#### Point-by-point reply to the Anonymous Referee #1 comments

We would like to thank the referee for their time and input. Their valuable comments and suggestions have been incorporated into our revised manuscript. The referee's comments are shown in *italics* whilst our response and details of the revisions are in normal blue type.

#### Referee's comment

This paper assesses the influence of thinning on the magnitude of CO2 exchanges between a temperate Oak forest and the atmosphere. Using a 9-year time series, the authors aim at evaluating the influence of a thinning event, comparing CO2 flux values before and after the thinning. A strength of the dataset is that the thinning was deliberately operated on a specific wind sector (Eastern part of the tower footprint) while the remaining sector (Western part of the footprint) remained unthinned. This experimental design helps disentangling the influence of meteorological variations from the proper influence of thinning.

#### Author's response

We are pleased to note that the referee acknowledges the strength of our data set and the benefits of the experimental design.

#### Referee's comment

However, this design is not able (as the authors acknowledge) to rule out the fact that easterly and westerly air masses are associated with different weather conditions. Flux data are not really easy to deal with, since they require many corrections, filtering and possibly gap-filling before being used. My main concern with the paper lies there. The authors base a large part of their analysis on the comparison (between sectors) of: 1. monthly or annual flux sums (Fig. 4-5), obtained from gap-filled and partitioned data. The intention is good, but the authors completely overlook the uncertainty associated to such aggregated data. As appears from Table 2, there are 72% and 65% gaps respectively in the half-hourly data from the West and East sectors. These gaps are filled before calculating time integrals, but the authors say nothing about the uncertainty associated to the gap-filling (GF), the precise method used (simply referring to the website of the online tool they used), nor do they intent to estimate the uncertainty on the time integrals induced by GF. This point must be addressed since the paper relies on comparisons between fluxes integrals (from the unthinned and thinned sectors) that may in fact not be significantly different. I perfectly understand that the same method was used to gap-fill the data from both sectors, but this is not an argument to insure that gap-filled fluxes can readily be compared.

#### Author's response

We acknowledge that our experimental design has resulted in large gaps in the half-hourly data and that this was of concern to both referees. We also acknowledge that using the standard MDS method to gap fill from low amounts of original data is likely to result in large uncertainty, which we did not address in the initial manuscript. Substantial revisions in order to address this concern have been incorporated in to the manuscript and are detailed below, following the referee's recommendations.

#### Referee's comment

I further question the absence of "bias in the data availability (e.g. day or night, or seasonal distribution) that might have affected our gap-filled annual total CO2 flux component estimates" the authors mention on p. 16212, since there is likely a seasonal predominance of W or E winds, associated to different meteorological conditions. A way to rule that out is to make a figure presenting the seasonal windrose / the availability of data from each sector along the season.

### Author's response

We acknowledge the referee's concern about, "bias in the data availability....." in the revised manuscript we have provided more information about the data availability and the relative contribution of fluxes during the day / night and during each season by sector as suggested, please see the modified table 2.

#### Referee's comment

I recommend the authors to either (1) stick to the comparison of average, non-filled, fluxes or (2) assess the influence of gap-filling methods on the calculated fluxes (see e.g. Moffat et al., 2007; van Gorsel et al., 2009). If the authors chose option (2), they should further consider the fact that the gap-filling of fluxes is influenced by EC "random errors" (e.g. Ollinger & Richardson, 2005), and run multiple iterations of each GF algorithm to account for such uncertainties.

I would recommend option (1). Do not forget that using gap-filled values translates in using a dataset that contains 65% to 72% of model-derived, uncertain data... Consider further the fact that the statistical model used to GF the data is data-driven...

#### Author's response

We acknowledge the weakness in gap filling this type of data where capture is low. In our revised manuscript we have therefore adopt recommendation (1) and removed all reference to the gap filled data including the integrated annual sums. In light of these changes, the manuscript has been substantially restructured to ensure a greater emphasis of the seasonal average fluxes and we have included more information about their uncertainty (see section 3.2 and 4). The restructured manuscript now focuses on the use these data to explain changes in the physiological processes during the pre- and post-thinning phases.

#### Referee's comment

2. parameters of light-response curves of NEE (Fig. 6) In that case, the reader wants to see the uncertainties, assessed as the standard error, of the parameters. Please make proper statistical comparisons of the values before jumping to the conclusion that NEE800 from one sector is different from NEE800 from the other.

### Author's response

We have substantially revised Fig. 6 (now Fig.5), which now includes error bars indicating the upper and lower 95% CI for each parameter.

### Referee's comment

The paper is divided in 2 parts: in the first part of results, the authors deal with flux data, while in the second part, they analyze Lidar data, to assess changes in the structure of the stand. The junction between the Lidar data and flux data is not obvious. Lidar provides assessment of changes in the spatial structuration of the stand, not of changes in the surface properties susceptible to impact CO2 fluxes. In other words, those Lidar data are interesting to view per se, but bring little to the comprehension of the influence of thinning on CO2 exchanges of the forest with the atmosphere. A more direct link with fluxes could be through Leaf Area Index (LAI) and biomass data.

### Author's response

With respect, we strongly disagree with this comment. The Lidar data was important to demonstrate a large increase in the frequency of gaps in the forest canopy gap fraction, an indication of the structural changes that occurred within the forest canopy during the thinning operation.

#### Referee's comment

B) The authors assessed the influence of thinning on CO2 fluxes. Why not on H2O fluxes?

#### Author's response

The focus of the paper from the title onwards is on C balance of this deciduous forest. Were we to include analysis and discussion of the H2O fluxes, we believe that this would not be beneficial as it would fundamentally change the nature and aims of the paper.

### Referee's comment

Last, throughout the text, some assertions or even conclusions of the paper are not really discussed, or stated gratuitously: - the authors refer several time to a caterpillar caused defoliation, occurring in 2010, which would have impacted more strongly fluxes from the thinned than from the unthinned sector (L20 P 16198, L10 P 16208). Are there data justifying

the differential influence of caterpillars among sectors? What may explain that the grazing was more important in the thinned sector?

#### Author's response

The reference to caterpillar caused defoliation has been substantially moderated, we have removed the sections of text that previously referred to caterpillar defoliation from our introduction, section 3.2 and from the discussion. The only remaining mention of caterpillar defoliation is across the whole forest which we reference using a previously published paper (Wilkinson et al., 20102).

#### Referee's comments

How to explain that the largest differences in Reco between sectors were observed 2 years after the thinning (L9 P 16209)? Due to a lagged respiration of debris? I notice from Fig. 5 that Reco is even lower in the E (thinned) sector in 2008, as compared to the W sector.

- Linked to what precedes: how to interpret the higher sensitivity of Reco to Temperature in 2009 in the E sector? It is not just a question of higher availability of substrate to decompose (i.e. woody debris) but of the sensitivity of Reco to T.

#### Author's response

We think that there has been a misunderstanding in the referee's interpretation of this graph. Fig. 5(b) (now figure 4 (b) shows 2009 not 2008, secondly Fig. 5(b) shows that Reco was *higher* (not lower) in the E (thinned) sector in 2009, as compared to the W sector, in the discussion we consider the possible reasons for this at length.

#### Referee's comment

Throughout the text, and on Figures, make clear and univocal distinction between west (=unthinned) and east (=thinned) sectors. Name them W/E or U/T once and keep on with that denomination. Confuse at the moment.

### Author's response

The manuscript has been revised, accordingly.

#### Referee's comment

P 16203, L2-6: Has the influence of instrument changes on flux calculation estimated? Measured through out an overlapping interval?

# Author's response

Logistically it was not possible to run both sets of instrumentation in parallel. However, we have analysed the data pre- and post-changes in detail and can find no evidence of instrument induced bias.

#### Referee's comment

P 16203 L21-23: U\* threshold depends on the surface structure. Hence it is not necessarily relevant to use the same threshold for both sectors. Can we see the u\*-threshold selection plots (e.g. as suppl. material)?

#### Author's response

We have provided the revised U\* thresholds for each sector and in each year, calculated using the method of Papale et al., (2006) as a suppl. table.

### Referee's comments

P. 16205, Eq. 2: Please revise the definition of NEEmax. In the actual form of the equation, the asymptote of the relationship is not NEEmax but GPPmax (be careful that there is an offset by Rd).

16205, Eq. 2: "incident" quantum yield should be replaced by "apparent" quantum yield. Remind that Quantum yield is originally defined at the leaf scale. So at the canopy scale, the relevant expression is "apparent QY". (Please check, there are other occurrences of "incident QY" in the text).

