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measurements

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# Calculations of automatic chamber flux measurements of methane and carbon dioxide using short time series of concentrations

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## Abstract

The closed chamber technique is widely used to measure the exchange of methane ( $\text{CH}_4$ ) and carbon dioxide ( $\text{CO}_2$ ) from terrestrial ecosystems. There is, however, large uncertainty about which model should be used to calculate the gas flux from the measured gas concentrations. Due to experimental uncertainties the robust linear regression model (first order polynomial) is often applied, even though theoretical considerations of the technique suggest the application of other, curvilinear models. High-resolution automatic chamber systems which sample gas concentrations several hundred times per flux measurement make it possible to resolve the curvilinear behavior and study the information imposed by the natural variability of the temporal concentration changes.

We used more than 50 000 such flux measurements of  $\text{CH}_4$  and  $\text{CO}_2$  from five field sites located in peat forming wetlands to calculate fluxes with different models. The flux differences from independent linear estimates are generally found to be smaller than the local flux variability on the plot scale. The curvilinear behavior of the gas concentrations within the chamber is strongly influenced by wind driven chamber leakage, and less so by changing gas concentration gradients in the soil during chamber closure.

Such physical processes affect both gas species equally, which makes it possible to isolate biochemical processes affecting the gases differently, such as photosynthesis limitation by chamber headspace  $\text{CO}_2$  concentrations under high levels of incoming solar radiation. We assess the possibility to exploit this effect for a partitioning of the net  $\text{CO}_2$  flux into photosynthesis and ecosystem respiration and argue that high-resolution automatic chamber measurements could be used for purposes beyond the estimation of the net gas flux.

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# 1 Introduction

To understand the role of wetlands within the global carbon cycle, accurate estimations of the fluxes of methane ( $\text{CH}_4$ ) and carbon dioxide ( $\text{CO}_2$ ) between the surface and the atmosphere are essential (McGuire et al., 2012). Gas exchange measurements are often made with the closed, non-steady state chamber technique whereby a chamber is placed on top of the soil for a short interval and the change in gas concentrations in the chamber headspace is monitored over time. The resulting time series of gas concentration measurements makes it possible to calculate an atmosphere–surface exchange with the plot on which the chamber was installed. This is often done using first order polynomial linear regression, even though the change in gas concentration might be curvilinear. A number of factors can influence the temporal changes in the gas concentration in a systematic manner that can lead to the development of the curvilinear change in the concentration. For example, the increase of temperature and humidity inside the closed chamber can affect biological processes (e.g. increase respiration, decrease photosynthesis) as well as the gas concentration measurements, which can lead to an apparent saturation of the increase. The same is true for the extraction of gas samples for analysis, and leaks in the chamber construction or installation by which enclosed air can mix with ambient air.

Also, the temporal increase might appear to saturate because the vertical concentration gradient between the soil and the chamber headspace lessens as a result of accumulation in the chamber. This effect was theoretically described using diffusion theory by Hutchinson and Mosier (1981). The more recent non-steady-state diffusive flux estimator (NDFE) model is built around the same argument of an altered gas concentration gradient in the soil and has proven to be perform well in computer simulations (Healy et al., 1996). The NDFE model captures the diffusive pathways of gas transport in the soil and has thus been applied in different experiments including flux measurements of  $\text{CO}_2$  (e.g., Kutzbach et al., 2007) and  $\text{CH}_4$  (e.g., Forbrich et al., 2010). The additional curvature parameter of such diffusion-based models is of particular interest, because

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it holds information about the processes of gas transport in the soil, which could be used to additionally characterize site conditions to e.g. assess the effect of vascular plant abundance on gas transport (Ström et al., 2005). Such flux models, however, disregard ebullitive gas transport, which has to be analyzed using different methods (e.g., Goodrich et al., 2011). Moreover, it is an open question whether the effect of an altered concentration gradient is important under field conditions, and it is hard to uncouple this effect from other episodic sources of changes.

The choice of flux model can be one of the largest sources of uncertainty for chamber flux measurements (Levy et al., 2011). In this process, log-linear or higher order polynomial models often yield significantly elevated fluxes but the additional parameter of the fit (curvature) makes them vulnerable to noise in the measurements. It has therefore been proposed to analyze the quality of fit of several models for every flux measurement, and use the result of the model which gives the best description of the gas concentration change (Forbrich et al., 2010; Pedersen et al., 2010; Kutzbach et al., 2007). The present study, on the other hand, analyses the resulting flux time series of different models separately, and compares them to independent flux estimates reported by the sites. We attempt to explain the apparent differences with environmental conditions, and thus investigate the processes affecting the evolution of the headspace gas concentrations. The simultaneous analysis of CH<sub>4</sub> and CO<sub>2</sub> curvatures could make it possible to isolate biological and physical processes, and thereby exploit the information for the purpose of CO<sub>2</sub> flux partitioning into photosynthesis and ecosystem respiration.

