

**Constraint of soil moisture on CO₂ efflux from tundra lichen, moss,
and tussock in Council, Alaska, using a hierarchical Bayesian model**

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Abstract

The tundra ecosystem is quite vulnerable to drastic climate change in the Arctic, and the quantification of carbon dynamics is of significant importance in response to thawing permafrost, changes to the snow-covered period and snow and shrub community extent, and the decline of sea ice in the Arctic. Here, CO₂ efflux measurements using a manual chamber system within a 40 m × 40 m (5-m interval; 81 total points) plot were conducted within dominant tundra vegetation on the Seward Peninsula of Alaska, during the growing seasons of 2011 and 2012, for the assessment of driving parameters of CO₂ efflux. We applied a hierarchical Bayesian (HB) model—a function of soil temperature, soil moisture, vegetation type, and thaw depth—to quantify the effects of environmental factors on CO₂ efflux, and to estimate growing season CO₂ emission. Our results showed that average CO₂ efflux in 2011 is 1.4 times higher than in 2012, resulting from the distinct difference in soil moisture between the two years. Tussock-dominated CO₂ efflux is 1.4 to 2.3 times higher than those measured in lichen and moss communities, reflecting tussock as a significant CO₂ source in the Arctic, with wide area distribution on a circumpolar scale. CO₂ efflux followed soil temperature nearly exponentially from both the observed data and the posterior medians of the HB model. This reveals that soil temperature regulates the seasonal variation of CO₂ efflux and that soil moisture contributes to the interannual variation of CO₂ efflux for the two growing seasons in question. Obvious changes in soil moisture during the growing seasons of 2011 and 2012 resulted in an explicit difference between CO₂ effluxes—742 and 539 gCO₂ m⁻² period⁻¹ for 2011 and 2012, respectively, suggesting the 2012 CO₂ emission rate was constrained by 27 % (95 % credible interval: 17-36 %) compared to 2011, due to higher soil moisture from severe rain. The estimated growing season CO₂ emission rate ranged from 0.86 in 2012 to 1.20 MgCO₂ in 2011 within a 40 m × 40 m plot, corresponding to 86 % and 80 % of annual CO₂ emission rates within the Alaska western tundra ecosystem, estimated from the temperature dependence of CO₂ efflux. Therefore, this HB model can be readily applied to observed CO₂ efflux, as it demands only four environmental factors and can also be effective for quantitatively assessing the driving parameters of CO₂ efflux.

1. Introduction

Carbon dioxide (CO₂) efflux from the soil surface to the atmosphere is important for estimating regional and global carbon budgets (Schlesinger and Andrews, 2000; Bond-Lamberty and Thomson, 2010), as well as being susceptible to increasing air temperature (Bond-Lamberty and Thomson, 2010), the degradation of permafrost (Schuur et al., 2009; Jensen et al., 2014), and the expansion of the shrub community (Sturm et al., 2005). All of which suggests alteration of the terrestrial carbon cycle in response to drastic climate change in the Arctic (ACIA, 2004).

The tundra ecosystem of Alaska has received increased attention for the enhanced greening in abundant Arctic coastal shrubs that has come with the decline of sea ice (Bhatt et al., 2010; 2013; Post et al., 2013), the shortened snow-covered period (Hinzman et al. 2005), thawing permafrost, and shrinking ponds and lakes (Romanovsky et al. 2002; Yoshikawa and Hinzman, 2003; Hinzman et al. 2005; Smith et al. 2005)—all reflecting the changes in terrestrial carbon and water cycles (Davidson et al., 1998; Oechel et al. 2000; Michaelson and Ping, 2003; ACIA, 2004; Oberbauer et al. 2007; Walter et al. 2007; Koven et al. 2011). Recently, Jensen et al. (2014) found a distinct difference in CO₂ efflux from undisturbed tundra during 2011 and 2012, resulting from greater rainfall in the growing season of 2012. This suggests that higher soil moisture from rainfall is a suppressant factor for soil-produced CO₂ emitted to the atmosphere (Davidson et al., 1998; Oberbauer et al. 2007), decreasing CO₂ emission by 43 % (Jensen et al., 2014). Davidson et al. (1998) reported that CO₂ efflux increased with soil moisture from 0 to 0.2 m³ m⁻³, steadily decreasing after with increasing soil moisture content beyond 0.2 m³ m⁻³. Hence, CO₂ efflux magnitude depends profoundly on the extent of soil moisture. Further, soil temperature is well known as a significant factor for regulating CO₂ efflux in worldwide terrestrial ecosystems, as reported by many researchers (Davidson et al., 1998; Xu and Qi, 2001; Davidson and Janssens, 2006; Rayment and Jarvis, 2000; Kim et al., 2007, 2013; Jensen et al., 2014). Q₁₀ value, which is a measure of the change in reaction rate at intervals of 10 °C (Lloyd and Taylor, 1994), has been effectively used to evaluate the temperature sensitivity of soil microbial activity as an exponential function (Davidson et al., 1998; Xu and Qi, 2001; Monson et al., 2006; Bond-Lamberty and Thomson, 2010; Kim et al., 2013). For example, Monson et al.

(2006) estimated their highest Q_{10} value, 1.25×10^6 , as the beneath-snowpack soil temperature warmed from -3 to 0 °C in a high-elevation subalpine forest in Colorado, reflecting higher CO₂ production by beneath-snow microbes (such as snow molds) during the end winter and early spring season. In the well-drained soil of Zachenberg, Greenland, higher CO₂ concentration in frozen soil came from a soil-thawing spring burst event, related to the trapping of produced CO₂ during winter. Subsequently, there is a distinct difference in Q_{10} value between above and below zero temperatures: Q_{10} value below zero was 430, even when water content was 39 % (Elberling and Brandt, 2003). Therefore, soil temperature, which is an analogue of soil microbial activity under the assumption that soil moisture and substrate availability are not limiting factors, is the most important factor in producing CO₂ in the soil.

Monthly CO₂ efflux measured in the tundra ecosystem has been further recognized as having insufficient spatiotemporal resolution and efflux data representativeness from the conventional dynamic chamber method (Hutchinson and Livingston, 2002; Savage and Davidson, 2003). Oberbauer et al. (1992) developed a mathematical model, which proved that soil temperature and water table depth might be used as efficient predictors of ecosystem CO₂ efflux in the riparian tundra of the northern foothills of Alaska. In order to overcome the weakness of monthly CO₂ efflux measurement in the field, the hierarchical Bayesian (HB) model framework can be applied for estimation of CO₂ efflux from the tundra ecosystem, as in Clark (2005) and Nishina et al. (2009; 2012). Their results indicated that the HB model is an effective tool for the estimation of fluxes and evaluation of parameters with less bias. Lately, free software such as WinBUGS (<http://www.mrc-bsu.ac.uk/bugs>) has resulted in the availability of a HB model using the Markov Chain Monte Carlo (MCMC) method (Spiegelhalter and Best, 2000). Clark (2005) described that the HB model reveals complex nonlinear relationships between efflux and environmental factors.

In this study, we modeled observed CO₂ efflux using a HB model with four explanatory variables: soil temperature, soil moisture, vegetation types, and thaw depth, all under the assumption of the lognormal distribution. The HB model used in this study accommodated nonlinear relationships between efflux and environmental factors. Therefore, the objectives of this study are to 1) evaluate the effects of dominant plants on CO₂ efflux; 2) quantitatively assess

driving parameters of CO₂ efflux simulated by a hierarchical Bayesian (HB) model; and 3) estimate growing season CO₂ emission rate within a 40 m × 40 m plot in the western Alaska tundra ecosystem.

2. Materials and Methods

2.1. Study Site and Experimental Methods

The study site is dominantly covered by typical tussock tundra. This site is located at the community of Council (64°51'38.3" N; 163°42'39.7" W; 45 m.a.s.l.) on the Seward Peninsula, about 120 km northeast of Nome, Alaska. This site was selected for its relatively smooth transition from forest to tundra, with underlying discontinuous permafrost regime. The monthly average air temperature of 1.2 °C at the Nome airport from 1971 to 2010 ranged from -10.5 °C in January to 14.6 °C in July. Annual average precipitation was 427 mm, including snowfall (Western Regional Climate Center). During the growing seasons (June to September) of 2011 and 2012, average ambient temperature and precipitation were 8.9 ± 1.0 °C (CV, coefficient of variance: 12 %) and 285 mm, and 8.5 ± 2.8 °C (CV: 33 %) and 380 mm, respectively, as shown in Figure 1. Precipitation in July-August of 2011 and 2012 were 231 and 299 mm, respectively, corresponding to 81 and 79 % of growing season precipitation. Under heavy precipitation in early July of 2011, CO₂ efflux measurement could not be conducted, unfortunately, due to underestimation of CO₂ efflux. The sampling periods were June 17-24, August 2-8, and September 9-15 for 2011, and June 20-29, July 14-21, August 11-18, and September 8-15 for 2012. The Alaska DOT (Department of Transportation) has maintained the access road from Nome to Council from late May to late September. Because this access road was closed during the snow-covered period (October to May), we could not conduct CO₂ efflux measurement during the non-growing season. The Council site has been managed by the WERC (Water Environmental Research Center) of UAF (University of Alaska Fairbanks) since 1999, for changes in permafrost and the water cycle (Yoshikawa and Hinzman, 2003).