P 16207, L4-6: this sentence is pleonastic. Rephrase.

P16208, L27: this statement is gratuitous.

P16209, L 4-11: what is Rs? Is it different from Reco? Use consistent denominations for the same flux throughout the manuscript

P16209, L21-22: tricky way to present those results... The sentence "values of NEEmax were generally larger (more negative)" is arithmetically wrong. Why not working with NEP=-NEE? It would simplify much the interpretation, and avoid sign confusion (same remark: rephrase L9 P 16214).

P 16211 L 19: the main result of the paper is not "surprising", considering what we know from the literature (and the authors remind: Vesala et al., Granier et al.)

### Author's response

All of these revisions have been incorporated into the revised the manuscript.

#### Referee's comment

P16213 L1: uncertainties of the light-response curve parameters do not appear in the paper (though it is much needed to allow the reader interpret the results).

#### Author's response

We have now incorporated uncertainty estimates to the figures and tables where appropriate including to Figure 5.

#### Referee's comment

Table 1: there clearly is a need for numbers before thinning. What about the influence of thinning in terms of removed (1) # of stems (2) biomass? A post-thinning W/E comparison is not enough.

## Author's response

Regrettably these data are not available, however the pre- and post-thinning west and east surveys did show a large and clear difference in tree density between the two sectors (Table.1).

### Referee's comment

Table 3: Q10: which base temperature?

Fig. 1: date the photograph. I assume this was taken after thinning.

Figs 3-7: blue (east= thinned) / green (west= unthinned). Use a systematic color code and make the legend apparent on each figure.

Fig. 6: indicate confidence intervals of the regressions on the graphs

### Author's response

All of these technical corrections have been incorporated into the revised manuscript. Fig. 6 has been removed as it is no longer necessary.

#### Point-by-point reply to the Anonymous Referee #2 comments

We would like to thank the referee for their time and input. Where appropriate their comments and suggestions have been incorporated into our revised manuscript. The referee's comments are shown in *italics* whilst our response is in normal blue type.

We acknowledge that the referee has some major criticism of our experimental set up, analysis of the data and of the manuscript. We hope that this detailed description of the changes that we have made, in addition to the other supporting documents will address these concerns.

#### Referee's comment

This manuscript attempts to quantify effects of thinning on CO2 exchange in deciduous forest canopy by analyzing one eddy-covariance tower with different wind directions. The authors argue that the effects of thinning on the carbon balance were not significant. The subject of this study will be of interest to scientific community because previous studies on impacts of thinning on carbon balance have been done in coniferous forests. However, I cannot confirm that the conclusions of this manuscript were drawn correctly because only one tower is located at the border between thinned and un-thinned sectors, which makes us difficult to test statistical significance and to properly interpret physical implications. Sufficient data and thorough investigation are needed more. I will not bring up specific issues and please consider major concerns below for giving more solid evidences to this study.

#### Author's response

We are pleased to note that the referee has commented that this study will be of interest to the scientific community and welcome this. Whilst we understand the referee's concern that about the interpretation of EC data split by wind sector, from only one tower, we would like to draw attention to the fact that this technique has been used successfully in previously published studies (e.g. Parmentier et al., 2011).

#### Referee's comment

1. It is misleading to discuss to differences of climatic conditions such as downward solar radiation, air temperature, wind and humidity between thinned and un-thinned sectors. For example, downward solar radiation should be same at these two sectors because the un-thinned and thinned sector are not hundreds kilometer away. Thinning management cannot make impacts on downward solar radiation! People may want to know changes in albedo and outgoing longwave radiation, and so net radiation more. But I am not quite sure physical meanings of radiative fluxes from radiometer close to the boundary of the thinned and un-thinned sectors.

#### Author's response

With respect, we think this referee has misunderstood the presentation and discussion of the met data (e.g. Fig 2 & 3a), which was separated according to when the wind was from either the E or W sector. We did not say that these differences were *because* of the thinning, but tried to make it clear that the differences in conditions *associated* with different wind directions (and therefore other weather conditions) will have affected the fluxes and needed to be taken into account. This is why in the revised manuscript we have placed greater emphasis on the examination of NEE response curves to light intensity and temperature sensitivity of respiration, rather than simply compare fluxes across time.

#### Referee's comment

2. The first issue is going to another issue. The different solar radiation between the two sectors indicates that solar radiation has been sampled on different time between thinned and un-thinned sectors. Let me show one example. 1) Flat and homogeneous surface without any disturbance like thinning. 2) Air temperature was higher on the first day than the second day because of different synoptic condition. 3) Main wind comes from the east on the first day but from the west on the second day. 4) If we compare air temperature between the east and west sectors, air temperature in the east sector is higher than the west sector. 5) Absolutely, thinning does not make this difference. We

need clear discrimination on these kinds of different from thinning effects but I am quite sure if one tower measurement can resolve this issue.

### Author's response

This point illustrates the referee's confusion noted in the previous point; we clearly stated that the different conditions are because of changing wind directions. We agree with the referee that thinning does not alter the meteorological conditions that the forest is exposed to, but the conditions *do* change with wind direction. We believe the existing ms was clear on this.

#### Referee's comment

3. The first and second issues are moving to another issue. The authors said that data retrieval rate is only 30%, indicating that 70% missing data are filled by the marginal distribution sampling (MDS). MDS is looking for the observed NEE values of similar climatic conditions. Therefore,

if more than 2/3 data are missed, MDS feel difficulties in finding the similar climatic conditions and will extend the time windows to find the similar climatic conditions. In this case, we expect that uncertainties in the gap-filled data increase dramatically. Furthermore

the gap-filled data strongly depends on climatic conditions which is related to the second issue above. How can the authors quantify these uncertainties and their impacts on data interpretation for the thinned and un-thinned sectors?

#### Author's response

The referee's concern about the use of gap filled data and the uncertainty associated with this method was also highlighted by Referee #1. We acknowledge this weakness, and as stated in our response to Referee #1 in our revised manuscript we have adopted their recommendation (1) and remove all reference to the gap filled data including the integrated annual sums. As previously stated in our response to Referee #1 we have restructured the manuscript with a greater emphasis on the seasonal average fluxes, and we have included significantly more information about the uncertainty in these values.

### Referee's comment

4. All interpretations of the authors are not based on solid statistical test. All figures and tables do not have any statistical test results (e.g., p value). For example, Figure 6 shows light response curves before and after the thinning management. But this figure only shows fitted curves without any error range and p-value. In addition to uncertainties in the measurements itself and data processing, it is difficult to say any difference or similarity with strong confidence.

# Author's response

Referee #1 also pointed out the present lack of statistical evaluation. Where appropriate we have included more information about the uncertainly and variation. In the restructuring of the paper, we have removed Fig. 6 which was of concern to both referees.

### Referee's comment

5. With the current experimental design, it is impossible to quantify changes in radiative fluxes, soil temperature and soil moisture, which are critical information on the thinning effects on carbon cycle.

### Author's response

It is true that we did not measure soil temperature / soil moisture at different locations within the forest and we agree that these variables are major determinants of the C cycle. However, we were not seeking to measure how these changing 'internal' (within-stand) factors affect the C cycle, but relating the observed C fluxes to the 'external' driving weather factors of Ta (at 26m) and incoming Srad.

#### Referee's comment

6. How can we separate disturbance by caterpillar from thinning management?

## Author's response

As pointed highlighted to Referee 1# the reference to caterpillar caused defoliation has been substantially moderated, we have removed sections of text that previously referred to caterpillar defoliation from our introduction, section 3.2 and from the discussion. The only remaining mention of caterpillar defoliation is across the whole forest which we reference using a previously published paper (Wilkinson et al., 20102).

# Reference

Parmentier, F. J. W., J. van Huissteden, M. K. van der Molen, A. J. Dolman, G. Schaepman-Strub, S. A. Karsanaev, and T. C. Maximov (2011), Spatial and temporal dynamics in eddy covariance observations of methane fluxes at a tundra site in northeastern Siberia, J. Geophys. Res., 116, G03016, doi: 10.1029/2010JG001637.

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not until two years after the thinning had been completed (2009); initially this was associated with		
an increase in ecosystem respir	ation $(R_{eco})$ .	

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. Ecosystem respiration fluxes increased in the thinned relative to the unthinned area in the post thinning phase.

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In subsequent years, NEE remained lower with reduced carbon sequestration in fluxes from the thinned area, which we suggest was in part due to heavy defoliation by caterpillars in 2010 reducing GPP in both sectors of the forest, but particularly in the east.