An analysis of concentration changes can only be meaningful if random experimental uncertainties are kept to a minimum. We achieve this by using data from high-resolution automatic chamber systems installed to monitor CH<sub>4</sub> and CO<sub>2</sub> fluxes at five natural wetland sites, ranging from the high Arctic down to the mid-latitudes. These sites feature comparable, but slightly different measurement configurations, and all have sufficient resolution in time and concentration to resolve the curvature within the concentration changes. Beside the ecological differences between sites, they also employ slightly

different methods to calculate the fluxes they report, which we use to assess the differences of the flux estimation methods.

## 2 Materials and methods

### 2.1 Study sites

5 The five study sites are all situated in peat forming wetlands where the water table is typically close to the soil surface. Table 1 shows an overview of their locations, long-term temperature and precipitation, the ecosystem type, as well as the year in which the data used in the present study was recorded. These sites span about 22 latitudinal degrees in the north atlantic region and hence cover a wide range of climatic  
10 conditions. The ground thermal regime at the sites ranges from continuous permafrost at Adventdalen (with ice-wedge polygons) and Zackenberg, to sporadic and isolated permafrost at Stordalen and Kobbefjord, to no permafrost at Fäjemyr. Apart from Fäjemyr, which is a mid-latitude bog, all sites are located in the arctic or subarctic tundra. The vegetation at all sites is dominated by typical wetland species such as *Eriophorum* spp. and *Dupontia* spp. with a varying subcanopy of mosses (*Sphagnum* spp.).  
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### 2.2 Experimental setup

All field sites are equipped with a similar automatic chamber system based on Goulden and Crill (1997). Adventdalen, Zackenberg, Kobbefjord and Fäjemyr all feature the same setup: A set of six transparent chambers (each covering a square of 60 cm by  
20 60 cm, with a height of 30 cm) are placed at representative locations at each site. Inside each chamber there is a fan for ventilation and gas mixing. A pair of high-density polyethylene tubes (4 mm inner diameter) connect each chamber to the gas analyzers, which consists of a nondestructive CO<sub>2</sub> analyzer (SBA-4, PP Systems, UK) and a like-wise nondestructive CH<sub>4</sub> analyzer (DLT100, Los Gatos Research, USA). Sample air is pumped from the chamber, through the gas analyzer and back to the chamber at a rate  
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of  $0.4 \text{ L min}^{-1}$ . Primary  $\text{CH}_4$  concentrations are recorded at  $1.0 \text{ Hz}$ , and primary  $\text{CO}_2$  concentrations are recorded at a slightly lower rate of  $0.625 \text{ Hz}$ . The computer running these automatic measurements activates the chambers in succession for 10 min. During the first 3 min the chamber is open for ventilation, then closed for 5 min, and then opened again for the last 2 min. Thus each chamber is activated once per hour while the five inactive chambers remain open.

At Stordalen there are nine transparent chambers that are activated for 18 min at a time. This results in a three-hour cycle (one 18 min slot is used as a control with ambient air). The chamber closure time is 5 min, between minute ten and 15 of each measurement. The construction of the chambers is different from the other sites. The entire chamber is lifted off plots with short canopies ( $< 20 \text{ cm}$ ) and a similar 20 cm portion is lifted off collars installed in habitats with taller vegetation. Another important difference to the other sites is that Stordalen does not use fans inside the chambers, which could lead to more variability in the measured concentrations. Mixing within the chamber is due to flow ( $2 \text{ L min}^{-1}$ ) between the sample return manifold and the sample outlet port. A small subflow is diverted to a cavity ring-down laser spectrometer (DLT-100, 908-011, Los Gatos Research, USA) used for concentration analysis at a rate of  $1.0 \text{ Hz}$  for both  $\text{CH}_4$  and  $\text{CO}_2$ .

Examples of the recorded data are shown in Fig. 1, for both  $\text{CH}_4$  and  $\text{CO}_2$  (see Supplement for more examples from other sites). An initial equilibration phase is apparent during the first few minutes after which the baseline stabilizes. Due to the distance between chambers and the gas analyzer there is a time delay between chamber closure and the start of the flux measurement. To allow for robust and automated processing we decided to use a fixed 3 min window when fitting models to the data. This window starts 2 min after closure (to account for the time delay) and ends at chamber opening, which ensures that all included concentration measurements were taken while the chamber was closed. This approach will always exclude parts of the flux measurement, but it still leaves 180 concentration measurements for  $\text{CH}_4$  and at least 112 for  $\text{CO}_2$ .

The air temperature ( $T$ ) and pressure ( $P$ ) used in the flux calculations were recorded by sensors in the vicinity of the chambers. For the sake of comparability, we only use flux measurements recorded in June, July and August of the respective year of each site.

