

1 **Export, biodegradation, and disinfection byproduct**
2 **formation of dissolved and particulate organic carbon in a**
3 **forested headwater stream during extreme rainfall events**

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15
16 **Abstract**

17 Despite an increasing recognition of the importance of extreme rainfall events for organic
18 carbon export to inland waters, little attention has been paid to the export and reactivity of
19 particulate organic carbon (POC) and dissolved organic C (DOC) in mountainous headwater
20 watersheds under monsoon climates. To investigate environmental implications of storm-
21 enhanced export of POC and DOC in mountainous headwater streams, we examined the
22 relationships between storm magnitude and C export from a forested watershed in the Haean
23 Basin, South Korea, during 13 storm events over four years and compared potentials of DOC
24 and POC for biodegradation and disinfection byproduct (DBP) formation during an extreme
25 rainfall event with a total rainfall of 209 mm. Event mean concentrations and export of POC
26 increased nonlinearly above thresholds of precipitation and discharge, far exceeding the
27 relatively small increases of DOC. The export of POC during a few storm events with a total
28 rainfall above 200 mm per event exceeded the annual organic C export during dry years.

1 During the 209 mm storm event, concentrations of total trihalomethanes formed by POC-
2 derived dissolved components changed synchronously with POC concentrations, exhibiting
3 lower levels than those formed by DOC. During a 30 day incubation at 25°C, DOC exported
4 during peak flow exhibited rapid biodegradation of labile components within seven days. In
5 contrast, the concentrations of DOC leached from POC gradually increased following the
6 initial decline. Gradual transformation of POC-derived dissolved materials resulted in greater
7 increases in the intensity of fulvic- and humic-like fluorescent components compared to the
8 DOC treatment. The results highlight the significance of extreme rainfall events as 'hot
9 moments' for POC export from mountainous watersheds and also suggest that storm pulses of
10 POC can provide potential sources of reactive organic components that can rapidly
11 biodegrade and form DBPs after being released into headwater streams.

12 **1 Introduction**

13 The transport of DOC and POC along streams and rivers represents a crucial linkage between
14 land and oceans in the global C cycle (Cole et al., 2007; Battin et al., 2008; Aufdenkampe et
15 al., 2011). Earlier estimates of the quantity of riverine organic C transport to the oceans range
16 from 0.19 to 0.33 Pg C, and have often been neglected in global C budgets because they
17 constitute only 1% to 2% of the net primary production of terrestrial ecosystems (Meybeck,
18 1982). Recent syntheses have provided higher estimates for the riverine C export, ranging
19 from 0.8 to 0.9 Pg C, and it has been suggested that the amount of C delivered to the oceans
20 might represent only a fraction of the total C received by inland waters (Battin et al., 2009;
21 Aufdenkampe et al., 2011). After receiving up to 2.7 Pg C from terrestrial sources, inland
22 waters do not transport it passively, but process it, and return substantial amounts of C to the
23 atmosphere either as CO₂ (Mayorga et al., 2005; Raymond et al., 2013) or as CH₄ (Bastviken
24 et al., 2011). Other studies have suggested a high C sequestration potential of inland water
25 systems including reservoirs, lakes, and floodplains (Stallard, 1998; Syvitski et al., 2005; Van
26 Oost et al., 2007; Goldsmith et al., 2008; Zehetner et al., 2009). Although studies conducted
27 in steep terrain such as the Himalayas have assumed mountainous rivers as a pipe transferring
28 organic C passively to the ocean and resulting in a preservation of terrigenous C in bottom
29 sediments along the coastal margins (Galy et al., 2007), little is known about the dynamics of
30 POC in mountainous headwater streams receiving storm pulses of POC.

31 Terrestrial mobilization and riverine processing of DOC have been studied over a wide
32 range of climate zones (del Giorgio and Davis, 2003), but less is known about the fate of POC

1 associated with suspended and deposited sediments in inland waters and floodplains. The
2 long-held assumption that POC is a minor component of aquatic organic C is based on
3 observations of large rivers, as exemplified by the high ratios of DOC to total organic carbon
4 (TOC) between 0.6 and 0.8 found in lowland rivers (Meybeck, 1982). However, unusually
5 high concentrations of POC that exceed DOC concentrations have been observed in low-order
6 streams and rivers, especially during storm events (Fisher and Likens, 1973; Bilby and Likens,
7 1979; Wallace et al., 1995; Coynel et al., 2005; Kim et al., 2010). Recent studies employing
8 intensive storm event sampling and high-frequency in situ instrumental measurements have
9 demonstrated that the pulsed export of POC during very short peak flow periods of intense or
10 extreme storms can account for a disproportionately large fraction of the annual C exported
11 from headwater watersheds (Jeong et al., 2012; Jung et al., 2012; Dhillon and Inamdar, 2013).

12 Studies of the biodegradation and transformation of POC have focused on the leaching
13 of DOM from different sources (Yoshimura et al., 2010) or the photochemically mediated
14 conversion of DOM from suspended sediments (Riggsbee et al., 2008). Although the bulk of
15 riverine POC at low flow has been considered aged and metabolically not reactive, it has been
16 suggested recently that not only labile components but also recalcitrant compounds such as
17 lignin and other phenolic compounds can be consumed rapidly by microorganisms in large
18 tropical rivers such as the Amazon (Ward et al., 2013). Studies of POC origin and lability in
19 some European estuaries have shown that the bulk of POC in estuaries might be refractory,
20 whereas seasonal changes in the input of organic matter of autochthonous or anthropogenic
21 origin can significantly increase the lability and biodegradability of POC even to a higher
22 degree than observed for DOC (Etcheber et al., 2007; Garnier et al., 2008). It remains
23 unexplored whether POC in headwater streams has comparable lability relative to DOC.

24 It was discovered in the 1970s that high levels of organic C in drinking water facilities
25 can lead to the formation of carcinogenic and mutagenic disinfection byproducts (DBPs)
26 (Rook, 1977). Much research has been conducted to characterize DOC components as DBP
27 precursors (e.g., Chow et al., 2007; Beggs et al., 2009), but little attention has been paid to the
28 reactive POC components responsible for DBP formation. Although storm pulses of
29 sediments have been suggested as a serious climate-induced risk to drinking water facilities
30 (Hurst et al., 2004; Park et al., 2010), rapid conversion of labile components from POC has
31 not yet been linked with DBP formation. The primary objective of this study was to evaluate
32 the relative importance of POC and DOC as a source of labile, reactive organic components in

1 headwater streams during monsoon rainfall events. First, we analyzed the relationship
2 between storm magnitude and organic C export using DOC and POC data collected from a
3 forested headwater stream in the Haean Basin, South Korea, during 13 storm events over four
4 years. We also compared the potentials of DOC and POC, released during a large storm event
5 with a total rainfall of 209 mm, for biodegradation and DBP formation, to redress the view of
6 low reactivity of POC in mountainous streams and rivers under monsoon climates.

