

1 **Microbial methanogenesis in the sulfate-reducing zone of surface sediments traversing the**  
2 **Peruvian margin**

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9

10 **Abstract**

11 We studied the concurrence of methanogenesis and sulfate reduction in surface sediments (0-25 cm  
12 below seafloor) at six stations (70, 145, 253, 407, 990 and 1024 m) along the Peruvian margin (12°S).

13 This oceanographic region is characterized by high carbon export to the seafloor creating an extensive  
14 oxygen minimum zone (OMZ) on the shelf, both factors that could favor surface methanogenesis.

15 Sediments sampled along the depth transect traversed areas of anoxic and oxic conditions in the  
16 bottom-near water. Net methane production (batch incubations) and sulfate reduction (<sup>35</sup>S-sulfate  
17 radiotracer incubation) were determined in the upper 0-25 cmbsf of multiple cores from all stations,  
18 while deep hydrogenotrophic methanogenesis (> 30 cmbsf, <sup>14</sup>C-bicarbonate radiotracer incubation)  
19 was determined in two gravity cores at selected sites (78 and 407 m). Furthermore, stimulation  
20 (methanol addition) and inhibition (molybdate addition) experiments were carried out to investigate  
21 the relationship between sulfate reduction and methanogenesis.

22 Highest rates of methanogenesis and sulfate reduction in the surface sediments, integrated over 0-25  
23 cmbsf, were observed on the shelf (70-253 m, 0.06-0.1 mmol m<sup>-2</sup> d<sup>-1</sup> and 0.5-4.7 mmol m<sup>-2</sup> d<sup>-1</sup>,  
24 respectively), while lowest rates were discovered at the deepest site (1024 m, 0.03 and 0.2 mmol m<sup>-2</sup> d<sup>-1</sup>,  
25 respectively). The addition of methanol resulted in significantly higher surface methanogenesis  
26 activity, suggesting that the process was mostly based on non-competitive substrates, i.e., substrates  
27 not used by sulfate reducers. In the deeper sediment horizons, where competition was probably

28 relieved due to the decrease of sulfate, the usage of competitive substrates was confirmed by the  
29 detection of hydrogenotrophic activity in the sulfate-depleted zone at the shallow shelf station (70 m).  
30 Surface methanogenesis appeared to be correlated to the availability of labile organic matter (C/N  
31 ratio) and organic carbon degradation (DIC production), both of which support the supply of  
32 methanogenic substrates. A negative correlation between methanogenesis rates and dissolved oxygen  
33 in the bottom-near water was not obvious, however, anoxic conditions within the OMZ might be  
34 advantageous for methanogenic organisms at the sediment-water interface.

35 Our results revealed a high relevance of surface methanogenesis on the shelf, where the ratio between  
36 surface to deep (below sulfate penetration) methanogenic activity ranged between 0.13 and  $10^5$ . In  
37 addition, methane concentration profiles indicated a partial release of surface methane into the water  
38 column as well as consumption of methane by anaerobic methane oxidation (AOM) in the surface  
39 sediment. The present study suggests that surface methanogenesis might play a greater role in benthic  
40 methane budgeting than previously thought, especially for fueling AOM above the sulfate-methane  
41 transition zone.

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43

44 *Keywords: Oxygen minimum zone, organic matter, competition, anaerobic oxidation of methane,*  
45 *emission*

46

## 47 **1. Introduction**

48 Microbial methanogenesis represents the terminal step of organic matter degradation in marine  
49 sediments (Jørgensen, 2006). The process is entirely restricted to a small group of prokaryotes within  
50 the domain of the Archaea (Thauer, 1998). Methanogens produce methane from a narrow spectrum of  
51 substrates, primarily carbon dioxide (CO<sub>2</sub>) and hydrogen (H<sub>2</sub>) (hydrogenotrophic pathway), as well as  
52 acetate (acetoclastic pathway) (Zinder, 1993). In addition, methanol or methylated compounds such as  
53 methylamine can be utilized (methylotrophic pathway) (Oremland & Polcin, 1982; Buckley et al.,  
54 2008; Zinder, 1993; King et al., 1983). Substrates for methanogenesis are produced during  
55 depolymerization and fermentation of organic macromolecules (e.g., sugars, vitamins, amino acids) to

56 smaller monomeric products (Jørgensen, 2006; Schink & Zeikus, 1982; Neill et al., 1978; Donnelly &  
57 Dagley, 1980).

58 Acetoclastic and hydrogenotrophic methanogenesis are predominantly found in deeper sediment zones  
59 below sulfate penetration, owing to the more effective utilization of H<sub>2</sub> and acetate by sulfate reducers  
60 due to their higher substrate affinity (Oremland & Polcin 1982; Jørgensen 2006). Methanogens can  
61 avoid competition with sulfate reducers by the utilization of non-competitive substrates, such as  
62 methanol or methylamines (Oremland & Polcin, 1982; King et al., 1983). Facilitated by the usage of  
63 such non-competitive substrates, sulfate reduction and methanogenesis were found to co-occur in  
64 sulfate-containing salt marsh sediments (Oremland et al., 1982; Buckley et al., 2008; Senior et al.,  
65 1982). Concurrent activity of sulfate reduction and methanogenesis in the marine environment has  
66 mostly been postulated for organic-rich sediments (Mitterer, 2010; Jørgensen & Parkes, 2010; Treude  
67 et al., 2009, 2005a; Hines & Buck, 1982; Crill & Martens, 1986); however, details on the magnitude  
68 and environmental controls of surface methanogenesis are still poorly understood (Holmer &  
69 Kristensen, 1994; Ferdelman et al., 1997).

70 In a study from Eckernförde Bay, southwestern Baltic Sea, considerable in vitro methanogenic activity  
71 was observed in samples taken from 5 to 40 cm sediment depth (Treude et al. 2005). Although in vitro  
72 activity was measured in sulfate-free setups, methanogenic activity coincided with zones of in-situ  
73 sulfate reduction. The authors concluded a coexistence of the two types of organisms, which could be  
74 enabled through either the usage of non-competitive substrates, dormancy of methanogens until phases  
75 of sulfate depletion, and/or temporal or spatial heterogeneity in the sediments. Eckernförde Bay  
76 sediments feature a high input of organic matter due to a shallow water depth (~30 m) and pronounced  
77 phytoplankton blooms in spring, summer, and fall (Smetacek, 1985). Furthermore, seasonal hypoxia  
78 (O<sub>2</sub> < 90 μM) or even anoxia (O<sub>2</sub>=0 μM) occur in the deep layers of the water column caused by  
79 stratification and degradation of organic matter ( Bange et al. 2011). Oxygen-depleted conditions in  
80 the bottom water together with frequent input of fresh organic matter possibly favors methanogenesis  
81 in surface sediment by offering reduced conditions and non-competitive substrates. As non-  
82 competitive substrates can be derived from organic osmolytes such as betaine or  
83 dimethylsulfoniopropionate (DMSP), a high load of organic matter (e.g. through sedimentation of

84 phytoplankton blooms) can increase the availability of non-competitive substrates (Zinder, 1993; Van  
85 Der Maarel & Hansen, 1997). In accordance, methanogenesis activity was observed within the sulfate-  
86 reducing zone of organic-rich sediments from the seasonally hypoxic Limfjorden sound, Northern  
87 Denmark (Jørgensen & Parkes, 2010; Jørgensen, 1977).

88

89 The environmental relevance of surface methanogenesis is hitherto unknown. Its closeness to the  
90 sediment-water interface makes it a potential source for methane emissions into the water column,  
91 unless the methane is microbially consumed before escaping the sediment (Knittel & Boetius, 2009).  
92 Methane escapes the sediment either by diffusion or, when methane saturation is exceeded, in the form  
93 of gas bubbles (Whiticar, 1978; Wever & Fiedler, 1995; Judd et al., 1997; Dimitrov, 2002). The  
94 fraction of methane released to the water column that reaches the atmosphere mainly depends on  
95 water depth, as methane is also consumed within the water column through aerobic microbial  
96 oxidation (Reeburgh, 2007; Valentine et al., 2001). Thus, shallow coastal areas have higher methane  
97 emission potentials than the open ocean (Bange et al., 1994) and a greater potential to contribute to  
98 methane-dependent atmospheric warming (IPCC, 2014).

99

100 In the present study we focused on the upwelling region off the Peruvian coast, which is another  
101 example of an environment where both factors that potentially favor surface methanogenesis convene,  
102 i.e., a high export of organic carbon and low dissolved oxygen concentrations in the bottom water.  
103 This upwelling region represents one of the most productive systems in the world oceans, creating one  
104 of the most intense oxygen minimum zones (OMZ, Kamykowski & Zentara 1990; Pennington et al.  
105 2006). Oxygen concentrations in waters impinging on the seafloor are below 20  $\mu\text{M}$  or even reach  
106 anoxia. Research on surface sediment methanogenesis in upwelling regions is scarce and its potential  
107 role in the carbon cycling of the Peruvian OMZ is completely unknown. In a study from the central  
108 Chilean upwelling area (87 m water depth, 0.5-6 cm sediment depth), low methane production rates  
109 were detected despite high sulfate reduction activity, when the non-competitive substrate  
110 trimethylamine was offered (Ferdelman et al., 1997). The authors concluded that the prevailing

111 methanogens were competing with sulfate reducers for H<sub>2</sub> and with acetogens for methylamines,  
112 explaining the overall low methanogenesis activity observed (Ferdelman et al., 1997).  
113 Even though the Chilean and Peruvian OMZs are connected, commonly known as OMZ in the Eastern  
114 South Pacific Ocean (ESP) (Fuenzalida et al., 2009), the core of the ESP-OMZ is centered off Peru  
115 with an upper boundary at < 100 m and a vertical distribution to > 600 m versus a thinner OMZ band  
116 off Chile constrained between 100-400 m water depth (Fuenzalida et al., 2009). Anoxic conditions in  
117 the water column of the OMZ core (and therewith a lack of bioirrigating macrofauna introducing  
118 oxygen into the sediments (Kristensen, 2000)) together with the high export rates of labile organic  
119 carbon to the seafloor (Reimers & Suess, 1983; Dale et al., 2015) provide favorable conditions for  
120 methanogenesis activity in surface sediments, thus increasing the potential for benthic methane  
121 emissions.

122 Here, we provide first insights into surface methanogenesis in sediment cores (< 30 cmbsf =  
123 centimeters below seafloor) taken along the Peruvian shelf and margin. We hypothesize that  
124 methanogenesis coexists with sulfate reduction through the utilization of non-competitive substrates.  
125 In addition, we postulate that surface methanogenesis depends on the quantity and quality (=   
126 freshness) of organic carbon, and the concentrations of dissolved oxygen in the bottom water. We  
127 therefore expect spatial variability of surface methanogenesis along the continental shelf and margin.  
128 The observed methanogenic activity will be compared to methane concentrations in the bottom-near  
129 water to discuss the potential relevance of surface methanogenesis for methane emissions into the  
130 pelagic zone.

131

## 132 **2. Material and Methods**

### 133 **2.1 Study site and sediment sampling**

134 Samples were taken during the R.V. Meteor cruise M92 between 5. Jan and 3. Feb 2013 along a depth  
135 transect off the Peruvian coast from the shelf (~70 m) to the continental margin (~1000 m). The  
136 transect was located in the central part of the ESP-OMZ (Fuenzalida et al., 2009) at 12°S. Further  
137 hydrographic details on the study area can be found elsewhere (Dale et al., 2015).

138 Sediment cores for the determination of near-surface methanogenesis were collected at six stations  
139 along the depth transect at 70, 145, 253, 407, 776 and 1024 m water depth (Fig.1), using a multiple  
140 corer with a mounted camera (TV-MUC). The MUC held seven cores (length: 60 cm, inner diameter:  
141 10 cm) and covered an area of  $\sim 1 \text{ m}^2$ . If necessary, a second MUC was deployed at the same station,  
142 thus sediment cores could originate from different MUC casts. Station numbers were assigned in  
143 accordance with Dale et al., (2015). After retrieval, sediment cores were transferred to a  $\sim 9^\circ\text{C}$  cold  
144 room and processed at the same day.

145 In addition to the MUC, a gravity corer was deployed at two stations (78 and 407 m) for determining  
146 deep methanogenesis. The total core length was 400 cm and 206 cm, respectively. The gravity corer  
147 was equipped with a 260 kg weight and a 5 m steel barrel (diameter: 14 cm). The replaceable core  
148 liner (PVC, diameter: 12.5 cm) was housed within the barrel and fixed with a core catcher. After  
149 retrieval, sediment cores from the gravity corer were sliced into 1-m sections, capped on both sides,  
150 and brought to the cold room ( $4^\circ\text{C}$ ) for further processing. Relevant station details for MUC and  
151 gravity cores are summarized in Table 1.

