

Dear Editor Professor Michael Weintraub,

Please receive our revised version of manuscript bg-2015-458 “Greenhouse gas emissions in natural and agricultural lands in sub-Saharan Africa: synthesis of available data and suggestions for further studies”. We thank you and the reviewers for constructive suggestions on the manuscript. We have addressed each of the comments as outlined below.

Anonymous Referee #1

This article is an interesting, novel review of greenhouse gas (GHG) emissions from natural and agricultural ecosystems in 22 countries in sub-Saharan Africa, compiling published data on CO₂, CH₄, and N₂O emissions. The authors summarize knowledge of the baseline (current) emissions from this region. They report measured emissions from a range of different ecosystem and land use types, and management practices. The variability in measured emissions is large, and the authors highlight important research gaps and the need for further studies to elucidate environmental and management drivers of emissions at multiple spatial and temporal scales. This paper fills an important knowledge gap. However, I think the authors could improve upon several aspects of their review.

Both the results and summary sections might be improved by including a framework that organizes or summarizes the suite of complex direct (e.g., oxygen and carbon availability) and indirect (e.g., root and microbial respiration, soil texture, temperature) controls on emissions across the studies and how those controls are affected by management (e.g., tillage, fertility inputs) and ecosystem state factors (e.g., parent material, climate, vegetation).

Related to this, the review should also include more synthesis, if possible, such as quantitatively summarizing findings regarding controls across studies.

As currently written, the results read as an inventory or list of emissions rates and key findings from individual studies (rather than a “synthesis,” which is in the title), depending on which factors individual studies addressed (e.g., temperature, moisture, vegetation type, pH, dynamics of C and N availability, etc.). The current presentation of results makes it difficult to discern

- on average or in aggregate for different ecosystem types or management systems
- the state of knowledge regarding relative importance of different drivers of variation.

Statistical analysis was performed on agricultural studies to fit models for emissions as a function of N inputs. I wonder what additional statistics might be performed on these data to understand the aggregate effect of controls on emissions rates across multiple studies or ecosystem types (i.e., how emissions vary with these different factors)? Are there consistent effects of soil texture across the studies? Such information (if available) would better direct future research efforts. For example, the authors could highlight whether more is known about some controls than others, or if there is a lack of information about interactions between different controls, etc. It seems that a key point from the findings is that there is a need for more studies that address questions about how interactions between management

(fertility practices, tillage) and environment (soil texture and type, etc.) drive GHG emissions, but this discussion could be strengthened.

Response:

There are fundamental challenges to address the comments due to lack of data in general and poor data quality specifically. For instance, few studies report GHG fluxes with respect to the research questions described above (e.g., mechanistic controls), or with suitable experimental designs (e.g., adequate replication). Therefore, it was difficult to synthesize beyond describing their findings relevant to key topical areas. Furthermore, available data was not a large enough to conduct valid statistical analysis. With the only exception of soil CO₂ fluxes, with which we were able to provide new statistical results as described below. Despite the difficulties, we have made significant efforts throughout the manuscript to improve the synthetic contribution of our effort to better describe and understand GHG emissions, mitigation potential and future challenges in SSA. Major changes were summarized as below:

1) New statistical results were added (see 3. 1. 1 CO₂ emissions) to show that observed annual soil CO₂ emissions in African natural terrestrial systems and agricultural lands had significant correlations with annual mean air temperature, annual rainfall, soil organic carbon and total nitrogen contents. Accordingly, Table 2 and relevant discussion were added.

2) We altered the previous descriptive list to a more thematically synthesized approach throughout Results and Discussion sections: the sub-title of the second section changed to 'Sources and drivers of greenhouse gas emissions in Africa' accordingly.

3) Summary of GHG emissions section (newly named as '3.1. Summary of greenhouse gas emissions in Africa') was revised and is now located right up front, first in Results and Discussion.

Second, more attention should be given to the disparate methods within the studies. The authors are clear that they selected only in situ studies, but then note that a wide range of methods was used in the studies they synthesize. Could this be accounted for in some way in the analysis (e.g., analyze emissions by measurement method)? Are some of the results presented likely more robust than others? More information could be added to the supplementary tables; for example, duration of the study (whether emissions were measured for one year, one growing season, multiple years, etc.), frequency of sampling events within a year, capturing major weather events, etc. Were any of the measurements for agricultural systems on actual farms, or were they in experiment stations? A methods column could also potentially list chamber type or other relevant information.

Response:

We newly assessed data quality of the cited studies using the criteria suggested by Rochette and Eriksen-Hamel (2008) and Barton et al. (2015). We categorized the studies as three different groups: the methods are 1) *poor to very poor*, 2) *marginal*

and 3) *good*. We newly added detail procedure of the assessment (see 2.1. Data collection), results (see 3.1.5 Data quality assessment) and discussion (see 3.3 Suggested future research) in the manuscript. We have recorded the assessment results in Supplementary Information Table S1 and S2.

If a paper provided detailed information on the method of gas collection, study periods and frequency, weather characteristics, and other environmental factors (soil, vegetation, management) then the information was recorded in the supplementary database (see Appendix A). However, few studies report detailed information, so it was not possible to analyze available data by measurement methods, frequency or periods as the reviewer suggested.

Third, the overall coherence would be improved by stronger links to theory, and by including broader discussion/interpretation of the summarized findings. For example, the authors could draw upon N saturation theory from N deposition studies in forest ecosystems (N surplus is mentioned in the discussion on page 16496, but might be better mentioned up front as a guiding framework for understanding a key driver of losses in systems with N inputs, and then woven throughout). For example, the finding that N₂O emissions increased exponentially when fertilizer applications exceed plant uptake (for the very high rates) is in line with N saturation theory. Another option is to link findings to an ecological nutrient management framework in the agriculture section, which aims to couple C and N cycles (e.g., by adding a C source such as a cover crop together with an N source, or using organic N sources) to reduce N surplus and balance N inputs with harvested exports.

Response:

To link collected data to relevant theories,

1. We newly conducted correlation analysis and found observed annual soil CO₂ emissions had significant correlations with annual mean air temperature, annual rainfall, soil organic carbon and total nitrogen contents. We found an unexpected result showing negative relation between annual soil CO₂ emissions and annual mean air temperature. We discussed the results based on theories on drought and water stress effects on carbon balance and ecosystem production as stated below (see 3. 1. 1 CO₂ emissions):

"We speculate that the generally high temperatures, and poor quality, of many African soils mean that air temperature increases frequently result in vegetation stress and/or soil aridity, hindering root and soil microbial activities (root and microbial respiration) and subsequent soil CO₂ flux (e.g., Thomas et al., 2011)"

2. We have provided two new insights in the paper related to N saturation as pointed out in the following lines:

First, we found relationship between nitrogen (N) input and nitrous oxide (N₂O) emissions observed in Africa (Figure 3.). Second, we found relationship between nitrogen (N) input and yield scaled nitrous oxide (N₂O) emissions (Figure 4).

3. To link ecological nutrient management framework in the agriculture section, we newly added the below text (see Agroforestry in 3.2.2. Greenhouse gas emissions from agricultural lands).

"Therefore, there may be potential to reduce N₂O emissions in the agroforestry practice, but it may require ecological nutrient management (i.e., reduced inorganic fertilizer N inputs accounting N input from the legume trees; adding a C source such as a cover crop together with an N source) and rotation planning."

Finally, the paper would improve with brief discussion throughout regarding why and how reported emissions for the different ecosystems might matter for current sustainability concerns, particularly regarding land use change. Linking emissions rates to crop productivity (the yield-scaled results) is an important start, but what other trade-offs are there? Vegetable systems with high emissions, for example, are likely a small proportion of total land use, and may contribute high nutritional value per area. Table 2 with the impacts of different management practices gets at this, but it would be useful to identify some potential tradeoffs more generally and better synthesize across studies.

Response:

We recognized the importance of this synthesis to speak to sustainability concerns, especially land use changes issues, we discuss sustainability issues in '3.2.3 Greenhouse gas emissions from land use change; and '3.3 Suggested future research'. We also added the below sentences in '3.3 Suggested future research'.

"Throughout the study, we identified various trade-offs including increased CO₂ emission following forest thinning management, increased GHG emissions in land-use changes, very high N₂O emissions in vegetable gardens due to excessive N input to get high yields, increased CO₂ and N₂O emission in incorporation of crop residues to the soil and agroforestry practices, and exponential increased of N₂O emission and yield-scaled N₂O emissions in excessive N input. Further studies are needed to assess and manage potential trade-offs and drivers."

Specific Comments

Page 16483, lines 7-13: How do these numbers compare to countries or regions with highly industrialized agricultural systems and higher average N fertilizer rates? This would help to place these figures in a broader context.

Response:

We revised the mentioned sentences accordingly:

"According to Lassaletta et al. (2014), mean N application rates in Africa were 34 kg N ha⁻¹ in 2009 and only 16 kg N ha⁻¹ in sub-Saharan African countries while the rate was 169.1 kg N ha⁻¹ in 2009 in the USA. Only Mauritius, Botswana and South Africa

had average N application rates exceeding 100 kg N ha⁻¹. Even with the low fertilizer rates used across the continent, agricultural GHG emissions in Africa are substantial; amounting to 26% of the continent's total GHG emissions (Valentini et al., 2014) while agricultural GHG emissions were responsible for 8.4% of total GHG emissions in the USA (US EPA, 2016)."

Page 16484, line 9: How many total papers did the initial search yield (from which the authors distilled the papers that met the criteria for inclusion)?

Response:

Over 300 peer-reviewed papers were acquired initially. We revised the sentence as below:

"Data were acquired by searching existing peer-reviewed literature (304 peer-reviewed papers) using the names of the sub-Saharan countries and the GHGs (i.e. CO₂, CH₄ and N₂O) as search terms (using Web of Science and Google Scholar; 1960 – 2015)."

Page 16485, line 6-11: Is there any reason to narrow your selection criteria? Can the authors analyze the results for different ecosystems by measurement method or frequency? Adding more information to the supplementary table on methods would help.

Response:

The paragraph was intended to note that the overall figures on GHG emissions shown are based on results achieved by different measurement techniques with inherent and contrasting sources of error.

If a paper provided detail information on gas collecting method, study periods and frequency, weather characteristics, and other environmental factors (soil, vegetation, management) we recorded the information in the supplementary database (see Appendix A). However, too few studies report sufficient descriptions and details of methodology to enable us to analyze available data by measurement methods, frequency or periods.

Page 16486, lines 7-10: Can the authors analyze the effect of soil moisture and temperature across the forest studies (e.g., more of a meta-analysis approach)? Or find ways to lump studies that measured or reported data on similar categories of controls?

Response:

We added new statistical analyses and discussion in section 3.1.1 as stated below:

"Observed annual soil CO₂ emissions in African natural terrestrial systems and agricultural lands showed significant correlations with annual mean air temperature ($r=-0.322$, $P=0.01$), annual rainfall ($r=0.518$, $P <0.001$), and SOC ($r=0.626$, $P <0.001$) and soil total N contents ($r= 0.849$, $P <0.001$) (Table 2). It was unexpected to find

negative relation between annual soil CO₂ emissions and annual mean air temperature in this study since positive relation between soil CO₂ flux and temperature has been well known (e.g., Bond-Lamberty and Thomson, 2010). We speculate that the generally high temperatures, and poor quality, of many African soils mean that air temperature increases frequently result in vegetation stress and/or soil aridity, hindering root and soil microbial activities (root and microbial respiration) and subsequent soil CO₂ flux. This would account for the negative relationship we observed between annual mean air temperature and annual soil CO₂ emissions, but is an unproven hypothesis that deserves further exploration."

Page 16487, lines 8-25: A mass balance, or C budget, perspective would help frame this paragraph. How do emissions relate to above- and belowground C inputs?

Response:

We newly conducted a correlation analysis for soil CO₂ flux with soil and environmental factors. We found an unexpected result showing negative relation between annual soil CO₂ emissions and annual mean air temperature. We discussed the results based on theories on drought and water stress on carbon balance and ecosystem production (see 3.1.1. CO₂ emissions). However, due to the dearth of data for above- and belowground C inputs, it was not feasible to test the relationship between soil CO₂ flux and above- and belowground C inputs as the reviewer suggested.

Page 16492, lines 7-27, Page 16493, lines 20-29; Page 16494, lines 7-9: Here are examples of where drawing on a mass balance perspective (and N saturation) would help provide a framework within which to interpret this list of results from individual studies. For example, in the case of the green beans, which did not increase emissions, much of the fixed N is harvested and exported from the system. There is also a need to understand relationships between management, N surplus, and emissions, which will depend on how loss pathways are partitioned (leaching v. gaseous losses).

Response:

We added the discussion in section 3.2.2 as below:

"Therefore, there may be potential to reduce N₂O emissions in the agroforestry practice, but it may require better management (i.e., reduced N inputs or adding a C source such as a cover crop together with an N source) and rotation planning."

Page 16493, line 3: I thought the review didn't include incubation studies. Or was this *in situ*?

Response:

For quantitative summary of GHG emissions, we only selected studies that reported *in situ* annual GHG emissions or those that provided enough information to estimate annual GHG emissions. So incubation studies were not included in the quantitative

summary of GHG emissions (section 3. 1 Summary of greenhouse gas emissions in Africa). However, results from incubation studies were included in the synthesis of results from greenhouse gas emissions studies (section 3.2 Sources and drivers of greenhouse gas emissions in Africa).

To clarify it, we revised a sentence in section '2.1. Data collection' as below:

"To produce the quantitative summary of GHG emissions, we selected studies that reported *in situ* annual GHG emissions or those that provided enough information to estimate annual GHG emissions through unit conversion and/or extrapolation of given data."

Page 16495, lines 22-24: The C isotope result comes a bit out of context here. Briefly explain why this was measured/objective of the study.

Response:

We revise the sentence as below:

In Kenya, CH₄ fluxes did not show any seasonal trend and did not indicate appreciable variability among two different strains of rice (Tyler et al., 1988).

Page 16497, line 6: What is meant by output here? Harvest, leaching, or gaseous losses?

Response:

The sentence was removed.

Page 16497, lines 12-21: In the agroforestry/maize systems, were fertilizer rates adjusted (reduced or eliminated) based on the N input from the legume trees? It seems that for some of these studies the N balance perspective would allow the authors to say whether there may be potential to reduce emissions (in line with theory, if N surplus is reduced), but may require better management (i.e., reduced inputs) and rotation planning.

Response:

We did not find fertilizer rates were adjusted based on the N input from the legume trees. We added the below sentence in the paragraph (see Agroforestry in 3.2.2. Greenhouse gas emissions from agricultural lands):

"Therefore, there may be potential to reduce N₂O emissions in the agroforestry practice, but it may require ecological nutrient management (i.e., reduced inorganic fertilizer N inputs accounting N input from the legume trees; adding a C source such as a cover crop together with an N source) and rotation planning."

Page 16498, line 5: Again, the discussion of incubation experiments is a bit confusing. Were these included in the selection criteria? Are they *in situ* rather than lab incubations? Perhaps

clarify in the methods.

Response:

For quantitative summary of GHG emissions, we only selected studies that reported *in situ* annual GHG emissions or those that provided enough information to estimate annual GHG emissions. So incubation studies were not included in the quantitative summary of GHG emissions (section 3. 1 Summary of greenhouse gas emissions in Africa). However, results from incubation studies were included in the synthesis of results from greenhouse gas emissions studies (section 3.2 Sources and drivers of greenhouse gas emissions in Africa).

To clarify it, we revised a sentence in section '2.1. Data collection' as below:

"To produce the quantitative summary of GHG emissions, we selected studies that reported *in situ* annual GHG emissions or those that provided enough information to estimate annual GHG emissions through unit conversion and/or extrapolation of given data."

Page 16499, line 8: Could place results in a broader sustainability context: soil CO₂ emissions are only one component of emissions from agricultural systems, which also have all of the CO₂ emissions from tillage, fuel use, and embodied emissions in chemical inputs, etc. (if used).

Response:

We added the prospective in '3.3 Suggested future studies' as below:

"Future research should consider the wider GHG budget of agriculture and include all the various (non-soil) components such as fuel use, and embodied emissions in chemical inputs."

Page 16500, lines 10-24 and Figure 5: Part (a) Can the authors separate the total N input by emissions graph by N source (e.g., manure, fertilizer, legume, or some combination of these)? It would be most interesting for part a, which is in a more realistic range of N input rates. For parts (b) and (c) it might be helpful to explain why these studies used such unrealistically high N rates, far outside of what would make economic sense for any farmer. What was the context of these studies?

Response:

In Fig. 5, we added N source information (control, organic fertilizer, inorganic fertilizer and mixture of organic and inorganic fertilizers) through showing different symbols for different N sources. The Fig. 5 (b) clearly indicated that very high N inputs came from mixture of organic and inorganic fertilizers and they were observed in vegetable gardens.

Page 16502, line 6: And N source (whether organic or inorganic).

Response:
It was revised as suggested.

Page 16502, line 15: Yes, and link new knowledge of microbial communities (e.g., functional gene abundance) to emissions rates (when talking about importance of identifying mechanisms/driving processes).

Response:
We revised the sentence as below:

"Where possible studies should seek to identify and separate driving processes contributing to efflux of soil CO₂ (e.g., autotrophic and heterotrophic sources), CH₄ (e.g., methanogenesis and methanotrophy) and N₂O (e.g., nitrification, denitrification, nitrifier denitrification) and link new knowledge of microbial communities (e.g., functional gene abundance) to GHG emissions rates."

Technical corrections

Page 16484, line 4: spell out AFOLU the first time

Response:
Corrected in line 12 in page 16482.

Page 16488, line 11: typo "this mechanisms"

Response:
The sentence was removed.

Page 16503, lines 22-23: two typos (advanced and higher)

Response:
We revised as it was suggested.

Anonymous Referee #2

The authors have done a notable job of bringing a lot of data into one article; however the structure at present is not acceptable. Due to the structure of the ‘results and discussion’ section it reads very much like a literature review made up of a list of examples which seem tediously linked. There has not been much actual synthesis, more just reporting on what individual studies have done. It would be far more informative to see more instances of ‘90% papers reviewed showed that: : :” as opposed to “x found Y, but Z found A”. I would suggest starting this section with the summary of GHG emissions section then go on to discuss individual findings with more actual synthesis.

Response:

We want to address that lack of studies and low data quality in the existing studies for this region made hard to synthesize information beyond reviewing their findings. There are fundamental challenges to address the comments due to lack of data in general and poor data quality specifically. For instance, few studies report GHG fluxes with respect to the research questions described above, e.g., mechanistic controls, or with suitable experimental designs, e.g., adequate replication. Therefore, it was difficult to synthesize beyond describing their findings relevant to key topical areas. Furthermore, the data available were not a large enough sample to conduct valid statistical analysis, except for soil CO₂ fluxes, with which we were able to provide new statistical results as described below. Despite the difficulties, we have made significant efforts throughout the manuscript to improve the synthetic contribution of our effort and to improve the MS in order to better describe and understand GHG emissions, mitigation potential and future challenges in SSA.

Major changes were summarized as below:

- 1) New statistical results were added (see 3. 1. 1 CO₂ emissions) to show that observed annual soil CO₂ emissions in African natural terrestrial systems and agricultural lands had significant correlations with annual mean air temperature, annual rainfall, and soil organic carbon and total nitrogen contents. Accordingly, Table 2 and relevant discussion were added.
- 2) We altered the previous descriptive list to a more thematically synthesized approach throughout Results and Discussion: the sub-title of the second section changed to ' Sources and drivers of greenhouse gas emissions in Africa' accordingly.
- 3) Summary of GHG emissions section (newly named as '3. 1 Summary of greenhouse gas emissions in Africa') was revised and is now located right up front, first in Results and Discussion.
- 4) We newly assessed data quality of the cited studies using the criteria suggested by Rochette and Eriksen-Hamel (2008) and Barton et al. (2015). We categorized the studies as three different groups: the methods are 1) *poor to very poor*, 2) *marginal* and 3) *good*. We newly added detail procedure of the assessment (see 2.1. Data collection), results (see 3.1.5 Data quality assessment) and discussion (see 3.3 Suggested future research) in the manuscript. We have recorded the assessment

results in Supplementary Information Table S1 and S2.

