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# Proximate and ultimate controls on carbon and nutrient dynamics of small agricultural catchments

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## Abstract

Direct and indirect effects from agriculture, urbanization, and resource extraction have dramatically increased nutrient loading to aquatic inland and estuarine ecosystems. The capacity of a watershed to remove or retain nutrients is a function of biotic and abiotic conditions across the terrestrial-aquatic gradient including soil, groundwater, riparian zone, and surface water. The goal of this study was to identify proximate and ultimate controls on dissolved organic carbon and nutrient dynamics in small agricultural catchments. We analysed a five-year, high frequency water chemistry dataset from 3 catchments ranging from 2.3 to 10.8 km<sup>2</sup> in northwestern France. Catchments differed in the relationship between hydrology and solute concentrations, associated with catchment characteristics such as hedgerow density, agricultural activity, and geology. The catchment with thicker soil and higher surface roughness appeared to have greater transient storage and residence time, buffering the catchment to fluctuations in water chemistry, reflected in relatively invariant carbon and nutrient chemistry across hydrologic conditions. Conversely, the catchments with smoother, thinner soils responded to both intra- and inter-annual hydrologic variation with high concentrations of PO<sub>4</sub><sup>3-</sup> and NH<sub>4</sub><sup>+</sup> during low flow conditions and strong increases in DOC, sediment, and particulate organic matter during high flows. Despite contrasting agricultural activity between catchments, the physical context (geology, topography, and land use) appeared to be the most important determinant of catchment solute dynamics based on principle components analysis. The influence of geology and accompanying topographic and geomorphological factors on elemental fluxes is both direct and indirect because the distribution of agricultural activity in these catchments is largely a consequence of the geologic and topographic context. This link between inherent catchment buffering capacity and probability of human disturbance provides a useful perspective for evaluating vulnerability of aquatic ecosystems to human disturbance.

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## 1 Introduction

Direct and indirect effects from agriculture, urbanization, and resource extraction have dramatically increased nutrient loading to aquatic inland and estuarine ecosystems. In the past 60 years, human activity has more than doubled global nitrogen fixation (Gruber and Galloway, 2008) and quadrupled phosphorus loading (Elser and Bennett, 2011), primarily due to agricultural activity and combustion of fossil fuels. At the same time human land-use has directly disturbed approximately half of global land surface (Vitousek et al., 1997), dramatically altering the capacity of ecosystems to buffer or process these nutrient inputs (Seitzinger et al., 2006). These changes in land use and nutrient flux have also altered carbon budgets, stimulating plant and algal growth in estuarine and marine ecosystems but accelerating organic matter decomposition in inland waters (Gruber and Galloway, 2008; Rosemond et al., 2015). Consequently, nitrogen and phosphorus pollution is considered to be one of the most urgent environmental issue currently facing humanity along with loss of biodiversity (Rockström et al., 2009).

The capacity of a watershed to remove or retain nutrients is a function of biotic and abiotic conditions across the terrestrial-aquatic gradient including soil, groundwater, riparian zone, and flow paths (Brookshire et al., 2009; Seitzinger et al., 2006; Sébilo et al., 2013; Pinay et al., 2015). For catchments larger than 100 km<sup>2</sup>, riverine nutrient fluxes are tightly associated with percentage of agricultural cover (Jordan et al., 1997; Omernik et al., 1981; Strayer et al., 2003). However in drainage basins smaller than 10 km<sup>2</sup>, nutrient fluxes vary widely despite similar land covers (Brookshire et al., 2009; Burt and Pinay, 2005; Lefebvre et al., 2007; Groffman et al., 2006; Sebilo et al., 2013). This breakdown of the relationship between land cover and nutrient flux represents an important ecological unknown since 90 % of global stream length occurs in catchments smaller than 15 km<sup>2</sup> (Bishop et al., 2008). It also highlights a practical problem, because though most water quality monitoring takes place in large rivers, most land-management decisions are made at the parcel or small-catchment scale (Thenail et al.,

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2009). This diversity in headwater-catchment response to nutrient loading also represents an opportunity to identify the mechanistic controls regulating nutrient processing and removal.

Two non-exclusive mechanisms may account for the change in variability of nutrient fluxes along stream networks. First, in-stream biogeochemical processes such as nutrient uptake and spiralling vary longitudinally in stream ecosystems, typically decreasing in importance as stream order or discharge increases (Alexander et al., 2009; Hall et al., 2013). Active in-stream removal or retention of nutrients in headwaters could decouple land-use from nutrient flux (Trauth et al., 2013). Second, catchment characteristics such as topography, stream network density, and surficial geology vary moving from uplands to lowlands (Brookshire et al., 2007; Sidle, 2006; Strayer et al., 2003; Pinay et al., 2015) potentially altering terrestrial-aquatic linkages, modulating transport and processing of carbon and nutrients.

We used a multiannual, high-frequency water chemistry dataset from three contrasting headwater catchments in western France to identify controls on carbon and nutrient dynamics. We predicted that in headwater catchments, landscape characteristics such as topography, surficial geology, and location within stream network would have a larger impact on carbon and nutrient concentrations than would land-use. We hypothesized that carbon and nutrient concentrations would show contrasting responses due to both distinct sources and dynamic flowpaths across scales. We expected phosphate ( $\text{PO}_4^{3-}$ ) and ammonium ( $\text{NH}_4^+$ ) to be associated with surface flow, dissolved organic carbon (DOC) with surface and subsurface flow, nitrate ( $\text{NO}_3^-$ ) to come primarily from the unsaturated zone and groundwater, and dissolved silica (DSi) to come from groundwater. To test these hypotheses, we investigated both high-frequency changes in solute concentration during discharge events and seasonal trends at multiple spatial scales.

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## 2 Methods

### 2.1 Study area

We tested our hypotheses with five years of water chemistry data from three small catchments located in the long-term socio-ecological research (LTSER) area Zone Ate-  
lier Armorique in northeast Brittany, France (Fig. 1). The research area is located on  
the French Massif Armoricain (48°36' N, 1°32' W) and is underlain by granite to the  
south and schist to the north (Fig. 2c). 20 years of meteorological data were used to  
estimate mean temperature and cumulative annual precipitation (Supplement Fig. S1).  
The climate is maritime with average monthly temperature ranging from 17.5 °C in July  
to 5 °C in December, and mean annual precipitation of 965 mm, a third of which occurs  
from October to December (Fig. S1). Since the 1950s, the area has been subject to  
intense agriculture with 90 % of arable land occupied by corn, wheat, and pastureland.  
The study area straddles the Couesnon and Guyoult River basins, which discharge into  
the bay of Mont-Saint-Michel and the bay of Le Vivier sur Mer, respectively.