## 5 2.3 Flux models

The linear model assumes a constant concentration change, i.e.

$$\frac{dc(t)}{dt} = \left(\frac{A}{V}\right) f_0, \quad (1)$$

10 where  $c(t)$  is the gas concentration in time,  $f_0$  is the (initial, pre-deployment) gas flux which is assumed to be constant during closure time,  $A$  is the area which is covered by the chamber, and  $V$  is the (effective, free) volume of the chamber. Note that gas concentrations are typically measured as a molar fraction (e.g. in units of ppm) and have to be converted to volumetric mass density (e.g.  $\text{mgm}^{-3}$ ) by means of the ideal gas (using  $T$  and  $P$ ) law before Eq. (1) can be applied. Solving this differential equation leads to the linear model

$$15 \quad c(t) = \left(\frac{A}{V}\right) f_0 \cdot t + c_0, \quad (2)$$

where the integration constant  $c_0$  represents the ambient atmospheric (pre-deployment) concentration of the respective gas.

We extend the linear model of Eq. (1) with a term counteracting any change of gas concentration from the ambient concentration in a linear fashion, i.e.

$$20 \quad \frac{dc(t)}{dt} = \left(\frac{A}{V}\right) f_0 - \lambda \cdot (c(t) - c_0), \quad (3)$$

where the constant  $\lambda$  (in units of  $\text{time}^{-1}$ ) describes the sum of all processes which are proportional to the concentration difference  $\Delta c(t) = c(t) - c_0$ . If no curvature is present,

i.e.  $\lambda = 0$ , this model reduces to the linear model. Equation (3) is solved by the function

$$c(t) = \left(\frac{A}{V}\right) \frac{f_0}{\lambda} \cdot (1 - e^{-\lambda t}) + c_0, \quad (4)$$

which defines the score function of this, hereafter referred to as, exponential model. It is based on the assumption that curvature is proportional to  $\Delta c(t)$ , but it does not a priori assume any process to be responsible for the curvature. Other authors have taken the opposite approach by identifying the relevant processes first, and through the assumption of their proportionality to  $\Delta c(t)$  derived the exponential form of the  $c(t)$  score function (e.g., Pedersen et al., 2010; Kutzbach et al., 2007). For example, the curvature of the  $\text{CO}_2$  flux measurement ( $\lambda_{\text{CO}_2}$ ) can be decomposed into three independent constants describing leakage, diffusivity in the soil profile and the saturation of photosynthesis under high sunlight conditions where photosynthesis is assumed to be limited by  $\text{CO}_2$  concentrations inside the chamber (Kutzbach et al., 2007). For this last effect, it has been shown that the relationship between the high sunlight photosynthetic flux,  $F_p$ , and the surrounding  $\text{CO}_2$  concentration is approximately linear in the relevant range of  $\text{CO}_2$  concentrations (Farquhar et al., 1980), i.e.  $F_p(t) = k_p \cdot c(t) \cdot \left(\frac{V}{A}\right)$ , where  $k_p$  is the constant of proportionality. As the  $\text{CO}_2$  concentration in the chamber headspace decreases during the closure time,  $F_p$  decreases correspondingly. This interaction is captured by the exponential model and would result in an increased  $\text{CO}_2$  curvature at high levels of sunlight, or photosynthetically active radiation (PAR). This also means that if  $k_p$  can be isolated from  $\lambda$ ,  $F_p$  can be estimated from the curvature of the measurement, and thereby achieve a  $\text{CO}_2$  flux partitioning.

The non-steady state diffusive flux estimator (NDFE) model (e.g., Healy et al., 1996) is implemented as

$$c(t) = \left(\frac{A}{V}\right) f_0 \tau \left[ \frac{2}{\sqrt{\pi}} \sqrt{t/\tau} + e^{t/\tau} \text{erfc} \left( \sqrt{t/\tau} \right) - 1 \right] + c_0, \quad (5)$$

where the curvature parameter  $\tau$  (in units of time) measures how fast the changed gas concentration gradient propagates through the soil. Like other authors (e.g., Kutzbach

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et al., 2007) we restrict the application of the NDFE model to exclusively positive fluxes (gas sources), i.e. our CH<sub>4</sub> measurements.

These models are optimized against the measured concentrations with a least-squares algorithm based on the Levenberg–Marquardt algorithm. The values of all other variables entering the flux calculation ( $A$ ,  $V$ ,  $T$ ,  $P$ ) are the same for all models.

## 2.4 Linear reference fluxes

We compare the curvilinear flux estimates derived from the fixed 3 min window of the flux measurement to independent flux estimates. Different versions of the linear regression method (cf. Eq. 2) were used to calculate these estimates at each site, which are hereafter referred to as reference fluxes.

At Zackenberg, a linear regression to the initial, most-linear, part of the gas concentration curve was applied by careful visual inspection of each measurement (Masstepanov et al., 2013). The same approach was used for the Kobbefjord (Jensen and Rasch, 2013) and Fäjemyr (Lund, 2009) reference fluxes.

At Stordalen, the algorithm first block-averages the raw data to 15 s resolution and then calculates eight sequential 2.25 min long fits starting every 15 s (Bäckstrand et al., 2008). The most linear (highest  $R^2$ ) of these eight fits is used for CH<sub>4</sub> flux calculation, and the steepest one for CO<sub>2</sub> uptake situation (usually during the day). This procedure is designed to avoid saturation effects.