7 **2 Materials and methods**

8 **2.1 Study site and storm event measurements**

9 The study was conducted at a forested watershed that represents steep headwater areas in the
10 Haean Basin, South Korea (38°15'–38°20' N; 128°05'–128°10' E; 400–1,304 m asl; Fig. 1).
11 The bedrock in the Haean Basin consists of highly weathered biotite granite at lower
12 elevations overlain by metamorphic rocks, which form the mountain ridges. The naturally
13 regrown, mixed deciduous forest on the steep (>20°) slopes is dominated by Mongolian oaks
14 (*Quercus mongolica*), Daimyo oaks (*Quercus dentata*), and Korean ashes (*Fraxinus*
15 *rhynchophylla*). Soils typical of the forested mountain slopes are dry to slightly moist brown
16 soils (acid Cambisols, according to the FAO World Reference Base for Soil Resources),
17 overlain by moder-like forest floors with a distinct Oi horizon and less distinct Oe/Oa
18 horizons. The annual mean temperature at the Haean Basin is around 9°C. The mean summer
19 (June to August) rainfall over the period from 1997 to 2010 was 939 mm, which accounted
20 for 61% of the mean annual precipitation of 1539 mm (Jeong, 2011).

21 Intensive storm event water samplings during the summer monsoon period were
22 conducted upstream of the V-notch weir located at the outlet of the stream draining the study
23 watershed during 13 storm events over four years from 2008 through 2011 (Fig. 1; refer to
24 Jeong et al., 2012 and Jung et al., 2012 for more details about sampling methods). During the
25 storm events, streamwater samples were usually collected every two hours using an
26 autosampler (6712 Portable Sampler, ISCO).

27 Water samples were first filtered through a 2 mm mesh sieve to remove large particles
28 and debris and then through a [pre-combusted GF/F filter \(Whatman; nominal pore size: 0.7](#)
29 [µm\)](#). The filters were then dried at 65 °C to a constant weight and reweighed to calculate the
30 total suspended solids (TSS). The dried filters were fumed with HCl in a sealed desiccator for
31 24 h to remove inorganic C prior to the analysis of POC. The concentrations of POC in the

1 acid-treated filters were measured with an elemental analyzer (Vario MAX CN, Elementar,
2 Germany). The DOC concentrations in the filtered water samples were measured with a TOC
3 analyzer using high-temperature combustion of organic matter (OM) followed by thermal
4 detection of CO₂ (TOC 5000a or TOC V_{CPH}, Shimadzu, Japan). For each batch of ten samples
5 measured for POC and DOC, we carried out continuous concentration verification with a
6 check standard and a baseline contamination check using both the field and the laboratory
7 blanks. As an additional measure for quality control, replicate analysis was conducted for
8 approximately 10% of all the measured samples. The relative standard deviations for the
9 repeated measurements of check standards were within 5%.

10 Discharge-weighted mean concentrations and exports of DOC and POC during each of
11 storm events were calculated using measured concentrations of DOC and POC and discharge
12 data that were either estimated using the HBV-Lite hydrologic model for the period from
13 2008 through July 2009 or measured with a V-notch weir thereafter (Jung et al., 2012). Storm
14 event periods were determined using a web-based hydrograph analysis tool (WHAT;
15 <http://cobweb.ecn.purdue.edu/~what/>; Lim et al., 2005). The relationships between the
16 measures of storm magnitude (event total precipitation or discharge) and the concentrations or
17 fluxes of DOC and POC were examined by establishing best-fit regressions using data
18 analysis software (SigmaPlot, version 10.0). Hourly precipitation data were obtained from an
19 automatic weather station located within the Haean Basin, operated by the Korea
20 Meteorological Administration.

21 **2.2 Laboratory incubation experiments**

22 Two laboratory incubation experiments were conducted with DOM and suspended sediment
23 (SS) samples collected during a large storm event from 22–27 June 2011 (cumulative
24 precipitation: 209 mm). According to widely used extreme precipitation indices such as upper
25 first to fifth percentiles of one day or five day cumulative precipitation, the monitored storm
26 event can be regarded as a “very wet” or “extremely wet” event (Choi et al., 2009). The first
27 experiment was performed using water samples collected at five different times under
28 different flow conditions during the first phase of the storm event to measure the DBP
29 formation potentials of DOM and SS-derived DOM (hereafter referred to as “SS-DOM”).
30 Streamwater samples were filtered through cellulose acetate membrane filters (Whatman;
31 nominal pore size: 0.45 μm) to separate particulates from DOM. We used membrane filters
32 instead of GF/F filters despite the slight difference in pore size, because the latter did not

1 allow us to scrape off particles from glass fiber. Filtered streamwater samples from five
2 sampling points were measured for DOC concentrations. SS samples collected on the filters
3 were transferred to ultrapure water (Milli-Q, Millipore) and then extracted for the dissolved
4 fraction on a shaker for 1 h. The SS extracts were again filtered through cellulose acetate
5 membrane filters and DOC concentrations were then measured. We used blank filters without
6 SS to check any contamination from filter materials.

7 Three replicates (150 mL each) of DOM and SS-DOM samples were treated with
8 NaOCl (final concentration after addition $\approx 4 \text{ mg L}^{-1}$) in the dark at 20°C for 24 h. Upon
9 completion of incubation, samples were quenched with ascorbic acid (0.04 g per sample). Six
10 volatile DBPs – chloroform (CF), bromodichloromethane (BDCM), dibromochloromethane
11 (DBCM), bromoform (BF), dichloroacetonitrile (DCAN), and 1,1-dichloro-2-propanone
12 (DCP) – were extracted according to EPA Method 501.2, and the extracts were analyzed by
13 gas chromatography-mass spectrometry (7890A GC/5975C MS; Agilent Technologies, Inc.,
14 Santa Clara, USA) in the selected ion monitoring mode. Ions for quantitative determination of
15 six DBPs were as follows – CF and BDCM: $m/z = 83, 85$; DBCM: 127, 129; BF: 173, 254;
16 DCAN: 74, 82; DCP: 43. Estimated method detection limits (MDL) were between 0.04 and
17 $0.1 \mu\text{g L}^{-1}$. As the concentrations of DBCM, BF, DCAN, and DCP were below the MDLs for
18 most samples, the combined concentrations of CF and BDCM are reported here as total
19 trihalomethanes (TTHMs).

20 Short-term changes in biodegradable DOC (BDOC) and optical properties of DOM
21 and SS-DOM sampled from the second peak flow period of the same storm event in June
22 2011 were examined in a 30 day incubation experiment. Water samples were filtered through
23 a $0.2 \mu\text{m}$ membrane filter (cellulose acetate membrane filter, Whatman) to remove
24 particulates and most of the microbial organisms, and then incubated in triplicates (400 mL
25 each) with an inoculum (1% of the final volume) under controlled laboratory conditions at
26 25°C in the dark. The inoculum was prepared using a method modified from Fellman et al.
27 (2009). Sediment samples collected from the streambank and bed were leached with an
28 unfiltered streamwater sample from the same site. The leachates were filtered through a $2 \mu\text{m}$
29 filter (Nuclepore polycarbonate filter, Whatman) to remove coarse particles $> 2 \mu\text{m}$. The
30 filtrates containing microbial cells ($< 2 \mu\text{m}$) were then incubated at 25°C for 7 days before
31 addition to the incubation samples. BDOC and optical properties were compared among three
32 treatments: (1) DOM that already existed at the time of sampling (“DOM” treatment), (2)

1 DOM added with freeze-dried SS samples collected at the same time as the DOM samples
2 (“DOM+SS” treatment), and (3) ultrapure water added with freeze-dried SS samples (“SS”
3 treatment). The three treatments were designed to evaluate biodegradation potentials of DOM
4 and SS-associated POM components in isolation (“DOM” and “SS” treatments) and
5 combined (“DOM+SS” treatment). The same inoculum (1% of the final volume) was added to
6 all samples. One hour after the addition of SS, a small portion of each sample was sampled
7 and filtered through a 0.2 µm membrane filter to measure the initial concentrations of DOC
8 leached from the SS. Subsamples were collected from the incubation bottles and analyzed for
9 DOC concentrations, UV absorbance at 254 nm (UVA₂₅₄), and fluorescence excitation
10 emission matrices (EEMs) on days 1, 3, 6, 13, 21, and 30 of the incubation period.