### 2.2 Catchment characteristics and experimental design

To compare the influence of catchment characteristics on carbon and nutrient concen-  
trations we monitored water chemistry at three headwater catchments ranging from 2.3  
to 10.8 km<sup>2</sup> (Table 1) with distinct topography, geology, soil characteristics, and land-  
use (Fig. 1). For the purposes of this study we named catchments by near-surface  
geology, with the catchment G-01 occurring on granite, catchment S-01 occurring on  
schist, and catchment GS-01 occurring on the boundary between the two geologies  
(detailed catchment characteristics summarized in Fig. 1 and Table 1). Soil depth,  
elevation, and land-use vary systematically between the surficial geologies, with the  
granite portion characterized by thicker soils ( $\geq 0.8$  m), higher elevation (85 to 110 m),  
and more permanent or semi-permanent pastureland (Table 1 and Fig. 2). The portion  
of the research area underlain by schist has shallower soils (0.45 to 0.7 m), a thinner

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5 weathered layer, lower elevation (14.5 to 60 m), and a larger proportion of arable land occupied by corn and wheat cultivation (Fig. 2e and Table 1). The catchment situated in the transition zone between granite and schist (catchment GS-01) has contrasted topography (from 105 m upstream to 12.5 m at the outlet) and a mix of pasture and cultivated land.

We measured water flow and chemistry over 5 hydrologic years (from 10 April 1996 to 20 August 2000) at the outlets of the S-01 and GS-01 catchments with an automated sampler (Isco 3700™) controlled by a data logger (Campbell Scientific CR10™). Stage was continuously monitored using pressure transducers and discharge was calculated from rating curves determined by manual gauging with an impellor flow meter and by salt dilution (Day, 1976; Hongve, 1987). Additionally discharge was determined at the gaged station situated at the outlet of the Guyoult bassin (station J0323010) as a part of national environmental monitoring (DREAL).

### 2.3 Water quality analyses

15 Water samples were collected every 12 h during baseflow for the first year, every 3 days for the following 2 years and then monthly for the last 2 years. We programmed the data loggers to trigger more frequent sampling during discharge events, with samples taken every 30 min during the rising limb and every 3 h during the falling limb. There were six high-discharge events captured by the automated samplers at GS-01 and S-01 from 20 June 1997 to April 1998. For some of the longer events sampling frequency slowed during the falling limb and for the largest event (April 1998) only the rising limb was sampled because the high water level damaged the autosampler.

25 Water samples were filtered to 0.45 µm with glass fiber filters and stored at 4 °C until analysis. Analysis of DOC, NO<sub>3</sub><sup>-</sup>, NH<sub>4</sub><sup>+</sup>, PO<sub>4</sub><sup>3-</sup>, DSi, particulate phosphorus (PP), dissolved organic phosphorus (DOP), dissolved organic nitrogen (DON), and chloride (Cl<sup>-</sup>) was performed over the study period. DOC was analyzed with a Shimadzu TOC analyzer and nutrients were analyzed on a Lachat Quick Chem autoanalyser. NO<sub>3</sub><sup>-</sup> was quantified with the modified Griess-Ilosvay method with copperized cad-

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mium reduction,  $\text{NH}_4^+$  by indophenol blue method,  $\text{PO}_4^{3-}$  by automatic ascorbic acid method, DSi by molybdosilicate colorimetric method, and  $\text{Cl}^-$  by colorimetric N-diethyl-p-phenylenediamine method. Total suspended sediment (TSS) was determined by filtering one liter to  $0.45\ \mu\text{m}$  and weighing the filter.

## 2.4 Spatial data and statistical analysis

A 2-m digital elevation model (DEM) based on airborne LiDAR data was used to extract morphological data used in the statistical analysis (Fig. 2a). Land use was mapped annually from aerial imagery from 1996–2000, with ground validation. Forest and pastureland were distinguished from cultivated land, and corn was distinguished from other cereal crops (predominantly wheat but also barley; hereafter wheat) based on color and timing of planting and harvest. Total agricultural coverage was calculated as the sum of all corn, wheat, and other crops. The Web Map Service (WMS) of the Bureau de Recherches Géologiques et Minières (BRGM) was used for the geological map of the study area.

We performed a principle components analysis (PCA) to explore correlations in the data and identify major sources of variability in water chemistry. We included morphological parameters (geology, topography, soil depth, density of hedgerows), land use, and solute concentrations. Specifically we included bedrock (granite, granite and schist, and schist), land use (wheat, corn, and pastureland), stream order (Strahler, 1952), drainage area, difference in elevation between upstream and downstream (dZ), and concentrations of DOC,  $\text{NO}_3^-$ ,  $\text{PO}_4^{3-}$ ,  $\text{Cl}^-$ , and DSi. We also performed Spearman's rank correlations to explore individual correlations between water chemistry parameters. All analyses were performed in R 3.0.2 (R Core Team, 2014) with the FactoMineR package for PCA (Lê and Husson, 2008).

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### 3 Results

#### 3.1 Hydrological and land-use analysis

Hydrologic conditions varied widely across the five years (Table S1). The first three hydrological years (1 September 1995 to 31 August 1998) were 145 to 180 mm below the 20-year average of 965 mm, while the last two years were 170 to 330 mm above average (Table S1 and Fig. S1 in the Supplement). Discharge from the three catchments was in good agreement with the Le Guyoult basin station though there were occasional phase shifts and departures, likely due to localized precipitation and differences in transient storage (Fig. S2).