For Adventdalen (the most recent site) we did not have independently calculated reference fluxes. Instead, we applied linear regression to the same three minute time window which was used for the curvilinear models. Consequently, Adventdalen yields the direct comparison between linear and curvilinear flux estimates, without additional effects of the fit window choice or block averaging.

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## 3 Results and discussion

### 3.1 Flux estimates

Figure 2 shows a typical example of the CH<sub>4</sub> flux estimates. Both curvilinear models give reasonable results with a comparable magnitude to the reference data. There are, however, clear spikes in the NDFE flux estimate which lead to a significantly higher temporal variability compared to both reference and exponential flux estimates. True natural CH<sub>4</sub> emissions are not expected to fluctuate so strongly under these conditions in summer time. The spikes do not relate to ebullition events but instead coincide with measurements with strong curvature (low  $\tau$ ), exemplified by the two examples from chamber 6 at Zackenberg (shown in Fig. 1) which are marked by the arrows in Fig. 2. This unrealistic CH<sub>4</sub> flux pattern of the NDFE model suggests a violation of the underlying assumption of the model, i.e. that curvature cannot generally be attributed to the altered gas concentration gradient in the soil profile.

Unlike in the NDFE model, curvature ( $\lambda$ ) and flux are uncoupled in our exponential model, demonstrated by the stable flux results, which are independent of curvature strength. In the example shown in Fig. 2 the exponential model yields on average about 7% flux increase compared to the reference data, while the NDFE model gives about 24% higher fluxes than the reference – more-or-less independent of the absolute flux magnitude.

An alternative way to quantify the differences between two flux models (for example reference and exponential) is to assume a constant ratio, i.e.  $f_0^{\text{ref}}(f_0^{\text{exp}}) = a \cdot f_0^{\text{exp}}$ , and estimate the ratio  $a$  by a least-squares fit. To avoid a strong influence of a few outliers on the fit we filtered out the highest and lowest 3% of the fluxes before fitting. Figure 3 shows the result for reference and exponential flux estimates for all chambers at Zackenberg combined (see Supplement for more examples from other sites). A high correspondence ( $R^2 > 0.9$ ) and an overall agreement of the flux magnitudes of about 3% for CH<sub>4</sub> and 9% for CO<sub>2</sub> is shown. Table 2 shows these summary (all chamber) statistics for all sites. It shows the effect of the different flux estimation procedures, as

well as site-specific differences. For example, the difference between reference and exponential estimates of CH<sub>4</sub> fluxes at Zackenberg, Kobbefjord and Fäjemyr is lower than at Adventdalen where the reference (linear regression) is applied for the full 3 min window (and not manually to the initial slope). For CH<sub>4</sub>, where the NDFE model can be applied, this model yields a significantly higher flux (and lower  $R^2$ ), which is probably caused by the above described problems of this model. At Fäjemyr, where CH<sub>4</sub> flux magnitudes are low compared to the other sites (hence lower signal-noise ratio), the  $R^2$  between reference and NDFE flux is particularly low. Nonetheless, the ratios between the different flux estimates are still below the typical spatial variability between the individual chambers of each respective site. So our findings suggest that the large uncertainty connected to the choice of the flux model is still exceeded by natural spatial variability on the plot scale.

### 3.2 Curvature parameter $\lambda$

We analyzed the dependency of the curvature parameter of the exponential model  $\lambda$  (cf. Eq. 3) to environmental conditions, such as air temperature, pressure, solar radiation and wind speed. As some of these variables may correlate amongst each other it can be difficult to identify the processes responsible for the observed curvature. However, throughout all sites, the ambient wind speed is found to have the strongest correlation to  $\lambda$ , as shown for CH<sub>4</sub> in Fig. 4a. We illustrate this with data recorded by chamber 3 in Adventdalen 2013, because it contains measurements taken with two different kinds of tape to seal the chamber on the edge of the automatically closing lid. But all other chambers show the same characteristic picture where curvature is influenced by chamber leakage driven by ambient wind speed. This experimental inevitability shows that curvature can be strongly related to other effects than the altered gas concentration gradient in the soil profile. To test whether this effect can nevertheless be seen in our dataset, we use CH<sub>4</sub> curvatures from Fäjemyr where water table height is also measured. Figure 4b shows that the curvature tends to increase when the water table

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drops, which could be explained by a change in the gas concentration gradient, which is supposedly faster in drier soil because of the increased effective diffusivity.

