#### 1 <u>REFEREE 1</u> (responses in bold)

This manuscript used the differences in  ${}^{13}C$  and  ${}^{14}C$  discrimination of the geogenic CO<sub>2</sub> emitted at mofette soils and atmospheric CO<sub>2</sub> in reference soils to distinguish C fixed by plants from C fixed by autotrophic microorganisms. The results show that CO<sub>2</sub> fixation by autotrophic microorganisms contributes significantly to soil organic matter formation and alters the isotope signatures in mofette soils. The experiment seems well designed and the study is overall well presented.

8

## 9 Thank you for handling our manuscript and for your constructive and helpful 10 comments. Replies to specific comments are given below.

11

12 However, a number of aspects were raised which were commented directly in the manuscript.

13

14 1. Page2 Line7-11. It is inconclusive to infer that the negative  $\delta^{13}$ C shift was caused

15 by the activity of chemolithoautotrophic microorganisms with the quantitative data even 16 under dark incubation. The soils used for isotope analysis (labelling experiment) were 17 different from those for quantitative PCR analysis, which makes me doubt more about the 18 consistency of the analysis. Furthurmore, I think it is inappropriate to classify aceto- genic 19 microorganisms. methanogenic microorganisms and chemo-lithoautotrophic mi-20 croorgansims into different catalogs.

We agree that the term chemolithoautotrophic microorganisms might be misleading 21 22 and does not represent the organisms that we wanted to target with qPCR analyses. Our 23 target was to get information about the potential of CO<sub>2</sub> fixation through the Calvin Benson Basham Cycle (CBB) in the mofette soils. This is important for our mass balance 24 approach, because the metabolic cycle determines the  $\delta^{13}$ C signature of the microbial 25 end-member in the isotope mass balance. These data should be complementary to 26 27 information about acetogenic and methanogenic pathways from previous studies (Beulig 28 et al. 2014, see manuscript for reference; see also discussion, 4.3). The term 29 chemoautotrophic microorganisms should refer to organisms using the CBB cycle for CO<sub>2</sub> fixation. We agree that other microbes also use the CBB Cycle. Similarly, not all 30

chemolithoautotrophs use the CBB Cycle, although it is the most common metabolic cycle. Therefore you are correct and the term chemolithoatutrophic microorganisms is not correct in this context. It was changed to "autotrophic microorganisms using the CBB cycle". The data for the qPCR was taken from same soil as for the labelling experiments (mofette soil 1), although they were sampled at different time points. This was clarified in the text.

7

8 2. Page2 Line 13. Change "organisms" to "microorganisms".

9 Done

10 3. Page2 Line 14. Consider adding "carbon" before "observed reservoir effects" here.

11 **Done** 

4. Page3 Line 2-4. Many published papers concerned this issue. Specify the differ- ences ofyour study from these published researches.

14 We added a sentence to clarify that we try to target the question of whether CO<sub>2</sub> fixation

15 influences carbon isotope signatures, and that we try to use a quantitative approach by

16 means of natural abundance carbon isotope values, that distinguishes it from other

17 studies dealing with CO<sub>2</sub> fixation in soils.

18

19 5. Page3 Line 30. "indicates"

20 Done

21

22 6. Page4 Line 23-25. Do the cbbL genes encode Form I RubisCO or Form II RubisCO, or

23 both Form I and Form II RubisCO? Make it clear.

7. Page 4 Line 25. How many subclasses do Form I RubisCO have? Why chose cbbL IA andcbbL IC in this study?

Answers to 6 and 7: We tried to improve this section according to the referee's suggestion. cbbL encodes for Form I RubisCO and we choose Form I genes encoding for subclasses 1A and 1C, because they can give us information as to weather CO<sub>2</sub> is assimilated by obligate or facultative microorganisms. Further, cbbM, encoding for Form II RubisCO, can give information if microbes with special adaption to anaerobic
 environments are important CO<sub>2</sub> fixing microorganisms.

3

8. Page5 Line 16-17. How far is Mofette 1 from Mofette2? If close, why the geochemicalproperties are so different? How to fix the sampling sizes of these two mofettes?

6 Mofette 2 is approximately 500 m distant from mofette 1. The geochemical properties 7 are very different, because the two mofettes differ in size. This is mentioned in the text 8 (page 5, line 25). Mofette 1 is considerably smaller than mofette 2. Thus, plant litter 9 inputs from the top is greater in mofette 1 compared to mofette 2, because plants 10 growing at the rim of the structure can more easily fall into the bare centre of the 11 mofette structure.

- 12
- 13

9. Page5 Line 29-30. Page 6 Line 12. Please clarify the sampling time of soil and plantsamples respectively.

16 10. Page6 Line14. Why take the vegetation samples at 2 meter intervals? Did you collect
17 Eriophorum vaginatum, Deschampsia cespitosa and Filipendula ulmaria from each intervals?

18 Answers to 9 and 10. Dates for all three sampling campaigns were clarified in the text, as 19 was the sampling strategy. We took vegetation samples (Eriophorum vaginatum) in the direct vicinity (rim) of both mofette structures. In addition, we took samples from a 20 21 transect crossing mofette 2, because mofette 2 represents an undisturbed hummock 22 structure without the disturbances associated with secondary exhalation structures found in the overall smaller mofette 1. In mofette 2, the vegetation changed with 23 24 increasing distance from the central CO<sub>2</sub> exhalation hummock (which was bare of 25 vegetation). This was clarified in the text.

The goal of the transect sampling was to have plant samples that experience a full mixture of the two presumed end members (geogenic and atmospheric CO<sub>2</sub>), so we could determine if  $\delta^{13}$ C and  $\Delta^{14}$ C values follow a linear trend with increasing CO<sub>2</sub> concentrations. This was necessary to make predictions for mofette SOM, because mofette SOM is derived from plants that are likely exposed to fluctuating CO<sub>2</sub>

concentrations. In order to make predictions for mofette SOM from vegetation  $\delta^{13}$ C and 1 2  $\Delta^{14}$ C values, it is therefore necessary to test, weather plants follow a linear trend in  $\delta^{13}$ C and  $\Lambda^{14}$ C values within a CO<sub>2</sub> gradient, as well as the if the relationship between  $\delta^{13}$ C 3 and  $\Lambda^{14}$ C of plants is linear with increasing CO<sub>2</sub> concentrations, or if fractionation of 4 plants is influenced by elevated CO<sub>2</sub> concentrations. The good correlation of  $\delta^{13}$ C and 5  $\Delta^{14}$ C, even with increasing CO<sub>2</sub> concentrations indicated that our model used for 6 prediction of <sup>13</sup>C values from <sup>14</sup>C is valid for plant-derived SOM (assuming, for the 7 moment, no radioactive decay influence on <sup>14</sup>C). 8

9

10

11 11. Page6 Line7-8. Mofette and reference soil cores used for geochemical properties analysis
were divided into depth intervals from 0-10cm, 10-25cm and 25-40cm according to the
sampling strategy. However, results (Table 1) showed these soil cores were sectioned into
0-5cm, 5-10cm, 10-20cm, 20-30cm and 30-40cm.

15 Table 1 shows the geochemical data from the sampling campaign conducted in September 2014, when we took samples for the second labelling experiment. The data 16 17 given in table 1 represent the unlabelled (time zero) geochemical data from soils 18 sampled for the second labelling experiment. We choose to use these data, to present  $\delta^{13}$ C isotope values with a higher spatial resolution, which gives a more detailed insight 19 20 into small scale variations in the geochemical and stable isotope changes in both, mofette and reference soil with depth. However, we do not have radiocarbon data from this 21 sampling campaign. Therefore, we included  $\Delta^{14}$ C values from the first sampling 22 campaign into table 1, which has a broader depth resolution. This is now clarified in the 23 table caption. 24

25

26

27 12. Page6 Line10-11. Confirm that the stable isotope and radiocarbon analysis were28 performed in triplicate for each mixed sample.

All stable isotope analyses were conducted in triplicate. For sampling campaign 1, we used a mixed homogenized sample composited from three cores. This homogenized sample was subsampled three times for stable isotope analyses. For radiocarbon, we

measured only one subsample. Therefore, uncertainties given for radiocarbon analyses
from sampling campaign 1 represent the analytical uncertainty (see revised figure
caption 3A)

4

5 13. Page6 Line 23-24. Why remove the top of the Oh horizon?

6 We removed the Oh layer because we wanted avoid CO<sub>2</sub> uptake by phototrophic
7 organisms like algae.

8

9 14. Page9 Line 4. The temperature may affect microbial activity. The incubation temperature
10 was10°C in the first labelling experiment and it was set to 12°C in the second labelling
11 experiment. How do you set the temperature?

12 The temperature given for experiment 1 is a mistake in the text. Both experiments were 13 conducted at 12°C. This has been corrected in the text.

14

15 15. Page10 Line 19-20. Was the DNA extraction performed in triple?

16 All DNA analyses were performed in triplicate.

17

18 16. Page11 Line1-7. What about the amplification reaction and program?

19 The exact protocol for amplification reactions and the program used have been inserted 20 in the text (page 11, line 22-30).

21

17. Page12 Line 12-13. It was inconsistent with the sampling method part, where you showedthat the replicates of respective depth intervals were mixed (Page6 Line 8-9).

24 This is correct. We did not perform a t-test on radiocarbon data, because the error in

25  $\Delta$ 14C values represent only the analytical precision. This was corrected in the text.

1 18. Page12 Line 20-24. The mofette soils and reference soils are close to each other (5 or 18
meters), however, the TOC contents in these soils are so different. And the variations of TOC
contents along soil depths are also quite different. Why?

4 The differences in C content, pH and C/N ratio are caused by permanently anoxic 5 conditions in mofette soil compared to the reference soil and by the addition of plant 6 material as well as microbial carbon that has been partly derived from autotrophic 7 organisms. The reasons for the observed differences are also discussed in detail in 8 section 4.1.

9

10 19. Page13 Line4-5. What about the radiocarbon and stable isotope concentrations of  $CO_2$ 

11 sampled at different depth (5cm, 15cm, 25cm, 40 cm)?

12 The radiocarbon and  $\delta^{13}$ C values of sampled CO<sub>2</sub> are given in line 27.

13

14 20. Page13 Line 6. I don't think "Measured" in this sentence is necessary.

- 15 **Done**
- 16

17 21. Page15 Line 12. What do CFE and bulk mean in table2?

18 CFE means values obtained from the first experiment, where uptake rates were 19 determined by CFE extractions and bulk refers to experiment 2, where <sup>13</sup>CO<sub>2</sub> 20 incorporation directly into bulk organic carbon was determined, without extracting 21 microbial biomass. It was clarified in the text.

22

22. Page16 Line2. The ratios of cbbL IC to 16s RNA in this study are almost two orders
magnitude higher than those in other published papers (table3), it is hard to believe that the
ratios range from 7% to 37% in these investigated soils. Even the cbbL IA copy numbers at 510cm in Mofette1 is 1.40E+08±1.69E+08. These results did throw my doubt on their
quantitative PCR.

The high number of 16s RNA and cbbL 1C and 1A genes correspond with the low C/N ratios and high CO2 uptake rates found in Mofette 1. Together with the geochemical data, and the activity measurements, the high number of 16S rRNA genes as well as
 genes encoding for RubisCO reflect a large number of microorganisms in these soils,
 which finally leads to an increased contribution of microorganisms to SOM

4

5 23. Page16 Line 23. I don't understand what "processed OM" are here.

#### 6 **Processed OM refers to partly degraded OM. This was clarified in the text.**

7

8 24. Page16 Line28-31. I didn't see the connection between RbusiCO encoding genes and
9 microbially assimilated carbon. This would depend more on the activity of CO2 assimilation
10 microorgansims.

11 We know from the study from Beulig et al. (2014), that acetogenic and methanogenic microorganisms are active in the mofette soil, especially in the top 10 cm. This was also 12 13 confirmed by a metatranscriptomic approach (Beulig et al., in press). In this study we 14 did complementary analyses to evaluate the importance of mciroorganisms using the 15 Calcin Benson Cycle. This is important, in order to derive the isotopic signature of 16 microbial carbon that is derived from CO2 fixation. This is discussed in detail in section 17 4.3 in the discussion. Obviously, RubisCO is also a pathway that contributes to CO2 18 assimilation and has to be considered by defining a microbial isotope end-member in the 19 isotope mass balance.

20

21 25. Page17 Line7-20. The discussion is too weak. Try not to repeat what was already22 presented in the methods and results.

#### 23 The discussion in this section was streamlined.

24

26. Page18 Line 19. For the correlation analysis in figure 4, the lateral axis rep- resents the
cbbL gene abundance, while the circle marker means cbbL, cbbM gene abundances. Which
marker gene do you use for this analysis?

Yes, the axis title is wrong. The data points represent total number of cbbL and cbbM
genes.

1

27. P19. Line 2-4 Which kinds of chemoautotrophic bacteria were involved in the Calvin
Benson Cycle according to the metatranscriptomic analysis? The good correlation of
cbbL/cbbM marker genes and CO2 fixation rates indicated that the fixed carbon derived from
autotrophic bacteria, not chemoautotrophic bacteria.

6

7 28. P19. Line 5-6 Is any data support your state that type I RubisCO is the dominant type in8 the mofette?

9 Type I RubisCO is the dominant type, because it is most abundant in the mofette soil.

#### 1 <u>REFEREE 2</u>

2

3 The manuscript describes a study investigating the autotrophic CO2 fixation by soil 4 microorganisms and their contribution to soil organic matter (SOM) in mofette soils 5 compared to reference soils. The particular composition of the geogenic CO<sub>2</sub> at these sites 6 allows estimating the contribution of plant-derived, SOM-derived and CO<sub>2</sub>-derived C in soil 7 microbial biomass and SOM. The approach taken by the authors includes tracing the isotopic composition (both <sup>13</sup>C and <sup>14</sup>C) into soil microbial biomass and SOM, but also molecular 8 analyses of genes involved in autotrophic CO<sub>2</sub> fixation. This allows the authors to study the 9 10 process and relate it to potentially responsible microorganisms.

11 methods The results show that a significant percentage of soil C in the investigated mofette soils is 12 derived from geogenic C, and that cbbL 1C was more abundant than the other genes 13 investigated. The manuscript thus describes an important process which has been mostly 14 neglected in the past. It is well written and organized, and thus it should be published. 15 However, some relatively minor revisions to the manuscript will further improve it.

