

1 **Comments from the Editor**

Dear Dr. Subke,

we would like to thank you for your effort. Your suggestions were very supportive and valuable to improve our manuscript. Below you find our answers point-by point and revisions under each of your comments.

Dear Dr Keidel,

Many thanks for submitting the revised manuscript. The two referees were supportive of your work, and you have managed to address the points they raised appropriately. However, I think that more changes are required for a coherent and more concise presentation of your results. Please address my own points below for further revisions before I can accept the manuscript for publication in Biogeosciences.

With best regards,

Jens-Arne Subke

1.1 Lines 27/28: The connection between added CO₂ and target year for scenario is confusing. At current concentrations, 20% are about 80 ppm, and with an increase of about 2 ppm per year, would not be reached for around 40 years. Please re-phrase, possibly by omitting a year for the scenario, to avoid confusion.

Response: Thank you for pointing out this inaccuracy. The scenario was based on an increase of 1-1.5 ppm CO₂ per year. Consequently, we omitted a year for the scenario.

1.2 Lines 150 to 155: A grammatical detail: There is no need to phrase hypotheses as predictions for the future – they are statements about conditions irrespective of when they are measured. So I suggest re-phrasing as e.g. “long-term (>10 years) moderate CO₂ enrichment causes increased soil respiration”, etc.

Response: We re-phrased our hypotheses in present time accordingly.

1.3 Lines 280 and following: The application of a temperature response model comes a little surprising, as this is not in your list of objectives. Please clarify that this was done to fill gaps in the measured data set at the beginning of the section. Also make it clearer what fraction of data on which annual and seasonal estimates are based was from measured or modeled data. (Possibly include an additional table).

36 **Response:** We added to the description of the temperature response model: “We applied a
37 temperature response model to fill gaps in the measured data set.” Moreover we added
38 Table 3 showing which results are based on modelled and/or observed data.

39
40 **1.4** Lines 291 to 293: You should clarify whether the data you used to fir the model were
41 independent from the data you used to validate your model against. This is critical, as a
42 lack of independence means that a slope of 1 would be expected, and does not hold
43 any information with regards to the quality of the model.

44
45 **Response:** We checked the data we used to validate our model and realized that it included
46 the complete measured dataset. Consequently, 20% of the dataset was not independent
47 from the data we used to fit the model. Therefore, we re-validated our model with the
48 remaining 80 % of the measured data which were not used for the model fit. The slope
49 of the significant linear relationship changed to 1.02 and R^2 to 0.75. We changed data
50 and Figure 5 accordingly.

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52 **1.5** Line 293: Full-stop needed after “1”.

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54 **Response:** We included the missing full-stop.

55
56 **1.6** 304: Please clarify how means were calculated. Presumably you calculated annual
57 sums per ring and averaged between rings as true reps ($n = 3$)? Otherwise repeated
58 measures would have to be used.

59
60 **Response:** We specified how means were calculated. “Estimates of annual sums were then
61 calculated with the observational data and the modelled data (Table 3) per ring and
62 averaged between treatments as true steps ($n=3$). Differences in annual soil respiration
63 between the CO_2 treatments were tested by using a paired t-test.

64
65 **1.7** The discussion is lengthy and draws on several aspects not directly covered by your
66 research. This is obviously appropriate as far as interpretation is concerned, but it
67 should be shorter. The discussion of N availability and frost effect allows only very
68 indirect inferences, for example and moisture effects have not been investigated
69 systematically, and should in my view be addressed more briefly.

70
71 **Response:** We reduced the discussion according your suggestions. Freeze/thaw cycles were
72 removed from the discussion completely.

73

74 **1.8** The discussion of temperature sensitivity is in my opinion not well suited to the paper.
75 There are a number of issues regarding changes in sources of CO₂ through the
76 seasons, reference temperature depth etc., which would have to be analysed in much
77 more detail to warrant this discussion. You did not introduce the issue of temperature
78 sensitivity and have no hypothesis that addresses the issue. In the interest of the
79 coherence of the paper, I recommend that you omit the discussion of the temperature
80 effects completely. Where the use of the temperature function for gap filling purposes
81 is concerned, this is appropriate to include.

82 **Response:** We omitted the discussion of temperature dependence of soil respiration and
83 Figure 4b.

84

85 **1.9** Lines 558 to 563: The sentence is awkward to read. Please simplify, e.g. to: "Previous
86 results from the GiFACE site show that in periods when soil moisture in the main
87 rooting zone was low (0.3 m³ m⁻³), soil continued to produce N₂O from deeper soil
88 layers (20 – 50 cm), where soil moisture remained high (c. 0.6 m³ m⁻³) (Müller et al.,
89 2004)."

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91 **Response:** We simplified the sentence according your suggestion.

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93 **1.10** Units should be in standard SI units, so m³ m⁻³, rather than cm³ cm⁻³.

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95 **Response:** We changed units from cm³ cm⁻³ to m³ m⁻³.

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97 **1.11** Table 1: Please either expand column headings, or explain shorthand in the legend.

98

99 **Response:** Table 1: We expanded column headings.

100

101 **1.12** Table 2 as well as in the text (e.g. lines 345-346 and 356): Please round figures to at
102 most 4 significant digits. (error terms should be rounded to same decimal place as
103 mean values). Explain what the p-value indicates (difference between treatments per
104 year obtained by which test). I also suggest referring to the table in the results section.

105

106 **Response:** Table 2 and text (e.g. lines 345-346 and 356): We rounded means to 4 significant
107 digits and error terms to the same decimal place. We added to the legend: "P-values
108 indicate the difference between treatments per year obtained by a paired t-test."

109

1.13 Fig. 2: Please separate measured CO₂ fluxes from modelled fluxes. I could not find any indication of the proportion of measured and modelled fluxes, and this would be an appropriate place to illustrate this. It may be best to use separate panels rather than changes in colour or symbols that may be hard to read in the figure.

Response: We separated measured CO₂ fluxes from modelled fluxes in Figure 2 using separate panels. By adding the figure of the measured dataset we realized some further data gaps that we added to lines 227-229. No measurements of soil respiration were made in November and December 2008 but in the following two years.

1.14 Fig. 3: Explain what the error bars show. Are these the errors associated by averaging across the three years (i.e. $n = 3$), or are errors of individual estimates propagated?

Response: We added to the legend of Figure 3: "Error bars show ± 1 SE associated by averaging across the three replicates per treatment ($n=3$). P-values indicate the difference between treatments obtained by a linear mixed-effect model analysis."

1.15 There is a mismatch between the legends for Figure 4 and 5.

Response: We exchanged legends of Figure 5 with legend of Figure 4.

1.16 Figure 5 shows one set of data, when it should show both ambient and elevated CO₂ data, I presume? The lines in the graph have to be explained in the legend.

Response: We distinguished between ambient and elevated CO₂ data in Figure 5 and removed lines from the figure which were not explained in the legend.

1.17 Figure 6 shows the same data as Table 6 and should be omitted.

Response: Figure 6 was omitted.

146 **Title:** **Positive feedback of elevated CO₂ on soil respiration in late autumn and**
147 **winter**

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159 **Keywords:** FACE, grassland, carbon cycle, seasonality, Li-8100, winter climate change,
160 winter dormancy, feedback effect, soil respiration, soil CO₂ efflux

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162 **Words:** ~~6419707~~ (main text)

163 **Figures:** ~~56~~

164 **Tables:** ~~32~~

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169 **Abstract**

170 Soil respiration of terrestrial ecosystems, a major component in the global carbon cycle is
171 affected by elevated atmospheric CO₂ concentrations. However, seasonal differences of
172 feedback effects of elevated CO₂ have rarely been studied. At the Giessen Free-Air CO₂
173 Enrichment (GiFACE) site, the effects of +20 % above ambient CO₂ concentration
174 ~~(corresponds to conditions reached 2035 – 2045)~~ have been investigated since 1998 in a
175 temperate grassland ecosystem. We defined five distinct annual seasons, with respect to
176 management practices and phenological cycles. For a period of three years (2008-2010),
177 weekly measurements of soil respiration were carried out with a survey chamber on
178 vegetation-free subplots. The results revealed a pronounced and repeated increase of soil
179 respiration under elevated CO₂ during late autumn and winter dormancy. Increased CO₂ losses
180 during the autumn season (September-October) were 15.7 % higher and during the winter
181 season (November – March) were 17.4 % higher compared to respiration from ambient CO₂
182 ~~control~~ plots.

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183 However, during spring time and summer, which are characterized by strong above- and
184 below-ground plant growth, no significant change in soil respiration was observed at the
185 FACE site under elevated CO₂. This suggests (i) that soil respiration measurements, carried
186 out only during the growing season under elevated CO₂ may underestimate the true soil-
187 respiratory CO₂ loss (i.e. overestimate the C sequestered) and (ii) that additional C assimilated
188 by plants during the growing season and transferred below-ground will quickly be lost via
189 enhanced heterotrophic respiration outside the main growing season.

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192 **1 Introduction**

193 The atmospheric concentration of CO₂ has increased from pre-industrial values of 275 - 285
194 ppm (Raynaud and Barnola, 1985) to 400 ppm in 2013 (Monastersky, 2013). Projections of
195 future atmospheric CO₂ concentration in the year 2100 range between 490 and 1370 ppm
196 depending on representative concentration pathways (Moss et al., 2010). As the major
197 radiative forcing component (IPCC, 2013), atmospheric CO₂ is positively correlated with air
198 temperature and is therefore an important component for global warming. Additionally,
199 indirect effects of elevated atmospheric CO₂ (eCO₂), which are altering carbon (C) fluxes in
200 ecosystems, may impose a feedback to climate change. About half of photosynthetically
201 assimilated C returns immediately to the atmosphere as plant-respired CO₂ (autotrophic
202 respiration) (Chapin et al., 2002). Portions of the net carbon gain (net primary production) are
203 transferred to the soil via root exudates, fine root growth and -turnover or other litter,
204 providing the substrate for soil organic carbon (SOC) buildup (Kirschbaum, 2000).

