

Dear Editor,

Thank you very much for your email with regard to our manuscript (**bg-2014-523R1**) together with the comments from the reviewers. The comments from the editors and reviewers were very helpful and we agree that the previous version needed revision. We take all of these comments into account in preparing the revised manuscript. We believe that manuscript has been improved satisfactorily and hopes it will be accepted for publication in **Biogeosciences**.

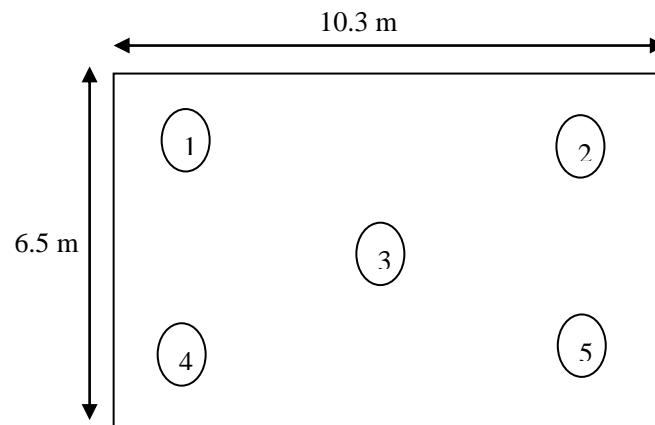
We thank again the reviewer for the helpful comments. Should you require any further information, please do not hesitate to ask.

To the Comments of Reviewer

QI-1: P6, L143-145: Please explain the method for measuring spatial variation in detail. For example, what time did you do the measurement and how long did you take for the measurement at 5 locations? Also, please explain how seriously the temporal fluctuation during the measurement affected on the estimation of spatial variation.

R: Five PVC collars were installed in our plots (attached pictures) for investigating the spatial variation of SOC mineralization rate in summer (July 11, 2008) and winter (November 18, 2008) (**Page 6, Line 140-142**). Although SOC mineralization rate highly varied with seasons, **the spatial variation of SOC mineralization rate is small enough with a variation coefficient of only 4% and 5% in summer and winter**, respectively (Table 1) (**page 20, line 497-500**). This implies that the temporal fluctuation during the

measurement have little effect on the spatial variation of SOC mineralization rate (**Page 7, Line 147-148**). This result is ascribed to the following season: 1) no vegetation or fresh C inputs in the plots since 1984 (absolute fallow); 2) the soil was derived aeolian deposit loess with relative uniform soil texture; and 3) flat terrain (**Page 7, Line 148-151**). Meanwhile, due to the small areas of our plots (66.95 m^2) and time constraints (5 min for measuring SOC mineralization rate in a given PVC collar), only twice measures was conducted in summer (July 11, 2008) and winter (November 18, 2008) for studying the spatial variation of SOC mineralization rate in our plots (**Page 7, Line 151-155**).



Attached pictures. The location of PVC collar in our plots

Table 1. SOC mineralization rate ($\mu \text{ mol m}^{-2} \text{ s}^{-1}$) in summer (July 11, 2008) and winter (November 18, 2008). Data are represented as mean \pm S.D of five collars (**page 20, line 497-500**).

Dates	SOC mineralization rate					Mean value
	Collar 1	Collar 2	Collar 3	Collar 4	Collar 5	
Summer	1.55 \pm 0.11	1.60 \pm 0.20	1.58 \pm 0.21	1.49 \pm 0.07	1.65 \pm 0.18	1.57 \pm 0.06
Winter	0.29 \pm 0.01	0.30 \pm 0.02	0.31 \pm 0.01	0.32 \pm 0.02	0.33 \pm 0.02	0.31 \pm 0.02

1 **Soil moisture influenced the interannual variation in**
2 **temperature sensitivity of soil organic carbon mineralization**
3 **in the Loess Plateau**

4

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25 **Abstract**

26 Temperature sensitivity of soil organic carbon (SOC) mineralization (i.e., Q_{10})
27 determines how strong the feedback from global warming may be on the atmospheric
28 CO₂ concentration, thus understanding the factors influencing the interannual
29 variation in Q_{10} is important to accurately estimate local soil carbon cycle. *In situ* SOC
30 mineralization rate was measured using an automated CO₂ flux system (Li-8100) in
31 long-term bare fallow soil in the Loess Plateau (35°12' N, 107°40' E) in Changwu,
32 Shaanxi, China from 2008 to 2013. The results showed that the annual cumulative
33 SOC mineralization ranged from 226 to 298 g C m⁻² y⁻¹, with a mean of 253 g C m⁻²
34 y⁻¹ and a CV of 13%, annual Q_{10} ranged from 1.48 to 1.94, with a mean of 1.70 and a
35 CV of 10%, and annual soil moisture content ranged from 38.6 to 50.7% soil
36 water-filled pore space (WFPS), with a mean of 43.8% WFPS and a CV of 11%,
37 which were mainly affected by the frequency and distribution of precipitation. Annual
38 Q_{10} showed a quadratic correlation with annual mean soil moisture content. In
39 conclusion, understanding of the relationships between interannual variation in Q_{10} ,
40 soil moisture and precipitation are important to accurately estimate the local carbon
41 cycle, especially under the changing climate.

42

43

44 **Keywords:** Soil temperature; SOC mineralization; distribution and frequency of
45 precipitation.