16

# Thank you for these encouraging and constructive comments. Our answers to specific comments are in bold face, below.

19

20 Here are some more detailed comments:

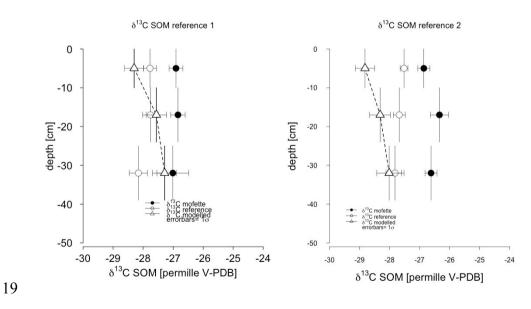
Labelling experiments: Were unlabelled controls included or are analyses of the starting materials used to correct for background values? Also, how were the different  $CO_2$ concentrations in the air during the labelling experiment accounted for when assessing the data?

We used  $\delta^{13}$ C values of autoclaved controls as background values. All autoclaved controls showed zero enrichment after labelling, compared to the samples that were not autoclaved. The non-autoclaved samples had enrichments of up to +80 % in unfumigated extracts, although enrichment was always higher after fumigation (see page 18, line 18-21). This is most probably caused by formation of 1 secondary metabolites of the microorganisms that were synthesised from labelled 2  $CO_2$  and excreted, presumably acetate. However, the  $\delta^{13}C$  value of autoclaved 3 samples does not accurately represent the natural abundance  $\delta^{13}C$  value of the 4 microbial biomass, because autoclaving disturbed the sample and also could have 5 made some of the plant material extractable. Natural abundance  $\delta^{13}C$  values of 6 microbial biomass are presumably more negative than the values obtained from 7 autoclaved samples and also more negative than bulk SOM.

p. 14571, top paragraph: When looking at Fig. 1, I wonder whether the effect of radioactive
decay need not be taken into account already here. For the reference soils, delta<sup>13</sup>C is almost
constant, whereas Delta<sup>14</sup>C varies, indicating different ages. How would this affect the
relationship between the two isotopic signatures?

12

13 It is true that the decreasing <sup>14</sup>C with depth in the reference soils indicate aging of SOM 14 with depth. However, interestingly, there is no change in  $\delta^{13}$ C values with depth. This 15 means, that  $\delta^{13}$ C values are not changed with increasing decomposition stages in these 16 soils, as observed in other soils, where there is usually an increase in  $\delta^{13}$ C values with 17 depth. Theoretically, it is possible to apply the model from eq. 9 to the reference soil. 18 This we have done in the attached plot (see figure 187 below).



20 Figure 184: Adapted model for reference soils.

The results indicate that measured  $\delta^{13}$ C values for both reference soils are more positive 1 2 in the first 5 cm, whereas they should increase with depth according to the model. However, in both soils  $\delta^{13}$ C values show the opposite trend and get more depleted with 3 4 depth. One has to consider that the process assumed responsible for the shift in the 5 mofette model is different than for the prediction for the reference soil in the figure 187. 6 In the reference soil the model assumes that  $\triangle 14C$  is mainly determined from the 7 amount of incorporated geogenic CO<sub>2</sub>, whereas radioactive decay is of minor importance compared to geogenic CO<sub>2</sub> and can be corrected with <sup>14</sup>C decay derived 8 9 from the reference soil. The model prediction shown in figure 187 shows implies that there is no linear relationship between aging of organic matter (as implied by  $\Delta^{14}$ C 10 values) and  $\delta^{13}$ C values. An increase in  $\delta^{13}$ C values with increasing decomposition of the 11 organic matter is therefore not supported by the relationship of  $\Delta^{14}$ C and  $\delta^{13}$ C values in 12 13 the reference soil. CO<sub>2</sub> fixation might be an explanation for this, because it adds depleted carbon via microbial biomass to the soil and might "shift"  $\delta^{13}$ C back towards 14 15 more negative values.

16

17 p. 14575, line 5-bottom of page: I certainly agree with the authors that radioactive has to be 18 taken into account when estimating the contribution of the different pools to SOM or microbial biomass. However, I have some doubts about how good their approach to do so 19 20 was. The authors claim that after correction for radioactive decay, all but one calculated value 21 for delta13C in the horizons deeper than 10 cm match the measured values. In total, however, 22 this means that 3 out of 6 calculated values match the measured ones, the other 3 don't. So it 23 is difficult to judge whether this correction really did a good job. The authors mention 24 different organic matter dynamics in the mofette and the reference soils as a potential source 25 of error. If there are any indications for this, this should be discussed more in detail.

26

Indeed, there is some evidence that carbon dynamics are slower in the deepest layer sampled in mofette 2. All soils in the floodplain are subjected to fluctuating water levels. However, in the mofettes these fluctuations are attenuated, because within the central part of the exhalation water table is elevated by the upstreaming CO<sub>2</sub>. Mofette 2 is considerably larger than mofette 1 and the CO<sub>2</sub> discharge is somewhat higher (see answer 8 to referee #1). Mofette soil 2 is therefore likely water-saturated throughout the 1 whole year, in contrast the all other soils, although we have no direct evidence, because 2 we did not measured the water level throughout the whole year. However, we have indirect evidence, because during the sampling campaign in September, it was not 3 possible the gain soil cores from the deepest point with our auger (this is the reason, why 4 5 this depth is missing from table 1 for that date). Permanently waterlogged conditions 6 might lead to much lower C turnover and the model-correction of radioactive decay with  $\Delta^{14}$ C from the reference soil might not be valid. This means that the modelled  $\delta^{13}$ C 7 8 values are biased toward too positive values.

9

p. 14578, line 1-11: An additional potential reason for the increase with depth of CO2 fixation
normalized to C in the reference soil, which could be included in the discussion, might be
that the deeper parts of the profile are adapted to higher CO2 concentrations in the soil air.

13 This is a very good point, we now include it in the text.

14

15 minor editorial comments:

p. 14556, line 2: I think it is a bit too ambitious to claim quantifying the actual con- tribution
of autotrophic microorganisms to SOM formation. I suggest to speak about "potential
contribution".

#### 19 Done. We changed the phrase according to the referees suggestion.

20

p. 14559, line 25 (and other places): Check for consistent use of either "form I Ru- bisCO" or
"type I RubisCO"

#### 23 OK, we checked and used the uniform terms Form I and II RubisCO

24

p. 14559, line 28: reword "cbbL 1A comprise obligate autotrophic bacteria"; this sounds odd
to me.

#### 27 Ok, we changed the sentence according to the referees suggestion

1	p. 14562, line 16: "unlabelled": up to now, no label was mentioned. Maybe: "To obtain
2	background values for the isotopic compositions"
3	This was rephrased according to the referee's suggestion
4	
5	p. 14564, equation 3: Check if this complex equation is printed correctly.
6	We checked the equation and compared to other published versions. It is correct
7	
8	p. 14564, line 20: Replace "Mofette" by "Soil"
9	Done
10	
11	p. 14567, line 13 (and other places): "anoxic restrictions": Should that be "anoxic conditions"?
12	We changed the paragraph according to referee #1.
13	
14	p. 14570, line 1 and 3: Not all of the values given here are consistent with Table 1.
15	The values were corrected, thank you for pointing this out.
16	
17	p. 14572, line 19: I could not find these numbers in the corresponding table.
18	This is true, for clarity we included solely values that were obtained after fumigation.
19	Values given in table 2 should illustrate the enrichment of the microbial biomass
20	compared to the background values.
21	
22	p. 14576, line 1: replace "alternation" by "alteration"
23	Done
24	
25	p. 14577, top paragraph: I suggest to focus on the decay corrected data here.

We think it is better to show both, the <sup>14</sup>C decay corrected and uncorrected preditions of <sup>13</sup>C. The predictions for using <sup>14</sup>C data uncorrected for radio-decay serve to illustrate the overall depletion of <sup>13</sup>C-SOM values compared to vegetation values. The decay corrected data in turn show, that after calibrating the model, <sup>13</sup>C- depletion is still occurring where we measure the greatest abundance and activity of autotrophic microorganisms.

Table 3: At least in the printed version, this table is difficult to read because the columns are
too small and therefore the averages and the standard deviations were printed in two lines.
Maybe a different layout would help.

10 **Done** 

11

12 General: The numbering of the Figures does not match their first occurrence in the text.

13 Apologies, this is now fixed.

- 1 List of relevant changes in the text:
- 2 (page numbers refer to pages in this document)

- 4 page 17, line 8
- 5 page 18, line 5-8
- 6 page 19, line 30 ff.
- 7 page 21, line 25 ff.
- 8 page 26, line 27 ff.
- 9 page 31, line 17 f.
- 10 page 32, line 21 ff.
- 11 page 33, line 5 ff.
- 12 page 34, line 10 ff.
- 13 page 35, line 31 ff.
- 14
- 15
- 16
- 17

# Autotrophic fixation of geogenic CO<sub>2</sub> by microorganisms contributes to soil organic matter formation and alters isotope signatures in a wetland mofette

4

## 5 M. E. Nowak<sup>1</sup>, F. Beulig<sup>2</sup>, J. von Fischer<sup>3</sup>, J. Muhr<sup>1</sup>, K. Küsel<sup>2</sup>, S. E. Trumbore<sup>1</sup>

6 [1]{Department for Biogeochemical Processes, Max-Planck Institute for Biogeochemistry,

- 7 Hans-Knöll Straße 10, 07745 Jena}
- 8 [2]{Aquatic Geomicrobiology, Institute of Ecology, Friedrich Schiller University Jena,
- 9 Dornburger Str. 159, 07743 Jena, Germany}
- 10 [3] {Department of Biology, Colorado State University, Fort Collins, CO 80523}
- 11 Correspondence to: M. Nowak (mnowak@bgc-jena.mpg.de)
- 12

#### 13 Abstract

To quantify the contribution of autotrophic microorganisms to organic matter formation (OM) in soils, we investigated natural CO<sub>2</sub> vents (mofettes) situated in a wetland in NW Bohemia (Czech Republic). Mofette soils had higher SOM concentrations than reference soils due to restricted decomposition under high CO<sub>2</sub> levels. We used radiocarbon ( $\Delta^{14}$ C) and stable carbon isotope ratios ( $\delta^{13}$ C) to characterize SOM and its sources in two moffetes and compared it with respective reference soils, which were not influenced by geogenic CO<sub>2</sub>.

The geogenic  $CO_2$  emitted at these sites is free of radiocarbon and enriched in <sup>13</sup>C compared to atmospheric  $CO_2$ . Together, these isotopic signals allow us to distinguish C fixed by plants from C fixed by autotrophic microorganisms using their differences in <sup>13</sup>C discrimination. We can then estimate that up to 27 % of soil organic matter in the 0-10 cm layer of these soils was derived from microbially assimilated  $CO_2$ .

Isotope values of bulk SOM were shifted towards more positive  $\delta^{13}$ C and more negative  $\Delta^{14}$ C values in mofettes compared to reference soils, suggesting that geogenic CO<sub>2</sub> emitted from the soil atmosphere is incorporated into SOM. To distinguish whether geogenic CO<sub>2</sub> was fixed by plants or by CO<sub>2</sub> assimilating microorganisms, we first used the

proportional differences in radiocarbon and  $\delta^{13}$ C values to indicate the magnitude of 1 2 discrimination of the stable isotopes in living plants. Deviation from this relationship was taken to indicate the presence of microbial CO<sub>2</sub> fixation, as microbial discrimination should 3 differ from that of plants. <sup>13</sup>CO<sub>2</sub>-labelling experiments confirmed high activity of CO<sub>2</sub> 4 assimilating microbes in the top 10 cm, where  $\delta^{13}$ C values of SOM were shifted up to 2 ‰ 5 towards more negative values. Uptake rates of microbial CO<sub>2</sub> fixation ranged up to  $1.59 \pm$ 6 0.16 ug gdw<sup>-1</sup> d<sup>-1</sup>. We inferred that the negative  $\delta^{13}C$  shift was caused by the activity of 7 8 chemo-lithoautotrophic autotrophic microorganisms using the Calvin Benson Basham Cycle, 9 as indicated from quantification of cbbL/cbbM marker genes encoding for RubisCO by 10 quantitative polymerase chain reaction (qPCR) and by acetogenic and methanogenic microorganisms, shown present in the moffettes by previous studies. Combined  $\Delta^{14}$ C and 11  $\delta^{13}$ C isotope mass balances indicated that microbially derived carbon accounted for 8 to 27 % 12 13 of bulk SOM in this soil layer.

14 The findings imply that autotrophic <u>micro</u>organisms can recycle significant amounts 15 of carbon in wetland soils and might contribute to observed <u>radiocarbon</u> reservoir effects 16 influencing  $\Delta^{14}$ Cradiocarbon signatures in peat deposits.

17

#### 18 **1 Introduction**

Microbial assimilation of CO<sub>2</sub> is a ubiquitous process in soils, and can be accomplished by a 19 20 wide variety of microorganisms using different metabolic pathways (Berg, 2011; Wood et al., 21 1941). RubisCO, the most important carboxylating enzyme for obligate and facultative 22 chemo- or photoautotrophic microorganisms that fix CO<sub>2</sub> using the Calvin Benson Bassham 23 Cylce (CBB) has been shown to be highly abundant in agricultural, forest and volcanic soils (Nanba et al., 2004; Tolli and King, 2005; Selesi et al., 2007). Direct uptake of CO<sub>2</sub> into 24 25 microbial biomass (MB) and soil organic matter (SOM) by photoautotrophic and 26 chemoautotrophic organisms has been measured in paddy rice and agricultural upland soils (Liu and Conrad, 2011; Wu et al., 2015; Wu et al., 2014), as well as under manipulating 27 28 experimental conditions, like H<sub>2</sub> amendment (Stein et al., 2005) or addition of reduced 29 sulphur compounds (Hart et al., 2013). Autotrophic acetogenic organisms, using the Wood-Ljungdahl Pathway for CO<sub>2</sub> fixation, are important groups in wetland and forest soils (Küsel 30 and Drake, 1995; Ye et al., 2014). In addition, many heterotrophic soil microorganisms fix 31 32 CO<sub>2</sub> in order to maintain their metabolic cycle by anaplerotic reactions, either to form new

sugars for cell wall synthesis or to excrete organic acids for nutrient mobilization (Feisthauer 1 2 et al., 2008; Miltner et al., 2005; Santruckova et al., 2005). Global estimates of microbial CO<sub>2</sub> fixation in soils range between 0.9 and 5.4 PgC per year (Yuan et al., 2012). However, it still 3 4 remains unclear how much of assimilated CO<sub>2</sub> is stored and contributes to the formation of 5 soil organic matter (SOM). In this study we aim at evaluating the impact of autotrophic 6 microorganisms on carbon isotope signatures of SOM. We further aim at quantifying the contribution of autotrophs to SOM by means of natural abundance <sup>14</sup>C and <sup>13</sup>C isotope 7 8 signatures in a unique environment.