### 3.3 Carbon dioxide flux partitioning from curvature differences

$\lambda_{\text{CH}_4}$  and  $\lambda_{\text{CO}_2}$  are largely affected by the same processes, as shown by their strong correlation in Fig. 5a with data of chamber 3 at Zackenberg (see Supplement for more examples from other sites). This can be explained by physical processes, such as wind driven leakage, which affect both gases equally. The difference  $\lambda_{\text{CO}_2} - \lambda_{\text{CH}_4}$ , on the other hand, should be sensitive to processes that affect the two gases differently. Analyzing the relationship of this curvature difference to environmental parameters, we noticed that it tends to increase above a certain level of incoming sunlight as shown in Fig. 5b. We hypothesize that this relationship is made up of a baseline, which is related to processes independent of incoming sunlight (such as the different diffusivity and gas concentration gradients), and a signal which sets in at higher levels of sunlight, when photosynthesis is supposedly limited by  $\text{CO}_2$  concentration in the chamber headspace rather than incoming sunlight. For the ecosystem of chamber 3 at Zackenberg, this increase in curvature difference starts at PAR of about  $500 \mu\text{mol m}^{-2} \text{s}^{-1}$  and levels off at about  $950 \mu\text{mol m}^{-2} \text{s}^{-1}$ . An indication of this effect could already be seen in the example of Fig. 1a, where PAR was  $917 \mu\text{mol m}^{-2} \text{s}^{-1}$  and  $\lambda_{\text{CO}_2} > \lambda_{\text{CH}_4}$ .

By subtracting the low PAR baseline from the curvature difference we can isolate the PAR-dependent signal in the curvature. Under conditions where photosynthesis is limited by  $\text{CO}_2$  concentrations, this can give an estimate of  $k_p$ , i.e. the rate at which the  $\text{CO}_2$  flux decreases as a response to the decreasing  $\text{CO}_2$  concentrations in the chamber headspace. This means that at pre-deployment conditions  $F_p(t=0) = k_p \cdot c_0 \cdot (\frac{V}{A})$ , given that all environmental variables are constant during closure time. Figure 6 shows the resulting  $F_p$  estimates, as well as ecosystem respiration,  $R_{\text{eco}}$ , calculated from the difference to the total  $\text{CO}_2$  flux (NEE). Due to unstable environmental conditions during the closure time some partitioned fluxes have too large standard errors to confine the partitioning (corresponding to  $\text{error}(R_{\text{eco}}) > 200 \text{ mg CO}_2 \text{ m}^{-2} \text{ h}^{-1}$ ), which

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were here filtered out. As no night-time fluxes are available during the summer at high Arctic sites, we compare these results to a commonly used day-time partitioning method (Lasslop et al., 2010), which models NEE as the sum of a rectangular hyperbolic light–response function (PAR-dependent) and the Lloyd–Taylor respiration model (temperature-dependent). Both estimates of  $F_p$  give a comparable flux, even though the uncertainty of the curvature derived estimates are high and only a few measurements are available (stable conditions and high PAR).

Another way of verifying the partitioned fluxes derived from the curvatures is to compare  $R_{\text{eco}}$  to dark measurements which were conducted during the field campaign at Zackenberg by putting a light-prove blanket over the chambers for one measurement per week. The resulting fluxes (labelled *Dark* in Fig. 6) tend to be lower than both model and curvature estimates, which could be explained by the elimination of photorespiration in dark measurements (which is included in the other two methods). On the other hand, it may also indicate the uncertainties that are connected to the different  $\text{CO}_2$  flux partitioning methods.

Note that the  $\text{CO}_2$  flux partitioning from curvature differences requires an accurate estimation of the curvature of both  $\text{CH}_4$  and  $\text{CO}_2$ . Even with high-quality measurements, this can be hindered by naturally low fluxes or unstable environmental conditions. Moreover, one needs enough measurements at all levels of sunlight to see the relationship between the curvature difference and PAR, and estimate the low PAR baseline. Our data of the other sites show the same characteristic picture described here, even though these limitation can impose significant uncertainty on the results and thereby limit the applicability of this partitioning method. Still, our data shows that it is in principle possible to partition NEE into  $F_p$  and  $R_{\text{eco}}$ , if enough accurate estimations of  $\text{CH}_4$  and  $\text{CO}_2$  curvatures can be obtained.

## 4 Conclusions

We analyzed short time series of concentrations of automatic chamber  $\text{CH}_4$  and  $\text{CO}_2$  flux measurements from natural wetlands using different flux estimation models. Throughout all five sites included in the study, the derived curvature parameters indicate that wind driven leakage has a strong effect on the concentration change within the chamber, which affects the various flux models differently. The linear regression model underestimates fluxes when leakage is strong, whereas the exponential model is better suited and yields fluxes very similar to those based on the initial slope. In other studies that report such fluxes, the use of linear regression is often motivated by short closure times and careful analysis. Indeed, the good accordance with the results of the exponential model justifies the careful application of linear regression on the basis of the large spatial variability present in nature.

The NDFE model, however, exemplifies that flux estimates can be overestimated and noisy when the assumptions of a process-based model are violated. The NDFE model should only be applied with outmost care, i.e. only if the analyst is sure that the altered gas concentration gradient is indeed the main reason for curvilinear concentration changes, such as it might be in controlled laboratory experiments or computer simulations. Direct measurements of the gas concentration at different depths in the soil under a chamber could in future studies quantify to what extent the concentration gradient is really altered by the presence of the chamber.

