

1 Response to referees' comments on "Temperature-mediated changes in microbial carbon use
2 efficiency and ^{13}C discrimination" by CA Lehmeier, F Ballantyne IV, K Min and SA Billings.

3

4 We thank the anonymous referees for their time and their efforts to improve the science and the
5 presentation of our work. Below, we respond to each of the referees' comments (which are in italics).
6 Please note that because of the structure of the Biogeosciences Discussions review process, we do not
7 provide a revised manuscript at this stage but instead provide an indication of where and how we will
8 revise the paper, if given the chance.

9

10 Anonymous referee #2:

11 *General comments:*

12 *"Temperature-mediated changes in microbial carbon use efficiency and ^{13}C discrimination" by Lehmeier
13 and co-authors is a well-written manuscript that addresses questions of interest to a wide range of BG
14 readers. The authors used stable isotope tracers and a flow-through chemostat with a single species and
15 single carbon substrate to identify temperature controls on microbial carbon use efficiency as well as the
16 discrimination against ^{13}C during respiration. This study offers unique insights into the role of
17 temperature for microbial carbon cycling and contributions to ^{13}C - CO_2 signatures.*

18

19 *In fact, while the authors focus on broader applications for soil and terrestrial C cycling, these results are
20 also extremely relevant for freshwater and marine biogeochemistry and microbial ecology. Perhaps even
21 more so given the chemostat conditions, which may be more appropriately applied to aquatic
22 ecosystems. The authors could reach a broader audience by acknowledging this in the language, scope,
23 and citations of the introduction/discussion (sometimes just a matter of deleting "soil").*

24

25 **Response:** Thank you to the reviewer for highlighting that the insights our approach reveals are
26 simultaneously unique yet applicable to diverse systems. We agree with the reviewer that the relevance
27 of our work for other disciplines should be acknowledged more in the manuscript. We will provide
28 context from other disciplines in the introduction, and weave this idea into our discussion as well. We
29 have selected multiple citations useful for these changes (e.g. Goldman & Dennet, 2000; Pomeroy &
30 Wiebe, 2001; Chrzanowski & Grover, 2008; Hall et al., 2008).

31

32

33

34 *The chemostat set-up and equilibrium assumptions are very clearly described. I do believe the authors
35 could be more up-front about the unknowns associated with equal labeling of cellulose and glucose
36 within the labeled cellobiose substrate, and what this might mean for the interpretation of the
37 discrimination results (there are hints of this in EEA methods and results/discussion, but this seems to be
38 an unknown with significant consequences for results).*

39

40 **Response:** The $\delta^{13}\text{C}$ value of repeated measurements of the cellobiose substrate was -24.2‰ with little
41 variation (± 0.04 1SE, $n=12$, Fig. 4), showing that the substrate was a homogeneous mixture. (at one
42 stage in the manuscript, we wrote -22.4‰ for cellobiose $\delta^{13}\text{C}$; this was a typo that will be changed.)
43 Measured against VPDB standard, the $\delta^{13}\text{C}$ of -24.2‰ implies a $^{13}\text{C}/^{12}\text{C}$ ratio in the sample of $\sim 1/91$.

44 Considering the molecular formula of cellobiose $C_{12}H_{22}O_{11}$, this means that not more than about one out
45 of eight cellobiose molecules in the supplied substrate had a ^{13}C atom incorporated.
46 We have not determined at which positions a cellobiose molecule may have had that ^{13}C atom
47 incorporated. We have, however, confirmation from the supplier (Sigma-Aldrich) that the cellobiose we
48 used was of biogenic origin. Evidence from work on carbon isotope distribution within carbohydrate
49 molecules (e.g., Rossmann et al., 1991; Gleixner & Schmidt, 1997; and references therein) suggests non-
50 random distribution of ^{13}C in biological molecules, probably mainly caused by kinetic isotope effects in
51 enzymatic reactions (see also Tcherkez et al., 2012). Based on this phenomenon, the ^{13}C atom in a ^{13}C -
52 containing cellobiose molecule we supplied to the microbes will not be randomly distributed within the
53 molecule but rather consistently on the same position. In our view, this rules out the possibility that any
54 changes in the ^{13}C distribution within the cellobiose substrate supplied to the microbes were responsible
55 for the $\delta^{13}C$ patterns in biomass and respired CO_2 we observed (Fig. 4).

56
57 There are still unknowns in isotope effects during glycolysis and respiration (Tcherkez et al., 2012), but
58 considering the literature that certain enzymatic reactions in metabolic pathways discriminate against
59 ^{13}C in organic molecules in a predictable way, we believe that our discussion offers parsimonious
60 explanations for the observed isotope patterns in biomass and respired CO_2 . In a revision, we will
61 include these considerations.

62
63
64

65 *Are there recommendations for how these results can be applied to non-steady state scenarios in*
66 *heterogeneous soil or biofilm matrices? This study is novel and useful, but drawing connections from*
67 *steady-state chemostat measurements to the real world remains a challenge.*

68

69 **Response:** We agree with the reviewer in that it would be very interesting to have more information
70 about how microbial C and C isotope fluxes during the transformation of organic matter change with
71 environmental/growth conditions.

72

73 In our experiments, maintaining microbial growth in steady-state was critical for permitting direct
74 measurement of respiration rates and the $\delta^{13}C$ of respired CO_2 , without the confounding influences that
75 measurements in less controlled environments entail. The steady-state growth was hence a means to
76 obtain data that would otherwise be only very difficult if not impossible to obtain, and to study the
77 effect of temperature on microbial C and C isotope fluxes in "isolation." That is, we strived to maintain
78 all other environmental factors constant except temperature.

79

80 Microbial experiments in controlled environments always reveal only a snapshot of specific growth and
81 environmental conditions, and accurate generalizations therefore are difficult to make. However, we
82 have no reason to assume that the principles and mechanisms about the temperature response of C and
83 C isotope fluxes identified in our work wouldn't happen in the "real world" (like, a heterogeneous soil or
84 an aquatic environment). Because our work used a simplified system to explore fundamental microbial
85 respiration rates and respiratory C isotope discrimination during organic matter transformations with
86 varying temperature, it serves as a starting point from which we can formulate new hypotheses, and
87 explore, for instance, the reason for the presumably strong discrimination against ^{13}C -containing
88 cellobiose molecules in the supplied substrate.

89

90

91

92 *Specific comments:*

93 *Page 17372, line 15 - Is a 1:1 respiratory quotient appropriate for both cellulose and glucose? I would*
94 *guess cellulose RQ > 1, while glucose RQ_ 1.*

95

96 **Response:** From our data we obtained and discussed some circumstantial evidence that Cellobiose was
97 cleaved intercellularly, and that the cleavage products, two Glucose units, then could serve as substrate
98 for respiration. Considering that the group of carbohydrates is generally ascribed an RQ of 1, we used
99 an RQ of 1 at that point of the manuscript to highlight that the supply rate of O₂ to the reactor should
100 have exceeded the consumption of O₂ by the microorganisms by far, and thus allow for aerobic
101 metabolism of the population in the reactor. If cellobiose had an RQ > 1, then the molar amount of O₂
102 consumed would be lower than the molar amount of CO₂ produced, which should support the point of
103 aerobic metabolism in our chemostat reactors even more.

104

105

106

107 *Figures –What is the uncertainty of the results presented in Figures 3, 4, 6?*

108

109 **Response:**

110 The data presented in these figures are from seven independent chemostat runs, conducted at seven
111 different temperatures. Therefore, we were unable to provide error bars that would indicate any
112 variation across true experimental replicates. However, in a revision, we could provide measurement
113 uncertainties for the concentration and δ¹³C of respired CO₂, based on large numbers of measurements
114 of reactor headspace CO₂ concentrations and δ¹³C at steady-state.

115

116 (please see also below our response to the referee #3 critique to the lack of replicates.)

117

118

119

120 *Figure 5 could more clearly identify the knowns/unknowns beyond boxed and unboxed. Perhaps two*
121 *panels to show the difference in (a) steady-state chemostat versus (b) soil measurements and the*
122 *unknowns/challenges for moving forward?*

123

124 **Response:** In a revision, we will modify Figure 5 in response to this comment so that differences
125 between steady-state chemostat vs. soil measurements and the unknowns/challenges for moving
126 forward become clearer.

127

128

129

130 *Anonymous referee #3:*

131 *This manuscript reports on results from a series of chemostat-based (at steady state) measurements of*
132 *respiration and d13C of CO2 from cultures of P. fluorescens at various temperatures. The objective is to*
133 *study the temperature dependence of microbial CUE and C isotope discrimination. It is clear that solid*

134 *measurements of CUE are very difficult and thus the understanding of the factors that influence it. It is*
135 *also clear that the microbial fractionation of C isotopes is a big unanswered question that keeps us*
136 *from being able to use it to help us understand the C cycle. I applaud the effort by the investigators to try*
137 *to tackle these very difficult questions and think that there is promise in the approach. However, I see a*
138 *fair number of aspects that add too much uncertainty to the findings and their interpretations.*

139
140 *The points I consider more critical are explained below and are followed by other secondary but also*
141 *important issues. SGR, SRR, CUE data: CUE is defined as the fraction of SGR out of the sum of SGR and*
142 *SRR. By definition, the reactor dilution rate is the SGR. (There is not a reference to support this*
143 *approach).*

144
145 **Response:** We appreciate the referee's thoughtful responses to our manuscript. We would like to
146 highlight here that on pages 17373 and 17374 of the submitted manuscript we provide references to
147 Bull (2010) and Ferenci (2008), in which multiple aspects of continuous culture experiments are
148 described, including the notion that the dilution rate of the reactor equals the growth rate of the
149 microbial population at steady-state. In crafting our manuscript, we attempted to find a balance
150 between repeating the theory already established from these and many earlier papers and the
151 application of the concept in our experiments to our readers; perhaps we erred in being too
152 conservative with our explanations.