9 Microbial utilization of  $CO_2$  and its incorporation into SOM is also potentially an 10 important mechanism influencing the isotope signatures of SOM (Ehleringer et al., 2000; 11 Kramer and Gleixner, 2006). Stable carbon ( $\delta^{13}C$ ) and radiocarbon ( $^{14}C$ ) isotope signatures 12 are important tools for determining turnover of soil organic matter and dating ancient 13 sediments (Balesdent et al., 1987; Hughen et al., 2004; Trumbore, 2000).

14 Stable isotope variations in soil reflect mass-dependent fractionation processes (Werth and Kuzyakov, 2010). In many well-drained soils, there is a well-documented increase in  $\delta^{13}C$ 15 with depth that has been variously attributed to selective preservation/decomposition of 16 different components of organic matter, recent declines in atmospheric  $\delta^{13}$ C due to the Suess 17 effect, or microbial fractionation (summarized in Ehleringer et al. 2000). Enzymatic 18 fractionation during assimilation of CO<sub>2</sub> can also lead to changes in  $\delta^{13}$ C values of 19 synthesized organic matter (Haves, 2001; Robinson and Cavanaugh, 1995; Whiticar, 1999). 20 Carboxylation processes by heterotrophic microorganisms have been hypothesized to be 21 responsible for the increase in  $\delta^{13}$ C values with depth in aerated upland soils (Ehleringer et 22 al., 2000). 23

Radiocarbon signatures reflect the time elapsed since the C being measured was fixed 24 from the atmosphere, and are corrected (using measured  $\delta^{13}C$  values) to remove mass 25 dependent fractionation effects. The radiocarbon signature of CO<sub>2</sub> in soil pore space can be 26 depleted or enriched in <sup>14</sup>C compared to organic matter found at the same depth, depending on 27 the age of C being mineralized (Trumbore, 2006). Because soil pore space CO<sub>2</sub> can have quite 28 different isotopic signatures compared to SOM at the same depth, microbial assimilation of 29 CO<sub>2</sub> may influence SOM <sup>14</sup>C signatures and therefore bias estimates of carbon turnover and 30 31 radiocarbon age by generating reservoir effects (Pancost et al., 2000).

In turn, comparing both, radiocarbon and stable isotope values of SOM, MB and their sources might allow quantifying the <u>potential</u> contribution of autotrophic organisms to SOM, because a mismatch of both isotopes in quantifying SOM sources indicates either fractionation of <sup>13</sup>C by carboxylation processes of different enzymes or depletion or enrichment of <sup>14</sup>C by the use of soil CO<sub>2</sub> (Kramer and Gleixner, 2006).

6 In order to test the hypothesis that microbial CO<sub>2</sub> fixation contributes to SOM 7 formation and alters isotope signatures in soil depth profiles, we investigated wetland 8 mofettes in NW Bohemia. Mofettes are cold exhalations of geogenic CO<sub>2</sub> from wetland soils with high CO<sub>2</sub> concentrations. The exhaling volcanic-derived CO<sub>2</sub> has a distinct isotopic 9 signature, is enriched in  $\delta^{13}$ C by about 5 ‰ and free of radiocarbon compared to atmospheric 10 11 CO<sub>2</sub>. This unique feature allows us to use geogenic CO<sub>2</sub> as a natural isotopic tracer, because CO<sub>2</sub> assimilating microorganisms take up an isotopically different CO<sub>2</sub> source compared to 12 13 plants growing in the area, which use a mixture of geogenic and atmospheric CO<sub>2</sub>. We used 14 three approaches to evaluate the importance of CO<sub>2</sub> fixation for SOM generation in mofettes 15 and its impact on carbon isotope values:

- 16 1) We measured natural abundance <sup>13</sup>C and radiocarbon signatures of SOM, CO<sub>2</sub> and 17 plant material in mofette and reference soils, in order to identify areas where C 18 derived from microbial CO<sub>2</sub> fixation altered isotope signatures of bulk SOM from 19 expected plant signals and quantified C derived from microbial CO<sub>2</sub> fixation by 20 isotope mass balances.
- 2) We conducted isotope-labelling experiments with <sup>13</sup>CO<sub>2</sub> in order to quantify the rate of
   CO<sub>2</sub> fixation by microorganisms in soil profiles of two CO<sub>2</sub> vents and compared these
   to reference soils away from the vents.
- 24 3) We complemented existing data about microbial community and activity in wetland 25 mofettes (Beulig et al., 2014), by assessing the importance of-microorganisms using the Calvin Benson Basham Cycle for CO<sub>2</sub> fixation<del>chemolithoautotrophic</del> 26 27 microorganisms. This waswas especially important to infer whether differences in 28 kinetic isotope effects compared to plants were feasible given the pathways of 29 microbial C fixation. Therefore, we quantified cbbL and cbbM marker genes encoding 30 for Form I and II RubisCO, respectively. Form I RubisCO consists of eight small and 31 eight large subunits. It can be subdivided into two groups, the "red" and "green" like groups, which can be further subdivided into Form 1A, 1B and 1C and 1D, 32

1 respectivelythe dominating forms in soils by gPCR (Yuan et al., 2012; Tolli and King, 2 2005). Form II RubisCO consists only of large subunits. Because of its low CO<sub>2</sub> affinity and high O<sub>2</sub> sensitivity, it represents an early form, evolved under anaerobic 3 conditions and high CO<sub>2</sub> concentrations (Alfreider et al., 2003). Form II RubisCO 4 5 might be favourable under conditions prevailing in mofettes. We investigated genes of two subclasses of Form I (cbbL 1A and cbbL 1C), as well as cbbM, encoding for 6 7 From II RubisCO. cbbL 1A was identified mainly in comprise obligate autotrophic 8 bacteria and cbbL 1C in facultative autotrophic bacteria (Tolli and King, 2005). cbbM 9 encodes for autotrophic organisms living under anaerobic conditionsrestriction (Selesi 10 et al., 2005).

Using this information, we aimed to quantify the amount of C derived from microbial assimilation of  $CO_2$  into soil organic matter within soil profiles, and assess its potential to alter isotope signatures of SOM.

#### 14 **2** Materials and methods

#### 15 **2.1 Site description**

The study site (50°08'48'' N, 12°27'03''E) is located in the northwestern part of the 16 17 Czech Republic (Bohemia). The area is part of a continental rift system, where deep tectonic 18 faults provide pathways for ascending gases and fluids from the upper earth's mantle (Kämpf 19 et al., 2013). Mofettes are surficial, low temperature exhalations of mantle derived CO<sub>2</sub>. 20 Macroscopically, they form a complex of landscape features. At centre is a spot of typically 0.5 to 1 meter bare soil. From this central spot, almost pure CO<sub>2</sub> emanates to the atmosphere. 21 22 The mofette centre is surrounded by a raised hummock that extends 1 to 20 m away from the 23 spot. The investigated mofettes are situated on the floodplain of the river Plesna and are part of a wetland. Geogenic CO<sub>2</sub> emanates with an average discharge of up to 0.62 tons CO<sub>2</sub>  $d^{-1}$ 24 25 per spot (Kämpf et al., 2013). The surrounding hummock is built up by different vascular 26 plant communities. Eriophorum vaginatum and Deschampsia cespitosa are dominating plant 27 species in the immediate proximity of the central vent and hummock structure, respectively. 28 Filipendula ulmaria represents typical floodplain vegetation.

We investigated two mofettes that differed in size. Mofette 1 had a spot-diameter of 0.6 m, whereas the diameter of Mofette 2 was 1.5 m. We also sampled soils away from the influence of the mofette-exhaled CO<sub>2</sub> (deemed reference soils). These soils are vegetated and

experience periodic anoxic conditions due to waterlogging, as evidenced by gleved soil 1 2 features and porewater geochemistry (Mehlhorn et al., 2014). In Mofettes 1 and 2, the local water table is elevated by ascending  $CO_2$  and  $O_2$  is mainly displaced by the  $CO_2$  stream, 3 4 leading to anoxic (but not necessarily water-logged) conditions (Bräuer et al., 2011). 5 According to the World Reference Base for soil resources (WRB, 2007), mofette soils are 6 characterized as Histosols with pronounced reductomorphic features (reduced Y horizons) 7 due to the influence of up-streaming CO<sub>2</sub>. Reference soils are classified as 'glevic' Fluvisols 8 (Beulig et al., 2014).

# 9 2.2 Sampling of soils, plants and gases for bulk geochemical and isotope 10 measurements

Soil and plant samples were acquired between November 2013 and September 2014. 11 For bulk  $\delta^{13}$ C and radiocarbon analyses soil cores were taken from the central, unvegetated 12 part of the mofette structure and reference soils. Reference soils lacking CO<sub>2</sub> emissions were 13 14 identified with a portable landfill gas analyser (Visalla GM70 portable CO<sub>2</sub> sensor) in close proximity to each vent structure. Reference soils 1 and 2 were defined 5 and 18 meters distant 15 16 from the central vent structures, respectively. Samples for bulk stable isotope and radiocarbon 17 analyses were taken in November 2013. In order to account for soil heterogeneity, three soil 18 cores (I.D. 5 cm) were taken from a plot of 50 x 50 cm from mofette and reference soils. 19 Because mofette and reference soils were characterised by very different soil features, soil 20 cores were not divided according to horizons, but depth intervals. Based on visual inspection, 21 soil cores were divided into depth intervals from 0-10 cm, 10-25 cm and 25-40 cm. Replicates 22 of the respective depth intervals were mixed and sieved to 2 mm. Roots and plant debris were removed by handpicking. The sieved soil was subsequently dried at 40° and prepared for 23 24 stable isotope, radiocarbon and C/N analysis.

In April 2014, vVegetation samples were taken from the same plot as soil cores, in order to characterize the isotopic composition of the plant material, contributing to mofette SOM. in April 2014. Vegetation samples in the direct proximity of both mofettes were represented by *Eriophorum vaginatum*. At each Additionally to vegetation sampling in the direct vicinity of the mofette, vVegetation samples were also taken by clipping plants at 2cm height at 2 meter intervals along a transect that crossed moffett 2, allowing us to test along a  $CO_2$ -gradient, in order to test weather thehow the isotope signalsignatures ( $\delta^{13}C$  and  $\Delta^{14}C$ ) of plants follows a linear relationship with increasingchanged with different mixtures of ambient

and geogenic CO<sub>2</sub>, concentrations and can therefore be used for a linear mixing model 1 2 sources.- Therefore, point, the dominant plant species was collected by clipping the plant at 2 em height, vVegetation was sampled at 2 meter intervals along a 20 m transect that transected 3 mofette 2. Mofette 2 representsis an exposed hummock, without disturbance of 4 5 smallerdominated by a un-vegetated central region of CO<sub>2</sub> exhalation (WITH NO <u>VEGETATION???</u>). <u>1One to to 2 meters away</u>distant from the central exhalation, the 6 7 dominant plant species changed was to-Deschampsia cespitosa, -and subsequently at greater 8 distances the dominant plant was-to Filipendula ulmaria, most distant from the central 9 mofette structure. At every sampling point, the dominant plant species was collected by clipping the plant at 2 cm height. The collected samples were dried at 40° C, ground and 10 11 prepared for stable isotope, radiocarbon and C/N analysis.

12 CO<sub>2</sub> was sampled from the centre of each mofette by filling 250 ml evacuated stainless 13 steel cylinders through a perforated lance from four different soil depths (5, 15, 25, 40 cm), in 14 order to determine its radiocarbon and stable isotope signature.

### 15 **2.3** Soil sampling for <sup>13</sup>CO<sub>2</sub> labelling experiments

Mofette soils were sampled for two labelling experiments in November 2013 and September 2014, respectively. For the first experiment, 10 x 10 cm soil monoliths, extending to 10 cm depth were sampled from each soil in November 2013. After removing the top of the Oh horizon (about 1 cm thickness), the remaining material was divided into three subsamples. Each replicate was homogenized within a sterilized plastic bag, put under an anoxic N<sub>2</sub> atmosphere and cooled at 4° until further processing in the lab within the same day.

For a second experiment, three soil cores (I.D. = 5 cm) were taken from 0 to 40 cm of each mofette and reference soil and subsampled from 0-5, 5-10 10-20, 20-30 and 30-40cm. 5g subsamples from each core were transferred immediately after core recovery to a sterilized 12 ml Labco<sup>®</sup> Exetainer, flushed with N<sub>2</sub> to preserve anoxia, sealed and brought to the laboratory at 4°C for further processing. To obtain background (i.e. with no influence of added label) values for isotopic compositions, oneAnother unlabelled set of subsamples was prepared-dried and prepared for TOC, C/N, pH and  $\delta^{13}$ C analyses as described above.

#### **2.4 Sampling for DNA extraction**

Samples for DNA extraction were taken in May 2014 from Mofette 1 and Reference 1. Samples were taken from 0-5, 5-10, 10-20, 20-30 and 30-40 cm. Three replicates of 30 g were sampled from each depth, and homogenized under anoxic conditions. Subsequently, subsamples of 5 g were transferred to 50 ml tubes, cooled with dry ice and transported under an Ar atmosphere to the laboratory for molecular analyses.

## 7 2.5 Analyses of geochemical parameters and natural abundance isotope 8 signatures of vegetation and soil samples

Soil pH was determined in a 0.01 M CaCl<sub>2</sub> solution with a soil:solution ratio of 1:2.5
using a WTW pH meter. The precision of pH measurements was better than 0.1 (n=3). Total
C and N concentration of soil and plant samples were determined on a "Vario EL" (Elementar
Analysesysteme GmbH, Germany). Gravimetric water content was determined after drying
soils for 48h at 105° and C and N content are reported per g dry soil weight.