152

## 153 **2.2 Water column sampling**

154 CTD/Rosette water column casts were conducted at the same station as sediment coring (for details  
155 see Table 1). Temperature and oxygen data were taken from Dale et al. 2015.

156 For the analysis of methane concentrations in the bottom-near water, water was sampled ca. 1.5 m  
157 above the seafloor from 10 L Niskin bottles mounted on the CTD/Rosette r. The collected water was  
158 filled bubble-free into 60 ml vials (triplicates), each vial containing 3 pellets of sodium hydroxide  
159 (NaOH,  $\sim 0.3 \text{ M}$  per vial) to stop microbial activity and force dissolved gas into the headspace. After  
160 closing the vials with a butyl rubber stopper and a crimp seal, 10 ml of water was removed with a  $\text{N}_2$ -  
161 flushed 10 ml syringe and replaced with  $\text{N}_2$  gas from a second syringe to create a headspace in the  
162 sampling vials. Samples were stored and transported at room temperature until further processing.

163 In the home laboratory, 100  $\mu\text{l}$  of the headspace volume was injected into a Shimadzu GC-2014 gas  
164 chromatograph equipped with a flame ionization detector and a HaySep-T 100/120 column (Length 3  
165 m, diameter: 2 mm). Gases were separated isothermally at  $75^\circ\text{C}$  with helium carrier gas. Methane

166 concentrations were calibrated against methane standards (Scotty gases). The detection limit was 0.1  
167 ppm with a precision of 2 %.

168

### 169 **2.3 Porewater geochemistry**

170 Porewater sampling for MUC cores has been previously described by Dale et al., (2015). In short, one  
171 MUC core per station was subsampled in an argon-filled glove bag to preserve redox sensitive  
172 constituents.

173 The gravity cores at St. 1 (78 m) and St. 8 (407 m) were subsampled at 10-12 different sediment  
174 depths (depending on core length) resulting in depth intervals of 20-33 cm. Before sampling, the  
175 plastic core liner was cut open with an electric saw at the specific depths. Porewater was extracted by  
176 using anoxic (flushed with argon), wetted rhizons (Rhizosphere Research Products, Seeberg-Elverfeldt  
177 et al., 2005).

178 Sulfate concentrations were determined by ion chromatography (Methrom 761) as described  
179 previously by Dale et al., (2015).

180 For DIC analysis, 1.8 ml of porewater was transferred into a 2 ml glass vial, fixed with 10 µl saturated  
181 mercury chloride solution and crimp sealed. Samples were stored at 4°C until further processing in the  
182 home laboratory. DIC concentration was determined as CO<sub>2</sub> with a multi N/C 2100 analyzer (Analytik  
183 Jena) following the manufacturer's instructions. Therefore the sample was acidified with phosphoric  
184 acid and the outgassing CO<sub>2</sub> was measured. The detection limit was 20 µM with a precision of 2-3%.

185

### 186 **2.4 Sediment porosity and particulate organic carbon/nitrogen**

187 Methodology and data for porosity, particulate organic carbon (POC) and particulate organic nitrogen  
188 (PON) have been previously described by Dale et al., (2015).

189 In short, wet sediment samples were taken from the porewater MUC core and the gravity cores for  
190 determination of porosity from the weight difference of wet and freeze-dried sediment. POC and PON  
191 were analyzed with a Carlo-Erba element analyzer (NA 1500). Ratios of POC:PON were calculated by  
192 division.

193

## 194 **2.5 Sediment methane**

195 For sediment methane concentration, one MUC core per station was sliced in 2 cm intervals until 20  
196 cm depth, followed by 5 cm intervals until the end of the core (maximum depth = 48 cm). Gravity  
197 cores were subsampled according to the above scheme (see 2.3). From each sampled sediment layer, 2  
198 cm<sup>3</sup> sediment were transferred into a 15 ml serum glass vial containing 5 ml of NaOH (2.5% w/w).  
199 The vial was closed with a butyl stopper, crimp sealed and shaken thoroughly to stop microbial  
200 activity and to force all methane into the headspace. Vials were stored upside down at room  
201 temperature until measurement in the home laboratory.  
202 Sediment methane concentration was determined by injecting 0.1 ml of headspace volume into a  
203 Shimadzu GC-2014 gas chromatograph as described under section 2.2.

204

## 205 **2.6 Net methanogenesis activity in MUC cores**

206 Sediment from MUC cores was used to determine net methanogenesis, which is defined as the sum of  
207 total methane production and consumption, including all available methanogenic substrates in the  
208 sediment. Net methanogenesis was determined by measuring the linear increase of methane  
209 concentration in the headspace of closed incubation vials over time. Therefore, one MUC core per  
210 station was sliced into 5 cm intervals, transferring 10 cm<sup>3</sup> of sediment in triplicates into N<sub>2</sub>-flushed 60  
211 ml serum glass vials. The sediment core lengths ranged from 25 up to 48 cm, resulting in maximum 10  
212 depth intervals. Ten ml of anoxic deep water overlying each MUC core was added to the vial and the  
213 slurry was mixed under a constant N<sub>2</sub> stream (Hungate, 1950) before being sealed with a butyl rubber  
214 stopper and crimped. The sediment slurry was repeatedly flushed with N<sub>2</sub> through the stopper to  
215 ensure fully anoxic conditions. The vials were incubated in the dark and at 9°C, which reflected the  
216 average in situ temperature along the depth transect (see Table 1). The first gas chromatographic  
217 measurement was done directly after preparation of the vials, by injecting 100 µl of headspace sample  
218 into the gas chromatograph. The on-board Hewlett Packard-5890 gas chromatograph was equipped  
219 with a flame ionization detector and a HaySep-T 100/120 column (Length 3m, diameter: 2mm). Gases  
220 were separated isothermally at 75°C with helium carrier gas. Methane concentrations were calibrated



221 against methane standards. The detection limit was 1 ppm with a precision of < 5%. Measurements  
222 were done in 2-4 day-intervals over a total incubation time of ~2 weeks.

223

## 224 **2.7 Potential non-competitive and competitive methanogenesis in sediment slurries from MUC** 225 **cores**

226 Sediment slurry experiments were conducted with sediment from St. 1 (70 m) to examine the  
227 interaction between sulfate reduction and methanogenesis, as this station revealed highest microbial  
228 activity of sulfate reduction and methanogenesis. On board, the sediment core was sliced in 5 cm  
229 intervals. Sediment from the 0-5 cm interval and the 20-25 cm interval was transferred into 250 ml  
230 glass bottles, which were then closed without headspace (filled to top) with a butyl rubber stopper and  
231 screw cap. Until further treatment, sediment was stored at 4°C on board and later in a 1°C cold room  
232 on shore.

233 Approximately 6 months after the cruise, sediment slurries from both depth intervals were prepared by  
234 mixing 5 ml sediment in a 1:1 ratio with artificial, fully marine seawater (Widdel & Bak, 1992) before  
235 further manipulations.

236 In total, three different treatments, each in triplicates, were prepared per depth: 1) sulfate-rich (28  
237 mM), serving as a control 2) sulfate-rich plus molybdate (22 mM) from now on referred to as  
238 molybdate-treatment, and 3) sulfate-rich plus methanol (10 mM) from now on referred to as methanol-  
239 treatment.

240 Molybdate was used as an enzymatic inhibitor for sulfate reduction (Oremland & Capone, 1988).

241 Methanol is a known non-competitive substrate used by methanogens, but not by sulfate reducers  
242 (Oremland & Polcin, 1982), which makes it suitable to examine non-competitive methanogenesis.

243 The sediment slurries were incubated at 9°C in the dark for 23 days and headspace concentration of  
244 methane was measured repeatedly over time on a gas chromatograph. Therefore, 100 µl of headspace  
245 was removed from the gas vials and injected into a Shimadzu gas chromatograph (GC-2014) equipped  
246 with a methanizer (inactive), a packed Haysep-D column and a flame ionization detector. The column  
247 temperature was 80°C and the helium flow was set to 12 ml min<sup>-1</sup>. Methane concentrations were  
248 measured against methane standards. The detection limit was 0.1 ppm with a precision of <5%. Rates

249 were determined from the linear increase of methane concentration over time. Due to differences in  
250 the linear increase between the three treatments, rates were determined at two different time points: the  
251 first period of incubation includes the starting point (day 0) until day 5, the second period includes day  
252 8 to day 23 (Supplement information, Fig. S1).

253 Student's t-test (independent, two-tailed,  $\alpha = 0.05$ ) was applied to detect significant differences  
254 between the three different treatments.

255

## 256 **2.8 Gross hydrogenotrophic methanogenesis activity in gravity cores**

257 For the determination of surface to deep methanogenesis activity in gravity cores the radiotracer  
258 technique using  $^{14}\text{C}$ -bicarbonate was applied (Jørgensen, 1978). With this method only  
259 hydrogenotrophic methanogenesis from  $\text{CO}_2/\text{H}_2$  can be determined, which is the expected main  
260 pathway in deeper sediment layers.

261 Sampled sediment depths were according to the sampling scheme described under section 2.3. Circa 5  
262  $\text{cm}^3$  of sediment was sampled in triplicates into glass tubes equipped with syringe plungers and then  
263 sealed headspace-free with butyl rubber stoppers. Then,  $^{14}\text{C}$ -bicarbonate-tracer (dissolved in water, pH  
264 = 8-9, injection volume 6  $\mu\text{l}$ , activity 222 kBq, specific activity 1.85-2.22 GBq/mmol) was injected  
265 through the stopper. The vials were incubated for 48 hours at  $9^\circ\text{C}$  before the reaction was stopped by  
266 transferring the sediment into 50 ml glass vials filled with 20 ml NaOH (2.5%), closed with butyl  
267 rubber stoppers and shaken thoroughly. Five controls were produced from various sediment depths by  
268 injecting the radiotracer directly into the NaOH with sediment.

269 In the home laboratory,  $^{14}\text{C}$ - methane production was determined with the slightly modified method by  
270 Treude et al., (2005a) used for the determination of anaerobic oxidation of methane. The method was  
271 identical, except that no unlabeled methane was determined by gas chromatography. Instead, DIC  
272 values were used to calculate hydrogenotrophic methane production (=  $\text{CO}_2$  reduction):

273

$$MG\ rate = \frac{{}^{14}\text{CH}_4 * [\text{DIC}]}{({}^{14}\text{CH}_4 + {}^{14}\text{C-DIC}) * t}$$

274

275

276 The methanogenesis rate (*MG rate*) is expressed in  $\text{nmol CH}_4 \text{ cm}^{-3} \text{ sediment d}^{-1}$ ,  $^{14}\text{CH}_4$  is the activity  
277 of produced  $^{14}\text{CH}_4$ ,  $^{14}\text{C-DIC}$  is the activity of residual radioactive dissolved organic carbon ( $\text{DIC} = \text{CO}_2$   
278  $+ \text{HCO}_3^- + \text{CO}_3^{2-}$ ),  $[\text{DIC}]$  is the concentration of dissolved organic carbon in  $\text{nmol cm}^{-3} \text{ sediment}$ , and  $t$   
279 is the incubation time in days.

280

## 281 **2.9 Sulfate reduction in MUC cores**

282 One MUC core per station was used for the determination of sulfate reduction. First, two replicate  
283 push cores (length 30 cm, inner diameter 2.6 cm) were subsampled from one MUC core. The actual  
284 core length varied from 23-25 cmbsf total length. Then, 6  $\mu\text{l}$  (~150 kBq) of carrier-free  $^{35}\text{SO}_4^{2-}$   
285 radiotracer (dissolved in water, specific activity 37 TBq  $\text{mmol}^{-1}$ ) was injected into the replicate  
286 pushcores in 1 cm intervals according to the whole-core injection method of Jørgensen, (1978). Push  
287 cores were incubated for ca. 12 h at 9°C. After incubation, bacterial activity was stopped by slicing the  
288 push core into 1-cm intervals and transferring each sediment layer into 50 ml plastic centrifuge tubes  
289 filled with 20 ml zinc acetate (20% w/w). Controls were done in triplicates from different depths.  
290 Here, the sediment was first fixed with zinc acetate before adding the tracer. Rates for sulfate  
291 reduction were determined using the cold chromium distillation procedure according to Kallmeyer et  
292 al., (2004).

293 The yielded sulfate reduction rates have to be treated with caution, due to long (up to 3 half-life times  
294 of  $^{35}\text{S}$ ) and unfrozen storage. Storage of sulfate reduction samples without freezing has recently been  
295 shown to result in the re-oxidation of  $^{35}\text{S}$ -sulfides, which results in an underestimation of sulfate  
296 reduction rates (Røy et al., 2014). During this reaction, zinc sulfide ( $\text{Zn}^{35}\text{S}$ ) and iron sulfide ( $\text{Fe}^{35}\text{S}$ ) are  
297 re-oxidized to sulfate by reactive Fe(III), which originates from the reaction of  $\text{Fe}^{2+}$  with oxygen.  $\text{Fe}^{2+}$   
298 is released during the gradual conversion of FeS to ZnS, which has the lower solubility product. Still,  
299 we do trust the relative distribution of activity along depth profiles and consider a potential  
300 underestimation of absolute rates.