205 Soil functions as an important C reservoir within the global carbon cycle and stores about
206 1500 Gt of C (Amundson, 2001; Lal, 2004; Batjes, 1996), which is about twice the amount of
207 C in the atmosphere (Schils et al., 2008).

208 Soil respiration, the sum of autotrophic root respiration and heterotrophic respiration from
209 microorganisms and soil meso- and macrofauna, accounts for two thirds of the total C loss from
210 terrestrial ecosystems (Luo, 2006). Enhanced net C losses under eCO₂ cause a positive feedback.

211 Many past studies focused on soil-atmosphere CO₂ exchange during the growing season.
212 However, soil respiration during vegetation dormancy may represent a significant component
213 of the annual C budget and contributes to the observed winter CO₂ maximum in the
214 atmosphere (Raich and Potter, 1995). Accordingly, analysis of CO₂ data from an air sampling
215 network identified seasonal oscillation with highest concentrations occurring each winter

when respiration exceeds photosynthesis (Keeling et al., 1996). This emphasizes the necessity to study seasonal dynamics of soil respiration under future CO₂ conditions to gain a better understanding of how soil respiration responds to changing atmospheric CO₂ concentrations. A meta-analysis of Zak et al. (2000) revealed a 51 % increase of soil respiration as a mean response in a grassland ecosystem under elevated CO₂, Janssens & Ceulemans (2000) provided evidence for consistent stimulation of soil respiration under a variety of tree species. However, the majority of studies, to date, are based on short-term exposure (less than five years) with *e*CO₂, often using open-top chamber experiments (Zak et al., 2000). Results from these experiments should be analyzed with appropriate caution because of the known “chamber effect” on the microclimate (Leadley and Drake, 1993) and their relevance to natural ecosystems in which longer-term biogeochemical feedbacks operate (Rastetter et al., 1991). Since soil respiration is a product of several rhizospheric processes i.e. root exudation, root respiration, and root turnover, as well as decomposition of litter and bulk soil organic matter from various pools with different characteristic turnover times, short- and long-term responses to *e*CO₂ may be quite different (Luo et al., 2001).

The most suitable approach for conducting ecosystem CO₂ experiments under natural conditions are FACE experiments, where intact ecosystems are exposed in-situ to a higher atmospheric CO₂ concentration. However, it has been reported that the sudden increase in atmospheric CO₂ (CO₂ step increase) at the beginning of a CO₂-enrichment, may cause certain short-term responses of the ecosystem that differ from long-term responses (Luo, 2001; Newton et al., 2001). Accordingly, Kammann et al. (2005) showed that yield responses to *e*CO₂, in the Giessen Free-Air CO₂ Enrichment (GiFACE), were different in the initial compared to the subsequent years. Moreover, plants may undergo micro-evolutionary changes in response to *e*CO₂ (Ward and Kelly, 2004), which may also be reflected in belowground processes (Klironomos et al., 2005). Consequently, to avoid misinterpretations due to insufficient experimental durations, results from long-term exposure studies are required. In the GiFACE this was after approximately 5-

242 6 years (Kammann et al., 2005). In the following we use the expression “short-term” for CO₂
 243 enrichment durations <5 years and “long-term” for durations >5 years.

244 Based on a literature overview, we found 13 other FACE studies, from a wide variety of
 245 ecosystems, where in-situ soil respiration under *e*CO₂ has been investigated. All of these
 246 FACE studies operated at higher CO₂ enrichment concentrations than the GiFACE
 247 experiment (with +20 % CO₂ above ambient), i.e. they imposed larger initial step increases
 248 (Klironomos et al., 2005). Klironomos et al.(2005) have demonstrated that ecosystem responses
 249 to *e*CO₂ may differ between using a sudden step increase and a gradual rise in the CO₂
 250 concentration. However, in any CO₂ enrichment study a step increase – also if lower than usual –
 251 cannot be avoided. Thus, experimental FACE results are more indicative for future predictions.

252 However; experimental studies with duration of > 10 years are scarce (Carol Adair et al.,
 253 2011;Jackson et al., 2009). To our knowledge, 10 of the 16 investigations on soil respiration
 254 across these 13 FACE studies were carried out within the first five years of exposure, thus
 255 reporting short-term responses (Craine et al., 2001;King et al., 2001;Allen et al., 2000;Andrews
 256 and Schlesinger, 2001;Selsted et al., 2012;Masyagina and Koike, 2012;Soe et al.,
 257 2004;Lagomarsino et al., 2013;Liu et al., 2006;Nakayama et al., 1994). All short-term study
 258 results pointed towards a consistent stimulatory effect of *e*CO₂ on soil respiration. The average
 259 increase ranged from 12 % under a sweetgum plantation (King et al., 2004) to 70 % under a mixed
 260 plantation of *Populus* species (Lagomarsino et al., 2013). In two of the short-term studies,
 261 significant effects were only observed on days with high photosynthetic activity (Masyagina and
 262 Koike, 2012;Soe et al., 2004); measurements during dormancy were not carried out.

263 Three of the short-term studies conducted measurements during winter dormancy with contrasting
 264 results (Allen et al., 2000;Andrews and Schlesinger, 2001;Selsted et al., 2012;Lagomarsino et al.,
 265 2013). In a temperate heathland (CLIMAITE study), soil respiration was significantly increased
 266 under *e*CO₂ during three consecutive winter seasons (Selsted et al., 2012). Allen et al. (2000)
 267 detected a significant effect of *e*CO₂ on soil respiration during December 1997 in the Duke Forest

268 FACE study but not during the previous growing season beneath the loblolly pine forest. Andrews
 269 and Schlesinger (2001) reported from the same site greater increases of soil respiration during
 270 fumigation periods (26-59 %) than during non-fumigated periods (8-15 %). Fumigation was
 271 stopped when ambient air temperature dropped below 5 °C for more than one hour. In line with
 272 these results, much larger percentage enhancements of the soil CO₂ efflux were observed during
 273 the growing season (up to 111 %) than during dormant season (40 %) from a mixed plantation of
 274 *Populus* species exposed to eCO₂ (EuroFACE) (Lagomarsino et al., 2013). CO₂ enrichment was
 275 provided from bud burst to leaf fall at this site.

276 Out of six long-term studies on soil respiration (Carol Adair et al., 2011; Pregitzer et al.,
 277 2008; Jackson et al., 2009; Pendall et al., 2001; Bader and Körner, 2010; Dawes et al., 2013), only
 278 one study reported measurements throughout the dormant season, showing that after 10 years of
 279 eCO₂ during the growing season at a loblolly pine forest (Duke FACE) soil respiration was
 280 consistently higher in midsummer to early fall and diminished or disappeared in winter (Jackson
 281 et al., 2009). This was explained by a reduction in assimilation and hence available root exudate
 282 during dormancy. If the fumigation may continue during the dormant season in an ecosystem with
 283 a green canopy e.g. in a permanent grassland, the stimulation may theoretically continue on a
 284 higher level.

285 Reports from other long-term FACE studies in temperate ecosystems (disregarding the dormant
 286 season) were consistent by reporting an increase in soil respiration under eCO₂, with the exception
 287 of the Swiss Canopy Crane experiment in an old-growth, mixed deciduous forest. Bader & Körner
 288 (2010) reported that soil respiration from the site was only stimulated when volumetric water
 289 content was ≤ 40 % at soil temperatures above 15 °C.

290 In summary, only fragmented information is available on how soil respiration responds to eCO₂
 291 during vegetation as well as dormant periods after long-term eCO₂. To our knowledge, no long-
 292 term FACE study in a grassland ecosystem exists which has investigated soil CO₂ fluxes across

293 several years. Consequently, it is difficult to generalize temporal patterns of soil respiration under
294 $e\text{CO}_2$, and thus the soil respiratory response to $e\text{CO}_2$ at all.

295 Based on the available studies and earlier observations at our site, where whole-ecosystem
296 respiration including the green canopy was increased under $e\text{CO}_2$, mainly during non-growing
297 season (Lenhart, 2008), we hypothesized that (i) long-term (>10 years) moderate CO_2
298 enrichment ~~will~~ causes increased soil respiration, (ii) soil respiration ~~is~~~~will~~~~be~~ more enhanced
299 in the growing season than during vegetation dormancy (*winter*) and (iii) soil respiration ~~will~~
300 ~~still be~~~~is~~ significantly enhanced in *winter* under $e\text{CO}_2$ in the GiFACE where the CO_2
301 enrichment is continuing during *winter*.

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317 2 Materials and methods

318 2.1 Study site and design

319 The Giessen Free Air Carbon Enrichment (GiFACE) experiment is located on permanent
320 semi-natural grassland. It is situated near Giessen, Germany (50°32'N and 8°41.3'E) at an
321 elevation of 172 m above sea level.

322 The set-up and performance of the GiFACE system has been described in detail by Jäger *et al.*
323 (2003). In brief, from May 1998 until present, atmospheric CO₂ concentrations were enriched
324 by 20 % above ambient, all-year-round during daylight hours. At present the GiFACE
325 experiment is still ongoing.

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326 The CO₂ enrichment was applied in three rings, each eight meter in diameter (E plots). Three
327 equally sized control plots were maintained at ambient atmospheric CO₂ levels (A plots). The
328 experimental design was a randomized block design. A block consisted of two plots to which
329 ambient and *e*CO₂ treatments were randomly assigned. A characteristic attribute of the study
330 site is a soil moisture gradient, resulting from a gradual terrain slope (2-3°) and varying depths
331 of a subsoil clay layer. Within each of the three blocks, soil moisture conditions were
332 relatively homogeneous (Jäger *et al.*, 2003).

