

Reply to the comments of Anonymous Referee #1 with regard to the performance of the dynamic flux chamber

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The main concern of the reviewer is that the chamber design may lead to biased fluxes. Here we provide a more extensive description of the chamber system and show atmospheric data obtained during the campaign supporting the flux dynamics observed with our chamber. We would like to point out that the main focus of this manuscript are i.) to show the strong imprint of the tides on trace gas fluxes in coastal ecosystems and ii.) to highlight the importance of accurately addressing the perturbations of turbulent flows in flux chamber studies. Both points may have severe implications for our understanding of trace gas dynamics in coastal ecosystems. Regarding the magnitude of the fluxes during submersion we agree with the reviewer and have clearly stated this in the manuscript on page 10587, line 9-12: *“As the strength of advection in our chamber system relative to ambient conditions is unknown we can currently not appraise the quality and reliability of the difference chamber systems”*.

Atmospheric mixing ratios of CO₂ and CH₄

In particular during periods of low wind speeds and during easterly winds the atmospheric mixing ratios of CH₄ and CO₂ support the observed tidal imprint on the trace gas fluxes (fig. 1 and 2). Throughout the campaign the atmospheric mixing ratios of CO₂ (average from both heights) ranged from 395.5 to 429.7 ppm (both heights) and averaged 400.3 ppm. The atmospheric mixing ratios of CH₄ ranged from 1.831 to 1.895 ppm (both heights) and averaged 1.861 ppm. Lowest mixing ratios of 395.8±0.2 ppm for CO₂ and of 1.834±0.004 ppm for CH₄ were observed between 8:00 pm on April 25th and 4:00 am on April 26th and coincided with westerly winds from the Open Ocean and wind speeds above 4m/s. With decreasing wind speeds and during easterly winds, when the air masses passed over large parts of the lagoon the atmospheric mixing ratios of CO₂ and CH₄ increased.

The close coupling between the measured fluxes and the atmospheric mixing ratios at low wind speeds becomes in particular evident at the end of the campaign. Over the last two tidal

cycles the atmospheric mixing ratios of CH₄ nicely resemble the enhanced emissions during immersion. The sharp methane emission peak observed when the water entered the chamber becomes diffuse under ambient conditions as bubble ebullition will occur throughout rising tide. On April 27th this coupling is somewhat confounded because of rapidly changing wind conditions. Nevertheless, elevated CH₄ mixing ratios coincide with elevated fluxes during tidal immersion. As for CH₄ elevated mixing ratios of atmospheric CO₂ coincide with periods of strong CO₂ emissions during tidal immersion at night. Notably on April 26th at noon the atmospheric CO₂ mixing ratios show a slight drop when carbon assimilation was largest. In conclusion the pattern of the atmospheric mixing ratios support the flux pattern observed with the chamber.

Operation Principle of the flux chamber

A During air exposure

During air exposure the chamber operates as a conventional dynamic flux chamber (DFC). Briefly, the surface of interest is enclosed with a chamber and air is drawn through the chamber at a predefined flow rate. Net fluxes above the covered surface are calculated from the concentration difference between the inlet and outlet of the chamber assuming steady state conditions:

$$F_{net} = \frac{Q_n}{A} (c_{out} - c_{in}) \quad (1)$$

where F_{Net} is the net flux [$M L^{-2} t^{-1}$], Q_N is the flushing flow rate through the chamber [$L^3 t^{-1}$, at 1013.25 mbar and 298.15 K], c_{out} and c_{in} are the concentration of target compounds [$M L^{-3}$] at the outlet and the inlet of the flux chamber and A is the bottom surface area of the flux chamber [L^2]. M, L, and t are dimensions of mass, length, and time, respectively.

A basic assumption of DFC-measurements is that the chamber is in steady state meaning that the time averaged net flux calculated from the inlet and outlet concentrations equals the time averaged net flux across the interface of interest. A convenient measure for the time required to establish steady state conditions is the response time (t_r) of the chamber denoting the time after disturbance (such as a sudden change in emission) required to achieve 37% ($e^{-1} \times 100$) of the initial perturbation. Provided that the change after a perturbation is exponential 95% of steady state will be reached after three times the response time. Under aerial conditions the response time can be estimated from the flushing flow rate and the chamber volume and is given by

$$t_r = \frac{V}{Q} \quad (2)$$

At a typical flushing flow rate of 3 L min⁻¹ the response time of our chamber system is 2 minutes. Pre-tests revealed that complete mixing of the chamber volume is achieved within 0.4 min. Hence with respect to our sampling frequency we can safely assume complete mixing of the air inside the chamber.