The distribution of agricultural land was associated with soil and topographical parameters with 55 % agricultural coverage of the total land surface in catchment S-01 in the northwest part of the study area, which has flat topography and thin soils on the schist bedrock. Conversely, catchments G-01 and GS-01 were dominated by pastureland and forest, with less than 38 % agricultural coverage (Fig. 1b, Table 1). The density of hedgerows also varied by a factor of two between catchments; with 105 m ha<sup>-1</sup> of hedgerows in catchment G-01 and 50 m ha<sup>-1</sup> in catchment S-01. Land-use changed little over the study period, with consistent differences between the 3 catchments. For the catchment S-01, variability in land use was highest for corn and wheat (SD 6.6 and 5%, respectively, Table 1).

#### 3.2 Effects of catchment characteristics on water chemistry

The relationship between specific discharge and solute concentration differed by solute and catchment (Figs. 2 and 3). Solutes typically expressed one of three responses to increases in discharge: an asymptotic increase (TSS, DOC, PP), an asymptotic decrease (DON, DOP, and PO<sub>4</sub><sup>3-</sup>), or a decrease in variability and a convergence to a moderate concentration (NO<sub>3</sub><sup>-</sup>, NH<sub>4</sub><sup>+</sup>, DSi, and Cl<sup>-</sup>). Solutes with similar response patterns were more tightly correlated, though significance and sometimes sign of in-

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dividual relationships varied somewhat by catchment (Table 2). The catchment G-01, which had low gradient, high roughness, and relatively less agriculture, had more stable chemistry during both baseflow and highflow conditions, particularly for DOC, TSS,  $\text{Cl}^-$ ,  $\text{NO}_3^-$ , and all phosphorus species. This lower amplitude of variation in water chemistry in catchment G-01 resulted in fewer significant correlations between water chemistry parameters, and notably discharge was not significantly correlated with any solute concentrations ( $p > 0.05$ , Spearman's rank correlation; Table 2). The steep and small catchment GS-01 and the highly agricultural catchment S-01 showed similar patterns of solute response to discharge, though nitrogen and phosphorus concentrations were generally lower in GS-01 than S-01, particularly during baseflow conditions. GS-01 had lower maximum specific discharge ( $47 \text{ L s}^{-1} \text{ km}^{-2}$ ) than catchments G-01 and S-01 ( $75$  and  $66 \text{ L s}^{-1} \text{ km}^{-2}$ , respectively).

Mean DOC concentration was highest in catchment G-01, but did not differ between catchments GS-01 and S, despite a 20 % difference in agricultural coverage (Fig. 4a). Mean  $\text{NO}_3^-$  concentration increased linearly with agricultural coverage from an average of  $5 \text{ mg NL}^{-1}$  in G-01 to an average of  $10 \text{ mg NL}^{-1}$  in S-01 (Fig. 4b). The highest concentration and variability in  $\text{PO}_4^{3-}$  occurred in S-01 where mean  $\text{PO}_4^{3-}$  concentration was nearly three-times higher than in the less agricultural catchments (Fig. 4c).

The PCA exploring the structure of the water chemistry data and catchment characteristics explained 80 % of the variability in the data with the first three axes (Fig. 5 and Fig. S6). The first axis explained 50 % of the overall variability and was most strongly correlated with hedgerow density ( $r = 0.98$ ), land use ( $r = -0.94, 0.92$ , and  $0.78$  for meadows, corn, and wheat, respectively), geology ( $r = -0.70$  for percent of catchment underlain by granite), and  $\text{NO}_3^-$  ( $r = 0.66$ ). The second axis explained 19 % of total variability and was positively associated with DOC ( $r = 0.88$ ) and discharge ( $R^2 = 0.69$ ) and negatively associated with  $\text{NO}_3^-$  ( $R^2 = -0.46$ ). The third axis explained 11 % of total variability and was associated primarily with  $\text{PO}_4^{3-}$  ( $R^2 = 0.79$ ). Overall, the first axis was determined by physical context (geology and topography represented by the difference in elevation and land use) and that the second axis was determined

by factors strongly associated with hydrology (discharge, TSS, and DOC). The three catchments showed clear separation along the first axis and vary along the second axis largely within their discrete, first-dimensional boundaries. The main departures along axis three (in the upper right corner of Fig. 5) are from  $\text{PO}_4^{3-}$  flushing events during dry years (years 1 and 3).

### 3.3 Solute dynamics during discharge events

The high-frequency samples collected during six discharge events at S-01 and GS-01 revealed a primarily counterclockwise hysteresis for DOC (higher concentration during the falling limb than at the equivalent discharge on the rising limb), a clockwise hysteresis for  $\text{NO}_3^-$  and  $\text{PO}_4^{3-}$ , and no clear pattern in  $\text{NH}_4^+$  except for large variations during the rising limb of the discharge events (Fig. 6). DOC and  $\text{NO}_3^-$  concentrations during discharge events were strongly negatively correlated (Fig. S5), with the elements showing nearly mirror-image responses to changes in discharge.  $\text{NO}_3^-$  concentration was highest and DOC was lowest at or immediately after the start of the discharge event, except for GS-01 during the largest two discharge events that were sampled in April 1998 where  $\text{NO}_3^-$  was higher and DOC was lower after the event (Fig. 6). DOC concentration was higher during the second storm pulse than the first for that compound event. Maximum  $\text{PO}_4^{3-}$  concentration typically occurred after the  $\text{NO}_3^-$  peak, but before the maximum discharge, except for S-01 in November 1997. Maximum ammonium concentration occurred during the rising limb. DOC was more strongly correlated with discharge for S-01, though DOC increased for both catchments during the rising limb of the hydrograph (Fig. 6).  $\text{PO}_4^{3-}$  concentration increased strongly at both sites at the onset of the rising limb, with similar peaks for the two subsequent discharge events in S-01 (Fig. 6).

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### 3.4 Inter-annual solute dynamics

The three catchments showed distinct inter-annual dynamics for both hydrology and solute concentrations across the contrasting hydrologic years (Fig. 7). Specific discharge was consistently lowest for catchment GS-01, the steep transition catchment.

DOC concentration in G-01 was invariant across the dry and wet years, whereas annual median DOC in GS-01 and S-01 generally tracked discharge. Highest median DOC concentration occurred in year 3 for all catchments, the transition from dry to wet conditions, and the year with the most high-flow events. Contrary to the high-frequency trends, annual  $\text{NO}_3^-$  was positively associated with annual discharge for GS-01 and S.