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333 The vegetation is an *Arrhenatheretum elatioris* Br.Bl. *Filipendula ulmaria* subcommunity,
334 dominated by *Arrhenatherum elatium*, *Galium mollugo* and *Geranium pratense*. At least 12
335 grass species, 15 non-leguminous herbs and 2 legumes are present within a single ring. For at
336 least 100 years, the grassland has not been ploughed. Since several decades, it was managed
337 as a hay meadow with two cuts per year, and fertilized in mid-April with granular mineral

calcium-ammonium-nitrate fertilizer at the rate of 40 kg N ha⁻¹ yr⁻¹. Before 1996, fertilizer was applied at a rate of 50–100 kg N ha⁻¹ yr⁻¹ (Kammann et al., 2008). The soil of the study site is classified as a Fluvic Gleysol (FAO classification) with a texture of sandy clay loam over a clay layer (Jäger et al., 2003). Observations in this study were carried out from January 2008 - December 2010 (i.e. more than 9 years after the onset of CO₂ enrichment). During the observation period the mean annual temperature was 9.2 °C and mean annual precipitation was 562 mm which was identical to the average rainfall since the beginning of recording in 1995. Rainfall was recorded at the site in 30-min intervals with 20 randomly distributed “Hellmann” samplers. Air temperature was recorded continuously at two locations at the site in 2 m height and averaged 9.5 °C since 1995.

2.2 Measurement of soil CO₂ fluxes at the field site

In each of the six FACE plots, soil respiration rates were measured using an automated closed dynamic chamber system with an infrared gas analyzer (LI-COR 8100, LI-COR, Inc., Lincoln, Nebraska, USA) with a patented vent for pressure equilibration between the closed chamber and the atmosphere (McDermitt et al., 2005). Carbon dioxide fluxes were reported in μmol CO₂ m⁻² s⁻¹. The measurements were performed at four permanently installed PVC soil collars per FACE ring, to cover the spatial heterogeneity within each ring. The soil collars had a diameter of 20.3 cm (8 inch) and were about 11 cm high. A beveled edge at one end facilitated the insertion into the soil, which took place on 9th May 2006 and the vegetation cover, including surficial rhizomes, was removed manually. Subsequently, the surface was held vegetation-free by removing germinated seedlings weekly. Due to uneven soil conditions, soil collars varied +/- 1 cm in their insertion depth. Generally, the insertion was chosen to be as shallow as possible, minimizing the trenching effect (Heinemeyer et al., 2011)

while maintaining an airtight connection between soil and chamber. A foam gasket and rubber seal between the bottom of the chamber and the top of the soil collar minimized leaks between the collar and the chamber. Before each measurement, the distance between the soil surface and the top of each soil collar (i.e. chamber offset) was measured and entered into the LICOR-software to enable correct flux calculations (= total chamber volume). After installation in May 2006, soil CO₂ efflux measurements were carried out over a period of one month to record the insertion and disturbance effects (Fig. S1). The investigation period spanned over three years (January 2008 until December 2010), after the collars were well established and held vegetation free for 1.5 years, allowing a die-back and decomposition of trenched roots, and in-growth of new roots from the outside vegetation. This ensured that soil respiration measurements in a dense, closed grassland canopy were taken as unbiased as possible.

Measurements of soil respiration were ~~made~~carried out weekly in the evening, except in July 2009. F₁ from May to July 2010 and from October to December 2010, ~~where~~ measurements were carried out every second week. No measurements were carried out in November and December 2008.

During the measurement, a pump provided circulating air flow from the closed chamber on its collar to the infrared gas analyzer for thorough mixing of the systems' inner volume. Chamber closure time was between 1 and 3 min., depending on the season (i.e. the strength of the CO₂ efflux and thus the detection limit). CO₂ and H₂O concentrations were measured simultaneously. The software calculated soil respiration rates by using the changes in CO₂ concentration over a period of time, taking the dilution of water vapor into account. Rates were calculated either by linear regression (lin_flux) or as the efflux rate at time t₀ at chamber closure using an exponential CO₂ efflux function (exp_flux) (LI-COR, 2007). The latter takes the diminishing CO₂ concentration gradient between the soil and the chamber headspace into account (Hutchinson and Mosier, 1981) and is implemented by LI-COR in the LI-8100 to

387 avoid underestimations of the CO₂ efflux. We used the following algorithm to choose between
388 these two types of flux calculation for the subsequent processing of all obtained flux data. The
389 use of the exp_flux calculation was only allowed when (1) the R² of the exp_flux calculation
390 was better than that of the lin_flux calculation, and (2) when the number of iterations
391 necessary for the exp_flux calculation was lower than 5. By applying these comparatively
392 strict criteria (stricter than those that are inbuilt by the manufacturer) we minimized
393 miscalculations caused either by large initial CO₂ concentration fluctuations at chamber
394 closure (when the exp_flux calculation is used) or underestimations of the true soil CO₂ efflux
395 (when only the lin_flux calculation is used). The algorithm was applied to each measurement
396 with the same settings. In general, CO₂ flux rates with an R² below 0.90 were excluded. This
397 was the case in 0.6 % of all measurements taken in this study throughout the three year
398 investigation period.

399 Soil moisture was measured in each FACE plot as the volumetric water content (VWC) with
400 time-domain-reflectometric (TDR) probes (Imko, Ettlingen, Germany, type P2G). The probes
401 were permanently installed (in March 1998) within the top 15 cm. The probes were monitored
402 manually once a day, except on weekends or holidays. Soil temperature was logged in every
403 plot at 10 cm depth as 30-min means (Imko, Ettlingen, Germany, Pt-100 sensors).

404 2.3 Data analyses

405 In order to describe changes in soil respiration during different seasons and to test for
406 differences in soil respiration between ambient and elevated CO₂, we performed a linear
407 mixed-effect model analysis with SPSS version 18. We used all observational data of three
408 years for the linear mixed-effect model analysis. CO₂ treatment was considered as a fixed
409 effect in the model. Coding variables were introduced to indicate the hierarchical order of the
410 data. The six mean fluxes taken in one measurement cycle received the same numerical code;

411 this variable (“measurement cycle”) was considered as a random effect in the linear mixed
 412 effect model. A further variable (“ringreplicate”) was introduced to define the ring where the
 413 measurement was taken (1-6). ”Ringreplicate” was selected as a repeated measure in the SPSS
 414 software using linear mixed effect model analysis. Maximum likelihood was used as the
 415 estimation method for the parameters in the model. The total observational data set was split
 416 by season to analyze seasonal CO₂-response patterns. Therefore, we distinguished the
 417 following five seasons (1 – 5), depending on major dates of phenology and management
 418 practices at the grassland study site (Fig. 1): **1** = *winter* (November – March); **2** = *start of*
 419 *vegetation period* up to the date of spring fertilizer application (March – middle of April); **3** =
 420 *spring* until first biomass harvest (middle of April – end of May); **4** = regrowth and *summer*
 421 growing season (end of May – beginning of September); **5** = regrowth and *autumn* growing
 422 season (beginning of September – end of October).

423 The start of the vegetation period for the grassland ecosystem was identified according to the
 424 calculations defined by Wasshausen (1987). The date of leaf discoloration of *Quercus robur*
 425 in the nearby phenological garden was used to identify the beginning of winter dormancy. All
 426 other dates were chosen according to the management practices at the study site (Fig. 1); the
 427 exact dates varied by a few days between the years.

428 2.4 Soil respiration model

429 We applied a temperature response model to fill gaps in the measured data set. Therefore

430 ~~In order to describe the dependence of soil respiration on temperature,~~ a function was fitted
 431 according to Lloyd & Taylor (1994) (Eq. 1) to 20 % of the data that were randomly selected.

432 We defined values for coefficients E0 (= 62.16), T0 (= 262.47) and R10 (= 2.85) for the first
 433 run of the model. Subsequently, E0, T0 and R10 were fitted for each treatment (ambient and

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$e\text{CO}_2$) by using the dynamic fit function in the SigmaPlot 11.0 software package (Systat Software, San Jose, CA, 2008). Mean soil temperature values were converted from °C to K.

$$f = R10e^{E0\left(\frac{1}{(283.15-T0)} - \frac{1}{(x-T0)}\right)} \quad \text{Eq. (1)}$$

with $E0$ = activation-energy-type empirical coefficient

$T0$ = lower temperature limit for soil respiration in K

$R10$ = respiration rate at 10 °C

Consequently, the quality of the soil respiration model was evaluated by plotting modelled soil respiration rates against the remaining 80% of the observed respiration values to test if the linear trend line meets the requested slope of 1 (Fig. 5). ~~We plotted the temperature relationship of soil respiration of the complete dataset, visualizing the different seasons to show seasonal differences (Fig. 5b) of the relationship. However, we did not include seasonal analyses due to the fact that in some seasons there were not enough data points and statistical power was not sufficient ($R^2=0.2$) to justify this kind of analysis.~~

2.5 Gap filling of soil respiration data

To obtain annual sums of soil respiration, a gap filling procedure was applied. Therefore modelled soil respiration rates were calculated, based on the almost continuous data set of soil temperature in 10 cm depth measured at 2-3 positions per ring. We received modelled fluxes for every 15 minutes over the three year period for all gaps where no observational data were available. Estimates of annual sums were then calculated with the observational data and the modelled data (Table 3) per ring and averaged between treatments as true steps (n=3). Differences in annual soil respiration between the CO_2 treatments were tested by using a paired t-test. Further, the absolute difference and relative change of monthly mean soil respiration rates under $e\text{CO}_2$ were calculated in comparison to soil respiration under ambient

457 | CO₂, based on observational and modelled data [\(Table 3\)](#). For calculating the relative change
458 | ambient soil respiration was set to 0 %.