This study determined CO₂ efflux and environmental factors in lichen-, moss-, and tussock-dominant tundra microsites within a 40 m × 40 m plot (5-m interval; 81 points) at this

1 site during the growing seasons of 2011 and 2012. Our plot was established for better
2 understanding of spatiotemporal variations of CO₂ efflux and environmental data. Within the
3 81-point area, on-ground dominant plants are lichen (*Cladonia mitis*, *Cladonia crispata*, and
4 *Cladonia stellaris*); moss, such as sphagnum (*Sphagnum magellanicum*, *Sphagnum angustifolium*,
5 and *Sphagnum fuscum*) and others (*Polytricum* spp., *Thuidium abietinum*, and *Calliergon* spp.);
6 and cotton grass tussock tundra (*Eriophorum vaginatum*). Dominant lichen, moss, and tussock
7 tundra were occupied the plot at 27, 53, and 20 % within the plot, respectively.

8 Soil temperatures were taken at 5 and 10 cm below the surface using a portable thermometer with
9 two probes (Model 8402-20, Cole-Palmer, USA), and soil moisture was measured at each point
10 with a portable soil-moisture logger (HH2, Delta-T Devices, UK) with sensor (ML2, Delta-T
11 Devices, UK). Thaw depth was measured with a fiberglass tile probe (1.5 m long), and pH with a
12 waterproof meter (IQ 160, Ben Meadows, USA) in September 2011, for soil characteristics. A
13 one- or two-way ANOVA (95 % confidence level) and data regression analysis using Microsoft
14 Excel Data Analysis software were performed.

15 **2.2. Estimation of CO₂ Efflux**

16 Our dynamic CO₂ efflux-measuring system was portable, convenient, and capable of calculating
17 efflux *in-situ*. The 81-cylinder chamber base (30-cm dia., 40-cm height) was fixed to the surface
18 at each point. This system consisted of a transparent-material chamber lid (35-cm dia., 0.3-cm
19 thickness) with input and output urethane tubing (6 mm OD; 4 mm ID) and a pressure vent, a
20 commercial pump (CM-15-12, Enomoto Micro Pump Co., Ltd., Japan), an NDIR CO₂ analyzer
21 (Licor-820, LICOR Inc., USA), a commercial 12-V battery, and a laptop computer for efflux
22 calculation (Kim et al., 2013). This system is similar to the manual system by Savage and
23 Davidson (2003; see Figure 1). To minimize the effect of pressure inside the chamber, the flow
24 rate of the pump was maintained at 0.5 Lmin⁻¹, due to under/overestimation of CO₂ efflux by
25 under- or over-pressurization of the chamber used and caused by flow restrictions in air
26 circulation design (Davidson et al., 2002). Efflux-measuring time was on a 5-10 minute interval,
27 depending on weather and soil surface conditions. For tussock CO₂ efflux estimates, the surface
28 area was variable and dependent on height; average tussock height in this case was 18.7 ± 5.1 cm

(CV: 27 %).

Efflux was calculated from the following equation, as described by Kim et al., (2013):

$$F_{CO_2} = \rho_a \times (\Delta C / \Delta t) \times (V/A), \quad (1)$$

where F_{CO_2} represents measured soil CO₂ efflux (g CO₂ m⁻² min⁻¹), ρ_a is the molar density of dry air (mol m⁻³), ΔC (ppmv) is the change in CO₂ concentration during measuring time (Δt , 5 to 10-min), V is chamber volume, and A is surface area (cross section = 0.070 m²). The height of each chamber was also measured alongside the chamber to allow calculation of the efflux.

To assess the response of temperature dependence on CO₂ efflux, the relationship was plotted, showing exponential curves for soil temperature at depths of 5 and 10 cm from this equation:

$$F_{CO_2} = \beta_0 \times e^{\beta_1 \times T}, \quad (2)$$

where T is soil temperature (°C), and β_0 and β_1 are constants. This exponential relationship is commonly used to represent soil carbon efflux as a function of temperature (Davidson et al., 1998; Xu and Qi, 2001; Davidson and Janssens, 2006; Rayment and Jarvis, 2000; Kim et al., 2007, 2013). Q_{10} temperature coefficient values were calculated as in Davidson et al. (1998) and Kim et al. (2013):

$$Q_{10} = e^{\beta_1 \times 10} \quad (3)$$

Q_{10} is a measure of the change in reaction rate at intervals of 10 °C, based on Van't Hoff's empirical rule that a rate increase on the order of two to three times occurs for every 10 °C rise in temperature (Lloyd and Taylor, 1994).

2.3. Description of Hierarchical Bayesian (HB) model

To evaluate the relationship between CO₂ efflux and environmental variables, we modeled observed CO₂ efflux using an HB model with four explanatory variables: soil temperature (ST), soil moisture (SM), vegetation types (Vege), and thaw depth (THAW).

1 First, CO₂ efflux (F_{CO_2}) was assumed normally distributed with mean parameter (μ_{flux}) and
 2 variance parameter (σ):

$$3 \quad F_{CO_2} \sim normal(\mu_{flux}, \sigma^2), \quad (4)$$

4 The scale parameter (μ_{flux}) was determined from the following equation:

$$5 \quad \mu_{flux} = f_P f_{ST} f_{SM} f_{THAW}, \quad (5)$$

6 where f_P represents the function of CO₂ efflux potential, f_T and f_{SM} are limiting response functions,
 7 ranging from 0 to 1. f_P was defined as follows:

$$8 \quad f_P = \beta_0 + Vege_{[k]} + Year_{[l]} + Posi_{[ij]}, \quad (6)$$

9 in which f_P is a linear predictor with intercept (β_0) and three random effects ($Vege$, $Year$, and
 10 $Posi$). The $Posi$ term represents the spatial random effect of the conditional autoregressive model
 11 (CAR) proposed by Besag et al. (1991).

12 Temperature (f_T) uses a modified Van't Hoff equation as follows:

$$13 \quad f_{ST} = e^{\frac{ST - ST_{ref}}{10} \log(Q_{10})}, \quad (7)$$

14 where f_{ST} is the temperature response function, varying from 0 to 1. The explanatory variable of
 15 this function, represented by ST and ST_{ref} , is a constant, set at 25 °C for this study. The
 16 temperature sensitivity parameter is shown by Q_{10} . The soil moisture limiting function (f_{SM}) is
 17 defined as follows:

$$18 \quad f_{SM} = \left(\frac{SM - a}{b - a} \right)^a \left(\frac{SM - c}{b - c} \right)^{-d(b - c)/(b - a)}, \quad (8)$$

19 where the soil moisture response function, f_{SM} , ranges from 0 to 1, and is the same as the
 20 temperature response function (Hashimoto et al., 2010). SM is the explanatory variable of this
 21 function, and a , b , c , and d are the parameters for determining the shape of the soil moisture
 22 function. The function has a convex shape, and values range from 0 to 1. Parameters a and c are

the minimum and maximum values of SM, respectively (i.e., $g(a) = g(c) = 0$). Parameter b , which ranges between a and c , is the optimum parameter (i.e., $g(b) = 1$). Parameter d controls the curvature of the function, though the three other parameters also affect the shape. This function was adopted from the DAYCENT model (Parton et al., 1996; Del Grosso et al., 2000).

f_{THAW} is a function of thaw depth. We modeled this as follows:

$$f_{THAW} = \frac{1}{1 + e^{k-r THAW}}, \quad (9)$$

where the thaw depth function also ranges from 0 to 1. $THAW$ is the explanatory variable of this function, and k and r are the parameters. We assumed CO₂ efflux to monotonically increase together with thaw depth (depth of active layer); however, these increases are not simply proportional, due to carbon depth distribution.

Finally, we modeled the priors of each parameter. For vegetation, we incorporated a random effect as follows:

$$Vege_k \sim normal(0, \sigma_{vege}); \quad (10)$$

$$Year_l \sim normal(0, \sigma_{year}). \quad (11)$$

For spatial explicit random effect, we used CAR modeling (Besag et al., 1991), as follows:

$$Posi_{il} \sim normal(b_{ij}, \frac{\sigma_{posi_{ij}}}{n})$$

$$b_{il} \sim \frac{1}{n_{ij}} \sum_{m=1}^{neighbors(ij)} b_m,$$

where n_{ij} is the number of neighbors for neighborhood ij .

For priors, we defined as follows:

$$\beta_0 \sim normal(0, 1000),$$

$$Q_{10} \sim uniform(1, 10),$$

$$\begin{aligned}
1 \quad & a \sim \text{uniform}(-2, 0), \\
2 \quad & b \sim \text{uniform}(0.1, 0.5), \\
3 \quad & c \sim \text{uniform}(1, 3), \\
4 \quad & d \sim \text{uniform}(0.01, 10), \\
5 \quad & k \sim \text{uniform}(0, 10), \\
6 \quad & r \sim \text{uniform}(0, 1), \\
7 \quad & \sigma^2 \sim \text{uniform}(0, 100), \\
8 \quad & \sigma_{vege}^2 \sim \text{uniform}(0, 100), \\
9 \quad & \sigma_{year}^2 \sim \text{uniform}(0, 100), \tag{12}
\end{aligned}$$

10 For β_0 , we used a normal distribution with mean 0 and a very large variance. Priors regarding
11 soil moisture function (a, b, c, d) are based on Hashimoto et al. (2012). We set priors for σ_{vege}^2
12 and σ_{year}^2 to be vague, meaning large enough in value to accommodate the actual observed CO₂
13 efflux of this study.