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Following an analysis of night-time NEE dependence on friction velocity using the method described by Papale et al. (2006), n

Page 7: Inserted **Matthew Wilkinson** 16/03/2016 08:31:00 a critical threshold (supplementary table 1).

Page 7: Deleted **Matthew Wilkinson** 15/03/2016 14:56:00 0.2 m s<sup>-1</sup>, following an analysis of night-time NEE dependence on friction velocity.

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Thirty minute NEE fluxes for both sectors were subsequently gap filled using the marginal distribution sampling method Reichstein et al., 2005) to allow the calculation of annual totals, and

Page 7: Deleted Matthew Wilkinson 14/03/2016 10:45:00 partitioned into the component fluxes of

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Ecosystem respiration (Reco) was calculated for each sector using the method proposed by Reichstein et al (2005). Here each dataset was split into ten-day consecutive periods and Reco was estimated using the Lloyd-Taylor regression model (Lloyd and Taylor, 1994) between night time CO<sub>2</sub> flux (global solar radiation < 20 W m<sup>-2</sup>) and air temperature. The estimated value of R<sub>eco</sub> was then assigned to the central time point of the averaging interval and linearly interpolated between time points.

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R<sub>eco</sub> and GPP using the on-line eddy covariance gap filling and partitioning tool (http://www.bgc-jena.mpg.de/~MDlwork/eddyproc/index.php). In brief, this uses the night-time fluxes and their response to temperature to estimate respiration during the day.

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The annual gap-filled totals of NEP and its component partitioned fluxes ( $R_{eco}$  and GPP) were calculated for both sectors of the forest (Fig. 4). Before the thinning the annual GPP was slightly higher (+6%) in fluxes from the west than those from the east sector (mean difference 2004-2007 = 121.3 g C m<sup>-2</sup> y<sup>-1</sup>; SE 33.2). After the thinning, the difference increased (mean difference 2008-2012 = 238 g C m<sup>-2</sup> y<sup>-1</sup>; SE 88.6) and annual GPP was higher in the west sector in all years, apart from 2009. Heavy defoliation by caterpillars (Wilkinson et al., 2012) occurred in 2010 reducing GPP in both sectors of the forest, but particularly in the east. Prior to thinning the annual  $R_{eco}$  was similar in both sectors of the forest (mean difference 2004-2007 = 26.3 g C m<sup>-2</sup> y<sup>-1</sup>; SE 58.1). In the first year after thinning (2008)  $R_{eco}$  was slightly lower (-12%) in the east than the west however, a year later, in 2009,  $R_{eco}$  increased substantially (+34%) in the east compared to the west, but in subsequent years it dropped in the east (mean difference 2010-2012 = 80.0 g C m<sup>-2</sup> y<sup>-1</sup>; SE 56.0).

NEP was also generally larger in the west sector than the east prior to the thinning (mean difference 2004-2007 = 95.3 g C  $m^{-2}$   $y^{-1}$ ; SE 36.09). In 2007, the year of the thinning, NEP was substantially higher for both sectors, which was a result of the

weather in that year, with a longer growing season (19 days longer than the average) and high peak LAI (see Wilkinson et al., 2012). From 2009 onwards there was a substantial increase in the ratio of  $R_{\rm eco}$  to GPP in the east which was not evident in the west sector (Fig. 4d), resulting in much larger differences in NEP between the two sectors after thinning (mean difference 2009-2012 = 318.3 g C m<sup>-2</sup> y<sup>-1</sup>; SE 39.5).

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this was not the case		
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there were deviations from t	this most notably	
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small (although non-signifian		
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The asymptotic net CO <sub>2</sub> assi	milation rate $(F^{\infty})$	
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The maximum rate of light s	aturated photosynthesis (NEE <sub>max</sub> )	
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in a 5 °C range (18 - 23 °C).		
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NEE <sub>max</sub>		
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reatment (mean difference
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but		
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significant		
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in fluxes from either sector		
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Prior to 2007, the magnitude	of	
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There was a small increase (	2.2 and 2.1 $\mu$ mol m <sup>-2</sup> s <sup>-1</sup> ) in NEE <sub>800</sub> in t	the east relative to the west in
2009 and 2010 respectively	. In 2011 and 2012 the two sectors r	eturned to their pre-thinning
ranking. Incident		
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much larger (more negative)		
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(except for 2007 & 2012) ind	dicating that this area of the forest u	
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two sectors converged in the	e post thinning phase.	
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sed low radiation levels more		

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Although

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post thinning relative to the west and remained higher through to 2012

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was usually about half the value of that in the west sector there was substantial inter-year variation and no clear change was associated with thinning

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confirming the results of R<sub>eco</sub> estimated using the Reichstein et al (2005) method

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c) the substantial inter-annual variation in component CO<sub>2</sub> fluxes (evident in Fig. 4);

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, which after classification and quality control can be as low as 22.5%

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However there was no evident bias in the data availability (e.g. day or night, or seasonal distribution) that might have affected our gap-filled annual total CO<sub>2</sub> flux component estimates. The main difference in CO<sub>2</sub> flux between the two sectors after thinning was substantially lower NEE in fluxes from the east (Fig. 4c) although this was only observed from 2009 onwards.

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, although there is substantial variation in the data and hence uncertainty in the fitted parameters.

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probably

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(Wilkinson et al., 2012)

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(more negative)

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the east in

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both sectors in		
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may be as a result of the thin	nning	
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is also likely to be an effect	of defoliation by caterpillars, possibly n	masking the effects of the tree
thinning but		
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as it		
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R <sub>d</sub> estimated from the light	response curves increased in the first y	ears after thinning in the eas
relative to the west.		
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We therefore expected an in	nitial stimulation of Reco in the first year	ar after thinning, but this wa
not observed, which may be l	because of slow colonisation by microflo	ora or because of the reduction
in living biomass (above an	d below ground) decreasing Ra. The la	ck of Reco response may als
have been because much of t	the coarse above	
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In addition, much of the lar	In addition, much of the large				
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ground					
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which may have been a substantial source of CO<sub>2</sub>, although we have no independent measurements of emission from them.

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therefore was not in direct contact with the soil surface and more likely to dry out, as well as reducing its availability to soil microbes.

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, which may affect decomposition		
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The low NEE in fluxes from the east after 2009 were consistent with a step-wise increase in the ratio of  $R_{\text{eco}}$  to GPP (Fig. 4e).

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ly		
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ecosystem respiration was higher		
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NEE was lower		
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east sector		
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onwards and remained higher  Page 15: Deleted  two years after the thinning.  Page 15: Inserted	er until the end of the study period.  Matthew Wilkinson	15/03/2016 15:44:00

The  $R_{eco}$  component increased in the thinned area but not until two years after the thinning had been completed and was associated with an increase in the sensitivity of  $R_{eco}$  to temperature.

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(taken in spring 2008)		

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, error bars represent 95% confidence intervals,		

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Figure 5. Summer (July and August) daytime modelled light response for the east sector (blue line) and west sector (green line) for the years of (a) 2006 (b) 2009 (c) 201 at the Straits Inclosure, Alice Holt Forest. Data limited to 18 -23 °C air temperature conditions.

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# Effects of management thinning on CO<sub>2</sub> exchange by

# 2 a plantation oak woodland in south-eastern England

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# **Abstract**

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Forest thinning, which removes some individual trees from a forest stand at intermediate stages of the rotation, is commonly used as a silvicultural technique and is a management practice that can substantially alter both forest canopy structure and carbon storage. Whilst a proportion of the standing biomass is removed through harvested timber, thinning also removes some of the photosynthetic leaf area and introduces a large pulse of woody residue (brash) to the soil surface which potentially can alter the balance of autotrophic and heterotrophic respiration. Using a combination of eddy covariance (EC) and aerial light detection and ranging (LiDAR) data, this study investigated the effects of management thinning on the carbon balance and canopy structure in a commercially managed oak plantation in the south-east of England. Whilst thinning had a large effect on the canopy structure, increasing canopy complexity and gap fraction, the effects of thinning on the carbon balance were not as evident. In the first year post thinning, peak summer photosynthetic rate was unaffected by the thinning, suggesting that the better illuminated ground vegetation and shrub layer partially compensated for the removed trees. Peak summer photosynthetic rate NEE was reduced in the thinned area between 2009 and 2011 but there was no significant difference between sectors not until two years after the thinning had been completed (2009); initially this was associated with an increase in ecosystem respiration (Reco). Ecosystem respiration fluxes increased in the thinned relative to the unthinned area in the post thinning phase. In subsequent years, NEE remained lower with reduced carbon sequestration in fluxes from the thinned area, which we suggest was in part due to heavy defoliation by caterpillars in 2010 reducing GPP in both sectors of the forest, but particularly in the east.