Catchment G-01 was again relatively distinct in its behavior, with stable  $\text{NO}_3^-$  concentrations across years (Fig. 7d). Annual  $\text{PO}_4^{3-}$  concentration was negatively correlated with annual discharge across sites, with significantly higher concentrations in dry years.

## 4 Discussion

To quantify the influence of physical, hydrologic, and anthropogenic controls on surface water quality, we monitored discharge and water chemistry from three agricultural catchments in northwestern France for five years. We hypothesized that carbon and nutrient concentrations would show contrasting responses to topographic and land-use differences due to both distinct sources and transport dynamics across scales. We found that carbon and nutrient dynamics differed between event and inter-annual temporal scales, supporting our hypothesis. However, spatially, the effect of hydrology on solute concentration was strongly modulated by catchment characteristics such as hedgerows density, agricultural activity, and geology. Because the distribution of agricultural activity in these catchments is largely a consequence of the geologic and topographic context, the interaction of these factors determines the retention and release of carbon and nutrients.

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## 4.1 Proximate and ultimate controls on water quality

The relationship between agricultural practice and hydrologic nutrient flux is strong at the large-basin scale but breaks down at the small-basin scale with widely different water chemistry in catchments with similar land covers (Brookshire et al., 2009; Burt and Pinay, 2005; Lefebvre et al., 2007). In our study, this phenomenon is apparent for catchments G-01 and GS-01 which have very similar land-use but distinct carbon and nutrient signatures, and for catchments GS-01 and S-01 which have distinct land-use but similar chemical dynamics (Table 2). These differences in carbon and nutrient dynamics can be attributed to catchment characteristics, which modulate carbon and nutrient flux independent of land-use. Granite parent material, such as underlies G-01, can give rise to thick but relatively acidic soils compared to schist substrate which produces thin and rich soils more conducive to row crop cultivation. Thicker soils and higher surface roughness in catchment G-01 likely increase transient storage and residence time, buffering the catchment to fluctuations in water chemistry. This is reflected in relatively invariant carbon and nutrient chemistry across hydrologic conditions. Conversely, catchments GS-01 and S-01, which are underlain primarily by schist, respond to both short- and long-term hydrologic changes with high concentrations of  $\text{PO}_4^{3-}$  and  $\text{NH}_4^+$  during low flow conditions and strong increases in DOC, sediment, and particulate organic matter during high flows. This pattern held on inter-annual timescales as well, where the G-01 showed remarkable stability, while DOC concentration decreased with discharge in the wettest years for the predominantly schist catchments (GS-01 and S-01). This is consistent with progressively smaller carbon pools in forested, pastured, and agricultural surfaces.

In addition to directly influencing catchment characteristics, geology and accompanying topographic and geomorphological factors exert a strong control on the distribution of human agriculture, indirectly influencing nutrient loading and disturbance regime. Because farmers and land managers do not randomly select surfaces for cultivation, land use in Brittany and throughout the world closely follows geologic and soil char-

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acteristics. Changes in soil fertility are a function of natural weathering processes and land use (Tye et al., 2013). Spatial variability of soil moisture, which is often controlled by topography and soil properties (Yeakley et al., 1998) play important roles in land use distribution and organization. In this case, the interactions between catchment context and human use have resulted in preferential agricultural development of schist catchments, which appear to be more prone to nutrient export.

In other contexts, this interaction between risk of human development and resilience to human disturbance can also mitigate impacts of agriculture (such as the preservation of forest in steep, erodible environments; Odgaard et al., 2013), but whether it has a net increase or decrease of human impacts on aquatic ecosystems at a global scale is an open question (Zabel et al., 2013; Ramankutty et al., 2008). We hypothesize that this effect of preferential development of surfaces with variable vulnerability would decrease in areas of intense anthropogenic pressure where selectivity decreases as the system reaches saturation (Li et al., 2014), but could strongly influence the distribution of human activity in systems that are expanding or contracting such as the developing world or areas of rural exodus such as northwestern France.

## 4.2 Controls on chemistry across scales

Different mechanisms can influence short- and long-term elemental fluxes (Moatar and Meybeck, 2012; Moatar et al., 2013), explaining the contrasting short- and long-term dynamics we observed within individual catchments. For example, conditions that favour frequent flushing of soil may decrease short-term  $\text{NO}_3^-$  concentration but result in larger overall fluxes (Figs. 3, 6, and 7).  $\text{NO}_3^-$  showed a non-linear decrease with discharge on short timescales but was higher in wetter years. This is indicative of changing  $\text{NO}_3^-$  sources. The convergence of  $\text{NO}_3^-$ ,  $\text{Cl}^-$ , and  $\text{DSi}$  concentrations across catchments during high-flow periods implies a shift from catchment-scale controls on water chemistry during low flows to larger inter-catchment controls during high flows. The upper layer of the Brioverian schist substrate beneath these drainage basins is composed of an unconsolidated weathered layer of variable thickness that could provide an

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inter-basin source of nutrients that could be activated during storms. This shallow, unconfined aquifer has  $\text{NO}_3^-$  values that correspond to those measured at the catchment outlets during high-water periods (Clément et al., 2003) and in other regional aquifers (Molénat et al., 2008). The lack of significant dilution of  $\text{NO}_3^-$  and  $\text{Cl}^-$  during high-discharge events, and the increase in DOC concentrations, suggest that high discharges are composed of mobilized groundwater and shallow subsurface flow (Grimaldi et al., 2009, 2012), including through the sub-soil weathered layer (Iwagami et al., 2010). The  $\text{NO}_3^-$  mobilized during storms may already be present or could result from mineralization and nitrification as the soils wetted up. The lower, but also constant, concentration of  $\text{NH}_4^+$  measured during high-water periods supports the hypothesis of high nitrogen mineralization from soil organic matter during these mild and very humid periods and high  $\text{NH}_4^+$  retention in soil. Nitrification of  $\text{NH}_4^+$  to  $\text{NO}_3^-$  during these same periods would maintain  $\text{NO}_3^-$  supply, i.e. constant concentration with increasing discharge. The relationship between  $\text{NH}_4^+$  and  $\text{NO}_3^-$  concentrations and discharge (Fig. 3) reveals either a logarithmic or linear trend, underlining the importance of soil as a constant nitrogen source during high-flow periods (Fig. 3) since the  $\text{NH}_4^+$  concentration in groundwater is low.