The authors also make the error of not addressing the massive elephant in the room as to WHY there is so little data from Africa. It's not simply a matter of scientific priorities but a massive socio-economic challenge! Mass poverty, extreme droughts, civil unrest, political instability, scientific funding/priorities etc. etc. are the main reason these data gaps exist. The authors seem to ignore this fact and suggest that it is as simple as someone deploying some cheap technologies such as chambers and IRGAS – noting that IRGAs are NOT a cheap technology! Unfortunately it is not that simple. There is certainly a point to be made that static chambers can be very cheap and require little know how to use but what about the analysis – where and how much will this cost?

Response:

We agree the current data gap is not only matter of research and science in the field but caused by long-lasting socio-economic issues in sub-Saharan Africa as well. We added the sentence below at the end of section 3.4.

"Beside, data acquisition will not be only determined by technical but also by socio-political (and economic) barriers in sub-Saharan Africa. These problems are not only affecting this process but are also driving forces for GHG emissions due to (e.g.) land-use change events. Therefore, the implication of social scientists on this kind of studies would be also needed."

Depending on countries in sub-Saharan Africa, different level of technology is applicable and approach and cost are very diverse. So we focused on providing a strategic plan for acquisition of soil GHG emission data such as prioritizing research topics and utilizing appropriate technology depending on level of scientific advance.

Specific comments

1. You need to establish some consistency with your units throughout the manuscript. It is confusing how you keep jumping from Pg to Gt to Kg etc... Pick one and stick with it through the manuscript using x10x where necessary. As it is it is very confusing and one must constantly be going back to check which unit you were in. It is best practice in science to use SI, in which case you should use kg and make use of x10x.

Response:

We modified the unit for CO₂ emissions. All CO₂ emissions were expressed as unit of Mg CO₂ throughout the text. In case of N₂O and CH₄ gases, some values were not large enough to apply 'Mg' unit so they were expressed as 'kg'.

2. Results/Discussion: Start this section with data from 3.4 so that it does not read like an introduction.

Response:

The summary of GHG emissions section (newly named as '3. 1 Summary of greenhouse gas emissions in Africa') is now located at the first section in Results and Discussion

Page 16481

Line 8: I would consider reporting these data in CO₂-eq. At present these GHG data are not comparable to each other.

Response:

Reporting CH₄ and N₂O emissions in CO₂ eq may have advantage and disadvantage at the same time so it can be applied depending on context. Providing CO₂ eq for CH₄ and N₂O gases would be convenient to compare them with CO₂ gas. However, it may cause unavoidable confusion to someone who wants to know the range of CH₄ and N₂O emissions. Considering the context in the referred line, we thought providing the values in both original units and CO₂ eq would be better since the sentence was intended to provide the range of GHG emissions as well as the comparisons of source by source. Therefore, we revised the sentences and provide both original units and CO₂ eq.

Line 11: Make use of abbreviation GHG

Response:

Changed to GHG

Line 16-18: How were they different?

Response:

Incorporation of crop residues or manure with inorganic fertilizers resulted in various change in CO₂ and N₂O- showing increase or decrease in CO₂ and N₂O depending on the studies. We revised the sentence as below:

"Incorporation of crop residues or manure with inorganic fertilizers resulted in significant changes in GHG emissions but these were different for CO₂ and N₂O either increasing or decreasing depending on studies."

Line 22: "croplands and type and..." does not read well. Please restructure

Response:

We revised the sentence as below:

"Improving fallow with nitrogen (N)-fixing trees increased CO₂ and N₂O emissions compared to conventional croplands. Type and quality of plant residue in the improving fallow is likely to be an important control factor affecting CO₂ and N₂O emissions."

Page 16482

Line 2: Change: “~~and~~ WITH natural and agricultural lands ~~contributed~~ CONTRIBUTING 76.3...”

Response:

We revised the sentence as below:

"Overall, total CO₂ eq emissions from African natural ecosystems and agricultural lands were $56.9 \pm 12.7 \times 10^9$ Mg CO₂ eq yr⁻¹ with natural ecosystems and agricultural lands contributing 76.3% and 23.7%, respectively."

Line 3: Change ‘Africa’ to ‘African’

Response:

Changed.

Line 5: Change: “options on emissions.” To “options for emissions”

Response:

Changed.

Line 8: Remove ‘and’ and change to ‘involving international’

Response:

We revised the sentence as below:

"There is also a need to develop a common strategy for addressing this data gap that may include identifying priorities for data acquisition, utilizing appropriate technologies, and involving international networks and collaboration."

Line 10: Redefine greenhouse gas as ‘GHG’

Response:

Changed.

Line 12: ‘land use’ to ‘land uses’

Response:

'agricultural, forestry and other land use (AFOLU)' has been commonly used in IPCC reports and other documents.

Line 20: Place comma after 'wetland'

Response:
Changed.

Page 16482

Line 2: "For example, CO₂ eq emissions from..." Are you talking just about CO₂ or about all the GHGs? You need to be clear. Using the terminology you have is not standard scientific practice and is confusing for the reader. CO₂-eq is a unit for standardising non-CO₂ GHGs for comparison to CO₂ and should not be used to describe the sum of all 3 GHG emissions. Additional confusion comes when you have stated 'CO₂ eq emissions' then report in terms of carbon! This section needs to be reworked to make it clear!

Response:

In the cited study (Borges et al., 2015), CO₂ eq emission was calculated by adopting 100-year global warming potentials (GWPs) of 28 and 265 for CH₄ and N₂O, respectively and then summing CO₂ emissions and GWPs of CH₄ and N₂O emissions. The method has been used in many other studies including a recent study (Tian et al., 2016, Nature).

We provided a modified unit (CO₂ eq per year) by multiplying 44/12 as below:

0.9 Pg C per year x (44/12) = 3.3 Pg CO₂ eq per year

Reference

Tian et al., 2016. The terrestrial biosphere as a net source of greenhouse gases to the atmosphere. *Nature*. 531, 225-228.

Line 17: I don't think fig 2 and 3 are particularly informative as you have stated all the information here in the text. Consider removing as they do not really add anything to your point.

Response
Removed.

Page: 16485

Line 19-23: Split this into 2 sentences.

Response:

We revised the sentence as below:

"Separate t-tests were used to assess significance of regression coefficients and intercepts in the fitted parametric models. Adjusted coefficients of determination

(adjusted R^2) of fitted parametric models were used as criteria for model selection: the model with the higher adjusted R^2 was selected.”

Line 24: Remove ‘These’ and start sentence with ‘Statistical...’

Response:
We removed it.

Page 16487

Line 1-7: There is no original hypoth testing or statistical analysis here. Merely a list of examples where other authors have found causes of fluxes. It seems the authors have not been systematic in their approach and are picking and choosing data to write about. It would be much more informative for a review such as this to say “70% of papers found temp affected CO2 flux in natural lands...”

Line 8 onwards: Much smarter analyses could have been done to summarise the data in the literature than just reporting a range of values

Line 8 onwards: None of this seems suitable to be called results or discussion...it reads like an intro. Where is your analysis?

Response:
We want to reiterated that lack of studies and low data quality in the existing studies for this region made hard to synthesize information beyond reviewing their findings. There are fundamental challenges to address the comments due to lack of data in general and poor data quality specifically. For instance, few studies report GHG fluxes with respect to the research questions described above, e.g., mechanistic controls, or with suitable experimental designs, e.g., adequate replication. Therefore, it was difficult to synthesize beyond describing their findings relevant to key topical areas. Furthermore, the data available were not a large enough sample to conduct valid statistical analysis, except for soil CO₂ fluxes, with which we were able to provide new statistical results as described below. Despite the difficulties, we have made efforts to improve the MS in order to better describe and understand GHG emissions, mitigation potential and future challenges in SSA.

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Page 16495

Line 3 and 10: Throughout MS you have used the American spelling of 'fertilizer' but on line 10 you use the British spelling. Be consistent through manuscript with you use of 'z' and 's'.

Response:

We changed to 'fertilized'.

Line 13-18: I would be VERY cautious to make these statements as you are reporting on 1 study. This tells us very little...it tells us about one place at one time and certainly no generalisations should be made about other grazing grasslands across Africa!! Acknowledge this as a limitation!

Response:

We revised the sentence as below:

"Only one study measured GHG emissions in grazing grasslands and there is a serious limitation in understanding GHG emissions in grazing grassland."

Page 16496

Line 1: Why have you suddenly switched to using kg CO₂ when everywhere else you have used Mg?! I have identified 4 different units being used through the MS (Mg, kg, Gt, Pg) when it should be 1! Do not be lazy and copy units from papers – make the conversions and the paper would be much easier to read.

Response:

We modified the unit for CO₂ emissions. All CO₂ emissions were expressed as unit of Mg CO₂ throughout the text. In case of N₂O and CH₄ gases, some values were not large enough to apply 'Mg' unit so they were expressed as 'kg'.

Line 10-15: I don't think you can make generalisations and draw conclusions from just 2 studies!

Response:

We revised to sentence to prevent over generalization as below:

"Greenhouse gas emissions from soils in vegetable gardens in peri-urban areas of Burkina Faso (Lompo et al., 2012) and Niger (Predotova et al., 2010) ranged from 73.3 to 132.0 Mg CO₂ ha⁻¹ y⁻¹ and 53.4 to 177.6 kg N₂O ha⁻¹ y⁻¹ (Table 1 and SI Table 1)."

Page 16499

This section needs to come first in the results/discussion section. This is your results, lead with this

Response:

The summary of GHG emissions section (newly named as '3. 1 Summary of greenhouse gas emissions in Africa') is now located at the first section in Results and Discussion

Page 16500

Line 5-9: I would be cautious about making these bold claims when gardens only used 2 studies!!!!

Response:

We recognize the limitation and revised them as below:

"The largest N₂O source in agricultural lands was vegetable gardens followed by agroforestry, cropland and rice fields (Table 1). The N₂O EF was 0.5 ± 0.2% and 3.5 ± 0.5% for cropland and vegetable gardens, respectively (Table 1 and SI Table 1). The N₂O EF of cropland is lower and the N₂O EF of vegetable gardens is higher than IPCC default N₂O EF (1%, IPCC, 2006). It is noticed that the results were made by limited number of studies and more research is needed to verify and update the results."

Page 16501

Line 9: Stop switching units!!

Response:

We modified the unit for CO₂ emissions. All CO₂ emissions were expressed as unit of Mg CO₂ through the text. In case of N₂O and CH₄ gases, some values were not large enough to apply 'Mg' unit so they were expressed as 'kg'.

Page: 16503

Line 3 onwards: But why is Africa studies less? Because it comes with more challenges...you need to acknowledge this!!

Response:

We surely understood the reviewer's concerns and also recognized the current data gap is not only matter of research and science in the field but caused by long-lasting socio-economic issues in sub-Saharan Africa as well. So we added the below sentence at the end of section 3.4.

"Beside, data acquisition will not be only determined by technical but also by socio-political (and economic) barriers in sub-Saharan Africa. These problems are not only affecting this process but are also driving forces for GHG emissions due to (e.g.) land-use change events. Therefore, the implication of social scientists on this kind of studies would be also needed."

Drs. Alberto Borges & Steven Bouillon' s comments

Kim and co-authors report an important data compilation of soil-atmosphere fluxes of greenhouse gases (GHGs) from the African continent that is probably the least studied on the globe despite the vital importance of the corresponding ecosystems such as the second largest evergreen tropical forest in the World. We would like to comment the way the river/stream data are classified per country in Table S1. The unit that matters for hydrology and river biogeochemistry (including ex-change of GHG with the atmosphere) is the river basin and not the country where the measurements were made. For instance, for the Congo River, the river basin comprises ten African countries (Angola, Burundi, Cameroon, Central African Republic, Democratic Republic of the Congo, Republic of the Congo, Rwanda, South Sudan, Tanzania, and Zambia). In Table S1, the data for Congo River are attributed to the Republic of the Congo although the data reported by Borges et al. (2015) were in fact acquired in the Democratic Republic of the Congo, a country that has the largest share of the Congo basin (60%). Similarly, the data on the Zambezi basin reported by Teodoru et al. (2015) were acquired in both Zambia and Mozambique, although the Zambezi basin comprises eight African countries (Angola, Namibia, Botswana, Zimbabwe, Zambia, Tanzania, Malawi, and Mozambique).

We would like to also highlight that lakes are important features of the African landscapes (in addition to rivers/streams) since these are among the largest in the world (Tanganyika, Victoria, Malawi, Kivu, Edward, Albert, etc. : :), and deserve further investigation with regards to GHG exchange. Some data are available for Lake Kivu (Borges et al. 2011; 2014). Data from Lake Kariba (Delsontro et al. 2011) and CH₄ from Ivory Coast lagoons (Koné et al. 2010) could also be included in the synthesis of aquatic fluxes.

References

Borges A.V., G. Abril, B. Delille, J.-P. Descy & F. Darchambeau (2011) Diffusive methane emissions to the atmosphere from Lake Kivu (Eastern Africa), *Journal of Geophysical Research - Biogeosciences*, 116, G03032, doi:10.1029/2011JG001673

Borges A.V., C. Morana, S. Bouillon, P. Servais, J.-P. Descy, F. Darchambeau (2014) Carbon cycling of Lake Kivu (East Africa): net autotrophy in the epilimnion and emission of CO₂ to the atmosphere sustained by geogenic inputs, *PLoS ONE* 9(10): e109500. doi:10.1371/journal.pone.0109500

Borges AV, Darchambeau F, Teodoru CR, Marwick TR, Tamooh F, Geeraert N, Omengo FO, Guérin F, Lambert T, Morana C, Okuku E & Bouillon S (2015) Globally significant greenhouse gas emissions from African inland waters, *Nature Geoscience*, 8, 637-642, doi:10.1038/NGEO2486

Delsontro T et al. (2011) Spatial Heterogeneity of Methane Ebullition in a Large Tropical Reservoir, *Environmental Science & Technology* (Impact Factor: 5.33). 12/2011; 45(23):9866-73.

Koné Y.J.M., G. Abril, B. Delille & A.V. Borges (2010) Seasonal variability of methane in the rivers and lagoons of Ivory Coast (West Africa), *Biogeochemistry*, 100, 21–37

Teodoru C. R., F. C. Nyoni, A. V. Borges, F. Darchambeau, I. Nyambe & S. Bouillon (2015) Dynamics of greenhouse gases (CO₂, CH₄, N₂O) along the Zambezi River and major tributaries, and their importance in the riverine carbon budget, *Biogeosciences*, 12, 2431–2453

Response:

We classified river and stream data per the river basin in Table S1. We newly added lake data (Lake Kivu (Borges et al. 2011; 2014), Lake Kariba (Delsontro et al. 2011), and Ivory Coast lagoons (Koné et al. 2010) in text and Table S1.

1 **Title**

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2 Greenhouse gas emissions ~~in~~from natural ecosystems and agricultural lands in sub-Saharan Africa:
3 synthesis of available data and suggestions for further studiesresearch

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4
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~~Abstract~~

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~~Acknowledgements~~

~~Reference~~

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Abstract

This paper summarizes currently available data on greenhouse gas (GHG) emissions from African natural ecosystems and agricultural lands, outlines the knowledge gaps and suggests future directions and strategies for GHG emission studies/research. GHG emission data were collected from 73 studies conducted in 22 countries in sub-Saharan Africa (SSA). Soil carbon dioxide (CO₂) emissions were by far the largest contributor to GHG emissions from and global warming potential (GWP) in African natural terrestrial systems. CO₂ emissions ranged from 3.3 to 57.0 Mg carbon dioxide (CO₂) ha⁻¹ yr⁻¹, methane (CH₄) emissions ranged from -4.8 to 3.5 kg methane (CH₄) ha⁻¹ yr⁻¹ and 0.1 to 13.7 kg (-0.16 to 0.12 Mg CO₂ equivalent (eq) ha⁻¹ yr⁻¹) and nitrous oxide (N₂O) ha⁻¹ yr⁻¹ emissions ranged from -0.1 to 13.7 kg ha⁻¹ yr⁻¹ (-0.03 to 4.1 Mg CO₂ eq ha⁻¹ yr⁻¹). Soil physical and chemical properties, rewetting, vegetation type, forest management and land-use changes were all found to be important factors affecting soil GHG emissions. Greenhouse gas emissions from natural terrestrial systems. In African aquatic systems, CO₂ was the largest contributor to total GHG emissions, ranging from 5.7 to 232.0 Mg CO₂ ha⁻¹ yr⁻¹, followed by -26.3 to 2741.9 kg CH₄ ha⁻¹ yr⁻¹ (-0.89 to 93.2 Mg CO₂ eq ha⁻¹ yr⁻¹) and 0.2 to 3.5 kg N₂O ha⁻¹ yr⁻¹ and (0.06 to 1.0 Mg CO₂ eq ha⁻¹ yr⁻¹). Rates of all GHG emissions from aquatic systems were all strongly affected by discharge. Soil GHG emissions from African croplands ranged, soil GHG emissions were dominated by CO₂, ranging from 1.7 to 141.2 Mg CO₂ ha⁻¹ yr⁻¹, with -1.3 to 66.7 kg CH₄ ha⁻¹ yr⁻¹ (-0.04 to 2.3 Mg CO₂ eq ha⁻¹ yr⁻¹) and 0.05 to 112.0 kg N₂O ha⁻¹ yr⁻¹ and the (0.015 to 33.4 Mg CO₂ eq ha⁻¹ yr⁻¹). N₂O emission factor/factors (EF) ranged from 0.01 to 4.1%. Incorporation of crop residues or manure with inorganic fertilizers invariably resulted in significant changes in GHG emissions but the magnitude and direction of changes were different for CO₂ and N₂O: as well as location. Soil GHG emissions from vegetable gardens ranged from 73.3 to 132.0 Mg CO₂ ha⁻¹ yr⁻¹ and 53.4 to 177.6 kg N₂O ha⁻¹ yr⁻¹ (15.9 to 52.9 Mg CO₂ eq ha⁻¹ yr⁻¹).

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1 and N₂O EFs ranged from 3 to 4%. Soil CO₂ and N₂O emissions from agroforestry were 38.6 Mg
2 CO₂ ha⁻¹ yr⁻¹ and 0.2 to 26.7 kg N₂O ha⁻¹ yr⁻¹; (0.06 to 8.0 Mg CO₂ eq ha⁻¹ yr⁻¹), respectively.
3 Improving fallow with nitrogen (N)-fixing trees led to increased CO₂ and N₂O emissions compared
4 to conventional croplands and. The type and quality of plant residue in the fallow is likely to be an
5 important control factor affecting on how CO₂ and N₂O emissions are affected. Throughout
6 agricultural lands, N₂O emissions slowly increased with N inputs below 150 kg N ha⁻¹ yr⁻¹ and
7 increased exponentially with N application rates up to 300 kg N ha⁻¹ yr⁻¹. The lowest yield-scaled
8 N₂O emissions were reported with N application rates ranging between 100 and 150 kg N ha⁻¹.
9 Overall, total CO₂ equivalent (eq) emissions from African ecosystems and agricultural lands
10 were 56.9 ± 12.7 Pgx 10⁹ Mg CO₂ eq yr⁻¹ andwith natural ecosystems and agricultural lands
11 contributedcontributing 76.3% and 23.7%, respectively. Additional GHG emission measurements
12 throughout AfricaAfrican agricultural and natural lands are urgently required to reduce uncertainty
13 on annual GHG emissions from the different land uses and identify major control factors and
14 mitigation options enfor emissions. There is also a need to develop a common strategy for
15 addressing this data gap that may involveinclude identifying priorities for data acquisition, utilizing
16 appropriate technologies, and establishinginvolving international networks and collaboration.
17
18 Key words: Africa, greenhouse gas, carbon dioxide, methane, nitrous oxide, natural lands,
19 agricultural lands.