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474 3 Results

475 3.1 Annual variability of soil respiration

476 From 2008 to 2010, soil respiration rates at the GiFACE experiment showed distinct annual
477 dynamics, following the seasonal temperature cycle with lowest soil respiration effluxes
478 during winter months and highest effluxes during mid-summer (Fig. 2c and 2g). Thus, soil
479 respiration rates responded to abiotic factors in particular temperature and moisture. This is
480 exemplified by the high CO₂ efflux rates in June 2009 which occurred shortly after a period of
481 high precipitation while soil temperatures were > 20 °C (Fig. 2g).

482 The relative and absolute change of soil respiration under *e*CO₂ (Fig 2d and 2e) followed a
483 seasonal pattern with greatest increases under *e*CO₂ during *autumn* and *winter*. During
484 midsummer, when the largest absolute soil respiration rates occurred, the relative increase due
485 to the CO₂ enrichment was lowest or non-existent. A linear mixed effect model analysis
486 confirmed that soil respiration rates under *e*CO₂ were significantly higher compared to rates
487 under ambient CO₂ during *autumn* (15.7 %) and *winter* (17.4 %) (Fig. 3). During all other
488 seasons (*beginning of vegetation period* (season 2), *spring* (season 3) and *summer* (season 4)),
489 covering most of the vegetation period, a trend towards higher soil respiration, but no
490 significant CO₂ effect was observed with *e*CO₂ (Fig. 3).

491

492 3.2 Model performance and parameter estimation

493 By comparing modelled soil respiration with observed soil respiration for all observation
494 dates from 2008 – 2010 a significant linear relationship was observed with a slope of 1.0²³
495 (Fig. 4⁵).

Based on the temperature-respiration function by Taylor & Lloyd (1994), soil respiration was significantly correlated to soil temperature under ambient as well as $e\text{CO}_2$ ($p = <0.0001$). From 2008 to 2010, 75 % of the variability of soil respiration rates was explained by soil temperature under ambient CO_2 and 82 % under $e\text{CO}_2$ (Fig. 4, Table 1). Soil respiration rates did not differ in their relationship to soil temperature between the treatments (Fig. 4a). In Fig. 5b we plotted the temperature relationship of soil respiration, visualizing the different seasons, which indicated that soil CO_2 efflux data from autumn imposed a different relationship to soil temperature compared to data from other seasons. During autumn, soil temperatures were within the same range as during spring and summer, but soil respiration was on average lower (Fig. 2).

3.3 Annual sums of soil respiration

Comparing annual sums of soil respiration, no mean treatment effect of elevated CO_2 (over all seasons) was observed in any of the observation years (Table 2, Fig. 6). Mean annual estimates of soil respiration under ambient CO_2 ranged from 12832.48 to 1344.00 g C [CO_2] $\text{m}^{-2} \text{yr}^{-1}$ and under $e\text{CO}_2$ from 1300.15 to 13521.56 g C [CO_2] $\text{m}^{-2} \text{yr}^{-1}$ (Table 2).

517 4 Discussion

518 4.1 Annual sums of soil respiration

519 In contrast to our initial hypotheses, annual estimates of soil respiration were not different
520 between the CO₂ treatments (Table 2, ~~Fig. 6~~). Mean annual sums of soil respiration were
521 $131\overline{76.76} \pm 18\overline{10}$ g C m⁻² yr⁻¹ under ambient CO₂ and $133\overline{10.58} \pm 1\overline{65.57}$ g C m⁻² yr⁻¹ under
522 elevated CO₂. Raich and Schlesinger (1992) estimated much lower rates of annual soil
523 respiration, reporting 400 to 500 g C m⁻² yr⁻¹ for temperate grasslands. Annual soil respiration
524 sums from a sandstone and serpentine grassland were 485 and 346 g C m⁻² yr⁻¹ (Luo et al.,
525 1996). These soil respiration rates were lower than those from the wet grassland site
526 investigated here due to the larger net primary productivity of the wet temperate grassland
527 with a year-round more or less moist climate, compared e.g. to a seasonally dry
528 Mediterranean-type grassland. A lower net ecosystem productivity (NEP) will automatically
529 result in lower overall soil respiratory C losses. Methodological differences may have been to
530 a lesser extent been responsible, because the studies of Luo et al. (1996) and Raich and
531 Schlesinger (1992) may have overestimated rather than underestimated the annual soil
532 respiration. Their measurements did not exceed 2 years in duration and soil respiration was
533 less frequently measured for a portion of the year. Other recent studies reported higher rates of
534 annual soil respiration which are closer to our estimates; however climatic factors are different
535 from our site: In a tallgrass prairie of Oklahoma annual soil respiration rates were 1131 and
536 877 g C m⁻² yr⁻¹ in 2002 and 2003 respectively (Zhou et al., 2006). In a Texas grassland
537 annual soil respiration rates increased with annual precipitation and were 1600, 1300, 1200,
538 1000, 2100 and 1500 g C m⁻² yr⁻¹ in 1993 through 1998 respectively (Mielnick and Dugas,
539 2000). At the Texas grassland site measurements were conducted year-round with a high time
540 resolution. Consequently annual rates could be estimated by more measured (than gap-filled)

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data compared to other studies. However the most important factors were likely the annual precipitation, its distribution over the year, and the annual mean temperature: High annual rainfall, a long growing season and large soil organic C contents explained the higher soil respiration rates (as a consequence of a higher NEP) at the Texas study site. Mean annual precipitation at the GiFACE study site (562 mm) was close to the mean precipitation reached in 1995 at the Texas grassland with 657 mm, when annual soil respiration averaged 1200 g C m⁻² yr⁻¹ at the Texas grassland.

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4.2 Seasonality of soil respiration

Also, contrary to our initial hypotheses is the observation that soil respiration was not significantly affected during the growing season (*start of vegetation period, spring and summer*) by the moderate long-term CO₂ enrichment. This indicates that any increase in the ecosystem respiration (Lenhart, 2008) during this season will not have been due to enhanced soil (root-derived) respiration but rather to increases in the respiration of the green canopy.

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The majority of long-term FACE studies reported significantly increased soil respiration under eCO₂ during the growing season (Pregitzer et al., 2008; Jackson et al., 2009; Pendall et al., 2001; Dawes et al., 2013; Carol Adair et al., 2011), whereas Bader & Körner (2010) reported that seven years of eCO₂ failed to stimulate cumulative soil respiration significantly during the growing season. Among the mentioned long-term FACE experiments, the GiFACE operates at the lowest CO₂ enrichment step increase (20 % above ambient CO₂), which may have contributed to this result.

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However, in line with our hypotheses, the results revealed that 10 years of moderate CO₂ enrichment increased soil respiration during *winter* and *autumn* (Fig. 3). These seasonal stimulations of soil respiration under eCO₂ were not observed by comparing the annual sums

564 of soil respiration (Table 2 Fig-6). This may be because soil respiration fluxes were lower in
 565 *winter* and *autumn* compared to fluxes from the other seasons where no differences in soil
 566 respiration between the CO₂ treatments were observed. However, within the *winter* and
 567 *autumn* season differences in soil respiration may play an important role concerning the
 568 global C balance. Increased rates of winter soil respiration under eCO₂ may increase the
 569 observed winter CO₂ maximum in the atmosphere (Raich and Potter, 1995; Keeling et al.,
 570 1996) when respiration exceeds photosynthesis. Another reason why annual sums of soil
 571 respiration were not different between the CO₂ treatments may be that our model
 572 underestimated high soil respiration fluxes (>10 μmol m⁻² s⁻¹). However these fluxes occurred
 573 only in 1.72 % of all observations. Our model did not take soil moisture into account. The
 574 high variability of observed soil respiration during summer may be partly due to differing soil
 575 moisture conditions, which were not significantly different between ambient and eCO₂ plots
 576 (Kammann et al., 2005; 2008).

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577 In most FACE studies which reported the effect of eCO₂ on soil respiration, the winter was
 578 excluded since fumigation during this period was mostly switched off (often in response to
 579 sub-zero freezing temperatures or deciduous forest ecosystems). This was the case in the
 580 Swiss FACE study, where seeded grassland was exposed to 600 ppm CO₂ (de Graaff et al.,
 581 2004), the BioCON FACE, also a grassland study (Craine et al., 2001; Carol Adair et al.,
 582 2011), the Aspen FACE, an aspen forest enriched with eCO₂ (Pregitzer et al., 2008; King et
 583 al., 2001), a Japanese model forest ecosystem exposed to 550 ppm CO₂ (Masyagina and
 584 Koike, 2012) and in a 9-year FACE study of an alpine treeline ecosystem (Dawes et al.,
 585 2013). In the Swiss Canopy Crane study soil respiration was measured during the beginning
 586 of the dormant season but not over the complete dormant season while fumigation was
 587 switched off (Bader and Körner, 2010). In the Maricopa FACE, where a wheat field was

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588 exposed to $e\text{CO}_2$, no winter measurements were carried out because this season was a fallow
589 season (Pendall et al., 2001). Outside the cultivation period no soil respiration measurements
590 were made on a cotton plantation exposed to $e\text{CO}_2$ (Nakayama et al., 1994).

