly shows that the capacity of soils to become carbon sources or remain sinks depends almost entirely on the Q_{10} parameter, and both states can be achieved for any value of k used while remaining within range of previous studies (Friedlingstein et al., 2014; Nishina et al., 2014). Figure 7b indicates that this is clearly a result of differences in the local response of model versions in the mid-latitudes as a function of Q_{10} . Such regional discrepancies leading to a change in the sign of global ΔC_s models have also been highlighted through a recent inter-comparison project that used a harmonised weather dataset to drive 7 biome models (Nishina et al., 2014). However, contrary to this previous study, none of our model versions accumulates soil carbon in the inter-tropical region during the 21st century. This is probably due to the fact that we use the same boundary conditions of NPP and T_s for all our model versions, while models used by Nishina et al. (2013) used a prescribed weather dataset but were left free to simulate their own NPP. Overall, the globally applied model parameter k drives the steady-state response of our

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Overall, the globally applied model parameter k drives the steady-state response of our reduced complexity system. However, the more conditions are changing (i.e. steady-state to historical to RCP 8.5 projections), the more the dynamic transition of the system towards a new equilibrium depends on the environmental scalar f_T and the specific value of Q_{10} . Although the same formulation of f_T is applied globally, differences in its response to local T_s

sum up to determine the sign of total soil carbon balance. We also note that model versions that equilibrate as a result of longer baseline residence time k have a tendency to produce a larger absolute response of total soil carbon balance. Therefore, the size of pools to which the change is applied seems to dominate the response even when higher values of k imply a smaller relative change in the decay rate $k^{-1} \times f_T \times f_W$ used in equation 2. This control of initial conditions obtained by spin-up on the response of the system is a critical aspect that needs to be better resolved, especially since recent inter-comparison experiments all exhibit huge discrepancies in equilibrium conditions of participating models (Anav et al., 2013; Todd-Brown et al., 2013; Nishina et al., 2014).

4.2 Discriminating between model versions

- Since k clearly influences the total soil carbon content at equilibrium in 1850, it is a good predictor of the current total soil carbon content. Therefore, k is the key parameter that decides how much carbon is active in the modelled system, and whether model versions fall within the CI_{95} of the HWSD. Here, all simulations with baseline residence time between 150 and 250 months fulfil this requirement regardless of which Q_{10} is used in f_T .
 - If we isolate these simulations, the range in total soil carbon change shrinks by 42% and 45% for the historical simulations and RCP 8.5 projections, respectively. However, while this selection dismisses outliers it does not increase confidence in the sign of the soil carbon change. This is because regional differences lead to similar values in total soil carbon for different values of Q_{10} . These regional differences translate into heterogeneous responses under RCP 8.5 forcing, especially in mid-latitudes. They are sufficient to induce a change of sign in the global soil carbon balance.

5 Conclusion

We have used a reduced complexity model, broadly representative of current state-of-the-art models of soil organic C decomposition used in CMIP5 and ISI-MIP experiments, to explore the response of microbial decomposition to climate change on soil C dynamics at regional and global scale. We have shown that key parameters in the first-order representation of decomposition interact in markedly different ways depending on the nature of forcing and

antecedent conditions. First, the time and space-invariant baseline residence time decides of the total soil carbon content at equilibrium after spin-up, typically the process used by CMIP5 models to initialise C pools. Next, the more boundary conditions imposed on the system move away from the equilibrium forcing, the more the environmental scalar describing the sensitivity of the system gains in importance. However, it is the size of the pool to which the change is applied that mostly controls the magnitude of the response.

Applying a constraint on total soil carbon that discriminates between acceptable simulations - of total soil carbon leads to a drastic reduction of the range of simulated change. Meanwhile, most of the remaining uncertainty in 21st century projections of total soil carbon can be attributed to zonal differences in the response to change, especially at mid-latitudes. These do not allow us to confidently project soil as either a global source or sink of carbon for the 21st century. However, it is clear that under RCP 8.5 tropical soils are not suited for long-term carbon storage while some more potential exists in high latitudes.

Finally, we suggest that future estimates of terrestrial, and especially soil, carbon responses to climate change should be more constrained by available datasets of carbon stocks. This is crticial as model structures describe fluxes as a fraction of the substrate pool size. So far, the process of spin-up has too many degrees of freedom that lead to model-specific amounts of active soil carbon.