Stable C isotope signatures of bulk soil and plant samples were determined on an isotope ratio mass spectrometer (DELTA+XL, Finnigan MAT, Bremen, Germany) coupled to an elemental analyser (NA 1110, CE Instruments, Milan, Italy) via a modified ConFloII<sup>TM</sup> interface (EA-IRMS). Stable carbon isotope ratios are reported in the delta notation that expresses <sup>13</sup>C/<sup>12</sup>C ratios as  $\delta^{13}$ C-values in per mil (‰) relative to the international reference material Vienna Pee Dee Belemnite (V-PDB, Coplen et al., 2006):

$$20 \qquad \delta^{13}C = \left[\frac{\frac{^{13}C}{^{12}C}}{\frac{^{13}C}{^{12}C}} - 1\right] \times 1000 \tag{1}$$

21 Analytical precision of all samples was better than 0.1 ‰.

22 For discussing microbially mediated isotope effects the isotope discrimination value  $\Delta$ 23 is used, which expresses the isotopic difference between two compounds in ‰:

 $24 \qquad \Delta_{x-y} = \delta_x - \delta_y \tag{2}$ 

25 Where  $\delta_x$  and  $\delta_y$  refer to  $\delta^{13}$ C values of the product and reactant, respectively.

1 The radiocarbon content of soil and plant samples was determined by accelerator mass 2 spectrometry at the Jena <sup>14</sup>C facilities (Steinhof et al., 2004). Subsamples of soil containing 1 mg of carbon were combusted quantitatively and the developed CO<sub>2</sub> was catalytically reduced 3 4 to graphite at 625°C by H<sub>2</sub> reduction. To simplify comparison with stable isotope ratios, radiocarbon activities are reported in  $\Delta^{14}$ C, which is the % deviation of the  ${}^{12}$ C/ ${}^{14}$ C ratio from 5 the international oxalic acid universal standard. The  $\Delta^{14}$ C value of the sample is corrected for 6 7 mass dependent isotope fractionation to a common value of -25 ‰ (Mook and van der Plicht, 8 1999). The standard is corrected for radioactive decay between 1950 and the year (y) of the 9 measurement (2014).

10 
$$\Delta^{14}C = \left[\frac{\frac{{}^{14}C}{{}^{12}C}_{sample,-25}}{0.95\frac{{}^{14}C}{{}^{12}C}_{OX1,-19}} \times \exp^{\left(y-\frac{1950}{8267}\right)}\right] \times 1000$$
(3)

Errors reported for radiocarbon measurements represent the analytical error of homogenized mixed samples in ‰. Analytical precision of all radiocarbon measurements was better than 3 ‰.

#### 14 **2.6 Labelling experiments**

15 The first labelling experiment traced the flow of fixed CO<sub>2</sub> directly into microbial biomass (MB), evaluated rates of CO<sub>2</sub> uptake associated with biological activity and 16 compared the proportion of labelled MB in mofettes with reference soils. From each field 17 18 replicate sample, 20 g aliquots were taken and put into sterilized 120 mL boro-silicate bottles with butyl rubber stoppers inside a glove box containing an N<sub>2</sub> atmosphere. From these 19 subsamples, three replicates were prepared for incubation with <sup>13</sup>CO<sub>2</sub>. In order to obtain 20 21 control samples without biological activity, an additional aliquot of each sample was prepared 22 and autoclaved for 2 hours at 160° and 60 bar.

23 SoilMofette samples were incubated under anoxic conditions with  ${}^{13}CO_2$  at N<sub>2</sub>:CO<sub>2</sub> 24 ratios equivalent to those experienced by the soils in the field: mofette soils were incubated 25 with a 100 vol. %  ${}^{13}CO_2$  atmosphere using sterile techniques and reference soils were 26 incubated with a 10 vol. %  ${}^{13}CO_2$  and 90 vol. % N<sub>2</sub> atmosphere. In order to account for soil 27 respiration and to maintain a constant label, the headspace of every sample was removed and 1 renewed every 3 days. The samples were incubated for 14 days in the dark at  $120^{\circ}$ C. Living 2 and autoclaved control samples were treated identically.

3 After 14 days, the jars were flushed with N<sub>2</sub> and the soil samples were homogenized and split. One part was air dried for bulk <sup>13</sup>C analysis and the other part was prepared for 4 extraction of the microbial biomass C by chloroform fumigation extraction (CFE) (Vance et 5 6 al., 1987). CFE extracts microbial biomass C by lysing the cells with chloroform and 7 releasing the products of cell lysis into a salt solution as dissolved organic carbon (DOC). In 8 order to enhance extraction efficiency and to minimize the losses for extracted C by microbial 9 degradation, the protocol from Vance et al. (1987) was slightly modified (Malik et al., 2013). The concentration of dissolved microbial biomass C (MB-DOC) and its stable carbon isotope 10 11 ratio were determined by a high performance liquid chromatography system coupled to an IRMS (HPLC/IRMS) system (Scheibe et al., 2012). This method allows direct determination 12 13 of concentration and carbon isotopic value of DOC in the liquid phase by coupling a LC-14 IsoLink system (Thermo Electron, Bremen, Germany) to a Delta+ XP IRMS (Thermo Fisher 15 Scientific, Germany). A detailed description of the apparatus and measurement procedure is 16 given in Scheibe et al. (2012).

17 The amount of microbial biomass was determined by subtracting the amount of MB-18 DOC of un-fumigated samples from MB-DOC of fumigated samples and dividing with a 19 proportionality factor  $K_c$  that accounts for the extraction efficiency:

$$20 C_{mic} = \frac{DOC_{fum} - DOC_{unfum}}{K_C} (4)$$

A value of 0.45 was used for *Kc* according to Amha et al. (2012). The isotope ratio of microbial biomass C can be derived by applying an isotope mass balance:

23 
$$\delta^{13}C_{MB} = \frac{\delta^{13}C_{fum} \times C_{fum} - \delta^{13}C_{unfum} \times C_{unfum}}{C_{fum} - C_{unfum}}$$
(5)

The net  $CO_2$  fixation rate was calculated by determining the increase in <sup>13</sup>C from the label compared to the unlabelled control, and is normalized for C content (either total soil or microbial-C). The excess <sup>13</sup>C can be derived from the <sup>13</sup>C/<sup>12</sup>C ratio of the sample before and after the labelling:

1 
$$ExcessC[mg] = \frac{{}^{13}C_{labeled}}{{}^{12}C_{labeled}} \times C_{sample}[mg] - \frac{{}^{13}C_{unlabeled}}{{}^{12}C_{unlabeled}} \times C_{sample}[mg]$$
(6)

2 The  ${}^{13}C/{}^{12}C$  ratio can be obtained from the measured  $\delta^{13}C$  as follows:

3 
$$\frac{{}^{13}C}{{}^{12}C}_{sample} = \left(\frac{\delta^{13}C_{measured}}{1000} + 1\right) \times 0.011237$$
 (7)

4 where 0.01123 is the  ${}^{13}C/{}^{12}C$  ratio of the international V-PDB standard (Craig, 1957).

5 A second labelling experiment was performed in order to obtain uptake rates as a function of depth for mofette and reference soils. After sampling 5 g of soil into 12 ml 6 Labco<sup>®</sup> Exetainers as described above, mofette samples were flushed with 100 vol. % <sup>13</sup>CO<sub>2</sub>, 7 and reference soils with 10 vol. % 13CO2 and 90 vol. % N2. Soils were incubated for 7 days in 8 9 the dark at 12°C. The headspace of all samples was exchanged after 3 days of incubation. After 7 days, vials were opened and flushed with N<sub>2</sub> for 2 min and evacuated to remove any 10 sorbed or dissolved <sup>13</sup>CO<sub>2</sub>. Soil samples were subsequently air dried at 60°C and prepared for 11 bulk <sup>13</sup>C analysis as described above. The measured enrichment in <sup>13</sup>C was used to measure 12 13 uptake rates according Eq. (6).

#### 14 2.7 DNA extraction and quantitative PCR

Total nucleic acid extractions of 0.7 g homogenised soil from mofette 1 and reference 16 | 1 were performed <u>in triplicates</u> according to the protocol of Lueders et al. (2004). Co-17 extracted organic soil compounds were removed by sequential purification with gel columns 18 (S-400 HR; Zymo Research, Irvine USA) and silica columns (Powersoil Total RNA Kit in 19 combination with the DNA Elution Accessory kit; MO BIO Laboratories, Carlsbad CA). 20 Nucleic acid extraction efficiency was checked by agarose gel electrophoresis.

Copy numbers of 16S rRNA, cbbL 1A, cbbL 1C and cbbM genes in extracted DNA were determined using quantitative PCR (qPCR). qPCR was performed on a Mx3000P instrument (Agilent, Santa Clara, CA, USA) by using <u>Maxima SYBR Green Mastermix</u> (Thermo Scientific) and the primer combinations Uni-338 F-RC and Uni-907 R (16S rRNA, (Weisburg et al., 1991), F-cbbM and R-cbbM (cbbM, (Alfreider et al., 2003)), F-cbbL and RcbbL (cbbL IA, (Alfreider et al., 2003)) as well as F-cbbL IC and R-cbbL IC (cbbL 1C, (Alfreider et al., 2003)) as described by Herrmann et al. (2012). Cycling conditions for 16S

rRNA genes as well as cbbL and cbbM genes consisted of denaturation for 10 min at 95°C, 1 2 followed by 50 cycles with 4 temperature steps (1. 95°C at 30 s; 2. 55 and 57°C at 30 s for cbbL and cbbM/16S rRNA genes, respectively; 3. 72°C at 45 s; 4. data acquisition at 78°C 3 and 15 s). Standard curves were constructed using plasmid CB54 for 16S rRNA and standard 4 5 curves for cbbL and cbbM marker genes were constructed from ten times dilution series of mixtures of plasmids containing cbbL and cbbM inserts, obtained from Herrmann et al. 6 (2015). PCR inhibitors were tested by ten times dilution series of representative samples. For 7 8 the investigated samples 5 ul of DNA was taken as template for gene copy quantification of 9 16S rRNA, cbbL and cbbM. The quantified functional marker genes are indicative for nitrifiers and sulphur-oxidizers (cbbL 1A), photosynthetic organisms (cbbl 1C) as well as 10 chemo- and phototrophic organisms living solely under anoxic restrictions (cbbM) (Selesi et 11 al., 2005). 12

#### 13 **2.8 Mass balance calculations**

The unique isotopic composition of geogenic CO2 and combined measurements of 14 radiocarbon and stable isotopes allows identification of plant and microbial end-members for 15 quantifying the importance of these two sources of SOM. Geogenic CO<sub>2</sub> ( $\Delta^{14}$ C = -1000‰, 16  $\delta^{13}C = -2$  ‰) is quite different from atmospheric CO<sub>2</sub> ( $\Delta^{14}C \sim +20$ ‰,  $\delta^{13}C = -7$  ‰) in both 17 isotopes. Therefore,  $\Delta^{14}$ C values can be used to determine the overall fraction of geogenic 18 CO<sub>2</sub> that is assimilated by plants or microorganisms in the mofette by using the end-members 19  $\Delta^{14}C_{geogenic CO2}$  and  $\Delta^{14}C_{air}$ . A conventional mixing model for determining the fraction of 20 geogenic CO<sub>2</sub> in SOM can be calculated according to: 21

22 
$$SOM_{geogenic} \left[\%\right] = \frac{\Delta^{14} C_{SOM} - \Delta^{14} C_{air}}{\Delta^{14} C_{geogenicCO2} - \Delta^{14} C_{air}} \times 100$$
(8)

This mass balance assumes that changes in  $\Delta^{14}C_{SOM}$  caused by radioactive decay of  ${}^{14}C$  are small compared to contributions from geogenic CO<sub>2</sub>.

The same mass balance can be applied for calculating the fraction of geogenic CO<sub>2</sub> with stable isotope values. The end-members for this calculation are  $\delta^{13}$ C values of plants, which grew solely on geogenic CO<sub>2</sub> or solely on ambient air CO<sub>2</sub>. Plant  $\delta^{13}$ C values are expected to be around 20 ‰ depleted in <sup>13</sup>C compared to the respective CO<sub>2</sub> source due to enzymatic fractionation, which has to be considered in determining the  $\delta^{13}$ C end-member value. 1 We used the correlations between  $\delta^{13}C$  and  $\Delta^{14}C$  of plant material to prove that 2 enzymatic discrimination of plants is constant in the vicinity of the mofette, despite 3 potentially fluctuating CO<sub>2</sub> concentrations. If  $\Delta^{14}C$  and  $\delta^{13}C$  values of plants show a linear 4 correlation,  $\Delta^{14}C$  values of SOM can be used to derive  $\delta^{13}C$  values that should be expected, if 5 the organic matter is solely derived from plants according the mixing model:

$$6 \qquad \delta^{13}C_{model} = \delta^{13}C_{plant\_geo} \times \left(\Delta^{14}C_{SOM\_mofette} \times m + t\right) + \delta^{13}C_{plant\_air} \times \left(1 - \left(\Delta^{14}C_{SOM\_mofette} \times m + t\right)\right)$$
(9)

where  $\delta^{13}C_{plant_{geo}}$  and  $\delta^{13}C_{plant_{air}}$  are the measured plant input end-members exhibiting the 7 8 most depleted (i.e. highest exposure to geogenic CO<sub>2</sub>) and most enriched (exposure to atmospheric CO<sub>2</sub>)  $\Delta^{14}$ C values, respectively.  $\Delta^{14}C_{SOM\_mofette}$  are measured radiocarbon values 9 at a certain depth within the mofette soil. m and t are the slope and intercept of the regression 10 between measured  $\delta^{13}C$  and  $\Delta^{14}C$  plant values. The model calculates the  $\delta^{13}C_{SOM}$  that 11 corresponds to measured  $\Delta^{14}C_{SOM}$  values, if all SOM would be derived from plant material. 12 13 Deviation from the model indicates input of C sources other than plants with distinct isotopic 14 compositions.

#### 15 2.9 Statistical analyses

16 Reported results (e.g.  $\delta^{13}$ C values, microbial biomass), represent the mean of three 17 independent replicates. Uncertainties reported for radiocarbon data represent analytical 18 precision of a homogenised sample comprised of three independent soil cores. Differences of 19  $\delta^{13}$ C and  $\Delta^{14}$ C values in mofette and reference soils as well as between soil depth intervals 20 were analysed using Student's t-test. Significant differences are reported at p< 0.05.

21

#### 22 3 Results

#### 23 3.1 pH, bulk TOC and C/N

Soil pH ranges from 3.0 to 3.5 in mofette soils and is higher in reference soils (averaging 4.4), without significant trends with depth (Table 1). Total organic carbon (TOC) contents are high (~12 - 20% C) in the surface 5 cm of both mofette and reference soils. In the reference soil, TOC decreases with depth to concentrations of 3 % C below 20 cm. In contrast, 1 TOC concentrations in both mofettes decrease below 5 cm (~6 to 16 %) and increase
2 subsequently to more than 30 % below 20 cm.