11 Fluorescence EEMs of the filtered water samples were obtained by simultaneous
12 scanning over excitation wavelengths from 200 to 400 nm at 5 nm steps and emission
13 wavelengths from 290 to 540 nm at 1 nm steps using a fluorescence spectrophotometer
14 (F7000, Hitachi, Tokyo, Japan). Scan speed was 2400 nm min⁻¹ and the bandwidth was set to
15 5 nm for both excitation and emission. A 290 nm cutoff filter was used for all the
16 measurements to minimize second order Rayleigh scattering. Blank runs were conducted with
17 Ultrapure water for every batch of 10 samples. The Raman peak of water at excitation 350
18 nm/emission 400 nm was used to assess the stability of the fluorescence spectrometer. To
19 account partially for the Rayleigh scattering, the machine response to blank runs was
20 subtracted from the EEMs of the samples. The inner-filter effect was corrected based on
21 McKnight et al. (2001). To deconvolute the major fluorescent components, PARAFAC
22 modeling was performed using MATLAB 7.0 (Mathworks, Natick, USA) with the DOM
23 Fluor toolbox (<http://www.models.life.ku.dk>) based on a method developed by Stedmon and
24 Bro (2008). Three PARAFAC components were identified and termed humic-, fulvic-, and
25 protein-like fluorescent components based on the similarity of the peak wavelengths to values
26 reported in the literature (Fellman et al., 2010).

27 **3 Results**

28 **3.1 The relationships between storm magnitude and C export**

29 Significant relationships were found between measures of storm magnitude (total event
30 rainfall and discharge) and discharge-weighted mean concentrations or exports of organic C
31 in the forest stream during 13 storm events (Fig. 2). The event mean concentrations of DOC
32 tended to increase linearly with increasing precipitation and discharge, but the relationships

1 were not statistically significant ($P = 0.15$). In contrast, event mean concentrations of POC
2 and exports of both DOC and POC had significant nonlinear relationships with event total
3 rainfall or discharge. Cubic polynomial equations, which represented best-fit regressions for
4 all these nonlinear relationships, better accounted for the variations of the observed nonlinear
5 relationships than power law functions or second-order polynomial functions that had been
6 used to describe discharge-DOC relationships in previous studies (e.g., Raymond and Saiers,
7 2010).

8 Across a range of small and moderate storm events with up to 100 mm of precipitation
9 per event, the concentrations and fluxes of DOC and POC increased gradually with increasing
10 precipitation or discharge, with higher values of DOC than POC for most events. However,
11 above the threshold precipitation greater than 100 mm per event, POC concentrations and
12 fluxes increased drastically, considerably exceeding the levels of DOC. Compared to the very
13 strong relationships with rainfall, POC concentrations and fluxes exhibited weaker
14 relationships with discharge. Below-average concentrations and fluxes of POC during a few
15 storm events with substantial amount of antecedent precipitation (e.g., event 4; Table 1)
16 resulted in large variations in POC concentration and export at similar ranges of discharge
17 (Fig. 2).

18 **3.2 DBP formation of DOC and POC exported during extreme storm event**

19 We examined short-term dynamics of DOC and POC export and their implications for C
20 biodegradation and DBP formation in the receiving headwater stream for two extreme storm
21 events that were part of 13 events analyzed in Fig. 2 (Fig. 3). Responses of discharge and
22 stream water chemistry differed between the two storm events. The second phase (total
23 rainfall: 120 mm; mean hourly rainfall: 2.1 mm h^{-1}) of event 12, which started on 25 June
24 2011, exhibited greater rainfall amount and higher intensity than the preceding phase (total
25 rainfall: 89 mm; mean hourly rainfall: 1.5 mm h^{-1}) that started on 22 June following a
26 relatively dry pre-monsoon period in June. As the cumulative rainfall approached 200 mm,
27 discharge and turbidity ($>1000 \text{ NTU}$) increased drastically toward the peak flow of the second
28 phase (Fig. 3). While DOC concentrations reached similar maximum levels ($2 - 3 \text{ mg C L}^{-1}$)
29 during all peak discharge periods, POC concentrations, along with TSS concentrations and
30 turbidity, exhibited large increases during the short peak flow periods of the intense rainfalls.

31 During the first phase of event 12, the concentrations of TTHMs formed by DOM and
32 SS-DOM changed in parallel with varying concentrations of DOC or POC, respectively (Fig.

1 4). The positive relationship between the concentrations of TTHMs and DOC ($r^2 = 0.894$; $P =$
2 0.015) or POC ($r^2 = 0.739$; $P = 0.062$) in the water samples collected over the course of the
3 first storm phase pointed to the importance of the amount of organic matter exposed to
4 chlorination for accounting for temporal variations in TTHMs formation. TTHMs formation
5 potentials differed substantially between DOM ($14.2\text{--}26.7 \mu\text{g L}^{-1}$) and SS-DOM ($1.5\text{--}7.5 \mu\text{g}$
6 L^{-1}).

7 **3.3 Biodegradation of DOC and POC exported during extreme storm event**

8 DOM and SS-DOM from the water samples collected during the second storm peak of event
9 12 (sampling time indicated by “BDOC sample” in Figs. 3, 4) exhibited relatively small
10 changes in DOC concentrations over the 30 day incubation (Fig. 5). DOC concentrations
11 decreased in the “DOM” by 7.1% and in the “DOM+SS” by 6.6%, but increased in the “SS”
12 treatment by 12.9% relative to the initial DOC concentrations (Fig. 5; Table 2). BDOC
13 concentrations in the “DOM” and “DOM+SS” treatments were similar to each other and also
14 over 7 and 30 days (Table 2). Although the absolute magnitude of BDOC was much smaller
15 in the “SS” treatment than in the other treatments, BDOC percentage values relative to the
16 initial DOC concentrations were comparable to those for other treatments during the first
17 half of the incubation and exhibited a shift from the initial rapid decreases (decrease of
18 10.6% over 7 days) to the gradual increases during the following period, resulting in an
19 overall increase of 12.9% over 30 days (Table 2).

20 Continuous transformations of OM between the dissolved and particulate phase were
21 also indicated by the “SS” samples exhibiting significantly greater changes in both SUVA_{254}
22 and specific fluorescence intensities of three PARAFAC components per unit mass of DOC
23 compared with the “DOM” samples (Fig. 5; Table 2). SUVA_{254} in the SS-DOM was initially
24 lower than in the other treatments, but exceeded the levels of the other treatments during the
25 second half of the incubation, whereas the specific fluorescence intensities of three
26 PARAFAC components were highest in the “SS” samples throughout most of the incubation
27 period (Fig. 5). For all treatments in general and the treatments containing SS in particular,
28 the intensities of SUVA_{254} and the humic- and fulvic-like fluorescence increased continuously
29 with time, whereas the intensities of protein-like fluorescence continued to decrease during
30 the course of the incubation, resulting in striking differences in the intensity over the entire
31 incubation between the protein-like fluorescence and other optical measurements (Table 2).