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Acknowledgements

This work was supported by the Australian Research Council ARC grant DP110102618 and the ARC Centre of Excellence for Climate System Science grant CE110001028. CSIRO Mk3L model runs were made possible by the NCI National Facility at the Australian National University via the provision of computing resources to the ARC Centre of Excellence for Climate System Science. We thank Dr. K. E. O. Todd-Brown for guidance in processing the

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References

HWSD dataset.

456 Ahlström, A., Smith, B., Lindström, J., Rummukainen, M., and Uvo, C. B.: GCM 457 characteristics explain the majority of uncertainty in projected 21st century terrestrial

- 458 ecosystem carbon balance, Biogeosciences, 10, 1517-1528, doi:10.5194/bg-10-1517-2013,
- 459 2013.
- 460 Allison, S. D., Wallenstein, M. D. and Bradford, M. A.: Soil-carbon response to warming
- dependent on microbial physiology, Nat. Geosci., 3(5), 336–340, doi:10.1038/ngeo846, 2010.
- 462 Anav, A., Friedlingstein, P., Kidston, M., Bopp, L., Ciais, P., Cox, P., Jones, C., Jung, M.,
- 463 Myneni, R., and Zhu, Z.: Evaluating the land and ocean components of the global carbon
- 464 cycle in the CMIP5 Earth systems models, J. Clim., 26, 6801–6843, doi:10.1175/JCLI-D-12-
- 465 00417.1, 2013.
- Bauer, J., Weihermüller, L., Huisman, J., Herbst, M., Graf, A., Séquaris, J. and Vereecken,
- 467 H.: Inverse determination of heterotrophic soil respiration response to temperature and water
- 468 content under field conditions, Biogeochemistry, 108(1), 119–134, doi:10.1007/s10533-011-
- 469 9583-1, 2012.
- 470 Davidson, E. A. and Janssens, I. A.: Temperature sensitivity of soil carbon decomposition
- 471 and feedbacks to climate change, Nature, 440(7081), 165–173, doi:10.1038/nature04514,
- 472 2006.
- Exbrayat, J.-F., Pitman, A. J., Abramowitz, G. and Wang, Y.-P.: Sensitivity of net ecosystem
- 474 exchange and heterotrophic respiration to parameterization uncertainty, J. Geophys. Res.
- 475 Atmospheres, 118(4), 1640–1651, doi:10.1029/2012JD018122, 2013a.
- Exbrayat, J.-F., Pitman, A. J., Zhang, Q., Abramowitz, G., and Wang, Y.-P.: Examining soil
- 477 carbon uncertainty in a global model: response of microbial decomposition to temperature,
- 478 moisture and nutrient limitation, Biogeosciences, 10, 7095-7108, doi:10.5194/bg-10-7095-
- 479 2013, 2013b.
- Exbrayat, J.-F., Pitman, A. J., and Abramowitz, G.: Response of microbial decomposition to
- spin-up explains CMIP5 soil carbon range until 2100, Geosci. Model Dev. Discuss., 7, 3481-
- 482 3504, doi:10.5194/gmdd-7-3481-2014, 2014.
- 483 Falloon, P. D., Jones, C. D., Ades, M. and Paul, K.: Direct soil moisture controls of future
- 484 global soil carbon changes: An important source of uncertainty, Glob. Biogeochem. Cycles,
- 485 25, GB3010, doi: 10.1029/2010GB003938, 2011.
- 486 FAO/IIASA/ISRIC/ISSCAS/JRC: Harmonized World Soil Database (version 1.21), FAO,
- 487 Rome, Italy and IIASA, Laxenburg, Austria, 2012.