It is moreover important that the used flux estimator is suitable for the resolution at which the primary gas concentrations are measured. The measurement precision in the present study was high enough for both time and concentration to perform an analysis of curvilinear behavior, and relevant information contained therein could be extracted. We have shown that the simultaneous measurement of  $\text{CH}_4$  and  $\text{CO}_2$  curvatures (as well as PAR) can be used to isolate leakage and estimate photosynthesis through its limitation by  $\text{CO}_2$  concentrations in the chamber headspace. Under stable, high PAR conditions this allows for  $\text{CO}_2$  flux partitioning, which is particularly relevant for high

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Arctic sites where night-time data is not available in summer time. Old datasets can be used to further compare the partitioned CO<sub>2</sub> fluxes of models to those derived from the measured curvatures. The potential of the curvature partitioning, as well as the large uncertainties still connected to it, provide an incentive for improvement in future measurement campaigns and analyses. The present study shows that the application of curvilinear models to high-resolution closed chamber measurements has the potential to provide additional insights to the different processes which give rise to the net gas flux in the chamber and govern ecosystem exchange at large.

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**Table 1.** Site overview, from north to south. Temperature and precipitation are average values of measurements by the respectively closest weather station in the period 1961–1990 (1958–1987 for Zackenberg).

Site	Location	Coordinates	Air temp.	Precipitation	Ecosystem type	Data year
Adventdalen	Svalbard	78°11′ N, 15°55′ E	−6.7 °C	190 mm yr <sup>−1</sup>	Fen	2013
Zackenberg	NE Greenland	74°30′ N, 21°00′ W	−9.9 °C	286 mm yr <sup>−1</sup>	Fen	2010
Stordalen	N Sweden	68°22′ N, 19°03′ E	−0.8 °C	304 mm yr <sup>−1</sup>	Mixed peatland	2012
Kobbefjord	W Greenland	64°08′ N, 52°23′ W	−1.4 °C	752 mm yr <sup>−1</sup>	Fen	2012
Fäjemyr	S Sweden	56°15′ N, 13°33′ E	6.2 °C	700 mm yr <sup>−1</sup>	Bog	2008

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**Table 2.** Summary statistics of all chambers. Temporal variability is expressed as daily standard deviation divided by daily mean (not shown for CO<sub>2</sub>). Spatial variability is expressed as the average over time of the ratio of standard deviation and mean of the individual chambers.

Site	Gas	Fluxes [#]	Difference to reference		Temporal variability			Spatial var.
			Exp. [%] ( $R^2$ )	NDFE [%] ( $R^2$ )	Ref. [%]	Exp. [%]	NDFE [%]	Ref. [%]
Adventdalen	CH <sub>4</sub>	1871	3.8 (0.99)	7.5 (0.97)	6.6	5.9	7.6	117.4
	CO <sub>2</sub>	1634	13.2 (0.98)	–	–	–	–	44.8
Zackenbergl	CH <sub>4</sub>	7092	3.1 (0.98)	22.1 (0.84)	14.2	15.2	26.1	93.0
	CO <sub>2</sub>	7809	9.1 (0.96)	–	–	–	–	46.5
Stordalen	CH <sub>4</sub>	1071	5.9 (0.73)	120.6 (0.20)	27.6	37.8	73.3	130.3
	CO <sub>2</sub>	1640	–15.5 (0.81)	–	–	–	–	82.2
Kobbefjord	CH <sub>4</sub>	8039	–0.4 (0.94)	10.0 (0.54)	7.0	7.8	13.8	28.2
	CO <sub>2</sub>	8839	–6.8 (0.98)	–	–	–	–	40.0
Fäjemyr	CH <sub>4</sub>	6986	–1.4 (0.83)	41.3 (0.00)	40.1	40.4	62.6	71.1
	CO <sub>2</sub>	6289	–19.1 (0.77)	–	–	–	–	64.3

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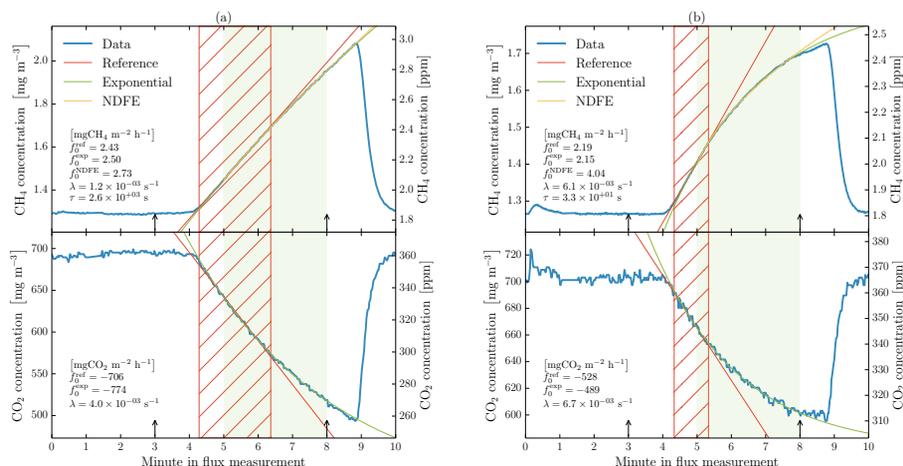
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**Figure 1.** Two examples of CH<sub>4</sub> (top) and CO<sub>2</sub> (bottom) flux measurements. **(a)** Chamber 6 at Zackenberg on 12 July 2010, 09:50 LT (hourly average wind speed 1.8 m s<sup>-1</sup>). **(b)** Same chamber on 2 July 2010, 13:50 LT (hourly average wind speed 4.5 m s<sup>-1</sup>). The arrows indicate chamber closing and opening time. The red hatched band indicates the time window used for the linear fit of the reference (Mastepanov et al., 2013). The shaded green band indicates the fixed 3 min window used for the curvilinear fits.