301

302

### 303 3. Results

#### 304 3.1 Water column oxygen and methane concentration

305 Dissolved oxygen in the bottom water was below detection limit from St.1 (70 m) to St. 8 (407 m),  
306 subsequently increasing with water depth to 53  $\mu\text{M}$  at the deepest site (see Table 1 and Dale et al.,  
307 2015). At the shallowest St. 1 (70 m) the water was turbid and smelled of sulfide.

308 Dissolved methane concentrations in the bottom water were high on the shelf (St.1-6, 70-253 m) and  
309 10 fold lower at the deeper sites (St. 8-10, 407-1024 m; Table 1). The highest measured methane  
310 concentration was detected at St. 6 (253 m,  $\sim 80$  nM) and lowest concentrations were detected at St. 10  
311 (1024 m,  $\sim 4$  nM).

#### 312 3.2 Sediment core description

313 A detailed sediment description for the porewater geochemistry cores has been already published in  
314 detail by Dale et al., (2015). In short, sediments revealed a grey color with a black surface layer at St.  
315 1 (70 m), a dark olive green color at St. 4-8 (145-407 m), and a green-brown color at St. 9 and 10 (770  
316 -1024 m). Sediment texture was soft and fluffy at St. 1-6 (70-253 m), and was less soft at the deeper  
317 sites. St. 8 (407 m) revealed a fluffy surface layer followed by a dense clay layer  $> 2$  cmbsf sediment  
318 depth. In addition, phosphorite nodules were found at the sediment surface (0-2 cmbsf) of St. 8 (407  
319 m).

320 Mats of the sulfur oxidizing bacteria *Thioploca spp.* (Gallardo, 1977) were visible at the sediment  
321 surface at St.1-6 (70-253 m), with the densest mat at St. 1 (70 m) continuously decreasing with  
322 increasing water depth. Sheaths of *Thioploca* were visible until 20-30 cmbsf at St. 1, 4 and 6 (70-253  
323 m).

324 Foraminifera could be observed at the sediment surface of St. 8 (407 m), St. 9 (770 m) and St. 10  
325 (1024 m). St. 8 (407 m) showed a thick layer of foraminifera ooze on the sediment surface (0-3 cmbsf)  
326 while St. 9 (770 m) and St. 10 (1024 m) showed only scattered foraminifera at the sediment surface (0-  
327 5 cmbsf).

328 Macrofauna (large polychaetes, oligochaetes, ophiuroids) were restricted to the sites below the OMZ  
329 at St. 9 (770 m) and St. 10 (1024 m), where deep waters were oxygenated. However, small snails (~ 1  
330 cm) were observed at St. 8 (407 m).

331

### 332 3.3 Geochemical parameters in MUC cores

333 Porewater and solid phase geochemistry of sediments retrieved by the MUC cores are shown in Fig. 2.

334 Surface sediment (0-0.5 cmbsf) POC content increased along the continental shelf from 1.6 wt % at  
335 the shallow St. 1 (70 m) to a maximum of 15 wt % at St. 8 (253 m). Surface POC content decreased  
336 again with increasing water depth showing the lowest POC content at St. 10 (1024 m, 2 wt %). While  
337 POC content showed more or less stable profiles throughout the sediment core at St. 1 (70 m, around 3  
338 wt %), St. 9 (770 m, around 4 wt %), and St. 10 (1024 m, around 3 wt %), POC content was stable  
339 only in the upper ~ 10 cmbsf at St. 4 (150 m, around 10 wt %) and St. 6 (253 m, around 15 wt %),  
340 followed by a decrease until the deepest sampled depth (2 wt % and 9 wt %, respectively). At St. 8  
341 (407 m), POC content increased with sediment depth below 3 cmbsf (from 4 wt % to 9 wt %), which  
342 consisted of dense clay (see above). In the upper 3 cmbsf, POC decreased from ~ 7 wt % to ~ 4 wt %,   
343 which was the sediment layer with a more fluffy appearance.

344 The sediment surface C/N ratio was lowest at St. 1 (70 m, 6.2) and increased along the continental  
345 shelf showing the highest surface C/N ratio at St. 10 (1024 m, 11). St. 8 (407 m) was exceptional, as it  
346 showed slightly lower surface C/N ratio (8) as at St. 6 (253 m, 9). St. 8 (407 m) was also the only site  
347 showing an increase of 4 units in the upper 0-5 cmbsf, followed by stable ratios around 12 throughout  
348 the rest of the core. St. 1 and 4 (70 and 145 m) showed shallower increases in C/N ratio in the upper ~  
349 2 cmbsf and upper 1 cmbsf, respectively, followed by stable ratios around 10 until the bottom of the  
350 core. At St. 9 and 10 (770 and 1024 m), C/N ratios ranged around 11 and 12, respectively.

351 The highest increase in methane concentration was observed at St. 1 (70 m). Here, methane increased  
352 linearly from the surface (1  $\mu\text{M}$ ) to the bottom of the core (100  $\mu\text{M}$ ). All other stations showed either  
353 no clear trend (St. 4= 145 m) or only slight methane increases with depth. At St. 9 (770 m), even a  
354 decrease in methane concentration was observed from the surface to the bottom of core.

355 Besides St. 1 (70 m), which showed a strong decrease in sulfate ( $\text{SO}_4^{2-}$ ) concentration with depth from  
356 about 28 mM at the top to about 9 mM at the bottom of the core (43 cmbsf), all other stations showed  
357  $\text{SO}_4^{2-}$  concentrations  $> 25$  mM throughout the cores. At St. 4, 6 and 9 (145, 253, 770 m),  $\text{SO}_4^{2-}$  showed  
358 very slight decrease with depth from about 28 mM at the top to about 25 mM at the bottom of the core.  
359 Porewater  $\text{SO}_4^{2-}$  concentrations were stable around 28 mM throughout the core at St. 8 and 10 (407  
360 and 1024 m).

361 Dissolved inorganic carbon (DIC) concentration increased with depth at St. 1- 6 (70 -253 m). St. 1 (70  
362 m) showed the steepest increase with depth, showing the lowest DIC concentration at the top (2.3  
363 mM) and the highest at the deepest sampled depth (21.6 mM). At St. 4 (153 m), maximum  
364 concentration was reached at  $\sim 23$  cmbsf with 4 mM. St. 6 (253 m) showed maximum concentration at  
365 the deepest sampled depth with 9 mM. St. 8 and 9 (407 and 770 m) showed stable DIC concentrations  
366 around 2.3 mM throughout the core. No DIC data was available for St. 10 (1024 m).

367

### 368 **3.4 Net methanogenesis and gross sulfate reduction in MUC cores**

369 Maximum net methanogenesis rates (Fig. 2) were detected at St. 1 (70 m,  $1.1 \pm 0.5 \text{ nmol cm}^{-3} \text{ d}^{-1}$ , 20-  
370 25 cmbsf) and St. 6 (253 m,  $1.3 \pm 0.65 \text{ nmol cm}^{-3} \text{ d}^{-1}$ , 25-30 cmbsf). At all other stations,  
371 methanogenesis was mostly below  $0.5 \text{ nmol cm}^{-3} \text{ d}^{-1}$  throughout the cores. St. 8 (407 m) showed  
372 methanogenesis activity only in the top 10 cmbsf with the maximum at 5-10 cmbsf ( $0.2 \pm 0.5 \text{ nmol cm}^{-3}$   
373  $\text{d}^{-1}$ ). At St. 9 and 10 (770 and 1024 m), maximum methanogenesis activity was found in the surface  
374 layer (0-5 cmbsf) with  $0.3 \pm 0.4 \text{ nmol cm}^{-3} \text{ d}^{-1}$  and  $0.4 \pm 0.6 \text{ nmol cm}^{-3} \text{ d}^{-1}$ , respectively. St. 10 (1024 m)  
375 also showed high average methanogenesis at 10-15 cmbsf ( $1.5 \pm 2.5 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ), which was caused  
376 by a single high replicate ( $4.3 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ). In the following, e.g., integration of rates, we will  
377 exclude this single high replicate, which will be further elaborated in the discussion.

378 At all stations beside St. 9 (770 m), sulfate reduction activity was highest in the 0-1 cmbsf horizon,  
379 followed by a sharp decrease in activity of 20-90% in the subsequent 1-2 cmbsf horizon. Highest  
380 measured rates at 0-1 cmbsf were observed at St. 4 (145 m,  $290 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ), followed by St. 1 (70  
381 m,  $270 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ). Surface (0-1 cmbsf) sulfate reduction activity decreased from St. 4 (145 m) to

382 St. 8 (407 m) with concomitant increase in water depth. St. 9 (770 m) was the only site without a  
383 surface sulfate reduction maximum. Here, highest rates were found at 7 cmbsf ( $11.2 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ).  
384 St. 6, 8 and 9 (253, 407, and 770 m) showed a second but smaller maximum of sulfate reduction  
385 activity. At St. 6 (253 m), this second maximum was situated at 20.5 cmbsf ( $6.2 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ). St. 8  
386 and 9 (407 and 770 m) showed additional maxima at 4.5 cmbsf ( $3.1 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ) and 2.5 cmbsf ( $1.5$   
387  $\text{nmol cm}^{-3} \text{ d}^{-1}$ ), respectively. At St. 9 (770 m), sulfate reduction activity was not detectable at most  
388 depth  $> 10$  cmbsf. At St.10 (1024 m), no sulfate reduction activity was detectable throughout the entire  
389 core. At St. 9 and 10 (770 and 1024 m) we cannot exclude that sulfate reduction was present but  
390 undetectable due to long, unfrozen storage of the samples (see 2.7).  
391 Figure 3 shows an overview of integrated methanogenesis and sulfate reduction rates (over the upper  
392 0-25 cm) along the depth transect on the Peruvian margin. Highest integrated surface methanogenesis  
393 activity was detected on the shelf (70, 145 and 253 m) with  $0.1 \pm 0.03 \text{ mmol m}^{-2} \text{ d}^{-1}$ ,  $0.06 \pm 0.02 \text{ mmol}$   
394  $\text{m}^{-2} \text{ d}^{-1}$ , and  $0.07 \pm 0.01 \text{ mmol m}^{-2} \text{ d}^{-1}$ , respectively. St. 8 (407 m) revealed the lowest integrated  
395 methanogenesis rate of all sites ( $0.02 \pm 0.00 \text{ mmol m}^{-2} \text{ d}^{-1}$ ). St. 9 (770 m) and St. 10 (1024 m) showed  
396 integrated methanogenesis activity around  $0.03 \pm 0.02 \text{ mmol m}^{-2} \text{ d}^{-1}$ , respectively.  
397 Integrated sulfate reduction activity decreased along the continental margin with increasing water  
398 depth, revealing the highest activity at the St. 1 (70 m,  $4.7 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) and the lowest activity at St.  
399 9 (770 m,  $0.2 \text{ mmol m}^{-2} \text{ d}^{-1}$ ). Please note again, that integrated sulfate reduction rates are probably  
400 underestimated due to long, unfrozen storage of the samples (see 2.7).

401

### 402 **3.5 Potential competitive and non-competitive methanogenesis in sediment slurries from MUC** 403 **cores**

404 Results from the sediment slurry experiments, in which we added either the sulfate reduction inhibitor  
405 molybdate, the non-competitive substrate methanol, or no additives (control), are shown in Fig. 4.  
406 During the first phase of incubation, all three treatments showed rates within the same order of  
407 magnitude. Nevertheless, potential methanogenesis rates were significantly higher ( $p < 0.05$ ) in all  
408 treatments in the shallow sediment horizon (0-5 cmbsf) compared to the deep horizon (20-25 cmbsf).

409 In addition, potential methanogenesis was always significantly higher in the molybdate and methanol  
410 treatment compared to the control.

411 During the second phase of the incubation (day 8-23), potential methanogenesis showed a different  
412 pattern. Rates in the methanol treatment were 350 and 4 times higher compared to the control and  
413 molybdate treatment in the 0-5 cm horizon and the 20-25 cm horizon, respectively ( $p < 0.05$ ). Control  
414 and molybdate treatments showed no significant difference ( $p > 0.05$ ) in the shallow and deep horizon.

415

### 416 **3.6 Geochemical parameters and gross hydrogenotrophic methanogenesis activity in gravity** 417 **cores**

418 At the shallow St. 1 (78 m), POC concentration slightly decreased with depth, from ~4 wt % at the  
419 surface to about 2-3 wt % at the bottom of the core (385 cmbsf, Fig. 5). At St. 8 (407 m), POC  
420 concentrations were slightly higher with values ranging around 8-9 wt % in the upper 120 cmbsf, and  
421 then decreasing with depth. The C/N ratio at St. 1 (78 m) remained around 10 throughout the core,  
422 while it showed slightly higher values around 12 throughout the core at St. 8 (407 m).