- 488 Friedlingstein, P., Cox, P., Betts, R., Bopp, L., von Bloh, W., Brovkin, V., Cadule, P., Doney,
- 489 S., Eby, M., Fung, I., Bala, G., John, J., Jones, C., Joos, F., Kato, T., Kawamiya, M., Knorr,
- 490 W., Lindsay, K., Matthews, H. D., Raddatz, T., Rayner, P., Reick, C., Roeckner, E.,
- 491 Schnitzler, K.-G., Schnur, R., Strassmann, K., Weaver, A. J., Yoshikawa, C. and Zeng, N.:
- 492 Climate–Carbon Cycle Feedback Analysis: Results from the C⁴MIP Model Intercomparison,
- 493 J. Clim., 19, 3337–3353, doi:10.1175/JCLI3800.1, 2006.
- 494 Friedlingstein, P., Meinhausen, M., Arora, V. K., Jones, C. D., Anav, A., Liddicoat, S. K.,
- and Knutti, R.: Uncertainties in CMIP5 climate projections due to carbon cycle feedbacks, J.
- 496 Clim., 27, 511-526, doi:10.1175/JCLI-D-12-00579.1, 2014.
- 497 Friend, A. D., et al: Carbon residence time dominates uncertainty in terrestrial vegetation
- responses to future climate and atmospheric CO₂, Proc. Natl. Acad. Sci., in press, 2013.
- 499 Kirschbaum, M. U. F.: Soil respiration under prolonged soil warming: are rate reductions
- 500 caused by acclimation or substrate loss?, Glob. Change Biol., 10(11), 1870–1877,
- 501 doi:10.1111/j.1365-2486.2004.00852.x, 2004.
- 502 Knorr, W., Prentice, I. C., House, J. I. and Holland, E. A.: Long-term sensitivity of soil
- 503 carbon turnover to warming, Nature, 433(7023), 298–301, doi:10.1038/nature03226, 2005.
- 504 Koven, C. D., Ringeval, B., Friedlingstein, P., Ciais, P., Cadule, P., Khvorostyanov, D.,
- 505 Krinner, G. and Tarnocai, C.: Permafrost carbon-climate feedbacks accelerate global
- 506 warming, Proc. Natl. Acad. Sci., 108(36), 14769–14774, doi:10.1073/pnas.1103910108,
- 507 2011.
- 508 Lloyd, J. and Taylor, J. A.: On the Temperature Dependence of Soil Respiration, Funct.
- 509 Ecol., 8(3), 315–323, doi:10.2307/2389824, 1994.
- 510 Luo, Y., Wan, S., Hui, D. and Wallace, L. L.: Acclimatization of soil respiration to warming
- in a tall grass prairie, Nature, 413(6856), 622–625, doi:10.1038/35098065, 2001.
- Mahecha, M. D., Reichstein, M., Carvalhais, N., Lasslop, G., Lange, H., Seneviratne, S. I.,
- Vargas, R., Ammann, C., Arain, M. A., Cescatti, A., Janssens, I. A., Migliavacca, M.,
- Montagnani, L. and Richardson, A. D.: Global Convergence in the Temperature Sensitivity
- 515 of Respiration at Ecosystem Level, Science, 329(5993), 838-840,
- 516 doi:10.1126/science.1189587, 2010.

- Nishina, K., Ito, A., Beerling, D. J., Cadule, P., Ciais, P., Clark, D. B., Falloon, P., Friend, A.
- 518 D., Kahana, R., Kato, E., Keribin, R., Lucht, W., Lomas, M., Rademacher, T. T., Pavlick, R.,
- 519 Schaphoff, S., Vuichard, N., Warszawaski, L., and Yokohata, T.: Quantifying uncertainties in
- 520 soil carbon responses to changes in global mean temperature and precipitation, Earth Syst.
- 521 Dynam., 5, 197-209, doi:10.5194/esd-5-197-2014, 2014.
- 522 Phipps, S. J., L. D. Rotstayn, H. B. Gordon, J. L. Roberts, A. C. Hirst and W. F. Budd: The
- 523 CSIRO Mk3L climate system model version 1.0 Part 1: Description and evaluation,
- 524 Geoscientific Model Development, 4, 483-509, doi:10.5194/gmd-4-483-2011, 2011.
- 525 Sarmiento, J. L., Gloor, M., Gruber, N., Beaulieu, C., Jacobson, A. R., Mikaloff Fletcher, S.
- 526 E., Pacala, S. and Rodgers, K.: Trends and regional distributions of land and ocean carbon
- 527 sinks, Biogeosciences, 7(8), 2351–2367, doi:10.5194/bg-7-2351-2010, 2010.
- 528 Tarnocai, C., Canadell, J. G., Schuur, E. a. G., Kuhry, P., Mazhitova, G. and Zimov, S.: Soil
- 529 organic carbon pools in the northern circumpolar permafrost region, Glob. Biogeochem.
- 530 Cycles, 23(2), GB2023, doi:10.1029/2008GB003327, 2009.
- Taylor, K. E., Stouffer, R. J. and Meehl, G. A.: An Overview of CMIP5 and the Experiment
- 532 Design, Bull. Am. Meteorol. Soc., 93(4), 485–498, doi:10.1175/BAMS-D-11-00094.1, 2012.
- Todd-Brown, K. E. O., Randerson, J. T., Post, W. M., Hoffman, F. M., Tarnocai, C., Schuur,
- E. A. G., and Allison, S. D.: Causes of variation in soil carbon simulations from CMIP5 Earth
- 535 system models and comparison with observations, Biogeosciences, 10, 1717-1736,
- 536 doi:10.5194/bg-10-1717-2013, 2013.
- Todd-Brown, K. E. O., Hopkins, F. M., Kivlin, S. N., Talbot, J. M. and Allison, S. D.: A
- 538 framework for representing microbial decomposition in coupled climate models,
- 539 Biogeochemistry, 109(1), 19–33, doi:10.1007/s10533-011-9635-6, 2012.
- Wang, Y. P., Kowalczyk, E., Leuning, R., Abramowitz, G., Raupach, M. R., Pak, B., Gorsel,
- 541 E. van and Luhar, A.: Diagnosing errors in a land surface model (CABLE) in the time and
- 542 frequency domains, J. Geophys. Res., 116, G01034, doi:10.1029/2010JG001385, 2011.
- Wang, Y. P., Law, R. M. and Pak, B.: A global model of carbon, nitrogen and phosphorus
- 544 cycles for the terrestrial biosphere, Biogeosciences, 7(7), 2261–2282, doi:10.5194/bg-7-2261-
- 545 2010, 2010.