The physical nature of trace gas fluxes across natural interfaces is commonly described in terms of multiresistance models Gao et al., 1997; Zhang et al., 2002):

$$F_i = (c_s - c_{ag}) \times h_i = \frac{c_s - c_a}{R_i} \quad (3)$$

Where F_i denotes the flux across the interface, c_s is the concentration in the sediment, c_g is the gas concentration on the air side of the interface h_i [L t⁻¹] is the overall exchange coefficient and R_i [t L⁻¹], the reciprocal of h_i , is the overall transport resistance. The overall transfer resistance is given by the sum of the transfer resistance of the quasi laminar boundary layer above the sediment (R_a) and the diffusive resistance of the sediment surface layer (R_s)

$$R_i = R_a + R_s \quad (4)$$

R_a is dependent on the aerodynamic properties of the chamber whereas R_s is dependent on the sediment properties. The sensitivity of the overall flux against the aerodynamic properties depends on the magnitude R_a and R_s . When both share the same magnitude the flux across the interface depends on R_a and R_s . On the other hand, when R_s becomes large relative to R_a the flux is governed by R_s and not affected by the aerodynamic properties of the chamber.

In order to determine R_a for our flux chamber we have determined the CO₂ flux above 1N NaOH. 4 L of 1N NaOH were carefully filled into a 10L Duran glass bottle avoiding wetting of the bottle walls and leaving a headspace of 6L being comparable to the headspace of the field chamber. The NaOH solution was gently stirred (50 RPM). Ambient air (407±3 ppm CO₂) was pumped through the chamber at flushing flow rates between 0.5 and 5 L min⁻¹ and the concentration was measured at the outlet of the chamber. Under the assumption that the NaOH solution provides a perfect sink for CO₂ R_s becomes negligible and R_a can be estimated according to:

$$R_a = \frac{-c_{out}}{c_{out}-c_{in}} \times \frac{A}{Q} \quad (5)$$

From 0.5 to 4 Lmin⁻¹ the uptake flux increased logarithmically from 16.4 to 57.4 mmol m²h⁻¹ ($F = 19.224 \cdot \ln(Q) + 29.697$; $R^2 = 0.9967$) and remained on this level with a further increased flushing flow rate of 5 L min⁻¹. Surprisingly R_a increased with increasing flushing flowrate from 0.162 hm⁻¹ ($Q = 0.5$ L min⁻¹) to 0.206 h m⁻¹ at $Q = 5.0$ L min⁻¹. We attribute this increase to an insufficient renewal of the liquid surface film limiting the CO₂ uptake at high CO₂ loads. However, these results provide an upper estimate of 0.162 hm⁻¹ for R_a . The sediment side transfer resistance is commonly computed from the diffusivity of the sediment surface layer and its thickness (Gao et al., 1997; Zhang et al., 2002).

$$R_s = \frac{l}{0.6D_0p_s} \quad (6)$$

Where l [L] denotes the thickness of the sediment surface layer, 0.6 is an empirical scaling factor accounting for the tortuosity of the pores (Penman 1940 a,b), D_0 [L² t⁻¹] is the diffusion coefficient in air and p_s is the fraction of the air filled pore space. Following Zhang et al. (2002) we used a thickness of 0.005 m to estimate R_s . With $D_0(\text{CO}_2) = 0.054$ m² h⁻¹ and p_s ranging from 0.01 to 0.1 for water logged intertidal sediments R_s is estimated to range from 1.54 to 15.4 h m⁻¹ being about one to two orders of magnitude larger than R_a . Given this it is reasonable to assume that the trace gas flux across the sediment surface is mainly governed by the sediment resistance. Further our tests suggest a minor effect of the flushing flow rate on the atmospheric transfer resistance making the overall transfer resistance insensitive against the aerodynamic properties of the chamber.