153 In a revision, we will add the citations of Egli (2015) and Dawson (1974) which are further sources of
154 information to continuous culture techniques, as well as a reference to Smith & Waltman (1995) who
155 provide a lucid description of the theory underlying chemostat dynamics. We also will rephrase our
156 sentences describing how dilution rate of a reactor equals microbial growth rate at steady state for
157 greater clarity.

158 Importantly, we had highlighted at several stages in the manuscript that our CUE calculations do not
159 take into account any possible exudation of organic compounds (see also below), particularly via Figure
160 5 and in the text. We discuss this omission in the text. We also highlight that this formulation permits
161 comparison with inferences about CUE in the soils literature that emanate from less well-controlled
162 environmental samples and that use the same formulation.

163
164
165
166 *The steady state of the culture is maintained by maintaining the SGR. Because SGR is maintained*
167 *constant, CUE will always be explained by changes in the SRR (CO₂ rate per unit of microbial*
168 *biomass). Is it really possible to independently assess effects of temperature on specific growth rates,*
169 *when they will be inevitably determined by the respiration rate and the microbial biomass? Thus is it*
170 *really possible to estimate CUE?*

171
172 **Response:** We agree with the reviewer that CUE of the continuous culture at steady-state, defined as
173 $SGR/(SGR+SRR)$, will always be explained by changes in SRR, when SGR is kept constant.
174 We would like to highlight that we intentionally took advantage of the established principles of
175 continuous culture techniques, that it is the dilution rate of the reactor medium with fresh tank medium
176 that determines the specific growth rate of the microbial population (see above and the cited
177 references).

178 At steady-state, as documented by constant reactor headspace CO₂ concentrations, we can thus
179 quantify temperature effects on metabolic/physiologic C partitioning i.e., the proportion of substrate
180 respired per unit time (SRR) vs. the proportion incorporated into biomass per unit time (SGR).

181 Measurements of SRR allowed us to compute the parameter CUE, as it is often done in the literature.
182 Because we use specific respiration rates, temperature effects on respiration rates of the populations
183 and on microbial biomass contents in the reactor at steady-state are both considered.

184
185
186

187 *The authors state that “The 50% reduction in steady state dry microbial biomass with increasing*
188 *temperature was due to 2.5 fold increase in SRR”. In my view, given the nature of the thermostat system,*
189 *the result is rather that SRR increased with temperature due to the decrease in microbial biomass with*
190 *temperature. There’s a circularity that complicates the interpretation of these variables when combined.*
191 *In my view, the relationship of SRR and microbial biomass with temperature can be explored with more*
192 *confidence than CUE and it is valuable that it was done at steady state.*

193

194 **Response:** We agree with the referee that the wording “the 50% reduction in steady-state microbial
195 biomass with increasing temperature was due to a 2.5 fold increase in SRR” may be overly constraining
196 cause and effect. Given that logic, we would also like to avoid saying that the increase in SRR is “due to
197 the decrease in microbial biomass with temperature”. In a revision, we would reword this sentence and
198 say that the increase in SRR with temperature integrates changes in absolute respiration rates of the
199 population as well as changes in microbial biomass with temperature.

200

201

202

203 *The extent of the impact of secretion/waste on the estimation of uptake (SGR+SRR) is difficult to*
204 *constrain realistically and also its variation with temperature without any measurement of what was*
205 *actually in the solution after filtration. How about the contribution of the further uptake and respiration*
206 *of those substrates? (on a somewhat related note: the lack of enzymatic activity in the solution may*
207 *mean that the enzymes are being quickly uptaken and thus are not detectable; the current interpretation*
208 *of the lack of enzymes is very speculative). Because of these uncertainties the overall interpretations and*
209 *general discussion on the effect of temperature on CUE are challenging to make.*

210

211 **Response:** We agree with the referee in that our estimation of uptake as SGR + SRR is an approximation
212 of total uptake, as it does not take into account any possible exudates. We had aimed at making this
213 very clear and upfront in the manuscript (pages 17377 and 17378). Our intended revisions to Fig. 5 (see
214 comment to referee #2, above) should help enhance the clarity of this point. Nevertheless, our
215 presentation of CUE reveals meaningful information, as it is defined in our manuscript in the same way
216 as in many other studies from diverse disciplines, yet it incorporates direct observations of fluxes that
217 cannot be easily measured using environmental samples, if at all.

218 Even if we don’t have measured rates of exudation, the isotope data and published literature provide
219 points of orientation (pages 17378 and 17379) to suggest that exudation was not the major sink for
220 exuded ¹³C; for this to happen, the degree of metabolic fractionation would have to be unnaturally high
221 (as discussed on page 17379). To fully counteract the negative correlation of temperature and CUE, the
222 exudation flux F_{SER} would have had to increase at the same magnitude with decreasing temperature as
223 F_{SRR} decreased with decreasing temperature. From our data, and from the viewpoint of temperature
224 effects on metabolic rates (e.g., Gillooly et al., 2001) this seems very unlikely.

225 We cannot rule out that the microbes may have taken up secreted C-based metabolic compounds,
226 although this does not appear to be an energetically favorable resource use strategy of the microbes in
227 these continuous culture conditions. Considering the points above, if such a “recycling flux” would have
228 occurred, its effect on the overall CUE estimation should be little, as it would represent a continuous
229 flow of C excreted from, and then re-consumed by, the microbes. Regardless, if the microbes exuded
230 meaningful amounts of the exo-enzymes whose activity we sought, the steady-state nature of the
231 system dictates that this exudation flux would be continuous, and hence detectable. However, whether
232 or not the microbes exuded these enzymes does not affect any major conclusion of this work.

233
234
235

236 *Isotopic discrimination data: in my view there is too much uncertainty in what happened with the C*
237 *during the experiments and this is combined with various unexpected hard- to- explain observations. A*
238 *full budget approach, accounting for all pools (both their size and isotopic composition), including*
239 *inorganic C, dissolved organic (not cellobiose) could have potentially allowed explanations to allow*
240 *confident interpretations. The uncertainty in the potential reasons for the very d13C values in the early*
241 *incubations, and more importantly the reasons for the gradual change towards ‘real’ values is a problem.*
242 *We don’t know to what extent the processes at play during the ‘climbing’ phase are still at play during*
243 *the plateau. The carbonate system explanation for the early stages would need measurement of the pool*
244 *of inorganic C.*

245
246 **Response:** We agree with the referee that more data would be helpful for constraining $^{13}\text{C}/^{12}\text{C}$ flows
247 during the microbial transformation of cellobiose in this system. However, we would like to differentiate
248 the two categories of data here – inorganic and organic C.

249 First, when designing the system, we recognized the importance of inorganic C fluxes as drivers of both
250 the concentration and the isotopic signature of the CO_2 we measured in the reactor headspace. As such,
251 we provided a detailed description of how the inorganic C system proceeds to chemical and isotopic
252 equilibrium in the reactor in the Supplementary Material part of the current submission. These efforts in
253 the Supplementary Material – including references to the literature and experimental data – clearly
254 show that knowing the size and C isotopic composition of the inorganic C pools (other than CO_2 in
255 reactor headspace) is not necessary for the accurate assessment of microbial respiration rates and the
256 isotopic signature of respired CO_2 at steady-state. The Supplementary Material also offers an
257 explanation for the gradual changes in concentration and $\delta^{13}\text{C}$ of the reactor headspace CO_2 during the
258 “climbing phase”, and that these changes do not pose any problem to our measurements of microbial C
259 fluxes at steady-state, where all C pools (including the inorganic C pools in the reactor medium) are in
260 chemical and isotopic equilibrium. In a revised manuscript version, we will add more references to this
261 Supplementary Material to ensure that readers are aware that we address this issue in full.

262 Second, we agree with the referee that knowing the pool sizes and $\delta^{13}\text{C}$ of any microbial exudates would
263 be ideal. Such data are not at all trivial to obtain. Nevertheless, we attempted to account for microbial
264 exudation by measuring the $\delta^{13}\text{C}$ of reactor filtrate (Fig. 4), but unfortunately we were searching for a
265 relatively small change in a large pool of unconsumed cellobiose, and we could not detect an increase in
266 the $\delta^{13}\text{C}$ of the filtrate compared to the $\delta^{13}\text{C}$ of the cellobiose (see below). Such an increase would
267 indicate what must be the presumed fate(s) of the ^{13}C in our system – isotopic enrichment of the
268 unconsumed cellobiose in the reactor waste line, and/or isotopic enrichment of exudates. The

269 discussion in the text describes how discrimination against ^{13}C during exudation is not a reasonable
270 means of generating the isotopic patterns we observed in isolation from discrimination during uptake,
271 but even knowledge of a very different $\delta^{13}\text{C}$ of the filtrate wouldn't have allowed us to parse these
272 processes (exudation and uptake fractionation).