- Wania, R., Meissner, K. J., Eby, M., Arora, V. K., Ross, I., and Weaver, A. J.: Carbon-
- 547 nitrogen feedbacks in the UVic ESCM, Geosci. Model Dev., 5, 1137-1160, doi:10.5194/gmd-
- 548 5-1137-2012, 2012.
- Wieder, W. R., Bonan, G. B and Allison, S. D., Global soil carbon projections are improved
- 550 by modelling microbial processes, Nature Clim. Change 3, 909–912,
- 551 doi:10.1038/nclimate1951, 2013.
- 552 Xenakis, G. and Williams, M., Comparing microbial and chemical approaches for modelling
- 553 soil organic carbon decomposition using the DecoChem v1.0 and DecoBio v1.0 models,
- 554 Geosci. Model. Dev. Discuss., 7, 33-72, doi:10.5194/gmdd-7-33-2014, 2014.
- Xia, J. Y., Luo, Y. Q., Wang, Y.-P., Weng, E. S., and Hararuk, O.: A semi-analytical solution
- 556 to accelerate spin-up of a coupled carbon and nitrogen land model to steady state, Geosci.
- 557 Model Dev., 5, 1259-1271, doi:10.5194/gmd-5-1259-2012, 2012.
- 558 Xia, J., Luo, Y., Wang, Y.-P. and Hararuk, O.: Traceable components of terrestrial carbon
- storage capacity in biogeochemical models. Glob. Change Biol., 19, 2104-2116. doi:
- 560 10.1111/gcb.12172, 2013.

- 561 Zhang, Q., Wang, Y. P., Pitman, A. J. and Dai, Y. J.: Limitations of nitrogen and
- 562 phosphorous on the terrestrial carbon uptake in the 20th century, Geophys. Res. Lett., 38,
- 563 L22701, doi:10.1029/2011GL049244, 2011.

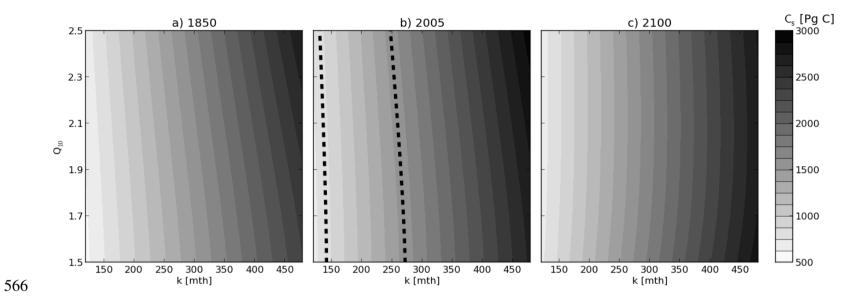


Figure 1. Snapshots of total soil carbon in the reduced complexity model as a function of parameter values. Dashed contours in panel b indicate the CI_{95} of the HWSD in 2005 (830 – 1550 Pg C).