423 At St. 1 (78 m), the methane concentration increased with depth from 0.1 mM at the surface to the  
424 highest measured concentration at 165 cmbsf (~5 mM), followed by a decrease to ~ 2 mM at 198  
425 cmbsf. Methane concentration stayed around 2 mM until the deepest measured depth (385 cmbsf).

426 Methane concentrations at St. 8 (407 m) ranged from 14 to 17  $\mu\text{M}$  in the upper 120 cmbsf, then  
427 increased to a maximum of 36  $\mu\text{M}$  at 180 cmbsf, followed by a decrease to 28  $\mu\text{M}$  at the deepest  
428 sampled depth (195 cmbsf).

429  $\text{SO}_4^{2-}$  concentration at St. 1 (78 m) decreased with depth with the highest concentration (10 mM) at the  
430 shallowest measured sediment depth (33 cmbsf) and the lowest concentration at 350 cmbsf (0.16 mM).

431 At St. 8 (407 m),  $\text{SO}_4^{2-}$  concentration decreased slightly from ~28 mM at the shallowest measured  
432 sediment depth (20 cmbsf) to ~24 mM at 145 cmbsf, followed by stable concentrations around 25 mM  
433 until the bottom of the core.

434 DIC concentrations were 5-8 times higher at St. 1 (78 m) compared to St. 8 (407 m) and increased  
435 with sediment depth from ~21 mM at 33 cmbsf to ~39 mM at 385 cmbsf. DIC concentrations at St. 8



436 (407 m) could only be measured at distinct sediment depths due to limited amounts of porewater but  
437 still revealed a slight increase with sediment depth (from ~3 mM to ~5 mM).  
438 Hydrogenotrophic methanogenesis at St. 1 (78 m) was present but low below 66 cmbsf until it reached  
439 a peak between 300 and 400 cmbsf ( $0.7 \text{ nmol cm}^{-3} \text{ d}^{-1}$ ). In contrast, no hydrogenotrophic  
440 methanogenesis activity was detected at St. 8 (407 m).

441

## 442 **4. Discussion**

### 443 **4.1 Concurrent activity of methanogenesis and sulfate reduction in surface sediments**

444 Before we discuss the distribution of methanogenesis in the collected sediment cores, it has to be  
445 pointed out that the top soft sediment layer (ca. 0-20 cm) of gravity cores is often disturbed or even  
446 lost during the coring procedure. Hence, surface parameters in the gravity cores should not be directly  
447 compared to the respective depth layers in MUC cores. According to this likely offset, we will use the  
448 term "surface methanogenesis/sediments" when referring to MUC cores and "deep  
449 methanogenesis/sediments" when referring to gravity cores.

450 We would further like the reader to keep in mind that we will compare two different types of rate  
451 determinations: radiotracer incubations of undisturbed sediments (deep hydrogenotrophic  
452 methanogenesis, surface sulfate reduction) and sediment slurry incubations (surface total  
453 methanogenesis). While the first method preserves the natural heterogeneity of the sediment, the latter  
454 homogenizes and dilutes sediment ingredients and organisms, which could have both negative and  
455 positive effects on the natural activity. As we are mainly interested in the vertical distribution of these  
456 processes within the sediment, these comparisons are justifiable.

457 In the present study, methanogenesis and sulfate reduction co-occurred in surface sediments along the  
458 entire depth transect (70-1024 m) on the Peruvian margin ( $12^{\circ}\text{S}$ ). Methanogenesis activity was  
459 detected in sediment layers that revealed high porewater sulfate concentrations and sulfate reduction  
460 activity (besides St. 10, where sulfate reduction was undetectable). Even though absolute sulfate  
461 reduction rates were most likely underestimated, we trust relative distribution pattern in the sediment  
462 and along the continental margin.

463 As the competition between methanogens and sulfate reducers for H<sub>2</sub> and acetate was probably never  
464 relieved, the detected surface methanogenesis was most likely based on non-competitive substrates  
465 such as methanol or methylated compounds including methylated amines or methylated sulfides  
466 (Oremland & Polcin, 1982; Oremland & Taylor, 1978; Kiene et al., 1986). Likewise, in a study off  
467 Chile (0-6 cm sediment depth, 87 m water depth), surface methanogenesis was found to be coupled to  
468 the non-competitive substrate trimethylamine, and not to CO<sub>2</sub>/H<sub>2</sub> or acetate, in sediments where sulfate  
469 and sulfate reduction was abundant (Ferdelman et al., 1997).

470 Non-competitive substrate utilization by methanogens in the present study was further confirmed by a  
471 significant increase of potential methanogenesis after the addition of methanol to sediment slurries  
472 from St. 1 (70 m) (Fig. 4 B). The delayed response of methanogenesis after methanol addition  
473 (Supplement, Fig. S1), however, suggests that the present microbial methanogenic community was not  
474 primarily feeding on methanol. Potentially, other non-competitive substrates like dimethyl sulfides  
475 were utilized predominantly. While most methylotrophic methanogens are able to use both methanol  
476 and methylated amines, growth on dimethyl sulfide appears to be restricted to only a few  
477 methylotrophic species (Oremland et al., 1989). Dimethyl sulfides could have accumulated during the  
478 long storage time (~ 6 months) before experimentation. Even though methylated sulfur compounds  
479 (e.g., dimethyl sulfide or methanthiol) are mainly produced by organisms in the marine photic zone  
480 (e.g., Andreae & Raemdonck 1983), it was recently postulated that these compounds may also be  
481 generated through nucleophilic attack by sulfide on methyl groups in the sedimentary organic matter  
482 (Mitterer, 2010). As sulfate reduction was a predominant process in the sediment, it could have  
483 delivered sufficient sulfide to produce methylated sulfur compounds. Consequently, results from the  
484 sediment slurry experiments might not reflect the activity of the in situ methanogenic community as  
485 we cannot exclude community shifts as a response to the availability of alternative substrates that were  
486 produced during the long storage.

487 The utilization of the competitive substrates H<sub>2</sub> and acetate by the methanogens probably only occurs  
488 when sulfate reducers are inhibited. Accordingly, potential methanogenesis rates in the molybdate  
489 treatment of the sediment slurry experiment were significantly higher in the two studied horizons (0-5  
490 and 20-25 cmbsf) compared to the controls during the first phase of the incubation (day 0-5),

491 indicating the usage of competitive substrate facilitated by the inhibition of sulfate reduction.  
492 However, in the second phase (day 8-23) of the incubation, rates were much lower in both the control  
493 and molybdate treatment and did not show significant differences in both horizons ( $p > 0.05$ ). In this  
494 second phase, methane production might have slowed down due to depletion of electron donors.  
495 Hydrogenotrophic methanogenesis in the gravity core from St. 1 (78 m) showed no activity at depths  
496 where porewater sulfate concentrations were  $> 0.7$  mM. Instead activity peaked where porewater  
497 sulfate was lowest (0.16 mM at 350 cmbsf), supporting the above conclusions regarding competition  
498 within the sulfate zone. The observation that sulfate was never completely depleted in the porewater  
499 until the bottom of the gravity core, in combination with an increase of iron (II) in the porewater at  
500 depths  $> 200$  cmbsf (data not shown), hint to the presence of a cryptic sulfur cycle that is responsible  
501 for deep formation of sulfate (Holmkvist et al., 2011; Treude et al., 2014) .  
502 In comparison, surface net methanogenesis activity along the Peruvian margin was similar to activities  
503 found off Chile at 87 m water depth (0-0.6  $\text{nmol cm}^{-3} \text{d}^{-1}$ ) (Ferdelman et al., 1997). The slightly higher  
504 rates determined in our study (St.1= 70 m; 0.4-1.7  $\text{nmol cm}^{-3} \text{d}^{-1}$ ) could be related to different  
505 approaches, as our rates represent the sum of net methanogenesis from all available substrates in the  
506 sediment, while rates off Chile were based only on  $\text{CO}_2$ , acetate, and trimethylamine utilization.  
507 Hence, total methanogenesis could have been easily underestimated, if methanogenesis was supplied  
508 by other substrates, which is not unlikely, as methylotrophic methanogens, which are able to use  
509 methanol or methylated amines, were the dominant type of methanogens in these sediments  
510 (Ferdelman et al., 1997). Interestingly, the authors detected a high number of acetogens, implying that  
511 acetogenesis competed for methylamines or other methylated compounds (Ferdelman et al., 1997). A  
512 competition with acetogens for methylated substrates is conceivable for our study, but would require  
513 further corroboration.

514

#### 515 **4.2 Surface vs deep methanogenesis**

516 Maximum single net surface methanogenesis activities detected in our study (0.3-4.3  $\text{nmol cm}^{-3} \text{d}^{-1}$ )  
517 were found to be at the very low end or even one order of magnitude lower than organic-rich, sulfate-  
518 depleted sediments (9.8-37  $\text{nmol cm}^{-3} \text{d}^{-1}$ , 0-40 cmbsf, Treude et al., 2005a, 10-17  $\text{nmol cm}^{-3} \text{d}^{-1}$ , 0-30

519 cmbsf, Schmaljohann 1996, 100-300 nmol cm<sup>-3</sup> d<sup>-1</sup>, 0-30 cmbsf, Crill & Martens, 1983, 1986, 100-400  
520 nmol cm<sup>-3</sup> d<sup>-1</sup>, 0-3 cmbsf, Alperin et al. 1992). To estimate the overall relevance of surface  
521 methanogenesis within the sulfate zone compared to deep methane production, we estimated the deep  
522 methane production in our study and compiled an overview of published deep methane production  
523 data from the sulfate-free zone of organic-rich sediments (Table 2). For this comparison, the deep  
524 methane production was assumed to equal the flux of methane into the sulfate-methane-transition zone  
525 (SMTZ), where it is consumed by anaerobic oxidation of methane (AOM). Within the SMTZ, both  
526 sulfate and methane are depleted steeply as a result of AOM, thus dividing the sulfate-reducing zone  
527 above from the methanogenic zone below. The SMTZ is the main niche for AOM in marine  
528 sediments, acting as an important filter for upwards migrating methane (Knittel & Boetius, 2009). The  
529 SMTZ can be found at decimeters to tens of meters below the seafloor, depending on the burial rate of  
530 reactive organic matter, the depth of the methane production zone, and the flux of methane and sulfate  
531 as well as their consumption rates (Knittel & Boetius, 2009).

532 In the present study, a SMTZ was only detected in the gravity core taken at St. 1 (78 m; Fig. 5), where  
533 it was located between 66 and 99 cmbsf, i.e., below the penetration depth of the MUC cores. We  
534 estimated a methane flux (= deep methane production) into the SMTZ (from 99 to 66 cmbsf)  
535 according to Iversen & Jørgensen, (1993) using a seawater methane-diffusion coefficient from Schulz,  
536 (2006) which was corrected for porosity resulting in a sediment-diffusion coefficient for methane of  
537  $D_s = 1.325 \times 10^{-5} \text{ cm}^2 \text{ s}^{-1}$  at 15 °C. The resulting deep methane production (0.8 mmol m<sup>-2</sup> d<sup>-1</sup>) was slightly  
538 higher (ratio of 0.13, surface vs. deep) but still in the same magnitude as the integrated surface  
539 methanogenesis at St. 1 (70 m; 0.1 mmol m<sup>-2</sup> d<sup>-1</sup>). Compared to a different study from the Peruvian  
540 OMZ, the ratio between shallow (0.07 to 0.1 mmol m<sup>-2</sup> d<sup>-1</sup>, this study) vs. deep ( $8.9 \times 10^{-8}$  to  $2.2 \times 10^{-7}$   
541 mmol m<sup>-2</sup> d<sup>-1</sup>; Arning et al., 2012) methanogenesis on the shelf (150-250 m) was  $3.2 \times 10^5$  to  $1.1 \times 10^6$ .  
542 Both examples highlight the significance of surface methanogenesis, especially on the Peruvian shelf.  
543 On the lower Peruvian slope (~3800 m water depth), deep methanogenesis increased (up to 0.017  
544 mmol m<sup>-2</sup> d<sup>-1</sup>; Arning et al., 2012). In contrast, surface methanogenesis at the deeper St. 10 (1024 m)  
545 was lower (0.02 mmol m<sup>-2</sup> d<sup>-1</sup>) compared to the shelf indicating a decreasing relevance of surface  
546 methanogenesis along the margin with increasing relevance of deep methanogenesis. The decrease of

547 surface methanogenesis with increasing water depth might be correlated to the decreasing organic  
548 carbon content and freshness in the sediment (Fig. 6), as further discussed in section 4.4.  
549 In comparison with other organic-rich sediments (Table 2), surface methanogenesis off Peru was in  
550 the same order of magnitude as most reported deep methanogenesis (e.g., off Namibia, off Chile,  
551 Limfjorden). The only exception was Eckernförde Bay (Baltic Sea), where surface methanogenesis off  
552 Peru was less than 15% of deep methanogenesis. Eckernförde Bay has a water depth of only ~30 m  
553 with high carbon export, featuring extremely high methanogenesis activity below the SMTZ, causing  
554 supersaturation and methane gas ebullition (Whiticar, 2002; Treude et al., 2005a).