1 **4 Discussion**

2 **4.1 Importance of extreme rainfall events for DOC and POC export**

3 Comparison of DOC and POC exports during 13 monsoon rainfall events of various
4 magnitude and duration well illustrates differential storm responses of DOC and POC (Figs. 2,
5 3). Although both DOC and POC exports increased with increasing rainfall and discharge,
6 much stronger responses of POC to intense rainfalls suggested the disproportionate
7 importance of a few extreme rainfall events as a ‘shortcut’ of organic C export to
8 mountainous headwater streams. During the one year period (17 July 2009 – 16 July 2010) for
9 which the annual C export from the same study watershed was estimated using continuous in
10 situ measurements, one short extreme storm event with a cumulative rainfall of 210 mm
11 comprised a disproportionately large proportion of the annual export of DOC (23% of 3.2 kg
12 C ha⁻¹ yr⁻¹) and POC (62% of 3.7 kg C ha⁻¹ yr⁻¹) (Jeong et al., 2012). In the current synthesis
13 of 13 storm events (Fig. 2), the export of DOC and POC during each of the four storm events
14 with cumulative precipitation exceeding 200 mm (209 – 292 mm) ranged from 0.99 – 4.91 kg
15 C ha⁻¹ and 3.15 – 15.55 kg C ha⁻¹, respectively. In particular, the large variations of POC
16 export across these storm events suggest that a few extreme events can account for a
17 substantial portion or even exceed the annual C export observed for relatively dry years such
18 as the year reported by Jeong et al. (2012).

19 The importance of extreme events for DOC export has been investigated recently by
20 synthesizing regional monitoring data sets (Raymond and Saiers, 2010) or by high-frequency
21 sampling during extreme hydrologic events such as hurricanes (Yoon and Raymond, 2012). In
22 a meta-data analysis using DOC monitoring data covering 30 small eastern US forested
23 watersheds, Raymond and Saiers (2010) found a nonlinear relationship between the annual
24 export of DOC and annual stream discharge, which was described by a second-order
25 polynomial function similar to the relationships found in this study. A high-frequency storm
26 event sampling in a forested watershed in Esopus Creek, New York State during Hurricane
27 Irene (total precipitation of 293 mm for 27–29 August, 2011) showed that the export of DOC
28 during this 200-year event accounted for 43% of the average annual DOC export (Yoon and
29 Raymond, 2012).

30 Stream exports of DOC and POC have been compared in diverse types of watersheds
31 under different rainfall regimes (Fisher and Likens, 1973; Bilby and Likens, 1979; Wallace et
32 al., 1995; Coynel et al., 2005). However, only a few studies have provided high-frequency

1 time series data of POC that also encompass extreme storm events (Kim et al., 2010, Jeong et
2 al., 2012; Dhillon and Inamdar, 2013). It has been suggested that large storm-induced exports
3 of POC are characteristic of erosion-prone mountainous watersheds, particularly in response
4 to intense storm events (Jung et al., 2012). However, large storm pulses of POC can also
5 occur in forested watersheds located on relatively gentle slopes, when extreme hydrologic
6 events such as tropical storms increase the vulnerability of streambanks and soils on steep
7 slopes to soil erosion (Dhillon and Inamdar, 2013). [Dhillon and Inamdar \(2013\) observed](#)
8 [large disparities between DOC and POC exported from a forest watershed in Maryland during](#)
9 [extreme rainfall events accompanying Hurricane Irene. Our results of large storm pulses of](#)
10 [POC exceeding the rather moderate increases of DOC \(Figs. 2, 3\) also suggest that the stream](#)
11 [C export regime would shift toward a greater dominance of POC over DOC, if watersheds](#)
12 [were subjected to larger and more extreme storm events.](#)

13 Relatively small variations in the peak DOC concentrations between the two periods of
14 event 12 (Fig. 3), together with small increases of DOC concentrations and fluxes with
15 increasing rainfall and discharge (Fig. 2), might imply a limited range of DOC supply from
16 the major sources in the upper soil horizons (Hornberger et al., 1994; Inamdar et al., 2004). In
17 contrast, POC export might have increased in a nonlinear relationship with increasing
18 cumulative precipitation after passing a threshold precipitation during short peak flow periods
19 of intense storm events. Nonlinear increases in POC export above a threshold precipitation
20 have been ascribed to a certain level of energy required to initiate soil erosion at sources such
21 as streambed, streambank, and bare soil surfaces on steep slopes (Jeong et al., 2012; Dhillon
22 and Inamdar, 2013). POC concentrations and export also exhibited threshold relationships
23 with discharge (Fig. 2). Larger variations of POC concentrations and export at similar levels
24 of discharge suggest that discharge-POC relationships depend on antecedent rainfalls more
25 strongly than rainfall-POC relationships. Reduced export of DOC and POC during falling
26 hydrograph events have been ascribed to temporary depletion of terrestrial supply of source
27 materials or shifts in the timing of runoff contributions from major source areas (Raymond
28 and Saiers, 2010; Jeong et al., 2012).

29 **4.2 Storm-enhanced DBP formation potentials of DOC and POC**

30 Although potentials of SS-DOM for THM formation were measured for only a small number
31 of samples due to difficulties associated with collecting large-volume samples under harsh
32 storm sampling conditions, simultaneous measurements of DOC and POC for THM formation

1 potentials provide rare empirical evidence for the rapid transformation of potentially soluble
2 components of particulate organic matter (POM) into DOM moieties that can actively
3 participate in chlorination reactions. The positive relationship between POC concentrations
4 and TTHMs formation potentials during the first phase of event 12 suggests that large
5 increases in POC concentrations during intense storm events, as observed during the second
6 phase of the same storm event and the following event 13 (Fig. 3), could result in much
7 higher levels of THMs than observed during the relatively moderate initial storm phase. DBP
8 formation potentials have been correlated with the concentrations of DOC as the primary
9 precursor of DBPs (Chow et al., 2007; Kraus et al., 2008; Nguyen et al., 2013). However,
10 there has been no attempt to measure potentials of DBP formation in unfiltered streamwater
11 that might contain labile components leached from SS.

12 Storm-induced increases in DOC input to drinking water source waters have been
13 suggested as a potential threat to the performance of treatment processes in drinking water
14 facilities (Hurst et al., 2004). Although a few studies have linked the storm pulses of DOC
15 with increasing potentials of DBP formation (Chow et al., 2007; Kraus et al., 2008; Nguyen et
16 al., 2013), there have been rare systematic approaches to examine the chemical reactivity and
17 biological transformation of both DOM and POM exported under storm flow conditions. Our
18 findings of rapid dissolution of soluble organic components from SS and the subsequent
19 responses to biological degradation and chlorination emphasize the hitherto neglected
20 reactivity of POM and its implications for both aquatic C dynamics and public health
21 concerns in the drinking water sector. Pulsed export of POC during intense storm events, as
22 shown in Fig. 3, might contain a higher proportion of old humic materials that are usually
23 enriched in ^{13}C than found at base or low flow (Jung et al., 2012). Humic materials in excess
24 of the usual coagulant dose can decrease drastically the removal of particulate turbidity due to
25 preferential formation of complexes between humic complexation sites and coagulant metals
26 (Hurst et al., 2004). Based on the strong positive correlations between DPB formation
27 potentials and optical measurements of DOM (UVA₂₅₄ and fluorescence EEM peaks related
28 to humic substances) in a forested watershed in Korea during two storm events, Nguyen et al.
29 (2013) argued that humic-like DOM components with more aromatic and condensed
30 structures have a higher potential to form DBPs upon chlorination than other DOM moieties.