46

47 **1. Introduction**

48 Temperature sensitivity of soil organic carbon (SOC) mineralization (hereafter
49 refer to as Q_{10}) is of critical importance because it determines how strong the
50 feedback from global warming may be on the atmospheric CO₂ concentration (Ågren
51 and Wetterstedt, 2007). However, this is an issue of considerable debatable (Davidson
52 et al., 2006; Kirschbaum, 2006), and the variations in Q_{10} are the main source of
53 controversies in this feedback intensity (Larionova et al., 2007; Karhu et al., 2010;
54 Conant et al., 2011; Sakurai et al., 2012). Therefore, understanding the factors
55 influencing Q_{10} is important to accurately estimate C cycle and the feedback from the
56 expected warmer climate.

57 Previous studies have shown that Q_{10} variations are closely related to soil
58 temperature (Kirschbaum, 2006; Von Lutzow and Kogel-Knabner, 2009), substrate
59 availability (Ågren and Wetterstedt, 2007; Gershenson et al., 2009), substrate quality
60 (Von Lutzow and Kogel-Knabner, 2009; Sakurai et al., 2012), and the size and
61 composition of microbial population (Djukic et al., 2010; Karhu et al., 2010). Soil
62 moisture is the most significant limiting factor for underground physiological
63 processes in dry and semi-dry ecosystems (Balogh et al., 2011; Cable et al., 2011;
64 Wang et al., 2014). Soil water availability may indirectly affect Q_{10} by influencing the
65 diffusion of substrates, because the diffusion of extracellular enzymes produced by
66 microorganisms and available substrates must conduct in the liquid phase (Davidson
67 et al., 1998; Illeris et al., 2004), but the response of Q_{10} to soil water availability is
68 extremely complex and controversial (Davidson et al., 2000; Davidson et al., 2006;
69 McCulley et al., 2007). For example, Gulledge and Schimel (2000) found that Q_{10} was
70 larger in wet years than in drought years, whereas the opposite result was found by
71 Dorr and Mdnich (1987). However, many other studies that mainly focused on the

72 short-term or seasonal variation in Q_{10} (Davidson et al., 2006) have showed that Q_{10}
73 was not affected by soil moisture (Fang and Moncrieff, 2001; Reichstein et al., 2002;
74 Jassal et al., 2008). Additionally, soil water availability experienced marked seasonal
75 and interannual fluctuations in these ecosystems due to uneven rainfall distribution
76 caused by the abnormal increase of atmospheric CO₂ concentrations (Solomon et al.,
77 2007). The uneven rainfall distribution inevitably influenced soil moisture availability
78 (Coronato and Bertiller, 1996; Qiu et al., 2001; Cho and Choi, 2014). Xiao et al.
79 (2014) have shown that the interannual changes in soil moisture storage in the Loess
80 Plateau were decided by the difference in soil moisture storage between October and
81 April, because precipitation from April to October of 2004 to 2010 accounted for at
82 least 86% of annual rainfall. However, to our knowledge, there have been few studies
83 investigating the relationship between interannual variation in Q_{10} and soil moisture
84 under natural conditions.

85 The Loess Plateau is located in northwest China covering an area of 640,000 km².
86 It has a continental monsoonal climate and shows a dramatically interannual
87 fluctuations in precipitation, with the highest precipitation of 1262 mm and the lowest
88 precipitation of only 80 mm, and a mean value of 150–750 mm (Lin and Wang, 2007).
89 The precipitation in the loess regions also shows a dramatically seasonal variation,
90 and approximately 60%–80% of the annual precipitation falls during the three
91 summer months from July to September (Guo et al., 2012). Several recent studies
92 have attempted to determine the dominant factors responsible for the variation of soil
93 respiration in vegetation ecosystems (Lafond et al., 2011; Shi et al., 2011; Jurasinski
94 et al., 2012). However, there have been no studies on the interannual variation in Q_{10} ,
95 nor the factors responsible for these changes. This highlights the need to accurately
96 evaluate the response of SOC mineralization to increasing temperature under warmer

97 climate scenarios in the eroded or degraded regions, because air temperature has been
98 increasing over the past decades (Fan and Wang, 2011; Wang et al., 2012). Thus, the
99 objectives of the present study are to (1) quantify the interannual variation in Q_{10} ; (2)
100 determine the effect of soil moisture on this interannual variation for the period
101 2008–2013 in the Loess Plateau, China.

102

103 **2. Materials and methods**

104 **2.1 Site description**

105 This study was a part of a long-term field experiment that began in 1984 in the
106 State Key Agro-Ecological Experimental Station in the Loess Plateau in Changwu,
107 Shaanxi, China (35°12' N, 107°40' E; 1,200 m above sea level) (Fig. 1). This region
108 had a continental monsoon climate with a mean annual precipitation of 560 mm for
109 the period 1984–2013, over 60% of which occurred from July to September. During
110 this 30-year period, the annual mean air temperature was 9.4 °C and the monthly
111 mean temperature between July and September was 19.4 °C. The study site is also
112 characterized by a ≥ 10 °C accumulated temperature of 3029 °C, an annual sunshine
113 duration of 2230 h, an annual total radiation of 484 kJ cm⁻², and a frost-free period of
114 171 days.

115 The site was located in a typical rain-fed cropping region of the Loess Plateau
116 highland in northwest China. The soil was classified as a loam (Cumulic Haplustoll,
117 USDA Soil Taxonomy System) developed from loess deposits. Soils collected at the
118 study site in 1984 at a depth of 0–20 cm contained 10.5% CaCO₃, 6.5 g organic C
119 kg⁻¹, 0.80 g total N kg⁻¹, and 200 mg NH₄OAc-extractable K kg⁻¹, 3.0 g kg⁻¹ available
120 phosphorus, and had a pH of 8.4 (with a 1: 1 ratio of soil: H₂O), a water-holding
121 capacity of 0.29 cm³ cm⁻³ (v/v), the wilting point of 11%, a soil bulk density of 1.3 g

122 cm⁻³, soil porosity of 51%, and a clay content of 24%.