The same principles apply to the trace gas exchange across the sea grass leaves. In contrast to higher plants on land seagrass lacks any stomata and gas transport is facilitated across the thin cuticle of the seagrass. MacFarlane (1992) reports a CO₂ permeability of 2.1 μm s⁻¹ for the cuticula of submersed plants. This corresponds to a transfer resistance of 132 h m⁻¹. Using a leaf area index of 3 - 5 for *Z. noltii* in the Ria Formosa (Pérez-Lloréns & Niell, 1993) we can normalize the transfer resistance to the enclosed surface area and estimate its range to 26.5 to 46 h m⁻¹. In summary we can safely assume that during air exposure the gas exchange across the sediment surface and the seagrass leaves is not dependent on the aerodynamic properties of the chamber.

The measured trace gas fluxes may be affected by the enclosure effect (i.e. the deviation of the internal trace gas concentration from ambient concentrations above the ground). The

effect of the flushing flow rate depends on the overall transfer resistance. When assuming a constant and uniform source (or sink) the effect of the flushing flow rate under steady state conditions is given by (eq. 22. in Zhang et al., 2002):

$$F_{steady} = \frac{c_s - c_i}{R_i} \times \frac{1}{1 + \frac{A}{R_i \times Q}} \quad (7)$$

Thus the enclosure effect is given by:

$$\frac{F_{steady}}{F_{limQ \rightarrow \infty}} = \frac{1}{1 + \frac{A}{R_i \times Q}} \quad (8)$$

With a bottom surface area of 0.037 m² and an overall transfer resistance ranging from 1.62 to 15.6 h m⁻¹ the steady state flux at a flushing flow rate of 3 L min⁻¹ [0.18 m³ h⁻¹] resembles more than 96% of the maximum flux ($F_{limQ \rightarrow \infty}$) suggesting a negligible effect of the flushing flow rate.

B During submersion

Under submerged conditions we have to consider an additional transfer resistance for the air water transfer of the dissolved trace gases. Thus we will start our discussion with this additional transfer resistance. The dissolved trace gas is stripped out and the flux will depend on its volatility (given by the inverse Henry's law constant) and the water air transfer resistance of the chamber system. In analogy to the air sea gas exchange the flux of a trace gas can be computed as:

$$F = k_{ch} \times (c_w \times H_{inv} - c_g) = \frac{(c_w \times H_{inv} - c_g)}{R_{ch}} \quad (9)$$

where k_{ch} is the gas exchange velocity [L t⁻¹] depending on the flushing flow rate (Q) and the chamber design (in particular the chamber geometry and the gas bubble geometry), $R_{ch} = 1/k_{ch}$ is the corresponding transfer resistance, c_w is the water concentration [m L⁻³], c_g is the concentration in the gas phase inside the chamber, and H_{inv} is the inverse dimensionless Henry's law constant. Under steady state conditions c_g equals c_{out} and F equals F_{net} . Combination of eq. 9 with eq.1 leads to:

$$\frac{(c_w \times H_{inv} - c_{out})}{R_{ch}} = (c_{out} - c_{in}) \frac{Q}{A} \quad (10)$$

and rearrangement leads to:

$$R_{ch} = \frac{(c_w \times H_{inv} - c_{out}) \times A}{(c_{out} - c_{in}) \times Q} \quad (11)$$

R_{ch} was determined in a similar manner as R_a . To mimic submersed conditions the chamber was filled with 1N NaOH. Then ambient air was pumped through the chamber with flushing flow rates ranging from 0.5 to 6 L/Min. Due to the rapid transformation of dissolved CO_2 to bicarbonate, $c_w \times H_{inv}$ is assumed to become negligible. Thus R_{ch} can be computed from eq. 5.

$$R_{ch} = \frac{-c_{out} \times A}{(c_{out} - c_{in}) \times Q} \quad (5)$$

R_{ch} averaged $0.031 \pm 0.006 \text{ h m}^{-1}$ and showed a slight decrease with increasing flushing flow rate that may be attributed to the purge gas dispersion inside the chamber. During submersion the interfacial resistance for diffusive transport across the sediment/water and water/leaf interface will be larger than during air exposure (1.54 to 15.4 h m^{-1} and 26.5 to 46 h m^{-1}) because of the lower diffusivity of trace gases in water. Thus as for aerial conditions we can assume that diffusive fluxes are primarily governed by the interfacial resistance while the transfer resistance of the chamber R_{ch} has a minor influence on diffusive trace gas fluxes. Our field data clearly show that during submersion the sediment/water flux is governed by advective transport. A treatment of these processes is beyond the scope of this paper as it would require detailed hydrodynamic data of the chamber system. On the other hand the fluxes across the water leaf interface are clearly governed by diffusion. Thus we limit the following discussion to diffusive exchange processes.