The various responses of chemistry across temporal scales can give insight into sources and pathways of carbon and nutrients. For catchments GS-01 and S-01, DOC increased strongly with discharge at both event and inter-annual scales. However, high concentrations of DON and DOP were only observed at low-flows. This shift in dissolved organic matter (DOM) stoichiometry indicates a change in DOC sources, with stormflow dominated by plant-derived DOM from surface soils and a larger proportion of microbial DOM from deeper soils and shallow groundwater during baseflow (Inamdar et al., 2012; Yang et al., 2015). The asymptotic increase of DOC during high discharges revealed that organic matter was released at very high rates early in the discharge event, reaching a constant concentration at high discharges. This DOC pattern agrees with observed trends in  $\delta^{13}\text{C}_{\text{DOC}}$  indicating that stream DOC concentration increases with discharge originating from riparian wetlands (Lambert et al., 2011) that

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are widespread in low-order streams in areas of colluvial soil including the current study area (Walter and Mérot, 2007; Laudon et al., 2011). High DOC concentration in these headwater agricultural streams highlights the importance of small streams in DOC production and transport (Agren et al., 2007) and more generally, the role of inland freshwater DOC production on the global carbon cycle (Cole et al., 2007).

We present a synthesis of our understanding of carbon and nutrient dynamics in different catchment components in Fig. 8. Based on the discharge deconvolution, the connectivity between unsaturated and saturated zones depends on water movement in the soil and the weathered zone. Nutrient concentration in streams depends on the thickness of those layers and typical water velocity. By modulating the connectivity between the hydrologic compartments, discharge regime controls the amount of nutrient exported (DOC and  $\text{PO}_4^{3-}$  increasing with discharge and  $\text{NO}_3^-$  decreasing). On shorter timescales, sinusoidal nutrient fluctuations are probably due to discrete storm events that are missed at the longer inter-annual scale. In contrast, various pathways and interaction between hillslope and stream are affected to a lesser extent by discharge regime (Fig. 8). Groundwater fluxes toward the stream may also control nutrients concentration especially for  $\text{NO}_3^-$  which is highest in this compartment (Fig. 8).

### 4.3 Hedgerow density and vegetation effect on soil and shallow groundwater

We found that the effect of hydrology on solute concentration was strongly modulated by hedgerow density, which was the strongest predictor of stream chemistry in the PCA. There was less variation in carbon and nutrient concentrations for the highest hedgerow density catchment G-01. Hedgerows exert multiple controls on watershed hydrology (Mérot and Bruneau, 1993). Because vegetation is one of the major controls on water and energy balance, the removal or redistribution of vegetation with land use alters albedo and evapotranspiration (Davin et al., 2007). Vegetation plays a central role in the interface between the atmosphere and groundwater via water uptake by roots and redistribution of water in the soil column, affecting soil moisture and groundwater recharge. Increased transpiration (Thomas et al. 2012) and interception



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(Ghazavi et al., 2008) by hedgerows can decrease soil moisture at a local scale, potentially reducing the transfer of carbon and nutrients from soils to groundwater or surface waters. Indeed, the relatively dry soil beneath hedgerows (Caubel et al., 2001; Thomas et al., 2008) corresponds with lower soil and shallow groundwater  $\text{NO}_3^-$  concentration (Grimaldi et al., 2012). Enhanced  $\text{NO}_3^-$  removal and or retention by hedgerows could be due to increased variability in soil moisture and longer residence time, creating local microsites for denitrification (Parkin, 1987) and increasing likelihood of uptake. In our study area, hedgerow density is also associated with geologic and topographic parameters. In catchments S-01 and GS-01, where soils are more suitable for intensive agriculture, hedgerows were removed to consolidate fields during the post-war period. While this study cannot untangle the relative impacts of hedgerow density and geology, it does suggest that hedgerow density can impact  $\text{NO}_3^-$  mass balance at larger scales.

## 5 Conclusions

Proximate and ultimate controls on carbon and nutrient concentrations differ across spatial and temporal scales, revealing distinct sources and transport dynamics for different elements. Thicker soils and higher surface roughness for catchments underlain by granite buffer fluctuations in water chemistry at both event and inter-annual scales, potentially due to increased transient storage and residence time. Conversely, catchments on schist substrates are highly sensitive to changes in hydrology. However, the convergence of water chemistry between catchments during discharge events suggests larger, regional influences independent of geology and topography.

Direct human impact on a catchment (fertilizer input, soil disturbance, urbanization) is asymmetrically linked with inherent catchment properties (geology, soil, topography), which together determine catchment resilience or vulnerability to human activity. The effect of hydrology on solute concentration is proximately controlled by catchment characteristics such as hedgerow density and agricultural activity, but because the distribution of agricultural land use in these catchments is largely a consequence of the geo-



logic and topographic context, these are the ultimate controls on retention and release of carbon and nutrients. This link between inherent catchment buffering capacity and probability of human disturbance provides a useful perspective for evaluating vulnerability of human impacts on aquatic ecosystems and for managing systems to maximize storage and minimize leakage of nutrients.

**The Supplement related to this article is available online at doi:10.5194/bgd-12-15337-2015-supplement.**

*Author contributions.* G. Pinay and J. Baudry designed the experiments. O. Troccaz carried them out. Z. Thomas and B. Abbott analyzed the data and prepared the manuscript with contributions from all co-authors.

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**Table 1.** Catchment characteristics (soil depth, topography, bedrock, hedge density, drainage area and land use. Mean and standard deviation of the % of maize, cereals, meadows and woods on arable land use were calculated from variation in land use over the 5-year study period.