### 555 **4.3 Potential consumption and emission of surface methane**

556 Due to its closeness to the sediment-water interface, surface methanogenesis along the Peruvian  
557 margin could lead to methane emissions from the sediment into the water column. A short diffusion  
558 distance, especially in the top most sediment layers, might facilitate a partial escape of methane from  
559 consumption by microbes. As surface methanogenesis decreased with water depth (Fig. 3), the  
560 methane emission potential appears to be highest on the shelf. Sediment methane concentrations in the  
561 0-2 sediment horizon of all sites along the margin were always higher than bottom-near water methane  
562 concentrations (~1.5 m above seafloor; Table 1, Fig. 2), hinting towards an efflux of methane from the  
563 sediment. However, more precise profiling of methane at the sediment-water interface would be  
564 necessary to confirm this hypothesis. Still, most of the sediment methane profiles suggest methane  
565 consumption close to the seafloor to some extent, which would reduce the amount of emitted methane  
566 (Fig. 2). AOM might act as an important methane filter at the sediment surface of the shelf stations,  
567 where anoxic conditions dominated, while aerobic oxidation might prevail at the deeper stations below  
568 the OMZ (St. 9 and 10). The presence of methane oxidation above the SMTZ of organic-rich  
569 sediments has been reported earlier (Treude et al., 2005a, 2005b), and could indeed be fueled by  
570 surface methanogenesis. An immediate oxidation of the produced methane would explain why  
571 sediment methane profiles did not necessarily correlate with peaks in surface methanogenesis (see,  
572 e.g., St 6, 253 m). The importance of AOM for the reduction of methane emissions from surface  
573 methanogenesis remains speculative, as explicit data is missing. On the basis of our findings, however,

574 we suggest to consider surface methanogenesis as a possible driver for AOM above the SMTZ in  
575 earlier and future studies.

576

#### 577 **4.4 Factors controlling methanogenesis along the Peruvian margin**

578 For this discussion, we excluded the high integrated methane production observed in one of the  
579 replicates at station 10 (1024 m) as we do not think that the detected activity ( $0.23 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) is  
580 representative for this deep site, especially as sediment POC content was lowest at station 10  
581 compared to the other stations ( $<4\%$ , Fig. 2). The outlier might have been caused by additional carbon  
582 sources in the sediment, e.g., from fecal pellets or organic carbon released from dead infauna, thus  
583 stimulating below-surface microbial activities during our incubations (Ziervogel et al., 2014; Bertics et  
584 al., 2013).

585

##### 586 **4.4.1 Oxygen**

587 Oxygen is an important controlling factor, as methanogenesis is an oxygen- and redox-sensitive  
588 process (Oremland, 1988). Some methanogens can tolerate oxygen exposure for several hours before  
589 they die, although no methane is produced in the presence of oxygen (Zinder, 1993).

590 Comparing integrated surface methanogenesis (over 0-25 cmbsf) from the shallowest to the deepest  
591 station (Fig. 3), highest rates ( $> 0.05 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) were detected on the shelf (St. 1, 4 and 6=70, 145,  
592 253 m), where oxygen concentrations were below detection (Fig.6), providing advantageous  
593 conditions for methanogenesis, particularly at the very sediment surface, where normally aerobic  
594 respiration dominates (Jørgensen, 2006). Below the OMZ, integrated methanogenesis decreased.  
595 Bioturbating macrofauna and megafauna (e.g., mussels, polychaetes, oligochaetes) were observed at  
596 these sites (St. 9 and 10, 770 and 1024 m) (Mosch et al. 2012), which could have transported oxygen  
597 into deeper sediment layer (Orsi et al., 1996), thus leading to less reduced conditions ( $> -200 \text{ mV}$ )  
598 unsuitable for methanogens (Oremland, 1988). However, integrated methanogenesis was lowest at St.  
599 8 (407 m), which still revealed anoxic bottom water. Thus, oxygen might just be advantageous but not  
600 the driving factor for surface methanogenesis.

601

#### 602 **4.4.2 Organic matter**

603 The most important factor controlling benthic methanogenesis activity is probably the POC content of  
604 the sediment, as it determines the substrate availability and variety, and can thus relieve the  
605 competitive situation between methanogens and sulfate reducers (Holmer & Kristensen, 1994; Treude  
606 et al., 2009). Consequently, high methanogenesis rates may be expected along the Peruvian margin at  
607 sites with high organic carbon load. However, integrated methanogenesis rates did not correlate with  
608 sediment POC content (Fig. 6). While POC content was increasing from St. 1 (70 m) to St. 6 (253 m),  
609 followed by a decrease until St. 10 (1024 m), integrated methanogenesis showed rather a decreasing  
610 trend with increasing water depth. This deviation from expectations might be caused by another factor,  
611 such as organic matter quality, i.e., freshness. Numerous studies have shown that the quality of the  
612 organic matter is important for the rate and magnitude of microbial organic matter degradation  
613 (Westrich & Berner, 1984; Canfield, 1994; Amon et al., 2001; Middelburg, 1989).  
614 Integrated methanogenesis and C/N ratios (indicating the freshness of organic matter) were negatively  
615 correlated along the Peruvian margin (Fig. 6), suggesting that fresh, labile organic matter is  
616 advantageous for surface methanogenesis. As methanogens consume mostly short, monomeric  
617 substrates, they depend on other microbial groups to break down large organic macromolecules  
618 (Zinder, 1993). Hence, labile organic matter offers an important supply of methanogenic substrates.  
619 In agreement with this hypothesis, highest integrated methanogenesis rates were observed at St. 1 (70  
620 m), which revealed the freshest organic matter (lowest C/N, Fig. 6) and the highest POC  
621 remineralization rates along the Peruvian margin (Dale et al., 2015). The degradation of organic matter  
622 within the water column was probably limited at St. 1 (70 m) due to anoxic conditions and high  
623 sedimentation rates (Dale et al., 2015); hence, labile organic matter accumulated at the seafloor,  
624 thereby increasing the benthic POC degradation and resulting in high substrate availability and variety  
625 for the methanogenic community.

626 Nevertheless, lowest methanogenesis rates were measured at St. 8 (407 m), which was neither the site  
627 of the highest C/N ratio, lowest POC content (Fig. 6), or the lowest POC mineralization (Dale et al.,  
628 2015). In this particular case, methanogenesis was most likely controlled by the sediment properties.  
629 Methanogenesis activity was undetectable below 10 cmbsf, which coincided with a very dense and

630 sticky clay layer. The POC profile at St. 8 (407 m) revealed lower concentrations in the upper 5 cmbsf,  
631 followed by an increase with depth, suggesting that either the organic matter at this station was  
632 resistant to microbial attack (indicated by the increase in C/N) or that microbes were not as  
633 frequent/active in the dense clay layer as at the surface. Similarly, sulfate reduction and microbial  
634 nitrogen fixation (Gier et al., 2015) showed very low activity at this site (Fig. 2).

635

## 636 **5. Conclusion**

637 The present study demonstrated that methanogenesis coincides with sulfate reduction in surface  
638 sediments (< 30 cmbsf) along the Peruvian margin. The competition with sulfate reducers was  
639 partially relieved due to the high load of organic carbon allowing both groups to show concurrent  
640 activity through the utilization of non-competitive substrates by the methanogens.

641 The significance of surface methanogenesis was high on the shelf, where ratios between surface and  
642 deep methanogenesis were around 0.13 (this study) or even as high as  $\sim 10^5$  (compared to Arning et al.  
643 2012), and decreased with increasing water depth. Accordingly, we assume that potential methane  
644 emissions into the water column, indicated by a higher methane concentration at the sediment surface  
645 compared to the bottom water, should be highest on the shelf, where surface methane production rates  
646 were highest. Our results further hint towards a partial consumption of methane before reaching the  
647 sediment-water interface, probably by anaerobic oxidation of methane (AOM). In this case, surface  
648 methanogenesis might act as important supplier of methane for AOM above the SMTZ, which has  
649 been largely overlooked previously.

650 We postulate that the dominant factor controlling surface methanogenesis is the availability of  
651 (primarily labile) organic matter. The high load of organic carbon and resulting high organic carbon  
652 mineralization rates secure the supply of methanogenic substrates, especially on the shelf, which  
653 mitigates the competition between sulfate reducers and methanogens. Anoxic conditions in the  
654 overlying water might be advantageous for the oxygen-sensitive process of methanogenesis, but does  
655 not appear to primarily control benthic rates, as they change within the anoxic zones.



656 Interestingly, organic matter made available by bioturbating infauna (e.g., fecal pellets or dead  
657 organisms) could be an important additional factor facilitating methanogenesis in surface sediments.  
658 As shown in this study, methanogenesis rates vary strongly in bioturbated sediments below the OMZ,  
659 sometimes exceeding all other observed methanogenic rates.  
660 Future studies should seek to (1) identify methanogens and their metabolic capabilities in surface  
661 sediments, (2) determine the direct interaction between surface methanogenesis and AOM, and (3)  
662 evaluate the effect of organic matter hot spots on total benthic surface methanogenesis in organic-rich  
663 sediments.

664

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675

#### 676 **Author contribution**

677 J.M. and T.T. designed the experiments. J.M. carried out all methanogenesis experiments, T.T.  
678 conducted sulfate reduction measurements. Porewater measurements of MUC cores were coordinated  
679 by A.D. and S.S. J.M. prepared the manuscript with contributions of all co-authors.

680

681 **References**

- 682 Alperin, M.J., Blair, N.E., Albert, D.B., Hoehler, T.M. & Martens, C.S. (1992). Factors that control  
683 the stable isotopic composition of methane produced in an anoxic marine sediment. *Global*  
684 *Biogeochemical Cycles*. 6 (3). pp. 271–291.
- 685 Amon, R.M.W., Fitznar, H.-P. & Benner, R. (2001). Linkages among the bioreactivity, chemical  
686 composition, and diagenetic state of marine dissolved organic matter. *Limnology and*  
687 *Oceanography*. 46 (2). pp. 287–297.
- 688 Andreae, M.O. & Raemdonck, H. (1983). Dimethyl sulfide in the surface ocean and the marine  
689 atmosphere: a global view. *Science (New York, N.Y.)*. 221 (4612). pp. 744–747.
- 690 Arning, E.T., Van Berk, W. & Schulz, H.M. (2012). Quantitative geochemical modeling along a  
691 transect off Peru: Carbon cycling in time and space, and the triggering factors for carbon loss and  
692 storage. *Global Biogeochemical Cycles*. 26 (4). pp. 1–18.
- 693 Bange, H.W., Bartell, U.H., Rapsomanikis, S. & Andreae, M.O. (1994). Methane in the Baltic and  
694 North Seas and a reassessment of the marine emissions of methane. *Global Biogeochemical*  
695 *Cycles*. 8 (4). pp. 465–480.
- 696 Bange, H.W., Hansen, H.P., Malien, F., Laß, K., Karstensen, J., Petereit, C., Friedrichs, G. & Dale, A.  
697 (2011). Boknis Eck Time Series Station ( SW Baltic Sea ): Measurements from 1957 to 2010.  
698 *LOICZ-Affiliated Activities*. Inprint 20. pp. 16–22.
- 699 Bertics, V.J., Löscher, C.R., Salonen, I., Dale, A.W., Gier, J., Schmitz, R.A. & Treude, T. (2013).  
700 Occurrence of benthic microbial nitrogen fixation coupled to sulfate reduction in the seasonally  
701 hypoxic Eckernförde Bay, Baltic Sea. *Biogeosciences*. 10 (3). pp. 1243–1258.
- 702 Buckley, D.H., Baumgartner, L.K. & Visscher, P.T. (2008). Vertical distribution of methane  
703 metabolism in microbial mats of the Great Sippewissett Salt Marsh. *Environmental*  
704 *microbiology*. 10 (4). pp. 967–77.
- 705 Canfield, D.E. (1994). Factors influencing organic carbon preservation in marine sediments. *Chemical*  
706 *geology*. 114 (93). pp. 315–329.
- 707 Crill, P. & Martens, C. (1983). Spatial and temporal fluctuations of methane production in anoxic  
708 coastal marine sediments. *Limnology and Oceanography*. 28. pp. 1117–1130.
- 709 Crill, P.M. & Martens, C.S. (1986). Methane production from bicarbonate and acetate in an anoxic  
710 marine sediment. *Geochimica et Cosmochimica Acta*. 50. pp. 2089–2097.
- 711 Dale, A.W., Sommer, S., Lomnitz, U., Montes, I., Treude, T., Liebetrau, V., Gier, J., Hensen, C.,  
712 Dengler, M., Stolpovsky, K., Bryant, L.D. & Wallmann, K. (2015). Organic carbon production,  
713 mineralisation and preservation on the Peruvian margin. *Biogeosciences*. 12. pp. 1537–1559.
- 714 Dimitrov, L. (2002). Mud volcanoes – the most important pathways for degassing deeply buried  
715 sediments. *Earth Science Review*. 59. pp. 49–76.
- 716 Donnelly, M.I. & Dagley, S. (1980). Production of Methanol from Aromatic Acids by *Pseudomonas*  
717 *putida*. *Journal of bacteriology*. 142 (3). pp. 916–924.