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21 **1. Introduction**

22 Global greenhouse gas (GHG) emissions were estimated to be 49 (± 4.5) Gtx 10⁹ Mg CO₂ eq
23 in 2010 (IPCC, 2014), with approximately 21.2 – 24% (10.3 – 12 Gtx 10⁹ Mg CO₂ eq) of emissions
24 originating from soils in agricultural, forestry and other land use (AFOLU) (Tubiello et al., 2015;
25 IPCC, 2014). Annual non-CO₂ GHG emissions (primarily CH₄ and N₂O) from agriculture were
26 estimated to be 5.2 – 5.8 Gtx 10⁹ Mg CO₂ eq yr⁻¹ in 2010 (FAOSTAT, 2014; Tubiello et al., 2013),

1 ~~kg N ha⁻¹ in 2009 and only 16 kg N ha⁻¹ in sub-Saharan African countries. Only Mauritius,~~
2 ~~Botswana and South Africa had average N application rates exceeding 100 kg N ha⁻¹. However, use~~
3 of synthetic fertilizers such as urea has increased in the last four decades as well as the number of
4 livestock (and their manure and urine products) in Africa (Bouwman et al., 2009 and 2013) (Figs. 2
5 ~~and 3~~). The increasing trend in N application rates is expected to cause a twofold increase in
6 agricultural N₂O emissions in Africa by 2050 (from 2000) (Hickman et al., 2011). In the case of
7 CH₄ emissions, there are important differences between ecosystems. Tropical humid forest,
8 wetlands, rice paddy fields, and termite mounds are likely sources of CH₄, while seasonally dry
9 forests and savannahs are typically CH₄ sinks (Valentini et al., 2014).

10 Our current understanding of GHG emissions in Africa is particularly limited when
11 compared to the potential the continent has as both a GHG sink and source. This lack of data on
12 GHG emissions from African natural and agricultural lands and the lack of a comprehensive
13 analysis of existing data hinder the progress of our understanding of GHG emissions on the
14 continent (Hickman et al., 2014; Valentini et al., 2014; Ciais et al., 2011; Bombelli et al., 2009). In
15 order to identify mitigation measures and other climate smart interventions for the region it is
16 important to quantify baseline GHG emissions, as well as understand the impacts of different land-
17 use management strategies on GHG emissions (e.g., Palm et al., 2010).

18 In this study our objectives are to synthesize currently available data on GHG emissions
19 from African AFOLU; create an inventory of information from studies on emissions; and select
20 priority topics for future GHG emission studies in natural and agricultural lands in SSA.

2. Methodology

2.1. Data collection

2. Methodology

2.1. Data collection

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1 Data were acquired by searching existing peer-reviewed literature (304 peer-reviewed
 2 papers) using the names of the sub-Saharan countries and the GHGs (i.e. CO₂, CH₄ and N₂O) as
 3 search terms (using Web of Science and Google Scholar; 1960 – 2015). We To produce the
 4 quantitative summary of GHG emissions, we selected studies that reported *in situ* annual GHG
 5 emissions or those that provided enough information to estimate annual GHG emissions through
 6 unit conversion and/or extrapolation of given data. Data from 7376 studies, conducted in 22
 7 countries (n=244) in SSA were used and were further categorized as GHG emission in natural
 8 landecosystems [n=117; Supplementary Information (SI) Table 1] and agricultural lands (n=127;
 9 SI Table 2) (Fig. 42). The category of GHG emissions in natural landecosystems were further
 10 divided into emissions from natural terrestrial systems [forest/ plantation/woodland (n=55),
 11 savannah/grassland (n=31), termite mounds (n=5), and salt pans (n=1)] and aquatic systems
 12 [streams/rivers (n=14), wetlands/ floodplains /lagoons/ reservoirs/lakes (n=11), termite mounds
 13 (n=5), and salt pans (n=1)] (Table 1). The category of GHG emission in agricultural lands, were
 14 subdivided into emissions from cropland (n=105), rice paddypaddies (n=1), vegetable garden (n=5),
 15 and agroforestry (n=16) (Table 1). Across all categories there were 174 CO₂, 201 CH₄ and 184 N₂O
 16 emissions measurements. To allow comparison between different GHG emissions CH₄ and N₂O
 17 emissions were converted to CO₂ eq assuming a 100 year global warming potential and values of 34
 18 and 298 kg CO₂ eq for CH₄ and N₂O, respectively (IPCC, 2013). Where N₂O emission studies
 19 included experimental data from control plots with no N fertilizer additions (i.e. for background
 20 N₂O emissions) and from plots with different levels of applied N, a N₂O emission factor (EF) was
 21 calculated following the IPCC (2006) Tier I methodology as follows:

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$$N_2O\ EF(\%) = \frac{N_2O\ emission_{N_{treatment}} - N_2O\ emission_{control}}{N_{input}} \times 100 \quad [1]$$

1 where, $N_2O\ EF\ (\%)$ is N_2O emission factor, $N_2O\ emission_{N\ treatment}$ is N_2O emission in N input, N_2O
2 $emission_{control}$ is control treatments with no N fertilizer additions, and N_{input} is the amount of added
3 N.

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4 It should be noted that our data compilation includes a wide variety of studies that were
5 conducted under diverse biophysical conditions using a range of methodologies for quantifying
6 GHG emissions (e.g., different sampling protocols, chamber design, and emission rate calculations),
7 soil properties, and climatic factors. Therefore, the overall figures on GHG emissions shown are
8 based on results achieved by different measurement techniques with inherent and contrasting
9 sources of error. To assess data quality of the cited studies we used the criteria (rank from “very
10 poor” to “very good”) suggested by Rochette and Eriksen-Hamel (2008). We went through the
11 methods of the papers used in the study (only those for terrestrial emissions, since these criteria do
12 not work for aquatic systems) where there was sufficient detail in the methods section. We
13 categorized the studies as three different groups: the methods are 1) *poor to very poor*, 2) *marginal*
14 and 3) *good*. Studies that were ranked “poor” on 3 or more criteria, or “very poor” on 2 or more
15 criteria were categorized as the methods were *poor to very poor*. In addition, we took into account
16 the importance of sampling frequency (Barton et al., 2015) and sampling periods. Studies
17 estimating annual GHG emissions with a sampling frequency lower than biweekly (i.e., less than 2
18 times per month) and sampling periods of less than 6 months (i.e., covering both rainy and dry
19 seasons) were categorized as the methods were *poor to very poor*. Studies that were ranked as
20 “poor” on 2 criteria, or “very poor” on 1 criterion, or with insufficient details on the methods were
21 ranked as *marginal*. The *good* studies were those with only 1 “poor” ranking, sufficient detail and a
22 sampling frequency of every 2 weeks or more frequent.

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24 2.2. Statistical analyses

25 To determine the relationship between annual soil CO_2 emissions and edaphic and climatic
26 factors (e.g., soil pH, soil bulk density, soil organic carbon (SOC), total N, and annual average air

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1 temperature and rainfall) in African natural terrestrial systems and agricultural lands, we used a
2 Pearson correlation analysis. The compiled datasets were used to examine the best model fit ~~and to~~
3 ~~derive the corresponding model parameters~~selection for N₂O emissions and yield-scaled N₂O
4 emissions as a function of the respective N input levels. Different data fitting models (linear,
5 nonlinear, natural log, logarithm and sigmoidal) were tested for each dataset. The regression models
6 were checked for violation of assumptions of normal distribution (Shapiro–Wilk test),
7 homoscedasticity (Breusch–Pagan test), and constant variance (Durbin–Watson statistic) (Motulsky
8 and Christopoulos, 2004). Separate t-tests were used to assess significance of regression coefficients
9 and intercepts in the fitted parametric models ~~and adjusted.~~ Adjusted coefficients of determination
10 (adjusted R²) of fitted parametric models were used as criteria for model selection: the model with
11 the higher adjusted R² was selected. Statistical significance was considered at the critical level of
12 5%. ~~These statistical~~Statistical analyses were conducted using SAS[®] ver. 9.2 (SAS Institute, Cary,
13 NC, USA) and SigmaPlot[®] ver. 11.0 (Systat Software Inc., San Jose, CA, USA).

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15 3. Results and discussionDiscussion

16 3. 1. Greenhouse Summary of greenhouse gas emissions in natural lands — Africa

17 3. 1. 1. Terrestrial systems CO₂ emissions,

18 ~~— Soil GHG emissions from African natural terrestrial systems such as natural forest,~~
19 ~~plantation, woodland, savannah, grassland, termite mounds and salt pans. Carbon dioxide~~
20 ~~emissions ranged from 3.3 to 130.9 Mg CO₂ ha⁻¹ y⁻¹, 4.8 in natural terrestrial systems and from -~~
21 ~~11.9 to 3.5 kg CH₄ 232.0 Mg CO₂ ha⁻¹ y⁻¹ and 0. yr⁻¹ to 13.7 kg N₂O in aquatic systems. The area~~
22 ~~weighted average was 27.6 ± 17.2 Mg CO₂ ha⁻¹ yr⁻¹ (Table 1 and SI Table 1). Aquatic systems~~
23 ~~such as water bodies or water submerged lands were the largest source of CO₂ followed by forest,~~
24 ~~savannah, termite mounds and salt pans (Table 1). Soil CO₂ emissions in agricultural lands were~~
25 ~~similar to emissions from natural lands and ranged from 6.5 to 141.2 Mg CO₂ ha⁻¹. The high~~
26 ~~variability in yr⁻¹ with an area weighted average of 23.0 ± 8.5 Mg CO₂ ha⁻¹ yr⁻¹ (Table 1 and SI~~

1 Table 2). Vegetable gardens were the largest sources of CO₂ emission rates was likely related to
2 differences in largely due to the large C inputs, followed by agroforestry, cropland and rice fields
3 (Table 1 and SI Table 2).

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4 Observed annual soil CO₂ emissions in African natural terrestrial systems and agricultural

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5 lands showed significant correlations with annual mean air temperature, moisture (r=-0.322,
6 P=0.01), annual rainfall (r=0.518, P <0.001), and SOC (r=0.626, P<0.001) and soil total N content

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7 and physical chemical properties as (r= 0.849, P <0.001) (Table 2). The negative relationship
8 between annual soil CO₂ emissions and annual mean air temperature was unexpected since positive

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9 correlations between soil CO₂ flux and temperature are well as the type of natural established (e.g.,

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10 Bond-Lamberty and Thomson, 2010). We speculate that the generally high temperatures, and poor

11 quality, of many African soils mean that air temperature increases frequently result in vegetation

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12 present. Within stress and/or soil aridity, hindering root and soil microbial activities (root and

13 microbial respiration) and subsequent soil CO₂ flux (e.g., Thomas et al., 2011). This would account

14 for the negative relationship we observed between annual mean air temperature and annual soil CO₂

15 emissions, but is an unproven hypothesis that deserves further exploration.

16 3. 1. 2 CH₄ emissions

18 Forest/plantation/woodland were sinks of CH₄ (-1.5 ± 0.6 kg CH₄ ha⁻¹ yr⁻¹) and savannah/

19 grassland, crop lands, termite mounds, and rice fields were low to moderate CH₄ sources (0.5 – 30.5

20 kg CH₄ ha⁻¹ yr⁻¹). Stream/river and wetland/floodplain/lagoon/reservoir were high CH₄ sources

21 (766.0 – 950.4 kg CH₄ ha⁻¹ yr⁻¹) (Table 1 and Table 1 in supplementary material). The area

22 weighted averages of CH₄ emissions from natural and agricultural lands were 43.0 ± 5.8 and 19.5 ±

23 5.6 kg CH₄ ha⁻¹ yr⁻¹, respectively.

24 3. 1. 3 N₂O emissions and emission factor (EF)

1 Nitrous oxide emissions in natural ecosystems ranged from -0.1 to $13.7 \text{ kg N}_2\text{O ha}^{-1} \text{ yr}^{-1}$ and
2 the area weighted average was $2.5 \pm 0.8 \text{ kg N}_2\text{O ha}^{-1} \text{ yr}^{-1}$ (Table 1 and SI Table 1). Our study
3 reveals that forest, plantation and woodland were the largest source of N_2O followed by rivers and
4 wetlands, savannah and termite mounds in natural ecosystems (Table 1). Soil N_2O emissions in
5 agricultural lands ranged from 0.051 to $177.6 \text{ kg N}_2\text{O ha}^{-1} \text{ yr}^{-1}$ and the area weighted average was
6 $4.5 \pm 2.2 \text{ kg N}_2\text{O ha}^{-1} \text{ yr}^{-1}$ (Table 1 and SI Table 2). The largest N_2O source in agricultural lands
7 was vegetable gardens followed by agroforestry, cropland and rice fields (Table 1). The N_2O EF
8 was $0.5 \pm 0.2\%$ and $3.5 \pm 0.5\%$ for cropland and vegetable gardens, respectively (Table 1 and SI
9 Table 1). The N_2O EF of cropland is lower and the N_2O EF of vegetable gardens is higher than
10 IPCC default N_2O EF (1%, IPCC, 2006). The number of studies on N_2O emissions in Africa is,
11 however, particularly low ($n=14$) and there are significant regional gaps leading to uncertainties in
12 the conclusions that can be currently drawn.

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13 N_2O emissions were significantly affected by N input levels (Fig. 3). N_2O emissions
14 increase slowly up to $150 \text{ kg N ha}^{-1} \text{ yr}^{-1}$, after which emissions increase exponentially up to 300 kg
15 $\text{N ha}^{-1} \text{ yr}^{-1}$ (Fig. 3 (A)). Consistent with earlier work by van Groenigen (2010) N inputs of over 300
16 $\text{kg N ha}^{-1} \text{ yr}^{-1}$ resulted in an exponential increase in emission (Fig. 3 (B)), slowing to a steady state
17 with N inputs of $3000 \text{ kg N ha}^{-1} \text{ yr}^{-1}$. Overall, the relationship between N input and N_2O emissions
18 shows a sigmoidal pattern (Fig. 3 (C)). The observed relationship is consistent with the proposed
19 hypothetical conceptualization of N_2O emission by Kim et al. (2013) showing a sigmoidal response
20 of N_2O emissions to N input increases. The results suggest that N inputs over $150 \text{ kg N ha}^{-1} \text{ yr}^{-1}$
21 may cause an abnormal increase of N_2O emissions in Africa. The relationship between N input and
22 N_2O emissions show that the lowest yield-scaled N_2O emissions were reported for N application
23 rates ranging from 100 to 150 kg N ha^{-1} (Fig. 4). The results are in line with the global meta-
24 analysis of Philiber et al. (2012) who showed that from an N application rate $\sim 150 \text{ kg N ha}^{-1}$ the
25 increase in N_2O emissions is not linear but exponential.

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3.1.4 CO₂ eq emission

Carbon dioxide eq emission (including CO₂, CH₄ and N₂O) in natural lands ranged from 11.7 to 121.3 Mg CO₂ eq ha⁻¹ yr⁻¹ and the area weighted average of CO₂ eq. emissions (excluding salt pans) was 29.9 ± 22.5 Mg CO₂ eq ha⁻¹ yr⁻¹ (Table 1). Water bodies or water submerged lands such as rivers and wetlands were the largest source of CO₂ eq. emissions followed by forest/ plantation/ woodland, savannah/ grassland and termite mounds (Table 1). Carbon dioxide eq. emissions in agricultural lands ranged from 7.3 to 26.1 Mg CO₂ eq ha⁻¹ yr⁻¹ and had an area weighted average of CO₂ eq. emissions (excluding vegetable gardens and agroforestry due to lack of data) of 25.6 ± 12.4 Mg CO₂ eq ha⁻¹ yr⁻¹ (Table 1).

Total CO₂ eq. emissions in natural lands (excluding salt pans) were 43.4 ± 9.3 x 10⁹ Mg CO₂ eq yr⁻¹ with forest/ plantation/ woodland the largest source followed by savannah/grassland, stream/river, wetlands/floodplains/lagoons/reservoir, and termite mounds (Table 1). Total CO₂ eq. emissions in agricultural lands (excluding vegetable gardens and agroforestry) were 13.5 ± 3.4 x 10⁹ Mg CO₂ eq yr⁻¹ with crop land the largest source followed by rice fields (Table 1). Overall, total CO₂ eq emissions in natural ecosystems and agricultural lands were 56.9 ± 12.7 x 10⁹ Mg CO₂ eq yr⁻¹ with natural and agricultural lands contributing 76.3% and 23.7%, respectively.

3.1.5 Data quality assessment

Twenty third of the 76 studies cited in the study were categorized as methods were *poor to very poor*, 19 studies were *marginal* and 14 studies were *good* (Table S1 and S2). Major reasons the studies were ranked as *poor to very poor* were because sampling periods were too short for calculating annual emissions (i.e., less than or only one season of data), sampling frequency was too low (i.e., monthly or less), or a combination of poor methods with the sample collection, primarily insufficient samples per gas collecting chamber and very long chamber deployment times.

3.2 Sources and drivers of greenhouse gas emissions in Africa

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1 **3.2.1 Greenhouse gas emissions in natural ecosystems**

2
3 **Natural terrestrial systems;**

4 A range of factors affect direct emissions of soil CO₂ in African natural terrestrial systems
5 such as natural forest, plantation, woodland, savannah, grassland, termite mounds and salt pans.
6 These factors can be grouped into i) climatic, ii) edaphic, iii) vegetation and iv) human
7 interventions via land management. Data on the effects of these variables on GHG emissions are
8 variable, with some much less well understood than others. In almost all cases data are limited to a
9 few studies, and there are large areas where there has been no research. This hinders our ability to
10 estimate the contribution of African landscapes to global GHG emissions.

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11
12 Soil CO₂ emissions were strongly related to both soil moisture and temperature in forest
13 systems. For example, soil moisture explained about 50% of the seasonal variability in soil CO₂
14 efflux in a *Croton macrostachys*, *Podocarpus falcatus* and *Prunus africana* forest in Ethiopia
15 (Yohannes et al., 2011), as well as much of the seasonal variation in soil CO₂ efflux in a 3-year-old
16 *Eucalyptus* plantation in Republic of Congo (Epron et al., 2004). Thomas et al. (2011) found that
17 the Q₁₀ of soil CO₂ efflux (a measure of the temperature sensitivity of efflux, where a Q₁₀ of 2
18 represents a doubling of efflux given a 10°C increase in temperature) was dependent on soil
19 moisture at sites across the Kalahari in Botswana, ranging from 1.1 in dry soils, to 1.5 after a 2mm
20 rainfall event and 1.95 after a 50mm event. Similarly, in a Zambian woodland, the main driving
21 factor controlling CO₂ emissions at a seasonal time scale was a combination of soil water content
22 and temperature (Merbold et al., 2011).

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23 Increased GHG emissions following soil rewetting were observed in various regions in
24 Africa. Soil rewetting has a significant and well documented impact on GHG emissions (e.g., Kim
25 et al., 2012b). Two broad mechanisms responsible for changed soil GHG flux following rewetting
26 have been hypothesized: (1) enhanced microbial metabolism by an increase in available substrate

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1 due to microbial death and/or destruction of soil aggregates (i.e. commonly known as the Birch
2 effect (Birch, 1964)), and (2) physical mechanisms that can influence gas flux, including infiltration,
3 reduced diffusivity, and gas displacement in the soil (e.g., Kim et al., 2012b). Soil CO₂ efflux
4 increased immediately after rainfall in a sub-tropical palm woodland in northern Botswana,
5 however the increase was short-lived (Thomas et al., 2014). Large pulses of CO₂ and N₂O, followed
6 by a steady decline were also observed after the first rainfall event of the wet season in a Kenyan
7 rainforest (Werner et al., 2007). Soil CO₂ efflux was strongly stimulated by addition of rainfall in a
8 South African savannah (Fan et al., 2015; Zepp et al., 1996). In Zimbabwe, the release of N₂O from
9 dryland savannahs was shown to constitute an important pathway of release for N, and emissions
10 were strongly linked to patterns of rainfall (Rees et al., 2006).

11 Soil physical (e.g., bulk density, porosity and soil texture) and chemical properties (e.g., pH,
12 C and N) also affected soil GHG emissions (e.g., Saggari et al., 2013; Smith, 2010; Snyder et al.,
13 2009). Soil CO₂ efflux was positively related to total soil C content in undisturbed *miombo*
14 woodland in Zambia, although not in an adjacent disturbed woodland (Merbold et al., 2011). In a
15 Kenyan rainforest, CO₂ emissions were negatively correlated with subsoil C and positively
16 correlated with subsoil N concentrations, while N₂O emissions were negatively correlated with clay
17 content and topsoil C:N ratios (Werner et al., 2007). However, soil bulk density and pH were the
18 most influential factors driving spatial variation of *in situ* N₂O emissions in a tropical highland
19 rainforest in Rwanda (Gharahi Ghehi et al., 2014). Similarly, a laboratory-based experiment using
20 soils from 31 locations in a tropical mountain forest in Rwanda showed that N₂O emissions were
21 negatively correlated with soil pH, and positively correlated with soil moisture, soil C and soil N
22 (Gharahi Ghehi et al., 2012).