1 **4.3 Storm pulses of DOC and POC as sources of labile C in headwater stream**

2 Unlike previous studies that focused on microbial transformation of labile DOM components,
3 our results provide rare insights into microbial alterations of the optical properties of particle-
4 derived DOM. Previous BDOC experiments that focused on DOM did not consider the rapid
5 conversion of OM between the dissolved and particulate phase in streams and rivers. The
6 exclusion of SS can lead to a significant underestimation of DOM biodegradation and
7 changes in optical properties, considering the role of SS surfaces as biofilm that can enhance
8 the attachment and growth of microbes (Garnier et al., 2008). In our study, the actual rate of
9 biodegradation of POC is not known, because we did not measure changes in POC
10 concentrations or CO₂ evolution as a measure of POC biodegradation. Given the unusually
11 high concentrations of POC during peak flow periods of intense or extreme storm events (Figs.
12 2, 3) and the fact that storm pulses of POC can contain a large pool of potentially soluble or
13 biodegradable OM, the actual biodegradation of POC might be much higher than estimated
14 from the small BDOC values observed for the SS samples.

15 Only a few studies have directly compared the biodegradation of DOM and POM (del
16 Giorgio and Pace, 2008; Garnier et al., 2008). In the turbidity maximum zone of the Seine
17 Estuary, France, Garnier et al. (2008) found much higher rates of biodegradation for POC (the
18 change in POC concentration between the start and end of 45 day incubation) than for DOC.
19 In a study comparing BDOC and bacterial respiration of DOC along the Hudson River, del
20 Giorgio and Pace (2008) also found a significant contribution of POC to the riverine bacterial
21 respiration of DOC and suggested that POC might affect bacteria not through direct
22 consumption of POC but by inputs of labile components to the DOC pool. In a 14 day
23 incubation experiment with water-soluble OM extracted from leaf litter, Hur et al. (2009)
24 found that microbial transformations of labile OM increased SUVA₂₅₄ and fulvic- and humic-
25 like fluorescence, but decreased protein-like fluorescence intensities. As suggested by Hur et
26 al. (2009), soluble components of POM that have simple structures can be degraded
27 preferentially by decomposers and hence enhance microbial activity, increasing the
28 production of dissolved humic materials as byproducts.

29 The BDOC percentage values observed in this study fall within the low range of the
30 BDOC values reported for various forested headwater streams (Servais et al., 1987: 11%;
31 Fellman et al., 2009: 20 – 23%; Kang and Mitchell, 2013: 6 – 18%). The almost overlapping
32 temporal patterns of DOC concentrations, observed in both the “DOM” and “DOM+SS”

1 treatments, might imply an absence of the influence of SS on DOM biodegradation. However,
2 the continuous increases in DOC concentration following the initial decreases over the first
3 seven days in the “SS” treatment (Fig. 5), along with increasingly higher humic- and fulvic-
4 like fluorescence of “DOM+SS” samples as compared with the “DOM” values (Fig. 5; Table
5 2), suggest that the DOC pool, reduced by the biodegradation of labile organic components in
6 the “DOM+SS” treatment, might have been replenished concurrently by the continuous
7 supply of DOC from potentially soluble OM pools in the SS, resulting in no apparent
8 differences in DOC concentrations with slight modifications of DOM optical properties. The
9 increasing DOC concentrations following the initial decreases in the “SS” treatment can be
10 explained by the shifting balance between the consumption and production of DOC towards
11 an increased dissolution of relatively recalcitrant organic components during the latter phase
12 compared with the initial, rapid biodegradation of labile OM dissolved from the “fresh” SS.

13 **5 Conclusions**

14 Repeated high-frequency storm event samplings at two-hour intervals allowed for a
15 systematic comparison of DOC and POC export in the mountainous headwater stream during
16 13 storm events over four years. The results show that POC export increases nonlinearly
17 above the thresholds of precipitation and discharge, significantly exceeding the relatively
18 small increases of DOC. Very large magnitudes and variations of the storm-induced export of
19 organic C in general, and POC in particular, imply that the export of POC during a few
20 extreme storm events can comprise a significant portion or the bulk of the annual soil C loss
21 and that storm pulses of POC can become increasingly important as a pathway of soil C loss
22 in mountainous watersheds under monsoon climates, in response to more frequent occurrence
23 of extreme rainfall events, as predicted for a large part of the Northern Hemisphere including
24 the study region as a consequence of global climate change (Choi et al., 2008; Min et al.,
25 2011). Results from a single study site cannot be extrapolated to estimate C mobilization from
26 terrestrial sources at the regional or global scale. However, this study provides rare high-
27 frequency monitoring data to reinforce the growing recognition that the export of terrigenous
28 organic C might be much greater than currently estimated, if we take into consideration the
29 “missing C” that cannot be captured with low-frequency sampling schemes.

30 This study also demonstrates that storm pulses of POC contain labile moieties that can
31 dissolve rapidly in streamwater and become subjected to microbial transformations, as
32 revealed by the greater changes in $SUVA_{254}$ and specific fluorescence intensities of SS-DOM

1 compared with DOM. While current debates on erosion-enhanced sinks of atmospheric CO₂
2 have focused on reduced decomposition of eroded C under depositional settings (Van Oost et
3 al., 2007; Lal and Pimentel, 2008), recent findings of high biodegradation potentials of POC
4 in large temperate and tropical rivers (Etcheber et al., 2007; del Giorgio and Pace, 2008; Ward
5 et al., 2013) raise an important question as to whether eroded C degrades faster after entering
6 streams and rivers than in source soils. Our measurements of BDOC and optical properties of
7 SS-DOM in the headwater stream suggest that rapid conversion of labile OM between the
8 DOC and POC pools starts to occur even in upstream source areas of mountainous river
9 systems in monsoon Asia, which have been suggested to transport organic C without
10 significant chemical transformations (Galy et al., 2007; Goldsmith et al., 2008). Future study
11 of C transformations and storage in inland waters should incorporate POC transformations to
12 improve predictions of C transport through inland waters under changing rainfall regimes.
13 The new finding of increasing potentials of THM formation with rising POC concentrations
14 during intense storms also underscores the importance of monitoring POC as well as DOC for
15 watershed-level climate change adaptation plans, which might involve monitoring drinking
16 water sources in response to the increasing frequency and intensity of storm events across
17 many parts of the world.

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24

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8

1 Table 1. Characteristics of 13 monsoon rainfall events. Original data of DOC and POC
 2 concentrations from eight events that were published elsewhere (references provided as notes)
 3 are here used to calculate discharge-weighted event mean concentrations and fluxes in Fig. 2.
 4 The events 12 and 13 were examined in detail to evaluate potentials of DOC and POC for
 5 DBP formation and biodegradation (Figs. 3 – 5).