14 The joint posterior probability was described as follows:

$$\begin{aligned}
& p(\theta|data) \propto \prod \text{Normal}(F_{CO_2}|\mu, \beta_{0,10}, a, b, c, d, k, r, \sigma_1, \sigma_{vege}, \sigma_{year}, \sigma_{posi}) \\
& \quad \times p(\beta_0) \times p(Q_{10}) \times p(a) \times p(b) \times p(c) \times p(d) \\
15 \quad & \quad \times p(k) \times p(r) \times p(\sigma_1) \times p(\sigma_{vege}) \times p(\sigma_{year}) \times p(\sigma_{posi}), \tag{13}
\end{aligned}$$

16 where $p(\theta)$ denotes priors. For this model, we used MCMC methods implemented with
17 Bayesian inference using the Gibbs sampling software WinBUGS (WinBUGS, version 1.4.3; D.
18 Spiegelhalter et al., 2007, available at <http://www.mrc-bsu.ac.uk/bugs>), and the Gelman-Rubin
19 convergence diagnostic as an index. For the model, we ran 20,000 Gibbs sampler iterations for
20 three chains, with a thinning interval of 10 iterations. We discarded the first 10,000 iterations as
21 burn-in, and used the remaining iterations to calculate posterior estimates. R was used to call
22 JAGS/WinBUGS and calculate statistics in R.

3. Results and Discussion

3.1. CO₂ Efflux and Environmental Factors

Table 1 shows monthly average \pm standard deviation (Coefficient of Variance, %) of CO₂ efflux, soil temperature at 5 and 10 cm below the surface, soil moisture, thaw depth, and pH in lichen, moss, tussock tundra, and grass during the growing seasons of 2011 and 2012. Annual growing season average CO₂ efflux is 4.6 ± 2.5 (54 %) and 3.1 ± 2.0 (66 %) mgCO₂ m⁻² min⁻¹ for 2011 and 2012, respectively. This indicates that growing season CO₂ efflux in 2011 was 1.5 times higher than in 2012, as well as the significance of heavy rainfall during the mid-growing season of 2012. CO₂ efflux in tussock tundra was approximately 1.8 times greater than in other plants, which may be due more to tussock's relatively wider surface area than others. While surface area in lichen and moss is 0.070 m²—the same surface area of the measurement chamber—average surface area of tussock is 0.090 ± 0.024 m², based on an average height of 19.2 ± 5.1 cm. CO₂ efflux in the Arctic tundra of Alaska ranged from 0.38 to 1.6 mgCO₂ m⁻² min⁻¹ in lichen and 0.44 to 4.3 mgCO₂ m⁻² min⁻¹ in tussock during the growing season (Poole and Miller, 1982). Within tundra near Barrow, Alaska, meanwhile, CO₂ efflux in tussock and wet sedge was 0.23 and 0.022 mgCO₂ m⁻² min⁻¹, respectively (Oechel et al., 1997), suggesting that CO₂ efflux in tussock is indeed a more significant atmospheric CO₂ source than wet sedge. Kim et al. (2013) reported that tussock is an important source of carbon efflux to the atmosphere, contributing 3.4-fold more than other vegetation types in Alaska tundra and boreal forest systems. Further, tussock-originated CO₂ efflux, which occupies a circumpolar area ranging from 9×10^{11} m² (Miller et al. 1983) to 6.5×10^{12} m² (Whalen and Reeburgh, 1988) when counted with moss species, provides a quantitative understanding of a significant atmospheric carbon source from the Arctic terrestrial ecosystem. Considering the circumpolar distribution of tussock tundra and moss in the Arctic tundra ecosystem, CO₂ efflux measured in this study should not be overlooked in the evaluation of the regional/global carbon budget regarding distribution characteristics of on-ground plants.

The spatial distribution of CO₂ efflux within a 40 m \times 40 m plot in 2011 and 2012 is shown in

Figure 2. CO₂ efflux in June 2011 was much higher than during other observation periods, reflecting the effects of higher air temperature and lower precipitation in June (see Figure 1). This further suggests an explicit difference in CO₂ efflux between June of 2011 and June 2012 within the plot, as shown in Table 1. We also note that CO₂ efflux in September 2012 rapidly decreased, due to heavy rainfall from mid-August to mid-September 2012. Within the plot, while the CV of monthly average CO₂ efflux in 2011 was prone to decrease, CV in 2012 tends to increase. This denotes a susceptibility to extreme environmental factors in 2012 compared to 2011.

Annual growing season average \pm standard deviation for soil temperatures at 5 and 10 cm below the soil surface were 9.0 ± 4.2 (47 %) and 5.9 ± 3.9 °C (66 %) for 2011 and 7.7 ± 4.5 (58 %) and 5.7 ± 3.5 °C (61 %) for 2012, respectively. This indicates soil temperature in 2011 was higher than in 2012, similar to annual average CO₂ efflux, suggesting soil temperature is likely to modulate CO₂ efflux, as largely reported in regions worldwide (Davidson et al., 1998; Xu and Qi, 2001; Davidson and Janssens, 2006; Rayment and Jarvis, 2000; Kim et al., 2007; 2013). The spatial distribution of high/low soil temperature for each month was identical to the pattern of high/low CO₂ efflux, as also shown in Figure 2.

Annual average soil moisture was 0.253 ± 0.158 (CV: 62 %) m³ m⁻³ in 2011 and 0.272 ± 0.180 m³ m⁻³ (66 %) in 2012, indicating moisture in 2011 was slightly lower than in 2012. Soil moisture in September 2011 was not measured, due to damage to the soil moisture sensor. Spatial distribution of soil moisture is related to geographical topography, such as slope and relief within the plot, reflecting spatial distribution of lower CO₂ efflux and lower soil temperature in the trough area (not shown). Soil moisture, along with soil temperature, is also an important factor in the control of CO₂ efflux (Davidson et al., 1998; Gaumont-Guay et al., 2006; Mahecha et al., 2010; Kim et al., 2013).

Average thaw depth was 39 ± 5 (15 %) cm in 2011 and 38 ± 6 (15 %) cm in 2012, showing no significant difference, based on a one-way ANOVA at the 95 % confidence level ($p < 0.001$). The distribution of thaw depth (not shown) appears similar to the soil moisture pattern, which is inversely related to CO₂ efflux and soil temperature. The average thaw rate over our 81 points

was 0.43 cm day^{-1} in 2011 and 0.41 cm day^{-1} in 2012, reflecting that thaw rate over time remains almost constant during the growing season, and that thaw depth is not considered to regulate CO_2 efflux. In general, the deeper the active layer in response to permafrost thaw in the Arctic (Marchenko et al., 2008), the more CO_2 emission from the soil to the atmosphere (Elberling et al., 2013), also suggesting the potential decomposition of frozen, higher-stocked soil organic carbon (Ping et al., 2008; Tarnocai et al., 2009; Grosse et al., 2011). However, temporal variation in thaw depth of the active layer may not stimulate CO_2 production. This suggests the strength of CO_2 production that depends on soil microbial metabolism is affected more by environmental factors than constant active layer depth for both years. The deeper active layer reached to nearly 80 cm below the surface with the soil temperature profile at 50, 70, 80, and 92 cm from July 2012 to October 2013 (not shown). When the soil contained much higher soil moisture and much deeper thaw depth for September, pH represented a similar value of 3.8 ± 0.4 (11 %), representing an acidic tundra soil ($\text{pH} < 5.5$; Walker et al., 1998) among all points. The pH measurement was not conducted during the growing season of 2012, due to near uniformity within the plot.

3.2. Environmental Factors Determining CO_2 Efflux

CO_2 efflux is potentially modulated by environmental factors such as soil temperature, soil moisture, and thaw depth. Q_{10} values were calculated using Eq. (3), based on the exponential relationship between CO_2 efflux and soil temperature at 5- and 10-cm depths for each plant. Table 2 shows Q_{10} values and correlation coefficients between CO_2 efflux and soil temperature at 5- and 10-cm depths in lichen, moss, grass, and tussock tundra during the growing season, based on a one-way ANOVA with a 95 % confidence level. Q_{10} is prone to increasing with time, suggesting that CO_2 production by soil microbes and roots has greater sensitivity to a narrower range of soil temperatures, such as in the spring and fall seasons (Rayment and Jarvis, 2000; Gaumont-Guay et al., 2006; Monson et al., 2006; Malcom et al., 2009). In this Alaska tundra ecosystem, average daily CO_2 efflux from wet sedge followed soil surface temperature closely, increasing exponentially as soil surface temperature increased, while efflux from the tussock tundra ecosystem followed soil surface temperature nearly logarithmically (Oechel et al., 1997).