23 In many temperate systems, vegetation type also affects soil GHG emissions, likely because
24 of differences in litter quality and production rate, amount of below-ground biomass, the structure
25 of root systems as well as plant-mediated effects on soil microclimate (e.g., D'áz-Pinés et al., 2014;
26 Masaka et al., 2014; Kim et al., 2010). This is consistent with findings from African systems where

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1 annual soil CO₂ efflux also varied with vegetation types. For example, annual soil CO₂ emissions
2 were significantly lower in N-fixing acacia monocultures than in eucalypt monocultures and mixed-
3 species stands in Republic of Congo (Epron et al., 2013). The differences were attributed to leaf
4 area index in another study from the Republic of Congo where they found 71% of seasonal soil CO₂
5 efflux variability was explained by the quantity of photosynthetically active radiation absorbed by
6 the grass canopy (Caquet et al., 2012). Also in the Republic of Congo, it was found that litterfall
7 accounted for most of the age-related trends after the first year of growth, and litter decomposition
8 produced 44% of soil CO₂ flux in the oldest stand (Nouvellon et al., 2012), strongly suggesting that
9 the amount and quality of litter plays a major role in determining soil CO₂ flux. However, the effect
10 of vegetation type can also interact with soil physical-chemical properties. For example in Benin,
11 root respiration contributed to 30% of total soil CO₂ efflux in oil palms when the soil was at field
12 capacity and 80% when soil was dry (Lamade et al., 1996).

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13 ~~Forest soils predominantly act as sinks for CH₄ (Werner et al., 2007).~~ ~~In Cameroon,~~
14 ~~the largest CH₄ oxidation rates were observed from relatively undisturbed near primary forest sites~~
15 ~~(-14.7 to -15.2 ng m⁻² s⁻¹) compared to disturbed forests (-10.5 to 0.6 ng m⁻² s⁻¹) (Maedonald et al.,~~
16 ~~1998).~~ ~~Savannah and grassland were found to be both a sink and source of CH₄.~~ ~~In Mali, CH₄~~
17 ~~uptake was observed in dry sandy savannah (Delmas et al., 1991), while a savannah in Burkina~~
18 ~~Faso was found to be both a CH₄ sink and source during the rainy season, although overall it was a~~
19 ~~net CH₄ source (Brümmer et al., 2009).~~

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20 ~~Soil rewetting typically has a large impact on GHG emissions.~~ ~~Two broad mechanisms~~
21 ~~responsible for changed soil GHG flux following rewetting have been hypothesized: (1) enhanced~~
22 ~~microbial metabolism by an increase in available substrate due to microbial death and/or destruction~~
23 ~~of soil aggregates (i.e. commonly known as the Birch effect (Birch, 1964)), and (2) physical~~
24 ~~mechanisms that can influence gas flux, including infiltration, reduced diffusivity, and gas~~
25 ~~displacement in the soil (e.g., Kim et al., 2012b).~~ ~~Consistent with this mechanisms of re-wetting~~
26 ~~effects in soils of other continents (e.g., Kim et al., 2012b), soil CO₂ efflux increased immediately~~

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1 after rainfall in a sub-tropical palm woodland in northern Botswana, however the increase was
2 short lived (Thomas et al., 2014), while large pulses of CO₂ and N₂O, followed by a steady decline
3 were also observed after the first rainfall event of the wet season in a Kenyan rainforest (Werner et
4 al., 2007). Soil CO₂ efflux in a South African savannah was strongly stimulated by addition of
5 rainfall (Fan et al., 2015; Zepp et al., 1996) and soil N₂O concentrations increased markedly 30
6 minutes after wetting and peaked between 2 and 5 hours after rainfall in a semi-arid savannah
7 (Scholes et al., 1997). In Zimbabwe, the release of N₂O from dryland savannahs was shown to
8 constitute an important pathway of release for N, and emissions were strongly linked to patterns of
9 rainfall (Rees et al., 2006). In Botswana, Thomas and Hoon (2010) reported large and short lived
10 pulses of soil CO₂ efflux after artificial wetting of dry soils: soil CO₂ efflux on dry soils was
11 between 2.8–14.8 mg C m⁻² h⁻¹ but increased to 65.6 mg C m⁻² h⁻¹ in the hour after light wetting
12 and 339.2 mg C m⁻² h⁻¹ in the hour after heavy wetting.

13 Forest management such as burning, which is a common practice in SSA, and thinning, affects
14 GHG emissions (Table 23). The IPCC Tier 1 methodology only calculates the amount of GHG
15 emissions as a percentage of the carbon that is released through the burning; however it may also
16 increase forest soil GHG emissions once the fire has passed. For example, soil CO₂ efflux
17 immediately increased after burning of woodland in Ethiopia (Andersson et al., 2004); also, five
18 days after burning rainfall resulted in a 2-fold increase in soil CO₂ efflux from the burned plots
19 compared to the unburned plots. In contrast, 12 days after burning soil CO₂ efflux was 21% lower
20 in the burned plots (Andersson et al., 2004). However, contrasting impacts of fire on soil GHG
21 emission were observed in a savannah/grassland in the Republic of Congo where fire did not
22 change soil CO₂, CH₄ and N₂O fluxes (Castaldi et al., 2010, Delmas et al., 1991). Similarly, in
23 South Africa, soil CH₄ efflux was not significantly affected by burning (Zepp et al., 1996). In
24 contrast, annual fires decreased soil CH₄ oxidation rates in a Ghanaian savannah (Prieme and
25 Christensen, 1999). These case studies demonstrate that fire impacts are not always consistent and
26 this is likely the result of different fire characteristics (e.g., intensity or frequency), soil type (e.g.,

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1 Kulmala et al., 2014; Kim et al., 2011) and post-fire weather conditions. Thinning forest cover can
2 also can increase soil CO₂ efflux. Yohannes et al. (2013) reported 24% and 14% increases in soil
3 CO₂ efflux in the first and second years following thinning of a 6 year old *Cupressus lusitanica*
4 plantation in Ethiopia.

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5 There is a particular paucity of data on sources and sinks of CH₄ in African natural
6 terrestrial systems. In Cameroon, the largest CH₄ oxidation rates were observed from relatively
7 undisturbed near-primary forest sites (-14.7 to -15.2 ng m⁻² s⁻¹) compared to disturbed forests (-
8 10.5 to 0.6 ng m⁻² s⁻¹) (Macdonald et al., 1998). Savannah and grassland were found to be both a
9 sink and source of CH₄. Termite mounds are known sources of CH₄ and CO₂, and a In Mali,
10 CH₄ uptake was observed in dry sandy savannah (Delmas et al., 1991), while a savannah in Burkina
11 Faso was found to be both a CH₄ sink and source during the rainy season, although overall it was a
12 net CH₄ source (Brümmer et al., 2009). Termite mounds are known sources of CH₄ and CO₂

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13 (References). A study in a Burkina Faso savannah found that CH₄ and CO₂ released by termites
14 (*Cubitermes fungifaber*) contributed 8.8% and 0.4% of total soil CH₄ and CO₂ emissions,
15 respectively (Brümmer et al., 2009). In Cameroon, the mounds of soil-feeding termites
16 (*Thoracotermes macrothorax* and *Cubitermes fungifaber*) were point sources of CH₄ ranging 53.4
17 to 636 ng s⁻¹ mound⁻¹, which at the landscape scale may exceed the general sink capacity of the soil
18 (Macdonald et al., 1998). In Zimbabwe, it was found that *Odontotermes transvaalensis* termite
19 mounds located in dambos (seasonal wetlands) were an important source of GHGs, and emissions
20 varied with catena position for CO₂ and CH₄ (Nyamadzawo et al., 2012).

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21 Compared to the other environments covered in this review there are very few studies from
22 salt pans. Thomas et al. (2014) however, found soil CO₂ efflux increased with temperature and also
23 increased for a few hours after flooding of the surface of the Makgadikgadi salt pan in Botswana.

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24 Annual CO₂ emissions in salt pan were estimated as 0.7 Mg CO₂ ha⁻¹ ~~yyr~~⁻¹ (Thomas et al., 2014).

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26 3.1.2. Aquatic systems

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1 ~~Greenhouse gas emissions from African aquatic systems such as streams, rivers, wetlands,~~
2 ~~floodplains, reservoir, and reservoirs, lagoons ranged from 5.7 to 232.0 Mg CO₂ ha⁻¹ y⁻¹, 26.3 to~~
3 ~~2741.9 kg CH₄ ha⁻¹ y⁻¹ and 0.2 to 4.5 kg N₂O ha⁻¹ y⁻¹, and lakes can be significant sources of GHG~~
4 ~~(Table 1 and SI Table 1). Differences in regional setting and hydrology mean that emissions are~~
5 ~~highly spatially and temporally variable and when combined with the Nyong River (Cameroon),~~
6 ~~CO₂ paucity of studies, it is challenging to identify clear control factors.~~

7 ~~Studies found African aquatic systems can be significant sources of GHG emissions (5.5 kg~~
8 ~~CO₂ m⁻² y⁻¹) were four times greater than the flux of dissolved inorganic carbon (Brunet et al.,~~
9 ~~2009). In Ivory Coast, three out of five lagoons were oversaturated in CO₂ during all seasons and~~
10 ~~all were CO₂ sources (3.1 – 16.2 g CO₂ m⁻² d⁻¹) due to net ecosystem heterotrophy and inputs of~~
11 ~~riverine CO₂ rich waters (Koné et al., 2009). In the flooded forest zone of the Congo River basin~~
12 ~~(Republic of Congo) and the Niger River floodplain (Mali), high CH₄ emissions (5.16 × 10²⁰ – 6.35~~
13 ~~× 10²² g CH₄ m⁻² d⁻¹) were recorded on flooded soils (Tathy et al., 1992; Delmas et al., 1991). In~~
14 ~~Zimbabwe, dambos can be major or minor sources of GHGs depending on catena position. Upland~~
15 ~~dambos were important sources of N₂O and CO₂, and a sink for CH₄; while those in a mid-slope~~
16 ~~position were a major source of CH₄, but a weak source of CO₂ and N₂O; and those at the bottom~~
17 ~~were a weak source for all GHGs (Nyamadzawo et al., 2014a). In the Congo Basin (Republic of~~
18 ~~Congo), streams and rivers in savannah regions had higher Nyong River (Cameroon), CO₂~~
19 ~~emissions (46.8 – 56.4 g 5.5 kg CO₂ m⁻² d⁻¹ yr⁻¹) were four times greater than swamps (13.7 – 16.3 g~~
20 ~~CO₂ m⁻² d⁻¹) and tropical forest catchments (37.9 – 62.9 g CO₂ m⁻² d⁻¹) (Mann et al., the 2014). In~~
21 ~~the Okavango Delta (Botswana), the average CH₄ flux in river channels (0.75 g CH₄ m⁻² d⁻¹) was~~
22 ~~higher than that in floodplains and lagoons (0.41 – 0.49 g CH₄ m⁻² d⁻¹) (Gondwe and Masamba,~~
23 ~~2014 of dissolved inorganic carbon (Brunet et al., 2009). In the Zambezi River (Zambia), while CO₂~~
24 ~~and CH₄ concentrations in the main channel were highest downstream of the floodplains, N₂O~~
25 ~~concentrations were lowest downstream of the floodplains (Teodoru et al., 2015). Overall, 38% of~~
26 ~~the total C in the Zambezi River is emitted into the atmosphere, mostly as CO₂ (98 %) (Teodoru et~~

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al., 2015). The source of CH₄ to the atmosphere from Lake Kivu corresponded to ~60% of the terrestrial sink of atmospheric CH₄ over the lake's catchment (Borges et al., 2011). A recent study of 10 river systems in SSA estimated water-air CO₂, CH₄ and N₂O fluxes to be 8.2 to 66.9 g CO₂ m⁻² d⁻¹, 0.008 to 0.46 g CH₄ m⁻² d⁻¹, and 0.09 to 1.23 mg N₂O m⁻² d⁻¹, respectively (Borges et al., 2015). The authors suggested that lateral inputs of CO₂ from soils, groundwater and wetlands were the largest contributors of the CO₂ emitted from the river systems (Borges et al., 2015).

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The magnitude of GHG emissions from African aquatic systems varied with type and location. Streams and rivers in savannah regions had higher CO₂ emissions (46.8 – 56.4 g CO₂ m⁻² d⁻¹) than swamps (13.7 – 16.3 g CO₂ m⁻² d⁻¹) and tropical forest catchments (37.9 – 62.9 g CO₂ m⁻² d⁻¹) in the Congo Basin (Mann et al., 2014). ~~The~~ The average CH₄ flux in river channels (0.75 g CH₄ m⁻² d⁻¹) was higher than that in floodplains and lagoons (0.41 – 0.49 g CH₄ m⁻² d⁻¹) in the Okavango Delta (Botswana) (Gondwe and Masamba, 2014). Methane emissions from river deltas were substantially higher (~103 mg CH₄ m⁻² d⁻¹) than these from non-river bays (<100 mg CH₄ m⁻² d⁻¹) in Lake Kariba (Zambia/Zimbabwe). It was found substantially higher CH₄ fluxes in river deltas (~103 mg CH₄ m⁻² d⁻¹) compared to non-river bays (<100 mg CH₄ m⁻² d⁻¹) in Lake Kariba (Zambia/Zimbabwe) (DelSontro et al., 2011). While CO₂ and CH₄ concentrations in the main channel were highest downstream of the floodplains, N₂O concentrations were lowest downstream of the floodplains in the Zambezi River (Zambia and Mozambique) (Teodoru et al., 2015). Dambos in Zimbabwe can be major or minor sources of GHGs depending on catena position. Upland dambos were important sources of N₂O and CO₂, and a sink for CH₄; while those in a mid-slope position were a major source of CH₄, but a weak source of CO₂ and N₂O; and those at the bottom were a weak source of all GHGs (Nyamadzawo et al., 2014a).

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Studies were conducted to identify control factors for concentration and flux of GHGs in African aquatic systems. Studies found the concentration and flux of GHGs are strongly linked to stream hydrological characteristics such as discharge, but clear patterns have not yet been identified. ~~In the Congo River, surface~~ Surface CO₂ flux was positively correlated with discharge in the Congo

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1 River (Wang et al., 2013), while in Ivory Coast, rivers were often oversaturated with CO₂ and the
2 seasonal variability of partial pressure of CO₂ ($p\text{CO}_2$) was due to dilution during the flooding period
3 (Koné et al., 2009). Similarly, CO₂ fluxes show a very pronounced seasonal pattern strongly linked
4 to hydrological conditions in the Oubangui River in the Central African Republic (Bouillon et al.,
5 2012). Although higher CH₄ concentrations were found during low-discharge conditions, N₂O
6 concentrations were lowest during low-discharge conditions (Bouillon et al., 2012). In the Zambezi
7 River (Zambia) In Lake Kivu, seasonal variations of CH₄ in the main basin were driven by deepening
8 of the mixolimnion and mixing of surface waters with deeper waters rich in CH₄ (Borges et al.,
9 2011). In the Zambezi River (Zambia and Mozambique), inter-annual variability was relatively
10 large for CO₂ and CH₄ and significantly higher concentrations were measured during wet seasons
11 (Teodoru et al., 2015). However, inter-annual variability of N₂O was less pronounced and generally
12 higher values were found during the dry season (Teodoru et al., 2015).

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13 The relationship between GHG fluxes from aquatic systems Studies found the concentration
14 and flux of GHGs are strongly linked to and water temperature is environment or quality but clear
15 patterns have not clear yet been identified. In the Okavango Delta (Botswana), CH₄ emissions were
16 highest during the warmer, summerrainysummer rainy season and lowest during cooler winter
17 season suggesting the emissions were probably regulated by water temperature (Gondwe and
18 Masamba, 2014). However, Borges et al., (2015) found no significant correlation between water
19 temperature and $p\text{CO}_2$ and dissolved CH₄ and N₂O in 11 SSA river systems, althoughbut there was
20 a positive relationship between $p\text{CO}_2$ and dissolved organic C in six of the rivers. They also found
21 the lowest N₂O values were observed at the highest $p\text{CO}_2$ and lowest % O₂ levels, suggesting the
22 removal of N₂O by denitrification (Borges et al., 2015). In Lake Kivu (East Africa), the magnitude
23 of CO₂ emissions to the atmosphere seems to depend mainly on inputs of dissolved inorganic
24 carbon from deep geothermal springs rather than on the lake metabolism (Borges et al., 2014).
25
26 **3-2.2. Greenhouse gas emissions from agricultural lands**

3.2.1. Croplands

Soil GHG emissions reported from African croplands ranged from 1.7 to 141.2 Mg CO₂ ha⁻¹ y⁻¹, 1.3 to 66.7 kg CH₄ ha⁻¹ y⁻¹ and 0.05 to 112.0 kg N₂O ha⁻¹ y⁻¹ (Table 1 and SI Table 1). The N₂O EF ranged from 0.01 to 4.1% (Table 1 and SI Table 1).

Identifying controls on the emission of GHG from African agricultural land is even more challenging because in addition to natural variations associated with climate and soil type, land management (particularly fertilization) and crop type have a dominant influence on GHG emissions.

Croplands

The effects of the amount and type of N input on N₂O emissions in croplands have been studied in several locations (Table 23). In western Kenya, the rate of N fertilizer application (0 to 200 kg N ha⁻¹) had no significant effect on N₂O emissions (620 to 710 g N₂O-N ha⁻¹ for 99 days) (Hickman et al., 2014), however). However, another study from western Kenya, found a relationship between N input and N₂O emissions that was best described by an exponential model with the largest impact on N₂O emissions occurring when N inputs increased from 100 to 150 kg N ha⁻¹ (Hickman et al., 2015). An incubation study in Madagascar demonstrated that application of mixed urea and di-ammonium -phosphate resulted in lower N₂O emissions (28 vs. 55 ng N₂O-N g⁻¹ h⁻¹ for 28 days, respectively) than a mixed application of urea and NPK fertilizer (Rabenarivo et al., 2014).

Incorporation of crop residues to the soil has frequently been proposed to increase soil fertility (Malhi et al., 2011), however incorporation of crop residues also affects CO₂ and N₂O emissions (Table 23). In Tanzania, incorporation of plant residue into soil increased annual CO₂ fluxes substantially (emissions rose from 2.5 to 4.0 and 2.4 to 3.4 Mg C ha⁻¹ y⁻¹ for clay and sand soils, respectively) (Sugihara et al., 2012), although a study in Madagascar showed that rice-straw residue application resulted in larger fluxes of CO₂ but reduced N₂O emissions due to N immobilization (Rabenarivo et al., 2014). In contrast, application of *Tithonia diversifolia* (tithonia) leaves led to greater N₂O emissions compared to urea application in maize fields in Kenya (Sommer

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1 et al., 2015; Kimetu et al., 2007). The higher N₂O emissions after application of *Tithonia*
2 *diversifolia* were attributed to high levels of nitrate and available carbon in the soil caused by the
3 application that subsequently enhanced denitrification rates. In incubation studies with cultivated
4 soil from Ghana, N₂O emissions were significantly higher from soils amended with low C:N ratio
5 clover residues compared to high C:N ratio barley residues (Frimpong et al., 2012) ~~and increasing~~.
6 Increasing the proportion of maize in a cowpea-maize residue significantly decreased N₂O
7 emissions compared to cowpea residue incorporation alone (Frimpong et al., 2011), again likely due
8 to the higher C:N ratio of the maize residue compared with the cowpea. Another incubation study
9 with cultivated soil from Ghana showed that N₂O emissions increased after addition of residues of
10 three tropical plant species (*Vigna unguiculata*, *Mucuna pruriens* and *Leucaena leucocephala*) and
11 emissions were positively correlated with the ~~residue~~ C:N ratio of the residue, and negatively
12 correlated with residue polyphenol content, polyphenol:N ratio and (lignin + polyphenol):N ratio
13 (Frimpong and Baggs, 2010). It is rare for N₂O emissions to be positively correlated to C:N ratio
14 and the authors of the study suggest that it was either because soil C was limiting denitrification
15 rates or that release of N from the residues was slow (Frimpong and Baggs, 2010). The results
16 demonstrate that the quality of residues (e.g., C:N ratio, N, lignin and soluble polyphenol contents)
17 affect GHG emissions and further studies are needed to clearly identify the relationship between
18 them (Snyder et al. 2009; Mafongoya et al., 1997).