123 **2.2 Experimental design and management**

124 This study was a part of a long-term field experiment established in June 1984.
125 The plot used in the present study is taken from a bare plot in a state of fallow since
126 June 1984 after the harvesting of winter wheat (*Triticum aestivum* L. ‘Chang Wu 131
127 series’), and living weed was artificially removed timely. Therefore, there were no
128 vegetation or inputs of aboveground and belowground litter, and then SOC
129 mineralization rates in the bare fallow soil did not include root respiration and litter
130 mineralization and decomposition. In this paper, three bare fallow plots were used to
131 investigate the mechanism of underground SOC mineralization rates. All plots of 10.3
132 m × 6.5 m (66.95 m²) were randomly arranged in three blocks. The plots were
133 separated by 0.5 m spaces, whereas the blocks were separated by 1 m strips.

134

135 **2.3 Measurements of SOC mineralization rate and soil microclimate**

136 SOC mineralization rate was measured using an automated closed soil CO₂ flux
137 system with a portable chamber (20 cm in diameter, Li-8100, Lincoln, NE, USA).
138 Approximately one day before the first measurement, a polyvinyl chloride (PVC)
139 collar (20 cm in diameter and 12 cm in height) was inserted to a depth of 2 cm into
140 each plot, and left in place throughout the experimental period from 2008 to 2013.

141 [Five PVC collars were installed in our plots for investigating the spatial variation of](#)
142 [SOC mineralization rate in summer \(July 11, 2008\) and winter \(November 18, 2008\),](#)
143 [respectively.](#) Although previous studies have demonstrated a significant spatial
144 variation of soil respiration, especially in the sites with complex terrain (causing the
145 redistribution of SOC) and different vegetation types (Epron et al., 2006; Luan et al.,
146 2012), the spatial variation of SOC mineralization rate in our sites is small with a

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147 variation coefficient of only 4% and 5% in summer and winter, respectively (Table 1).

148 The results implied that the temporal fluctuation during the measurement have little
149 effect on the spatial variation of SOC mineralization rate. This can be attributed to

150 that there have been no vegetation or inputs of (aboveground and belowground) litter

151 in our plots since 1984 (absolute fallow), and the soil was derived aeolian deposit

152 loess and flat terrain. Due to the small areas of our plots (66.95 m²) and time

153 constraints (5 min for measuring SOC mineralization rate in a given PVC collar), only

154 one PVC collar was used in each plot for measuring SOC mineralization rate and only

155 twice measures were conducted for studying the spatial variation of SOC

156 mineralization rate in our plots. All visible living organisms were removed before the

157 measurement. If necessary, one or more additional measurements would be taken until

158 the variations between two consecutive measurements were less than 15%. The final

159 instantaneous soil respiration for a given collar was the average of the two

160 measurements with a 90 s enclosure period and 30 s delay between them. Field

161 measurements were performed between 09:00 and 11:00 AM from March 2008 to

162 November 2013, except in December, January, and February because of cold weather.

163 A total of 17, 25, 26, 22, 26 and 17 SOC mineralization measurements were made in

164 2008–2013, respectively.

165 Soil temperatures and water contents at a 5-cm depth were measured at a

166 distance of 10 cm from the chamber collar at the same time as the SOC mineralization

167 rates using a Li-Cor thermocouple probe and a Theta Probe ML2X with a HH2 water

168 content meter (Delta-T Devices, Cambridge, England), respectively. Daily mean soil

169 temperature and moisture data were provided by the State Key Agro-Ecological

170 Experimental Station, both of which were measured at 5 cm below the surface using a

171 Hydra soil moisture sensor (Hydra Data Reader and Hydra Probe II Soil Moisture

172 Sensor (SDI-12/RS485); Precision: Moisture, $\pm 0.5\%$ vol; Temperature, ± 0.6 °C;
173 Stevens Water Monitoring Systems Inc., Australia). Soil water-filled pore space
174 (WFPS) was calculated as follows: $\text{WFPS} (\%) = 100 \times [\text{volumetric water content} /$
175 $(2.65 - \text{soil bulk density}) / 2.65]$, with 2.65 being the particle density of the soil (g
176 cm^{-3}).

177

178 **2.4 Data analysis**

179 An exponential (or “ Q_{10} ”) function was used to simulate the relationship between
180 SOC mineralization rate and soil temperature (Xu and Qi, 2001):

$$181 \quad F = \beta_0 e^{\beta_1 T} \quad (1)$$

$$182 \quad Q_{10} = e^{10\beta_1} \quad (2)$$

183 Where F ($\mu \text{ mol m}^{-2} \text{ s}^{-1}$) is the SOC mineralization rate, T (°C) is the soil temperature
184 at a depth of 5 cm, and β_0 and β_1 are the fitted parameters.

185 A quadratic polynomial function was used to simulate the relationship between
186 SOC mineralization rate and soil moisture content (Tang et al., 2005):

$$187 \quad F = \beta_3 \theta^2 + \beta_2 \theta + \beta_4 \quad (3)$$

188 Where θ is the soil moisture at a depth of 0–5 cm, and β_2 , β_3 , and β_4 are the fitted
189 parameters.

190 The interactions of soil temperature with moisture content can more accurately
191 simulate soil respiration than either soil temperature or moisture alone (Tang et al.,
192 2005). Our data indicated that SOC mineralization rate increased with increasing soil
193 moisture content to a maximum at approximately 46% WFPS, and then decreased
194 with further increase of soil moisture content. After comparing different functions and
195 resulting residual plots, a bivariate model was used to simulate the effect of soil
196 moisture content and temperature on SOC mineralization rate:

197
$$F = \beta_0 e^{\beta_1 T \theta + \beta_2 T \theta^2} \quad (4)$$

198 The annual cumulative SOC mineralization rate was estimated by linear
199 interpolating between measurement dates to obtain the mean daily SOC
200 mineralization rate for each plot, and then summing the mean daily SOC
201 mineralization rate for a given year.