273

274

275 *The observation that microbial biomass was depleted in ^{13}C relative to the substrate is surprising given*
276 *findings of previous much simpler studies. If they are depleted we would expect enrichment of the*
277 *respiration, which what not the case. What accounts for the further depletion of the respiration is too*
278 *hard to explain and it is seriously speculative. It is surprising that the $d^{13}\text{C}$ values of the filtrate are not*
279 *presented and that an attempt to partition is not done. They suggest microbial discrimination against*
280 *heavy (enriched) substrate. With a $d^{13}\text{C}$ value of -24, the atom percent of ^{13}C is very low. Could the*
281 *actual amount of potential ^{13}C to discriminate against explain the actual degree of enrichment? Again, a*
282 *budget approach would have helped here. The observation of strong fluctuations in the microbial and*
283 *respiration with temperature is very (very) hard to explain and grasp and the current attempt is highly*
284 *speculative.*

285

286 **Response:** We agree that the fate of the “missing” ^{13}C in our system is not directly observed. We address
287 this in our response to the “Isotopic discrimination” comment above. Briefly, we attempted to do what
288 the referee calls for, and indeed present $\delta^{13}\text{C}$ values of the filtrate from four out of the seven chemostat
289 runs in Fig. 4a of the manuscript. As described in the manuscript, the values do not differ significantly
290 from the $\delta^{13}\text{C}$ values of the cellobiose. This reflects our estimates that roughly 95% of the cellobiose that
291 is fed to the cultures is not consumed by the reactor populations at all temperatures, hence, the $\delta^{13}\text{C}$
292 value of the filtrate must be very close to that of the main organic component cellobiose (see lines 15
293 and following on page 17378). Thus, we were searching for a small change in a relatively large pool. Our
294 data suggest that 1) the amount of exudates in the filtrate was low compared to the amount of
295 cellobiose and 2) any ^{13}C enrichment of exudates (as one possible fate of the “missing” ^{13}C) would not
296 have been large enough to make the $\delta^{13}\text{C}$ value of the filtrate significantly different from that of the
297 cellobiose (see also our isotopic mass balance considerations on page 17379).

298

299 We agree that $\delta^{13}\text{C}$ data for organic compounds other than cellobiose in the filtrate would be most
300 desirable and would allow us to come closer to the full budget approach the referee mentions.
301 However, we emphasize that our experimental system provides data that are the closest our community
302 has come to direct observations of these C and C isotope fluxes. Because there is not much comparable
303 information in the literature, opportunities for discussion are minimal.

304 We contend that our work provides a robust example for how science progresses: it produces novel data
305 that have not been presented in such a manner before, and it offers both novel information for
306 researchers and hypotheses for further experimentation. Indeed, we are currently expanding on the
307 work by performing experiments in which we explore the reason for the presumably strong
308 discrimination against ^{13}C -containing cellobiose molecules in the substrate.

309

310

311 *I also wonder what would have happened if the runs had been replicated and the contribution of*
312 *experimental error to the ups and downs.*

313
314 **Response:** We agree with the reviewer that performing replicate chemostat runs would be desirable.
315 With a certain capacity for experimentation, we decided to increase the range of independent
316 chemostat runs at different temperatures instead of performing a higher number of experimental
317 replicates at fewer temperatures. This trade-off is very common in chemostat-studies, suggested by the
318 fact that a lack of true replicates can frequently be found in chemostat literature. Though perhaps not
319 ideal, the approach has permitted chemostat enthusiasts to draw viable conclusions about microbial
320 populations for decades.
321 Importantly, we performed the chemostat runs in random order of temperature (page 17372). This
322 means that the temperature response of specific respiration rate, CUE and $\Delta^{13}\text{C}$ (Figs. 3, 4 & 6) were not
323 due to any systematic drifts in measurements or experimentation with time.

324
325
326
327 *The combined uncertainties in the isotope data and metabolism data then make the discussion on the*
328 *relationship between them a bit of a stretch.*

329 **Response:** We agree with the referee that our manuscript highlights several points that deserve further
330 exploration. We never make the claim that our results are ubiquitous. However, rather than highlighting
331 the lack of complete answers, we agree with the referee in that there is promise in our novel approach.
332 Please see our responses to the above concerns about our statements on isotopic discrimination and
333 CUE.

334

335

336 *Introduction -More background on the connection between metabolism and isotopic discrimination*
337 *would be nice to have. -There's not sufficient background on the factors that may drive microbial C*
338 *discrimination.*

339
340 **Response:** In a revised version, we will aim to satisfy this comment in the Introduction and provide a bit
341 more information about C isotope discrimination and metabolism (see also the responses to referee #2,
342 above). However, we will refrain from expanding on this point in the discussion given that we do not
343 provide data showing metabolic fluxes.

344

345

346

347 *Methods -what is the material of the filters? Could filter adsorb some molecules that the "removal" of*
348 *material from the filter would not get?*

349 **Response:** The material of the filters we used was polyethersulfone, and we will add this information to
350 the methods section. For the collection of dry microbial biomass for elemental analysis, we did not aim
351 at removing all microbial biomass from a filter, as this could have potentially contaminated the biomass
352 with filter material and hence affected measured C contents and $\delta^{13}\text{C}$ of microbial biomass. For the
353 determination of microbial dry weight on a filter, we subtracted the mass of the dry, clean filter from
354 the mass of the dry filter including dried microbial biomass. Whether there was microbial biomass

355 adsorbed by the filters or not is hence irrelevant for the estimation of microbial dry weight on a filter. In
356 a revision, we will amend the section accordingly.

357

358

359 Cited literature:

360 Bull, A.T.: The renaissance of continuous culture in the post-genomics age, *J. Ind. Microbiol. Biotechnol.*,
361 37, 993-1021, 2010.

362 Chrzanowski, T.H., and Grover, J.P.: Element content of *Pseudomonas fluorescens* varies with growth
363 rate and temperature: A replicated chemostat study addressing ecological stoichiometry, *Limnol.*
364 *Oceanogr.*, 53, 1242-1251, 2008.

365 Cotner, J.B., Makino, W., and Biddanda, B.A.: Temperature affects stoichiometry and biochemical
366 composition of *Escherichia coli*, *Microb. Ecol.*, 52, 26-33, 2006.

367 Dawson, P.S.S.: *Microbial Growth*, Dowden, Hutchinson & Ross, Stroudsburg, 1974.

368 Egli, T.: Microbial growth and physiology: a call for better craftsmanship. *Front. Microbiol.*, 6:287, DOI:
369 10.3389/fmicb.2015.00287, 2015.

370 Ferenci, T.: Bacterial physiology, regulation and mutational adaption in a chemostat environment, *Adv.*
371 *Microb. Physiol.*, 53, 169-230, 2008.

372 Gillooly, J.F., Brown, J.H., West, G.B., Savage, V.M., Charnov, E.L.: Effects of size and temperature on
373 metabolic rate, *Science*, 293, 2248-2251, 2001.

374 Gleixner, G., and Schmidt, H.-L.: Carbon isotope effects on the Fructose-1,6-bisphosphate aldolase
375 reaction, origin for non-statistical ¹³C distributions in carbohydrates, *J. Biol. Chem.*, 272, 5382-5387,
376 1997.

377 Goldman, J.C., and Denner, M.R.: Growth of marine bacteria in batch and continuous culture under
378 carbon and nitrogen limitation, *Limnol. Oceanogr.*, 45, 789-800, 2000.

379 Hall, E.K., Neuhauser, C., and Cotner, J.B.: Toward a mechanistic understanding of how natural bacterial
380 communities respond to changes in temperature in aquatic ecosystems. *ISME J.*, 2, 471-481, 2008.

381 Pomeroy, L.R., and Wiebe, W.J.: Temperature and substrates as interactive limiting factors for marine
382 heterotrophic bacteria. *Aquat. Microb. Ecol.*, 23, 187-204, 2001.

383 Rossmann, A., Butzenlechner, M., and Schmidt, H.-L.: Evidence for a nonstatistical carbon isotope
384 distribution in natural glucose, *Plant Physiol.*, 96, 609-614, 1991.

385 Smith, H., and Waltman, P.: The theory of the chemostat: dynamics of microbial competition. Cambridge
386 University Press, 1995.

387 Tcherkez, G., Mahé, A., and Hodges, M.: 12C/13C fractionations in plant primary metabolism. Trends
388 Plat Sci., 16, 499-506, 2012.

389

390

391

392

393

394

395

396

397

398

399

400

401

402

403

404 **Title:** Temperature-mediated changes in microbial carbon use efficiency and ^{13}C discrimination

405

406 **Author list and affiliations:**

407 C. A. Lehmeier¹, F. Ballantyne IV^{1,2}, K. Min¹, S. A. Billings^{1*}

408 ¹Department of Ecology and Evolutionary Biology, Kansas Biological Survey, University of
409 Kansas, 2101 Constant Ave., Lawrence, KS 66047, USA.

410 ²now: Odum School of Ecology, University of Georgia, 140 E. Green St., Athens, GA 30602,
411 USA.