1 In this study, the response from CO₂ efflux in tussock tundra to soil temperature depicts an
2 almost linear relationship; however, it shows an exponential curve for Q₁₀ values, listed in Table
3 2. Soil temperature at 5-cm depth explained 86 % and 70 % of the variability in CO₂ efflux for
4 2011 and 2012, respectively, from the linear relationships, demonstrating that soil temperature is
5 a significant factor in driving CO₂ efflux in dominated tundra plants during the growing season.
6 The Q₁₀ value for soil temperature at 5-cm depth for the moss regime in August 2012 was the
7 lowest, at 1.15, resulting from higher soil temperature and higher soil moisture in August 2012
8 (Table 1).

9 Figure 3 shows the responses from monthly averaged CO₂ efflux to soil temperature at 5- and
10 10-cm depths (a1 and b1), soil moisture (a2 and b2), and thaw depth (a3 and b3), and the
11 responses from soil temperature at 5 cm to soil moisture (a4 and b4) and thaw depth (a5 and b5)
12 during the growing seasons of 2011 and 2012. Except for a1 and b1, these relationships were
13 each negatively related during the growing season of 2011-12. However, except for data
14 measured in September 2012, these relationships denoted positive lines from June to August
15 2012, as also shown in Figure 3 (b2-5). This seems to be the effect of heavy rainfall beginning in
16 August 20, 2012, as shown in Figure 1, which represents daily and cumulative precipitation in
17 2011 and 2012. Interestingly, cumulative rainfall indeed began to surpass 2011 cumulative
18 precipitation on August 20, 2012 (not shown). The correlation coefficient (R^2) from June to
19 August 2012 ranged from 0.01 in Figure 3 (b3) to 0.32 in (b2). Hence, soil moisture elucidated
20 32 % of the variability in CO₂ efflux before the severe rainfall event of the fall season of 2012,
21 demonstrating that soil moisture is another important factor aside from soil temperature. Jensen
22 et al. (2014) estimated 2.3 ± 0.2 and 1.3 ± 0.11 mgCO₂ m⁻² min⁻¹ in the northwestern tundra of
23 Alaska in July of 2011 and 2012, respectively, suggesting lower carbon flux results from the
24 stronger rainfall event in 2012 (see Figure 3a, Jensen et al., 2014), with a similar trend in air
25 temperature between both years. This rainfall may have possibly inhibited 43 % of CO₂ emission
26 from the soil surface with increasing soil moisture in 2012, indicating a similar result to those
27 observed in this study (Davidson et al., 1998).

3.3. Simulated CO₂ Efflux from a Hierarchical Bayesian Model

We used 486 datasets of CO₂ efflux, soil temperature, soil moisture, vegetation types, and thaw depth for adjusting the parameters of a hierarchical Bayesian (HB) model, and the posterior distribution of the parameter for the CO₂ efflux is summarized in Table 3. Potential CO₂ effluxes from the dominated plants calculated from posterior medians of the model were 16.8 mgCO₂ m⁻² min⁻¹ in grass (95 % predicted credible intervals (CI), 13.7-20.4 mgCO₂ m⁻² min⁻¹), 15.3 mgCO₂ m⁻² min⁻¹ in lichen (95 % predicted CI, 11.1-16.8 mgCO₂ m⁻² min⁻¹), 14.8 mgCO₂ m⁻² min⁻¹ in moss (95 % predicted CI, 10.2-15.9 mgCO₂ m⁻² min⁻¹), and 21.9 mgCO₂ m⁻² min⁻¹ in tussock (95 % predicted CI, 24.0-31.0 mgCO₂ m⁻² min⁻¹). This suggests that the contribution of atmospheric carbon from tussock tundra should receive attention when it comes to the tundra ecosystem and a circumpolar scale response to the changing climate in the higher northern hemisphere (Oechel et al., 1997; Bhatt et al., 2010; 2013; Kim et al., 2013). We computed limiting functions for soil temperature, soil moisture, and thaw depth of CO₂ efflux simulated by posterior distributions (n = 1,000), as shown in Figure 4, for the quantitative assessment of the driving parameters for CO₂ efflux. Because changes in vegetation within the plot were not observed during this study period, these two parameters are not correlated with one another. In actuality, there was very low correlation ($R^2 = 0.019$) between t_{veg} and t_{year} in our results.

For soil temperature limiting functions, the parameter simulated from the posterior median followed soil temperature nearly exponentially (Figure 4a), demonstrating the definite temperature dependency of CO₂ efflux (Raich and Schlesinger, 1992; Davidson et al., 1998; Gaumont-Guay et al., 2006; Mahecha et al., 2010; Kim et al., 2013), as shown in Figure 3a1 and b1. For soil temperature response, the parameter Q_{10} value was 2.52 ± 0.12 (95 % predicted CI, 2.29-2.75). For soil moisture limiting functions (Figure 4b), optimum soil moisture value was 0.228 m³ m⁻³ (95 % predicted CI, 0.184-0.238 m³ m⁻³). CO₂ efflux tended to increase with an increase in soil moisture when soil moisture value was at the optimum, as shown in Figure 3a2 and b2. On the other hand, the response from CO₂ efflux to soil moisture changed to a negative trend beyond the optimum value for soil moisture. The results from Jensen et al. (2014) proved

the findings observed in this study, in which CO₂ efflux was relatively lower with the higher soil moisture observed in 2012, compared to 2011 (see Figure 4b, Jensen et al., 2014). Davidson et al. (1998) reported the correlation between soil water content and CO₂ efflux in different drainage classes. CO₂ efflux increased when soil water content was less than 0.2 m³ m⁻³; on the other hand, higher soil moisture resulted in a decrease in CO₂ efflux (see Figure 7, Davidson et al., 1998). For thaw depth limiting functions, the parameter increased to 20 cm, which represents the optimum thaw depth value (Figure 4c). While CO₂ efflux increased with the rise in thaw depth in June until optimum thaw depth value, efflux was constant, despite an increase in thaw depth with time. The response from CO₂ efflux to thaw depth turned to a negative trend during the growing seasons of 2011 and 2012, as shown in Figure 3a3 and b3. These findings suggest that thaw depth may not be a significant parameter in influencing CO₂ efflux in the tundra ecosystem, in spite of a deeper active layer over time.

Spatial distribution of simulated CO₂ efflux, calculated from the posterior medians of the hierarchical Bayesian model during the growing seasons of 2011 and 2012 and excluding July and September of 2011, is similar to that of measured CO₂ efflux, as shown in Figure 2. The pattern of simulated CO₂ efflux is nearly identical to the spatial distribution of measured CO₂ efflux (Figure 3), as simulated CO₂ efflux is a function of soil temperature, soil moisture, and thaw depth. Of these, we consider soil temperature the most important parameter in modulating CO₂ efflux in the tundra ecosystem during the growing season. We compared measured CO₂ efflux to predicted CO₂ efflux using posterior medians in the HB model at each sampling period of 2011 and 2012 (Figure 5), denoting that CO₂ efflux simulated by a nonlinear equation is consistent with measured data. Using the HB model, cumulative predicted CO₂ emission rates from June 28 to September 30 of 2011 and 2012—based on monitored soil temperature and soil moisture in the Council area—were 742 gCO₂ m⁻² period⁻¹ (95 % predicted CI, 646-839 gCO₂ m⁻² period⁻¹) and 539 gCO₂ m⁻² period⁻¹ (95 % predicted CI, 460-613 gCO₂ m⁻² period⁻¹), respectively. These findings suggest that the 2012 CO₂ emission rate is constrained by 27 % (95 % CI, 17-36 %) compared to the 2011 emission, demonstrating that higher soil moisture from severe rain constrains the emission of soil-produced CO₂ to the atmosphere (Jensen et al., 2014).

During the study periods (DOY: 179-273; Figure 6) of 2011 and 2012, average soil temperature was 9.3 ± 3.8 (CV: 41 %) and 8.6 ± 4.8 (CV: 56 %) °C, respectively, showing that there is no significant difference between the years, based on a one-way ANOVA 95 % confidence level. Trends in soil temperature during the periods of 2011 and 2012 were $ST = -0.135 \times DOY + 5522$ ($R^2 = 0.70$), and $ST = -0.093 \times DOY + 3781$ ($R^2 = 0.45$), respectively. On the other hand, trends for soil moisture were $SM = 0.0025 \times DOY - 103.5$ ($R^2 = 0.37$) in 2011 and $SM = -0.0008 \times DOY + 33.2$ ($R^2 = 0.31$) in 2012, as shown in 6. Average soil moisture was 0.260 ± 0.040 (15 %) and 0.493 ± 0.124 (25 %) $m^3 m^{-3}$ in 2011 and 2012, respectively, suggesting a distinct difference in soil moisture between the two years. Soil moisture during the 2012 period did not change with time, resulting from heavy rainfall events (Figure 1) during the growing season (Jensen et al., 2014). When soil temperature at the end of September in 2012 was below zero (Figure 6a), soil moisture sharply decreased, suggesting the frozen layer reached to soil moisture measuring depth (e.g., near surface), as shown in Figure 6b. The 2012 weather conditions may represent an episodic event, requiring additional monitoring for several representative points within the plot. Nevertheless, the higher CO₂ emission rate simulated by the HB model in 2011 is considered likely the result of CO₂ efflux increases until soil moisture reached optimum value, as shown in Figures 4b. Therefore, soil moisture plays an important parameter in constraining CO₂ emission in this tundra ecosystem, when the soil moisture is over the optimum value. When the annual simulated CO₂ emission rate was estimated from the relationship between CO₂ efflux and air temperature using equation (2), the annual emission rates were 827 and 609 gCO₂ m⁻² year⁻¹ in 2011 and 2012, respectively, corresponding to 86 and 80 % of annual CO₂ emission rates. Kim et al. (2013) estimated growing season CO₂ emission in the foothill tundra north of Brooks Range, Alaska was 645 gCO₂ m⁻² period⁻¹ during 2006-2010, despite the difference in latitudinal distributions for CO₂ efflux and parameter. This value is situated between the 2011 and 2012 emission rates simulated in Council in this study. That is, the simulated CO₂ emission rates were 0.86-1.20 MgCO₂ within a 40 m × 40 m plot during the growing seasons of 2012 and 2011, respectively.