- 718 Ferdelman, T.G., Lee, C., Pantoja, S., Harder, J., Bebout, B.M. & Fossing, H. (1997). Sulfate  
719 reduction and methanogenesis in a Thioploca-dominated sediment off the coast of Chile.  
720 *Geochimica et Cosmochimica Acta*. 61 (15). pp. 3065–3079.
- 721 Fuenzalida, R., Schneider, W., Garcés-Vargas, J., Bravo, L. & Lange, C. (2009). Vertical and  
722 horizontal extension of the oxygen minimum zone in the eastern South Pacific Ocean. *Deep-Sea*  
723 *Research Part II: Topical Studies in Oceanography*. 56. pp. 992–1003.
- 724 Gallardo, V.A. (1977). Large benthic microbial communities in sulphide biota under Peru-Chile  
725 subsurface countercurrent. *Nature*. 268. pp. 331–332.
- 726 Gier, J., Sommer, S., Löscher, C.R., Dale, A.W., Schmitz, R.A. & Treude, T. (2015). Nitrogen fixation  
727 in sediments along a depth transect through the Peruvian oxygen minimum zone. *Biogeosciences*  
728 *Discussions*. 12 (17). pp. 14401–14440.
- 729 Hines, M.E. & Buck, J.D. (1982). Distribution of methanogenic and sulfate-reducing bacteria in near-  
730 shore marine sediments. *Applied and environmental microbiology*. 43 (2). pp. 447–453.
- 731 Holmer, M. & Kristensen, E. (1994). Coexistence of sulfate reduction and methane production in an  
732 organic-rich sediment. *Marine Ecology Progress Series*. 107. pp. 177–184.
- 733 Holmkvist, L., Ferdelman, T.G. & Jørgensen, B.B. (2011). A cryptic sulfur cycle driven by iron in the  
734 methane zone of marine sediment (Aarhus Bay, Denmark). *Geochimica et Cosmochimica Acta*.  
735 75 (12). pp. 3581–3599.
- 736 IPCC (2014). *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to*  
737 *the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. T. core writing  
738 Team, R. K. Pachauri, & L. A. Meyer (eds.). Geneva, Switzerland.
- 739 Iversen, N. & Jørgensen, B.B. (1993). Diffusion coefficients of sulfate and methane in marine  
740 sediments: Influence of porosity. *Geochimica et Cosmochimica Acta*. 57 (3). pp. 571–578.
- 741 Jørgensen, B.B. (1978). A comparison of methods for the quantification of bacterial sulfate reduction  
742 in coastal marine sediments: I. Measurements with radiotracer techniques. *Geomicrobiology*  
743 *Journal*. 1. pp. 11–27.
- 744 Jørgensen, B.B. (2006). Bacteria and marine Biogeochemistry. In: H. D. Schulz & M. Zabel (eds.).  
745 *Marine Geochemistry*. Berlin/Heidelberg: Springer-Verlag, pp. 173–207.
- 746 Jørgensen, B.B. (1977). The sulfur cycle of a coastal marine sediment (Limfjorden, Denmark).  
747 *Limnology and Oceanography*. 22 (5). pp. 814–832.
- 748 Jørgensen, B.B. & Parkes, R.J. (2010). Role of sulfate reduction and methane production by organic  
749 carbon degradation in eutrophic fjord sediments (Limfjorden, Denmark). *Limnology and*  
750 *Oceanography*. 55 (3). pp. 1338–1352.
- 751 Judd, A., Davies, G., Wilson, J., Holmes, R., Baron, G. & Bryden, I. (1997). Contributions to  
752 atmospheric methane by natural seepages on the UK continental shelf. *Marine Geology*. 137 (1-  
753 2). pp. 165–189.
- 754 Kallmeyer, J., Ferdelman, T.G., Weber, A., Fossing, H. & Jørgensen, B.B. (2004). Evaluation of a  
755 cold chromium distillation procedure for recovering very small amounts of radiolabeled sulfide  
756 related to sulfate reduction measurements. *Limnology and Oceanography: Methods*. 2. pp. 171–  
757 180.

- 758 Kamykowski, D. & Zentara, S. (1990). Hypoxia in the world ocean as recorded in the historical data  
759 set. *Deep-Sea Research*. 37 (12). pp. 1861–1874.
- 760 Kiene, R.P., Oremland, R.S., Catena, a, Miller, L.G. & Capone, D.G. (1986). Metabolism of reduced  
761 methylated sulfur compounds in anaerobic sediments and by a pure culture of an estuarine  
762 methanogen. *Applied and environmental microbiology*. 52 (5). pp. 1037–1045.
- 763 King, G.M., Klug, M.J. & Lovley, D.R. (1983). Metabolism of acetate, methanol, and methylated  
764 amines in intertidal sediments of lowes cove, maine. *Applied and environmental microbiology*.  
765 45 (6). pp. 1848–1853.
- 766 Knittel, K. & Boetius, A. (2009). Anaerobic oxidation of methane: progress with an unknown process.  
767 *Annual review of microbiology*. 63. pp. 311–34.
- 768 Kristensen, E. (2000). Organic matter diagenesis at the oxic/anoxic interface in coastal marine  
769 sediments, with emphasis on the role of burrowing animals. *Hydrobiologia*. 426 (1). pp. 1–24.
- 770 Van Der Maarel, M.J.E.C. & Hansen, T. a. (1997). Dimethylsulfoniopropionate in anoxic intertidal  
771 sediments: A precursor of methanogenesis via dimethyl sulfide, methanethiol, and  
772 methiolpropionate. *Marine Geology*. 137 (1-2). pp. 5–12.
- 773 Middelburg, J.J. (1989). A simple rate model for organic matter decomposition in marine sediments.  
774 *Geochimica et Cosmochimica Acta*. 53 (7). pp. 1577–1581.
- 775 Mitterer, R.M. (2010). Methanogenesis and sulfate reduction in marine sediments: A new model.  
776 *Earth and Planetary Science Letters*. 295 (3-4). pp. 358–366.
- 777 Neill, a R., Grime, D.W. & Dawson, R.M. (1978). Conversion of choline methyl groups through  
778 trimethylamine into methane in the rumen. *The Biochemical journal*. 170 (3). pp. 529–535.
- 779 Niewöhner, C., Hensen, C., Kasten, S., Zabel, M. & Schulz, H.D. (1998). Deep sulfate reduction  
780 completely mediated by anaerobic oxidation in sediments of the upwelling area off Namibia.  
781 *Geochimica et Cosmochimica Acta*. 62 (3). pp. 455–464.
- 782 Oremland, R.S. (1988). Biogeochemistry of methanogenic bacteria. In: A. J. B. Zehnder (ed.). *Biology*  
783 *of Anaerobic Microorganisms*. New York: J. Wiley & Sons, pp. 641–705.
- 784 Oremland, R.S. & Capone, D.G. (1988). Use of specific inhibitors in biogeochemistry and microbial  
785 ecology. In: K. C. Marshall (ed.). *Advances in Microbial Ecology*. Advances in Microbial  
786 Ecology. Boston, MA: Springer US, pp. 285–383.
- 787 Oremland, R.S., Kiene, R.P., Mathrani, I., Whiticar, M.J. & Boone, D.R. (1989). Description of an  
788 estuarine methylotrophic methanogen which grows on dimethyl sulfide. *Applied and*  
789 *environmental microbiology*. 55 (4). pp. 994–1002.
- 790 Oremland, R.S., Marsh, L. & Desmarais, D.J. (1982). Methanogenesis in Big Soda Lake , Nevada : an  
791 Alkaline , Moderately Hypersaline Desert Lake. *Applied and environmental microbiology*. 43  
792 (2). pp. 462–468.
- 793 Oremland, R.S. & Polcin, S. (1982). Methanogenesis and Sulfate Reduction : Competitive and  
794 Noncompetitive Substrates in Estuarine Sediments. *Applied and Environmental Microbiology*. 44  
795 (6). pp. 1270–1276.

- 796 Oremland, R.S. & Taylor, B.F. (1978). Sulfate reduction and methanogenesis in marine sediments.  
797 *Geochimica Cosmochimica Acta*. 42. pp. 209–214.
- 798 Orsi, T.H., Werner, F., Milkert, D., Anderson, a. L. & Bryant, W.R. (1996). Environmental overview  
799 of Eckernförde Bay, northern Germany. *Geo-Marine Letters*. 16 (3). pp. 140–147.
- 800 Pennington, J.T., Mahoney, K.L., Kuwahara, V.S., Kolber, D.D., Calienes, R. & Chavez, F.P. (2006).  
801 Primary production in the eastern tropical Pacific: A review. *Progress in Oceanography*. 69 (2-  
802 4). pp. 285–317.
- 803 Reeburgh, W. (2007). Oceanic methane biogeochemistry. *Chemical Reviews*. 107. pp. 486–513.
- 804 Reimers, C.E. & Suess, E. (1983). The partitioning of organic carbon fluxes and sedimentary organic  
805 matter decomposition rates in the ocean. *Marine Chemistry*. 13. pp. 141–168.
- 806 Røy, H., Weber, H.S., Tarpgaard, I.H., Ferdelman, T.G. & Jørgensen, B.B. (2014). Determination of  
807 dissimilatory sulfate reduction rates in marine sediment via radioactive <sup>35</sup>S tracer. *Limnology  
808 and Oceanography: Methods*. 12. pp. 196–211.
- 809 Schink, B. & Zeikus, J.G. (1982). Microbial Ecology of Pectin Decomposition in Anoxic Lake  
810 Sediments and in defined laboratory cultures of species prevalent in the lake sediment . The  
811 turnover. *Journal of General Microbiology*. 128 (393-404). pp. 393–404.
- 812 Schmaljohann, R. (1996). Methane dynamics in the sediment and water column of Kiel Harbour  
813 (Baltic Sea). *Marine Chemistry*. 131. pp. 263–273.
- 814 Schulz, H.D. (2006). Quantification of early diagenesis: dissolved constituents in marine pore water.  
815 In: H. D. Schulz & M. Zabel (eds.). *Marine Geochemistry*. Berlin, Heidelberg: Springer Berlin  
816 Heidelberg, pp. 75–124.
- 817 Seeberg-Elverfeldt, J., Schluter, M., Feseker, T. & Kolling, M. (2005). Rhizon sampling of porewaters  
818 near the sediment-water interface of aquatic systems. *Limnology and Oceanography-Methods*. 3.  
819 pp. 361–371.
- 820 Senior, E., Lindström, E.B., Banat, I.M. & Nedwell, D.B. (1982). Sulfate reduction and  
821 methanogenesis in the sediment of a saltmarsh on the East coast of the United kingdom. *Applied  
822 and environmental microbiology*. 43 (5). pp. 987–996.
- 823 Smetacek, V. (1985). The Annual Cycle of Kiel Bight Plankton: A Long-Term Analysis. *Estuaries*. 8  
824 (June). pp. 145–157.
- 825 Thauer, R.K. (1998). Biochemistry of methanogenesis : a tribute to Marjory Stephenson.  
826 *Microbiology*. 144. pp. 2377–2406.
- 827 Treude, T., Krause, S., Maltby, J., Dale, A.W., Coffin, R. & Hamdan, L.J. (2014). Sulfate reduction  
828 and methane oxidation activity below the sulfate-methane transition zone in Alaskan Beaufort  
829 Sea continental margin sediments: Implications for deep sulfur cycling. *Geochimica et  
830 Cosmochimica Acta*. 144. pp. 217–237.
- 831 Treude, T., Krüger, M., Boetius, A. & Jørgensen, B.B. (2005a). Environmental control on anaerobic  
832 oxidation of methane in the gassy sediments of Eckernförde Bay ( German Baltic ). *Limnology  
833 and Oceanography*. 50 (6). pp. 1771–1786.