412 ***Corresponding author:**

413 Sharon A. Billings, Kansas Biological Survey, University of Kansas, 2101 Constant Ave.,
414 Lawrence, KS 66047, USA. Tel.: 001-785-864-1560, Fax: 001-785-864-1534, email:

415 Sharon.Billings@ku.edu

416

417

418

419

420

421

422 Abstract

423 | Understanding how carbon dioxide (CO₂) flux from ~~soils~~soils-ecosystems feeds back to climate
424 | warming depends in part on our ability to quantify the efficiency with which microorganisms
425 | convert ~~soil~~ organic carbon (C) into either biomass or CO₂. Quantifying ecosystem-level
426 | respiratory CO₂ losses often also requires assumptions about stable C isotope fractionations
427 | associated with the microbial transformation of ~~soil~~ organic substrates. However, the diversity of
428 | organic substrates' δ¹³C and the challenges of measuring microbial C use efficiency (CUE) in
429 | their natural environment ~~soils~~ fundamentally limit our ability to project ~~soil, and thus~~
430 | ecosystem, C budgets in a warming climate. Here, we quantify the effect of temperature on C
431 | fluxes during metabolic transformations of cellobiose, a common microbial substrate, by a
432 | cosmopolitan ~~soil~~ microorganism growing at a constant rate. Specific respiration rate increased
433 | by 250% between 13 °C and 26.5 °C, decreasing CUE from 77% to 56%. Specific respiration
434 | rate was positively correlated with an increase in respiratory ¹³C discrimination from 4.4‰ to
435 | 6.7‰ across the same temperature range. This first demonstration of a direct link between
436 | temperature, microbial CUE and associated isotope fluxes provides a critical step towards
437 | understanding δ¹³C of respired CO₂ at multiple scales, and towards a framework for predicting
438 | future ~~soil~~soil-ecosystem C fluxes.

439

440 1 Introduction

441 | Because Earth's C cycle is a key regulator of climate, a central goal of biogeochemistry is to
442 | understand ~~terrestrial~~ biosphere-atmosphere C exchange. Globally, almost all C initially

443 | assimilated via photosynthesis is respired ~~by plants and soil microorganisms~~ back to the
444 | atmosphere as CO₂ by auto- and heterotrophic organisms (Schimel, 1995; Trumbore, 2006).
445 | Though we have a reasonably comprehensive understanding of how environmental conditions
446 | influence ~~photosynthetic~~ CO₂ uptake by ~~plants~~photosynthetic organisms, our understanding of
447 | how respiratory CO₂ fluxes respond to environmental conditions significantly lags behind. This
448 | is especially true for respiratory CO₂ derived from heterotrophic ~~soil microorganisms in aerobic~~
449 | ~~conditions~~, which may account for ~~~65%~~more than half of ~~the respiratory~~ CO₂ losses ~~es~~
450 | and aquatic systems (Kucera and Kirkham, 1971; Hanson et al., 2000; Cotner and Biddanda,
451 | 2002; Subke et al., 2006). Metabolic rates of ~~heterotrophs soil microbes~~ are expected to
452 | increase with rising temperatures (Gillooly et al., 2001; Pomeroy and Wiebe, 2001; Hall et al.,
453 | 2008), which is of great concern given ~~the Earth's~~ large reservoir of ~~C substrates in soil reduced~~
454 | organic matter (~~SOM~~) that may be mineralized to CO₂ via metabolism (Hedges et al., 2000;
455 | Kirschbaum, 2006). The influence of temperature on the physiology of ~~soil heterotrophic~~
456 | microbes must therefore be well understood to project shifts in the ~~terrestrial global~~ C balance in
457 | a warmer climate.

458 | Existing knowledge of Earth's terrestrial C balance has been bolstered by the use of stable
459 | isotopes. A milestone for progress was when photosynthetic responses to environmental
460 | conditions were linked to differences between the stable C isotopic composition ($\delta^{13}\text{C}$) of
461 | atmospheric CO₂ and that of plant products (Farquhar et al., 1982). These differences, caused by
462 | C isotope fractionation during CO₂ diffusion into leaves and subsequent carboxylation (Park and
463 | Epstein, 1961; O'Leary, 1981), impart an isotopic fingerprint on ecosystem C pools and permit
464 | inference about C fluxes from $\delta^{13}\text{C}$ of ecosystem C pools at multiple spatio-temporal scales
465 | (Farquhar and Richards, 1984; Pataki et al., 2003; Dijkstra et al., 2004; Barbosa et al., 2010).

466 Recent studies remind us that respiratory C losses also leave an isotopic fingerprint on $\delta^{13}\text{C}$
467 values of plant tissues via respiration of substrates with distinct $\delta^{13}\text{C}$ (Bathellier et al., 2009;
468 Brüggemann et al. 2011; Ghashghaie and Badeck, 2014), and via C isotope fractionation during
469 decarboxylation in respiratory pathways (Werner and Gessler, 2011; Werner et al., 2011;
470 Tcherkez et al., 2012). Though not all C isotope fractionations during metabolism are well-
471 characterized, $\delta^{13}\text{C}$ of metabolic reaction substrates and products can vary predictably, caused by
472 kinetic or thermodynamic isotope effects (Rossmann et al., 1991; Gleixner and Schmidt, 1997;
473 Cleland, 2005; Tcherkez et al., 2012). Accounting for isotope effects in plant respiratory C
474 losses improves our ability to quantify the contributions of different pools to CO_2 fluxes and thus
475 our predictions of terrestrial ecosystem C budgets under changing environmental conditions.
476 Using $\delta^{13}\text{C}$ of heterotrophically respired CO_2 holds similar promise, but if and how changing
477 environmental conditions influence any fractionation factors for the fluxes associated with the
478 liberation of C from SOM is unknown.

479 Significant uncertainty about the direction and magnitude of C isotope fractionation during
480 microbial C transformations (Bowling et al., 2008; Werth and Kuzyakov, 2010) renders
481 quantifying microbial CO_2 fluxes in ~~terrestrial~~-ecosystems difficult. Difficulties arise because
482 ~~soil~~-microbes in natural systems can access a diverse array of SOM-organic substrates with
483 distinct $\delta^{13}\text{C}$ signatures (Park and Epstein, 1961; Billings, 2006), the respiration of which
484 influences $\delta^{13}\text{C}$ of respired CO_2 . Though we know the growth rate of microbial populations
485 influences C flux into and through biomass (Kayser et al., 2005), it is impossible to directly
486 quantify microbial growth *in situ*. Furthermore, absence of steady state conditions over a course
487 of ~~soil~~- CO_2 flux measurements makes the interpretation of temperature effects on the magnitude
488 and the $\delta^{13}\text{C}$ of soil-ecosystem respiration an even greater challenge (Gamnitzer et al., 2011;

489 Nickerson et al., 2013). Thus, establishing a mechanistic understanding of the links between
490 temperature, microbial respiration rates and C isotope fractionation during substrate
491 transformations at a fundamental level requires that we characterize these processes as
492 temperature changes in isolation from other factors that influence microbial C transformations.

493 To assess the influence of temperature on microbial growth and respiration rates, we grew a
494 widely distributed Gram-negative, heterotrophic ~~soil~~-bacterium (*Pseudomonas fluorescens*) in
495 continuous culture bioreactors (chemostats; ~~Ferenci, 2008; Bull, 2010~~) at seven temperatures
496 ranging from 13 °C to 26 °C (Fig. 1) at reactor dilution rates of approximately 0.14 h⁻¹, which is
497 equivalent to the relative growth rates of the microbial populations (Dawson, 1974; Smith and
498 Waltman, 1995; Goldman and Dennett, 2000; Chrzanowski and Grover, 2008; Ferenci, 2008;
499 Bull, 2010; Egli, 2015). We measured microbial respiration rates and $\delta^{13}\text{C}$ of respired CO_2 in
500 this open, ~~flow-flow~~-through system at steady-state (Craig and Gordon, 1965; Fry, 2006; see
501 Supplementary Material for a detailed elaboration of this approach). We computed the
502 temperature dependence of a widely applied metric of microbial C use efficiency (CUE), defined
503 as $\text{SGR} / (\text{SGR} + \text{SRR})$, where SGR and SRR are specific growth and specific respiration rates
504 respectively, with units of C per microbial biomass-C and time. Our simplified system
505 eliminates factors present in ~~real soils~~natural environments that preclude accurate assessment of
506 specific growth and respiration rates, and thus accurate estimates of CUE as defined above. _
507 Obtaining accurate estimates of microbial CUE is critical for projecting C fluxes into the future
508 because the particular value of CUE significantly influences CO_2 loss rates from ~~soils-~~
509 ecosystems in models of SOM decomposition (Allison et al., 2010; Wieder et al., 2013). Finally,
510 simultaneously quantifying differences in $\delta^{13}\text{C}$ of organic substrate, microbial biomass and

511 respired CO₂ along a temperature gradient is critical for partitioning synoptic CO₂ measurements
512 into ~~its~~ component fluxes.