202 The relationships between Q_{10} and meteorological factors were investigated
203 using the SAS software (version 8.0; SAS Institute, Cary, NC). All other statistical
204 analyses were performed with ANOVA at $P = 0.05$.

205

206 **3. Results**

207 **3.1 Interannual variation in Q_{10}**

208 The temporal variation in SOC mineralization rate was correlated with that of
209 soil temperature in all six years (Figs. 2b and c), and it increased exponentially with
210 soil temperature ($P < 0.01$). The mean annual SOC mineralization rate ranged from
211 0.83 (2012) to 1.22 $\mu\text{mol m}^{-2} \text{s}^{-1}$ (2008), with a mean of 0.99 $\mu\text{mol m}^{-2} \text{s}^{-1}$ and a CV
212 of 17%; the annual cumulative SOC mineralization ranged from 226 (2012) to 298 g
213 $\text{C m}^{-2} \text{y}^{-1}$ (2009), with a mean of 253 $\text{g C m}^{-2} \text{y}^{-1}$ and a CV of 13% (Table 2), and the
214 annual Q_{10} in our sites was 1.65 in 2008, 1.94 in 2009, 1.72 in 2010, 1.48 in 2011,
215 1.86 in 2012, and 1.55 in 2013, respectively, with a mean Q_{10} of 1.72 and a CV of
216 10% (Table 3).

217

218 **3.2 Interannual variation in soil microclimate**

219 Annual precipitation showed a significant annual variation (Fig.1 and Table 2; P
220 < 0.05). Rainfall ranged from 481 (2009 and 2012) to 644 mm (2011), with a 6-year
221 mean of 540 ± 64 mm and a CV of 12%. Annual rainfall days ranged from 71 (2013) to

222 105 days (2008), with a 6-year mean of 96 ± 12 days and a CV of 13%. Interannual
223 variation in air temperature was not significant (Fig.1 and Table 2; $P > 0.05$). It ranged
224 from 9.43 (2011 and 2012) to 11.08 °C (2013), with a 6-year mean of 10.1 ± 0.6 °C and
225 a CV of only 6%.

226 Soil temperature and soil moisture at a depth of 0–5 cm showed significant
227 temporal variations over the six-year observation period (Fig. 2b). The seasonal mean
228 soil moisture content was 49.2% WFPS in the wet season (July to September in each
229 year) and 38.6% WFPS in the dry season (other months). The mean annual soil
230 moisture content ranged from 38.6% WFPS (2013) to 50.7% WFPS (2011), with a
231 mean of 43.8% WFPS and a CV of 11%. The seasonal mean soil temperature was
232 14.50 °C in the dry season and 20.39 °C in the wet season. The mean annual soil
233 temperature ranged from 14.90 °C (2011) to 18.42 °C (2009), with a mean of
234 17.05 °C and a CV of only 7%.

235

236 **3.3 Effect of soil moisture on the interannual variation of Q_{10}**

237 Annual Q_{10} showed a negative quadratic correlation with annual mean soil
238 moisture (Fig. 3b). Additionally, the seasonal SOC mineralization rate increased
239 exponentially with soil temperature, and showed a negative quadratic correlation with
240 soil moisture content (Table 3). The response surface of SOC mineralization rate to
241 soil temperature and moisture including both seasonal and interannual scales clearly
242 described how soil microclimate influenced SOC mineralization rate (Fig. 4).

243

244 **4. Discussion**

245 **4.1 Soil moisture influenced the interannual variation in Q_{10}**

246 The range of annual Q_{10} (1.48–1.94, with a CV of 10%) in our sites for the

247 period 2008–2013 was within the limits reported for annual Q_{10} (1.20–4.89) at global
248 scale (Boone et al., 1998; Zhou et al., 2007; Gaumont-Guay et al., 2008; Zhu and
249 Cheng, 2011; Zimmermann et al., 2012). However, the mean annual Q_{10} in our sites
250 (1.70) was lower than the global mean (2.47) (Boone et al., 1998; Zhou et al., 2007;
251 Gaumont-Guay et al., 2008; Zhu and Cheng, 2011; Zimmermann et al., 2012),
252 probably due to low SOC contents, small microbial communities, dry soil conditions
253 in semi-arid regions (Conant et al., 2004; Gershenson et al., 2009; Cable et al., 2011),
254 and different methods used for separating SOC mineralization rate (Boone et al., 1998;
255 Zhu and Cheng, 2011; Zimmermann et al., 2012).

256 Annual Q_{10} was negatively linearly correlated with annual mean precipitation,
257 but this correlation did not reach statistical significance ($P>0.05$); whereas it was
258 significantly related to soil moisture content (Fig. 3). This was in agreement with
259 previous studies (Suseela et al., 2012; Poll et al., 2013). However, Q_{10} was found to be
260 negatively correlated with mean annual precipitation ($P<0.01$) in different forest
261 ecosystems in China, which could be due to the relatively abundant rainfall in the
262 forest ecosystems (700–1956 mm) (Peng et al., 2009). Soil moisture was the major
263 limiting factor for the underground biological processes, especially in water-limited
264 regions (Reth et al., 2005; Balogh et al., 2011; Wang et al., 2014). Although
265 precipitation was the only source of water for soil moisture underneath long-term bare
266 soil, there was no significant relationship between annual mean soil moisture and
267 annual precipitation amount ($P>0.05$) (Fig. 5a), but rainfall frequency and distribution
268 were closely related to annual mean soil moisture content (Fig. 5b). Similar results
269 have also been found in other studies (Coronato and Bertiller, 1996; Qiu et al., 2001;
270 Cho and Choi, 2014). The annual precipitation during the six-year observation period
271 of 2008–2013 ranged from 481 (2009) to 644 mm (2011), with a CV of 12% (Table 2).