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591 Increased winter soil CO_2 fluxes are in line with results from Selsted et al.(2012), who
592 reported stimulated rates during three consecutive winter periods in a Danish N-limited
593 *Calluna-Deschampsia*-heathland exposed to FACE at 510 ppm (CLIMAITE study).
594 Fumigation was carried out all year-round except during periods with full snow cover.
595 Contrary to our results, in the CLIMAITE study, the stimulatory effect of $e\text{CO}_2$ on soil
596 respiration persisted throughout most of the year, i.e. also in summer and not only during
597 winter. However, in the CLIMAITE study, monthly soil respiration measurements were
598 carried out within the first three years after the experimental start and may therefore reflect
599 short-term responses, driven by the initial CO_2 step increase (Klironomos et al., 2005). Thus
600 the results are not completely comparable to this study where measurements were carried out
601 in the 11th – 13th year of CO_2 enrichment.

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602 To our knowledge, the Duke Forest FACE is the only other FACE experiment where soil
603 respiration was measured in an evergreen ecosystem year-round for several years and after
604 long-term fumigation with $e\text{CO}_2$ (+200 ppm). On average, soil respiration was significantly
605 higher by 23 % under $e\text{CO}_2$. Jackson et al. (2009) summarized, after 10 years of CO_2
606 enrichment, that the greatest stimulation of soil respiration under $e\text{CO}_2$ occurred from
607 midsummer to early fall, in contrast to our observations, during winter the CO_2 response of
608 soil respiration was weakest. However, fumigation was stopped at the Duke Forest FACE
609 when ambient air temperature dropped below 5°C for more than one hour.

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610 After short-term enrichment with $e\text{CO}_2$ (550 ppm) on a mixed plantation of *Populus* species
611 (EuroFACE; in the 4th and 5th year of enrichment), Lagomarsino et al. (2013) recorded much
612 larger stimulation of soil respiration during the vegetation (up to 111 % enhancement) than
613 dormant season (40 % enhancement), when fumigation was stopped, which is also contrary to
614 our results. However, experimental setup and climate differed from our site. While minimum
615 soil temperatures reached -1.7 °C in the GiFACE experiment during winter (Fig. 2b),
616 comparably warm and mild winters without sub-zero temperatures were typical at the
617 EUROFACE site located in Italy. Moreover, the *Populus* plantation was a fertilized agro-
618 ecosystem, where coppicing was carried out every three years, while the GiFACE was an old
619 established, species-rich ecosystem where N-supply was limited.

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620 In line with results from the EuroFACE but in contrast to our findings, Volk & Niklaus (2002)
621 did not observe any wintertime increase in the ecosystem CO_2 efflux from a calcareous
622 grassland in response to three years of CO_2 enrichment (600 ppm) with a screen-aided CO_2
623 enrichment facility.

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624 Investigations from the GiFACE experiment showed that N_2O emissions also exhibited a
625 “seasonality response”, with the greatest stimulation of N_2O emission under $e\text{CO}_2$ being
626 observed in late-summer and autumn (Kammann et al., 2008). These findings support the
627 hypothesis that the driving mechanism of the $e\text{CO}_2$ seasonality responses of enhanced
628 microbial activity may have been related to the mineralization of previously accumulated
629 organic matter, fuelling denitrification (Kammann et al., 2008).

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632 4.3 Root derived soil respiration

633 Increased root biomass was frequently recorded under $e\text{CO}_2$ (Rogers et al., 1994; Jastrow et
634 al., 2000; Lukac et al., 2009), potentially affecting soil respiration rates (Zak et al., 2000).

635 However, at the GiFACE, root biomass, picked with forceps (for set time intervals per
636 sample, $n=3$ per FACE ring), was only different in December 2005 between the CO_2
637 treatments but not at other dates during 2004 – 2007 (Lenhart, 2008) or in November 2011
638 (unpublished results). Lenhart (2008) observed in the GiFACE $e\text{CO}_2$ plots, using Keeling
639 plots and two-component mixing models that the fraction of root-derived CO_2 (root- and root-
640 exudate respiration and fine root decay), as part of the total soil CO_2 efflux was lower in
641 winter than during the growing season. Accordingly, during *winter*–, the soil CO_2 efflux
642 originated mainly from microbial soil respiration.

643 Higher fine root turnover under $e\text{CO}_2$, resulting in higher C input via root necromass could
644 explain increased *autumn* soil respiration but unlikely the *winter* increase in soil CO_2 efflux at
645 the GiFACE since root necromass was not changed under $e\text{CO}_2$ in November 2011
646 (unpublished results). Alternatively, differences in the root necromass could already have
647 been decomposed at this time of sampling or may be observed later in the year, so that
648 “enhanced fine root decomposition” as cause of the *autumn* and *winter* soil respiration
649 increase under $e\text{CO}_2$ cannot be ruled out.

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654 ~~4.4 Temperature dependence of soil respiration~~

655 ~~We observed that the temperature dependence of soil respiration was different in autumn~~
656 ~~compared to other seasons, whereas $e\text{CO}_2$ did not change the relationship of soil respiration to~~
657 ~~temperature. During autumn, soil temperatures were at the same range as during spring and~~
658 ~~summer, but soil respiration was on average lower (Fig.5a). This pattern could reflect the~~
659 ~~higher proportion of root respiration (due to active root growth and assimilate allocation to~~
660 ~~exudates) during spring and summer, as observed by Lenhart (Lenhart, 2008). Boone et al.~~
661 ~~(1998) found a greater temperature sensitivity of root respiration than microbial respiration,~~
662 ~~whereas, Bååth et al. (2003) contradicted this finding in a microcosm experiment where~~
663 ~~different fractions of soil respiration had the same Q_{10} relationship. They suggested that the~~
664 ~~intensity of light, and thus the intensity of photosynthetic carbon gain and its availability for~~
665 ~~root derived soil respired C, may co-vary with temperature in field studies, probably~~
666 ~~explaining different temperature dependencies of soil respiration between seasons. In~~
667 ~~summary, the lack of a difference between ambient and $e\text{CO}_2$ soil respiration temperature~~
668 ~~functions suggests that there is no need to account for a special “ $e\text{CO}_2$ temperature sensitivity~~
669 ~~effect” in larger scale models of temperate grassland CO_2 exchange under future CO_2 -~~
670 ~~enriched atmospheres.~~

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671 4.5 N availabilty

672 Since soil microorganisms require C as well as N for maintenance and growth (De Graaff et
673 al., 2006; Zak et al., 1993), N availability plays an important role in determining soil CO_2
674 efflux. ~~Moreover, the build-up of stable humus compounds (C:N ratio of 10-11), as a potential~~
675 ~~negative (dampening) feedback of rising CO_2 atmospheres, requires sufficient quantities of N.~~
676 Root respiration rates were observed to correlate with tissue nitrogen concentration (Burton et

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677 al., 1996, 1998). ~~whereas nitrogen affected microbial respiration in a complex pattern (Magill~~
 678 ~~and Aber, 1998; Saiya-Cork et al., 2002; Ågren et al., 2001).~~ In the Giessen-FACE, $e\text{CO}_2$
 679 caused reduced tissue N concentrations and higher C:N-ratios of aboveground plant biomass
 680 (Kammann et al., 2008). ~~In line with these findings is the observation of Lenhart (2008), who~~
 681 ~~found a lower fraction of root-derived CO_2 on soil respiration with increasing atmospheric~~
 682 ~~CO_2 . Furthermore, $e\text{CO}_2$ induced a shift of available NO_3^- towards NH_4^+ at the study site~~
 683 ~~(Müller et al., 2009), a typical feature of N-limited ecosystems to retain mineral N (Rütting et~~
 684 ~~al., 2008; Huygens et al., 2008).~~ Through freezing effects in winter, mineral N, which was
 685 immobilized into the microbial biomass shortly after fertilizer application in spring, became
 686 partly available again (Müller et al., 2003). It is possible that N, as a limiting factor in the
 687 temperate grassland, may partly be responsible for the increase in soil C loss during the
 688 *autumn* and *winter* season under $e\text{CO}_2$.

689 4.6 Microbial community

690 Multiple observations from the GiFACE indicated that increases in winter soil respiration
 691 under $e\text{CO}_2$ were largely associated with microbial respiration (including rhizosphere
 692 microbiota). Recent studies from other FACE sites detected differences between microbial
 693 communities at $e\text{CO}_2$ compared to ambient CO_2 (Drigo et al., 2008; Drigo et al., 2009). At the
 694 GiFACE, stimulated rhizosphere-C utilization by arbuscular mycorrhizal fungi were found
 695 under $e\text{CO}_2$ by a ^{13}C -PLFA study (Denef et al., 2007), which may have contributed to altered
 696 soil respiration. Recent measurements in 2013 did not indicate any differences in the
 697 abundance of bacteria and archaea between the ambient and $e\text{CO}_2$ plots (K. Brenzinger,
 698 personal communication) so that this can be ruled out as a cause for differed soil respiration
 699 between the CO_2 treatments if this observation persists throughout *autumn* and *winter*.