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19 Adding an additional ~~source of~~ N (mineral or organic) when crop residues are incorporated
20 into the soil could stimulate mineralization of crop residues, increase N-use efficiency and produce
21 higher yields (e.g., Garcia-Ruiz and Baggs, 2007) (Table 23). It was found that application of mixed
22 crop residue or manure and inorganic fertilizers resulted in different response of CO₂ and N₂O
23 emissions. In maize (*Zea mays* L.) and winter wheat (*Triticum aestivum* L.) fields in Zimbabwe,
24 application of inorganic fertilizer (ammonium nitrate, NH₄NO₃-N) with manure increased CO₂
25 emissions (26 to 73%), compared to sole application of manure (Nyamadzawo et al., 2014a).
26 However, the mixed application resulted in lower N₂O emissions per yield (1.6–4.6 g N₂O kg⁻¹

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1 yield), compared to sole application of inorganic fertilizer (6–14 g N₂O kg⁻¹ yield) (Nyamadzawo et
2 al., 2014a). Similarly, in a maize field in Zimbabwe, N₂O emissions were lower after the application
3 of composted manure and inorganic fertilizer (NH₄NO₃-N) compared to sole application of
4 inorganic fertilizer. The same treatments, however, led to the opposite results for CO₂ emissions
5 (Mapanda et al., 2011). In Mali, pearl millet (*Pennisetum glaucum*) fields treated with both manure
6 and inorganic fertilizer urea emitted significantly less N₂O than plots receiving only urea fertilizer
7 (Dick et al., 2008). The lower N₂O emissions in soils amended with manure were attributed to the
8 initial slow release and immobilisation of mineral N and the consequently diminished pool of N
9 available to be lost as N₂O (Nyamadzawo et al., 2014a, b; Mapanda et al., 2011; Dick et al., 2008).
10 In an incubation study with cultivated soils from Zimbabwe, Ghana and Kenya, combining organic
11 residue (maize, calliandra, and tithonia) and urea fertilizers decreased N₂O emissions in coarse-
12 textured soils but it increased N₂O emissions in fine-textured soils due to the higher level of
13 available N (Gentile et al., 2008).

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14 The effects of crop type and management on GHG emissions have also been studied by
15 several groups (Table 23). In Uganda, there were no significant differences in soil CO₂ effluxes
16 from different crops (lettuces, cabbages, beans) (Koerber et al., 2009). However, in Zimbabwe, rape
17 production resulted in greater N₂O emissions (0.64 – 0.93% of applied N was lost as N₂O) than
18 tomatoes (0.40 – 0.51% of applied N was lost as N₂O) (Masaka et al., 2014). ~~The results suggest
19 that the effect of crop type on GHG emissions is difficult to predict and more research is needed to
20 elucidate the relationship between crops, crop management and GHG emissions.~~

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21 ~~In Mali, growing N fixing haricot beans in rotation did not significantly increase N₂O
22 emissions (Dick et al., 2008).~~ In Madagascar, N₂O emissions were not significantly affected by
23 management practices such as direct seeding mulch-based cropping and traditional hand-ploughing
24 after harvesting (Chapuis-Lardy et al., 2009). However, the authors admitted the lack of difference
25 between treatments may be partially due to the short duration of the experiment and suggested more
26 complete monitoring to validate the observation. In highland Tanzanian maize fields, GHG fluxes

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1 were similar from soils under conventional and various conservation agriculture practices (Kimaro
 2 et al., 2015). However, when fluxes were yield-scaled the global warming potential (Mg CO₂ eq Mg
 3 grain⁻¹) was lower from fields with reduced tillage plus mulch and leguminous trees (2.1–3.1 Mg
 4 CO₂ eq Mg grain⁻¹) and from fields with reduced tillage plus mulch and nitrogen fertilizer (1.9–2.3
 5 Mg CO₂ eq Mg grain⁻¹) compared to fields under conventional agriculture (1.9–8.3 Mg CO₂ eq Mg
 6 grain⁻¹) (Kimaro et al., 2015). The results suggest that the effect of crop type and management on
 7 GHG emissions is difficult to predict and more research is needed to elucidate the relationship
 8 between crops, crop management and GHG emissions.

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9 Croplands were found to be both a sink and a source of CH₄. In Burkina Faso, CH₄ flux
 10 rates from croplands ranged from -0.67 to 0.70 kg CH₄-C ha⁻¹ yr⁻¹ (Brümmer et al., 2009), while
 11 in Republic of Congo, CH₄ uptake was observed in cassava and peanut fields and a recently
 12 ploughed field (Delmas et al., 1991). However, cropped and fertilised dambos in
 13 Zimbabwe were consistently sources of CH₄ (13.4 to 66.7 kg CH₄ ha⁻¹ yr⁻¹) (Nyamadzawo et al.,
 14 2014b).

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16 3.2.2. Grazing grassland

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17 Only one study measured GHG emissions in grazing grasslands and there is a serious
 18 limitation in understanding GHG emissions in grazing grassland. Thomas (2012) found that soil
 19 CO₂ efflux from a Botswana grazing land was significantly higher in sandy soils where the
 20 biological soil crust (BSC) was removed and on calcrete where the BSC was buried under sand. The
 21 results indicated the importance of BSCs for C cycling in drylands and indicate that intensive
 22 grazing, which destroys BSCs through trampling and burial, will adversely affect C sequestration
 23 and storage (Thomas, 2012).

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25 3.2.3. Rice paddies

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1 Rice paddies are ~~well known to be~~ sources of CH₄ (e.g., Linquist et al., 2012). Experiments
 2 measuring GHG emissions in rice paddies were conducted in Kenya (Tyler et al., 1988) and
 3 Zimbabwe (Nyamadzawo et al., 2013). In Kenya, ~~the range of δ¹³C in CH₄ for rice paddies was~~
 4 ~~from -57 to -63‰ and δ¹³C_{CH₄} fluxes~~ did not show any seasonal trend and did not indicate
 5 appreciable variability among two different strains of rice (Tyler et al., 1988). In Zimbabwe,
 6 intermittently saturated dambo rice paddies were a source of GHG and annual emissions from these
 7 rice paddies (150 day growing season and 126 kg of applied N ha⁻¹) were estimated as ~~2680 kg~~2.7
 8 Mg CO₂ ha⁻¹ yyr⁻¹, 12.5 kg CH₄ ha⁻¹, and 0.12 kg N₂O ha⁻¹ (Nyamadzawo et al., 2013). The IPCC
 9 (2006) use a CH₄ emission factor of 1.30 kg CH₄ ha⁻¹ day⁻¹ for rice cultivation. The CH₄ emissions
 10 in the dambo rice paddies referred to here are much lower than the IPCC estimate (195 kg CH₄ ha⁻¹
 11 ¹=1.3 kg CH₄ ha⁻¹ day⁻¹ × 150 days). The corresponding IPCC (2006) N₂O EF is 0.3% for rice
 12 cultivation and thus the N₂O emissions in the dambo rice paddies are also much lower than the
 13 IPCC estimate (0.40 kg N₂O-N ha⁻¹ = 126 kg N ha⁻¹ × 0.003; 0.63 kg N₂O ha⁻¹).

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15 ~~3.2.4. Vegetable gardens~~

16 Greenhouse gas emissions from soils in vegetable gardens in peri-urban areas of Burkina
 17 Faso (Lompo et al., 2012) and Niger (Predotova et al., 2010) ~~were much higher than all other land~~
 18 ~~uses, ranging ranged~~ from 73.3 to 132.0 Mg CO₂ ha⁻¹ yyr⁻¹ and 53.4 to 177.6 kg N₂O ha⁻¹ yyr⁻¹
 19 (Table 1 and SI Table 1).

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20 In Burkina Faso, annual CO₂ and N₂O emissions from the garden soils were 68 to 85% and
 21 3 to 4% of total C and N input, respectively (Lompo et al., 2012). The N₂O EFs (3 to 4%) were
 22 higher than the IPCC default value of 1.0% for all cropping systems (IPCC, 2006) and the global
 23 N₂O EF of vegetable fields (0.94%) (Rezaei Rashti et al., 2015). The high N₂O EFs may be
 24 attributed to the large amount of applied N in vegetable gardens (2700 – 2800 kg N ha⁻¹ yyr⁻¹) since
 25 surplus N will stimulate N₂O production and also indirectly promote N₂O production by inhibiting
 26 biochemical N₂O reduction (e.g., Shcherbak et al., 2014; Kim et al., 2013). In vegetable gardens of

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1 Niger, a simple plastic sheet roofing and addition of ground rock phosphate to stored ruminant
2 manure decreased N₂O gaseous losses by 50% in comparison to dung directly exposed to the sun
3 (Predotova et al. 2010). The authors argued that a decreased evaporation rate was behind this
4 abating effect.

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6 ~~3.2.5. Agroforestry~~

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7 Soil CO₂ and N₂O emissions from African agroforestry were 38.6 Mg CO₂ ha⁻¹ ~~yyr⁻¹~~ and
8 0.2 to 26.7 kg N₂O ha⁻¹ ~~yyr⁻¹~~, respectively (Table 1 and SI Table 1). ~~In agroforestry homegardens in~~
9 ~~Sudan, CO₂ (16.6 Mg CO₂ ha⁻¹ from June to December) and N₂O emissions (17.3 kg N₂O ha⁻¹ from~~
10 ~~June to December) accounted for two thirds of total C output and one third of total N output,~~
11 ~~respectively and the CO₂ and N₂O fluxes were positively correlated with soil moisture (Goenster et~~
12 ~~al., 2015).~~

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13 Improving fallow with N-fixing trees is a common agroforestry practice in several areas of
14 Africa since it provides additional N to the soil that can be utilised by the subsequent cash crop (e.g.,
15 Makumba et al., 2007; Chikowo et al., 2004; Dick et al., 2001). However, the practice is also
16 thought to increase CO₂ and N₂O emissions compared to conventional croplands (Table 2). ~~In an~~
17 ~~intercropping system with a N fixing tree (*Gliricidia*) and maize in southern Malawi, soil C was~~
18 ~~depleted as a result of enhanced CO₂ emissions, with over 67% of soil C lost over the first 7 years~~
19 ~~of intercropping (Kim, 2012a). In Zimbabwe, N₂O emissions in improved fallow agroforestry~~
20 ~~systems were 7 times higher than emissions in maize monoculture (Chikowo et al., 2004). In~~
21 ~~Senegal, soil collected under the N fixing tree (*Acacia raddiana*) emitted significantly more N₂O~~
22 ~~than soil collected under the N fixing crop (*Arachis hypogaea*) and non N fixing tree (*Eucalyptus*~~
23 ~~*camaldulensis*) (Dick et al., 3). Nitrous oxide 2006). In western Kenya, N₂O emissions increased~~
24 after incorporation of fallow residues and emissions were higher after incorporation of improved-
25 fallow legume residues than natural-fallow residues (Baggs et al., 2006; Millar and Baggs, 2004;
26 Millar et al., 2004). It was found that N₂O emissions were positively correlated with residue N

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1 content (Baggs et al., 2006; Millar et al., 2004) and negatively correlated with polyphenol content
2 and their protein binding capacity (Millar and Baggs, 2004), soluble C-to-N ratio (Millar and Baggs,
3 2005) and lignin content (Baggs et al., 2006). While high residue N content likely leads to more
4 available soil N and consequently increased N₂O production (Baggs et al., 2006; Millar and Baggs,
5 2005; Millar et al., 2004), polyphenols and lignins are both resistant to decomposition and could
6 result in N immobilization resulting in less labile soil N and less N₂O production (Baggs et al., 2006;
7 Millar and Baggs, 2004). ~~The type and quality of plant residue is likely to be an important control~~
8 ~~factor affecting N₂O emissions~~ Therefore, there may be potential to reduce N₂O emissions in the
9 agroforestry practice, but it may require ecological nutrient management (i.e., reduced inorganic
10 fertilizer N inputs accounting N input from the legume trees; adding a C source such as a cover crop
11 together with an N source) and rotation planning.

12 As in natural systems, improved fallow with N fixing trees agroforestry also results in
13 increased N₂O emissions following rainfall events. In an incubation experiment in Uganda, N₂O
14 emissions following simulated rainfall were at least 4 times larger for soils from under N-fixing
15 trees (*Calliandra calothyrsus*) compared to soils with non-N fixing trees (*Grevillea robusta*) (Dick
16 et al., 2001). Similarly, in Mali, N₂O emissions were around six times higher from improved fallow
17 with N-fixing trees (*Gliricidia sepium* and *Acacia colei*) following a simulated rainfall event,
18 compared with the emissions from soil under traditional fallow and continuous cultivation (Hall et
19 al., 2006). ~~Replacing traditional natural fallow with improved fallow systems in the humid tropics~~
20 ~~of Kenya also increased N₂O emissions by up to 3.9 kg N₂O-N ha⁻¹ over a 122 day maize cropping~~
21 ~~season (Millar et al., 2004).~~ In agroforestry homegardens in Sudan, CO₂ and N₂O fluxes were
22 positively correlated with soil moisture (Goenster et al., 2015).

24 ~~3.2.3~~ **Greenhouse gas emissions from land use change**

25 Land-use change affects soil GHG emissions due to changes in vegetation, soil, hydrology
26 and nutrient management (e.g., Kim and Kirschbaum, 2015) and the effects of land-use change on

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1 soil GHG emissions have been observed in African woodlands and savannah. In Zimbabwe,
2 clearing Clearing and converting woodlands to croplands increased soil emissions of CO₂, CH₄ and
3 N₂O (Mapanda et al., 2012) and soil CO₂ emissions from the converted croplands were higher than
4 Eucalyptus plantations established in former natural woodlands (Mapanda et al., 2010). In Republic
5 of Congo, early rotation changes Changes in soil CO₂ efflux after afforestation of a tropical
6 savannah with *Eucalyptus* were mostly driven by the rapid decomposition of savannah residues and
7 the increase in *Eucalyptus* rhizospheric respiration (Nouvellon et al., 2012).

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9 **3.4. Summary of greenhouse gas emissions in natural and agricultural lands in Africa**

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10 **3.4.1. CO₂ emissions**

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11 ~~Carbon dioxide emissions ranged from 3.3 to 130.9 Mg CO₂ ha⁻¹ y⁻¹ in natural terrestrial~~
12 ~~systems and from 11.9 to 232.0 Mg CO₂ ha⁻¹ y⁻¹ in aquatic systems. The area weighted average~~
13 ~~was 27.6 ± 17.2 Mg CO₂ ha⁻¹ y⁻¹ (Table 1 and SI Table 1). Aquatic systems such as water bodies or~~
14 ~~water submerged lands were the largest source of CO₂ followed by forest, savannah, termite~~
15 ~~mounds and salt pans (Table 1). Soil CO₂ emissions in agricultural lands were similar to emissions~~
16 ~~from natural lands and ranged from 6.5 to 141.2 Mg CO₂ ha⁻¹ y⁻¹ with an area weighted average of~~
17 ~~23.0 ± 8.5 Mg CO₂ ha⁻¹ y⁻¹ (Table 1 and SI Table 2). Vegetable gardens were the largest sources of~~
18 ~~CO₂ emission largely due to the large C inputs, followed by agroforestry, cropland and rice fields~~
19 ~~(Table 1 and SI Table 2).~~

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21 **3.4.2. CH₄ emissions**

22 ~~Forest/plantation/woodland were sinks of CH₄ (-1.5 ± 0.6 kg CH₄ ha⁻¹ y⁻¹) and savannah/~~
23 ~~grassland, crop lands, termite mounds, and rice fields were low to moderate CH₄ sources (0.5—30.5~~
24 ~~kg CH₄ ha⁻¹ y⁻¹). Stream/river and wetland/floodplain/lagoon/reservoir were high CH₄ sources~~
25 ~~(766.0—950.4 kg CH₄ ha⁻¹ y⁻¹) (Table 1 and SI Table 1). The area weighted averages of CH₄~~

emissions from natural and agricultural lands were 43.0 ± 5.8 and 19.5 ± 5.6 kg CH_4 ha^{-1} y^{-1} , respectively.

3.4.3. N_2O emissions and emission factor (EF)

Nitrous oxide emissions in natural lands ranged from 0.1 to 13.7 kg N_2O ha^{-1} y^{-1} and the area weighted average was 2.5 ± 0.8 kg N_2O ha^{-1} y^{-1} (Table 1 and SI Table 1). Our study reveals that forest, plantation and woodland were the largest source of N_2O followed by rivers and wetlands, savannah and termite mounds (Table 1). Soil N_2O emissions in agricultural lands ranged from 0.051 to 177.6 kg N_2O ha^{-1} y^{-1} and the area weighted average was 4.5 ± 2.2 kg N_2O ha^{-1} y^{-1} (Table 1 and SI Table 2). The largest N_2O source in agricultural lands was vegetable gardens followed by agroforestry, cropland and rice fields (Table 1). The N_2O EF was $0.5 \pm 0.2\%$ and $3.5 \pm 0.5\%$ for cropland and vegetable gardens, respectively (Table 1 and SI Table 1). The results indicate that the N_2O EF of African cropland is lower and the N_2O EF of African vegetable gardens is higher than IPCC default N_2O EF (1%, IPCC, 2006).