272 The annual mean soil moisture content was high (51% WFPS) in 2011 due to
273 relatively uniform distribution of precipitation, and low (38% WFPS) in 2010 and
274 2013 due to relatively uneven distribution of precipitation. For example, the rainfall
275 amount on July 23, 2010 (118 mm) and July 22, 2013 (121 mm) was about 20% and
276 23% of that in 2010 (588 mm) and 2013 (523 mm), respectively. The annual mean
277 soil moisture was moderate (43–47% WFPS) in 2008, 2009 and 2012 due to the
278 normal distribution of precipitation. Similarly, the interannual soil moisture regulation
279 in the forest ecosystems in the Loess Plateau was determined not only by rainfall
280 amount but also by rainfall distribution (Li et al., 1998).

281 Annual Q_{10} showed a negative quadratic relationship with soil moisture content,
282 as it increased with increasing soil moisture content to a maximum at approximately
283 42% WFPS, and then decreased with further increase of soil moisture content (Fig.
284 3b), which was in agreement with other studies (Bowden et al., 1998; Conant et al.,
285 2004; Smith, 2005). This could be attributed to the following reasons: Firstly, lower
286 soil water availability could reduce Q_{10} by limiting respiration substrate availability
287 and soil pore water became increasingly disconnected, thus slowing down the
288 diffusion rate of solutes (Wan et al., 2007; Balogh et al., 2011), and decreasing the
289 activity and quantity of organisms due to drought stress (Davidson et al., 2006).
290 Secondly, higher soil moisture could also reduce Q_{10} by limiting O_2 diffusion rate
291 (Davidson et al., 1998; Byrne et al., 2005; Saiz et al., 2007) because of low effective
292 soil porosity, as the diffusion rate of O_2 through water was much slower than that
293 through air (Cook and Knight, 2003; Manzoni et al., 2012), thus the decomposition
294 activity of aerobic microbes was inhibited due to lack of oxygen (Davidson et al.,
295 2000). Finally, the diffusion rate of both soluble organic matter and O_2 were not
296 inhibited, also the survival of microorganisms not subject to water stress at suitable

297 soil water content, instead increasing temperature increased the diffusion of soluble
298 organic matter, thus resulting in an increase in Q_{10} (McCulley et al., 2007). Overall,
299 soil moisture content may be the most important factors that affected the interannual
300 variation in Q_{10} .

301 The variation in the temperature sensitivities of SOC mineralization could have
302 potential implications for climate carbon modeling (Davidson and Janssens, 2006;
303 Conant et al., 2011), as uncertainty remains regarding environmental controls over
304 SOC mineralization (Larionova et al., 2007; Karhu et al., 2010; Conant et al., 2011;
305 Sakurai et al., 2012). The previous results have emphasized the importance of
306 seasonal variation in precipitation and soil moisture in determining Q_{10} (Xu and Qi,
307 2001; Davidson et al., 2006; Davidson and Janssens, 2006), but have rarely taken into
308 account the interannual variation in soil moisture resulting from the uneven
309 distribution of precipitation. Carbon cycle modeling without considering this
310 interannual variation in soil moisture may produce misleading conclusions.

311

312 **4.2 Comparison with annual cumulative SOC mineralization rate estimated by** 313 **different methods**

314 Annual cumulative SOC mineralization rate was estimated by different methods,
315 including linear interpolation method, modeled method, and unit conversion method.
316 The results clearly showed that there was no significant difference in the estimates of
317 annual cumulative SOC mineralization rate between linear interpolation and modeled
318 method, and the modeled method could well predict the SOC mineralization rate in
319 most cases from 2008 to 2013 (Fig. 6), which was in line with the previous studies
320 (Tang et al., 2005). However, unit conversion method seriously overestimated annual
321 cumulative SOC mineralization rate (Table 4). This can be attributed to the following

322 reasons: 1) the study site has a continental monsoon climate with 60% of rainfall
323 occurring from July to September (rainy season), thus the study site is hot and rainy in
324 the rainy season, but cool and dry in the non-rainy season; and 2) SOC mineralization
325 rate in the rainy and non-rainy season is largely the same, but the duration of rainy
326 season is only a quarter of a year. Thus, the SOC mineralization rate was much greater
327 in rainy season than in non-rainy season, thus resulting in an overestimation of
328 cumulative SOC mineralization rate in a given year.

329 In conclusion, linear interpolation method is a simple and controllable method
330 for estimating annual cumulative SOC mineralization rate (Schindlbacher et al., 2014;
331 Shi et al., 2014). Although the modeled method can well estimate annual cumulative
332 SOC mineralization rate, it is limited in practice as it needs daily soil temperature and
333 moisture. Unit conversion method may seriously overestimate annual cumulative
334 SOC mineralization rate unless the SOC mineralization rate is very uniform in a given
335 year.