Under submersed conditions the response time of the chamber will apparently depend on the solubility of the trace gases and in case of CO_2 also on the carbonate system. The response time of the submersed chamber was derived from the slope of concentration over time at the outlet of the chamber after injection of seawater enriched in methane and CO_2 respectively. For these tests artificial seawater with 35 PSU was prepared from commercially available sea salt. Prior to the experiments the pH was adjusted to 8.1 ± 0.1 . DIC was adjusted to yield an atmospheric equilibrium concentration of 350 ± 20 ppm. The chamber was continuously purged either with synthetic air (Westfalen, Germany) or synthetic air doped with CO_2 (Westfalen Germany) providing a pCO_2 of 350 ± 20 ppm in the purge gas. The purge flow was set to 3.0 L min^{-1} . 50 ml of seawater were withdrawn from the chamber and were equilibrated with either methane (30% in synthetic air) or CO_2 (30% in synthetic air) for 24h at 1.3 bar and 4°C . These solutions were injected into the chamber and the decline in pCH_4 and pCO_2 was measured over time. A second set of experiments was carried out to monitor the CO_2 uptake. Prior to these experiments the chamber was flushed with pure synthetic air for several hours

to lower the DIC inside the chamber. During the CO₂ measurements the pH inside the chamber was measured with a Schott CG 841 pH meter. The evolution of the CO₂ outlet concentration and the CO₂ flux is exemplified for CO₂ uptake in figure 3. Complete mixing of the water inside the chamber as indicated by the pH was achieved within 1.10 minutes. The response time of the chamber towards changes in the pCH₄ was 1.20± 0.20 min and the response time of the chamber towards changes in the DIC system was dependent on the pH, reflecting the residence time of CO₂ in water with respect to hydration and reaction with OH⁻, and ranged from 49 to 67 min. with pH ranging from 7.80 to 8.40. Thus we can estimate the time for re-establishing a steady state to 3.60±1.0 min. for CH₄ and to about 170±30 min. for DIC. When assuming a constant source / sink inside the chamber and a response time of 60 min., 37% of the source / sink will be replenished after 1h and 80% after 6h of incubation. Thus our dynamic chamber can in principal capture changes in the DIC pool over longer incubations times but will smooth the initial signal.

During tidal inundation the chamber is connected to the surrounding waters via U-tube at the bottom (stainless steel tube 50 cm length, 4 mm i.d.) to assure pressure equilibrium. This inevitably leads to an exchange of water between the chamber and the surrounding water body that may thus affect the flux measurements. The flux through the U-tube depends on the pressure gradient between the chamber and the surrounding water and is thus mainly modulated by wave actions. The flux measurements were carried out under calm conditions with wind speeds rarely exceeding 5 ms⁻¹ and wave heights below 10cm. For a pressure difference of 1 kPa is the measured flow was 308±7 ml min⁻¹. This is about 46% of the theoretical flow expected from Hagen-Poiseilles law reflecting the narrowing of the tube at the bendings. The wave induced pressure oscillations cause an oscillating flow between the chamber and the surrounding water body. For sinusoidal waves with an amplitude of 5cm the water exchange rate is 46 ml min⁻¹ resulting in a response time towards water exchange of 2.15h being substantially larger than the respective response times for the gas exchange. For CH₄ we can safely assume that the bias due to water exchange is less than 1%, regardless of the concentration difference between the chamber and the surrounding water. For CO₂ the bias only is less than 3%. Due to the much slower response of the carbonate system the bias with respect to DIC may become substantially larger. From the respective response times we can estimate an upper limit for the bias of 32±3% e.g. when the chamber exchanges with DIC free water. Benthic respiration and assimilation will shift both DIC pools into the same direction. Due to the enclosure effect the shift is more pronounced inside the chamber. In typical coastal waters (Alkalinity = 2200 µeq, pH (8.185, DIC = 1999 µmol, pCO₂ = 380

ppm) a change of 1% in the DIC changes the pCO₂ by about 10%. As a consequence, gas exchange is more sensitive to changes in DIC than the water exchange. For a difference in the DIC concentrations of 20% we have carefully estimated the bias to be in the range of 12 to 18%. We found this acceptable for a first tentative assessment of the DIC dynamics over full tidal cycles as was the primary goal of our study.

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