- 834 Treude, T., Niggemann, J., Kallmeyer, J., Wintersteller, P., Schubert, C.J., Boetius, A. & Jørgensen,  
835 B.B. (2005b). Anaerobic oxidation of methane and sulfate reduction along the Chilean  
836 continental margin. *Geochimica et Cosmochimica Acta*. 69 (11). pp. 2767–2779.
- 837 Treude, T., Smith, C.R., Wenzhöfer, F., Carney, E., Bernardino, A.F., Hannides, A.K., Krgüer, M. &  
838 Boetius, A. (2009). Biogeochemistry of a deep-sea whale fall: Sulfate reduction, sulfide efflux  
839 and methanogenesis. *Marine Ecology Progress Series*. 382. pp. 1–21.
- 840 Valentine, D.L., Blanton, D.C., Reeburgh, W.S. & Kastner, M. (2001). Water column methane  
841 oxidation adjacent to an area of active hydrate dissociation, Eel River Basin. *Geochimica et*  
842 *Cosmochimica Acta*. 65 (16). pp. 2633–2640.
- 843 Westrich, J.T. & Berner, R. a. (1984). The role of sedimentary organic matter in bacterial sulfate  
844 reduction: The G model tested. *Limnology and Oceanography*. 29 (2). pp. 236–249.
- 845 Wever, T.F. & Fiedler, H.M. (1995). Variability of acoustic turbidity in Eckernförde Bay ( southwest  
846 Baltic Sea ) related to the annual temperature cycle. *Marine Geology*. 125. pp. 21–27.
- 847 Whiticar, M.J. (2002). Diagenetic relationships of methanogenesis, nutrients, acoustic turbidity,  
848 pockmarks and freshwater seepages in Eckernförde Bay. *Marine Geology*. 182. pp. 29–53.
- 849 Whiticar, M.J. (1978). *Relationships of interstitial gases and fluids during early diagenesis in some*  
850 *marine sediments*.
- 851 Widdel, F. & Bak, F. (1992). Gram-Negative Mesophilic Sulfate-Reducing Bacteria. In: A. Balows, H.  
852 G. Trüper, M. Dworkin, W. Harder, & K.-H. Schleifer (eds.). *The Prokaryotes*. New York, NY:  
853 Springer New York, pp. 3352–3378.
- 854 Ziervogel, K., Joye, S.B. & Arnosti, C. (2014). Microbial enzymatic activity and secondary production  
855 in sediments affected by the sedimentation pulse following the Deepwater Horizon oil spill.  
856 *Deep-Sea Research Part II: Topical Studies in Oceanography*. pp. 1–8.
- 857 Zinder, S.H. (1993). Physiological ecology of methanogens. In: J. G. Ferry (ed.). *Methanogenesis*.  
858 New York, NY: Chapman & Hall, pp. 128–206.
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866 **Figure Captions**

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868 **Figure 1:** Location of sampling sites off Peru along the depth transect at 12° S. Source: Schlitzer, R.,  
869 Ocean Data View, <http://odv.awi.de>, 2014

870 **Figure 2:** Profiles of particulate organic carbon (POC), C/N ratio, methane (CH<sub>4</sub>), sulfate (SO<sub>4</sub><sup>2-</sup>), DIC  
871 (dissolved inorganic carbon), net methanogenesis (MG) rates and sulfate reduction (SR) rates in the  
872 MUC cores along the depth transect. For MG, triplicates (symbols) and mean (solid line) are shown.  
873 For SR, duplicates are shown. Data points from the overlaying water in the MUC core (OLW) are set  
874 to 0 cm. Note deviant scale dimension for MG at St. 6 and for SR at St. 1 and 2.

875

876 **Figure 3:** Integrated methanogenesis and sulfate reduction rates (0-25 cm) along the depth transect.  
877 For methanogenesis rates (black bars), average values are shown with standard deviation. Note for St.  
878 10 a mean from two replicates is shown without standard deviation (pattern-filled bar) and the outlier  
879 is shown separately (cross). For sulfate reduction rates (blue bars), means from two replicates are  
880 shown without standard deviation.

881 **Figure 4:** Potential methanogenesis rates in sediment slurry experiments from the two sediment  
882 intervals (0-5 cm and 20-25 cm) at St. 1 (70 m). The first phase of the incubation shows rates  
883 calculated from day 0 to 5 (A), while the second phase of the incubation summarizes the rates from  
884 day 8-23 (B). "Control" is the treatment with sulfate-rich (28 mM) artificial seawater medium, "plus  
885 Mb" is the treatment with sulfate-rich artificial seawater medium plus molybdate (Mb, 22mM), and  
886 "plus Meth" is defined as the treatment with sulfate-rich artificial seawater medium plus methanol  
887 (Meth, 10 mM). Per treatment, average values are shown with standard deviation. Please note the  
888 split-up in the diagram in part B and the different x-axis for methanogenesis

889

890 **Figure 5:** Profiles of particulate organic carbon (POC), C/N ratio, methane (CH<sub>4</sub>), sulfate (SO<sub>4</sub><sup>2-</sup>),  
891 dissolved inorganic carbon (DIC), and hydrogenotrophic methanogenesis (MG) rates in the gravity  
892 cores at two stations within the depth transect. For MG, triplicates (symbols) and mean (solid line) are  
893 shown.

894

895 **Figure 6:** Bottom-near water methane (CH<sub>4</sub>) and oxygen (O<sub>2</sub>) concentrations along the depth transect  
896 (above). Surface sediment particulate organic carbon (POC) content and C/N ratio together with  
897 integrated methanogenesis (MG) rates (0-25 cmbsf) along the depth transect (below). For MG rates,  
898 averages are shown with standard deviation beside St. 10, where a mean from two replicates is shown  
899 (see text). Please note the secondary y-axis.

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902

903 **Tables**

904 Table 1: Stations, instruments, chemical/physical parameters in the bottom-near water, and analyses  
 905 applied to samples along the depth transect on the Peruvian margin (12°S). For abbreviations see  
 906 footnote.

Station No	Instrumen t	Latitude (S)	Longitude (W)	Water depth (m)	O <sub>2</sub> (μM)	Temp. (°C)	CH <sub>4</sub> (nM)	Type of analysis
1	MUC 13	12°13.492	77°10.511	70				All
	MUC 38	12°13.517	77°10.084	70				SE
	GC 8	12°14.500	77°9.611	78				GC-All
	CTD 9	12°13.535	77°10.522	73	bdl	14	38.6	WC
4	MUC 10	12°18.704	77°17.790	145				All
	CTD 14	12°18.697	77°18.004	145	bdl	13.4	24.4	WC
6	MUC 5	12°23.321	77°24.176	253				Gas+PW
	MUC 6	12°23.322	77°24.181	253				nMG
	CTD 6	12°24.904	77°26.314	305	bdl	12	79.6	WC
8	MUC 23	12°27.198	77°29.497	407				Gas+ PW
	MUC 24	12°27.197	77°29.497	407				nMG
	GC 3	12°27.192	77°29.491	407				GC-All
	CTD 37	12°29.502	77°29.502	407	bdl	10.6	7.3	WC
9	MUC 17	12°31.374	77°35.183	770				Gas+ PW
	MUC 18	12°31.373	77°35.184	770				nMG
	CTD 27	12°31.327	77°35.265	770	19	5.5	8.4	WC
10	MUC 28	12°35.377	77°40.975	1024				Gas+ PW
	MUC 29	12°35.377	77°40.976	1024				nMG
	CTD 11	12°34.863	77°38.954	1010	53	4.4	3.9	WC

907 MUC = multicorer, GC= gravity corer, CTD = CTD/Rosette, bdl= below detection limit (5μM), All = methane  
 908 gas analysis, porewater analysis, net methanogenesis analysis, SE = slurry experiment, GC-All= analysis for  
 909 gravity cores including methane gas anaylsis, porewater analysis, hydrogenotrophic methanogenesis analysis,  
 910 WC= Water column analyses, Gas = methane gas analysis, PW= porewater analysis, nMG= net methanogenesis  
 911 analysis,  
 912

**Table 2:** Comparison of deep methanogenesis in organic-rich sediments from different regions with surface methanogenesis ( $0.02-0.1 \text{ mmol m}^{-2} \text{ d}^{-1}$ ) determined in the present study. The ratio range was achieved by dividing the lowest surface by the highest deep and the highest surface by the lowest deep methanogenic activity, respectively.

	<b>Water Depth (m)</b>	<b>Depth of SMTZ (mbsf)</b>	<b>Methane flux into the SMTZ = integrated deep methanogenesis (<math>\text{mmol m}^{-2} \text{ d}^{-1}</math>)</b>	<b>Ratio between surface methanogenesis (present study) and deep methanogenesis</b>	<b>Reference</b>
<b>Namibia (SE Atlantic)</b>	1312-2060	3-10	0.07-0.15	0.13-1.43	Niewöhner et al., (1998)
<b>Eckernförde Bay (SW Baltic Sea)</b>	25-28	0.5-1.5	0.66-1.88	0.01-0.15	Treude et al., (2005a)
<b>Chile (SE Pacific)</b>	797-2746	3-4	0.068-0.13	0.15-1.47	Treude et al., (2005b)
<b>Limfjorden (North Sea)</b>	7-10	1-1.5	0.076	0.03-1.32	Jørgensen & Parkes, (2010)
<b>Peru (SE Pacific)</b>	150-3819	2-50	$2.2 \times 10^{-7}$ -0.017	$1.18-4.55 \times 10^5$	Arning et al., (2012)
<b>Peru (SE Pacific)</b>	70-1024	0.7-1	0.8	0.03-0.13	present study

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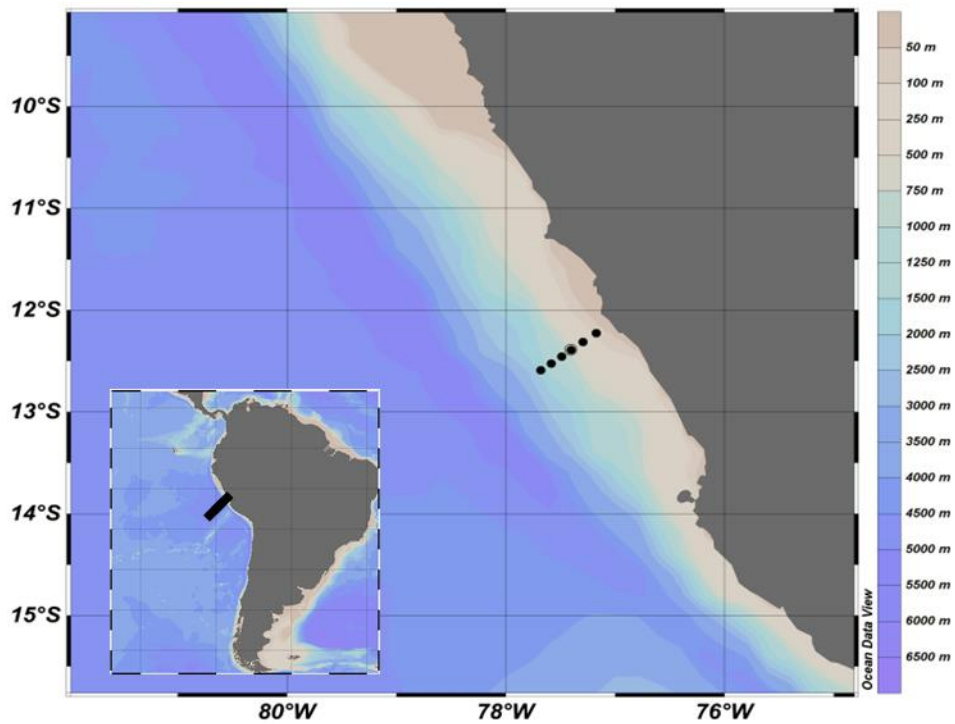
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918 **Figures**