# 1 Introduction

In England, woodlands cover 10.0% of the land surface area, with the majority (0.78 Mha) comprising broadleaved woodland (Forestry Commission, 2013). The total carbon stock in the forest biomass was estimated to be 105.4 MtC in 2011 (Forestry Commission, 2014), with 27.7 and 77.7 MtC in conifer and broadleaved woodlands respectively. Much of this broadleaved woodland is in small areas, with 51% in woodlands < 20 ha (Forestry Commission, 2013) and a little more than half (57%) of these deemed to be under active management (Forestry Commission England, 2014). As forests are such large stores of carbon, the effects of disturbance (such as harvesting) are of considerable interest (e.g. Amiro et al., 2010). If more woodlands are brought back into management and thinning or felling is carried out, then the carbon balance may be affected.

Thinning is a forestry practice that aims to manage competition between trees in order to improve the quality, productivity, yield and form of the final tree crop, and to provide an economic return before final felling. In Britain, two main types of thinning are practised: low thinning and crown thinning, with intermediate thinning a combination of these. In low thinning, suppressed and sub-dominant trees are removed, along with those from the smaller diameter classes, thereby reducing the competition experienced by the larger, more valuable, trees. Crown thinning aims to reduce the competition from other larger trees (dominant and co-dominant). When trees of poorer growth are removed along with some dominant individuals to open the canopy, it can be classed as intermediate thinning (Kerr and Haufe, 2011).

A few studies have considered the impacts of thinning and other aspects of the forest management cycle on forest carbon balances using the eddy covariance technique (EC) (e.g. Vesala et al., 2005; Payeur-Poirier et al., 2011; Saunders et al., 2012). However, it is logistically challenging to manipulate forest stands at the scale required to facilitate EC studies. One approach is to thin the entire forest stand and analyse the pre- and post-thinning phases separately (e.g. Saunders et al., 2012), However, large inter-annual variation in forest C fluxes is common (e.g. Allard et al., 2008; Granier et al., 2008; Wilkinson et al., 2012) which makes unequivocal determination of the effect of thinning difficult from short time series. Alternatively, if only a portion of the forest stand is subjected to the thinning, contemporaneous treatment and control plots are possible, and paired EC systems may be used to detect the fluxes from each section, (e.g. Moreaux et al., 2011), although this approach requires extensive and homogeneous forest areas. For this study, neither of these approaches were available, and so the area

and extent of the thinning operation was deliberately manipulated so that the EC tower was sited near to the line dividing the treatment and control portions of the forest.

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Assessing the impacts of management thinning on the Net Ecosystem Exchange (NEE) of a forest stand is further complicated because NEE is the small difference between ecosystem respiration (Reco) and Gross Primary Productivity (GPP), both of which are much larger components; a small shift in the balance between these will therefore have a large effect on NEE. Furthermore, the ways Reco and GPP are affected by thinning will differ; for example, thinning changes the canopy density, altering the soil temperature and moisture conditions (e.g. Tang et al., 2005; Olajuyigbe et al., 2012) and affecting the soil component of Reco. Vesala et al. (2005) found that whilst there was no reduction in the size of the carbon sink of a boreal Scots pine (Pinus sylvestris L.) stand in Finland following thinning, increases in ground vegetation photosynthesis and heterotrophic respiration were offset by decreases in canopy GPP and in both above- and below-ground autotrophic respiration. Amiro et al. (2010) published a comprehensive study tracking changes in net ecosystem productivity (NEP) across a variety of different forest types following a range of disturbance events. All three conifer forests studied that were subjected to thinning showed relatively short-term impacts on the carbon balance following a decrease in NEP in the year of disturbance. Other studies in managed forests have shown that NEP rates are sustained following thinning of canopy trees (e.g. Granier et al., 2008), which is often attributed to increased growth by sub-canopy plants after dominant canopy trees have been removed (Moreaux et al., 2011; Dore et al., 2012). Many of these studies are concerned with coniferous forests with very different seasonal dynamics to the deciduous oak woodland found in much of lowland England.

Aerialirborne light detection and ranging (LiDAR) is a remote sensing method capable of producing three-dimensional models of large areas of landscape with submetre accuracy and has been used to measure forest height for more than a decade (e.g. Yu et al., 2003). In recent years, its application in forest inventories has become common practice, particularly in northern European countries, where the method is used to quickly cover large areas at a high spatial resolution (Næsset, 2004; Maier et al., 2006). Additionally, the ability to view the resulting data in a variety of ways removes the problems associated with illumination and shadowing seen with standard aerial photography. By carrying out aerialairborne LiDAR surveys before and after a management thinning operation, it is possible to quantify the changes in the forest canopy structure.

- 1 The aim of this study was to examine the effects of management thinning on the factors
- 2 determining the carbon balance of a plantation deciduous oak woodland in southern
- 3 England. Our hypotheses were that the removal of pre-selected trees from the woodland
- 4 during a thinning operation would lead to an initial reduction in GPP. As thinning also
- 5 increases the amount of woody debris and other litter components added to the forest
- 6 floor, an increase in R<sub>eco</sub> was also expected. Together, these changes would result in a
- 7 large decrease of NEE during the immediate period after thinning, which would be
- followed by a recovery of NEE to pre-thin rates over a period of time, possibly several
- 9 years, although we could not predict the timescale.

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## 2 Materials and Instrumentation

# 2.1 Site description

- 13 The eddy covariance measurement site is located in the Straits Inclosure, Alice Holt
- Research Forest, UK (51° 09' N; 0° 51' W), close to the Alice Holt Research Station in
- south-eastern England (Fig. 1a). The inclosure is a flat area with an elevation of 80 m
- above mean sea level; the surrounding landscape consists of mixed lowland woodland
- and both arable and pasture agricultural land. The whole 90 ha inclosure was planted in
- the 1820s with oak (Schlich, 1905) and then replanted in the 1930s. The main tree
- 19 species is Quercus robur L., but other species, including European ash (Fraxinus
- 20 excelsior L.), Q. petraea (Mattuschka) Liebl. and Q. cerris L., are present. There is a
- small area (4.6 ha) of mixed conifers consisting of Corsican pine (*Pinus nigra* subsp
- 22 laricio Maire.) and Scots pine (Pinus sylvestris L.) at the north-west edge of the
- 23 woodland and isolated pockets of Japanese red-cedar (Cryptomeria japonica (L.f.)
- 24 D.Don) are also present in the eastern area. The understorey is dominated by hazel
- 25 (Corylus avellana L.) and hawthorn (Crataegus monogyna Jacq.) (Pitman and
- 26 Broadmeadow, 2001). Prior to this study, the whole of the stand was previously thinned
- 27 in 1995.
- The climate regime is mild temperate oceanic, the long term mean (1971-2000)
- 29 screen annual air temperature was 9.6 °C and the mean annual precipitation 779 mm at
- 30 the UK Meteorological Office affiliated weather station, Alice Holt, Farnham (51° 10'
- N; 0° 51' W), approximately 1.8 km from the measurement site. Further site-specific
- details can be found in Wilkinson et al. (2012).
- Between June and August 2007, the eastern half of the woodland (approx. 47.5
- ha) was selectively thinned (Fig. 1a) using an 'intermediate' thinning procedure, (see

1 introduction) resulting in an open forest canopy with a uniform stand structure (Kerr 2 and Haufe, 2011). Pre-selected trees (based on stem form and position within the 3 canopy) were felled, de-limbed and sectioned using mechanical harvesters. The 4 merchantable stem wood with a diameter > 7.0 cm was subsequently collected and transported to the forest roadside using a forwarder, before being removed from the 5 forest by timber haulage lorries. This harvesting technique resulted in substantial 6 7 disturbance to the understorey and shrub layer. Whilst all of the remaining woody debris was left on the site, some of it was collected and used to construct 'brash mats' for 8 9 machinery movement in order to minimise damage and compaction to the soil, 10 especially in areas of heavy traffic. Mensuration surveys carried out after the thinning 11 in 2009 (western sectorhalf) and 2011 (eastern sectorhalf) showed 453 and 354 trees ha<sup>-1</sup> respectively, a difference in stand density of approx 22% (Table 1). 12

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# 2.2 Micrometeorological measurements and flux calculations