4. Summary and Future Works

Here, CO₂ efflux measurement was conducted with a manual chamber system in the tundra ecosystem of the Seward Peninsula of western Alaska, during the growing seasons of 2011 and 2012, to evaluate the significant parameter(s) controlling CO₂ efflux, as well as the effect(s) on the soil-produced CO₂ emission rate to the atmosphere, using a hierarchical Bayesian (HB) model within a 40 m × 40 m plot (5-m interval; 81 points). Tussock tundra is an atmospheric carbon source in the tundra ecosystem year-round (Oechel et al., 1997; Kim et al., 2007, 2013). Considering the wide-ranged distribution of tussock in the high Northern Hemisphere, tussock- and moss-originated CO₂ efflux should not be overlooked as a significant carbon source in the estimation of regional and global carbon budgets. The response from CO₂ efflux in tussock to soil temperature denoted a linear relationship; meanwhile, effluxes observed in lichen and moss regimes increased exponentially as soil temperature increased. This finding suggests that soil temperature is a key environmental factor in modulating CO₂ efflux, as many scientists have also reported around the world. Except for data observed in September 2012, soil moisture played an important factor in controlling CO₂ efflux. For 2012, higher soil moisture, resulting from the heavy rainfall in the end of August, was a constraining factor for the transport of soil-produced carbon to the atmosphere (Davidson et al., 1998; Jensen et al., 2014).

Using the HB model, we computed limiting functions for soil temperature, soil moisture, and thaw depth of CO₂ efflux simulated by posterior distribution. Simulated CO₂ efflux increased both 1) exponentially as soil temperature increased and 2) nearly linearly until soil moisture was at optimum values (0.228 m³ m⁻³); however, efflux decreased 3) logarithmically when soil moisture was beyond the optimum, and 4) nearly linearly until thaw depth was at optimum value (20 cm). Finally, efflux remained constant when thaw depth increased with time. These simulated findings show similar patterns to the data obtained in this study, as well as to Jensen et al. (2014)'s results observed in the northwestern tundra of Alaska during the growing seasons of 2011 and 2012. During these growing seasons of 2011 and 2012, the difference in soil temperature between the two years was not significant; however, there was a distinct difference in soil moisture between them, resulting in the inhibition of CO₂ emissions due to higher soil

moisture. This demonstrates that higher soil moisture is constrained to 27 % of the CO₂ emission in 2012 compared to 2011. However, to prove the effect of soil moisture on controlling CO₂ emission in the tundra ecosystem, additional study must monitor the profiles of soil moisture and soil temperature at representative points from lichen, moss, and tussock tundra regimes within the plot. As conducted by Risk et al. (2011), the monitoring of soil CO₂ efflux must also show representative points, along with the monitoring of environmental factor profiles within the plot.

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Figure Legends

Figure 1. Average daily ambient temperature and precipitation in Council, Seward Peninsula, Alaska during April-October of 2011 and 2012 (Western Regional Climate Center). Dotted arrows denote that cumulative rainfall in 2012 exceeds that of 2011, beginning August 20, 2012.

Figure 2. Spatial distribution of CO₂ efflux (mgCO₂ m⁻² min⁻¹) within a 40 m × 40 m plot (5-m interval; 81 points), Council, Seward Peninsula, Alaska during the growing seasons of 2011 (upper panel) and 2012 (lower). Due to heavy rain in early July 2011, this data could not be measured, as shown in Figure 1.

Figure 3. Responses from monthly average CO₂ efflux to (1) average soil temperature at 5 and 10 cm (open and solid circles), (2) average soil moisture, and (3) thaw depth, as well as responses from average soil temperature at 5 cm to (4) average soil moisture and (5) average thaw depth during the growing seasons of (a) 2011 and (b) 2012. Dashed curves (a1 and b1) and dotted lines indicate the relationship between the two. Furthermore, solid lines in b2-5 denote the relationship between factors, except for data measured in September.

Figure 4. Limiting functions for (a) soil temperature, (b) soil moisture, and (c) thaw depth of CO₂ efflux simulated by posteriors (n = 1000). Red solid lines are simulated from posterior median.

Figure 5. Response from measured CO₂ efflux on simulated CO₂ efflux by posterior medians in the HB model as a function of soil temperature, soil moisture, and thaw depth within a 40 m × 40 m plot (5-m interval; 81 points), Council, Seward Peninsula, Alaska during the growing seasons of 2011 and 2012. (Reviewer #2)

Figure 6. Temporal variations in (a) soil temperature (°C) and (b) soil moisture (m³ m⁻³), measured for tundra sites during the growing seasons of 2011 (black) and 2012 (red). When soil temperature was below zero, at the end of September 2012, soil moisture dropped rapidly, as shown in Figure 1.

*Supplementary material: Convergence plot of all HB model parameters

Table 1. Average \pm standard deviation (coefficient of variation, %) of CO₂ efflux, soil temperature at 5 and 10 cm below the surface, soil moisture, thaw depth, and pH in lichen, moss, and tussock tundra, Council, Seward Peninsula, Alaska during growing seasons of 2011 and 2012