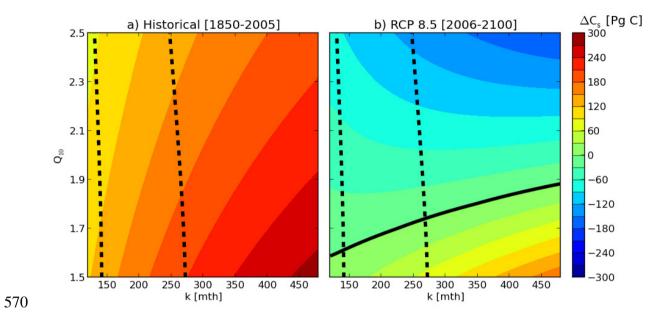


Figure 2. Change in total soil carbon in the reduced complexity model as a function of parameter values for each period as indicated. Dashed contours in panel b indicate model versions that produced soil stocks within the CI_{95} of the HWSD in 2005 (830 – 1550 Pg C). The thick black line represents no change.

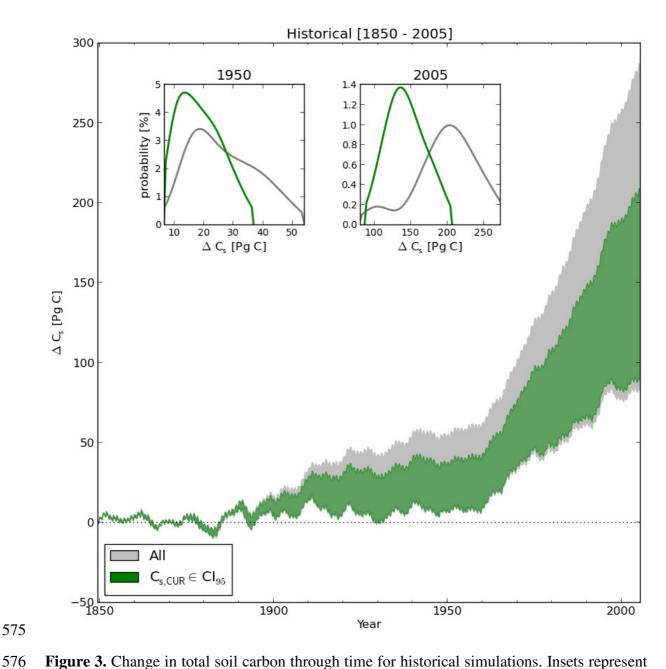


Figure 3. Change in total soil carbon through time for historical simulations. Insets represent the probability density function of the change since 1850 for the period indicated. Grey is for all simulations while green is used to distinguish simulations for which total soil carbon is within the CI_{95} of the HWSD in 2005.

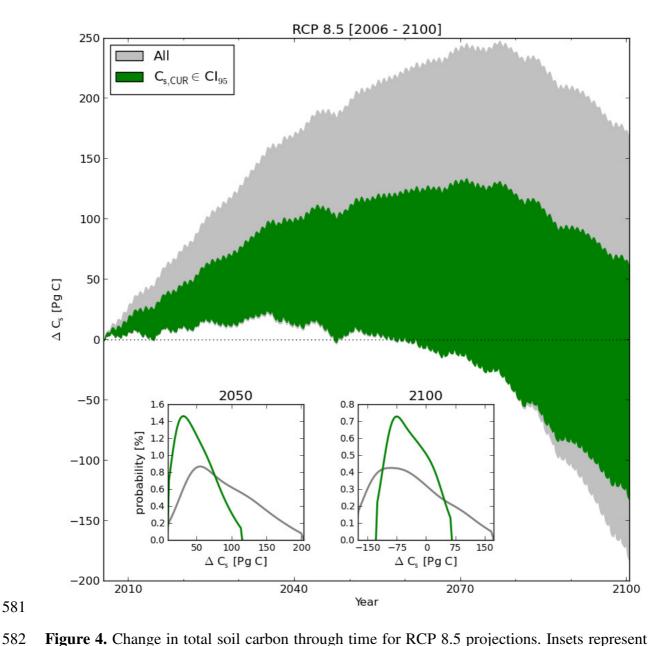


Figure 4. Change in total soil carbon through time for RCP 8.5 projections. Insets represent the probability density function of the change since 2005 for the indicated year. Grey is for all simulations while green is used to distinguish simulations for which total soil carbon is within the CI₉₅ of the HWSD in 2005.