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700 4.7 Soil moisture

701 Several studies showed that $e\text{CO}_2$ can affect soil moisture (Niklaus et al., 1998; Field et al.,
702 1995; Hungate et al., 1997), which in turn regulates soil respiration. However, large effects are
703 only expected and were detected at the dry end of the spectrum (Moyano et al., 2012; Guntinas
704 et al., 2013; Rodrigo et al., 1997). During the investigation period, the volumetric water content
705 ranged from 20 to 80 vol.% at the GiFACE site, with an average of 44% during 2008-2010,
706 and 39% over the vegetation periods of these years. Thus, ~~based on previous studies,~~ the soil
707 moisture effect is likely not to be large ~~(i.e. soil moisture was not the limiting factor).~~
708 ~~Therefore, we focused in our study on the soil temperature effect.~~ Moreover, ~~no significant~~
709 effect of $e\text{CO}_2$ on the soil water content was observed either during the first 5 years of
710 enrichment (Kammann et al., 2005) or after 13 years of enrichment (Meine, 2013).
711 Consequently, a CO_2 -induced soil moisture effect is unlikely governing increased soil
712 respiration rates, ~~but still, enhanced anaerobicity due to enhanced microbial activity, as~~
713 ~~experimentally produced e.g. by Sehy et al. (2004), cannot completely be ruled out. However~~
714 ~~any hypothetical aerobicity change, if present at all in the GiFACE, was not large enough to~~
715 ~~affect the performance and composition of the methanogenic community in the 11th year of~~
716 ~~CO_2 enrichment (Angel et al., 2012), which is a sensitive indicator for aerobicity changes.~~
717 However, it can be assumed that annual dynamics of soil moisture with wettest conditions in
718 winter, i.e. close to saturation, and driest conditions in summer (Fig. 2a) contributed to the
719 seasonal dynamics of soil respiration under $e\text{CO}_2$ due to diffusion limitations. ~~Analysis of~~
720 ~~stable isotopes revealed a distinctive $\delta^{13}\text{CO}_2$ gradient in soil during winter with decreasing~~
721 ~~signatures with depth but a homogenous $\delta^{13}\text{CO}_2$ profile during vegetation period at our study~~
722 ~~site (Lenhart, 2008). The absence of a $\delta^{13}\text{CO}_2$ gradient during summer was likely due to~~
723 ~~improved diffusive mixing of soil air in the profile during aerobic soil conditions. Previous~~

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724 results from the GiFACE site show that in periods when soil moisture in the main rooting
 725 zone was low ($0.3 \text{ m}^3 \text{ m}^{-3}$), soil continued to produce N_2O from deeper soil layers (20 – 50
 726 cm), where soil moisture remained high (c. $0.6 \text{ m}^3 \text{ m}^{-3}$) (Müller et al., 2004). Based on
 727 previous studies on this grassland (e.g. Müller et al., (2004) it was shown that during summer,
 728 when soil moisture content was relatively low ($0.3 \text{ cm}^3 \text{ cm}^{-3}$) in the main rooting zone (top 10
 729 cm) of the GiFACE site, the site of production for gaseous emissions (e.g. N_2O) occurred at
 730 deeper soil layers (20–50 cm depth) where the soil moisture content was still high (0.6 cm^3
 731 cm^{-3}). The production of N_2O at deep soil layers seemed to coincide with the production of
 732 CO_2 during summer, which was also characterized by a homogenous $\delta^{13}\text{CO}_2$ profile during
 733 vegetation period at our study site (Lenhart, 2008). However, a detailed investigation on
 734 layer-specific CO_2 production was beyond the scope of this study. At times of high soil
 735 moisture Accordingly, CO_2 diffusion was slowed down at times of high soil moisture,
 736 coinciding with limited oxygen supply (Skopp et al., 1990). At these times, soil respiration
 737 was likely originating to a major part from the topsoil. However, increased autumn soil
 738 respiration under $e\text{CO}_2$ cannot be attributed to this phenomenon since soil water content is
 739 relatively low at this season (Fig. 2a). We suggest that increased substrate supply under $e\text{CO}_2$
 740 from end-of-season dieback of roots and the enhanced root-associated microbiome activity
 741 may explain stimulated soil respiration rates in autumn.

742 4.8 Freeze/thaw cycles

743 Freeze/thaw cycles are known to mobilize previously inaccessible C and N substrates
 744 (Goodroad and Keeney, 1984; Kammann et al., 1998; Röver et al., 1998; Müller et al.,
 745 2002; Edwards and Cresser, 1992), providing substrates for heterotrophic activity. Frost events
 746 occurred during the study at the GiFACE from end of December 2008 to February 2009 (Fig.
 747 2e). The relative change of soil respiration under $e\text{CO}_2$ was 17 %, 12 % and 5 % from January

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748 ~~to March 2009 respectively (Fig. 2d), showing a more pronounced stimulation in these~~
749 ~~seasons than during the growing season, apart from October 2010 (12 % increase under~~
750 ~~$e\text{CO}_2$).~~

751 4.9 Plant community

752 Another aspect which may have contributed to altered soil respiration rates under $e\text{CO}_2$ is a
753 shift in the plant community composition. Grütters et al. (2006) observed that summer-greens
754 decreased, whereas evergreens increased under $e\text{CO}_2$ in the GiFACE experiment. Since soil
755 respiration is controlled by substrate supply via rhizodeposition (Verburg et al., 2004; Wan
756 and Luo, 2003; Craine et al., 1999), higher photosynthetic activity in $e\text{CO}_2$ plots during mild
757 *winter* may have contributed to the observed increase in soil respiration. In addition, since the
758 vegetative aboveground growth is dormant and does not provide an assimilate sink, the
759 relative proportion of assimilate partitioned below-ground towards the root-associated micro-
760 biota may increase, contributing to the relative increase under $e\text{CO}_2$ during *winter*. The higher
761 abundance of evergreens at $e\text{CO}_2$ also underlines the importance of a year-round CO_2
762 enrichment strategy in such ecosystems with the respective climatic conditions. To date,
763 increased winter soil respiration at $e\text{CO}_2$ was only found in FACE experiments with year-
764 round fumigation and a photosynthesizing at least partly green canopy, i.e. in the CLIMATE
765 study (Selsted et al., 2012) and in this study.

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770 5 Conclusions

771 | In conclusion, our results demonstrated the importance of winter soil respiration
772 | measurements, by showing that soil respiration was increased during *autumn* and *winter* after
773 | moderate long-term $e\text{CO}_2$. Measurements and year-round CO_2 enrichment should not be
774 | neglected, at least in winter-green temperate ecosystems. Studies in such ecosystems
775 | excluding measurements during the dormant season may thus underestimate the effect of
776 | $e\text{CO}_2$ on annual soil-respiratory CO_2 losses (i.e. leading to an overestimation of C
777 | sequestered). Consequently, winter soil CO_2 fluxes may play a crucial role in determining the
778 | carbon balance and dynamics of temperate grassland ecosystems. Our results indicate that
779 | temperate European grasslands which are characterized by a greenhouse gas balance near zero
780 | (Soussana et al., 2007) may gradually turn into greenhouse gas sources with rising
781 | atmospheric CO_2 due to enhanced CO_2 losses during *autumn* and *winter*, in particular if N_2O
782 | emissions are significantly increased as well as observed in the GiFACE (Kammann et al.,
783 | 2008; Regan et al., 2011).

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784 | To generalize and explain the variation in the temporal dynamics of soil respiration under
785 | $e\text{CO}_2$ more studies of winter C dynamics under long-term $e\text{CO}_2$ are required. For such future
786 | studies it is advisable to include frequent samplings of root biomass, including the fine root
787 | fraction and necromass, in particular during the *autumn/winter* period under $e\text{CO}_2$. Another
788 | beneficial research strategy may be combined (pulse) labelling of ^{15}N and ^{13}C to elucidate
789 | gross C and N turnover processes after long-term (>10 years) of CO_2 enrichment to study the
790 | C-N gross dynamics and associated carbonaceous gas losses.

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1059 **Tables**

1060 **Table 1**

1061 Results of fitting the temperature-dependence model after Lloyd and Taylor (Lloyd and
1062 Taylor, 1994) to 20% of our observation data under ambient and elevated CO₂.

| CO ₂ treatment | R | R²_{seq} | <u>Adjusted</u> R²_{seq} | Standard Error of Estimate |
|---------------------------|------|--|---|----------------------------------|
| Ambient CO ₂ | 0.87 | 0.75 | 0.75 | 1.35 |
| Elevated CO ₂ | 0.91 | 0.82 | 0.82 | 1.19 |

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Formatierte Tabelle

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Table 2
 Annual sums of soil respiration under ambient and eCO₂ from 2008 – 2010. Data are presented as averages (n=3) ± standard error (SE). P-values indicate the difference between treatments per year obtained by a paired t-test.

| Year | CO ₂ treatment | Mean annual sum of soil respiration (g CO ₂ m ⁻² yr ⁻¹) | Mean annual sum of soil respiration (g C[CO ₂] m ⁻² yr ⁻¹) | Relative change to control (%) | P value |
|------|---------------------------|---|---|--------------------------------|---------|
| 2008 | Ambient CO ₂ | 485 43.93 ± 343.84 | 132 43.80 ± 9.23 | 1.22 | 0.17 |
| | Elevated CO ₂ | 4913.38 ± 14.20 | 1340.01 ± 43.87 | | |
| 2009 | Ambient CO ₂ | 4928.00 ± 48.34 | 1344.00 ± 13.18 | 0.56 | 0.64 |
| | Elevated CO ₂ | 495 65.74 ± 39.08 | 135 21.56 ± 110.66 | | |
| 2010 | Ambient CO ₂ | 4702.44 ± 376.69 | 128 32.48 ± 10.01 | 1.38 | 0.23 |
| | Elevated CO ₂ | 4767.22 ± 121.47 | 1300.15 ± 3.13 | | |

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1110 **Figure legends**

1111 **Fig. 1** Seasonal patterns and the five defined seasons at the GiFACE grassland study site.

1112 **Fig. 2** Volumetric water content under ambient and elevated CO₂ (a), daily sums of
1113 precipitation at the GiFACE (b), mean soil temperature during soil respiration measurements
1114 and minimum daily soil temperature at 10 cm depth (c), the relative mean monthly change of
1115 soil respiration under elevated CO₂ based on ~~observed-measured~~ and modelled data (d), the
1116 absolute mean monthly difference in soil respiration under elevated CO₂ based on ~~observed~~
1117 ~~measured~~ and modelled data (e), measured soil respiration under ambient and elevated CO₂
1118 from 2008 to 2010 (f) and modelled soil respiration under ambient and elevated CO₂ per
1119 measurement from 2008 to 2010 based on observed and modelled data (gf). Data are
1120 presented as averages (n=3) ± 1 SE.