336

337 **5. Conclusions**

338 Understanding the factors influencing the temperature sensitivity of SOC
339 mineralization is important to accurately estimate local carbon cycle. The results of
340 this study showed that the annual cumulative SOC mineralization ranged from 226 to
341 298 g C m⁻² y⁻¹, with a CV of 13%, annual Q_{10} ranged from 1.48 to 1.94, with a CV
342 of 10%, and annual soil moisture content ranged from 38.6 to 50.7% WFPS, with a
343 CV of 11%. Annual Q_{10} showed a negative quadratic correlation with annual mean
344 soil moisture, which was determined by uneven distribution and frequency of rainfall.
345 In conclusion, the interannual variation in soil moisture content should be considered
346 in carbon cycle models in semi-arid areas.

347

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499 Table 1. SOC mineralization rate ($\mu \text{ mol m}^{-2} \text{ s}^{-1}$) in summer (July 11, 2008) and winter
 500 (November 18, 2008). Data are represented as mean \pm S.D of five collars.

SOC mineralization rate						
Dates	Collar 1	Collar 2	Collar 3	Collar 4	Collar 5	Mean value
Summer	1.55 \pm 0.11	1.60 \pm 0.20	1.58 \pm 0.21	1.49 \pm 0.07	1.65 \pm 0.18	1.57 \pm 0.06
Winter	0.29 \pm 0.01	0.30 \pm 0.02	0.31 \pm 0.01	0.32 \pm 0.02	0.33 \pm 0.02	0.31 \pm 0.02

501 Note: SOC mineralization rate was measured on July 11, 2008 and November 18, 2008 (representing
 502 summer and winter) using 5 PVC collars installed in our plots

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521 Table 2. Cumulative SOC mineralization rate ($\text{g C m}^{-2} \text{year}^{-1}$), annual precipitation amount
 522 (mm), annual precipitation days, and air temperature ($^{\circ}\text{C}$) from 2009 to 2013. Data are
 523 represented as mean \pm S.D.

Years	Cumulative SOC mineralization rate	Precipitation amount	Precipitation days	Air temperature
2008	293 \pm 10	520	105	9.76
2009	298 \pm 9	481	99	10.26
2010	238 \pm 50	588	101	10.39
2011	234 \pm 48	644	100	9.43
2012	226 \pm 19	481	98	9.43
2013	240 \pm 30	523	71	11.08
Mean	253 \pm 32	540 \pm 64	96 \pm 12	10.1 \pm 0.6

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538 Table 3. Relationships between SOC mineralization rate and soil temperature (F-T) or soil
 539 moisture (F- θ) for each year from 2008 to 2013.

Years	F-T			F- θ			
	Functions	R^2	P	Q_{10}	Functions	R^2	P
2008	$F=0.49e^{0.0499T}$	0.56	<0.01	1.65	$F=-0.0008\theta^2 + 0.10\theta - 1.52$	0.53	<0.01
2009	$F=0.34e^{0.0661T}$	0.63	<0.01	1.94	$F=-0.0001\theta^2 - 0.02\theta + 2.63$	0.61	<0.01
2010	$F=0.35e^{0.0544T}$	0.47	<0.01	1.72	$F=0.0002\theta^2 - 0.04\theta + 2.15$	0.86	<0.01
2011	$F=0.45e^{0.0395T}$	0.47	<0.01	1.48	$F=-0.0008\theta^2 + 0.06\theta + 0.06$	0.46	<0.01
2012	$F=0.27e^{0.0623T}$	0.67	<0.01	1.86	$F=-0.0019\theta^2 + 0.14\theta - 1.71$	0.35	<0.05
2013	$F=0.52e^{0.0441T}$	0.32	<0.01	1.55	$F=-0.001\theta^2 + 0.08\theta - 0.60$	0.36	<0.05

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557 Table 4. Annual cumulative SOC mineralization rate ($\text{g C m}^{-2} \text{ year}^{-1}$) estimated by linear
558 interpolation method, modeled method, and unit conversed method from 2008 to 2013.

Years	Annual cumulative SOC mineralization rate		
	Linear interpolation	Soil temperature and moisture modeled	Unit conversion
2008	293	258	462
2009	298	272	460
2010	238	268	344
2011	234	260	325
2012	226	271	314
2013	240	284	348
Mean	255 \pm 32	269 \pm 6	374 \pm 65

559 Note: Modeled method: using the interactions of soil temperature with moisture for estimating annual cumulative SOC
560 mineralization rate with Eq. 4 (2.4 sections); Unit conversion method: estimating annual cumulative SOC mineralization rate
561 with mean SOC mineralization rate in a given year.

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574 Figure captions

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576 Fig. 1

577 Location of the State Key Agro-Ecological Experimental Station (Changwu Station).

578

579 Fig. 2

580 Temporal variations of (a) precipitation and air temperature, (b) soil moisture and soil
581 temperature, and (c) SOC mineralization rate from 2008 to 2013.

582

583 Fig. 3

584 Regression analysis performed between (a) Q_{10} and annual precipitation amount, and (b) Q_{10}
585 and annual mean soil moisture.

586

587 Fig. 4

588 Response surface of SOC mineralization rate as a function of soil moisture and soil
589 temperature from 2008 to 2013.

590

591 Fig. 5

592 Regression analysis performed between (a) annual mean soil moisture and annual
593 precipitation amount, and (b) annual mean soil moisture and annual precipitation days.

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595 Fig. 6

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597 Estimated daily (2008–2013) SOC mineralization rate (solid line) with periodic measurement
598 values (filled circles).