513

514 **2 Materials and Methods**

515 **2.1 Pre-cultivation of microorganisms for chemostat inoculation**

516 We pre-cultivated *Pseudomonas fluorescens* (Carolina Biological Supply, USA) in nutrient
517 solution containing 10 mM NH₄Cl, 1.6 mM KNO₃, 2.6 mM K₂HPO₄, 1.0 mM KH₂PO₄, 0.8 mM
518 MgSO₄, 0.2 mM CaCl₂, 0.1 mM CuCl₂, 0.04 mM FeSO₄, 0.03 mM MnCl₂ and 0.02 mM ZnSO₄,
519 modified from Abraham et al. (1998). The sole C source in the nutrient medium was 10 mM
520 cellobiose (C₁₂H₂₂O₁₁; with a δ¹³C of ~~-22.4~~ -24.2‰); cellobiose is a disaccharide consisting of
521 two glucose molecules and a basic module of cellulose. Thus, the C to N to P atomic ratio of the
522 autoclaved, sterile nutrient solution was 100 to 10 to 3.3; its pH was adjusted to 6.5. The
523 bacteria grew for a few days in batch culture in a flask fitted with a vent for air exchange covered
524 by a 0.22 μm filter (Fisher Scientific, USA) to avoid contamination. Vessel contents were stirred
525 continuously in an incubator maintained at 10 °C.

526 **2.2 The laboratory mesocosm – the chemostat**

527 The chemostat system was composed of two 1.9 L vessels, a medium reservoir tank and a
528 reactor, each maintained on separate heating/stirring plates (Fig. 1) in separate incubators. The
529 reactor volume was on average 870 mL (Supplementary Table 1). The reservoir tank was
530 connected via a flexible tube to the reactor (Tygon E-LFL pump tubing, Masterflex, USA),
531 which in turn had an outlet tube (Fig. 1; both tubes had a 1.6 mm inner diameter). When the

532 chemostat was operated in “continuous culture mode” a peristaltic pump transported fresh
533 medium to the reactor and simultaneously removed medium from the reactor at the same rate.
534 Thus, reactor volume remained constant during all chemostat runs. The 0.22 μm filter in the
535 reservoir tank lid allowed for pressure compensation during withdrawal of nutrient solution in
536 the continuous flow mode. Experimental temperatures were continuously measured with a
537 thermometer (Oakton, USA) placed in the reactor medium (Fig. 1). This thermometer was
538 routinely compared against an internal laboratory standard mercury thermometer, before and at
539 the end of each experiment. The reactor temperatures were adjusted with heating/stirring plate
540 and incubator settings, and kept constant during all experimental runs.

541 The reactor lid had two ports for gas lines. The outlet port tube was connected to a $^{13}\text{CO}_2/^{12}\text{CO}_2$
542 analyzer (G2101-i, Picarro, USA) containing a pump that continuously removed air from the
543 reactor headspace at an average rate of 0.025 L min^{-1} . A water trap (magnesium perchlorate,
544 Costech, USA) was installed between outlet port of the reactor and the gas analyzer. The CO_2
545 analyzer recorded the concentration and the $\delta^{13}\text{C}$ of the reactor headspace CO_2 ~~about of nee-~~
546 ~~every two seconds~~ 0.5 Hz. The reactor’s inlet tube was connected to a mass flow controller (MC-
547 50SCCM, Alicat Scientific, USA), which in turn, was connected to a gas cylinder containing
548 CO_2 -free air (Fig. 1). The mass flow controller was programmed to maintain the reactor
549 headspace at constant atmospheric pressure; thus, the 0.025 L min^{-1} headspace air removed by
550 the $^{13}\text{CO}_2/^{12}\text{CO}_2$ analyzer pump was instantaneously replaced with CO_2 -free air flowing from the
551 gas cylinder into the reactor medium. Considering-Assuming (1) that 1 mol of O_2 is consumed
552 per 1 mol of CO_2 produced in aerobic respiration, (2) a typical reactor headspace CO_2
553 concentration of around 2000 ppm at steady state (see Fig. 2A and below), and (3) an O_2
554 concentration of 21% in the air supply to the reactor, the air supply permitted continuous aerobic

555 metabolism. Routine tests with CO₂-free air in sterile chemostats were performed to ensure there
556 were no leaks in the system.

557 **2.3. The chemostat run – standardized protocol and description of events**

558 We conducted seven independent chemostat runs, at temperatures of 13, 14.5, 16, 18, 21, 23.5
559 and 26.5 °C, in random temporal order. For each of the chemostat runs, we inoculated the
560 reactor with a 10 mL aliquot of the *P. fluorescens* pre-culture and activated the flow of CO₂-free
561 air through the reactor; this was considered time 0. At the initial stage of a chemostat run, the
562 bacteria grew in batch culture, that is, there was no flow of fresh nutrient medium from the
563 reservoir tank to the reactor, and no removal of medium from the reactor (Fig. 1).

564 **2.3.1 Respiration measurements at chemical and isotopic equilibrium in the continuous** 565 **flow chemostat at steady-state**

566 At the initial pH of 6.5, inorganic C in the fresh reactor medium was mainly in the form of
567 H₂CO₃ (aq) and HCO₃⁻ (Stumm and Morgan, 1981). By continuously bubbling CO₂-free air into
568 the reactor, we expelled these initial inorganic C pools from the reactor medium. This was
569 evident by concentrations of reactor headspace CO₂ of virtually zero in the early stages of batch
570 culture after each run's inoculation (Fig. 2A). During the phase of rising reactor headspace CO₂
571 via respiratory activity of the exponentially growing population (Fig. 2A), inorganic C in the
572 reactor medium accrued with the increasing addition of CO₂ from microbial respiration. That is,
573 at any point in time during the phase of increasing reactor headspace CO₂ concentration, the
574 nutrient medium acted as a sink for respired CO₂- (see also Supplementary Material).

575 Once the respiratory activity of the growing microbial population pushed the reactor headspace
576 CO₂ concentration above 500 ppm, we transferred the chemostat into the “continuous culture,
577 open system” mode (Figs. 1, 2; Ferenci, 2008; Bull, 2010). The peristaltic pump henceforth
578 transported fresh nutrient medium from the reservoir tank to the reactor at a constant rate of, on
579 average, 118 mL h⁻¹ (Supplementary Table 1), and simultaneously removed medium from the
580 reactor at the same rate so that the reactor volume remained constant. Initial chemostat
581 experiments indicated that when headspace CO₂ concentrations reached 500 ppm, the bacterial
582 population was sufficiently dense to maintain itself without being washed out via ~~medium-~~
583 flowdilution. Depending on the reactor temperature, the onset of the continuous culture mode
584 occurred between 40 h (at 26.5 °C) and 72 h (at 13 °C) after inoculation.

585 After the switch from batch to continuous culture, the rate of increase in reactor headspace CO₂
586 concentration gradually slowed because cells were continuously diluted into the waste stream
587 (Fig. 1), and approached a phase where the CO₂ concentration became stable (Fig. 2A). At this
588 point, bacteria grown in continuous culture had reached the phase of steady-state growth and
589 physiology (see Ferenci, 2008; Bull, 2010). A key feature of the continuous culture chemostat
590 relevant to our study is that at this steady-state, the constant dilution rate of the reactor (the
591 medium flow rate divided by the reactor volume) is equivalent to the specific growth rate of the
592 microbial culture (Bull, 2010). That is, washout of cells with the nutrient medium flow is
593 balanced by cell division so that the size of the population in the reactor can be expected to be
594 reasonably constant in the time frames employed here (see discussion in Ferenci, 2008; Bull,
595 2010).

596 Critically, when reactor headspace CO₂ concentrations approached the steady-state, inorganic C
597 pools came to their respective equilibria as well (Stumm and Morgan, 1981). At this point, pools
598 of H₂CO₃ (aq) and HCO₃⁻ were no longer a *net* sink for respired CO₂. As reactor headspace CO₂
599 concentrations reached steady state, the system supported constant microbial CO₂ production
600 reflective of steady-state growth under constant environmental conditions, and reflected
601 chemical equilibrium (i.e., constant size) of the dissolved inorganic C pools. Thus, the rate of
602 CO₂ addition to the reactor headspace volume at steady-state accurately represented the CO₂
603 released during microbial respiration [\(see also Supplementary Material\)](#).

604 We calculated the molar CO₂ production rate of the microbial population as the product of the
605 average molar CO₂ concentration measured by the ¹³CO₂/¹²CO₂ analyzer for 5 hours at steady
606 state (Fig. 2A) multiplied by the molar air flow rate through the reactor, which was calculated as

607
$$\text{Air flow (mol min}^{-1}\text{)} = 0.96 \text{ atm} * 0.025 \text{ L min}^{-1} / (0.082 \text{ atm L mol}^{-1} \text{ K}^{-1} * 296 \text{ K}),$$

608 with 0.96 atm and 296 K being the barometric pressure and the temperature in the lab where the
609 experiments were performed, 0.025 L min⁻¹ the average volumetric headspace flow rate through
610 the reactor and 0.082 atm L mol⁻¹ K⁻¹ the gas constant.