6

| Event | Start time | Duration (h) | Total rainfall (mm) | Mean rainfall intensity (mm h ⁻¹) | 7-d antecedent rainfall (mm) | Note |
|-------|------------------|-----------------|---------------------------|--------------------------------------------------------|---------------------------------------|-------------------|
| 1 | 18-06-2008 03:00 | 19 | 72 | 3.8 | 0 | Jung et al. 2012 |
| 2 | 16-07-2008 13:00 | 6 | 17 | 2.8 | 40 | Jung et al. 2012 |
| 3 | 24-07-2008 02:00 | 33 | 292 | 8.9 | 86 | Jung et al. 2012 |
| 4 | 26-07-2008 04:00 | 26 | 137 | 5.3 | 378 | Jung et al. 2012 |
| 5 | 20-06-2009 06:00 | 11 | 15 | 1.4 | 55 | |
| 6 | 09-07-2009 05:00 | 14 | 167 | 11.9 | 14 | |
| 7 | 18-07-2009 00:00 | 36 | 73 | 2.0 | 298 | |
| 8 | 11-08-2009 14:00 | 23 | 210 | 9.1 | 6 | Jeong et al. 2012 |
| 9 | 02-07-2010 03:00 | 31 | 44 | 1.4 | 10 | Jeong et al. 2012 |
| 10 | 16-07-2010 16:00 | 51 | 83 | 1.6 | 7 | Jeong et al. 2012 |
| 11 | 02-09-2010 00:00 | 41 | 67 | 1.6 | 45 | Jeong et al. 2012 |
| 12 | 22-06-2011 22:00 | 93 | 209 | 2.1 | 0 | Figs. 3–5 |
| 13 | 26-07-2011 18:00 | 44 | 244 | 5.5 | 9 | Fig. 3. |

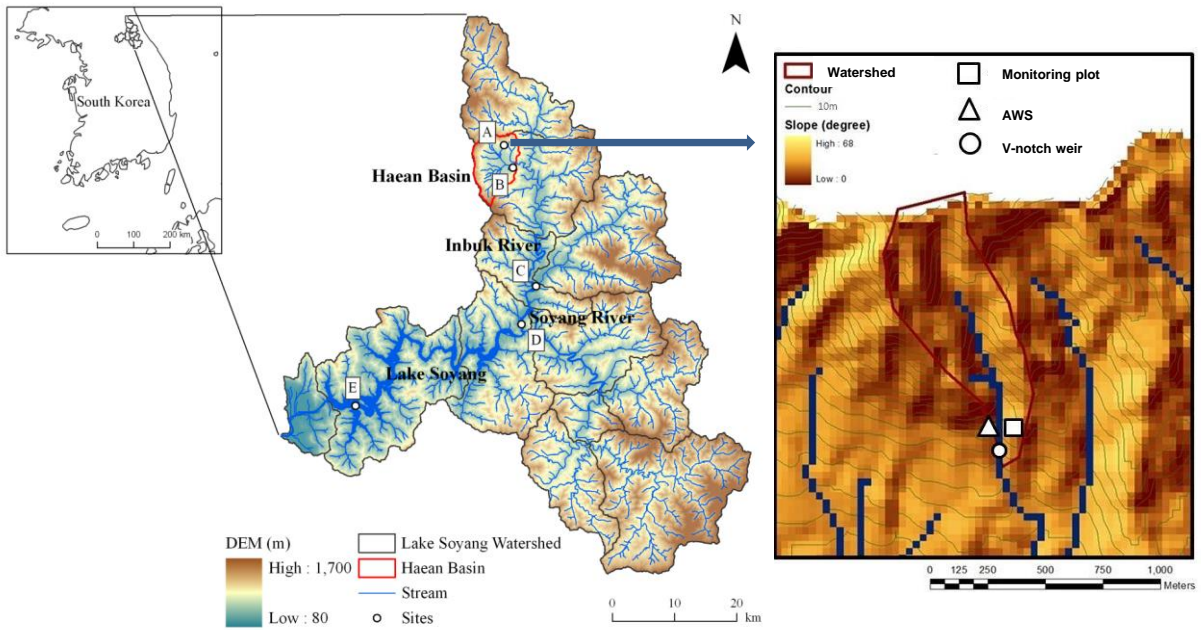
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8 **Table 2.** Changes in DOC concentrations (mg C L^{-1}) over 7 and 30 days and SUVA_{254} ($\text{L mg C}^{-1} \text{m}^{-1}$) and specific fluorescence (AU mg C^{-1}) over 30
9 days of incubation.

| Sample | Initial DOC (mg C L^{-1}) | BDOC (7 d) | | BDOC (30 d) | | SUVA_{254} (30 d) ($\text{L mg C}^{-1} \text{m}^{-1}$) | Specific fluorescence (30 d) (AU mg C^{-1}) | | |
|--------|-----------------------------------------|-----------------------------|-----------------|------------------------------|-------------------------------|----------------------------------------------------------------------|-----------------------------------------------------------|------------------------------|----------------|
| | | (mg C L^{-1}) | (%) | (mg C L^{-1}) | (%) | | C1 | C2 | C3 |
| DOM | 2.37 ^a (0.03) | 0.21 ^a (0.02) | 8.67 (0.67) | 0.17 ^a (0.02) | 7.09 ^a (0.93) | -0.21 ^a (0.04) | -0.14 ^a (0.02) | -0.21 ^a (0.02) | 0.13 (0.03) |
| DOM+SS | 2.36 ^a (0.01) | 0.20 ^a (0.03) | 8.37 (1.24) | 0.16 ^a (0.01) | 6.61 ^a (0.33) | -0.19 ^a (0.02) | -0.40 ^a (0.01) | -0.31 ^a (0.00) | - - |
| SS | 0.44 ^b (0.01) | 0.05 ^b (0.01) | 10.55 (2.63) | -0.06 ^b (0.01) | -12.86 ^b (0.95) | -1.60 ^b (0.28) | -1.45 ^b (0.17) | -0.85 ^b (0.12) | 0.59 (0.23) |

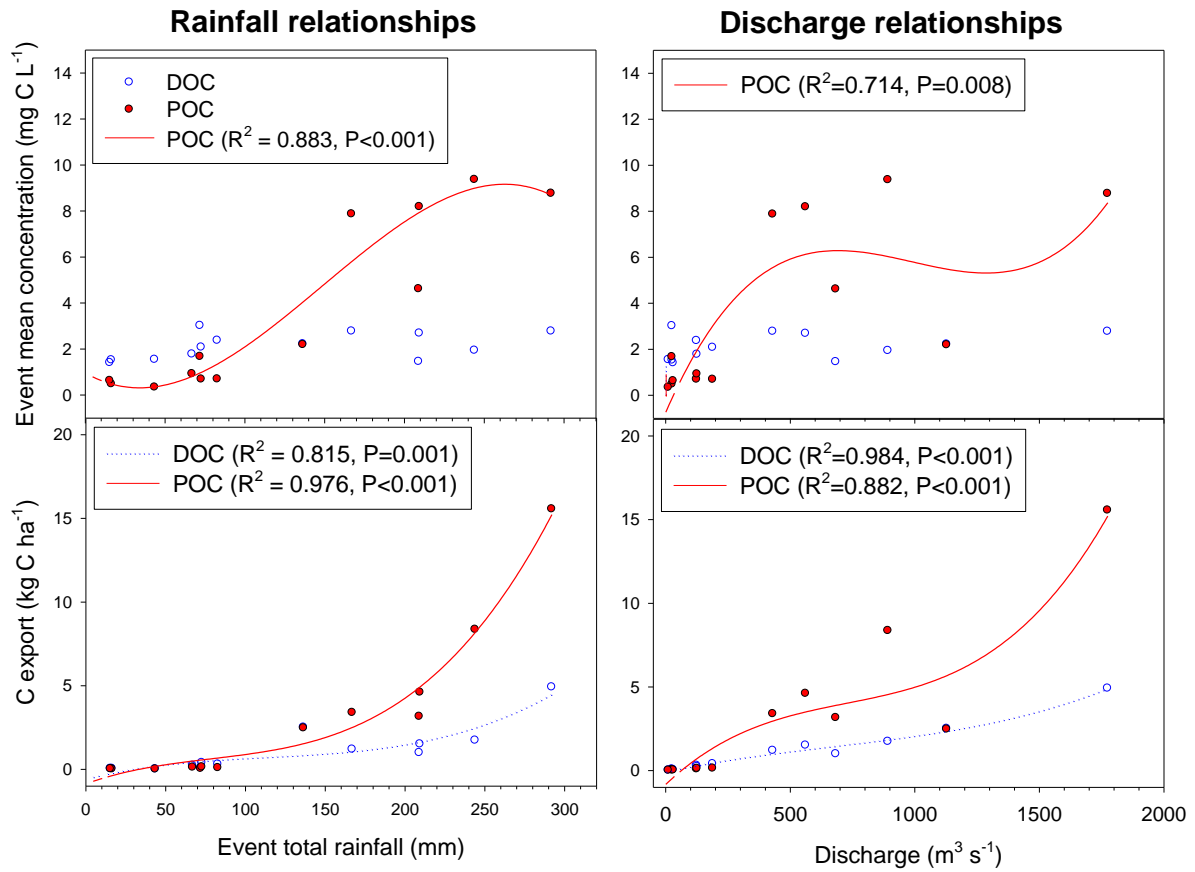
10 Values are means followed by standard errors in parentheses ($n = 3$). Different letters indicate significant differences between means at $P < 0.05$.