Organic matter quality as indicated by C/N ratio also highlights differences between mofette and reference soils. High C/N ratios ranging from 25 to 30 are found below 20 cm depth in both mofettes, whereas C/N ratios decrease rapidly as low as 16.5 to 9 (for mofette 1 and 2, respectively) in the upper 10 cm (Table 1). In both reference soils, C/N ratios remain constant throughout the profile at 10 to 14 (Table 1).

#### 8 3.2 Radiocarbon and stable isotope ratios of bulk SOM, plants and CO<sub>2</sub>

9 Consistent with our expectation, we found that geogenic CO<sub>2</sub> is free of radiocarbon (-10 1000 ‰) and has an average  $\delta^{13}$ C value of -2.36 ± 0.6 ‰.

11 Measured <u>R</u>radiocarbon concentrations of SOM for <u>SOM range between -550 ‰ and</u> 12 <u>800 ‰</u> in both mofettes are generally more depleted <u>by several hundred</u> ‰ <u>than-compared to</u> 13 reference soils <u>(table 1)</u>. In reference soils,  $\Delta^{14}$ C values decrease uniformly with depth from -14 60 ‰ and -34 ‰ in the top 10 cm to values of -280 ‰ and -163 ‰ at 40 cm depth in 15 reference soil 1 and 2, respectively, reflecting radioactive decay (table 1).

16  $\delta^{13}C_{SOM}$  in mofettes has an average values of  $-26.99 \pm 0.33$  ‰ and  $-26.38 \pm 0.54$  ‰ in 17 mofette 1 and 2, respectively. In both mofettes  $\delta^{13}C_{SOM}$  decreases slightly (but not 18 significantly) below 20 cm depth (p = 0.39 and 0.49 in mofette 1 and 2, respectively) (table 19 1). Both reference soils have  $\delta^{13}C_{SOM}$  of  $-28.08 \pm 0.4$  ‰ with no distinct depth trend in 20 reference 1 (p = 0.96) and a slight but not significant decrease in reference 2 (p = 0.35) below 20 cm. At every depth, reference soils are 1 to 2 ‰ depleted in <sup>13</sup>C compared to mofette 22  $\delta^{13}C_{SOM}$  throughout the soil profile (p < 0.05) (table 1).

23 Carbon isotope signatures in vegetation samples surrounding the mofette range from - $29.95 \pm 0.16$  ‰ to  $-23.81 \pm 0.30$  ‰ in  $\delta^{13}$ C and from -10.3 ‰ to -807.7 ‰ in  $\Delta^{14}$ C. 24 Variations in the two isotopes are highly correlated, and plants with most positive  $\delta^{13}C$  and 25 most negative  $\Delta^{14}$ C were found closest to the mofette and vice versa (figure 1). The linear fit 26 to the strong ( $R^2 = 0.86$ ) relationship between <sup>13</sup>C and <sup>14</sup>C found in vegetation material 27 28 (fFigure 1) is used to determine parameters for the mixing model (Eq. 9). The intercept of the line with the y-axis yields a value of -22.79 ‰ and represents the  $\delta^{13}$ C end-member value of 29 plant material which is fully labelled with geogenic CO<sub>2</sub> ( $\delta^{13}C_{plant geo}$ ), or t in 30

Eq. (9)). For the other endmember,  $\delta^{13}C_{plant\_air}$ , we used the  $\delta^{13}C$  value of plants from the reference site that exhibited the most positive  $\Delta^{14}C$  value, which yields  $\delta^{13}C_{plant\_air}$  of -29.15 %0. The corresponding  $\Delta^{14}C$  value, i.e. the value closest to atmospheric radiocarbon concentrations, was -10.3 ‰ (=  $\Delta^{14}C_{plant\_air}$ ). This is less than  $\Delta^{14}C$  measured in CO<sub>2</sub> in clean background air in the year of sampling (~+20‰) and indicates either that the reference site experiences some influence of geogenic CO<sub>2</sub> or the influence of local fossil fuel release in the region.

8 The slope of the relationship fit to plant samples (m in Eq. (9)) is what would be 9 expected for a linear mixture of plant material of the two end-member atmospheres (pure geogenic and pure air). Plant derived SOM would be expected to fall with this mixing line. 10 The majority (71 %) of reference soil values are within the 95 % confidence interval of this 11 expected slope  $\delta^{13}C/\Delta^{14}C$  relationship of plants (figure 1). In reference soils, but in general 12 have relatively constant <sup>13</sup>C values, while <sup>14</sup>C declines with soil depth, while <sup>13</sup>C remains 13 nearly constant. Mofette SOM generally has lower <sup>13</sup>C values than would be expected if they 14 had the same linear relationship as plant material, and <sup>14</sup>C signatures are all much lower than 15 those of the reference soil (figure 1). Only 5 % of mofette SOM values fall within the 95 % 16 17 confidence interval of the regression line.

18

#### 3.3 Mass balance calculations

19 Radiocarbon signatures of SOM indicate that, on average, 55 to 65 % of carbon accumulated in the mofette is derived from geogenic CO<sub>2</sub> (assuming end-members of -10 ‰ 20 for  $\Delta^{14}$ C air and -1000‰ for  $\Delta^{14}$ C geogenic CO<sub>2</sub>). The calculated proportion increases with 21 depth. By doing the same mass-balance calculation with  $\delta^{13}$ C values, (with - 22.47 ‰ as 22 geogenic CO<sub>2</sub> end-member and - 29.15 ‰ as reference end-member), one obtains lower 23 proportions of 34 - 44 % geogenic C compared to the radiocarbon mass balance. Thise 24 mismatch in quantifying the proportion of geogenic C suggests that  $\delta^{13}C_{SOM}$  values differ 25 from what we would expect if they were completely derived from plant inputs. 26

Equation (9) can be used to predict  $\delta^{13}$ C SOM values corresponding to measured radiocarbon values, assuming that all carbon would be derived from unaltered plant material. Calculated  $\delta^{13}$ C<sub>SOM</sub> values are 1-2 ‰ more positive at all depths (p < 0.05) compared to 1 observations (figure 23 B), i.e. measured  $\delta^{13}C_{SOM}$  values are depleted in  $^{13}C$  compared to a 2 signal that would be expected, if SOM would have preserved its original plant  $\delta^{13}C$  signature.

3

#### 3.4 Quantification of microbial CO<sub>2</sub> fixation activity

4 The analysis of bulk SOM and plant material revealed that mofette and reference soils 5 are distinct in their radiocarbon as well as stable isotope values, indicating incorporation of 6 geogenic CO<sub>2</sub> into mofette SOM either by plants or by microorganisms. Both isotopes show a bias in quantifying the amount of SOM derived from geogenic CO<sub>2</sub> by the same isotope mass 7 8 balance, which suggests the presence of another source of carbon than plants, presumably microorganisms, that depletes  $\delta^{13}$ C values. CO<sub>2</sub> fixing microorganisms might be a potential 9 source with a distinct  $\delta^{13}$ C value. In order to assess the activity of CO<sub>2</sub> fixing microorganisms 10 as well as their spatial distribution along the soil profile, we conducted two isotope-labelling 11 12 experiments.

In the first experiment we traced <sup>13</sup>CO<sub>2</sub> directly into microbial biomass (MB) within 13 the first 10 cm of the soil profile. After incubating the soils with <sup>13</sup>CO<sub>2</sub>, MB within all soils 14 showed high enrichment in <sup>13</sup>C, except in autoclaved control soils. Microbial biomass extracts 15 of autoclaved controls had  $\delta^{13}$ C values ranging between -24.10 ± 0.38 to -27.55 ± 0.14 ‰, in 16 both, furnigated and unfurnigated samples, which is close to bulk  $\delta^{13}$ C values obtained from 17 bulk soil measurements (table 2). This confirms that mainly biological processes mediated 18 CO<sub>2</sub> incorporation. In un-sterilized samples, unfumigated extracts showed enrichment in <sup>13</sup>C 19 in all mofette and reference soils. The  $\delta^{13}$ C of unfumigated samples ranged from -14.29 ± 0.8 20 % to  $+80.47 \pm 9.46$  % and are therefore enriched in <sup>13</sup>C compared to controls (p < 0.05). 21 However, in all cases <sup>13</sup>C enrichment was higher after fumigation (p < 0.05).  $\delta^{13}$ C values of 22 fumigated samples ranged between  $143.76 \pm 3.93$  ‰ and  $227.04 \pm 2.63$  ‰. 23

The calculated rate of CO<sub>2</sub> uptake expressed per gram microbial biomass in the top 10 cm of soil (table 2) was higher in mofettes compared to reference soils (p < 0.05) ranging between 287 ± 85 and 271 ± 58 ug<sup>-1</sup> gMB<sup>-1</sup> d<sup>-1</sup> in mofettes compared to 139 ± 32 and 99 ± 36 ug<sup>-1</sup> gMB<sup>-1</sup> d<sup>-1</sup> in reference soils (table 2).

The second labelling experiment measured  $CO_2$  fixation activity along the whole soil profile with samples taken from depth intervals between 1 to 40 cm. Tracer uptake was measured only in bulk SOM. In both soils, uptake rates decrease with depth (figure 32). In the top 5 cm, uptake rates were higher in mofette soils compared to reference soils. Below 20 cm, 1 rates decrease to values of  $0.14 \pm 0.03$  ug gdw<sup>-1</sup> d<sup>-1</sup> in both mofettes and  $0.09 \pm 0.02$  ug gdw<sup>-1</sup> 2 d<sup>-1</sup> in reference soils. Normalizing the uptake rates to soil carbon content (ug gC<sup>-1</sup> d<sup>-1</sup>) instead 3 of soil mass, removes the depth-dependence of uptake rates in reference soils (p <0.05), but 4 | not in mofette soils (figure <u>32</u>).

#### 5 3.5 Quantification of 16s rRNA and marker genes for RubisCO

6 Results of 16S rRNA and RubisCO encoding marker genes are listed in table 3. The 7 abundance of 16S rRNA genes per gram soil is a measure of the total abundance of 8 microorganisms in the soil (Fierer et al., 2005). Gene copy numbers per gram soil of 16S 9 rRNA genes were more abundant in the top 5 cm of the mofette soil. They decrease with 10 depth, in both, mofette and reference soil (p < 0.05), but the decrease is more rapid in the 11 mofette. The same holds true for marker genes encoding for RubisCO. CbbL IC is the most 12 abundant marker gene in both soils, whereas it is more abundant in the reference soil 13 compared to the mofette. CbbL 1C is one order of magnitude more abundant than cbbL 1A 14 and cbbM in both, reference and mofette soils. cbbL:16S rRNA ratios range between  $0.07 \pm$ 15 0.03 and  $0.19 \pm 0.04$  in the mofette soil and stays fairly constant with depth (p = 0.61). In the reference soil the ratio decreases slightly with depth from  $0.37 \pm 0.16$  to  $0.17 \pm 0.04$ , but 16 17 values are consistently greater than in the mofette soil.

18

#### 19 **4 Discussion**

20

#### 4.1 Carbon sources in mofette soils

21 The investigated mofettes are characterized by low pH values, permanently anoxic 22 conditions and TOC accumulation throughout the soil profile, in contrast to nearby reference soils, where C contents accumulate preferentially in the organic rich A horizon and pH values 23 24 are higher. pH values in mofette soils are lower than organic acid buffers. Based on odour, H<sub>2</sub>S oxidation might be responsible for observed low pH values. C/N ratios in both mofette 25 26 soils indicate a change in SOM quality with depth. Low C/N ratios, as found in the top 10 cm 27 of both mofettes, reflect biologically microbially degraded highly processed OM (Rumpel and 28 Kogel-Knabner, 2011) and C/N ratios as low as 9 (top 10 cm of mofette 2) suggest a high 29 contribution of microbial biomass to bulk SOM (Wallander, 2003). A significant contribution 30 of microbial biomass carbon at these depths is also supported by very high 16S rRNA copy

numbers, extracted from mofette 1, which are one order of magnitude higher than known 1 2 from other soils (Fierer et al., 2005). Also numbers of RubisCO encoding genes are two orders of magnitude more abundant than in agricultural soils (Selesi et al., 2007) and twice as 3 4 high as in organic rich paddy rice fields (Wu et al., 2015), suggesting microbial carbon 5 derived from CO<sub>2</sub> assimilation as an important carbon source. Further evidence is given by 6 the isotope data, as mofette SOM at 0 to 10 cm differs from a pure plant signal. The deviation of  $\delta^{13}C_{SOM}$  towards more negative values compared to plant signatures suggests that 7 microbialy derived carbon in shallower depths is fractionated against <sup>13</sup>C, which provides 8 9 further evidence that autotrophic microorganisms contribute significantly to mofette SOM.

Below 20 cm, inceasing C contents in both mofettes are accompined with a steep increase in C/N, which is attributed to lower proportions of microbial carbon and accumulation of undecomposed plant organic matter, as suggested from studies at other mofette sites (Rennert et al., 2011).

## 4.2 Quantification of SOM isotope shifts by combined $\Delta^{14}$ C and $\delta^{13}$ C massbalances

16 TOC, C/N ratios and the abundance of 16S rRNA genes in mofette soils all suggest 17 that microbial carbon might constitute a significant part of bulk SOM. The isotope mass balance model can be used to assess the contribution of plant vs. microbial derived carbon. 18 The approach assumes that microbially derived carbon is distinct either in its  ${}^{14}C$  or its  ${}^{13}C$ 19 isotope ratio compared to plant carbon. The isotope mass balance model derived from 20 equation 9 shows that microbial carbon that is added to SOM has to be depleted in  $\delta^{13}$ C 21 compared to plant inputs, leading to an overall negative  $\delta^{13}$ C shift in bulk SOM of 1-2 ‰ 22 23 compared to a pure plant signal at all depths (figure 23 B).

However, the model assumes that the radiocarbon content of mofette SOM solely depends on the amount of fixed geogenic CO<sub>2</sub> and does not consider radioactive decay. <sup>14</sup>C depletion by radioactive decay, especially with soil depth, can lead to an overestimation of fixed geogenic CO<sub>2</sub> and consequently to an overestimation of the shift in  $\delta^{13}$ C values. In order to account for <sup>14</sup>C depletion by radioactive decay,  $\Delta^{14}$ C values of reference soil SOM can be subtracted from  $\Delta^{14}C_{SOM_mofette}$  in Eq (9).