611 The δ¹³C of the reactor headspace CO₂ during the [early-earliest](#) batch culture phase was generally
612 very negative; ~~due to the inability of~~ the ¹³CO₂/¹²CO₂ analyzer ~~cannot to~~ accurately measure ¹³C
613 and ¹²C in very low CO₂ concentrations (Fig. 2B). The δ¹³C of reactor headspace CO₂ became
614 less negative as the CO₂ concentration increased (Fig. 2B). During the “climbing” phase of the
615 reactor headspace CO₂, the δ¹³C of the CO₂ pool was influenced by isotopic fractionation among
616 gaseous CO₂, H₂CO₃ (aq) and HCO₃⁻ (Vogel et al., 1970; Mook et al., 1974; Stumm and Morgan,

617 1981; Szaran, 1997), because the dissolved inorganic C pools functioned as a net sink for
618 respired CO₂. At steady-state, with constant headspace CO₂ concentrations and constant size of
619 the dissolved inorganic C pools (see above), isotopic equilibrium was achieved, evidenced by
620 constant δ¹³C readings of reactor headspace CO₂ (Fig. 2B). As such, in this open system at
621 steady-state, the δ¹³C of the CO₂ leaving the reactor (the CO₂ measured by the analyzer) is
622 identical to the δ¹³C of microbial respiration (Craig and Gordon, 1965; Fry, 2006). Importantly,
623 this principle is valid irrespective of temperature, microbial growth rate or microbial biomass in
624 the reactor. (See Supplementary Material for an elaboration of the principle of chemical and
625 isotopic equilibrium.)

626 We used the average δ¹³C measurement of reactor headspace CO₂ over the same five hours in the
627 stable phase employed for calculations of microbial respiration rates (see above) as the isotopic
628 signature of CO₂ respired by the microbial culture at each temperature. Any measurements of
629 headspace CO₂ and δ¹³C during the climbing phase before steady-state (Fig. 2) were not used in
630 these calculations.

631 For the example chemostat at 23.5 °C, the half-life of the reactor ($t_{1/2}$), i.e., the time it took until
632 50% of the reactor medium was exchanged with fresh tank medium, was 5.2 h (with $t_{1/2} = \ln(2) /$
633 (medium flow rate / reactor volume); Supplementary Table 1). In a homogeneous, well-mixed
634 system such as that employed here, 95% of the pool (i.e., the reactor) is exchanged with new
635 medium within approximately five times the half-life. Thus, during the respiration
636 measurements between time 70 h and 74 h (in the example time course in Fig. 2), any “leftovers”
637 from the batch culture mode were insignificant, and the microbial culture could be considered
638 homogeneous. This principle was applicable to all chemostat runs we performed.

639 After the 5-hour respiration measurements were completed, we disconnected the gas lines from
640 the reactor, connected the mass flow controller directly to the $^{13}\text{CO}_2/^{12}\text{CO}_2$ analyzer, and replaced
641 the CO_2 -free air cylinder with a reference gas cylinder containing 1015 ppm CO_2 at a $\delta^{13}\text{C}$ of -
642 48.9‰ (Matheson, USA). This laboratory standard gas was previously calibrated against
643 secondary CO_2 standards (Oztech, USA) and served for any necessary corrections of the $\delta^{13}\text{C}$ of
644 the reactor headspace CO_2 measurements. Across the seven standard measurement procedures
645 after each individual chemostat run, the $\delta^{13}\text{C}$ measured for the laboratory standard gas showed
646 only slight variation (1 SD = 0.16‰). CO_2 concentration measurements needed no correction;
647 measurements of lab-internal gases with previously determined CO_2 concentrations between
648 chemostat runs showed very stable and accurate analyzer performance.

649 **2.3.2 Measurements of extracellular enzyme activities at steady-state**

650 Using principles detailed by Lehmeier et al. (2013) and Min et al. (2014), we tested reactor
651 medium for activity of the extracellular enzymes β -glucosidase and β -N-acetyl glucosaminidase
652 across all chemostat temperatures; we never detected extracellular activity of either enzyme. The
653 lack of extracellular β -glucosidase activity indicates that the sole C source, ~~of~~ cellobiose, was
654 directly taken up by microbes and cleaved intracellularly into glucose monomers for further
655 metabolism. The lack of extracellular β -N-acetyl glucosaminidase activity suggests that the
656 inorganic N provided in nutrient medium was the sole source of N taken up by *P. fluorescens*.

657 These findings do not rule out the possibility that *P. fluorescens* may have taken up (i.e. recycled)
658 any exuded C-based metabolic compounds, although such a scenario in continuous culture
659 conditions may not appear to be energetically favorable. ~~These~~ Thus, ~~inferences strengthened~~

660 | ~~our~~the assumption that the sole resources used by *P. fluorescens* were the cellobiose and the
661 | nutrient medium appears reasonable.

662 | 2.3.3 Harvest of microbial biomass at steady-state

663 | Immediately after completing the 5 hour respiration measurements, we filtered approximately
664 | 300 mL of reactor medium for steady-state microbial biomass using 0.2 μm filters made of
665 | polyethersulfone (Pall, USA) and a vacuum pump. The filters had previously been oven-dried
666 | for 48 h at 75 °C and their dry weight determined. We then oven-dried the filters post filtration
667 | for 48 h at 75 °C, removed some of the dry biomass and weighed 1.2 mg of the ~~dry~~
668 | biomass material into tin cups for subsequent combustion in an elemental analyzer (1110 CHN
669 | Combustion Analyzer, Carlo Erba Strumentazione, Italy) coupled to a ThermoFinnigan DeltaPlus
670 | mass spectrometer (Finnigan MAT, Germany) at the Keck Paleoenvironmental and
671 | Environmental Stable Isotope Laboratory (The University of Kansas, USA). We thus determined
672 | biomass C (and N) elemental content, as well as the $\delta^{13}\text{C}$ of the biomass. In this analysis, the
673 | samples were compared against a laboratory standard CO_2 previously calibrated against the same
674 | secondary CO_2 standards as used in calibration of the CO_2 standard used for respiration
675 | measurements (see above). The $\delta^{13}\text{C}$ of the substrate cellobiose was measured likewise. Dry
676 | weight of the sampled reactor aliquots and the C content obtained from elemental analysis served
677 | to calculate total microbial C content in the steady-state reactor and to calculate specific
678 | respiration rates.

679 | At all temperatures studied, biomass C and N contents were virtually the same, on average 27%
680 | and 8% of microbial dry mass, respectively (Supplementary Table 1). From mass balance
681 | calculations, we determined that only a small percentage of the C and N supplied via fresh

682 | medium from the tank was taken up to fuel microbial growth (~~1.8~~2.8 % and 4.3 % on average
683 | for C and N, respectively). This suggests that the observed temperature effects on specific
684 | respiration rates and CUE were not confounded by any differences in C and N limitations at the
685 | different temperatures (Goldmann and Dennet, 2000; Cotner et al., 2006; Chrzanowski and
686 | Grover, 2008).

687

688 **3 Results and Discussion**

689 | For *P. fluorescens* grown in continuous culture, CUE, defined as $SGR / (SGR + SRR)$, declined
690 | with increasing temperature, from 77% at 13 °C to 56% at 26.5 °C (Fig. 3A). Because specific
691 | growth rates were similar across the experimental temperatures ($137 \text{ mg g}^{-1} \text{ h}^{-1}$, ± 8 (1 SD); or
692 | 13.7 % h⁻¹ in relative terms; Fig. 3B), the decline in CUE was due to the 2.5 fold increase of SRR
693 | with temperature, which rose from $45 \text{ mg g}^{-1} \text{ h}^{-1}$ at 13 °C to $113 \text{ mg g}^{-1} \text{ h}^{-1}$ at 26.5 °C (Fig. 3B).
694 | The decline in CUE is also evident in the ~~the~~ more than 50% reduction in steady-state dry
695 | microbial biomass with increasing temperature (Fig. 3A) ~~was driven by the 2.5 fold increase in~~
696 | ~~SRR, from $45 \text{ mg g}^{-1} \text{ h}^{-1}$ at 13 °C to $113 \text{ mg g}^{-1} \text{ h}^{-1}$ at 26.5 °C (Fig. 3B).~~ For example, though
697 | SGR was approximately the same and thus the fraction of biomass replaced per time similar
698 | across all experimental temperatures (0.147 and 0.141 at 13 °C and 26.5 °C, respectively; Fig.
699 | 3B) microbes at 13 °C generated $20.9 \text{ mg biomass h}^{-1}$ while those at 26.5 °C generated less than
700 | half the amount ($9.5 \text{ mg biomass h}^{-1}$). –

701 | Because we did not quantify possible C losses from the population at steady-state such as
702 | secretion of organic acids or other compounds (El-Mansi and Holms, 1989; Nanchen et al.,

703 | 2006), gross rates of steady-state cellobiose C uptake may have been slightly higher than what
704 | was calculated from the sum of SGR and SRR. However, the direct observation of *P.*
705 | *fluorescens*' CUE is consistent with the negative effect of increasing temperature on microbial
706 | CUE widely reported in literature for soils and aquatic ecosystems (del Giorgio and Cole, 1998;
707 | Gillooly et al., 2001; Rivkin and Legendre, 2001; Apple et al., 2006; Manzoni et al., 2012; Frey
708 | et al., 2013; Tucker et al., 2013).