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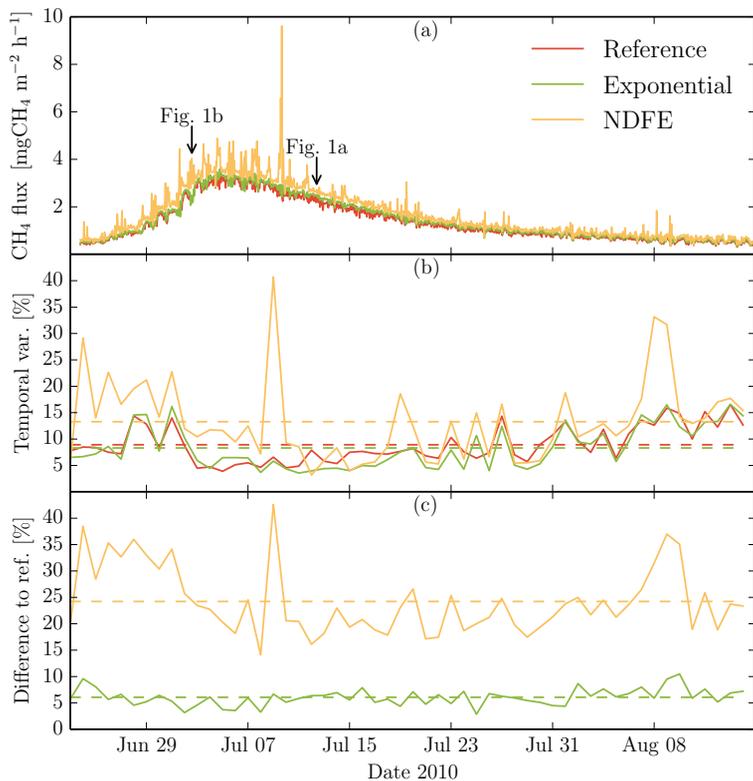
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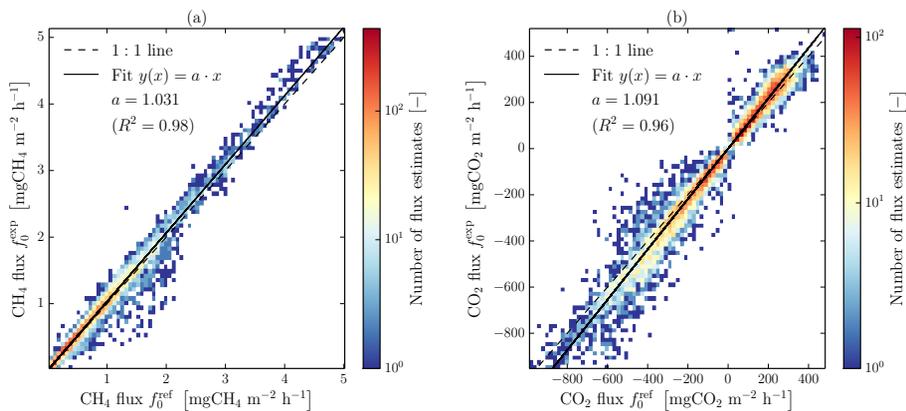




**Figure 2.** Results of chamber 6 at Zackenberg. **(a)** CH<sub>4</sub> flux in measurement time resolution (hourly). The arrows indicate the two examples of Fig. 1. **(b)** Flux temporal variability expressed as daily standard deviation divided by daily mean. **(c)** Mean daily ratio with respect to the reference data. Dashed lines indicate mean values of the entire time series.

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**Figure 3.** Example histograms of the relationship between reference and exponential flux estimates for all chambers of Zackenberg. **(a)** CH<sub>4</sub>. **(b)** CO<sub>2</sub>.