30 After correcting the model for radioactive decay, the calculated  $\delta^{13}C_{SOM}$  depletion still 31 matches the data for the first 10 cm of both mofettes, where measured  $\delta^{13}C$  values are more

negative than calculated ones (figure 23 C). Below 10 cm. the calculated  $\delta^{13}C_{SOM}$  coincides 1 with measured values in both mofettes, suggesting that SOM  $\delta^{13}$ C preserved the signal of the 2 plant source and only radioactive decay lead to the initial  $\delta^{13}C$  shift in the model (figure 23) 3 C). This supports findings from previous studies, where carbon accumulation accompanied 4 5 with high C/N ratios was attributed to accumulation of poorly decomposed plant material (Rennert et al., 2011). The only exception from this pattern is at 30 - 40 cm in mofette 2, 6 where measured  $\delta^{13}$ C values are still more negative than calculated ones, even after correction 7 8 for radioactive decay (figure 23 C). This might be caused by extremely low carbon dynamics, e.g. due to permanently waterlogged conditions, which would lead to an overestimation of the 9  $\delta^{13}$ C isotope shift in the model. Although water levels fluctuate in the floodplain, permanently 10 11 waterlogged conditions are likely to occur at lower depthsdeeper in moffete 2, where high CO<sub>2</sub> discharge rates might lead to an elevation of the water table. Waterlogged conditions 12 13 lead to low carbon turnover, and correction of radioactive decay with reference soil values might not be sufficient, because reference soils at these depths are only temporally 14 waterlogged. This might explain the mismatch of measured and calculated  $\delta^{13}C$  values at the 15 lowest depth of deepest sampling point in mofette 2 and would indicate a potential bias of 16 modelled C-isotope signatures towards too positive  $\delta^{13}$ C values. 17

18 Another source of error in the model is accumulation of recalcitrant compounds within the SOM pool, like lignin or lipids, which might also lead to a shift in  $\delta^{13}$ C values compared 19 to the original bulk plant material (Benner et al., 1987Alewell et al., 2011; Werth and 20 Kuzyakov, 2010). The accumulation of phenolic compounds is usually accompanied with an 21 increase in C/N ratios (Hornibrook et al., 2000; Werth and Kuzyakov, 2010), which is not the 22 23 case in the top 10 cm of the mofette soil. Therefore, lignin accumuation is not likely to have caused the depletion in the top 10 cm of both mofettes. Nevertheless, increased lignin 24 accumulation might also be the reason for the observed depletion in  $\delta^{13}$ C below 20 cm depth 25 in mofette 2. 26

27 Therefore, the model shows that  $\delta^{13}$ C values in the top 10 cm of both mofettes are 28 significantly depleted compared lower than expected for SOM derived from to a pure plants 29 signal-alone, indicating significant due to addition of  $\delta^{13}$ C depleted carbon. Below, whereas 20 below-10 cm depth, the calculated and measured  $\delta^{13}$ C values coincide agree after correcting 31 for possible sources of error, like radioactive decay and alteration of  $\delta^{13}$ C due to 32 decomposition processes. Microbial carbon that is added to mofette SOM by several CO<sub>2</sub> fixation pathways is likely to be depleted in  $\delta^{13}$ C because of enzymatic fractionation processes (Fuchs, 2011). The deviation in  $\delta^{13}$ C in the top 10 cm of both mofettes also <u>coincides wellis in accord</u> with high CO<sub>2</sub> fixation rates and the abundance of functional marker genes for CO<sub>2</sub> fixation <u>at this</u> <u>depth (figure 4)</u>. This implies that microbial carbon derived from CO<sub>2</sub> assimilating organisms is a major driver of the observed  $\delta^{13}C_{SOM}$  depletion.

#### 7 4.3 Quantification of microbial carbon C derived from CO<sub>2</sub> fixation

8 In order to quantify the proportion of CO<sub>2</sub>-derived microbial carbon from the observed 9 isotope shift, it is important to know the metabolic pathway that was used for CO<sub>2</sub> fixation 10 and its respective corresponding isotope fractionation factor. Beulig et al. (in prepin press, 11 2016-) investigated by metatranscriptomic and metagenomic approaches microbial key 12 processes in mofette soil 1. Consistent with our quantification of cbbL/cbbM marker genes, 13 Beulig et al., (in preppress, 2016.) detected high frequencies of transcripts encoding key 14 enzymes for the Calvin Benson Cycle as well as the Reductive Acetyl CoA Cycle. The Acetyl 15 CoA Cycle is used by acetogens, methanogens and sulphate reducers for catabolism and anabolism (Drake et al. 2006). According to Beulig et al. (in preppress, 2016-), transcripts of 16 17 key enzymes for the Acetyl CoA pathway in the mofette soil are also related to these groups. Most transcripts encoding for the Calvin Benson Cycle were related to chemoautotrophic 18 19 bacteria and algae, living under anaerobic restrictions. The activity of chemoautotrophic 20 bacteria using the Calvin Benson Cylce is also supported by our data, as shown by the good 21 correlation of cbbL/cbbM marker genes and uptake rates (figure 4).

22 Carbon that is fixed by chemoautotrophs or algae using Formtype I RubisCO, the 23 dominant formtype in the mofette, is depleted by -27 to -30 % compared to the source  $CO_2(\Delta)$ 24  $\approx$  -27 to -30 ‰)(Hayes, 2001; Pancost and Damste, 2003). A similar value can be expected 25 for acetate formed from geogenic CO<sub>2</sub> during acetogenesis. In systems where acetate is not limiting, depletion is less pronounced ( $\Delta \approx -32$  ‰) than in acetate-limited systems ( $\Delta \approx -58.6$ 26 27 ‰) (Conrad, 2005; Gelwicks et al., 1989). A value of -32 ‰ is in accordance with acetate  $\delta^{13}$ C values measured by Beulig et al. (2014) in a mofette study from the same area. 28 Therefore, given a  $\delta^{13}$ C value of geogenic CO<sub>2</sub> of around -2 ‰, the C end-member derived 29 from microbial CO<sub>2</sub> fixation adds carbon with an average  $\delta^{13}$ C value of -30 to -34 % to 30 31 bacterial biomass and SOM in mofettes. Taking the differences between measured and

1 calculated  $\delta^{13}$ C (with and without correction for radioactive decay, respectively) for mass 2 balance calculation according to equation 8, microbially fixed geogenic CO<sub>2</sub> carbon in the top 3 10 cm of the mofette soil can make up between 8 ± 2 % and 15 ± 4 % in mofette 1 and 4 between 23 ± 4 % and 27 ± 5 % in mofette 2.

#### 5

#### 4.4 Importance of microbial CO<sub>2</sub> fixation for isotope ratios in peat soils

6 Our data provides evidence that assimilation of CO<sub>2</sub> by several groups of autotrophic microorganisms contributes to SOM formation derived from CO<sub>2</sub>. Recycling of CO<sub>2</sub> in peat 7 deposits has been proposed to cause 'reservoir' effects in radiocarbon, biasing dating of peat 8 (Kilian et al., 1995). As an explanation, Pancost et al. (2000) proposed recycling of  $\Delta^{14}$ C 9 depleted methane that diffuses from the catotelm layer up the peat profile, where it is oxidized 10 11 by methanotrophic organisms and subsequently assimilated by mycorrhizal fungi living in 12 association with Ericaceae rootlets. However, the authors could not find evidence from biomarker analyses of methanotrophic or fungal organisms and attributed recycling of <sup>14</sup>C 13 depleted CO<sub>2</sub> to plants. Our findings suggest that other groups besides fungi are involved in 14 CO<sub>2</sub> recycling, namely CO<sub>2</sub> utilizing autotrophic microorganisms. Pancost et al. (2000) 15 estimated that 20 % of C in the investigated peat is derived from this recycling process. This 16 17 proportion is very similar to our estimates for autotrophic fixation of CO<sub>2</sub> in the 0-10 cm of 18 mofette soil. Hence we would propose that direct fixation of CO<sub>2</sub> could be a major process 19 influencing peat radiocarbon signatures.

20

#### 4.5 Importance of CO<sub>2</sub> fixation in for soil carbon in reference soils

21 When normalized for the mass of carbon (as opposed to mass of soil), rates of CO<sub>2</sub> 22 fixation in the reference soil at depth remain similar to values at the surface (figure 32). While W www. We cannot use the isotope-mixing model to estimate the amount of C derived from CO<sub>2</sub> 23 24 fixation in the reference soil, because the soil atmosphere as well as plants at the reference 25 soilis are not directly influenced by geogenic CO<sub>2</sub>. However, the rate measurementss suggest increasing importance of CO<sub>2</sub> assimilating microorganisms for carbon stocks with depth. Also 26 27 In addition, the high relative abundance of RubisCO marker genes relative to 16S rRNA 28 genes suggest that autotrophic organisms constitute a substantial part of the microbial community throughout the soil profile. Their activity is also indicated by the strong 29 correlation between RubisCO marker genes and uptake rates ( $R^2 = 0.94$ , p < 0.05) (figure 4). 30 Higher CO<sub>2</sub> concentrations, which are usually observed with depth, might also lead to an 31

increase of CO<sub>2</sub> assimilation, because of a higher substrate availability for RubisCO or other
 carboxylases with depth.

In contrast to the mofette soil, which is characterized as an organic rich histosol, the reference soils are classified as gleysols, with high organic carbon contents only in the A horizon. They are characterized by frequently changing redox conditions due to groundwater fluctuations, which might provide sufficient electron donors and acceptors for chemolithoautotrophic microorganisms (Akob and Küsel, 2011).

Beulig et al. (2014) characterized the microbial community of a reference soil at the same study site. The authors found that Proteobacteria constituted a substantial part of the microbial community. Many Proteobacteria are facultative autotrophs using the CBB cycle and have a facultative anaerobe metabolism (Badger and Bek, 2008). They would be therefore able to assimilate  $CO_2$  also under the experimental conditions.

A contribution of phototrophic and chemoautotrophic microorganisms to SOM has been demonstrated already by other studies (Hart et al., 2013; Yuan et al., 2012), but solely for top soils. Wu et al. (2014) and Wu et al. (2015) investigated soil depth profiles up to 15 cm depth , but found no significant incorporation below 5 cm depth in upland and paddy soils under not manipulating experimental conditions, like illumination .

18 Our data suggest that autotrophic microorganisms are active even in the referencesubsoil. Microorganisms using the CBB cycle would add <sup>13</sup>C-depleted carbon to SOM. 19 20 Indeed,  $\delta^{13}$ C profiles of both reference soils do not show shifts towards more positive values 21 with depth, as is usually observed from other Gleysols, although radiocarbon data 22 proofs indicates and increasing ageing of that SOM becomes older with depth (Alewell et al., 2011; Bol et al., 1999).- Further, both reference soils have C/N ratios close to 10 throughout 23 the soil profile, which normally indicates a higher contribution of microbial C to SOM 24 25 (Rumpel and Kogel-Knabner, 2011). This strongly suggests a contribution of autotrophic microorganisms to carbon stocks in the subsoil, though ultimately its influence on the C 26 27 isotopic signature of SOM at depth must be further evaluated.

28

## 29 Conclusions

30  $\delta^{13}$ C and  $\Delta^{14}$ C values of SOM in wetland mofettes were are influenced by 31 incorporation of geogenic CO<sub>2</sub> fixed not only by plants, but also by microbial CO<sub>2</sub>

fixation microbes, as indicated by deviation of  $\delta^{13}$ C values from those expected from if plant C 1 2 inputs alonewere the sole source of SOM-C. The unique isotopic composition of geogenic CO<sub>2</sub> and the different enzymatic fractionation of plants and microorganisms allowsed us to 3 quantify microbially derived C by using combined <sup>14</sup>C and <sup>13</sup>C mass balances, because 4 5 microbial carbon is more depleted than plant C. Other parameters, like C/N ratio, 16S rRNA 6 and cbbL gene abundance also indicate addition of C fixed from geogenic CO<sub>2</sub> by microbes. According to the isotope mass balances, microbial carbon derived from CO<sub>2</sub> fixation accounts 7 for 8 - 27 % of bulk SOM in mofette soils. The significant contribution of autotrophic 8 9 microorganisms to SOM also implies that they might be able to cause reservoir effects in 10 radiocarbon by recycling of old CO<sub>2</sub>, as has been already suggested for peat soils.

Further, high CO<sub>2</sub> fixation rates, especially in mineral horizons of the reference soil, as
 well as the high of RubisCO marker genes indicate a significant contribution of autotrophic
 microorganisms to subsoil carbon.

14

## 15 Acknowledgements

We thank Heike Geilmann and Steffen Rühlow for assistance with  $\delta^{13}$ C analysis of bulk soil 16 and CFE extracts. We thank Heike Machts and Axel Steinhof for radiocarbon analysis of soil 17 18 and plant samples. Further we thank Iris Kuhlmann for assistance in CFE extractions, as well 19 as Julia Kuhr for helping in DNA extraction and soil sampling. We-also kindly acknowledge 20 Gerd Gleixner for helpful discussions and comments on the manuscript. We would also like 21 to acknowledge two anonymous reviewers, whose comments highly improved the original 22 version of the manuscript. This project was supported by the graduate research training group "Alteration and element mobility at the microbe- mineral interface" (GRK 1257), which is 23 24 part of the Jena School for Microbial Communication (JSMC) and funded by the Deutsche 25 Forschungsgemeinschaft (DFG).

## 1 References

- Akob, D. M. and Küsel, K.: Where microorganisms meet rocks in the Earth's Critical Zone,
  Biogeosciences, 8, 3531-3543, 2011.
- Alewell, C., Giesler, R., Klaminder, J., Leifeld, J., and Rollog, M.: Stable carbon isotopes as
  indicators for environmental change in palsa peats, Biogeosciences, 8, 1769-1778, 2011.
- 6 Alfreider, A., Vogt, C., Hoffmann, D., and Babel, W.: Diversity of ribulose-1,5-bisphosphate
- 7 carboxylase/oxygenase large-subunit genes from groundwater and aquifer microorganisms,
- 8 Microb. Ecol., 45, 317-328, 2003.
- 9 Amha, Y., Bohne, H., and Alsanius, B.: Comparison of physiological and biochemical
- 10 methods for assessing microbial activity and biomass of peats. In: Peat: Formation, Uses and
- 11 Biological Effects, 2012.
- 12 Badger, M. R. and Bek, E. J.: Multiple Rubisco forms in proteobacteria: their functional
- significance in relation to CO<sub>2</sub> acquisition by the CBB cycle, J. Exp. Bot., 59, 1525-1541,
  2008.
- Balesdent, J., Mariotti, A., and Guillet, B.: Natural C-13 Abundance as a Tracer for Studies of
  Soil Organic-Matter Dynamics, Soil Biology & Biochemistry, 19, 25-30, 1987.
- 17 Benner, R., Fogel, M. L., Sprague, E. K., and Hodson, R. E.: Depletion of C-13 in Lignin and
- 18 Its Implications for Stable Carbon Isotope Studies, Nature, 329, 708-710, 1987.
- Berg, I. A.: Ecological Aspects of the Distribution of Different Autotrophic CO<sub>2</sub> Fixation
  Pathways, Applied and Environmental Microbiology, 77, 1925-1936, 2011.
- 21 Beulig, F., Heuer, V. B., Akob, D. M., Viehweger, B., Elvert, M., Herrmann, M., Hinrichs,
- 22 K.-U., and Küsel, K.: Carbon flow from volcanic CO<sub>2</sub> into soil microbial communities of a
- 23 wetland mofette, ISME J, doi: 10.1038/ismej.2014.148, 2014. 2014.
- Beulig, F., Urich, T., Nowak, M., Gleixner, G., Trumbore, S. E., Küsel, K.: Soil carbon
  accumulation under high levels of CO<sub>2</sub> is associated with reduced foodweb complexity, in
  prep, Nature Microbiology, in press, 2016.
- 27 Bol, R. A., Harkness, D. D., Huang, Y., and Howard, D. M.: The influence of soil processes
- on carbon isotope distribution and turnover in the British uplands, European Journal of Soil
- 29 Science, 50, 41-51, 1999.