1121 **Fig. 3** Mean soil respiration rates during the five defined seasons under ambient and elevated
1122 CO₂ averaged over three years from 2008 – 2010. Error bars show ± 1 SE associated by
1123 averaging across the three replicates per treatment (n=3) (a); (1) = winter dormancy; (2) =
1124 start of vegetation period; (3) = spring; (4) = summer; (5) = autumn (for details see methods).
1125 P-values indicate the difference between treatments obtained by a linear mixed-effect model
1126 analysis.

1127 **Fig. 4** Relationship between soil respiration rate and soil temperature under ambient and
1128 elevated CO₂ (a) and temperature dependence of soil respiration under ambient and elevated
1129 CO₂ during different seasons (b). Equation of dynamic fit (Lloyd and Taylor, 1994): $f =$
1130 $R10e^{E0\left(\frac{1}{(283.15-T0)} - \frac{1}{(x-T0)}\right)}$
1131

1132 ~~Observed versus modelled soil respiration rates under ambient and elevated CO₂.~~

1133 **Fig. 5** Observed versus modelled soil respiration rates under ambient and elevated CO₂.

1134 Relationship between soil respiration rate and soil temperature under ambient and elevated
1135 CO₂ (a) and temperature dependence of soil respiration under ambient and elevated CO₂
1136 during different seasons (b). Equation of dynamic fit (Lloyd and Taylor, 1994): $f =$
1137 $R10e^{E0\left(\frac{1}{(283.15-T0)} - \frac{1}{(x-T0)}\right)}$
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1139 **Fig. 6** Annual sums of soil respiration under ambient and elevated CO₂ for 2008—2010 based
1140 on observed and modelled data. Error bars represent ± 1 SE of the mean.

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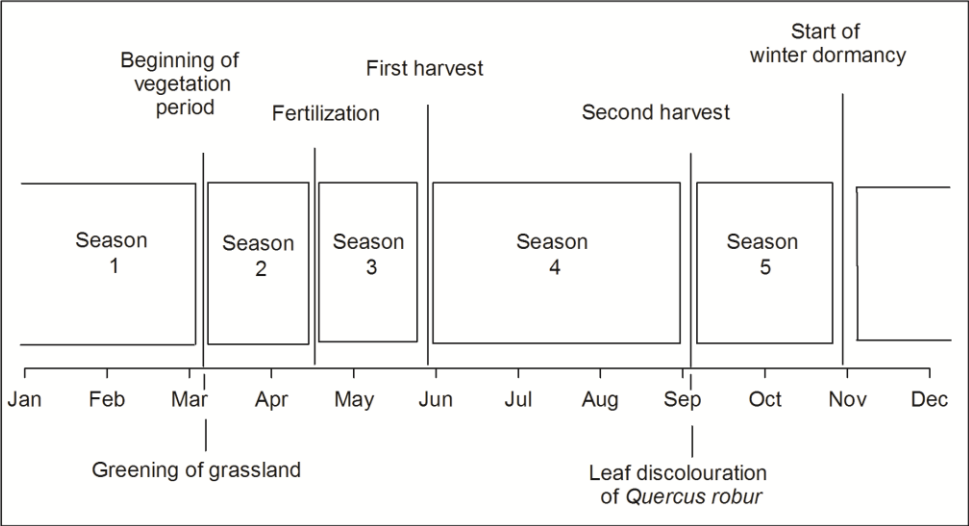
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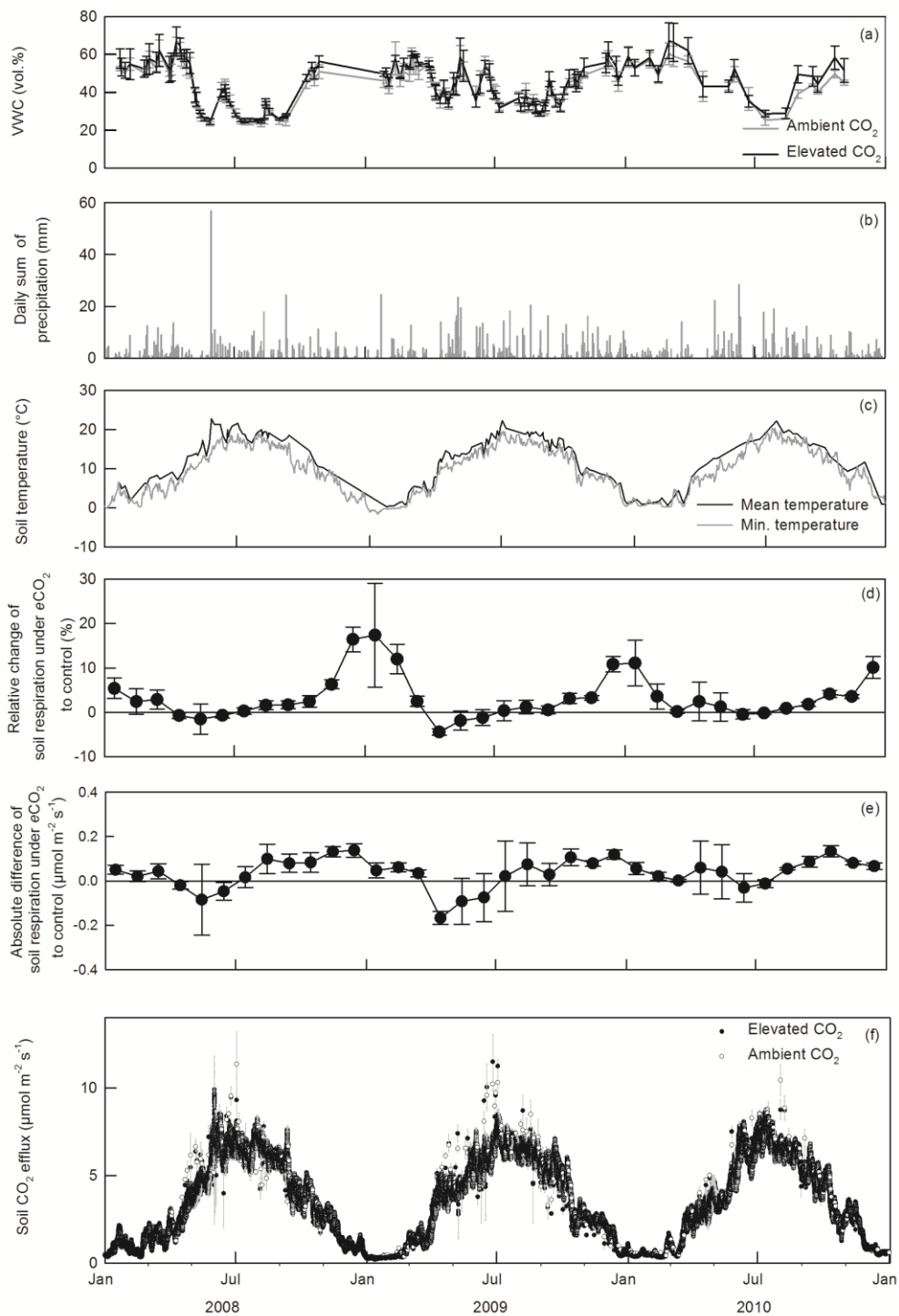
1145 **Figures**