Eddy covariance (EC) measurements of energy flux (sensible and latent heat), momentum, net ecosystem exchange (NEE) and water vapour flux have been made above the forest canopy at the site since 1998. The flux tower is located close to the boundary of the thinned and un-thinned sectors (Fig. 1a). The EC instrumentation consisted of a three-dimensional sonic anemometer (model Solent R2 until September 2011, model Solent R3 thereafter, Gill Instruments, Lymington, UK) and a closed-path infrared CO<sub>2</sub> and H<sub>2</sub>O analyser (model LI-6262 until October 2005, model LI-7000 thereafter, LI-COR Biosciences, Lincoln, Nebraska, USA), sampling air at 28 m height. Raw high frequency data (20.8 Hz) were logged using the Edisol software package (Moncrieff et al., 1997). Further details of the instrumentation can be found in Wilkinson et al. (2012). For that previous paper, post processing of the raw high frequency data was performed using the Edinburgh University micrometeorological software tool EdiRe (http://www.geos.ed.ac.uk/abs/research/micromet/EdiRe/); here we used the EddyPro software package (Ver 4.2.1, LI-COR Biosciences, Lincoln, Nebraska, USA) but with similar processing options. Angle of attack correction (specific to Gill anemometers) was applied according to Nakai et al. (2006). Double axis rotation tilt correction was also applied to ensure that the vertical velocity signal was orthogonal to the plane of mean air flow. The lag time of the sample from the intake point to the measurement cell of the infrared analyser was determined by maximising the covariance between the vertical wind velocity and scalar concentration. In order to

account for flux loss caused by signal damping inside the tube, limited time response and sensor separation, etc, spectral corrections were applied using the fully analytical approach of Moncrieff et al. (1997). Following an analysis of night-time NEE dependence on friction velocity using the method described by Papale et al. (2006), night-time NEE data were rejected where friction velocity was less than a critical threshold (supplementary table 1). 0.2 m s<sup>-1</sup>, following an analysis of night-time NEE dependence on friction velocity. Since CO<sub>2</sub> profile data were not available for the entire measurement period, we have made no corrections for CO<sub>2</sub> storage below the EC instruments. Footprint analysis was performed based on the flux footprint model of Kljun et al. (2004) and the half-hourly flux measurements were rejected when more than 10% of the measured flux was derived from outside the woodland, our area of interest.

# 2.3 Flux data processing and treatment separation

Following the calculation of corrected NEE and in order to remove extreme spikes, which were assumed not to be biologically valid, a data filter was applied using an approach similar to that proposed by Papale et al. (2006) and Thomas et al. (2011). For each calendar year, NEE data were first split into positive or negative values. Positive values more than the mean positive value for the whole year plus three standard deviations were removed and the same approach applied to all negative values. A secondary stage data filter was subsequently applied, which removed positive values more than the mean monthly value for that half hourly period plus three standard deviations, and negative values less than the mean monthly value minus three deviations.

Thirty minute average flux data (including additional meteorological data such as air temperature, humidity and incident solar radiation  $(S_g)$ ) were separated into two sectors according to wind direction: data that were collected when the wind direction was between 315° and 170° were classified as 'east sector' (the area that was thinned in 2007), and data collected when the wind direction was between 170° and 315° were classified as 'west sector' (unthinned area). Table 2 summarises the data availability after this classification into the two sectors. Thirty minute NEE fluxes for both sectors were subsequently gap filled using the marginal distribution sampling method Reichstein et al., 2005) to allow the calculation of annual totals, and partitioned into the component fluxes of Ecosystem respiration ( $R_{eco}$ ) was calculated for each sector using the method proposed by Reichstein et al. (2005). Here each dataset was split into ten-

- day consecutive periods and  $R_{eco}$  was estimated using the Lloyd-Taylor regression
- 2 model (Lloyd and Taylor, 1994) between night time CO<sub>2</sub> flux (global solar radiation <
- 3 20 W m<sup>-2</sup>) and air temperature. The estimated value of  $R_{eco}$  was then assigned to the
- 4 central time point of the averaging interval and linearly interpolated between time
- 5 points.

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- 6 Reco and GPP using the on-line eddy covariance gap filling and
- 7 partitioning tool (http://www.bgc-
- 8 jena.mpg.de/~MDlwork/eddyproc/index.php). In brief, this uses the night-
- 9 time fluxes and their response to temperature to estimate respiration
- 10 during the day.

# 2.4 Model parameters

- In order to examine changes in the physiological drivers of the carbon balance,
- original quality controlled, but not gap filled or flux partitioned) daytime and night-
- time 30 minute average NEE data were separated and analysed independently. The
- temperature sensitivity of ecosystem respiration for each sector of the forest was
- determined using an exponential equation fitted to the average half hourly night-time
- 17 NEE and air temperature for each corresponding period:

$$R_{s} = K_{1} \exp(K_{2}T_{air})$$
 (1)

- where  $R_s$  is the night-time NEE and  $T_{air}$  is the night-time air temperature at 26 m. Data
- 20 fitted to this function were limited to night-time condition only where the mean half
- hourly  $Sg < 20 \text{ W m}^{-2}$  and the quality control flag calculated by EddyPro according to
- 22 the Mauder and Foken (2004) method was equal to zero.
- The relationship between summer (July and August) daytime NEE and  $S_g$  was modelled
- 24 using a rectangular-hyperbolic function:

25 NEE = 
$$\left[\frac{\left(\varepsilon \cdot F^{\infty} \cdot S_{g}\right)}{\left(\varepsilon \cdot S_{g} + F^{\infty}\right)}\right] + R_{d}$$
 (2)

- Where  $F^{\infty}$  NEE<sub>max</sub> is the asymptotic *net* CO<sub>2</sub> assimilation rate,  $\varepsilon$  is the incident quantum
- 27 yield (initial slope of the light response curve), and R<sub>d</sub> is respiration in the dark. Data
- 28 fitted to the light response model were limited to periods where the mean half hourly
- 29  $Sg > 20 \text{ W m}^{-2}$ , air temperature was between 18–23 °C and the quality control flag
- 30 (Mauder and Foken, 2004) was equal to zero.

# 2.5 LiDAR measurements and calculation of vegetation structure

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The aerial photograph taken after thinning (Fig. 1a) and mensuration surveys revealed substantial spatial heterogeneity within the forest block and showed large differences in forest structure between the two sectors. Changes in canopy top height and gap fraction were assessed using aerialairborne LiDAR surveys conducted over two flight campaigns for the whole of Alice Holt forest (800 ha) by the Unit for Landscape Modelling (ULM), (Dept. of Geography, University of Cambridge). The first was in early November 2006, prior to the thinning and the second in August 2009, two years after the thinning. Due to the mild autumn in 2006 both surveys were completed whilst the forest had a fully developed canopy. A LiDAR system (ALTM 3033, Optech Incorporated, Ontario, Canada) flown at an altitude of 1000 m above ground level and with a scan angle +/- 15° was used along a series of overlapping transects designed to cover the whole forest. The system combined a pulse rate of 33 kHz and an overlap of 50% between swaths, resulting in a point density of 2 to 4 points m<sup>-2</sup>, which was used to generate a virtual cloud of 3D data points with an accuracy of +/- 15 cm root mean square (RMS). The first and last pulse return data were used to generate a Digital Surface Model (DSM) which included the tree cover and a Digital Terrain Model (DTM) representing the ground surface. These data were provided by the ULM as raster elevation models with a 0.5 m cell size. By subtracting the DTM from the DSM using GIS software (ArcGIS 10, Esri, Redlands, California, USA), Canopy Height Models (CHM) for each survey were created. Furthermore, by subtracting the 2006 DSM from the 2009 DSM, a model of change between the two surveys was also created (Fig. 1b). To allow a detailed analysis of the vertical change in forest height and gap frequency between 2006 and 2009, each CHM was converted to a 1m cell size and then spatially split into a 1 ha grid. Canopy top height histograms (bin size = 50 cm) were calculated for each grid cell, based on the 10,000 values per hectare. Frequencies were then averaged for all the grid cells within each sector. Grid cells at the interface between the eastthinned and westunthinned areas of the forest were excluded from the analysis, as were those cells that contained, either wholly or partially, areas of the surrounding agricultural land. All analyses were conducted using R software (R Development Core Team, 2011).