The relationship between N input and N_2O emissions varied depending on N input level (Fig. 4). N_2O emissions increase slowly up to 150 kg N ha^{-1} y^{-1} , after which emissions increase exponentially up to 300 kg N ha^{-1} y^{-1} (Fig. 5 (A)). Consistent with van Groenigen (2010) N inputs of over 300 kg N ha^{-1} y^{-1} resulted in an exponential increase in emission (Fig. 5 (B)), slowing to a steady state with N inputs of 3000 kg N ha^{-1} y^{-1} . Overall, the relationship between N input and N_2O emissions shows a sigmoidal pattern (Fig. 5 (C)). The observed relationship is consistent with the proposed hypothetical conceptualization of N_2O emission by Kim et al. (2013) showing a sigmoidal response of N_2O emissions to N input increases. The results suggest that N inputs over 150 kg N ha^{-1} y^{-1} may cause an abnormal increase of N_2O emissions in Africa. The relationship between N input and N_2O emissions show that the lowest yield scaled N_2O emissions were reported for N application rates ranging from 100 to 150 kg N ha^{-1} (Fig. 6). The results are in line with the global

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1 ~~meta-analysis of Philiber et al. (2012) who showed that from an N application rate 150 kg N ha⁻¹~~
2 ~~the increase in N₂O emissions is not linear but exponential.~~

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3.4.4. CO₂ eq emission

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4 ~~Carbon dioxide eq emission (including CO₂, CH₄ and N₂O) in natural lands ranged from~~
5 ~~11.7 to 121.3 Mg CO₂ eq. ha⁻¹ y⁻¹ and the area-weighted average of CO₂ eq. emissions (excluding~~
6 ~~salt pans) was 29.9 ± 22.5 Mg CO₂ eq. ha⁻¹ y⁻¹ (Table 1). Water bodies or water submerged lands~~
7 ~~such as rivers and wetlands were the largest source of CO₂ eq. emissions followed by forest/~~
8 ~~plantation/ woodland, savannah/ grassland and termite mounds (Table 1). Carbon dioxide eq.~~
9 ~~emissions in agricultural lands ranged from 7.3 to 26.1 Mg CO₂ eq. ha⁻¹ y⁻¹ and had an area~~
10 ~~weighted average of CO₂ eq. emissions (excluding vegetable gardens and agroforestry) of 25.6 ±~~
11 ~~12.4 Mg CO₂ eq. ha⁻¹ y⁻¹ (Table 1).~~

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12 ~~Total CO₂ eq. emissions in natural lands (excluding salt pans) were 43.4 ± 9.3 Pg CO₂ eq. y⁻¹~~
13 ~~with forest/ plantation/ woodland the largest source followed by savannah/ grassland, stream/ river,~~
14 ~~wetlands/ floodplains/ lagoons/ reservoir, and termite mounds (Table 1). Total CO₂ eq. emissions in~~
15 ~~agricultural lands (excluding vegetable gardens and agroforestry) were 13.5 ± 3.4 Pg CO₂ eq. y⁻¹~~
16 ~~with crop land the largest source followed by rice fields (Table 1). Overall, total CO₂ eq. emissions~~
17 ~~in natural and agricultural lands were 56.9 ± 12.7 Pg CO₂ eq. y⁻¹ with natural and agricultural lands~~
18 ~~contributing 76.3% and 23.7%, respectively.~~

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3.5. Suggested future studies research

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21 Despite an increasing number of published estimates of GHG emissions in the last decade,
22 there remains a high degree of uncertainty about the contribution of AFOLU to emissions in SSA.
23 due to lack of studies and uncertainty in the limited number of existing studies. To address this and
24 reduce the uncertainty surrounding the estimates, additional GHG emission measurements across
25 agricultural and natural lands throughout Africa are urgently required. Identifying controlling
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1 factors and their effects on GHG fluxes is a pre-requisite to enhancing our understanding of efflux
 2 mechanisms and a necessary step towards scaling up the field-scale data to landscape, national and
 3 continental scales. It is important to know how GHG fluxes can be affected by management
 4 practices and natural events such as logging (e.g., Yashiro et al., 2008), thinning (e.g., Yohannes et
 5 al., 2013), storms (e.g., Vargas, 2012), pest outbreaks (e.g., Reed et al., 2014), fires (e.g., Andersson
 6 et al., 2004), and wood encroachment (e.g., Smith and Johnson, 2004) in natural terrestrial systems
 7 and changing discharge (e.g., Wang et al., 2013) and water table (e.g., Yang et al., 2013) in aquatic
 8 systems. It is also important in agricultural lands to know how GHG fluxes are affected by
 9 management factors such as soil compaction (e.g., Ball et al., 1999), tillage (e.g., Sheehy et al.,
 10 2013), removal of crop residues (Jin et al., 2014), incorporation of crop residues and synthetic
 11 fertilizer (e.g., Nyamadzawo et al., 2014a), N input (whether organic or inorganic) (e.g., Hickman et
 12 al., 2015) and crop type (e.g., Masaka et al., 2014). However, because management and soil
 13 physical/chemical interactions cause different responses in soil GHG emissions (e.g. Pelster et al.,
 14 2012), it is critical to measure these interaction effects in the African context. The effect of
 15 predicted climatic change in Africa such as increased temperature (e.g., Dijkstra et al., 2012),
 16 changing rainfall patterns (e.g., Hall et al., 2006), increase in droughts incidence (e.g., Berger et al.,
 17 2013), rewetting effects (e.g., Kim et al., 2012b) and increased atmospheric CO₂ concentration (e.g.,
 18 Lane et al., 2013) also require further testing using laboratory and field experiments. Future
 19 research should consider the wider GHG budget of agriculture and include all the various (non-soil)
 20 components such as fuel use, and embodied emissions in chemical inputs.
 21 Where possible studies should seek to identify and separate driving processes contributing to
 22 efflux of soil CO₂ (e.g., autotrophic and heterotrophic sources), CH₄ (e.g., methanogenesis and
 23 methanotrophy) and N₂O (e.g., nitrification, denitrification, nitrifier denitrification) and link new
 24 knowledge of microbial communities (e.g., functional gene abundance) to GHG emissions rates.
 25 This is important because the consequences of increasing GHG emissions depend on the
 26 mechanism responsible. For example, if greater soil CO₂ efflux is primarily due to autotrophic

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1 respiration from plant roots, then it simply reflects greater plant growth. If however, it is due to
2 heterotrophic microbial respiration of soil organic carbon then it represents a depletion of soil
3 organic matter and a net transfer of C from soil to the atmosphere. Currently there are very few
4 studies that differentiate these sources making it impossible to truly determine the consequences
5 and implications on changes in soil GHG efflux.

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6 Land-use change has been recognized as the largest source of GHG emission in Africa
7 (Valentini et al., 2014). Hence, various types of conversion from natural lands to different land-use
8 types should be assessed to know how these changes may affect the GHG budget (e.g., Kim and
9 Kirschbaum, 2015). The focus of the assessment should be on deforestation and wetland drainage,
10 followed by a conversion to agricultural lands, since they are dominant types of land-use change in
11 Africa (Valentini et al., 2014).

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12 Throughout the study, we identified various trade-offs including increased CO₂ emission
13 following forest thinning management, increased GHG emissions in land-use changes, very high
14 N₂O emissions in vegetable gardens due to excessive N input to get high yields, increased CO₂ and
15 N₂O emission in incorporation of crop residues to the soil and agroforestry practices, and
16 exponential increased of N₂O emission and yield-scaled N₂O emissions in excessive N input.
17 Further studies are needed to assess and manage potential trade-offs and drivers.

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18 3.6.4 Strategic approaches for data acquisition

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19 A strategic plan for acquisition of soil GHG emission data in sub-Saharan Africa is required.
20 The success of any plan is dependent on long-term investment, stakeholder involvement, technical
21 skill and supporting industries, which have not always been available in the region (Olander et al.,
22 2013; Franks et al., 2012). A major challenge is to address the lack of consistency in the various
23 methodologies used to quantify GHG emissions (Rosenstock et al. 2013). Relatively low cost and
24 simple techniques can be used to determine GHG emission estimates in the first instance. Soil CO₂
25 fluxes can be quantified with a soda lime method (Tufekcioglu et al., 2001; Cropper et al., 1985;
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1 Edwards, 1982) or an infra-red gas analyzer (Bastviken et al., 2015; Verchot et al., 2008; Lee and
2 Jose, 2003) and these do not require advanced technology or high levels of resource to undertake.
3 Later, other GHG such as N₂O and CH₄ fluxes in addition to CO₂ flux can be measured with more
4 advanced technology (e.g., gas chromatography, photo-acoustic spectroscopy, or laser gas
5 analyzers). Initially, the measurement can be conducted using manual gas chambers with periodical
6 sampling frequencies. The sampling interval can be designed so that it is appropriate to the
7 particular type of land-use or ecosystem, management practices and/or for capturing the effects of
8 episodic events (e.g., Parkin, 2008). For example, GHG measurement should be more during
9 potentially high GHG emission periods following tillage and fertilizer applications and rewetting by
10 natural rainfalls or irrigation. With more ~~advanee~~advanced technology and utilisation of automatic
11 chamber systems measurements can be conducted at a much ~~high~~higher frequency with relative
12 ease. ▲

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13 ▲ In order for the challenges associated with improving our understanding of GHG emissions
14 from African soils it is critical to establish networks of scientists and scientific bodies both within
15 Africa and across the world. Good communication and collaboration between field researchers and
16 the modelling community should also be established during the initial stages of research, so results
17 obtained from field scientists can be effectively used for model development and to generate
18 hypotheses to be tested in the field and laboratory (de Bruijn et al., 2009). ▲

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19 ▲ Furthermore, lessons learned from scientific experiments can only really be successfully
20 implemented by farmers if local stakeholders are involved from the start and throughout (see for
21 example Stringer et al., 2012). Interviews, focus-groups, on-site or farm demonstrations, local
22 capacity building training, local farmers and extension staff can all improve dialogue and
23 understanding between local communities and scientists, ultimately improving the likelihood of
24 successful GHG emission and mitigation strategies. These will equip local researchers and
25 stakeholders (including farmers and extension staff) with state of art methodologies and help
26 motivate them to develop their GHG mitigation measures and assist them in understand their roles

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1 and contributions to global environmental issues. Beside, data acquisition will not be only
2 determined by technical but also by socio-political (and economic) barriers in sub-Saharan Africa.
3 These problems are not only affecting this process but are also driving forces for GHG emissions
4 due to (e.g.) land-use change events. Therefore, the implication of social scientists on this kind of
5 studies would be also needed.

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4. Conclusions

8 This paper synthesizes the available data on GHG emissions from African agricultural and
9 natural lands. Emissions of CO₂, CH₄ and N₂O in a variety of environments (forests, savannahs,
10 termite mounds, salt pans, agricultural areas and water bodies) were considered. Two broad
11 conclusions can be drawn from the work. The first one is that African natural and agricultural lands
12 may be a significant source of GHG and that the emissions may increase through land-use change
13 and management strategies. Secondly, there are huge research gaps. Africa is a vast continent, with
14 a multitude of land uses, climates, soils and ecosystems. Field-based data on soil GHG emissions
15 from many areas, soil types and environments are extremely sparse and as a result our
16 understanding of Africa's contribution to global GHG emissions remains incomplete and highly
17 uncertain. There is an urgent need to develop and agree on a strategy for addressing this data gap.
18 The strategy may involve identifying priorities for data acquisition, utilizing appropriate
19 technologies, and establishing networks and collaboration.

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Appendix A

A Blog for open discussion and web based open databases

23 We have created a Blog entitled 'Greenhouse gas emissions in Africa: study summary and
24 database' (<http://ghginafrica.blogspot.com/>) and an open-access database, which can be modified by
25 the users, entitled 'Soil greenhouse gas emissions in Africa database' (linked in the Blog) based on
26 this review. In the Blog, we have posted a technical summary of each section of this review, where

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1 comments can be left under the posts. The database contains detailed information on the studies
2 reported on GHG emissions, such as ecosystem and land use types, location, climate, vegetation
3 type, crop type, fertilizer type, N input rate, soil properties, GHGs emission measurement periods,
4 N₂O EF, and corresponding reference. The database is hosted in web based spreadsheets and is
5 easily accessible and modified. The authors do not have any relationship with the companies
6 currently being used to host the Blog and databases.

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8 **Acknowledgements**

9 We are grateful for the numerous researchers and technicians who provided invaluable data.
10 It is impossible to cite all the references due to limited space allowed and we apologize for the
11 authors whose work has not been cited. We are also grateful to [Benjamin Bond-Lamberty and](#)
12 [Rodrigo Vargas for insightful comments](#), Luis Lassaletta for providing raw data of N application
13 rates in Africa and Antony Smith for creating maps showing studies sites. A. S.-C. gratefully
14 acknowledges to the Spanish Ministry of Science and Innovation and the Autonomous Community
15 of Madrid for their economic support through the NEREA project (AGL2012-37815- C05-01,
16 AGL2012-37815-C05-04), the Agrisost Project (S2013/ABI-2717) and the FACCE JPI MACSUR
17 project. D.-G.K. acknowledges support from Research and Development Office, Wondo Genet
18 College and IAEA Coordinated Research Project (CRP D1 50.16).

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1 Table 1 Summary of greenhouse gas carbon dioxide (CO₂), methane (CH₄), nitrous oxide (N₂O) emissions and CO₂ equivalents (CO₂ eq) in natural
 2 ecosystems and agricultural lands in sub-Saharan African countries. Mean ± standard error (number of data) are shown.

Type	Area (Mha)	CO ₂ emission	CH ₄ emission	N ₂ O emission	N ₂ O emission factor	CO ₂ eq emission	Total CO ₂ eq emission
		Mg CO ₂ ha ⁻¹ yr ⁻¹	kg CH ₄ ha ⁻¹ yr ⁻¹	kg N ₂ O ha ⁻¹ yr ⁻¹	%	Mg CO ₂ eq. ha ⁻¹ yr ⁻¹	$\text{Pg} \times 10^9 \text{ Mg CO}_2 \text{ eq. yr}^{-1}$
▲ Forest/ plantation/ woodland	740.6 [#]	32.0 ± 5.0 (34)	-1.5 ± 0.6 (15)	4.2 ± 1.5 (10)	*	34.0 ± 5.7	25.2 ± 4.2
▲ Savannah /grassland	638.9 [#]	15.5 ± 3.8 (11)	0.5 ± 0.4 (18)	0.6 ± 0.1 (6)	*	15.8 ± 3.8	10.1 ± 2.4
▲ Stream/river	28.2 [#]	78.1 ± 13.2 (27)	436.3 ± 133.8 (24)	1.6 ± 0.3 (17)	*	93.4 ± 17.9	2.8 ± 1.0
▲ Wetlands/floodplains/lagoons/reservoir	43.8 [#]	96.6 ± 31.0 (7)	950.4 ± 350.4 (5)	2.0 ± 1.5 (2)	*	121.3 ± 39.7	5.3 ± 1.7
▲ Termite mounds	0.97 [†]	11.6 ± 6.2 (3)	2.3 ± 1.1 (3)	0.01 (1)	*	11.7 ± 6.3	0.01 ± 0.01
▲ Salt pan	*	0.7 (1)	*	*	*	*	*
▲ Total natural <u>lands¹ ecosystems¹</u>	1452.5	27.6 ± 2.9 ^s	43.0 ± 5.8 ^s	2.5 ± 0.4 ^s	*	29.9 ± 22.5 ^s	43.4 ± 9.3 (76.3%) ^{††}
▲ Cropland	468.7 [#]	23.4 ± 5.1 (45)	19.3 ± 4.2 (26)	4.0 ± 1.5 (83)	0.5 ± 0.2 (24)	26.1 ± 6.0	12.2 ± 2.8
▲ Rice field	10.5 ^{##}	6.5 (1)	30.5 (1)	0.19 (1)	*	7.3	1.3 ± 0.6
▲ Vegetable gardens	*	96.4 ± 10.2 (5)	*	120.1 ± 26.1 (5)	3.5 ± 0.5 (2)	*	*
▲ Agroforestry	190 [‡]	38.6 (1)	*	4.7 ± 2.2 (15)	*	*	*
▲ Total agricultural lands ²	479.2	23.0 ± 8.5 ^s	19.5 ± 5.6 ^s	4.5 ± 2.2 ^s	*	25.6 ± 12.4 ^s	13.5 ± 3.4 (23.7%) ^{††}
▲ Total natural <u>ecosystems</u> and agricultural lands ³	1931.7						56.9 ± 12.7

4 [#] GlobCover 2009

5 [†] 0.07% of savanna and rainforest (Brümmer et al., 2009)

6 ^{##} FAO STAT (<http://faostat3.fao.org/home/E>), year 2012

7 ^{*} No data available

8 ^s Area weighted average

9 [‡] Zomer et al., 2009

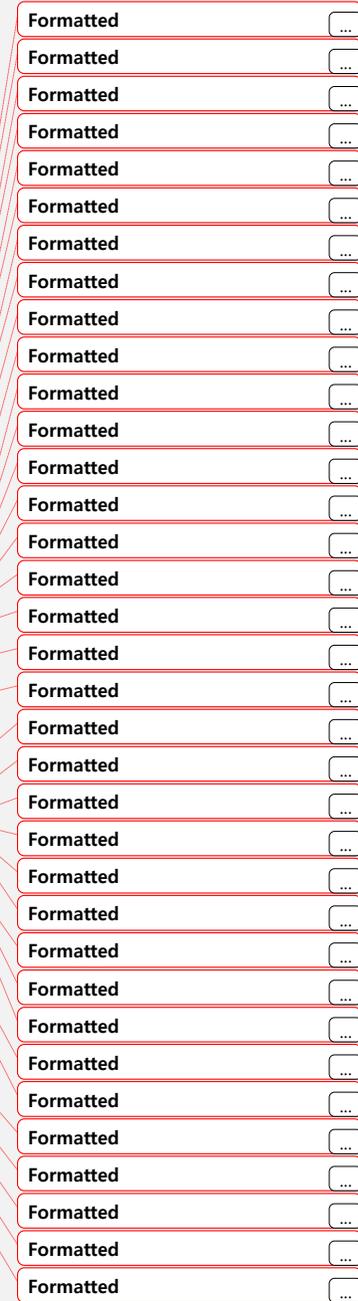
10 ^{††} Contribution to CO₂ eq. emission in total natural and agricultural lands

11 ¹ except salt pan

12 ² except vegetable gardens and agroforestry

13 ³ except salt pan, vegetable gardens and agroforestry

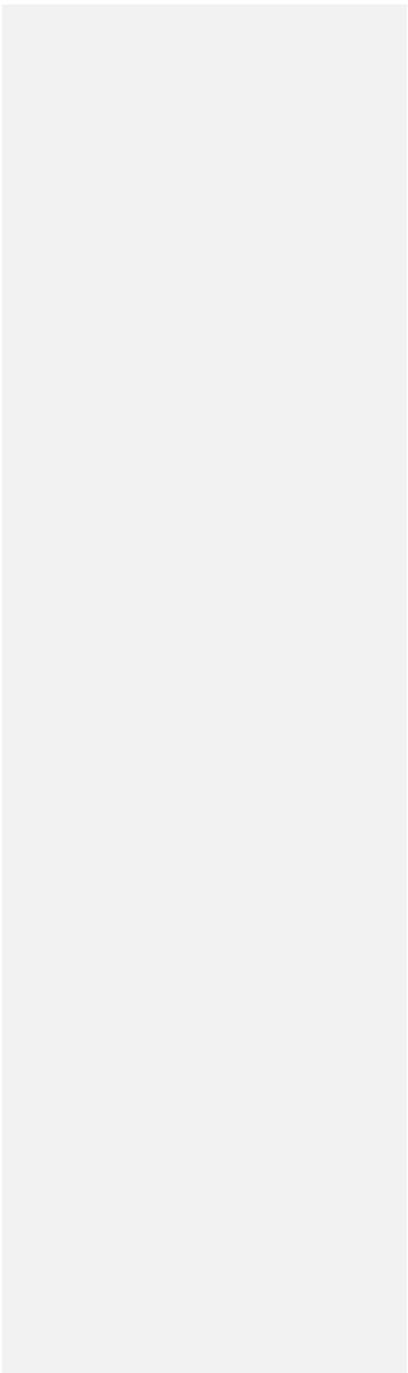
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Table 2 Correlation between annual soil CO₂ emissions (Mg CO₂ ha⁻¹ yr⁻¹) and environmental factors in African natural terrestrial systems

	<u>Annual mean Air temperature (°C)</u>	<u>Annual rainfall (mm)</u>	<u>Soil organic carbon (%)</u>	<u>Soil total nitrogen (%)</u>
<u>Correlation coefficient</u>	<u>-0.322</u>	<u>0.518</u>	<u>0.626</u>	<u>0.849</u>
<u>P- value</u>	<u>0.01</u>	<u>< 0.001</u>	<u>< 0.001</u>	<u>< 0.001</u>
<u>Number of samples</u>	<u>60</u>	<u>61</u>	<u>31</u>	<u>26</u>



1 **Table 3** Summary of the effect of management practices on greenhouse gas (GHG) emissions. + indicates increasing, · indicates no change, and –
 2 indicates decreasing.
 3

Land use/ecosystem type	Management practices	Impact on GHG			Country (data source)	
		CO ₂	N ₂ O	CH ₄		
Forest/ plantation/ Woodland	Burning	+			Ethiopia (Andersson et al., 2004)	
	Thinning	+			Ethiopia (Yohannes et al., 2013)	
	Land use change (cleaning and conversion to croplands)	+	+	+	Zimbabwe (Mapanda et al., 2012; Mapanda et al., 2010)	
	Flooding			+	Cameroon (Macdonald et al., 1998); Republic of Congo (Tathy et al., 1992); Mali (Delmas et al., 1991)	
Savannah/grassland	Burning	·	·	·	Republic of Congo (Castaldi et al., 2010; Delmas et al., 1991); South Africa (Zepp et al., 1996)	
	Land use change (cleaning and conversion to croplands)	+			¹ Republic of Congo (Nouvellon et al., 2012)	
Croplands	Increase in N fertilization rate		+		Kenya (Hickman et al., 2015)	
	Type of synthetic fertilizer		·		Madagascar (Rabenarivo et al., 2014)	
	Application of plant residues			–		Tanzania (Sugihara et al., 2012); ² Madagascar (Rabenarivo et al., 2014)
			+	+		Kenya (Kimetu et al., 2006); ³ Ghana (Frimpong et al. 2012)
	Crop residues + N fertilizer			+		⁴ Zimbabwe (Nyamadzawo et al., 2014a,b)
				–		⁵ Zimbabwe, Ghana and Kenya (Gentile et al., 2008)
	Combination of synthetic and organic fertilizers		+	–		⁶ Zimbabwe (Mapanda et al., 2011)
				–		⁷ Mali (Dick et al., 2008)
	Crop type		·			⁸ Uganda (Koerber et al., 2009)
				–		⁹ Zimbabwe (Masaka et al., 2014)
	Introducing N fixing crops in rotations			–		Mali (Dick et al., 2008)
	Direct seeding mulch-based			·		Madagascar (Chapuis-Lardy et al., 2009)
	Hand-ploughing after harvesting			·		Madagascar (Chapuis-Lardy et al., 2009)
Intensive grazing		+			Botswana (Thomas, 2012)	
<u>Reduced tillage + mulch, leguminous crop/tree, or N fertilizer</u>		<u>±</u>	<u>±</u>	<u>±; –</u>	<u>Tanzania (Kimaro et al., 2015)</u>	
Vegetable gardens	Plastic cover for ruminant manure		–		Niger (Predotova et al. 2010)	

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	Incorporation of fallow residues		+		Kenya (Baggs et al., 2006; Millar and Baggs, 2004; Millar et al., 2004)
▲ Agroforestry	Improving fallow with N-fixing crops		+		Zimbabwe (Chikowo et al., 2004)
	Cover crops		+		Kenya (Millar et al., 2004)
	N-fixing tree species	+	+		Malawi (Kim, 2012; Makumba et al., 2007); Senegal (Dick et al., 2006)

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1 ¹U+DAP instead U+NPK; ²N₂O study; ³Low C:N ratio clover residues compared to high C:N ratio barley residues; ⁴Application of ammonium nitrate
2 with manure to maize (*Zea mays* L.) and winter wheat (*Triticumaestivum* L.) plant residues; ⁵Plant residues of maize, calliandra, and tithonia + urea;
3 ⁶Mixed application of composted manure and inorganic fertilizer (AN); ⁷Manure and urea; ⁸Lettuces vs cabbages vs beans; ⁹Tomatoes vs rape

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4 **Figure captions**
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6 Figure 1. Change of areas of agricultural land and forest in Africa. Data source: FAOSTAT, <http://faostat.fao.org/site/377/default.aspx#ancor>, Access
7 [23 April 2015](#).