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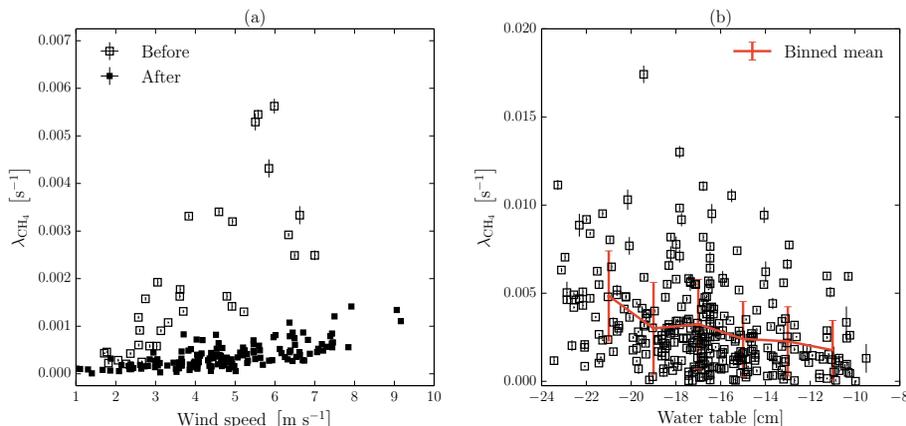
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**Figure 4.** Curvature parameter  $\lambda_{\text{CH}_4}$  against environmental parameters. **(a)** Wind speed. Data recorded by chamber 3 in Adventdalen between 26 July 2013 and 21 August 2013. On 4 August 2013 the sealing of the chamber was improved, so different markers are used here for times *Before* and *After* the improvement. **(b)** Water table position. Data from Fäjemyr between 1 June 2008 and 31 July 2008. Error bars indicate standard errors as calculated by the least-squares fit.

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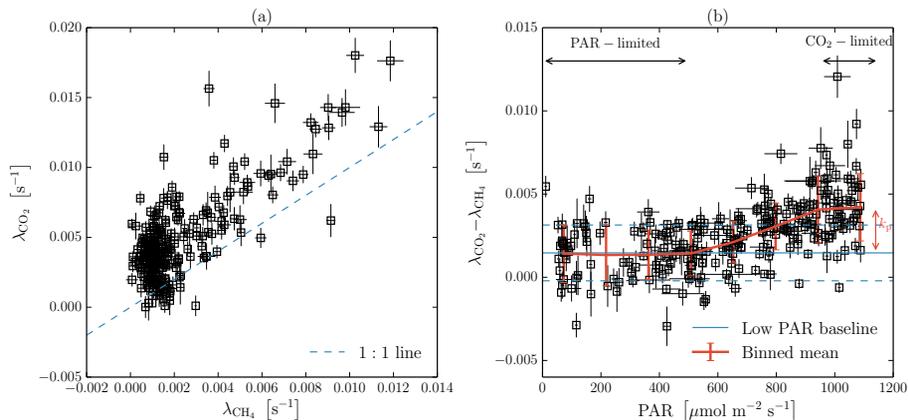
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**Figure 5.** Example of curvature correlation **(a)**, and curvature difference against PAR **(b)**. All data taken from chamber 3 at Zackenberg between 17 July 2010 and 5 August 2010. Error bars indicate standard errors as calculated by the least-squares fit.

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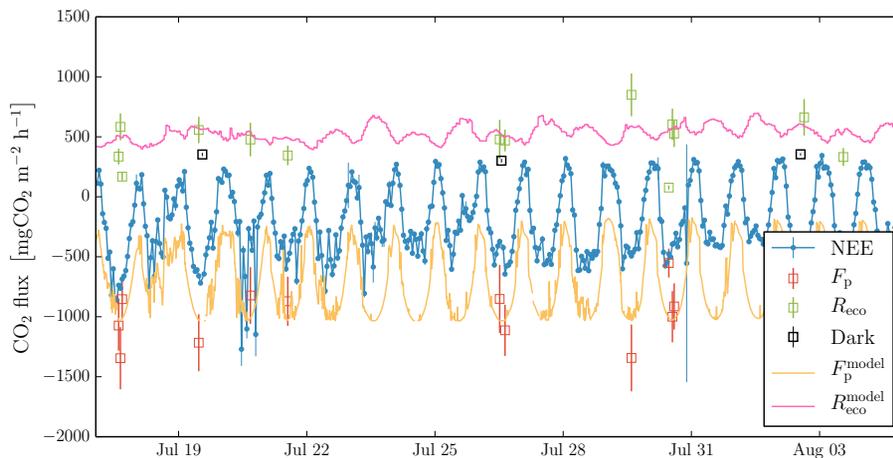
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**Figure 6.** Example of  $\text{CO}_2$  flux partitioning based on high PAR data points of Fig. 5, i.e. from chamber 3 at Zackenberg, 2010. NEE is the total  $\text{CO}_2$  flux,  $F_p$  the photosynthesis estimate derived from the curvatures, and  $R_{\text{eco}}$  their difference. *Dark* measurements are taken with a light-prove blanket over the chamber. Modeled lines are estimates from the day-time partitioning method of Lasslop et al. (2010).

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