### 3 Results

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# 3.1 Climatic conditions

- The prevailing wind direction at the site is from the south west, so more of the data come from periods when the wind is from the west-(unthinned) sector (Table 2). As the meteorological conditions associated with easterly and westerly winds differ, the flux data recorded from the two sectors did not reflect the same meteorological conditions (Fig. 2). Mean canopy level annual air temperature (2004-2012) was slightly warmer when air flow was from the west sector (10.8 °C) than from the east sector (9.6 °C).
- 9 The mean diurnal course of air temperature over the bi-monthly winter periods of November - December and January - February were generally warmer when airflow 10 was from the more usual west rather than the east. The largest difference in winter air 11 12 temperature was observed in January - February 2012: a period of cold weather from 13 the start of February, dominated by easterly conditions, persisted over the southern UK 14 for about two weeks this was and was also associated with snowfall to parts of the 15 region (Fig. 3a). Conversely, mean summer air temperatures (during daylight hours) 16 were generally higher when airflow was from the east than from the west, as occurred 17 in 2004. Incident solar radiation and relative humidity were generally very similar when 18 air flow was from either sector (Fig. 2).

# 3.2 Variation in NEE and Reco

generally acted as a CO<sub>2</sub> source for the bi-monthly periods of November - December,
January - February and March - April, although in exceptionally warm and early springs
such as 2007 the forest became a weak CO<sub>2</sub> sink for a few hours around noon. Both
sectors of the forest were a strong CO<sub>2</sub> sink from May through to October, although

The mean (not gap filled) diurnal course of NEE (Fig. 3b) indicated that the forest

- 25 there was considerable variation between years. In some periods, when temperature and
- 26 insolation conditions were very similar for each sector, NEE patterns were also similar
- 27 (e.g. May June 2007, and March April, 2012). In other periods with similar
- temperature and insolation, NEE was different, for example, in July August 2012.
- 29 The annual gap-filled totals of NEP and its component partitioned fluxes
- 30 (Reco and GPP) were calculated for both sectors of the forest (Fig. 4).
- 31 Before the thinning the annual GPP was slightly higher (+6%) in fluxes
- 32 from the west than those from the east sector (mean difference 2004-2007

= 121.3 q C m<sup>-2</sup> v<sup>-1</sup>; SE 33.2). After the thinning, the difference increased (mean difference 2008-2012 = 238 g C m<sup>-2</sup> y<sup>-1</sup>; SE 88.6) and annual GPP 2 was higher in the west sector in all years, apart from 2009. Heavy 3 defoliation by caterpillars (Wilkinson et al., 2012) occurred in 2010 4 reducing GPP in both sectors of the forest, but particularly in the east. 5 Prior to thinning the annual Reco was similar in both sectors of the forest 6 (mean difference 2004-2007 = 26.3 g C  $m^{-2}$   $v^{-1}$ ; SE 58.1). In the first year after thinning (2008) Reco was slightly lower (-12%) in the east than the 8 9 west however, a year later, in 2009, Reco increased substantially (+34%) in the east compared to the west, but in subsequent years it dropped in the 10 11 east (mean difference 2010-2012 = 80.0 g C m<sup>-2</sup> y<sup>-1</sup>; SE 56.0).

NEP was also generally larger in the west sector than the east prior to the 12 thinning (mean difference 2004-2007 = 95.3 g C m<sup>-2</sup> y<sup>-1</sup>; SE 36.09). In 2007, 13 14 the year of the thinning, NEP was substantially higher for both sectors, 15 which was a result of the weather in that year, with a longer growing season (19 days longer than the average) and high peak LAI (see 16 Wilkinson et al., 2012). From 2009 onwards there was a substantial 17 increase in the ratio of Reco to GPP in the east which was not evident in 18 the west sector (Fig. 4d), resulting in much larger differences in NEP 19 20 between the two sectors after thinning (mean difference 2009-2012 = 318.3 g C m<sup>-2</sup> y<sup>-1</sup>; SE 39.5). 21

# 3.3 Effects of thinning on ecosystem respiration

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23 As expected for a temperate, deciduous forest, there was a large annual cycle in R<sub>eco</sub>, 24 with a peak in May - August, (Fig. 4a - 4c) but varying substantially year to year. with 25 weather conditions, particularly precipitation and temperature. Before thinning, annual 26 R<sub>eco</sub> patterns were similar between sectors (e.g. Fig. 4a, 2006) but in the immediate 27 period after thinning R<sub>eco</sub> was usually higher in the east sector (e.g. Fig. 4b, 2009), particularly in the warmer summer period. As weather conditions differed for fluxes 28 29 measured for east and west, we compared the underlying relationships of Reco with 30 temperature between sectors. 31

As an assessment of the sensitivity of  $R_s$  to air temperature using the coefficients of the exponential function (Eq.(1)) revealed important differences between sectors (Table 3). Overall  $Q_{10}$  was generally higher and more variable between years, when airflow was from the west sector (mean = 2.924.14; SD = 0.741.42) than from the east

- (mean = 2.0841; SD = 0.2351) however this was not the case there were deviations
- 2 from this most notably in 2009 and 2010. The largest differences in R<sub>s</sub> (highest in the
- 3 east) between the two sectors occurred in 2009, two years after the thinning. This was
- 4 the only year during which there was a constant *small (although non-signifiant)* positive
- offset in the sensitivity of  $R_s$  to air temperature between the two sectors (Fig. 4h).

# 3.4 Effects of thinning on canopy NEE light response

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7 The asymptotic net CO<sub>2</sub> assimilation rate (F<sup>∞</sup>) The maximum rate of light saturated 8 photosynthesis (NEE<sub>max</sub>) and apparent incident quantum yield ( $\varepsilon$ ) were determined from 9 a light response function (Eq.(2)) fitted to the summer (July and August) daytime NEE 10 flux data for both forest sectors (Fig. 6). in a 5 °C range (18 - 23 °C). Differences in NEE<sub>max</sub>  $F^{\infty}$  were observed between the east and west sectors both before and after 11 12 thinning (data not shown). Although both sectors followed the same general inter-13 annual pattern, there was no clear change (in either sector) after thinning. The 14 magnitude values of F<sup>\*</sup>NEE<sub>max</sub> wasere generally larger (more negative) than the 15 maximum observed rates of daytime NEE, due to an over estimation of F NEE by the rectangular-hyperbolic model, therefore NEE at  $S_g$  800 Wm<sup>-2</sup> (NEE<sub>800</sub>) was 16 17 considered a better indication of the maximum rate of light saturated NEE. NEE<sub>800</sub> was 18 consistently smaller (lowerless negative) in the fluxes observed from the east sector 19 (Fig. 5a) than from the west from the start of the observation period through to 2007, the year of the treatment (mean difference 2004 - 2007 = 3.85 µmol m<sup>-2</sup> s<sup>-1</sup>; sd = 20 21 0.92), for the entire measurement period, but there was no significant reduction in 22 NEE<sub>800</sub> in 2008 in fluxes from either sector. Prior to 2007, the magnitude of There was 23 a small increase (2.2 and 2.1 µmol m<sup>-2</sup>s<sup>-1</sup>) in NEE<sub>800</sub> in the east relative to the west in 2009 and 2010 respectively. In 2011 and 2012 the two sectors returned to their pre-24 25 thinning ranking. Incident apparent quantum yield (Fig. 5b) was generally higher much 26 larger (more negative) when fluxes were from the west than from the east; -the(except 27 for 2007 & 2012) indicating that this area of the forest u two sectors converged in the post thinning phase, sed low radiation levels more efficiently. Although R<sub>d</sub> (respiration 28 in the dark) estimated from the light response curves increased in the east sector post 29 thinning relative to the west and remained higher through to 2012 was usually about 30 half the value of that in the west sector there was substantial inter-year variation and no 31 32 clear change was associated with thinning (Fig. 5c) confirming the results of Reco 33 estimated using the Reichstein et al (2005) method.

# 3.5 Changes in canopy height and gap fraction

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The canopy top height derived from the first return data from the LiDAR survey showed 2 3 that the two sectors of forest had similar canopy height distributions in 2006, before 4 thinning (Fig. 6a & b), but with some differences in detail. The small peak in frequency 5 between 5 and 10m height in the west in 2006 (Fig. 6a) is from areas of the forest which were undergoing succession development following previous disturbance events. By 6 7 2009 these areas of the forest had grown and are evident as heterogeneous patches in Fig. 1b. In both sectors, the canopy height distribution profile changed, in the west this 8 9 was because of growth, whilst in the east the thinning operation had a substantial effect. 10

Prior to the thinning both maximum and mean canopy heights were similar in both sectors (Table 4). Between 2006 and 2009, the maximum canopy height increased in the west sector by 0.9 m, but was reduced slightly in the east sector by 0.1 m. Over the same time period mean canopy top height also increased in the west sector by 0.95 m and reduced in the east sector by 1.4m.