11 All values were calculated as the initial minus the final value, so positive and negative values indicate decreases and increases relative to the initial
12 values, respectively. BDOC was also calculated as % of the initial DOC concentration. C1, C2, and C3 indicate humic-like, fulvic-like, and protein-
13 like PARAFAC components, respectively.



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Fig. 1. Monitored headwater watershed (right) within the Haean Basin in the Lake Soyang Watershed, South Korea (left; “A” indicates the location of the headwater watershed in the Haean Basin).



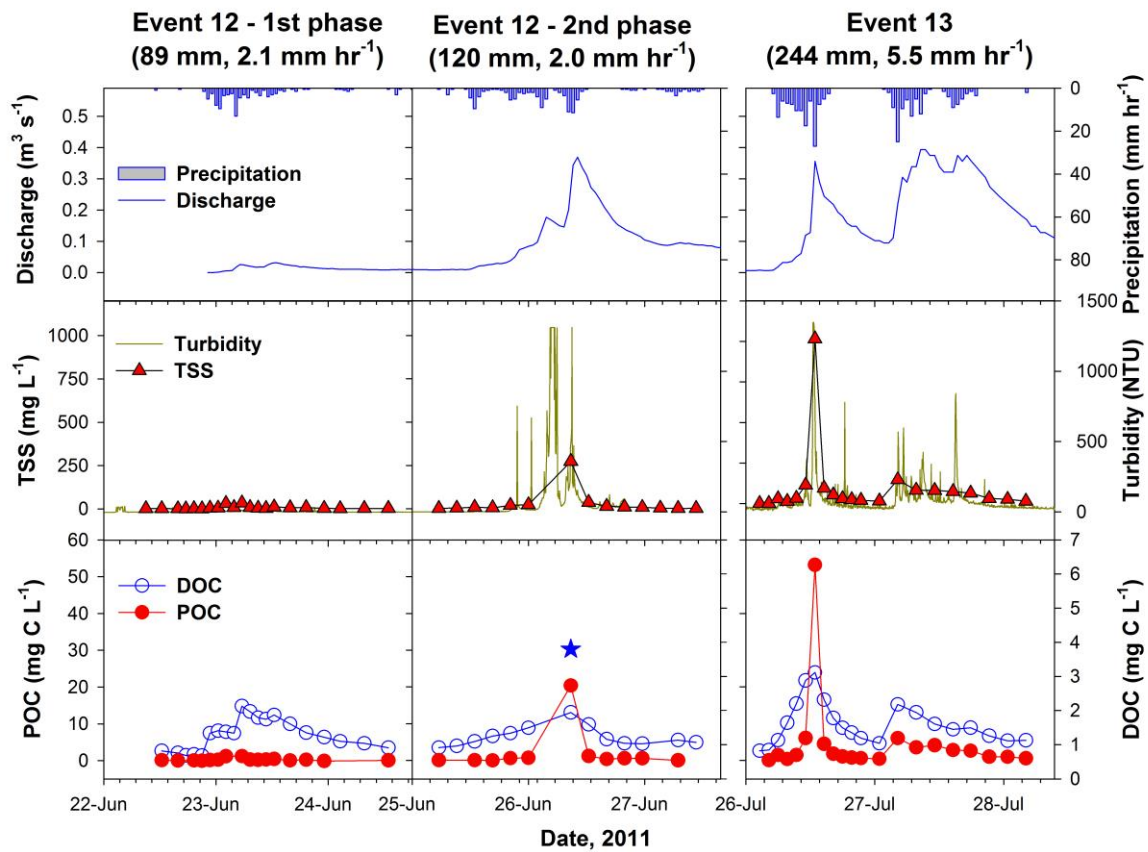
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3 **Fig. 2.** Relationships between measures of storm magnitude (event total rainfall or discharge)

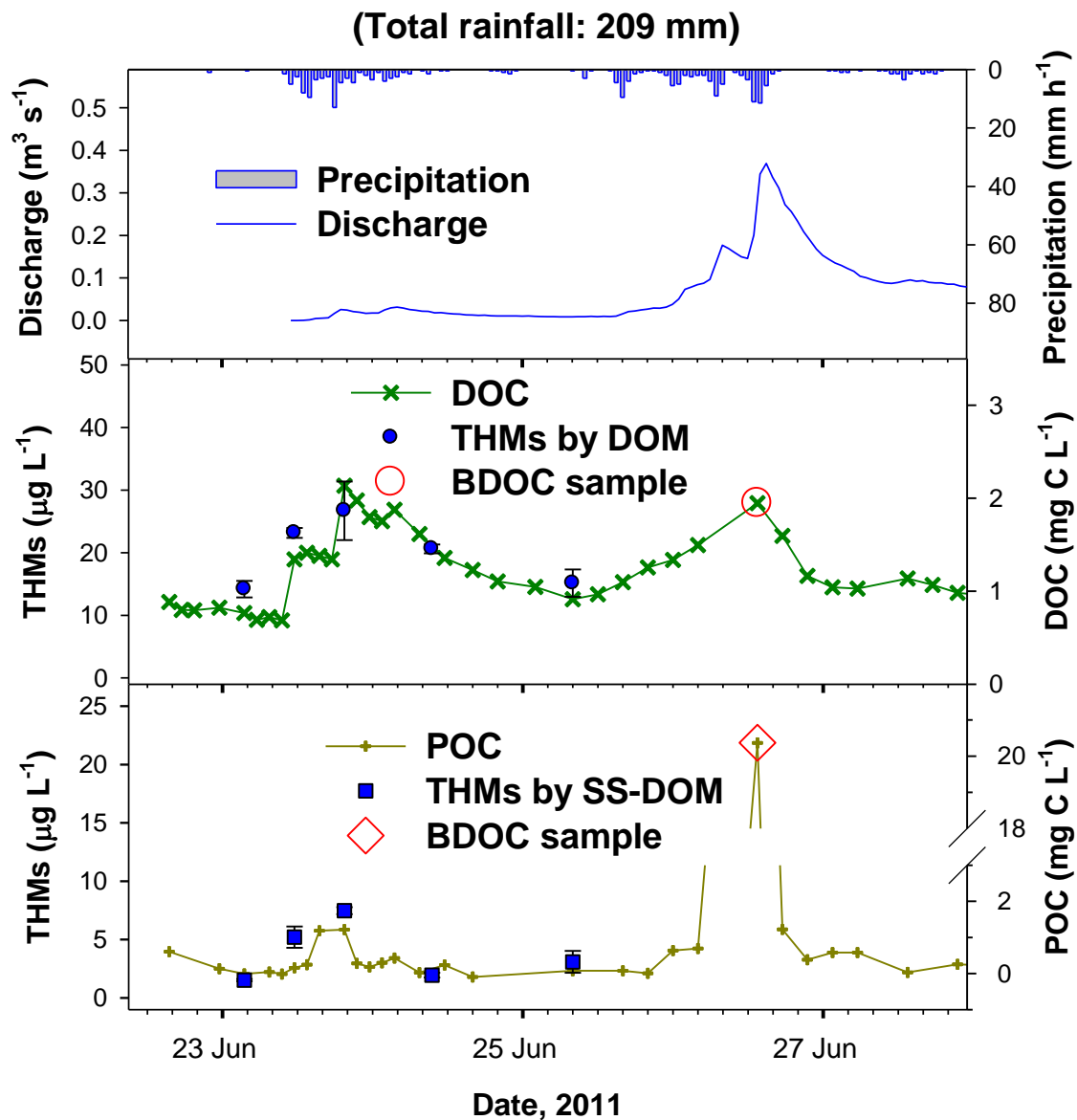
4 and the discharge-weighted event mean concentrations or the export of DOC and POC in the