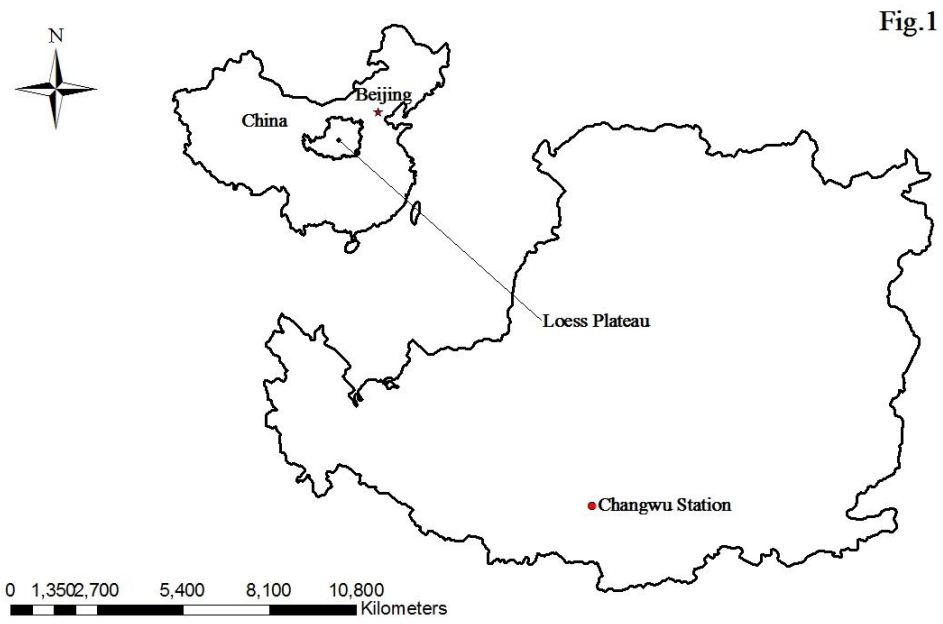


Fig.1

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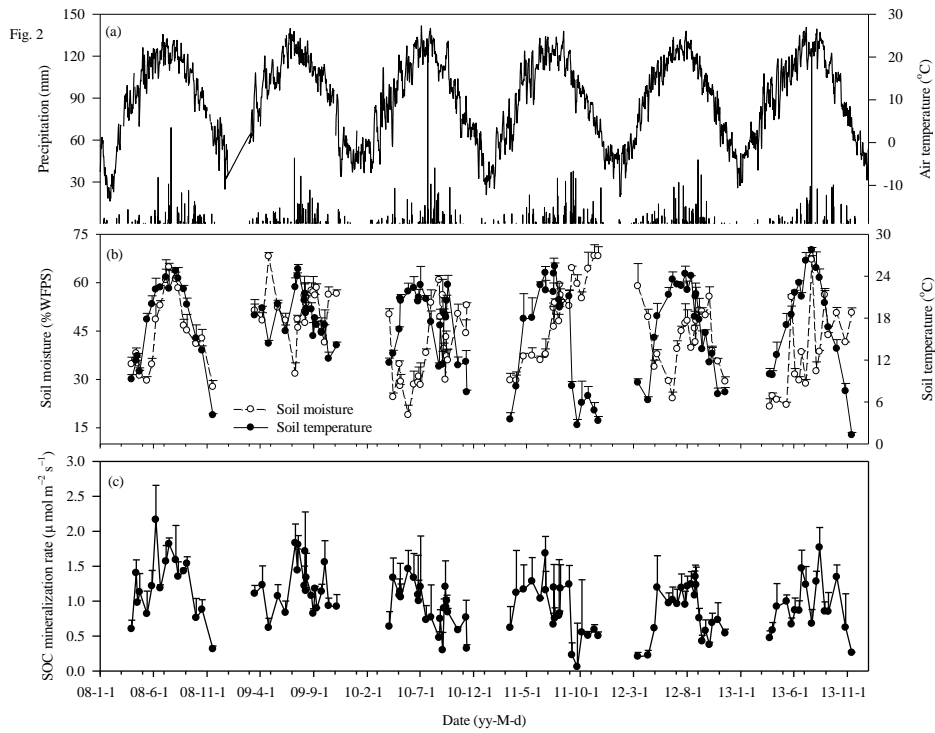
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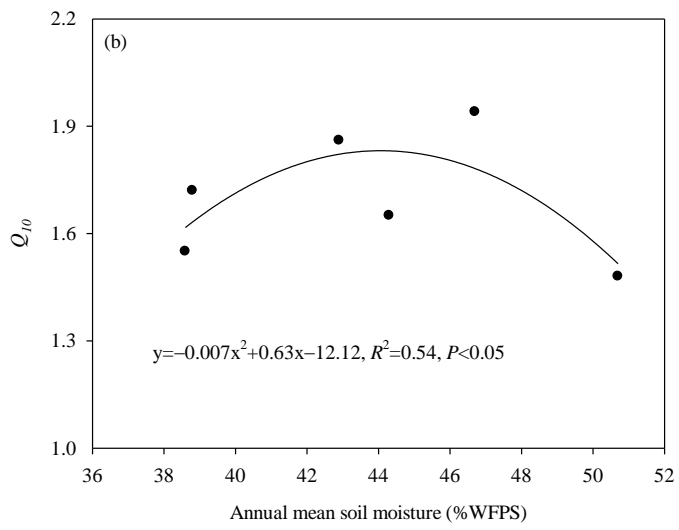
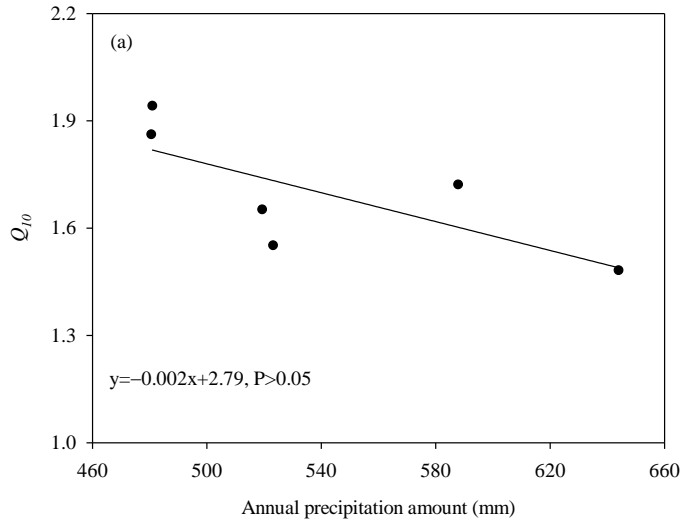
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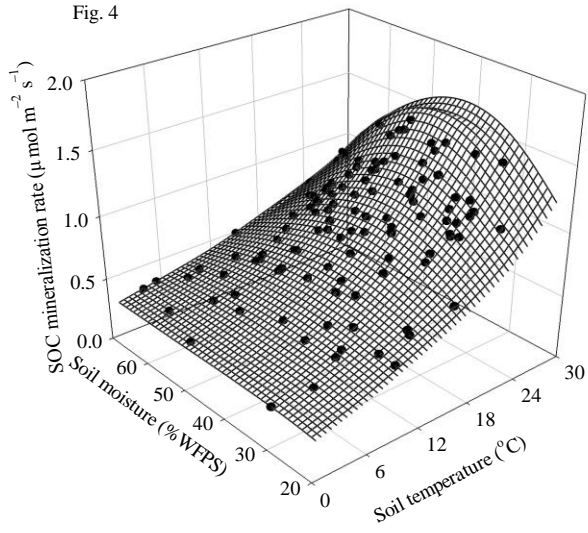
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Fig. 3