709 | Across the chemostat runs, we observed strong C isotope fractionations, which created
710 | pronounced differences in $\delta^{13}\text{C}$ between microbial biomass and the sole C substrate, cellobiose,
711 | and between microbial biomass and respired CO_2 (Fig. 4). Microbial biomass exhibited 5.5 to
712 | 10.5‰ more negative $\delta^{13}\text{C}$ values than the cellobiose and respired CO_2 was even more ^{13}C
713 | depleted, at least 4.4‰ more negative than the biomass (Fig. 4A). Because each chemostat was
714 | at steady-state, isotopic mass balance dictates that ^{13}C “missing” from cellobiose had to
715 | accumulate in another pool in the reactor. The only pool that could have been enriched with the
716 | “missing” ^{13}C was reactor DOC, which we analyzed for $\delta^{13}\text{C}$ in four out of the seven chemostat
717 | runs (Fig. 4A). Reactor DOC consisted of a large pool of cellobiose (because the rate of C
718 | consumption by the chemostat cultures was, on average, only ~~+2.8%~~ of the rate of C supply) and
719 | presumably a pool of additional organic compounds (e.g. such as acetate). Such compounds
720 | appear to be typically secreted from microbial cells at low rates in aerobic chemostats operated at
721 | dilution rates similar to those of our runs (El-Mansi and Holms, 1989; Nanchen et al., 2006), and
722 | have been shown to be enriched in ^{13}C compared to cellular biomass (Blair et al., 1985).
723 | However, because such a small fraction of the available cellobiose was taken up by *P.*
724 | *fluorescens*, the fraction of total DOC comprised of secreted organic compounds was small. As a

725 result, ^{13}C enrichment of any microbial exudates was insufficient to enrich bulk DOC to an
726 extent detectable by the isotope-ratio mass spectrometer (Fig. 4A).

727 The majority of the fractionation between *P. fluorescens* biomass and the substrate was most
728 likely due to discrimination against ^{13}C during cellobiose uptake. If we assume that *P.*
729 *fluorescens* secreted organic compounds at a rate of 10% of the sum of SGR and SRR (El-Mansi
730 and Holms, 1989) and that the bacteria did not discriminate against ^{13}C -containing cellobiose
731 during uptake (and thus assimilated cellobiose possessed a $\delta^{13}\text{C}$ of -24.2‰), isotopic mass
732 balance dictates that the $\delta^{13}\text{C}$ of the C secretion flux (Fig. 5) would have to be +70‰, at
733 minimum, across all temperatures. To our knowledge, such high metabolic discrimination
734 against ^{13}C would be very unusual for biological systems (O’Leary, 1981). An alternative and
735 more likely scenario is therefore that *P. fluorescens* took up less ^{13}C -containing cellobiose than
736 was supplied as substrate, and that discrimination during uptake contributed substantially to *P.*
737 *fluorescens* biomass and respired CO_2 being more ^{13}C depleted than the cellobiose supplied.

738 This conclusion holds for all temperatures studied. If we assume for the example 23.5 °C
739 chemostat run at steady-state (Figs. 2, 5) that *P. fluorescens* secreted organic compounds at a rate
740 of 10% of the sum of SGR and SRR, and we further assume that the $\delta^{13}\text{C}$ of secreted compounds
741 was 11.7‰ less negative than that of the biomass (Blair et al., 1985), that is -18.1‰, then did not
742 discriminate against ^{13}C during secretion but only during uptake, the $\delta^{13}\text{C}$ of the cellobiose taken
743 up would have been, ~~on average, -34.2-31.1‰~~, which is only a 10‰ difference from the
744 cellobiose provided, and therefore probably a more likely scenario (Fig. 5).

745 Substantial ^{13}C depletion of respired CO_2 relative to microbial biomass has not, to our
746 knowledge, been reported in other studies. With the microbial C consumption rate amounting to

747 only 1.82.8% of the rate of C supply, C availability was high compared to what ~~soil~~-microbes in
748 their natural environments typically experience (Tempest and Neijssel, 1978; Cole et al., 1988;
749 Hobbie and Hobbie, 2013), potentially promoting enzymatic discrimination. The cellobiose $\delta^{13}\text{C}$
750 of -24.2 ‰ implies a $^{13}\text{C}/^{12}\text{C}$ ratio of ~1/91. Considering the molecular formula of cellobiose
751 $\text{C}_{12}\text{H}_{22}\text{O}_{11}$, this means that not more than about one out of eight cellobiose molecules in the
752 supplied substrate included a ^{13}C atom. Faster diffusion of the isotopically lighter cellobiose
753 molecules may have contributed to a lower probability of ^{13}C -containing cellobiose approaching
754 bacterial membrane uptake sites, and hence, to the differences between $\delta^{13}\text{C}$ of substrate and
755 biomass (Fig. 4A). However, isotope fractionation during diffusion – a physical process
756 dependent on compound mass – would likely exhibit a continuous temperature response. Thus,
757 it seems unlikely that fractionation during diffusion was the primary driver of the pronounced,
758 discontinuous changes in the difference between substrate and biomass $\delta^{13}\text{C}$, which ranged from
759 5.5 to 10.5‰ (Fig. 4A). Rather, this variation, with one apparently linear part between 13 °C and
760 16 °C and another between 18 °C and 23.5 °C, may be explained parsimoniously by a
761 significant, discontinuous reorganization of enzyme-mediated C fluxes into and out of bacterial
762 cells (see Nanchen et al., 2006), induced by differences in temperature at which *P. fluorescens*
763 was growing and the related differences in substrate uptake rates.

764 Evidence from work on C isotope distribution within carbohydrate molecules (e.g., Rossmann et
765 al., 1991; Gleixner and Schmidt, 1997) suggests non-random distribution of ^{13}C in biological
766 molecules (such as cellobiose). Based on such phenomena, it is probable that the ^{13}C atom in
767 ^{13}C -containing cellobiose was consistently at the same position within the molecule, which rules
768 out the possibility that any changes in the intramolecular ^{13}C distribution of cellobiose were
769 responsible for the observed $\delta^{13}\text{C}$ patterns in biomass and respired CO_2 (Fig. 4). Hence, ~~T~~the

770 discontinuous pattern of $\delta^{13}\text{C}$ of respired CO_2 with temperature, similar to the pattern for $\delta^{13}\text{C}$ of
771 the biomass (Fig. 4A), presumably reflects the downstream consequence of an upstream change
772 in $\delta^{13}\text{C}$ of the metabolic substrate taken up and ultimately respired. However, the more negative
773 $\delta^{13}\text{C}$ of respired CO_2 compared to that of biomass is, to our knowledge, the most direct evidence
774 to date for ^{13}C discrimination during respiration of a heterotrophic ~~soil~~-microbe. The observation
775 of a substantial respiratory ^{13}C discrimination corroborates inferences drawn in earlier studies
776 (Šantrůčková et al., 2000; Fernandez and Cadisch, 2003) and is also consistent with plant studies
777 reporting C isotope discrimination during dark respiration in roots (Klumpp et al., 2005;
778 Bathellier et al., 2009; Ghashghaie and Badeck, 2014). Our observations of respiratory
779 discrimination against ^{13}C highlight the similarity of heterotrophic, aerobic respiratory pathways,
780 and isotope effects within them, across life's domains.

781 In contrast to the discontinuous relationship between biomass $\delta^{13}\text{C}$ and temperature, we observed
782 a comparably continuous and linear increase in respiratory discrimination against ^{13}C with
783 temperature (Fig. 4B). This increase generated a marginally positive significant ($P=0.08$)
784 correlation with SRR (Fig. 6), and hence a marginally significant ($P=0.07$) correlation with CUE.
785 A physiological interpretation of this finding is not straightforward, as multiple, possibly
786 simultaneous enzymatic fractionations may have contributed to the observed $\delta^{13}\text{C}$ of respired
787 CO_2 (Dijkstra et al., 2011; Tcherkez et al., 2012). It could simply result from a proportionally
788 increasing flux through respiratory pathways, with associated stronger expression of ^{13}C
789 discrimination by the enzymes involved (Tcherkez et al., 2012), or could result from increasing
790 temperatures altering the relative fluxes through respiratory pathways (Chung et al., 1976;
791 Wittmann et al., 2007; Dijkstra et al., 2011) such that the overall observed respiratory ^{13}C
792 discrimination increased with temperature. This may be possible given that respiratory pathways

793 can exhibit distinct fractionation factors (Bathellier et al., 2009 and references therein) and
794 prompt different, specific C atoms to undergo decarboxylation from the two glucose units of the
795 substrate cellobiose, which contain non-randomly distributed ^{13}C atoms (Rossmann et al., 1991;
796 Gleixner and Schmidt, 1997). If relative fluxes through different respiratory pathways changed
797 with temperature, the continuous nature of the relationship between temperature and respiratory
798 ^{13}C discrimination suggests a smooth transition compared to the abrupt and discontinuous shifts
799 in apparent uptake and/or secretion discrimination described above. Future metabolic flux
800 analyses linked to isotopic approaches sensitive enough to quantify C isotopes in microbial
801 exudation will be well-suited to explore how C allocation to distinct, aerobic respiratory
802 pathways may vary with temperature and result in varying $\delta^{13}\text{C}$ of respired CO_2 .

803

804 **4. Conclusions**

805 Our observations clearly show a decline in microbial CUE with increasing temperature when C
806 substrate is plentiful and demonstrate the mechanism driving it – an increase in SRR. The
807 relationship between CUE and temperature underscores the importance of incorporating variable,
808 temperature dependent SRR, which influences CUE and growth efficiency, in ecosystem process
809 models. The temperature-driven changes in SRR and respiratory discrimination against ^{13}C were
810 not independent of each other, suggesting that increasing SRR, to some degree, drives enhanced
811 C isotopic discrimination. We demonstrate that C isotope discrimination associated with
812 | microbial decomposition of SOM can impart large and variable isotopic signatures on C pools
813 typically characterized and interpreted in biogeochemical studies at any scale. To date, efforts to
814 partition flux components of net ecosystem exchange have assumed little to no fractionation

815 between respired substrates and the resultant CO₂. Our results suggest that this assumption must
816 be reevaluated, and represent a first step towards an isotopically explicit, mechanistic framework
817 for microbial C isotope fluxes in Earth system models.