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9 Figure 2. ~~Change of use of urea fertiliser in Africa. Data source: FAOSTAT, <http://faostat.fao.org/site/422/default.aspx#ancor>, Access 23 April~~
10 ~~2015. Maps showing study sites of CO₂, CH₄ and N₂O fluxes.~~

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12 ~~Figure 3. Trends of African livestock population. Data source: FAOSTAT, http://faostat3.fao.org/faostat_gateway/go/to/download/Q/QA/E, Access 23~~
13 ~~April 2015.~~

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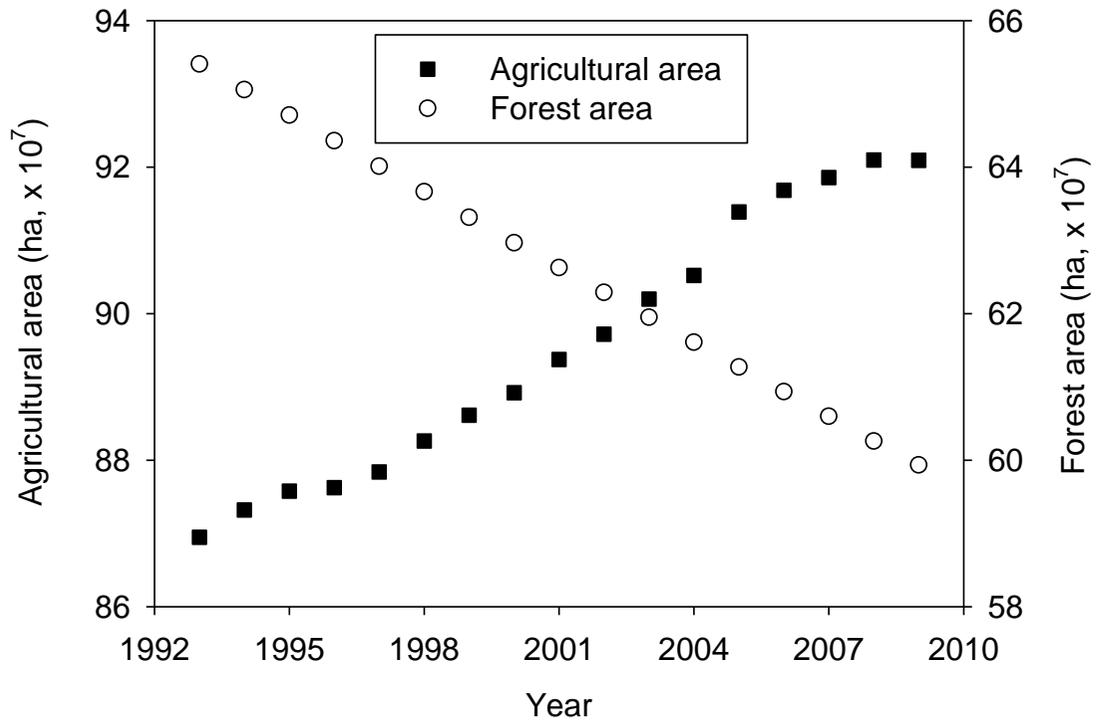
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15 ~~Figure 4. Maps showing study sites of CO₂, CH₄ and N₂O fluxes~~

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17 ~~Figure 5.~~

18 ~~Figure 3. Relationship between nitrogen (N) input and nitrous oxide (N₂O) emissions observed in Africa. N input ranged from 0 to 300 (A), 300 to~~
19 ~~4000 (B) and 0 to 4000 kg N ha⁻¹ yr⁻¹ (C). The dashed lines indicate 95% confidence intervals. *Control indicates no fertilizer application, Organic*~~
20 ~~*fertilizer is manure, Inorganic fertilizer includes NPK, ammonium nitrate and urea fertilizers, and Mixture indicated mixed application of organic and*~~
21 ~~*inorganic fertilizers.*~~

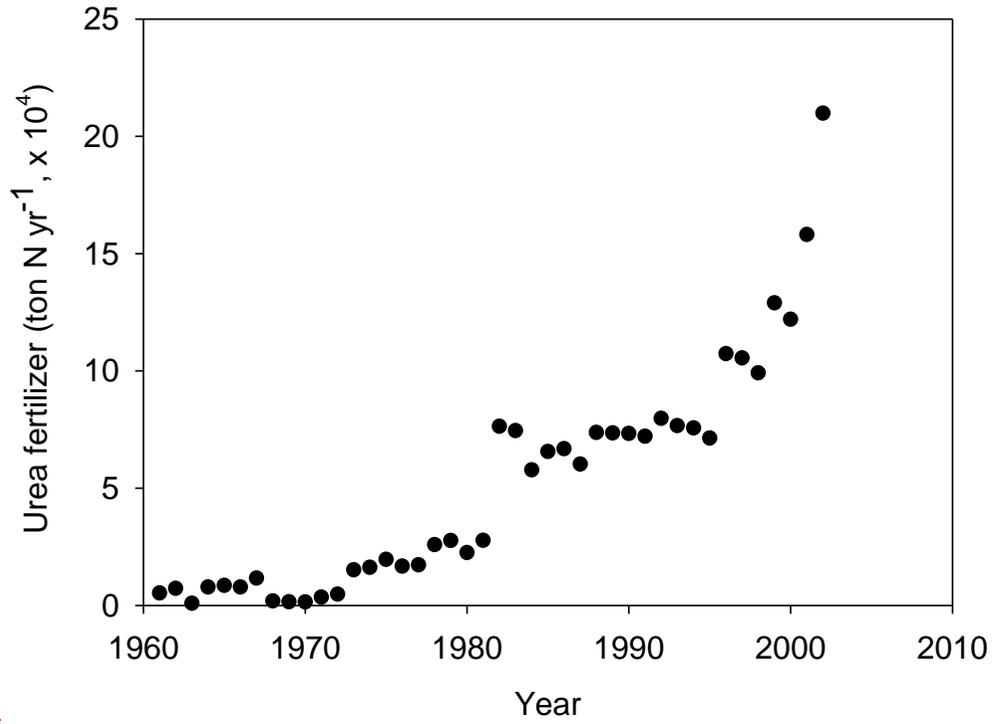
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23 ~~Figure 64. Relationship between nitrogen (N) input and yield scaled nitrous oxide (N₂O) emissions. Grain type: (A) rape (*Brassica napus*) and (B) and~~
24 ~~(C) maize (*Zea mays* L.). Data sources: (A) from Nyamadzawo et al. (2014), (B) from Hickman et al. (2014) and (C) from Hickman et al. (2015). The~~
25 ~~dashed lines indicate 95% confidence intervals. Note the different scales across panels.~~

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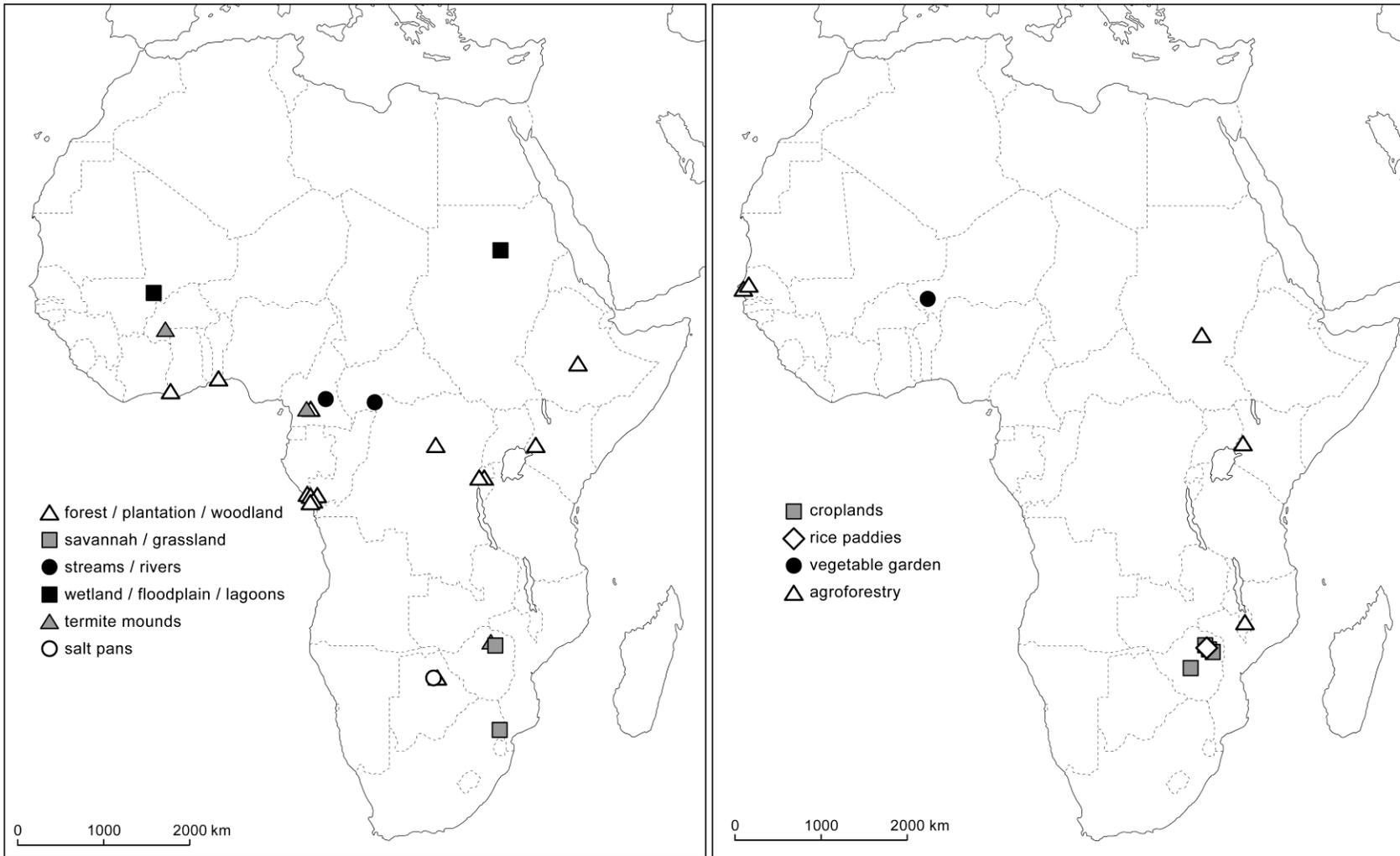
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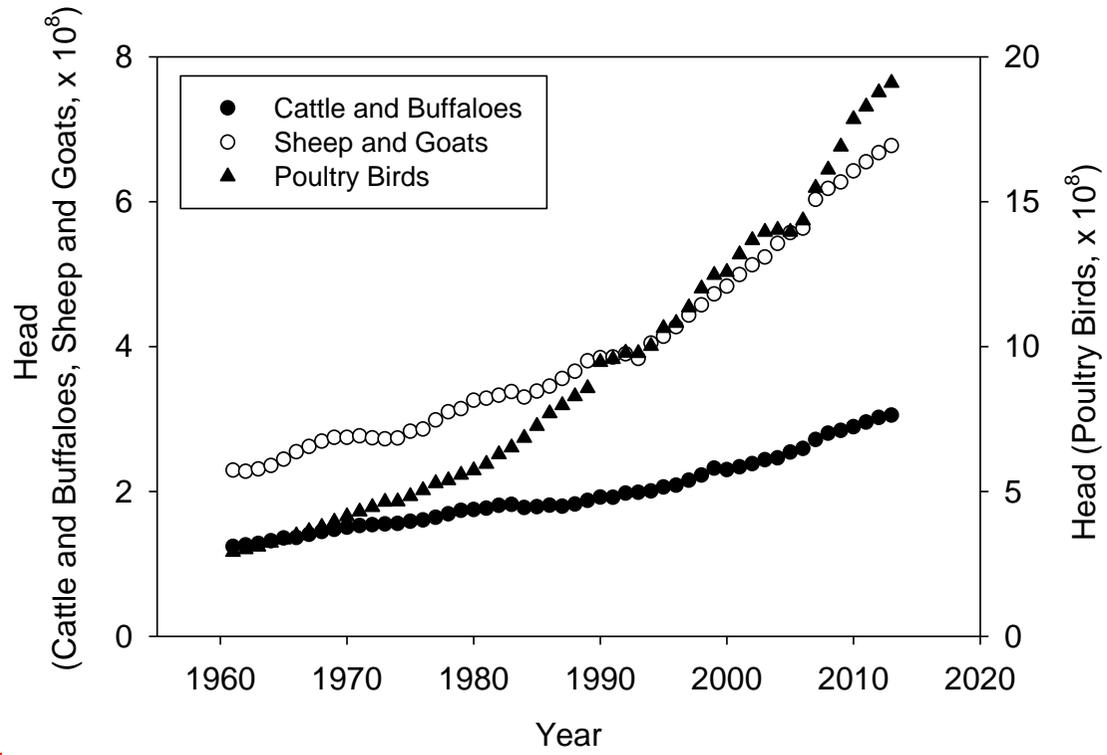
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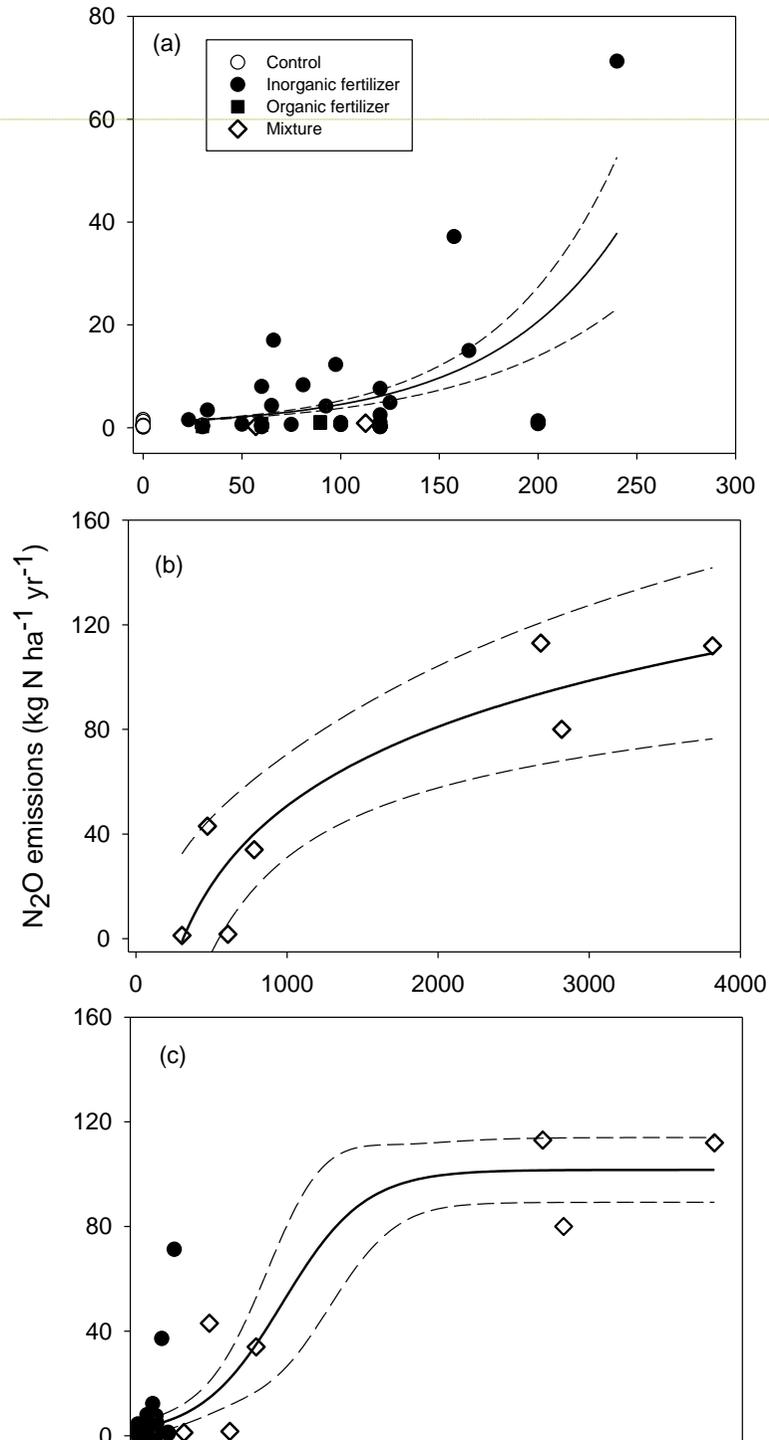
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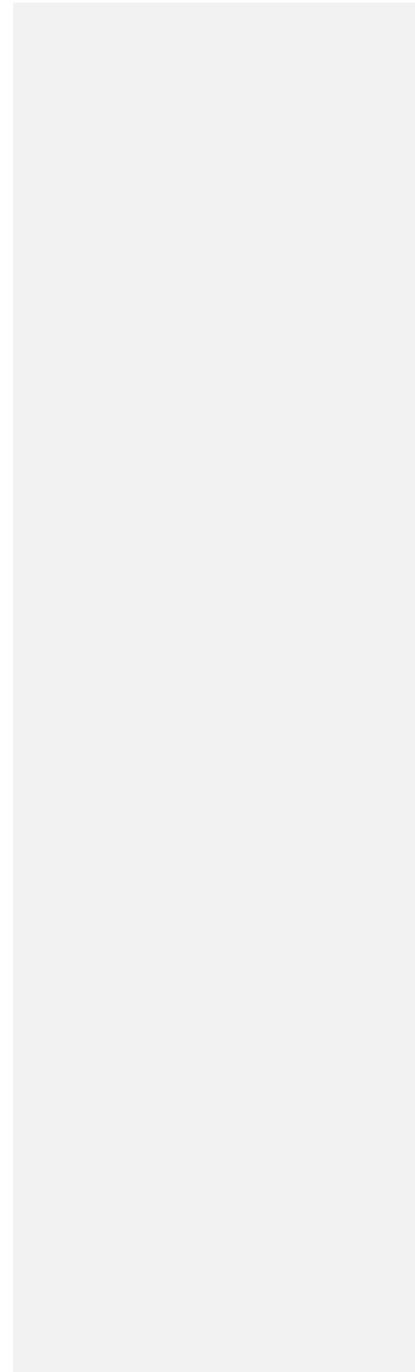
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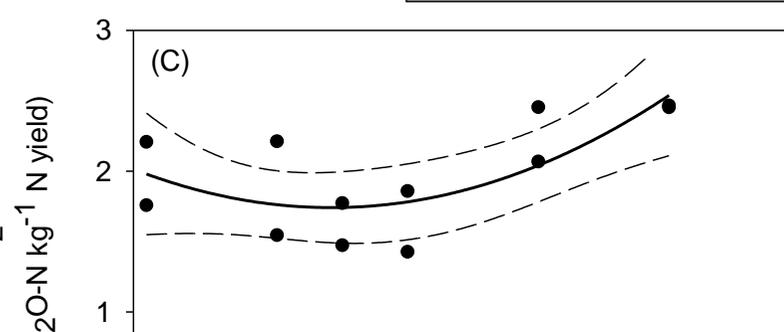
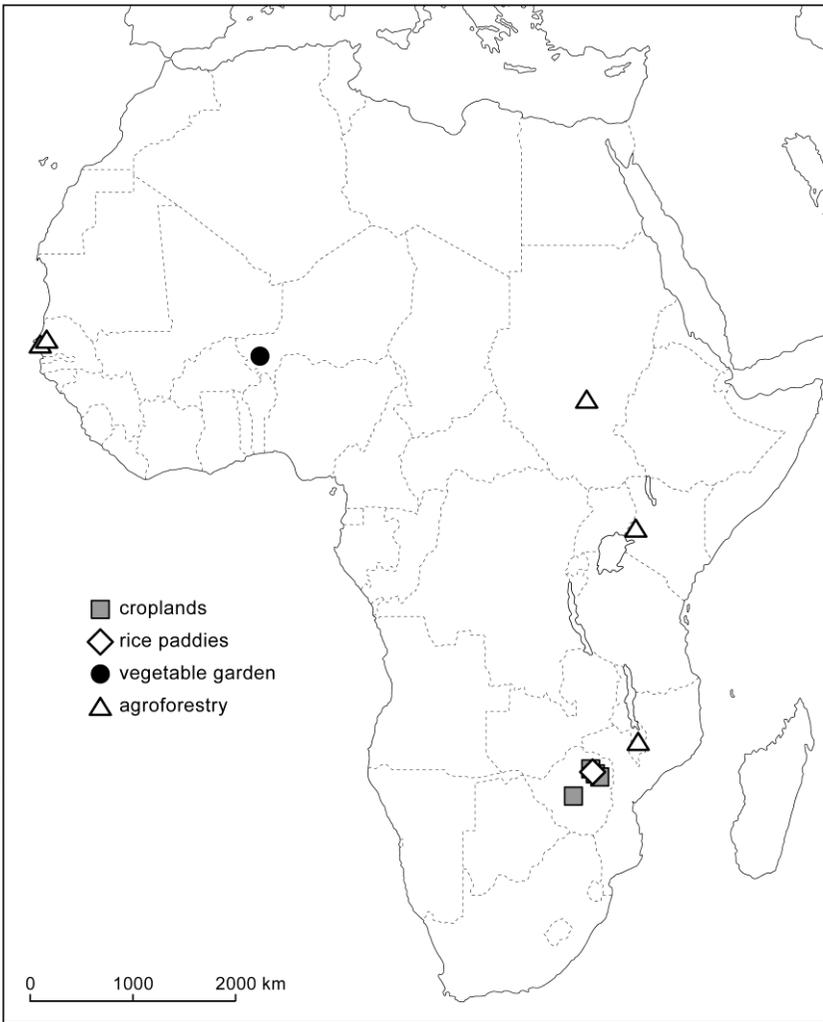
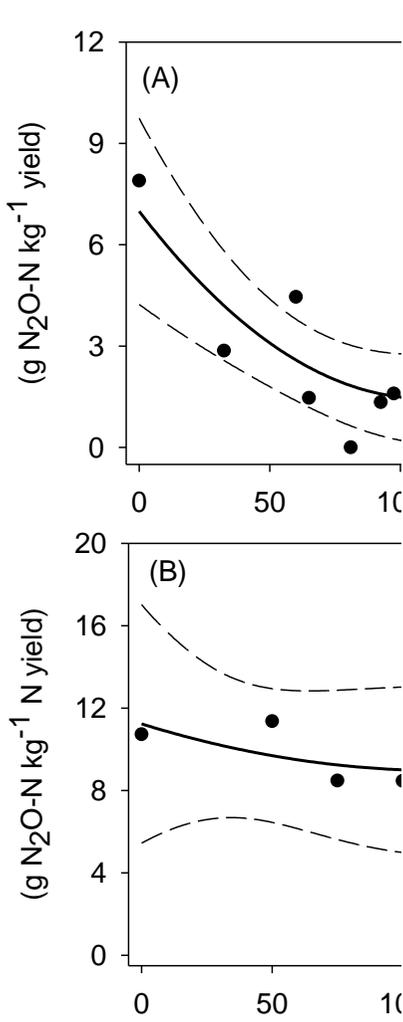
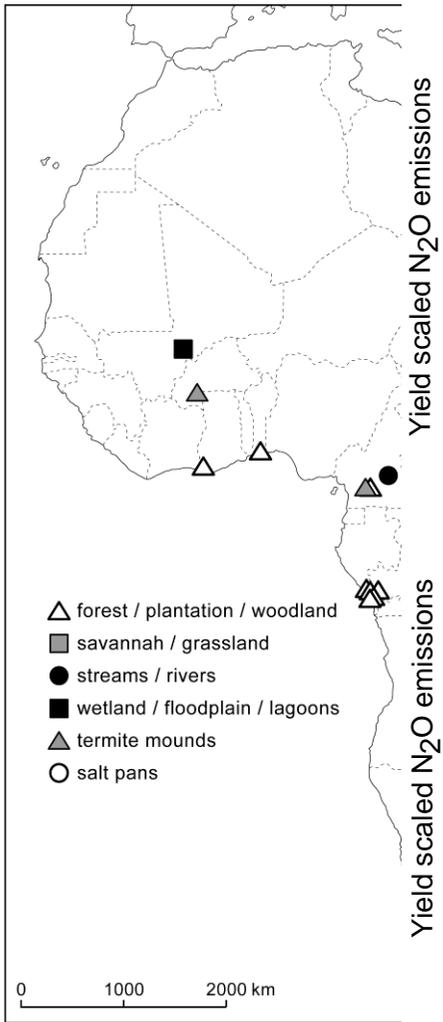
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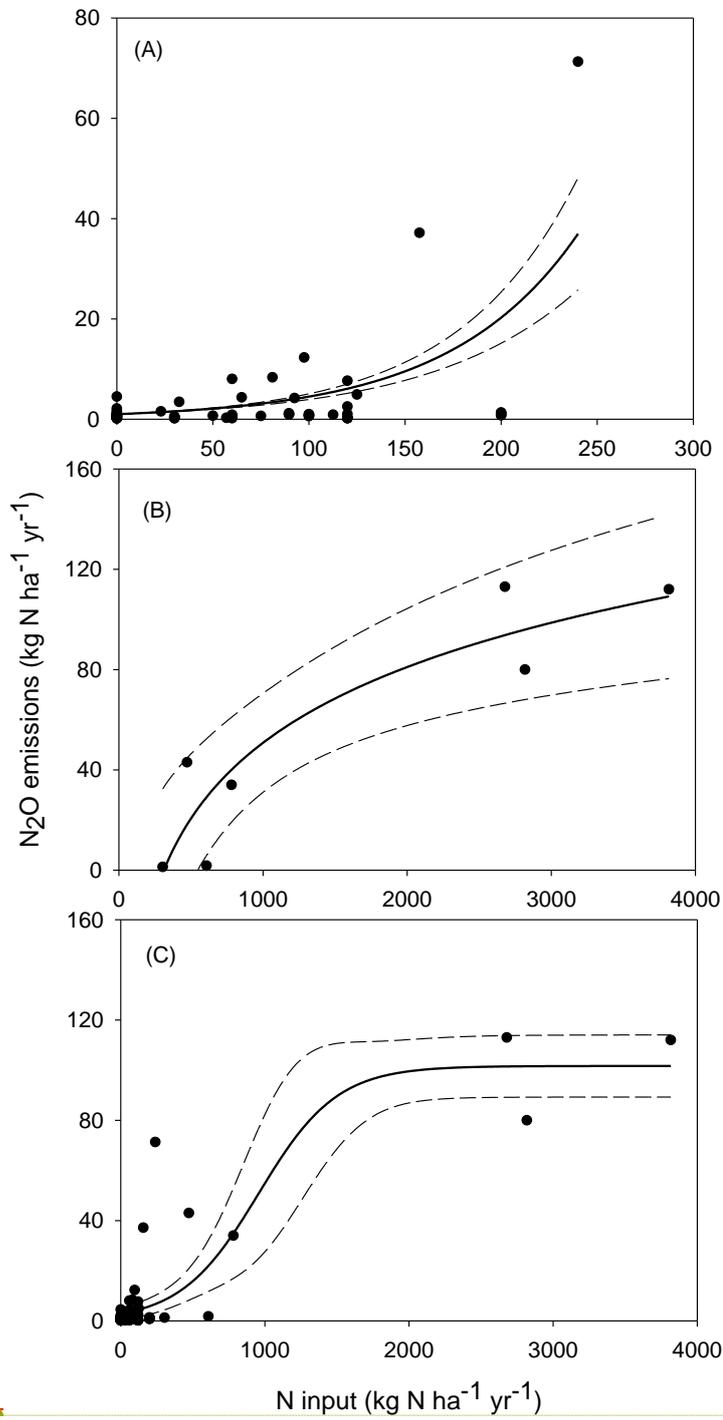