Month	Vegetation	n	CO ₂ efflux	Soil temperature (°C)		Soil moisture	Thaw depth	pH
			(mgCO ₂ /m ² /min)	5 cm	10 cm	(m ³ /m ³)	(cm)	
June, 2011	Lichen	22	5.7 \pm 3.6 (63)	10.1 \pm 2.5 (25)	3.3 \pm 1.4 (42)	0.270 \pm 0.162 (60)	22 \pm 3 (12)	n.m. [#]
	Moss	43	7.8 \pm 2.2 (29)	13.2 \pm 2.9 (22)	6.7 \pm 2.8 (42)	0.224 \pm 0.122 (54)	21 \pm 3 (14)	n.m.
	Tussock	16	12.9 \pm 6.2 (48)	12.7 \pm 3.3 (26)	7.6 \pm 3.7 (48)	0.301 \pm 0.116 (39)	22 \pm 2 (11)	n.m.
	Average	81*	8.0 \pm 3.6 (45)	12.3 \pm 3.2 (53)	6.0 \pm 3.1 (51)	0.255 \pm 0.127 (49)	21 \pm 3 (14)	
August, 2011	Lichen	24	2.5 \pm 1.2 (47)	6.9 \pm 1.5 (22)	4.4 \pm 1.1 (25)	0.297 \pm 0.200 (67)	38 \pm 5 (14)	n.m.
	Moss	41	3.3 \pm 1.7 (52)	9.0 \pm 1.6 (18)	6.2 \pm 1.7 (27)	0.264 \pm 0.237 (90)	41 \pm 8 (19)	n.m.
	Tussock	16	5.1 \pm 2.7 (53)	9.4 \pm 2.4 (25)	7.0 \pm 2.1 (30)	0.256 \pm 0.141 (55)	40 \pm 5 (12)	n.m.
	Average	81*	3.3 \pm 1.3 (39)	8.6 \pm 1.9 (22)	5.8 \pm 1.4 (24)	0.272 \pm 0.180 (66)	40 \pm 6 (15)	
September, 2011	Lichen	23	2.3 \pm 0.9 (40)	6.2 \pm 1.0 (16)	4.6 \pm 1.0 (21)	- **	57 \pm 8 (13)	3.7 \pm 0.4 (7)
	Moss	43	2.5 \pm 1.2 (50)	6.9 \pm 1.4 (20)	5.6 \pm 1.3 (23)	-	58 \pm 12 (20)	3.8 \pm 0.4 (11)
	Tussock	15	3.5 \pm 1.5 (43)	6.5 \pm 1.4 (22)	5.2 \pm 1.3 (25)	-	55 \pm 5 (8)	3.8 \pm 0.3 (8)
	Average	81*	2.6 \pm 0.8 (30)	6.0 \pm 1.6 (23)	5.3 \pm 1.1 (21)	-	57 \pm 9 (16)	3.8 \pm 0.4 (11)
June, 2012	Lichen	25	3.7 \pm 2.0 (53)	11.1 \pm 3.0 (27)	5.9 \pm 2.6 (44)	0.213 \pm 0.113 (53)	22 \pm 3 (12)	- **
	Moss	38	4.7 \pm 1.8 (39)	12.7 \pm 2.4 (19)	7.1 \pm 2.3 (32)	0.189 \pm 0.097 (51)	21 \pm 3 (16)	-
	Tussock	14	5.6 \pm 1.9 (33)	12.2 \pm 2.4 (19)	8.8 \pm 2.5 (29)	0.339 \pm 0.136 (40)	21 \pm 2 (11)	-
	Grass	4	5.2 \pm 2.1 (40)	10.4 \pm 3.0 (28)	6.4 \pm 2.1 (33)	0.304 \pm 0.149 (49)	21 \pm 2 (8)	-
	Average	81*	4.8 \pm 2.0 (42)	11.5 \pm 2.6 (23)	6.6 \pm 2.5 (38)	0.224 \pm 0.125 (56)	21 \pm 3 (14)	
July, 2012	Lichen	25	4.0 \pm 1.5 (38)	10.1 \pm 2.1 (21)	6.9 \pm 1.8 (26)	0.165 \pm 0.088 (53)	33 \pm 3 (9)	- **
	Moss	38	4.3 \pm 1.5 (35)	11.2 \pm 2.4 (22)	7.9 \pm 1.9 (25)	0.243 \pm 0.086 (60)	31 \pm 4 (13)	-
	Tussock	14	5.9 \pm 2.8 (48)	10.5 \pm 2.5 (23)	7.9 \pm 2.5 (31)	0.268 \pm 0.140 (52)	31 \pm 2 (8)	-
	Grass	4	5.6 \pm 1.9 (34)	9.9 \pm 1.1 (11)	6.6 \pm 1.0 (15)	0.208 \pm 0.088 (42)	36 \pm 6 (16)	-
	Average	81*	5.0 \pm 2.0 (40)	11.3 \pm 2.2 (19)	7.2 \pm 2.4 (33)	0.191 \pm 0.118 (62)	33 \pm 6 (18)	
August, 2012	Lichen	25	3.3 \pm 1.1 (33)	13.0 \pm 2.6 (20)	9.3 \pm 2.2 (23)	0.201 \pm 0.117 (58)	45 \pm 4 (10)	- **
	Moss	38	4.7 \pm 1.6 (35)	16.0 \pm 2.5 (15)	11.9 \pm 2.7 (22)	0.258 \pm 0.115 (73)	44 \pm 7 (15)	-
	Tussock	14	6.4 \pm 2.1 (33)	16.2 \pm 2.5 (15)	12.6 \pm 4.0 (32)	0.288 \pm 0.120 (42)	43 \pm 3 (7)	-
	Grass	4	5.5 \pm 2.4 (43)	13.2 \pm 0.8 (6)	9.3 \pm 1.2 (13)	0.199 \pm 0.069 (35)	47 \pm 11 (22)	-
	Average	81*	4.8 \pm 1.9 (40)	15.0 \pm 2.9 (19)	11.0 \pm 3.2 (29)	0.246 \pm 0.126 (51)	45 \pm 6 (13)	
September, 2012	Lichen	25	1.6 \pm 0.9 (54)	3.5 \pm 1.9 (55)	2.1 \pm 1.6 (75)	0.465 \pm 0.260 (56)	59 \pm 7 (11)	- **
	Moss	38	1.8 \pm 0.8 (44)	4.9 \pm 2.3 (47)	3.1 \pm 1.8 (59)	0.340 \pm 0.264 (78)	60 \pm 9 (16)	-
	Tussock	14	2.3 \pm 1.0 (44)	5.9 \pm 2.5 (42)	4.1 \pm 2.0 (48)	0.427 \pm 0.121 (28)	57 \pm 4 (7)	-
	Grass	4	2.2 \pm 0.9 (40)	2.9 \pm 2.5 (26)	2.0 \pm 1.6 (82)	0.456 \pm 0.378 (82)	64 \pm 9 (14)	-
	Average	81*	1.9 \pm 0.8 (42)	4.4 \pm 2.2 (50)	2.7 \pm 1.8 (65)	0.424 \pm 0.262 (62)	60 \pm 8 (13)	

* denotes total measured points.

** - is not conducted.

[#] n.m. indicates not measured.

Table 2. Q_{10} values and correlation coefficient between CO_2 efflux and soil temperature at 5 and 10 cm below the soil surface in lichen, moss, and tussock during the growing season based on a one-way ANOVA with a 95% confidence level

Vegetation, Year	Month	5 cm			10 cm		
		Q_{10}	R^2	p	Q_{10}	R^2	p
Lichen, 2011	June	2.05	0.10	<0.001	1.68	0.01	0.018
	August	8.58	0.36	<0.001	2.47	0.04	<0.001
	September	10.59	0.43	<0.001	6.87	0.32	<0.001
	Total	4.97	0.34	<0.001	1.06	0.01	0.032
Moss, 2011	June	1.58	0.26	<0.001	1.54	0.15	0.073
	August	6.59	0.40	<0.001	5.88	0.41	<0.001
	September	7.54	0.28	<0.001	10.10	0.78	<0.001
	Total	5.05	0.62	<0.002	4.46	0.21	<0.001
Tussock, 2011	June	2.68	0.54	0.890	2.01	0.33	0.005
	August	8.66	0.68	<0.001	11.70	0.66	0.041
	September	10.74	0.58	<0.001	9.64	0.44	0.008
	Total	6.15	0.73	0.018	5.44	0.39	0.467
Lichen, 2012	June	4.03	0.66	<0.001	1.40	0.24	<0.001
	July	5.04	0.69	<0.001	0.57	0.65	<0.001
	August	2.41	0.46	<0.001	2.50	0.35	<0.001
	September	6.17	0.57	<0.001	9.55	0.59	<0.001
	Total	2.86	0.65	<0.001	1.09	0.19	<0.001
Moss, 2012	June	2.62	0.37	<0.001	0.95	0.01	<0.001
	July	3.82	0.66	<0.001	3.51	0.51	<0.001
	August	1.15	0.01	<0.001	1.14	0.01	<0.001
	September	2.10	0.16	<0.001	2.18	0.11	<0.001
	Total	2.44	0.54	<0.001	2.35	0.33	<0.001
Tussock, 2012	June	5.06	0.77	<0.001	4.59	0.68	<0.001
	July	3.78	0.73	<0.001	2.78	0.50	<0.001
	August	2.98	0.77	<0.001	1.59	0.37	<0.001
	September	4.12	0.72	<0.001	5.01	0.59	<0.001
	Total	3.11	0.76	<0.001	3.00	0.62	<0.001
Grass, 2012	Total	2.28	0.41	<0.001	3.11	0.38	<0.001

Table 3. Summary of the posterior distribution for each parameter

Parameter	Mean	S.D.	2.5%	97.5%	Rhat
β_o	16.55	3.975	8.393	23.837	1.006
Q_{10}	2.517	0.115	2.291	2.752	1.009
a	-1.276	0.476	-1.966	-0.312	1.003
b	0.249	0.111	0.105	0.483	1.003
c	2.043	0.573	1.054	2.956	1.001
d	3.657	2.816	0.098	9.460	1.006
k	4.973	2.900	0.287	9.722	1.001
r	0.511	0.289	0.025	0.976	1.003
τ_{vege}	0.176	0.126	0.024	0.501	1.007
τ_{year}	0.072	0.059	0.006	0.226	1.006
τ^*	0.225	0.013	0.199	0.252	1.001

* τ indicates $1/\sigma$.