Changes in the canopy height distribution profiles were also observed (Fig. 6c & d). The elevation relief ratio, E (Pike and Wilson, 1971) reflects the degree to which the outer canopy surfaces are in the upper (E > 0.5) or lower (E < 0.5) portion of the height range is defined as:

$$E = \frac{h_{\text{mean}} - h_{\text{min}}}{h_{\text{max}} - h_{\text{min}}}$$
(3)

where  $h_{\text{mean}}$ ,  $h_{\text{min}}$ , and  $h_{\text{max}}$  are the mean, minimum and maximum canopy heights 20 21 respectively. E was reduced substantially in the east because of the larger proportion of 22 lower top heights, while there was only a small increase in E in the west (Table 4). The canopy top height distribution also showed a relatively small increase in the proportion 23 24 of canopy > 15 m in height between 2006 and 2009 in the west (+6.2%) but a substantial 25 reduction in the east (-13.7%) as a result of the thinning operation. 26 The LiDAR survey also showed that canopy complexity across the upper-most surface 27 of the forest in the east sector increased following the thinning operations. The relative 28 variability in canopy height (indicated by the coefficient of variation) increased 29 substantially (Table 4) in the east but not in the west. After thinning there was a large 30 increase in the frequency of gaps in the forest canopy (canopy top height < 1 m) in the 31 east sector but not the west, because of the removal of canopy trees (compare Fig. 6d). 32 Gaps in the forest canopy were relatively uniformly distributed throughout the whole east sector and increased from a total area of 1.13 ha (3.1 % of the eastern area) in 2006 33 34 to 2.16 ha (6.6 %) in 2009. Over the same period there was a small decrease in the total

- area of gaps in the forest canopy in the west, which measured 0.89 ha (2.47% of the
- 2 total western area) in 2006 and 0.85 ha (2.35%) in the 2009 surveys (Fig. 1b).

### 4. Discussion

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4 Surprisingly, effects of the thinning procedure in 2007 on carbon balance were not 5 clearly evident. In part, this might have been because of our experimental approach. We 6 used eddy covariance measurements at one location near the boundary between the thinned and unthinned sectors in order to determine the CO2 fluxes, because of the 7 8 relatively small size of the forest block and being restricted to only one tower and EC 9 system. The effects of thinning are partly obscured by: a) the differences in weather 10 conditions when airflow is from either sector (Fig. 2); b) existing heterogeneity in fluxes 11 from different parts of the forest prior to thinning (Fig. 3b) e) the substantial inter-12 annual variation in component CO<sub>2</sub> fluxes (evident in Fig. 4); and d) the limited data 13 availability for each sector, which after classification and quality control can be as low 14 as 22.5% (Table 2). However there was no evident bias in the data availability (e.g. day or night, or seasonal distribution) that might have affected our gap-filled annual 15 16 total CO2-flux component estimates. The main difference in CO2 flux between the two 17 sectors after thinning was substantially lower NEE in fluxes from the east (Fig. 4c) 18 although this was only observed from 2009 onwards.

The pre- and post-thinning LiDAR surveys indicated that whilst canopy top height distributions were comparable in 2006, the thinning operations in 2007 had a large effect on the canopy structure of the east sector, resulting in a more complex canopy with a wider range of top heights and a larger total area of gaps. The complexity of the forest canopy at our site, as a result of variability in gaps and a dense understorey, contrasts with other published studies using LiDAR at other deciduous forest sites (Wasser et al., 2013). Whilst we acknowledge that the 2009 LiDAR survey was not immediately after the thinning, our estimate of the change in canopy gap fraction may be an under representation. Firstly, LiDAR pulses have a relatively large footprint (~25 cm in diameter) and therefore gaps in the canopy would need to be larger than this in order to be recognised as a gap. Secondly, off-nadir pulses are more likely to produce a canopy height return than they are to penetrate to ground level. Our approach used only the first and last return signals of the LiDAR data, so the canopy height model showed only the uppermost component of the forest canopy. As such, some of the changes in the understorey canopy during thinning may have been masked by the vertical overlap of the understorey vegetation and upper canopy. Whilst we

acknowledge that an analysis of full wave-form or multiple return data (Mallet and Bretar, 2009) may provide more detailed information about the canopy's 3D structure, we maintain that the approach adopted here provided a useful assessment of the changes to the forest canopy due to the thinning operations.

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The parameters obtained from the summer light response curves did not support our hypothesis that tree thinning would lead to a reduction in NEE through a loss of canopy photosynthetic area., although there is substantial variation in the data and hence uncertainty in the fitted parameters. Contrary to expectation, there was no clear difference in NEE<sub>800</sub> (Fig.5) in 2008 for the east sector relative to the west. We suggest that this apparent insensitivity in 2008 to the thinning indicates that in the first year after thinning the newly exposed ground vegetation and shrub layer, and better illumination of the remaining crowns compensates for the removed trees. From 2009 to 2011, NEE<sub>800</sub> was reduced in both sectors *probably* as a result of defoliation by caterpillars (Wilkinson et al., 2012). The increase in  $\varepsilon$  (more negative) in the east in both sectors in 2008 and especially 2009 may be as a result of the thinning is also likely to be an effect of defoliation by caterpillars, possibly masking the effects of the tree thinning but as it consistent with the earlier work of Niinemets (2007) and Pangle et al. (2009) who demonstrated that as forest canopies become more structurally diverse, light efficiency increases because of a more even distribution of radiation throughout the tree canopy and better light penetration to sub-canopy species with a higher ε. Our findings however contrast with results from thinning studies carried out on evergreen conifer sites (with presumably little or no understorey vegetation). For example, Saunders et al. (2012) attributed observed changes in the photosynthetic efficiency of a Sitka spruce stand following thinning to inherent change in the photosynthetic efficiency of the remaining trees, rather than being due to increased light absorption.

The impacts of thinning on respiration are complicated by the fact that  $R_{eco}$  consists of  $CO_2$  derived from both heterotrophic respiration ( $R_h$ ) largely in the soil and from autotrophic respiration ( $R_a$ ), both above and below ground. Both of these  $CO_2$  sources comprise a number of processes and components which are likely to be influenced by both time and forest management in different ways.  $R_d$  estimated from the light response curves increased in the first years after thinning in the east relative to the west. In the first years' after thinning (2008 - 2010) the initial supply of fine roots, small twigs, leaves and other easily degradable fractions of litter would be a major new source of carbon and nitrogen for the decomposition system. Soil disturbance from machinery might also be expected to increase  $R_h$  as was demonstrated by Concilio et

al. (2005) at a mixed species conifer site. We therefore expected an initial stimulation of R<sub>eee</sub> in the first year after thinning, but this was not observed, which may be because of slow colonisation by microflora or because of the reduction in living biomass (above and below ground) decreasing R<sub>a</sub>. The lack of R<sub>eee</sub> response may also have been because much of the coarse above/*ln addition, much of the large-ground* woody debris had been gathered together to form brash mats and which may have been a substantial source of CO<sub>2</sub>, although we have no independent measurements of emission from them. therefore was not in direct contact with the soil surface and more likely to dry out, as well as reducing its availability to soil microbes. Thinning is also likely to cause local increases in temperature, increased throughfall, reductions in humidity and probably higher evaporation rates in gaps (Vesala et al., 2005), which may affect decomposition. However, we cannot quantify such effects as the climatic data we recorded was only that from the central instrument tower.

The low NEE in fluxes from the east after 2009 were consistent with a step-wise increase in the ratio of Reco to GPP (Fig. 4e). After thinning there is likely to be a succession of changes in the relative contributions of R<sub>a</sub> and R<sub>h</sub> to total R<sub>eco</sub>, which may be associated not only with changes to soil conditions but also with biomass removal (Anderson-Teixeira et al., 2011) and a reduction in GPP (Woodward et al., 2010). Although we do not have independent measures for R<sub>a</sub> and R<sub>h</sub> throughout the period of the present study, work at the site in 2008-2010 (Heinemeyer et al., 2012) demonstrated that in an unthinned area the largest proportion of total soil efflux was from R<sub>a</sub> (56%) compared with R<sub>h</sub> (44%). Importantly for this study, Heinemeyer et al. (2012) demonstrated a stronger temperature response for R<sub>h</sub> than for either R<sub>a</sub> (roots) or R<sub>a</sub> (mycorrhizae). After thinning the proportion of total soil CO<sub>2</sub> efflux derived from R<sub>h</sub> is likely to increase, which may result in an increased temperature sensitivity of CO<sub>2</sub> efflux by forest soils. In 2011 From 2010 onwards there was no clearly discernable difference in R<sub>eco</sub> between the two sectors, and we therefore assume that any increase in below ground R<sub>h</sub> is likely to be cancelled out by a corresponding reduction in R<sub>a</sub>, which is consistent with the findings of Tang et al. (2005).