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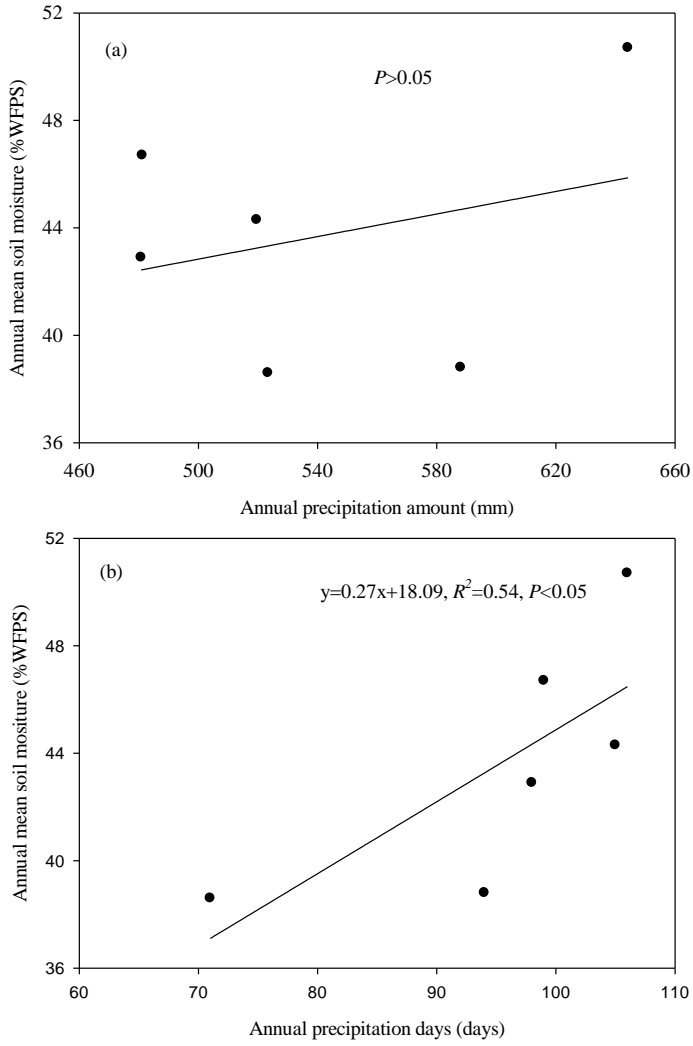
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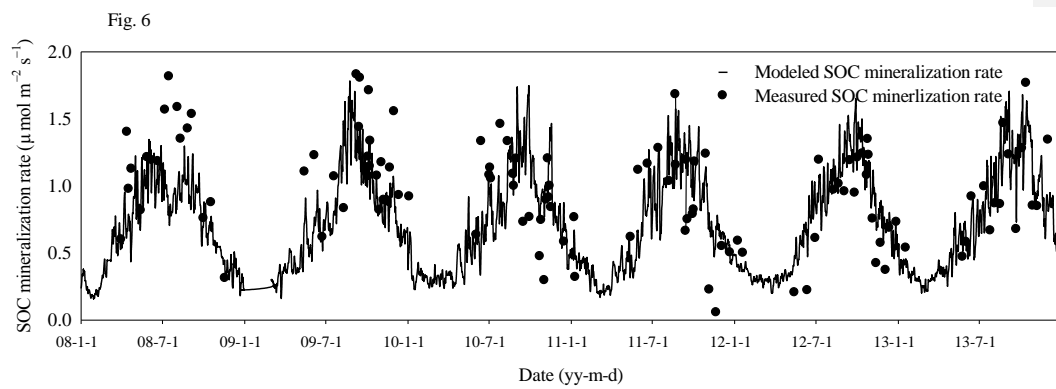
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Fig.5



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