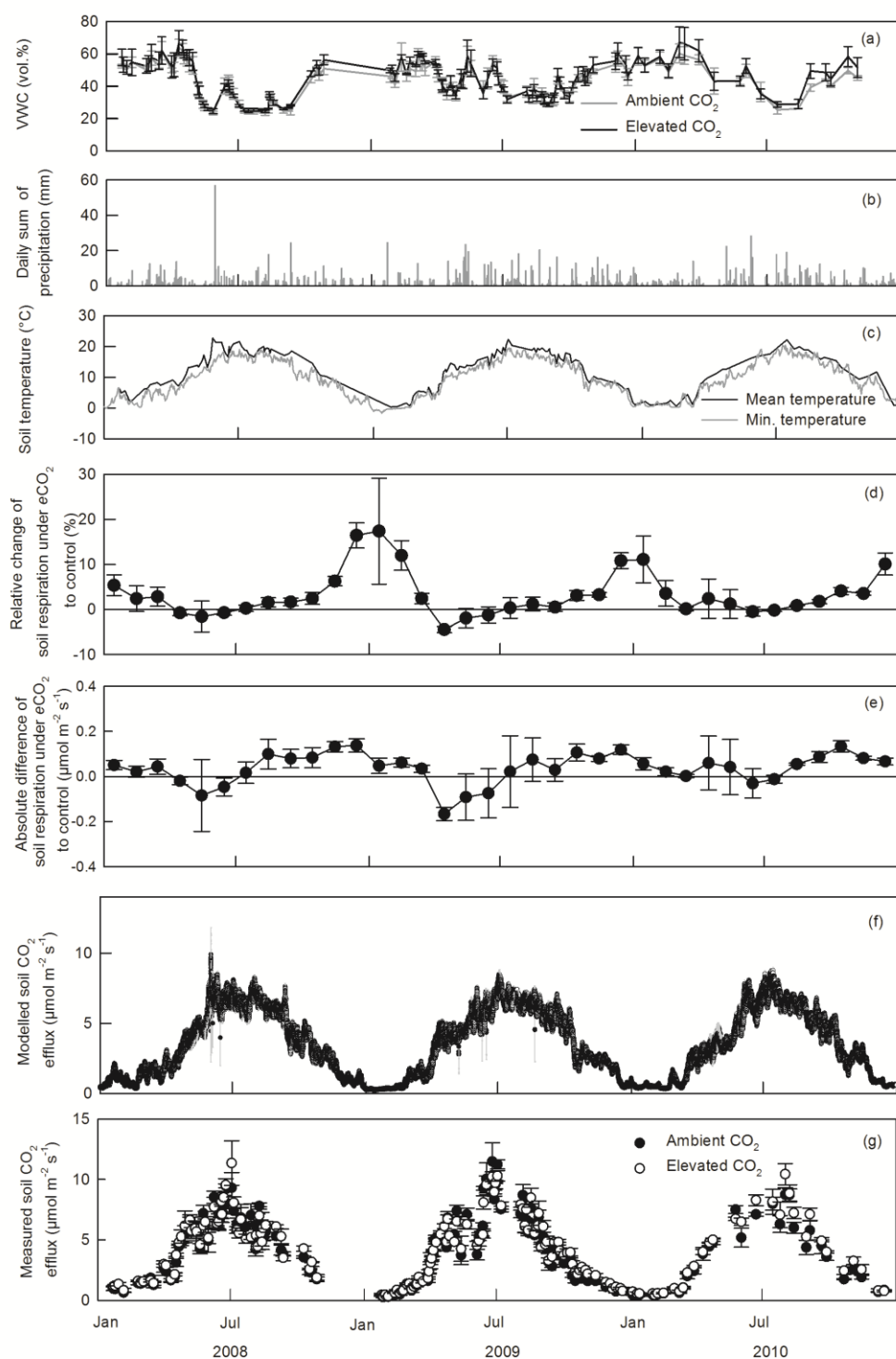
1146 Fig. 1



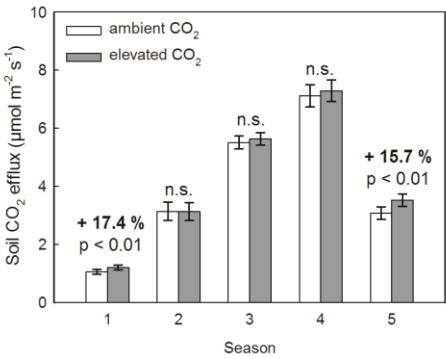
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1159 Fig. 2



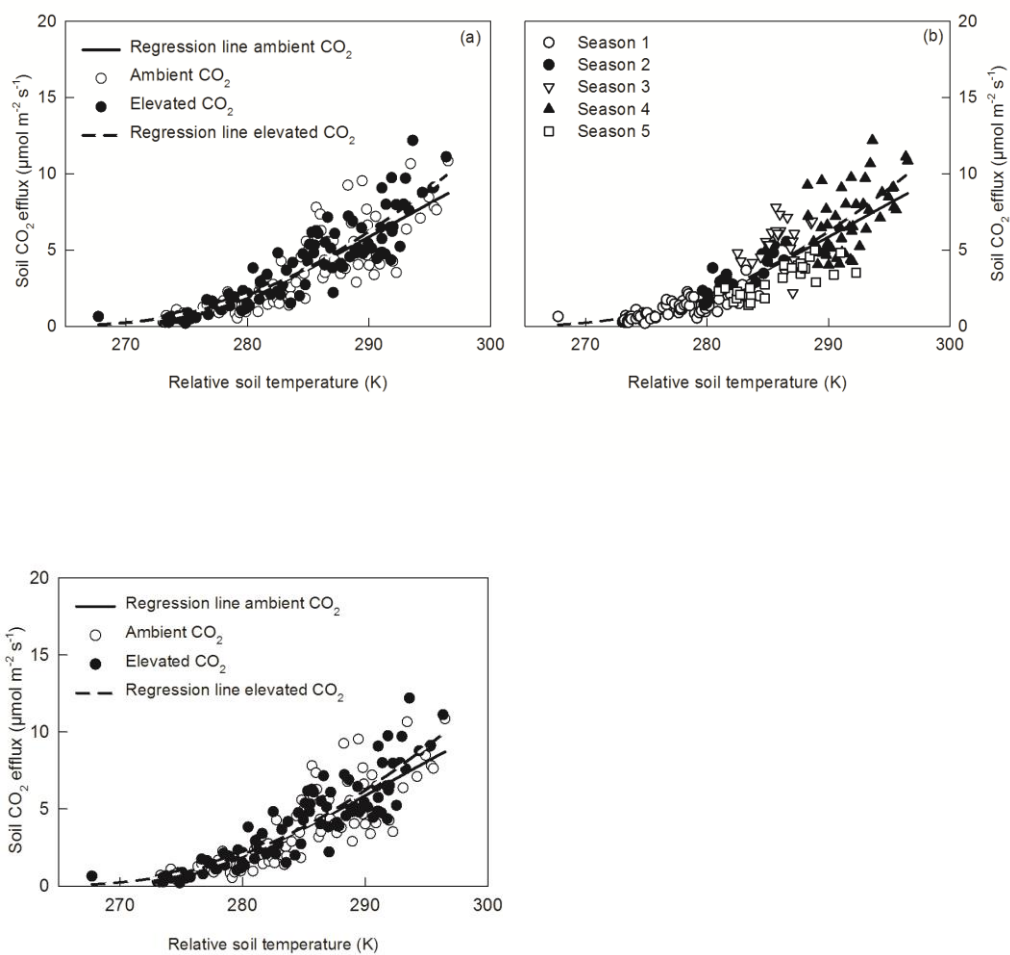


1162 Fig. 3

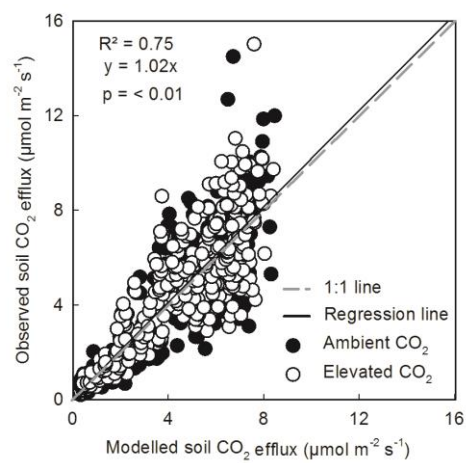
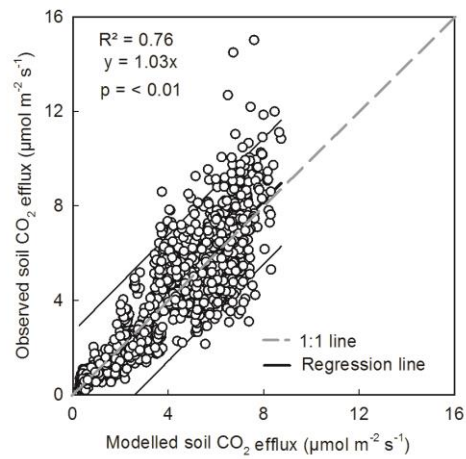


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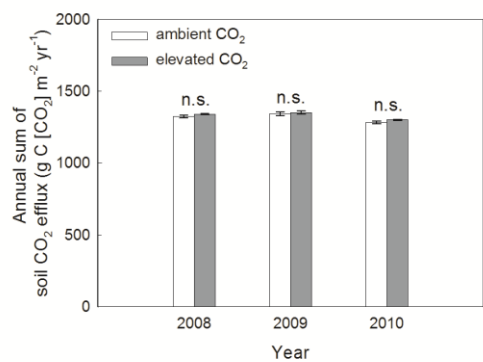
1180 Fig. 4



1189 Fig. 5

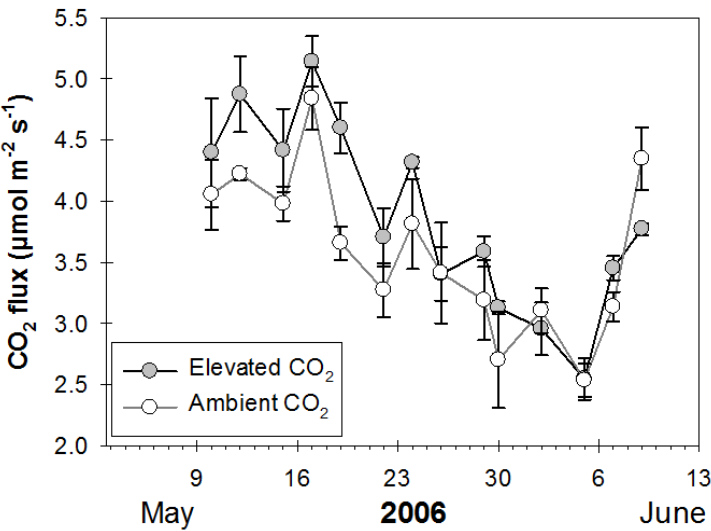


1198 **Fig. 6**



1216 **Supporting Information**

1217 **Fig. S1**



1218
1219 **Fig S1:** Mean CO₂ efflux +/- standard error (n=3) after installation of the frames and removal
1220 of the aboveground biomass on 9th May 2006.

1221
1222 On 11 out of 14 measurement occasions all three E-plot fluxes were higher than those of
1223 their corresponding A-plot partner. A mixed Model analysis (SPSS version 18) with the
1224 factors CO₂-treatment and time revealed that the soil CO₂ efflux was significantly increased
1225 by CO₂ enrichment.

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1235 **Table S1**
1236 Parameter estimates of the temperature-dependence model after Lloyd and Taylor (1994)

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| CO ₂ treatment | Model parameter | Coefficient | P value |
|---------------------------|-----------------|-----------------|---------|
| Ambient CO ₂ | E0 | 61.92 ± 33.59 | 0.07 |
| | R10 | 3.00 ± 0.19 | < 0.001 |
| | T0 | 261.18 ± 6.53 | < 0.001 |
| Elevated CO ₂ | E0 | 143.68 ± 103.57 | 0.17 |
| | R10 | 3.11 ± 0.17 | < 0.001 |
| | T0 | 248.72 ± 13.35 | < 0.001 |

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