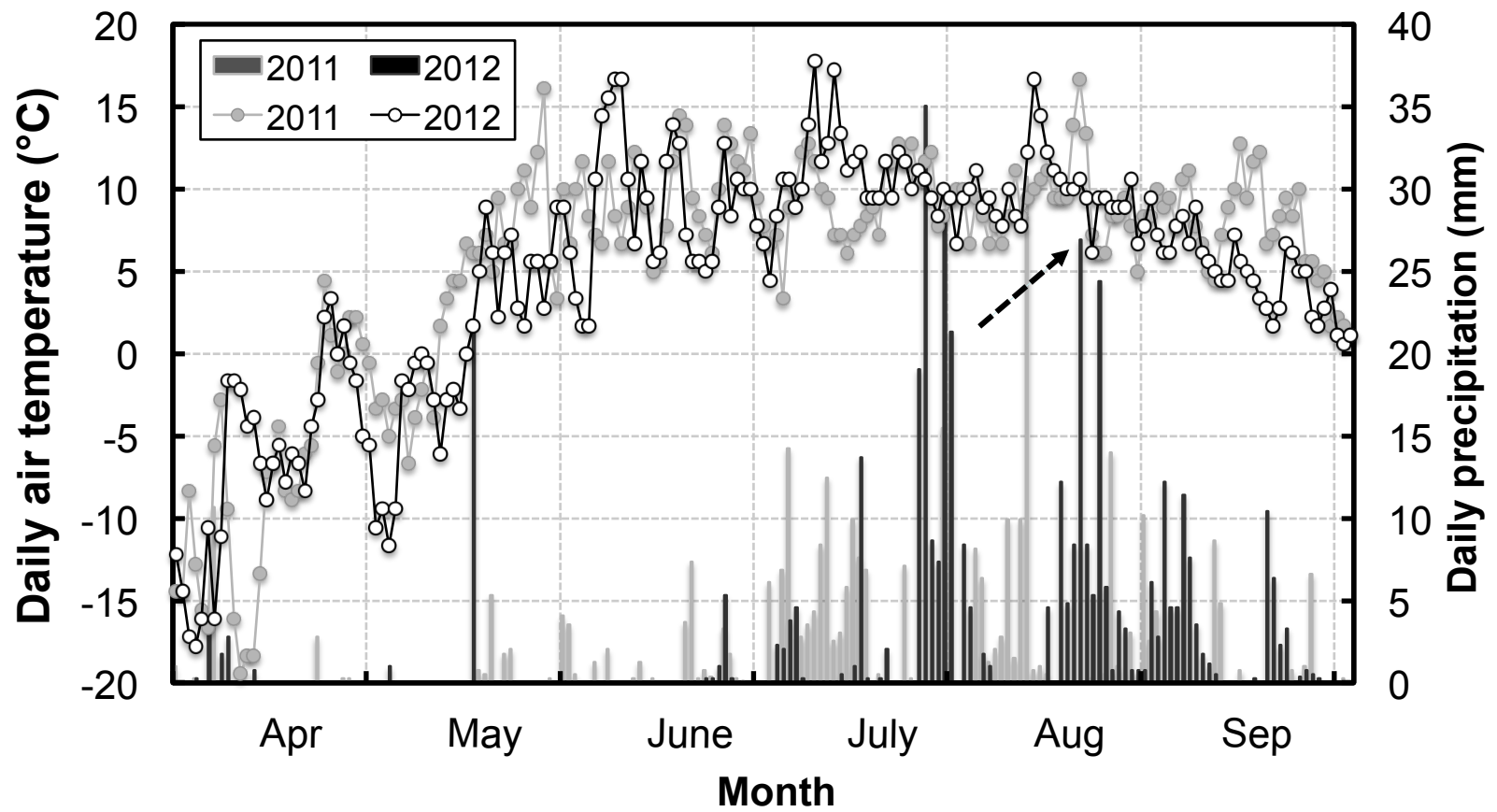


Figure 1

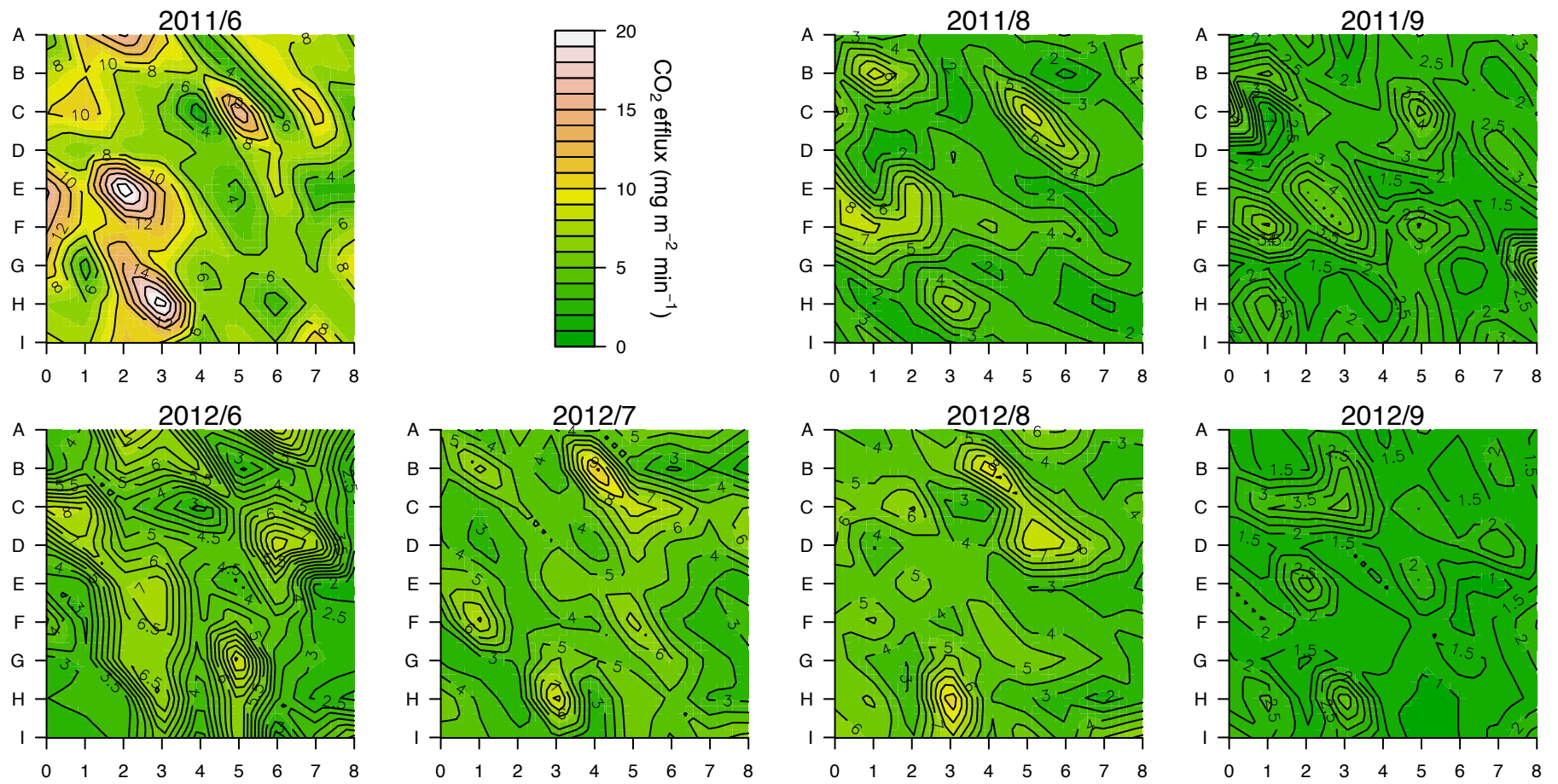


Figure 2

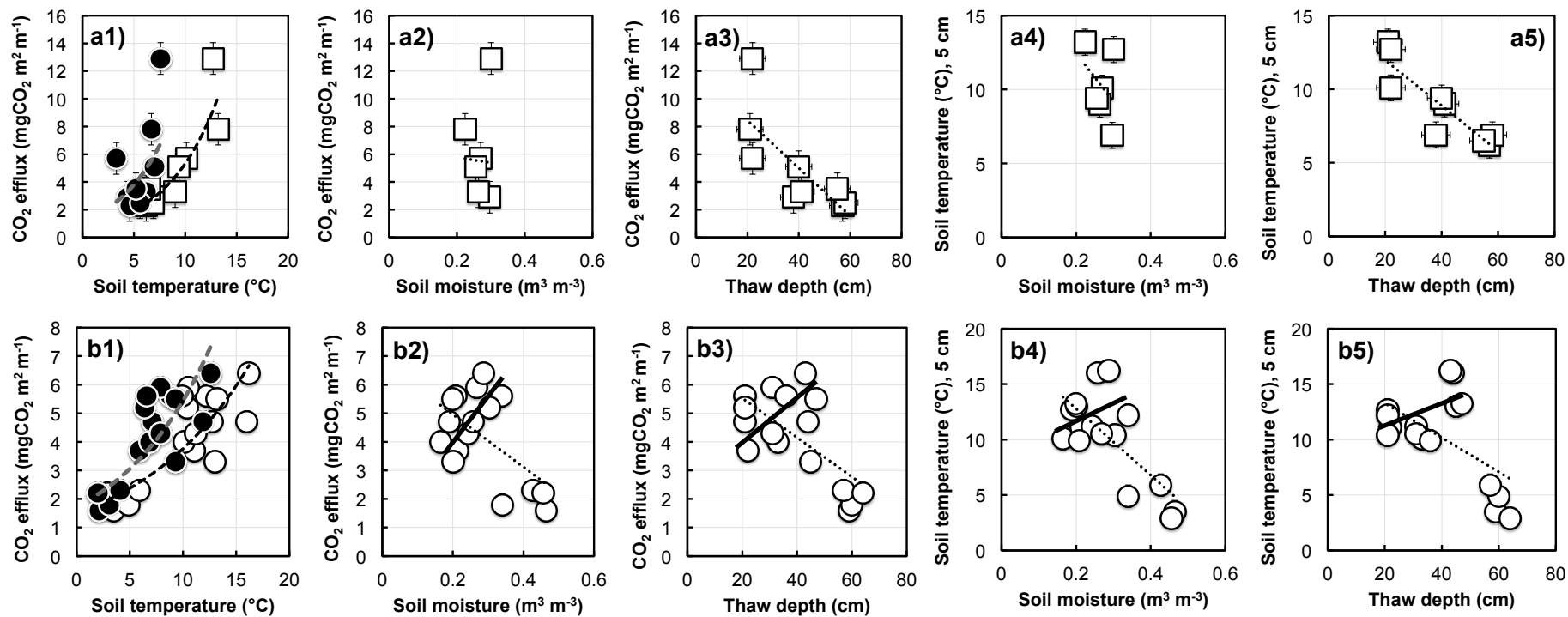


Figure 3