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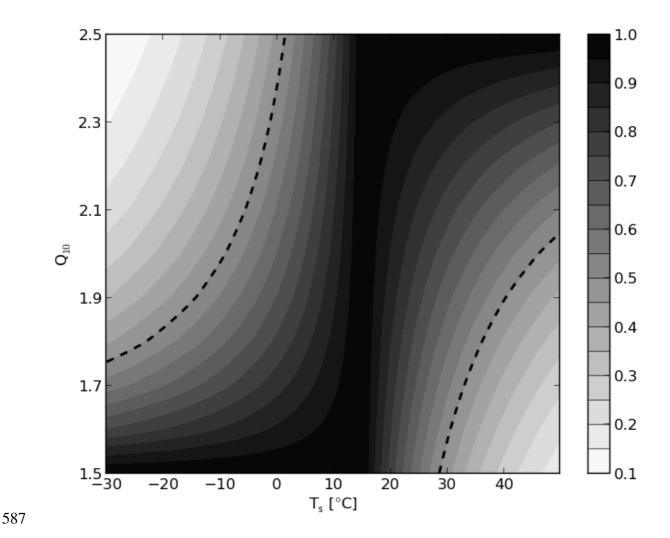


Figure 5. Values of f_T as a function of T_s and Q_{10} . For each temperature, the value is expressed as the proportion of the maximum value achieved for any value of Q_{10} . Areas outside of the dashed lines represent where f_T is less than 50% of the maximum for the same temperature.

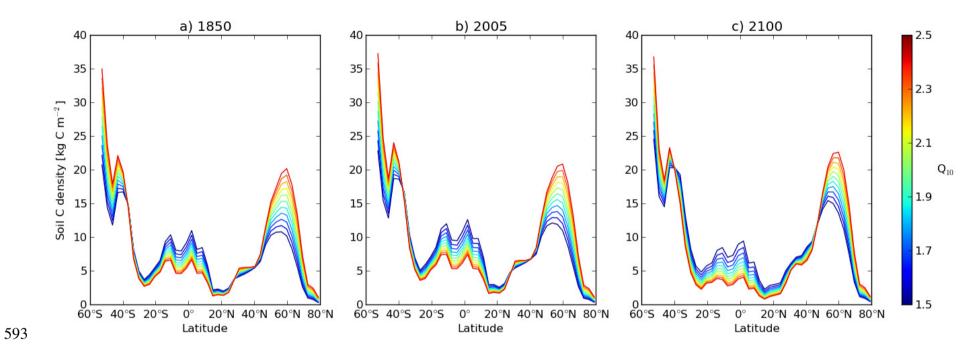


Figure 6. Zonal average soil carbon density in the reduced complexity model with k=180 months and various values of Q_{10} as indicated by the colour bar.

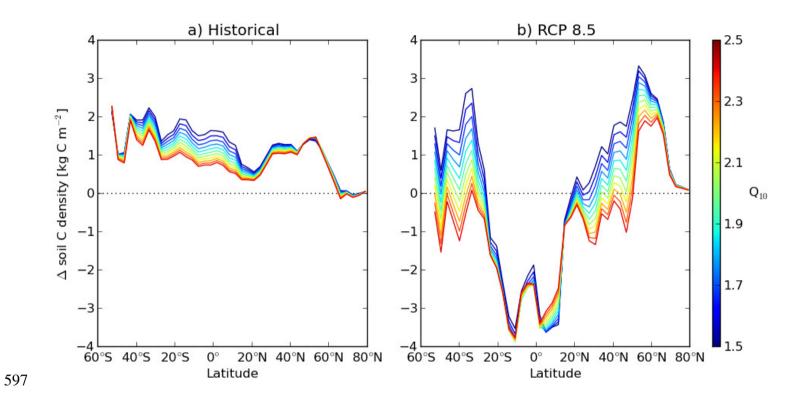


Figure 7. Zonal change in soil C density during historical simulations (a) and RCP8.5 (b)