		Catchments		
		G-01	GS-01	S-01
Soil	Depth (m)	≥ 0.8	0.7–0.8	0.45–0.7
Topography	Upstream Elevation (m)	110.0	105.0	60.0
	Downstream Elevation (m)	85.0	12.5	14.5
	Difference in elevation (m)	25.0	92.5	45.5
Bedrock	Granite (% of area)	100.0	17.0	6.0
	Schist (% of area)	0.0	83.0	94.0
Hedge	Density (m ha <sup>-1</sup> )	104.72	82.76	49.77
Hydrology	Drainage area (km <sup>2</sup> )	6.4	2.3	10.8
	Basin	Le Guyoult	Couesnon	

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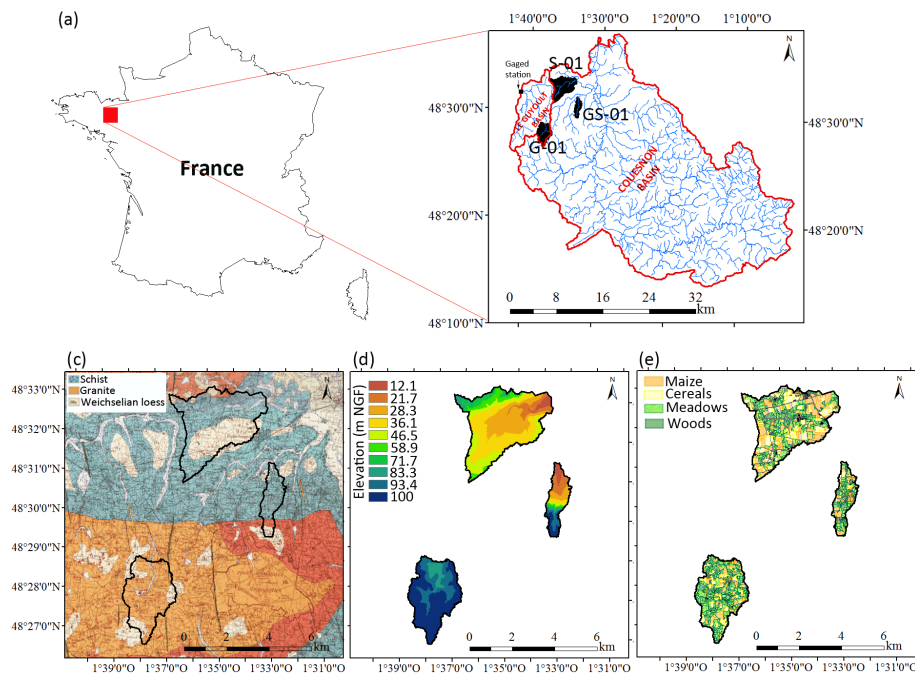


**Table 2.** Spearman's rank correlations for water chemistry parameters. Significant correlations ( $p < 0.05$ , Bonferroni-corrected) are in bold.

Catchment G-01	Q	N-NO <sub>3</sub> <sup>-</sup>	N-NH <sub>4</sub> <sup>+</sup>	P-PO <sub>4</sub> <sup>3-</sup>	PP	DSi	DOC
N-NO <sub>3</sub> <sup>-</sup>	-0.06						
N-NH <sub>4</sub> <sup>+</sup>	0.14	-0.08					
P-PO <sub>4</sub> <sup>3-</sup>	-0.1	-0.09	0.1				
PP	-0.04	-0.29	0.04	<b>0.44</b>			
DSi	-0.11	0.23	0.18	0.28	-0.18		
DOC	0.1	-0.33	-0.12	<b>0.48</b>	0.12	0.14	
TSS	-0.04	-0.16	-0.01	<b>0.37</b>	0.14	0.22	<b>0.46</b>
Catchment GS-01	Q	N-NO <sub>3</sub> <sup>-</sup>	N-NH <sub>4</sub> <sup>+</sup>	P-PO <sub>4</sub> <sup>3-</sup>	PP	DSi	DOC
NO <sub>3</sub> N	<b>-0.24</b>						
NH <sub>4</sub> N	-0.1	0.01					
PO <sub>4</sub> P	<b>0.18</b>	<b>-0.30</b>	0.1				
PP	<b>0.22</b>	<b>-0.23</b>	-0.09	<b>0.64</b>			
DSi	-0.08	-0.13	0.04	0.05	<b>-0.14</b>		
DOC	<b>0.49</b>	<b>-0.62</b>	0.02	<b>0.62</b>	<b>0.43</b>	0.13	
TSS	<b>0.23</b>	-0.08	0.12	<b>0.59</b>	<b>0.55</b>	<b>-0.16</b>	<b>0.42</b>
Catchment S-01	Q	N-NO <sub>3</sub> <sup>-</sup>	N-NH <sub>4</sub> <sup>+</sup>	P-PO <sub>4</sub> <sup>3-</sup>	PP	DSi	DOC
NO <sub>3</sub> N	<b>-0.14</b>						
NH <sub>4</sub> N	-0.07	<b>-0.16</b>					
PO <sub>4</sub> P	<b>-0.48</b>	<b>-0.31</b>	<b>0.23</b>				
PP	<b>-0.22</b>	<b>-0.36</b>	0.06	<b>0.70</b>			
DSi	<b>-0.16</b>	<b>-0.11</b>	<b>0.25</b>	<b>0.43</b>	<b>0.16</b>		
DOC	<b>0.51</b>	<b>-0.28</b>	<b>0.28</b>	<b>0.22</b>	<b>0.21</b>	<b>0.31</b>	
TSS	<b>0.22</b>	<b>-0.17</b>	<b>0.13</b>	<b>0.18</b>	<b>0.35</b>	<b>0.23</b>	<b>0.48</b>

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**Figure 1.** Location and physical characteristics of the three study catchments. **(a)** The red square indicates the Zone Atelier of Pleine Fougères location in the map of France. **(b)** The catchment G-01 is a subcatchement of Le Guyoult basin where the gaged station (black square) is located. The two catchments GS-01 and S-01 are subcatchments of the Couesnon basin. **(c)** Geological overview showing granite in the south and schist in the north. **(d)** Digital Elevation Model from LIDAR of 2 m resolution. **(e)** Land use from the last year of studied period (i.e. 2000).