- 1 Bräuer, K., Kämpf, H., Koch, U., and Strauch, G.: Monthly monitoring of gas and isotope
- 2 compositions in the free gas phase at degassing locations close to the Novy Kostel focal zone
- 3 in the western Eger Rift Czech Republic, Chemical Geology, 290, 163-176, 2011.
- 4 Conrad, R.: Quantification of methanogenic pathways using stable carbon isotopic signatures:
  5 a review and a proposal, Organic Geochemistry, 36, 739-752, 2005.
- 6 Conrad, R.: Soil microorganisms as controllers of atmospheric trace gases (H<sub>2</sub>, CO, CH<sub>4</sub>,
  7 OCS, N<sub>2</sub>O, and NO), Microbiol Rev, 60, 609-640, 1996.
- 8 Craig, H.: Isotopic standards for carbon and oxygen and correction factors for mass9 spectrometric analysis of carbon dioxide, Geochimica et Cosmochimica Acta, 12, 133-149,
  10 1957.
- Ehleringer, J. R., Buchmann, N., and Flanagan, L. B.: Carbon isotope ratios in belowground
  carbon cycle processes, Ecol Appl, 10, 412-422, 2000.
- 13 Feisthauer, S., Wick, L. Y., Kastner, M., Kaschabek, S. R., Schlomann, M., and Richnow, H.
- 14 H.: Differences of heterotrophic (CO<sub>2</sub>)-C-13 assimilation by Pseudomonas knackmussii strain
- 15 B13 and Rhodococcus opacus 1CP and potential impact on biomarker stable isotope probing,
- 16 Environmental Microbiology, 10, 1641-1651, 2008.
- Fierer, N., Jackson, J. A., Vilgalys, R., and Jackson, R. B.: Assessment of soil microbial
  community structure by use of taxon-specific quantitative PCR assays, Applied and
  Environmental Microbiology, 71, 4117-4120, 2005.
- Fuchs, G.: Alternative Pathways of Carbon Dioxide Fixation: Insights into the Early
  Evolution of Life? In: Annual Review of Microbiology, Vol 65, Gottesman, S. and Harwood,
  C. S. (Eds.), Annual Review of Microbiology, 2011.
- Gelwicks, J. T., Risatti, J. B., and Hayes, J. M.: Carbon isotope effects associated with autotrophic acetogenesis, Organic Geochemistry, 14, 441-446, 1989.
- 25 Hart, K. M., Kulakova, A. N., Allen, C. C. R., Simpson, A. J., Oppenheimer, S. F., Masoom,
- 26 H., Courtier-Murias, D., Soong, R., Kulakov, L. A., Flanagan, P. V., Murphy, B. T., and
- 27 Kelleher, B. P.: Tracking the Fate of Microbially Sequestered Carbon Dioxide in Soil Organic
- 28 Matter, Environ Sci Technol, 47, 5128-5137, 2013.
- 29 Hayes, J. M.: Fractionation of carbon and hydrogen isotopes in biosynthetic processes, Rev
- 30 Mineral Geochem, 43, 225-277, 2001.

- 1 Herrmann, M., Hädrich, A., and Küsel, K.: Predominance of thaumarchaeal ammonia
- 2 oxidizer abundance and transcriptional activity in an acidic fen, Environmental Microbiology,
- 3 14, 3013-3025, 2012.
- 4 Hornibrook, E. R. C., Longstaffe, F. J., Fyfe, W. S., and Bloom, Y.: Carbon-isotope ratios and
- carbon, nitrogen and sulfur abundances in flora and soil organic matter from a temperate-zone
  bog and marsh, Geochemical Journal, 34, 237-245, 2000.
- 7 Hughen, K., Lehman, S., Southon, J., Overpeck, J., Marchal, O., Herring, C., and Turnbull, J.:
- 8 C-14 activity and global carbon cycle changes over the past 50,000 years, Science, 303, 2029 207, 2004.
- 10 Kämpf, H., Bräuer, K., Schumann, J., Hahne, K., and Strauch, G.: CO<sub>2</sub> discharge in an active,
- 11 non-volcanic continental rift area (Czech Republic): Characterisation ( $\delta^{13}$ C,  ${}^{3}$ He/ ${}^{4}$ He) and
- 12 quantification of diffuse and vent CO<sub>2</sub> emissions, Chemical Geology, 339, 71-83, 2013.
- Kilian, M. R., VanDerPlicht, J., and VanGeel, B.: Dating raised bogs: New aspects of AMS
  C-14 wiggle matching, a reservoir effect and climatic change, Quaternary Science Reviews,
  14, 959-966, 1995.
- Kramer, C. and Gleixner, G.: Soil organic matter in soil depth profiles: Distinct carbon
  preferences of microbial groups during carbon transformation, Soil Biology and
  Biochemistry, 40, 425-433, 2008.
- Küsel, K. and Drake, H. L.: Effects of environmental parameters on the formation and
  turnover of acetate by forest soils, Applied and Environmental Microbiology, 61, 3667-3675,
  1995.
- Liu, F. H. and Conrad, R.: Chemolithotrophic acetogenic H<sub>2</sub>/CO<sub>2</sub> utilization in Italian rice
  field soil, Isme J., 5, 1526-1539, 2011.
- Lueders, T., Manefield, M., and Friedrich, M. W.: Enhanced sensitivity of DNA- and rRNAbased stable isotope probing by fractionation and quantitative analysis of isopycnic centrifugation gradients, Environmental Microbiology, 6, 73-78, 2004.
- Malik, A., Blagodatskaya, E., and Gleixner, G.: Soil microbial carbon turnover decreases with
  increasing molecular size, Soil Biology & Biochemistry, 62, 115-118, 2013.
- Mehlhorn J, Beulig F, Küsel K, Planer-Friedrich B. (2014). Carbon dioxide triggered
  metal(loid) mobilisation in a mofette. *Chemical Geology* 382:54–66.

- 1 Miltner, A., Kopinke, F. D., Kindler, R., Selesi, D. E., Hartmann, A., and Kastner, M.: Non-
- 2 phototrophic CO<sub>2</sub> fixation by soil microorganisms, Plant Soil, 269, 193-203, 2005.
- Mook, W. G. and van der Plicht, J.: Reporting C-14 activities and concentrations,
  Radiocarbon, 41, 227-239, 1999.
- Nanba, K., King, G. M., and Dunfield, K.: Analysis of facultative lithotroph distribution and
  diversity on volcanic deposits by use of the large subunit of ribulose 1,5-bisphosphate
  carboxylase/oxygenase, Applied and Environmental Microbiology, 70, 2245-2253, 2004.
- Pancost, R. D. and Damste, J. S. S.: Carbon isotopic compositions of prokaryotic lipids as
  tracers of carbon cycling in diverse settings, Chemical Geology, 195, 29-58, 2003.
- 10 Pancost, R. D., van Geel, B., Baas, M., and Damste, J. S. S.: delta C-13 values and
- 11 radiocarbon dates of microbial biomarkers as tracers for carbon recycling in peat deposits,
- 12 Geology, 28, 663-666, 2000.
- 13 Rennert, T., Eusterhues, K., Pfanz, H., and Totsche, K. U.: Influence of geogenic CO<sub>2</sub> on
- 14 mineral and organic soil constituents on a mofette site in the NW Czech Republic, European
- 15 Journal of Soil Science, 62, 572-580, 2011.
- 16 Robinson, J. J. and Cavanaugh, C. M.: Expression of Form I and Form II Rubisco in
- 17 Chemoautotrophic Symbioses: Implications for the Interpretation of Stable Carbon Isotope
- 18 Values, Limnology and Oceanography, 40, 1496-1502, 1995.
- Rumpel, C. and Kogel-Knabner, I.: Deep soil organic matter-a key but poorly understood
  component of terrestrial C cycle, Plant Soil, 338, 143-158, 2011.
- Santruckova, H., Bird, M. I., Elhottova, D., Novak, J., Picek, T., Simek, M., and Tykva, R.:
  Heterotrophic fixation of CO2 in soil, Microb. Ecol., 49, 218-225, 2005.
- Scheibe, A., Krantz, L., and Gleixner, G.: Simultaneous determination of the quantity and
  isotopic signature of dissolved organic matter from soil water using high-performance liquid
  chromatography/isotope ratio mass spectrometry, Rapid Communications in Mass
  Spectrometry, 26, 173-180, 2012.
- 27 Selesi, D., Pattis, I., Schmid, M., Kandeler, E., and Hartmann, A.: Quantification of bacterial
- RubisCO genes in soils by cbbL targeted real-time PCR, J Microbiol Meth, 69, 497-503,
  2007.

- 1 Selesi, D., Schmid, M., and Hartmann, A.: Diversity of green-like and red-like ribulose-1,5-
- 2 bisphosphate carboxylase/oxygenase large-subunit genes (cbbL) in differently managed
- 3 agricultural soils, Applied and Environmental Microbiology, 71, 175-184, 2005.
- 4 Stein, S., Selesi, D., Schilling, R., Pattis, I., Schmid, M., and Hartmann, A.: Microbial activity
- 5 and bacterial composition of H-2-treated soils with net CO<sub>2</sub> fixation, Soil Biology &
- 6 Biochemistry, 37, 1938-1945, 2005.
- 7 Steinhof, A., Adamiec, G., Gleixner, G., van Klinken, G. J., and Wagner, T.: The new C-14
- 8 analysis laboratory in Jena, Germany, Radiocarbon, 46, 51-58, 2004.
- 9 Tolli, J. and King, G. M.: Diversity and structure of bacterial chemolithotrophic communities
- 10 in pine forest and agroecosystem soils, Applied and Environmental Microbiology, 71, 8411-
- 11 8418, 2005
- 12 Trumbore, S. E.: Age of soil organic matter and soil respiration: Radiocarbon constraints on
- 13 belowground C dynamics, Ecol Appl, 10, 399-411, 2000.
- Trumbore, S. E.: Carbon respired by terrestrial ecosystems recent progress and challenges,
  Global Change Biology, 12, 141-153, 2006.
- 16 Vance, E. D., Brookes, P. C., and Jenkinson, D. S.: An extraction method for measuring soil
- 17 microbial biomass-C, Soil Biology & Biochemistry, 19, 703-707, 1987.
- 18 Wallander, H., Nilsson, L. O., Hagerberg, D., and Rosengren, U.: Direct estimates of C:N
- 19 ratios of ectomycorrhizal mycelia collected from Norway spruce forest soils, Soil Biology and
- 20 Biochemistry, 35, 997-999, 2003.
- Weisburg, W. G., Barns, S. M., Pelletier, D. A., and Lane, D. J.: 16s Ribosomal DNA
  Amplification for Phylogenetic Study, J Bacteriol, 173, 697-703, 1991.
- Werth, M. and Kuzyakov, Y.: C-13 fractionation at the root-microorganisms-soil interface: A
  review and outlook for partitioning studies, Soil Biology & Biochemistry, 42, 1372-1384,
  2010.
- 26 Whiticar, M. J.: Carbon and hydrogen isotope systematics of bacterial formation and 27 oxidation of methane, Chemical Geology, 161, 291-314, 1999.
- Wood, H. G., Werkman, C. H., Hemingway, A., and Nier, A. O.: Heavy carbon as a tracer in
  heterotrophic carbon dioxide assimilation, J. Biol. Chem., 139, 365-376, 1941.