In a previous paper describing the pattern of CO<sub>2</sub> fluxes at this site between 1999 and 2010 (Wilkinson et al., 2012) we noted the substantial inter-annual variation in NEE. The analysis presented here (e.g. Fig. 3) suggest that part of that may be caused by inter-annual differences in the contribution from the east and west areas of the forest, which differed even before the thin.

## 5. Conclusion

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This study has investigated the effects of management thinning on the carbon balance 2 3 of deciduous oak woodland in south-eastern England. LiDAR data were used to assess changes in the forest canopy, while EC was used to measure changes in the carbon 4 5 balance. Management thinning reduced the mean canopy top height and resulted in a forest canopy with a wider top height range and more gaps. The impacts of management 6 7 thinning on the carbon balance were not clearly evident although ecosystem respiration 8 was higher NEE was lower in fluxes from the east sectorthinned area from 20089 onwards, onwards and remained higher until the end of the study period. two years after 9 10 the thinning. The insensitivity of the summer photosynthetic parameters in the first year 11 after thinning, 2008, suggests that newly exposed ground vegetation and shrub layers receiving better illumination compensated for the removed trees. The Reco component 12 13 increased in the thinned area but not until two years after the thinning had been 14 completed and was associated with an increase in the sensitivity of Reco to temperature.

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- 1 Table 1.
- 2 Results of tree mensuration surveys carried out in 2009 (west sector) and 2011 (east
- 3 sector) at the Straits Inclosure, Alice Holt Forest. In the east sector 26 circular plots
- 4 were measured each with a radius of 12.6 m, whilst in the west sector, 18 plots were
- 5 measured each with a plot radius of 8 m. Plots locations were selected using a stratified
- 6 grid basis to ensure the heterogeneity of the forest structure was measured; figures in
- 7 brackets are standard error.

	All trees		Oak trees only	
	Density	Mean diameter	Density	Mean diameter
	(trees ha <sup>-1</sup> )	at breast height	(trees ha <sup>-1</sup> )	at breast height
		(cm)		(cm)
East	354	23.9 (0.55)	217	30.0 (0.53)
West	450	26.6 (0.57)	423	26.8 (0.57)

- 1 Table 2.
- 2 Annual eddy covariance CO<sub>2</sub> flux data capture and quality controlled data availability
- 3 following de-spiking, footprint and u\* quality checks for each sector by time of day (all
- 4 in %) over the period 2004 2012 at the Straits Inclosure, Alice Holt Forest.

Year	Total data	QC East	QC East	QC West	QC West
	Capture	day	night	day	night
2004	79.6	8.9	6.1	18.7	14.7
2005	92.5	11.6	8.3	21.6	16.3
2006	74.3	10.7	8.6	16.6	11.4
2007	92.5	9.9	6.3	18.6	12.8
2008	81.4	10.7	6.4	26.0	21.o
2009	77.3	11.9	10.0	15.7	12.9
2010	93.0	15.9	10.5	18.9	14.1
2011	86.7	12.6	9.5	18.4	15.1
2012	82.1	11.9	8.6	18.0	14.9
Mean	84.4	11.6	8.3	19.2	14.8

- 1 Table 3.
- Night-time ecosystem respiration ( $R_s$ ) coefficients and the estimated  $Q_{10}$  values (base
- 3 temperature = 0°C) derived from fitting an exponential equation to half hourly night-
- 4 time NEE and air temperature vales over the period 2004 2012 at the Straits Inclosure,
- 5 Alice Holt Forest.

Year	K1 East	K2 East	K1 West	K2 West	Q10 East	Q10 west
2004	2.22 (0.07)	0.064 (0.003)	1.22 (0.06)	0.120 (0.003)	1.90	3.32
2005	2.14 (0.06)	0.063 (0.002)	1.59 (0.07)	0.091 (0.003)	1.88	2.48
2006	1.82 (0.07)	0.068 (0.003)	1.67 (0.08)	0.082 (0.003)	1.97	2.27
2007	2.08 (0.10)	0.061 (0.004)	1.11 (0.06)	0.122 (0.004)	1.84	3.39
2008	1.82 (0.07)	0.078 (0.003)	0.81 (0.04)	0.140 (0.003)	2.18	4.06
2009	1.71 (0.07)	0.089 (0.003)	1.37 (0.06)	0.089 (0.089)	2.44	2.44
2010	1.70 (0.05)	0.072 (0.002)	1.74 (0.05)	0.064 (0.002)	2.05	1.90
2011	1.62 (0.08)	0.071 (0.004)	1.15 (0.05)	0.098 (0.004)	2.03	2.66
2012	1.72 (0.08)	0.088 (0.004)	0.93 (0.05)	0.134 (0.004)	2.41	3.82

Figures inside brackets are one standard error (SE).

1 Table 4.

5

- 2 Results of aerial LiDAR surveys before and after thinning calculated from first and last
- 3 return data at a point density of 2 points m<sup>-2</sup> and extracted from a 1ha gridded canopy
- 4 height model at the Straits Inclosure, Alice Holt Forest

Year	Sector	Maximum height (m)	Mean height (m)	S.D of mean height	C.V.	Elevation relief ratio (E)	% of canopy > 10m	% of canopy > 15m
2006	West	25.7	15.0	5.04	0.34	0.58	81.9	66.3
2006	East	26.0	15.0	5.03	0.34	0.57	84.5	65.3
2009	West	26.6	15.9	4.99	0.32	0.59	85.8	72.5
2009	East	25.9	13.6	6.19	0.46	0.52	73.6	51.7

# Figure captions

2 3	Figure 1a. Aerial photograph (taken in spring 2008) of the Straits Inclosure, Alice Holt
4	Forest. © Bluesky International Ltd/Getmapping PLC
5	
6	Figure 1b. Change in canopy height between November 2006 and August 2009
7	calculated using aerial LiDAR data at the Straits Inclosure, Alice Holt Forest.
8	
9	Figure 2. Average bi-monthly values (2004-2012) for the key climatic variables of (a)
10	air temperature (b) incident solar radiation, Sg and (c) relative humidity for the east
11	(blue) and west (green) sectors, error bars represent $\pm$ 1 standard deviation (n=7) at the
12	Straits Inclosure, Alice Holt Forest.
13	
14	Figure 3a. Average bi-monthly diurnal curve of incident solar radiation, $S_g$ for east
15	sector (blue solid line) and west sector (green solid line) and air temperature for east
16	sector (blue open circles) and west sector (green open circles) for 2004, 2007 and 2012
17	at the Straits Inclosure, Alice Holt Forest.
18	
19	Figure 3b. Mean bi-monthly diurnal curve of net ecosystem exchange for east sector
20	(blue solid line) and west sector (green solid line) for 2004, 2007 and 2012, + symbols
21	represent $\pm$ 1SE at the Straits Inclosure, Alice Holt Forest.
22	
23	Figure 4. Monthly estimated R <sub>eco</sub> for the east sector (blue solid line with open circles)
24	and west sector (green solid line with open circles) for (a) 2006 (b) 2009 (c) 2012;
25	monthly mean air temperature (at 26 m height) and monthly precipitation total for (d)
26	2006, (e) 2009 (f) 2012; modelled temperature response (Rs derived from night-time
27	NEE fluxes only) for east sector (blue sold line) and west sector (green solid line) for
28	(g) 2006 (h) 2009 (i) 2012, error bars represent 95% confidence intervals, at the Straits
29	Inclosure, Alice Holt Forest.
30	
31	Figure 5. Summer (July and August) daytime modelled light response for the east
32	sector (blue line) and west sector (green line) for the years of (a) 2006 (b) 2009 (c)
33	201 at the Straits Inclosure, Alice Holt Forest. Data limited to 18 23 °C air
2.4	tonen quetano, con ditions

- Figure 5. Inter-annual variation in summer (July and August) daytime light response
- model parameters for (a) NEE<sub>800</sub> (b)  $\varepsilon$  and (c) Rd for the east sector (blue line with
- open circles) and west sector (green line with open circles) error bars represent  $\pm$  1SE,
- 4 at the Straits Inclosure, Alice Holt Forest.

5

- 6 Figure 6. Histograms of canopy top height (m) derived from *aerialairborne* LiDAR for
- the east sector (blue bars) and west (green bars) for (a) 2006 and 2009 west sectors, (b)
- 8 2006 and 2009 east sectors, cumulative frequency of canopy top height for (c) west
- 9 sector in 2006 & 2009 and (d) east sector in 2006 and 2009 at the Straits Inclosure,
- 10 Alice Holt Forest.

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