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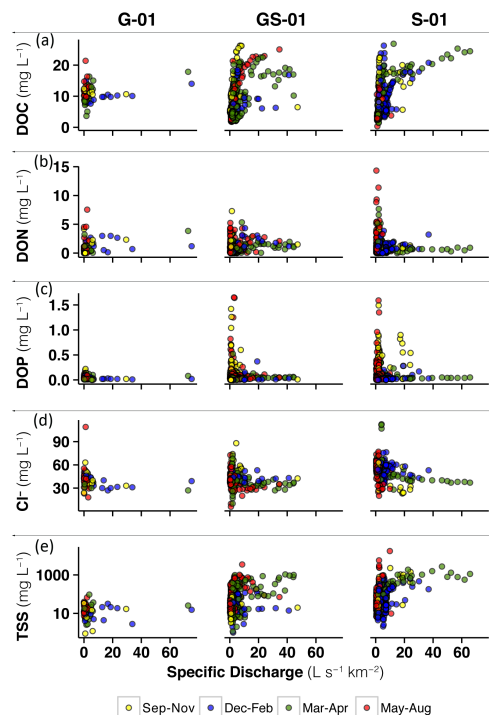
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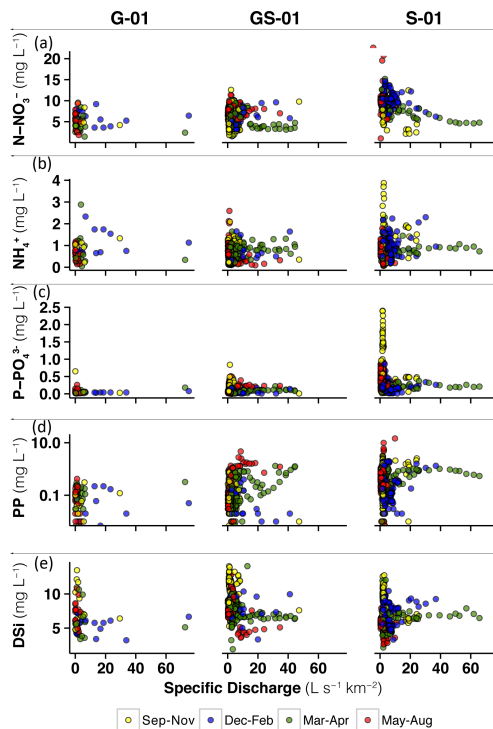


**Figure 2.** Relationship between discharge and (a) dissolved organic carbon, (b) nitrogen, (c) phosphorus, (d) chloride, and (e) total suspended sediment for three headwater catchments (G-01, GS-01 and S-01) in Brittany France. Chemistry data are from daily automatic sampling supplemented by sub-daily sampling for catchments GS-01 and S-01 during discharge events (see methods for detailed sampling description). Note the log scale for TSS.

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**Figure 3.** Relationship between discharge and (a and b) nitrogen, (c and d) phosphorus, and (e) dissolved silica species for three headwater catchments (G-01, GS-01 and S-01) in Brittany France. Chemistry data are from daily automatic sampling supplemented by sub-daily sampling for catchments GS-01 and S-01 during discharge events (see methods for detailed sampling description). Note the log scale for PP.

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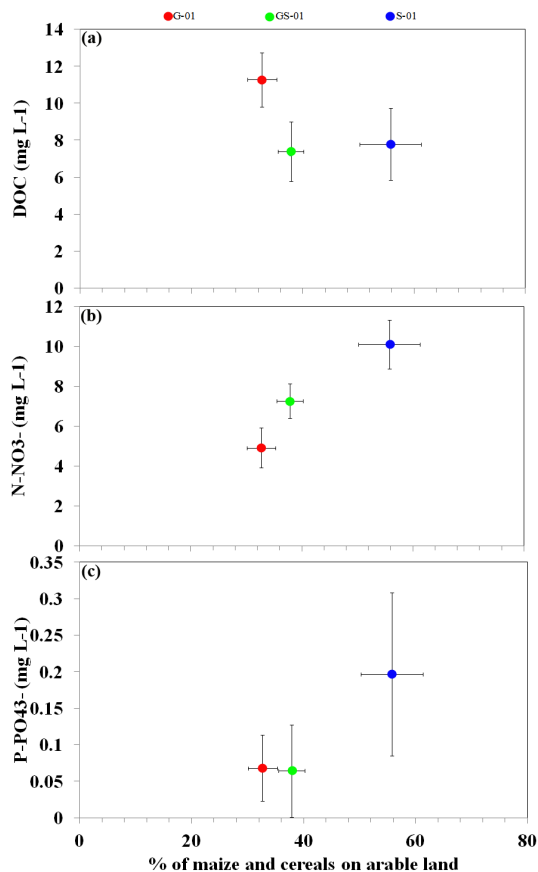
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**Figure 4.** Relationship between (a) dissolved organic carbon, (b) nitrate, and (c) phosphate concentration and land use (% of maize and cereals on arable land) over the studied period. Error bars represent standard error of chemistry parameters and land use.

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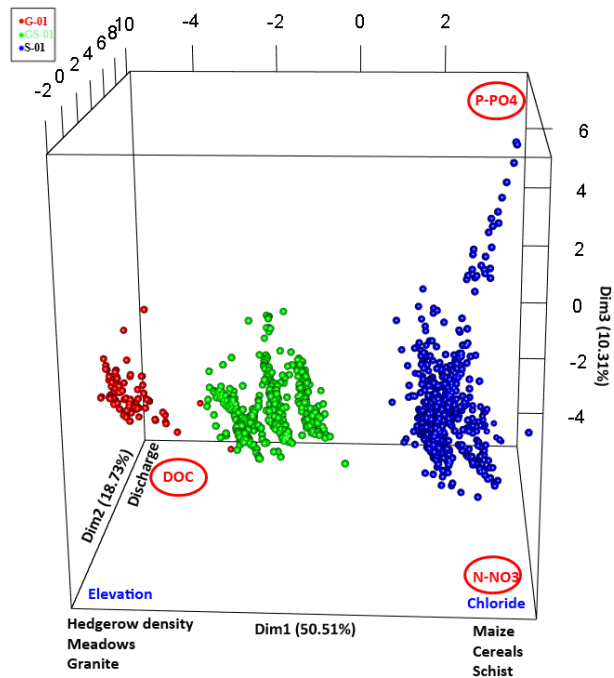
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**Figure 5.** 3-D PCA of the major sources of variability in water chemistry. Together, the first-three axes explain 80 % of the total variability of the data.