919 **Figure 1**



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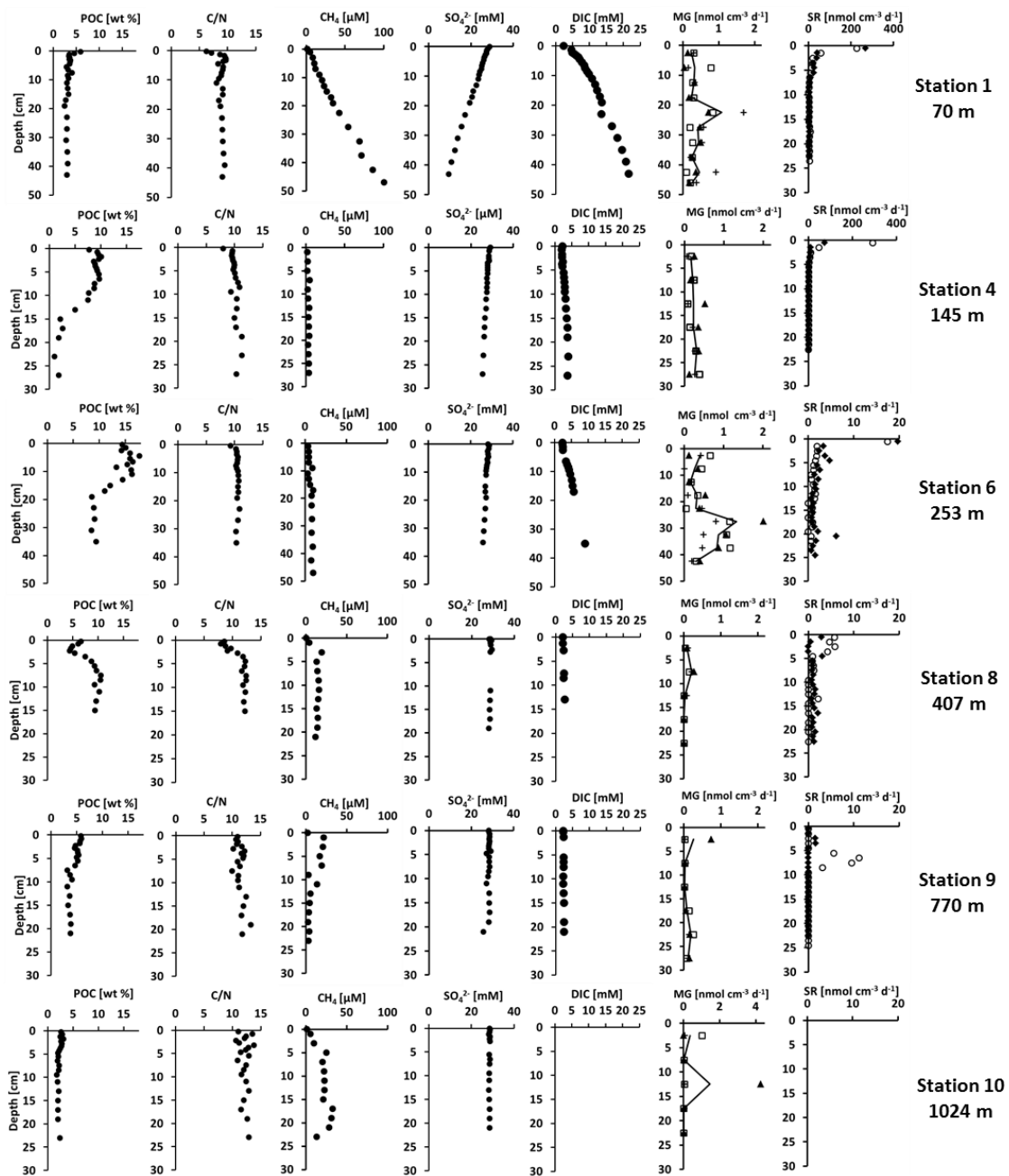
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934 **Figure 2**



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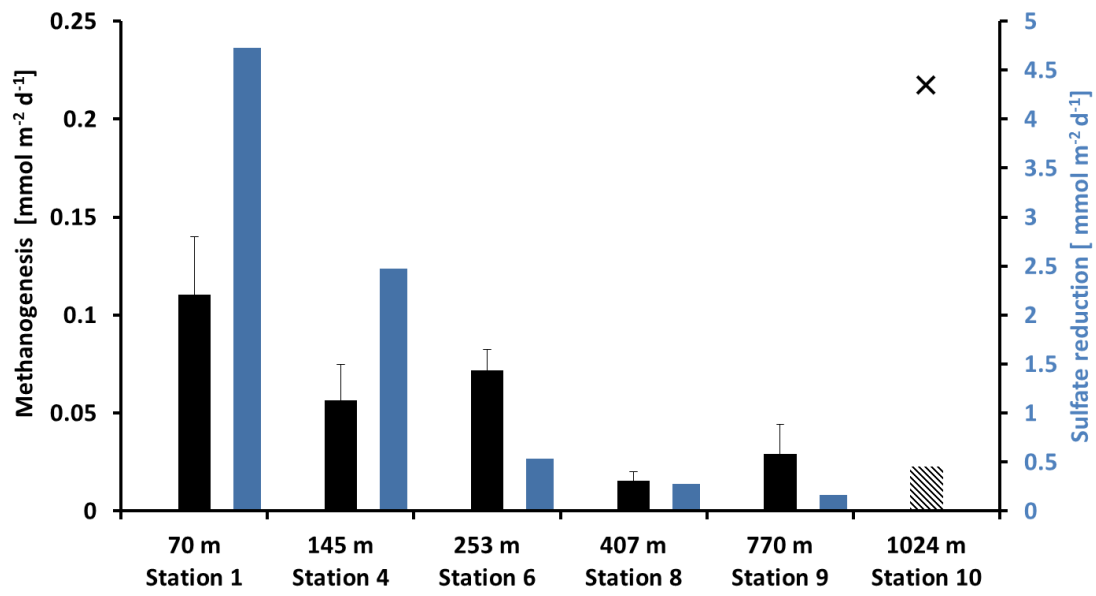
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942 **Figure 3**



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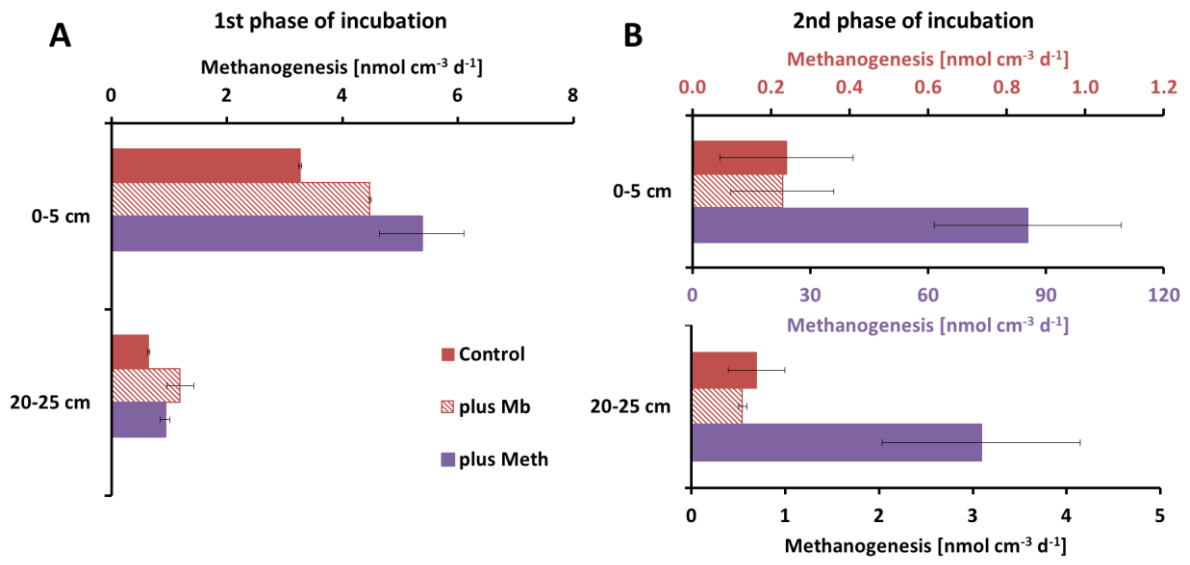
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959 **Figure 4**



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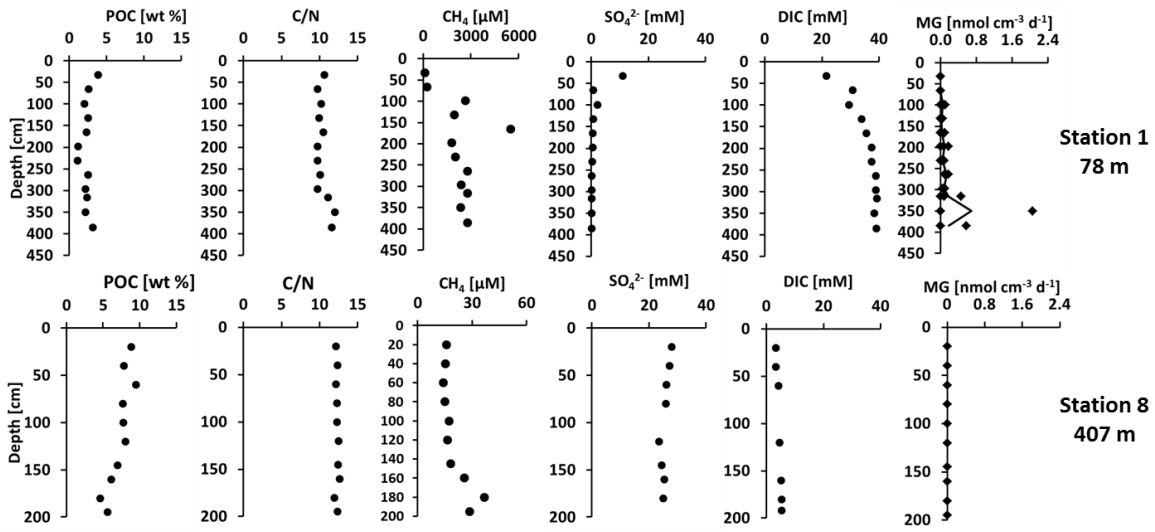
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977 **Figure 5**



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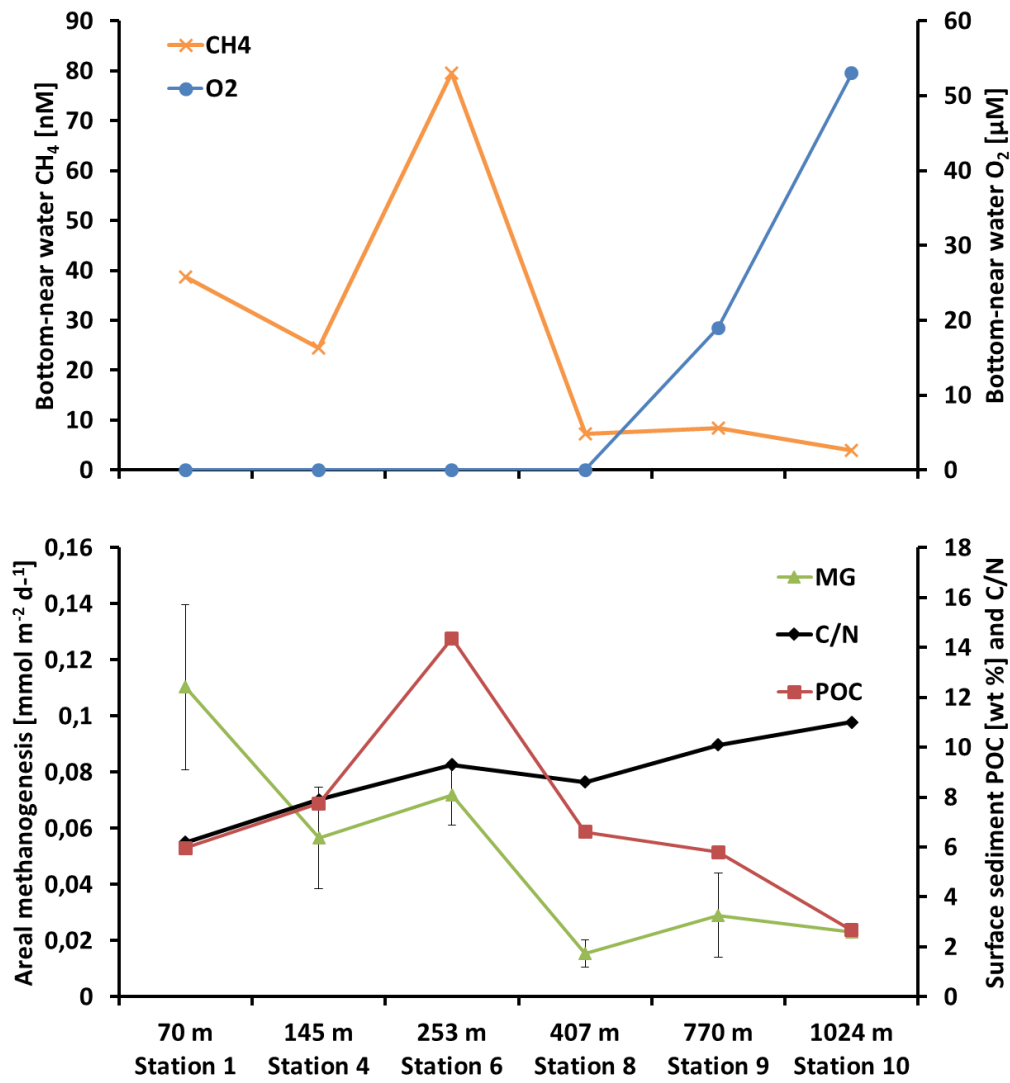
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993 **Figure 6**



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