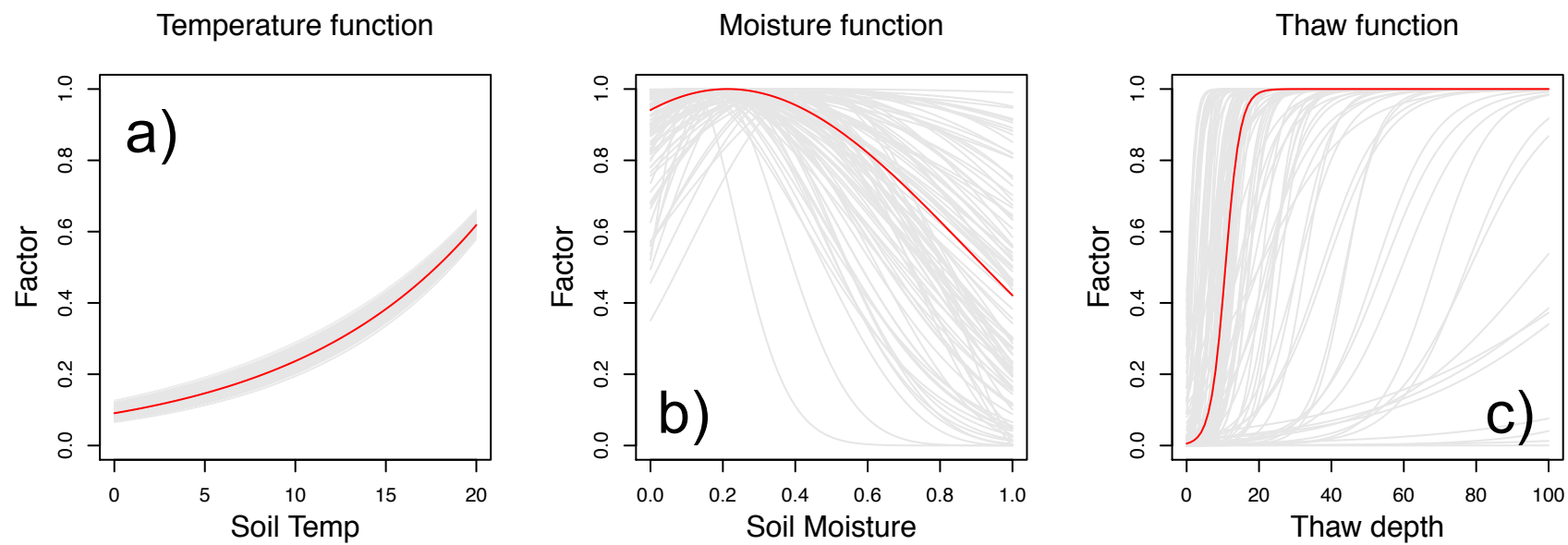


Figure 4

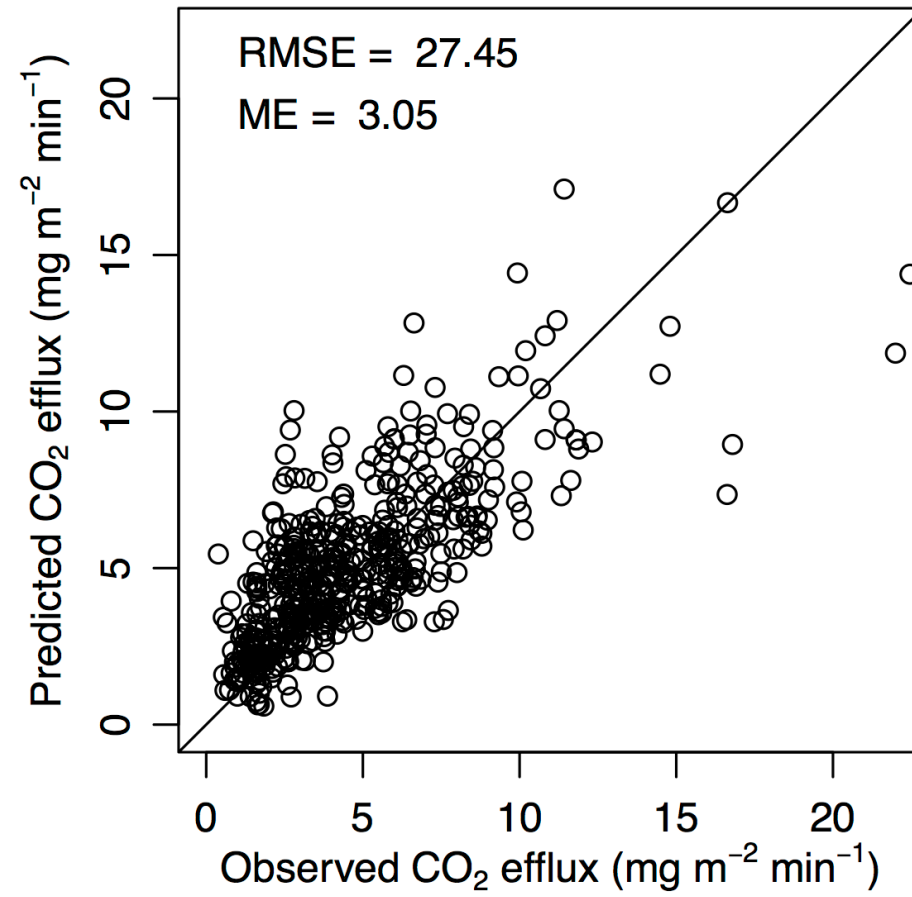


Figure 5

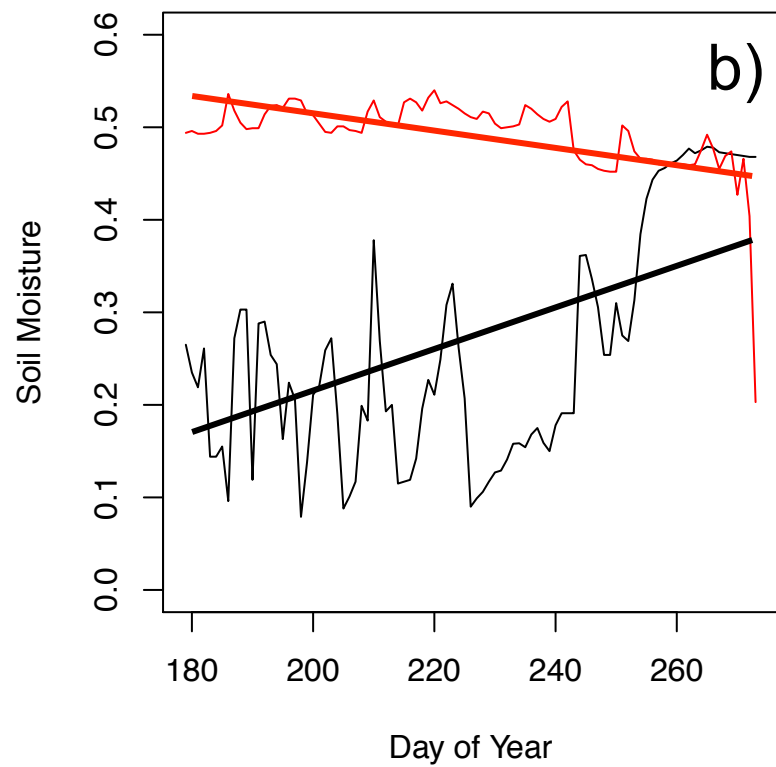
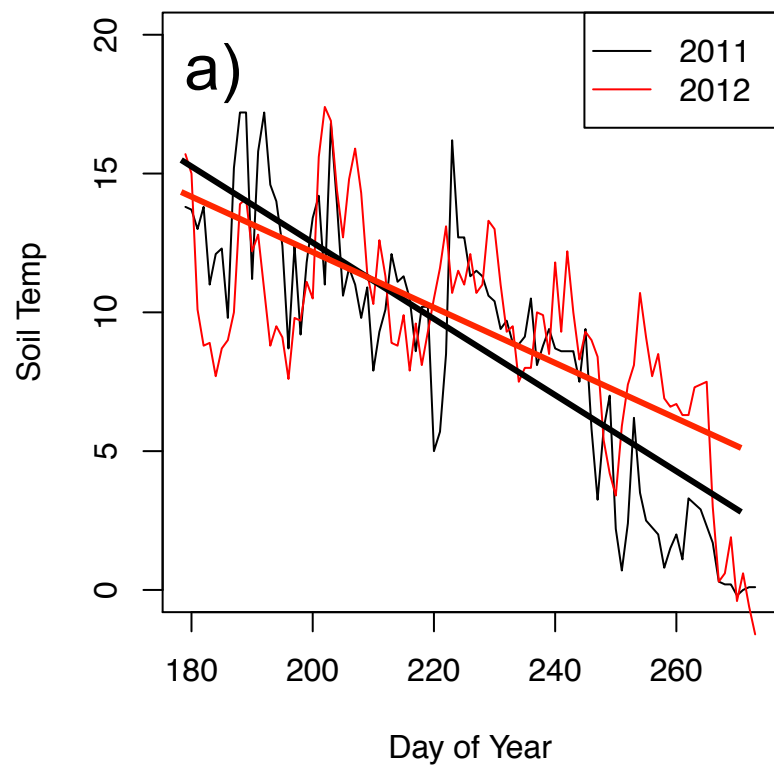


Figure 6