- WRB, I. W. G.: World Reference Base for Soil Resources 2006, World Soil Resources
   Reports No. 103., FAO, Rome, 2007.
- 3 Wu, X. H., Ge, T. D., Yuan, H. Z., Li, B. Z., Zhu, H. H., Zhou, P., Sui, F. G., O'Donnell, A.
- 4 G., and Wu, J. S.: Changes in bacterial CO2 fixation with depth in agricultural soils, Appl
- 5 Microbiol Biot, 98, 2309-2319, 2014
- 6 Wu, X. H., Ge, T. D., Wang, W., Yuan, H. Z., Wegner, C. E., Zhu, Z. K., Whiteley, A. S., and
- 7 Wu, J. S.: Cropping systems modulate the rate and magnitude of soil microbial autotrophic
- 8 CO<sub>2</sub> fixation in soil, Frontiers in Microbiology, 6, 2015.
- 9 Ye, R. Z., Jin, Q. S., Bohannan, B., Keller, J. K., and Bridgham, S. D.: Homoacetogenesis: A
- potentially underappreciated carbon pathway in peatlands, Soil Biology & Biochemistry, 68,
  385-391, 2014.
- 12 Yuan, H. Z., Ge, T. D., Chen, C. Y., O'Donnell, A. G., and Wu, J. S.: Significant Role for
- 13 Microbial Autotrophy in the Sequestration of Soil Carbon, Applied and Environmental
- 14 Microbiology, 78, 2328-2336, 2012.
- 15

Table 1. Geochemical soil properties of mofette and reference soils.  $\delta^{13}C$  and geochemical data represents background (i.e. without addition of label) data obtained from sampling in September 2014. Radiocarbon data was obtained in November 2013. Uncertainties ofin

geochemical and  $\delta^{13}C$  data represent  $\pm 1\sigma$  standard deviation (n=3). Uncertainties of-in

radiocarbon values represent analytical precision of a homogenized mixed sample. 

	pН	TOC [w-%]	C/N	Water content [%]	$\delta^{13}C$	$\Delta^{14}C$
Mofette 1						
0-5	3.68	$19.64 \pm 1.20$	15.95	53	$-26.90 \pm 0.15$	5542 + 20
5-10	3.59	$26.54 \pm 0.08$	16.52	52	$-27.55 \pm 0.21$	$-554.3 \pm 2.0$
10-20	3.68	$11.53 \pm 0.18$	15.12	57	$-26.71 \pm 0.18$	5507 + 21
20-30	3.43	$16.33 \pm 0.59$	21.65	51	$-26.79 \pm 0.12$	$-559.7 \pm 2.1$
30-40	3.40	$34.00 \pm 1.25$	31.40	56	$-27.01 \pm 0.23$	$-640.2 \pm 1.9$
Reference 1						
0-5	4.13	$25.85 \pm 1.72$	14.37	69	$-27.98 \pm 0.32$	1175 1 2 9
5-10	4.07	$12.40 \pm 0.60$	14.18	49	$-28.10 \pm 0.24$	$-117.5 \pm 2.8$
10-20	4.00	$3.16 \pm 0.26$	14.52	42	$-27.80 \pm 0.13$	22(2 + 27)
20-30	3.91	$3.14 \pm 0.13$	12.93	31	$-27.79 \pm 0.16$	$-236.3 \pm 2.7$
30-40	3.69	$2.81 \pm 0.50$	15.88	30	$-28.23 \pm 0.09$	$-280.2 \pm 2.5$
Mofette 2						
0-5	3.80	$8.66 \pm 0.69$	8.95	52	$-26.01 \pm 0.14$	$-648.1 \pm 1.2$
5-10	3.76	$5.87 \pm 1.11$	8.97	53	$-26.26 \pm 0.24$	$- 648.1 \pm 1.2$
10-20	3.79	$11.41 \pm 0.95$	9.72	50	$-26.76 \pm 0.19$	$(10.7 \pm 1.2)$
20-30	3.52	$28.72 \pm 1.42$	19.74	56	$-27.10 \pm 0.59$	- 618.7 + 1.3
30-40	-	-	-	-	-	-
Reference 2						
0-5	4.50	$12.48 \pm 0.31$	12.16	45	$-27.91 \pm 0.12$	241 + 2.2
5-10	4.51	$7.59 \pm 0.21$	11.52	42	$-28.85 \pm 0.21$	$-34.1 \pm 2.2$
10-20	4.48	$2.94 \pm 0.15$	10.30	46	$-28.11 \pm 0.05$	$114.7 \pm 1.0$
20-30	4.46	$1.91 \pm 0.10$	11.85	40	$-27.82 \pm 0.30$	$-114.7 \pm 1.9$
30-40	4.43	$1.80 \pm 0.04$	10.19	35	$-28.23 \pm 0.06$	$-162.9 \pm 1.9$

1 Table 2: Microbial biomass C and comparison of uptake rates determined during experiment

	$\delta^{13}$ C extract (after fumigation) [‰]	$\delta^{13}$ C control [‰]	Uptake rate/ <u>g</u> soil [ug gdw <sup>-1</sup> d <sup>-1</sup> ]	Uptake rate/g MB-CFE [ug gMB <sup>-1</sup> d <sup>-1</sup> ]	% labelled MB
Mofette 1					
CFE 0 - 10 cm	$233.24 \pm 11.19$	$-25.94 \pm 0.36$	$0.17\pm0.03$	$287 \pm 85$	0.88 + 0.33
Bulk 0 -10 cm	$-21.19 \pm 0.62$	$-26.28 \pm 0.10$	$0.77 \pm 0.23$	-	-
Reference 1					
CFE 0 - 10 cm	$182 \pm 5.44$	$-23.65 \pm 0.54$	$0.59 \pm 0.05$	$139 \pm 32$	$0.40 \pm 0.13$
Bulk 0 -10 cm	$-12.82 \pm 0.95$	$-27.55 \pm 0.14$	$2.65 \pm 0.36$	-	-
Mofette 2					
CFE 0 - 10 cm	$124.51 \pm 10.66$	$-24.10 \pm 0.38$	$0.06 \pm 0.02$	$271 \pm 58$	$0.8 \pm 0.16$
Bulk 0 -10 cm	$-21.37 \pm 0.99$	$-26.49 \pm 0.08$	$0.66 \pm 0.15$	-	-
Reference 2					
CFE 0 - 10 cm	$158.05 \pm 4.01$	$-26.46 \pm 0.21$	$0.25\pm0.09$	$99 \pm 36$	$0.20 \pm 0.10$
Bulk 0 -10 cm	$-17.44 \pm 0.81$	$-27.21 \pm 0.22$	$0.71 \pm 0.16$	-	-

2 1 with CFE and bulk measurements. Uncertainties represent  $\pm 1\sigma$  standard deviation (n=3).

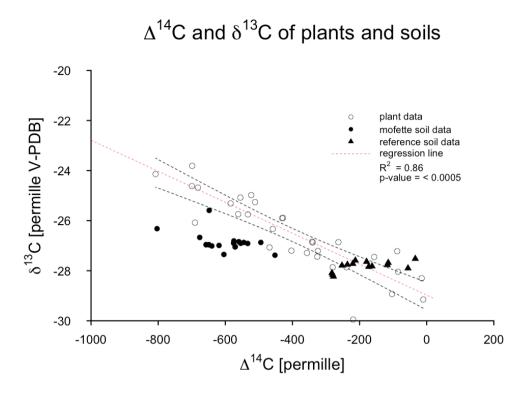
3

1
I

Table 3: Quantification of 168s RNA, cbbL and cbbM marker genes. Uncertainties represent

 $\pm 1\sigma$  standard deviation (n=3).

		× ,				
	Depth [cm]	16S rRNA	cbbM	cbbL 1A	cbbL 1C	cbbL 1C/ 16sRNA
Mofette 1	0 - 5	7.50E+10 ± 1.42E+07	5.70E+08 ± 3.21E+08	9.45E+08 ± 4.86E+08	9.23E+09 ± 4.55E+09	$0.12 \pm 0.06$
	5 - 10	1.65E+10 ± 5.35E+06	2.21E+08 ± 1.28E+08	1.40E+08 ± 1.69E+08	1.46E+09 ± 1.20E+09	0.11 ± 0.04
	10 - 20	3.35E+09 ± 0.51E+06	1.49E+07 ± 8.45E+06	1.83E+07 ± 1.22E+07	6.02E+08 ± 1.25E+08	$0.17 \pm 0.03$
	20 - 30	5.94E+09 ± 9.02E+05	1.62E+07 ± 1.23E+07	1.12E+07 ± 4.07E+06	3.98E+08 ± 1.53E+08	$0.07 \pm 0.03$
	30 - 40	7.62E+08 ± 9.39E+04	8.53E+05 ± 3.02E+05	1.71E+06 ± 5.23E+05	7.91E+07 ± 2.18E+07	$0.10 \pm 0.03$
Reference 1	0 - 5	4.63E+10 ± 3.01E+07	3.43E+08 ± 3.18E+08	1.14E+09 ± 4.74E+08	1.58E+10 ± 7.20E+09	$0.37 \pm 0.23$
	5 - 10	2.98E+10 ± 2.02E+07	2.01E+08 ± 5.98E+07	2.69E+08 ± 1.52E+08	7.78E+09 ± 8.12E+08	$0.28 \pm 0.08$
	10 - 20	2.81E+10 ± 4.83E+07	1.31E+08 ± 4.73E+07	3.06E+08 ± 1.59E+08	5.95E+09 ± 1.50E+09	$0.21 \pm 0.06$
	20 - 30	1.24E+10 ± 4.37E+07	9.75E+07 ± 3.99E+07	9.11E+07 ± 3.90E+07	2.25E+09 ± 6.84E+08	0.18 ± 0.03
	30 - 40	4.65E+09 ± 9.61E+07	1.57E+08 ± 9.26E+07	3.47E+07 ± 2.20E+07	5.95E+08 ± 1.78E+08	0.10 ± 0.06



2

Figure 1. Correlation between  $\delta^{13}C$  and  $\Delta^{14}C$  of plants growing around the mofette structure. 3 Dependent on the exposure to geogenic CO<sub>2</sub>, plants incorporate different amounts of geogenic 4 CO<sub>2</sub>, which complicates isotope mass balance calculations for mofette SOM. However, both 5 6 isotopes are highly correlated in sampled plant vegetation material, which allows to predictprediction of  $\delta^{13}$ C comparing SOM isotope values with the aid of with from plant  $\Delta^{14}$ C 7 values. Most data points measured from mofette SOM fall outside 95% confidence levels of 8 the regression, which suggests a deviation of mofette SOM  $\delta^{13}$ C values from a pure 9 vegetation signal. Reference SOM  $\delta^{13}$ C values fall mainly within the observed plant  $\delta^{13}$ C 10 valuessignal, althoughand they do not show a increase with depth, as is usually often observed 11 in soil depth profiles. Parameters of the regression can be used for calculating are used to 12 predict the  $\delta^{13}C_{SOM}$  values expected in mofette soils that correspond to measured radiocarbon 13 14 values, assuming that all carbon would be plant derived (Eq. 9). 15

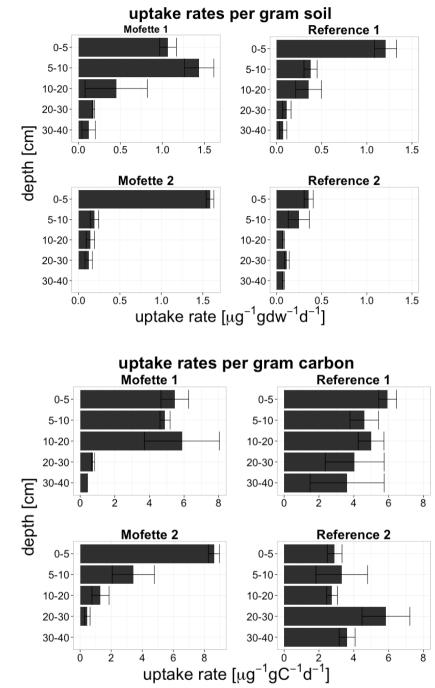
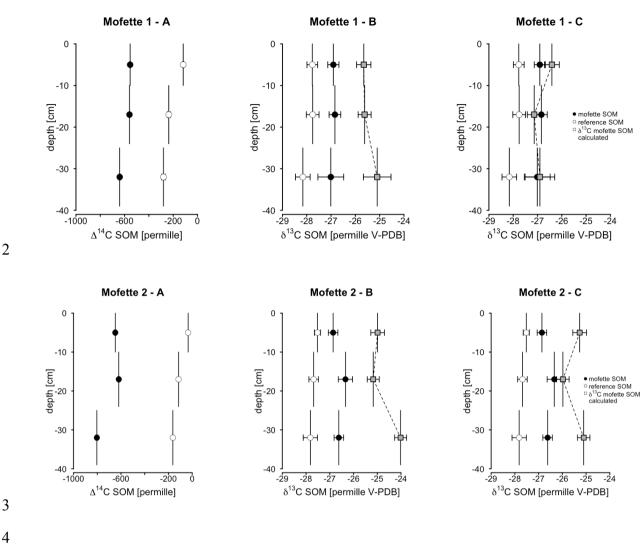


Figure <u>32</u>: <u>CO<sub>2</sub></u> <u>uU</u>ptake rates along depth profiles of mofette and reference soils as determined by bulk measurements from experiment 2. In both mofettes, uptake rates are highest in the top 10 cm and show a trend towards decreasing values at lower depths, especially below 20 cm. Uptake rates in reference soils <u>are</u>-also decreaseing with depth, but stay fairly the same are nearly constant if normalized to organic carbon content. In contrast, <del>5</del> which is not true for mofette soils (uptake rates per organic carbon decline with depth in the

- 1 <u>mofette soils</u>). This suggests an-increasing importance of autotrophic organisms with soil
- 2 depth in <u>the</u> reference soil.



4

5

Figure 23: Depth profile of <sup>14</sup>C and <sup>13</sup>C signatures of SOM in mofette and reference soils.

6 A) Radiocarbon values in mofette soils are more depleted than reference soils, reflecting 7 incorporation of geogenic CO<sub>2</sub> either by plants or by microorganisms. Error bars reflect 8 analytical precision because only one homogenized sample was runanalyzed.

B)  $\delta^{13}C$  values in both mofettes are also shifted towards geogenic CO<sub>2</sub>, but to a smaller 9 extented than radiocarbon values. Gray squares in  $\delta^{13}C$  depth profiles show calculated values 10 of  $\delta^{13}$ C values of in mofette SOM estimated using according to Eq (9). Measured  $\delta^{13}$ C values 11 are more depleted than calculated estimated values at all depths. 12

C) Calculated Estimated  $\delta^{13}$ C values, assuming eq (9) but with 14C values that have been -are 13 14 additionally corrected for radioactive decay assuming that SOM ages with depth in the same

way as the reference soil. Radiocarbon values of mofette SOM depends not only on 1 incorporated geogenic CO<sub>2</sub> but also on depletion of <sup>14</sup>C by radioactive decay. These estimated 2  $\delta^{13}$ C values , which were corrected for radioactive decay, correspond wellagree with 3 measured values below 20 cm depths but remain still-depleted compared to a-what is expected 4 from a pure plant signal-SOM source in the top 10 cm. This suggests that the observed 5 depletion in the top 10 cm of both mofette soils is caused by addition of <sup>13</sup>C depleted 6 microbial carbon, derived from fixed CO<sub>2</sub>. In contrast, the mismatch between calculated 7 8 estimated and measured values below 20 cm depth in (B) can be explained by radioactive 9 decay.

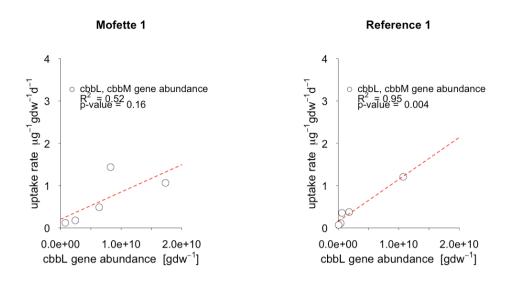




Figure 4: Correlation of marker genes encoding for RubisCO and measured uptake rates in mofette soil 1 and reference soil 1 in the soil depth profile from 0 to 40 cm depth. The good correlation in the reference soil indicates high contribution of chemolithoautotrophic microorganisms to measured uptake rates. In the mofette soil  $R^2$  is considerable lower, most probably, because also other CO<sub>2</sub> fixation cycles that the CBB cycle, like the Acetyl-CoA cycle, are important pathways in these soils.