5 forest stream during 13 storm events. Best-fit regressions are indicated only if $P < 0.05$.



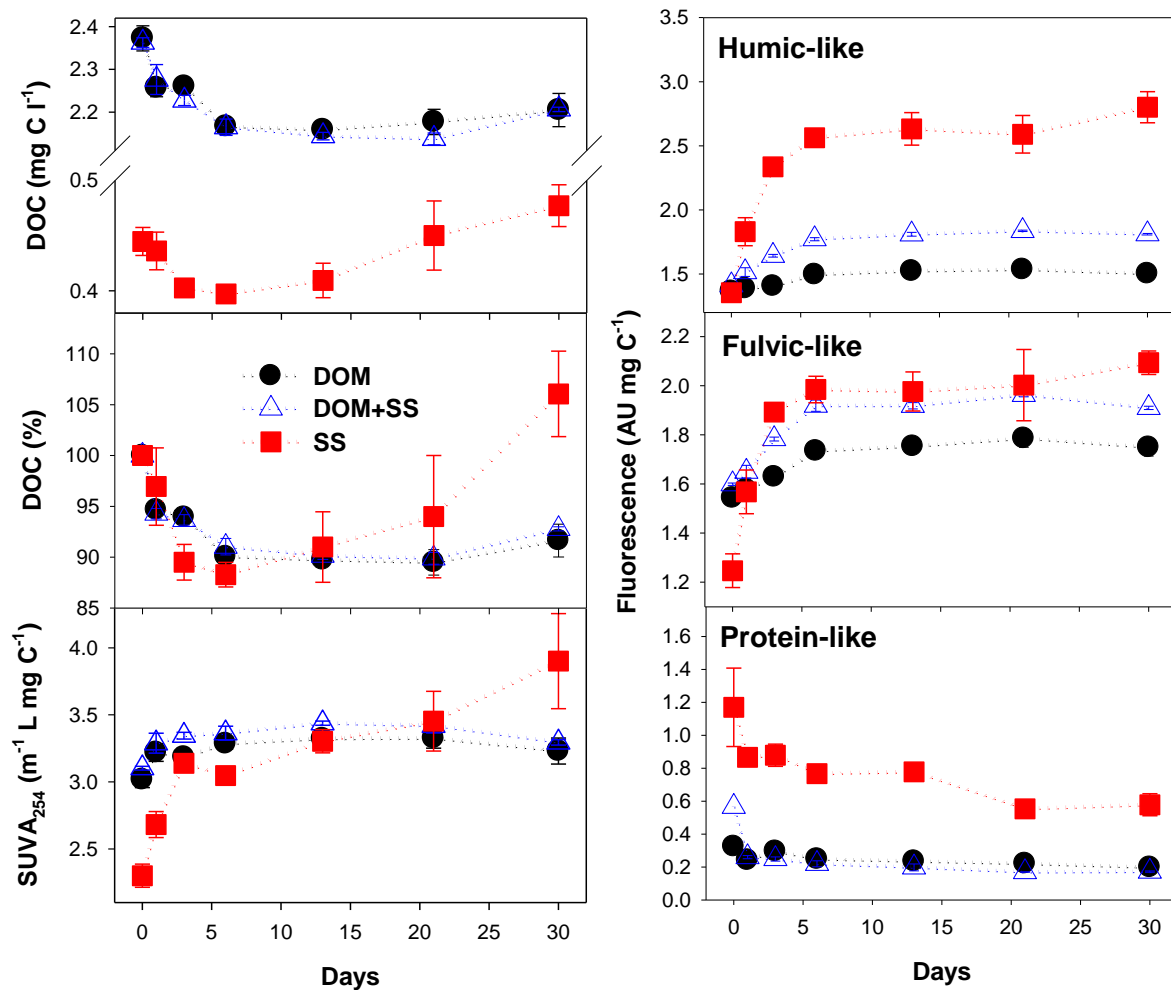
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Fig. 3. Temporal variations in hourly precipitation (mm h⁻¹), discharge (m³ s⁻¹), turbidity (NTU), and the concentrations (mg L⁻¹) of TSS, DOC, and POC in the forest stream during two extreme storm events in 2011. Concentrations of TSS, DOC, and POC were measured for water samples collected at two-hour intervals. The star symbol above the peak concentrations of DOC and POC on 26 June indicates grab sampling conducted for the laboratory BDOC incubation experiment.



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Fig. 4. Temporal changes in the concentrations (mg C L⁻¹) of DOC and POC, and the potentials (μg L⁻¹) of TTHM formation by either DOM or suspended sediment-derived DOM (SS-DOM) over the course of the first storm event in 2011. Error bars indicate ± one standard error (n = 3). ‘BDOC sample’ refers to the samples of DOC or POC collected for the biodegradation incubation experiment (Fig. 5).



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3 **Fig. 5.** Temporal changes in DOC concentrations (mg C L^{-1} or % of the initial concentration),
 4 SUVA_{254} ($\text{L mg C}^{-1} \text{ m}^{-1}$), and specific fluorescence ($\text{AU mg}^{-1} \text{ DOC}$) of three PARAFAC
 5 components (C1: humic-like fluorescence; C2: fulvic-like fluorescence; C3: protein-like
 6 fluorescence) during the 30 day incubation of DOM (filtered streamwater), DOM+SS (filtered
 7 streamwater added with suspended sediment), and SS (ultrapure water added with suspended
 8 sediment) samples. Error bars indicate \pm one standard error ($n = 3$).