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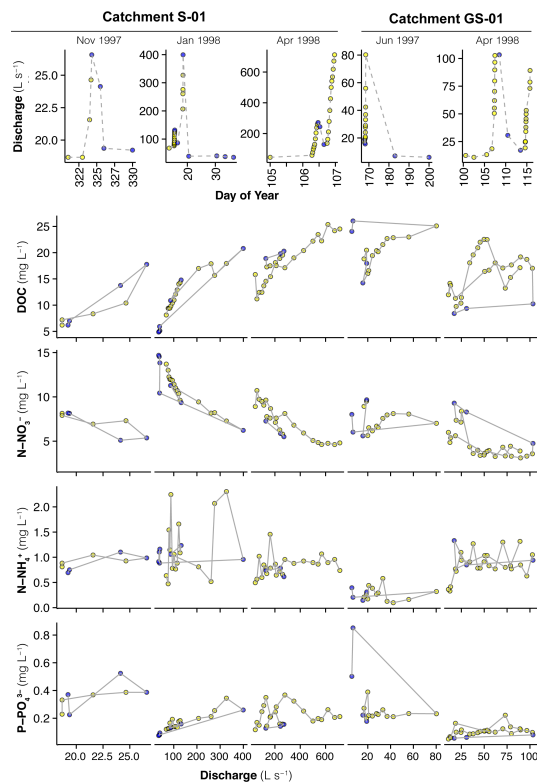
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**Figure 6.** Response of Discharge,  $\text{NO}_3^-$ ,  $\text{PO}_4^{3-}$  and DOC, during storm events for catchment GS-01 and catchment S-01. Yellow points represent the rising limb and blue points represent the falling limb of the hydrograph.

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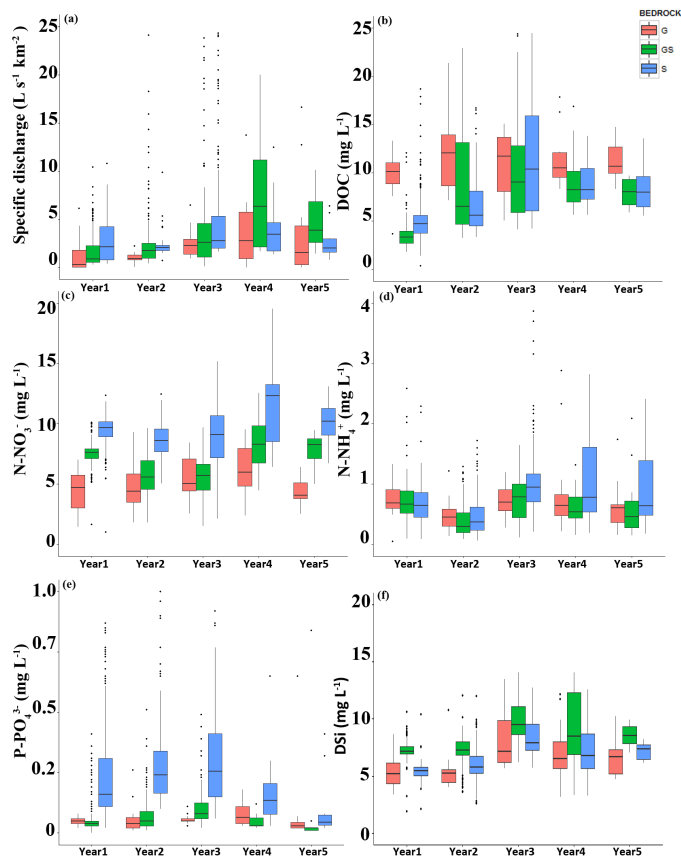
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**Figure 7.** Relationship between water chemistry and hydrological context analyzed using 5 contrasting hydrological years. Boxplot of (a) Specific discharge, (b) DOC, (c)  $\text{NO}_3^-$ , (d)  $\text{NH}_4^+$ , (e)  $\text{PO}_4^{3-}$ , and (f) DSI for the catchments G-01 (red box), GS-01 (green box) and S-01 (blue box). Box plots represent median, quartiles, minimum and maximum within 1.5 times the interquartile range, and outliers beyond 1.5 interquartile range.

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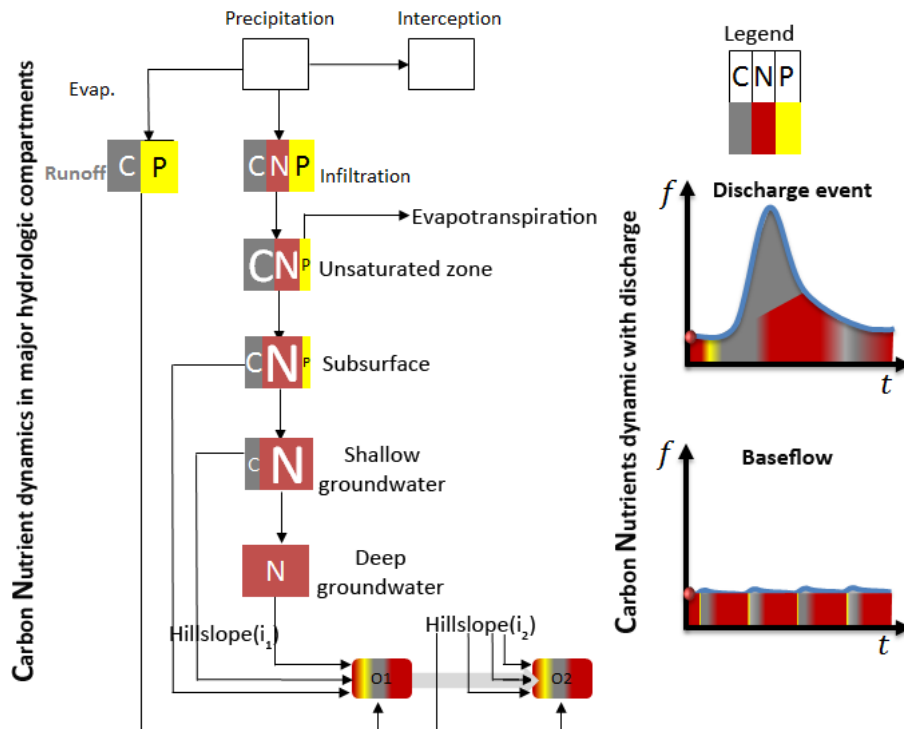
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Interactive Discussion



Proximate and ultimate controls on carbon

Z. Thomas et al.



**Figure 8.** Schematic of carbon and nutrient dynamics in major hydrologic compartments during transient (discharge events) and steady state (baseflow) discharge regimes.

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