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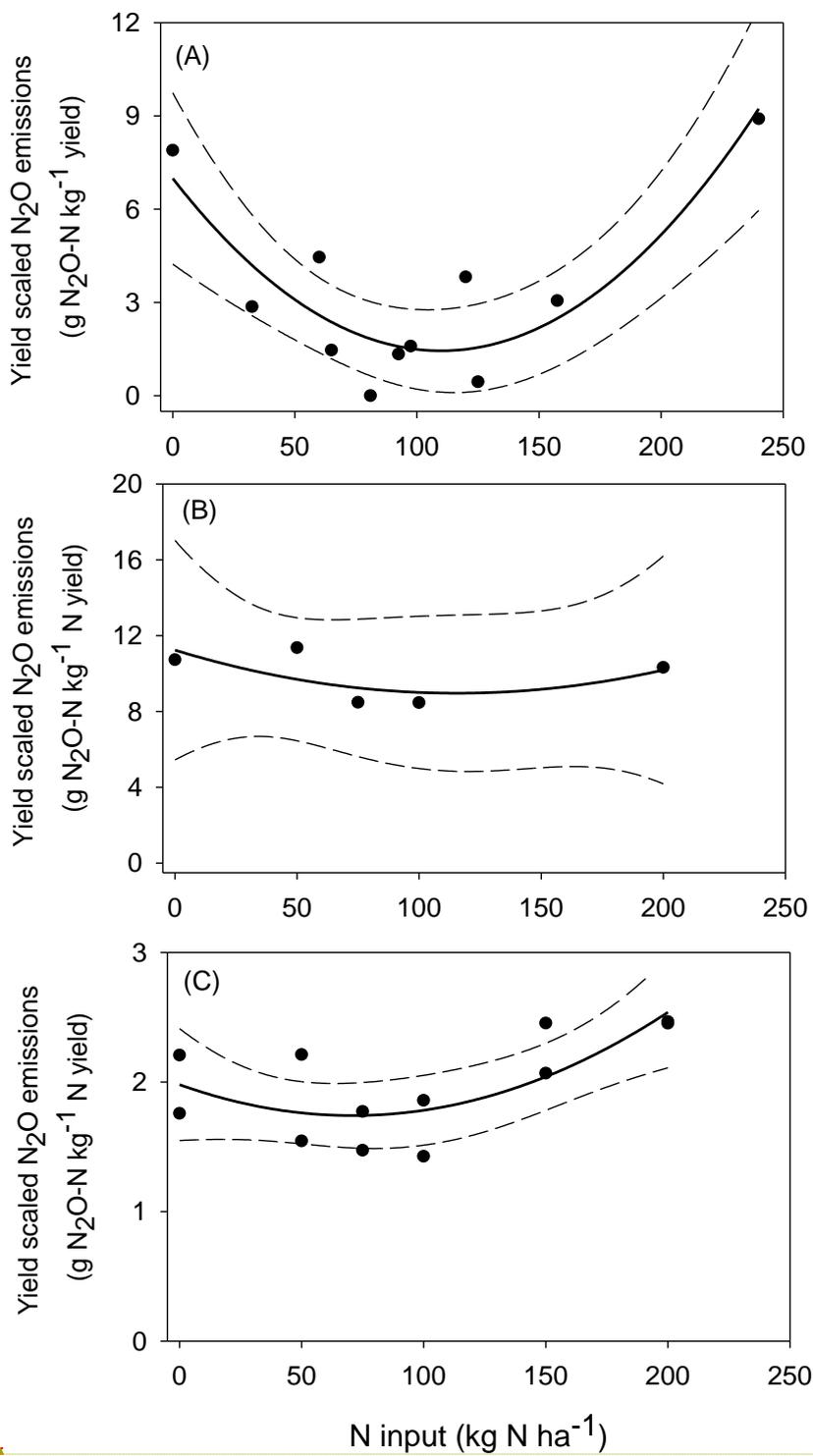
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Figure 5



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Figure 6

Wetlands/floodplains/lagoons /reservoir/lake	<u>Botswana</u> <u>d</u>	<u>*Okavango Delta</u> <u>Botswana</u>	*	*	<u>1480.4 to 1787.0*</u>	<u>≈1480.4 to 1787.0</u>	*	Gondwe and Masamba, 2014
Wetlands/floodplains/lagoons /reservoir/lake	<u>Ivory Coast</u> <u>nd</u>	<u>*Ivory Coast</u>	*	<u>-11.9 to 161.7*</u>	<u>≈-11.9 to 161.7</u>	*	*	<u>Koné</u> et al., 2009
Wetlands/floodplains/lagoons /reservoir/lake	<u>Republic of Congo</u> <u>nd</u>	<u>*Congo River</u>	*	*	<u>246.4*</u>	<u>≈246.4</u>	*	Tathy et al., 1992
Wetlands/floodplains/lagoons /reservoir/lake	<u>Mali</u> <u>nd</u>	<u>*Mali</u>	*	*	<u>3.1*</u>	<u>≈3.1</u>	*	Delmas et al., 1991
Wetlands/floodplains/lagoons /reservoir/lake	<u>Zimbabwe</u> <u>nd</u>	<u>*Zimbabwe</u>	*	<u>65.0 to 232.0*</u>	<u>-26.365.0 to 1235.232.0</u>	<u>0.5-26.3 to 3.51235.2</u>	<u>≈0.5 to 3.5</u>	Nyamadzawo et al., 2014
Wetlands/floodplains/lagoons /reservoir/lake	<u>nd</u>	<u>Lake Kivu</u>	*	*	<u>1.7 to 85.8</u>	<u>2.1 to 6.0</u>	*	<u>Borget et al., 2011 and 2014</u>
Wetlands/floodplains/lagoons /reservoir/lake	<u>nd</u>	<u>Lake Kariba</u>	*	*	<u>*</u>	<u>11 to 7665</u>	*	<u>Delsonro T et al., 2011</u>
Wetlands/floodplains/lagoons /reservoir/lake	<u>nd</u>	<u>Ivory Coast</u>	*	<u>1500 to 1800</u>	<u>*</u>	<u>4.4 to 19.3</u>	*	<u>Koné et al., 2010</u>
Termite mound	<u>Burkina Faso</u> <u>+</u>	<u>29.5Burkina Faso</u>	<u>92629.5</u>	<u>13.5 to 21.3926</u>	<u>3.013.5 to 21.3.7</u>	<u>≈3.0 to 3.7</u>	*	Brümmer et al., 2009
Termite mound	<u>Zimbabwe</u> <u>-</u>	<u>18Zimbabwe</u>	<u>85018</u>	<u>0.002850</u>	<u>0.4002</u>	<u>0.041</u>	<u>≈0.01</u>	Nyamadzawo et al., 2012
Salt pan	<u>Botswana</u> <u>+</u>	<u>*Botswana</u>	*	<u>0.7*</u>	<u>≈0.7</u>	*	*	Thomas et al., 2014

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Symbols: +: methods are good; *: methods are marginal; -: methods are poor to very poor; ?: methods are unclear; nd: cannot comment due to no available criteria

Table S2 Summary of *in situ* carbon dioxide (CO₂), methane (CH₄) and nitrous oxide (N₂O) fluxes and N₂O emission factor (%) in agricultural ecosystems. More detail information is available in 'Soil greenhouse gas emissions in Africa database' (<http://ghginafrica.blogspot.com/>).

Ecosystem type	<u>Quality check†</u>	<u>Country</u> <u>Location</u>	Temperature (°C)	Rainfall (mm)	CO ₂ (Mg CO ₂ ha ⁻¹ y ⁻¹)	CH ₄ (kg CH ₄ ha ⁻¹ y ⁻¹)	N ₂ O (kg N ₂ O ha ⁻¹ y ⁻¹)	N ₂ O emission factor (%)	Reference
Croplands	<u>±</u>	Burkina Faso	29.5	926	9.2 to 16.5	-0.9	*	*	Brümmer et al., 2009

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Croplands	+	Kenya	*	1750	*	*	1.0 to 1.3	*	Hickman et al., 2014
Croplands	+	Kenya	*	1750	*	*	0.2 to 0.5	0.01 to 0.1	Hickman et al., 2015
Croplands	+	Madagascar	16	1300	*	*	0.4	0.47	Chapuis-Lardy et al., 2009
Croplands	+	Malawi	24	930	15.0	*	*	*	Kim, 2012
Croplands	-	Mali	27.6	1100	*	*	0.9 to 2.4	0.3 to 4.1	Dick et al., 2008
Croplands	-	Republic of Congo	23.6	1600	1.7 to 3.7	-1.3 to -1.8	*	*	Delmas et al., 1991
Croplands	-	Senegal	29.7	670	*	*	0.05 to 0.1	*	Dick et al., 2006
Croplands	+	Tanzania	24.5	626 to 905	3.4 to 14.8	*	*	*	Sugihara et al., 2012
Croplands	+	Tanzania	*	*	17.6 to 20.2	-1.7 to 5.6	0.6 to 1.1		Kimaro et al., 2015
Croplands	*	Uganda	21.9	1224	111.1 to 141.2	*	*	*	Koerber et al., 2009
Croplands	*	Zimbabwe	19.1	940	*	*	0.5 to 1.4	*	Rees et al., 2012
Croplands	*	Zimbabwe	*	*	*	*	0.9 to 7.1	*	Chikowo et al., 2004
Croplands	*	Zimbabwe	19.1	940	*	*	0.3 to 0.8	*	Rees et al. 2013
Croplands	-	Zimbabwe	*	*	*	*	0.5	*	Mapanda et al., 2010
Croplands	-	Zimbabwe	18.6	750	19.0 to 44.9	13.4 to 66.7	0.3 to 112.0	*	Nyamadzawo et al., 2014b
Croplands	-	Zimbabwe	18.9	748	1.9 to 10.4	-0.04 to 49.1	0.2 to 3.9	*	Mapanda et al., 2012
Croplands	*	Zimbabwe	21	725	*	*	0.5 to 2.7	0.3 to 1.0	Masaka et al., 2014
Croplands	-	Zimbabwe	*	*	*	3.2 to 11.9	0.8 to 3.5	*	Mapanda et al., 2010
Rice paddy	-	Zimbabwe	18.6	750	6.5	12.5	0.2	*	Nyamadzawo et al., 2013
Vegetable gardens	*	Burkina Faso	27	900	80.7 to 132.0	*	125.7 to 177.6	*	Lompo et al., 2012
Vegetable gardens	*	Niger	30.3	542	73.3 to 100.8	*	53.4 to 176.0	*	Predotova et al., 2010
Agroforestry	-	Senegal	25.8	370	*	*	0.2 to 2.7	*	Dick et al., 2006
Agroforestry	*	Sudan	28.2	698	*	*	23.6 to 26.7	*	Goenster et al., 2014
Agroforestry	+	Kenya	24	1880	*	*	0.3 to 6.4	*	Millar et al., 2004
Agroforestry	+	Malawi	24	930	38.6	*	*	*	Kim, 2012

†Symbols: +: methods are good; *: methods are marginal; -: methods are poor to very poor.

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