818

819 **Author contribution:**

820 C.A.L. and K.M. performed the experiments; all authors contributed to all other parts and stages
821 of the manuscript.

822

823 **Data availability:** The data presented in this study are available for collaborative use by anyone
824 interested; contact the corresponding author for access to the data.

825

826 **Acknowledgements:**

827 We thank Dr. Susan Ziegler, Dr. Jarad Mellard and Chao Song for helpful discussions during the
828 design of the experiments, Greg Caine for expert stable isotope analysis and the provision of
829 isotopic standards, Drs. Susan Ziegler, Mike Burrell, Karl Auerswald, Hanns-Ludwig Schmidt
830 and John Kelly for comments on the manuscript, Gil Ortiz with assistance generating Figure 5,

831 anonymous referees for their time and constructive comments, and the National Science

832 Foundation of the USA for funding (grant no. DEB-0950095 and EAR-1331846).

833

834 **References:**

835 Abraham, W.-R., Hesse, C., and Pelz, O.: Ratios of carbon isotopes in microbial lipids as an
836 indicator of substrate usage, *App. Environ. Microbiol.*, 64, 4202-4209, 1998.

837 Allison, S.D., Wallenstein, M.D., and Bradford, M.A.: Soil-carbon response to warming
838 dependent on microbial physiology, *Nat. Geosci.*, 3, 336-340, 2010.

839 Apple, J.K., del Giorgio, P.A., and Kemp, W.M.: Temperature regulation of bacterial production,
840 respiration, and growth efficiency in a temperate salt-marsh estuary, *Aquat. Microb. Ecol.*, 43,
841 243-254, 2006.

842 Barbosa, Î.C.R., Köhler, I.H., Auerswald, K., Lüps, P., and Schnyder, H.: Last-century changes of
843 alpine grassland water-use efficiency: a reconstruction through carbon isotope analysis of a time-
844 series of *Capra ibex* horns, *Glob. Change Biol.*, 16, 1171-1180, 2010.

845 Bathellier, C., Tcherkez, G., Bligny, R., Gout, E., Cornic, G., and Ghashghaie, J.: Metabolic
846 origin of the $\delta^{13}\text{C}$ of respired CO_2 in roots of *Phaseolus vulgaris*, *New Phytol.*, 181, 387-399,
847 2009.

848 Billings, S.: Soil organic matter dynamics and land use change at a grassland/forest ecotone, *Soil*
849 *Biol. Biochem.*, 38, 2934-2943, 2006.

850 Blair, N., Leu, A., Muñoz, E., Olsen, J., Kwong, E., and Des Marais, D.: Carbon isotopic
851 fractionation in heterotrophic microbial metabolism, *Appl. Environ. Microbiol.*, 50, 996-1001,
852 1985.

853 Bowling, D.R., Pataki, D.E., and Randerson, J.T.: Carbon isotopes in terrestrial ecosystem pools
854 and CO₂ fluxes, *New Phytol.*, 178, 24-40, 2008.

855 Brüggemann, N., Gessler, A., Kayler, Z., Keel, S.G., Badeck, F., Barthel, M., Boeckx, P.,
856 Buchmann, N., Brugnoli, E., Esperschütz, J., Gavrichkova, O., Ghashghaie, J., Gomez-
857 Casanovas, N., Keitel, C., Knohl, A., Kuptz, D., Palacio, S., Salmon, Y., Uchida, Y., and Bahn,
858 M.: Carbon allocation and carbon isotope fluxes in the plant-soil-atmosphere continuum: a
859 review, *Biogeosci.*, 8, 3457-3489, 2011.

860 Bull, A.T.: The renaissance of continuous culture in the post-genomics age, *J. Ind. Microbiol.*
861 *Biotechnol.*, 37, 993-1021, 2010.

862 Chrzanowski, T.H., and Grover J.P.: Element content of *Pseudomonas fluorescens* varies with
863 growth rate and temperature: A replicated chemostat study addressing ecological stoichiometry,
864 *Limnol. Oceanogr.*, 53, 1242-1251, 2008.

865 Chung, B.H., Cannon, R.Y., and Smith, R.C.: Influence of growth temperature on glucose
866 metabolism of a psychrotrophic strain of *Bacillus cereus*, *Appl. Environ. Microb.*, 31, 39-45,
867 1976.

868 Cleland, W.W.: The use of isotope effects to determine enzyme mechanisms, *Arch. Biochem.*
869 *Biophys.*, 433, 2-12, 2005.

870 Cole, J.J., Findlay, S., and Pace M.L.: Bacterial production in fresh and saltwater ecosystems: a
871 cross-system overview, *Mar. Ecol. Prog. Ser.*, 43, 1-10, 1988.

872 | [Cotner, J.B., and Biddanda, B.A.: Small players, large role: microbial influence on](#)
873 | [biogeochemical processes in pelagic aquatic ecosystems, *Ecosystems*, 5, 105-121, 2002.](#)

874 | Cotner, J.B., Makino, W., and Biddanda, B.A.: Temperature affects stoichiometry and
875 | biochemical composition of *Escherichia coli*, *Microb. Ecol.*, 52, 26-33, 2006.

876 | Craig, H., and Gordon, L.I.: Deuterium and 18 oxygen variations in the ocean and the marine
877 | atmosphere. In: Proceedings of a conference on stable isotopes in oceanographic studies and
878 | paleotemperatures. Ed: E. Tongioli, Lischi & Figli, Pisa, 1965.

879 | [Dawson, P.S.S.: Microbial growth, Dowden, Hutchinson & Ross, Stroudsburg, 1974.](#)

880 | del Giorgio, P.A., and Cole, J.J.: Bacterial growth efficiency in natural aquatic systems, *Annu.*
881 | *Rev. Ecol. Syst.*, 29, 503-541, 1998.

882 | Dijkstra, F.A., Hobbie, S.E., Knops, J.M.H., and Reich, P.B.: Nitrogen stabilization and plant
883 | species interact to influence soil carbon stabilization, *Ecol. Lett.*, 7, 1192-1198, 2004.

884 | Dijkstra, P., Thomas, S.C., Heinrich, P.L., Koch, G.W., Schwartz, E., and Hungate, B.A.: Effect
885 | of temperature on metabolic activity of intact microbial communities: evidence for altered
886 | metabolic pathway activity but not for increased maintenance respiration and reduced carbon use
887 | efficiency, *Soil Biol. Biochem.*, 43, 2023-2031, 2011.

888 | [Egli, T.: Microbial growth and physiology: a call for better craftsmanship, *Front. Microbiol.*,](#)
889 | [6:287, DOI: 10.3389/fmicb.2015.00287, 2015.](#)

890 | El-Mansi, E.M.T., Holms, W.H.: Control of carbon flux to acetate excretion during growth of
891 | *Escherichia coli* in batch and continuous cultures, *J. General Microbiol.*, 135, 2875-883, 1989.

892 Farquhar, G.D., O'Leary, M.H., and Berry, J.A.: On the relationship between carbon isotope
893 discrimination and the intercellular carbon dioxide concentration in leaves, *Aust. J. Plant*
894 *Physiol.*, 9, 121-137, 1982.

895 Farquhar, G.D., and Richards, R.A.: Isotopic composition of plant carbon correlates with water-
896 use efficiency of wheat genotypes, *Aust. J. Plant Physiol.*, 11, 539-552, 1984.

897 Ferenci, T.: Bacterial physiology, regulation and mutational adaptation in a chemostat
898 environment, *Adv. Microb. Physiol.*, 53, 169-230, 2008.

899 Fernandez, I., and Cadisch, G.: Discrimination against ^{13}C during degradation of simple and
900 complex substrates by two white rot fungi, *Rapid Commun. Mass Spectrom.*, 17, 2614-2620,
901 2003.

902 Frey, S., Lee, J., Melillo, J.M., and Six, J.: The temperature response of soil microbial efficiency
903 and its feedback to climate, *Nat. Clim. Change*, 3, 395-398, 2013.

904 Fry, B.: *Stable Isotope Ecology*, Springer, New York, 2006.

905 Gamnitzer, U., Moyes, A.B., Bowling, D.R., and Schnyder, H.: Measuring and modeling the
906 isotopic composition of soil respiration: insights from a grassland tracer experiment, *Biogeosci.*,
907 8, 333-1350, 2011.

908 Ghashghaie, J., and Badeck, F.W.: Opposite carbon isotope discrimination during dark
909 respiration in leaves versus roots – a review, *New Phytol.*, 201, 751-769.

910 Gillooly, J.F., Brown, J.H., West, G.B., Savage, V.M., and Charnov, E.L.: Effects of size and
911 temperature on metabolic rate, *Science*, 293, 2248-2251, 2001.

912 Gleixner, G., and Schmidt H.-L.: Carbon isotope effects on the Fructose-1,6-bisphosphate
913 aldolase reaction, origin for non-statistical ¹³C distributions in carbohydrates, J. Biol. Chem.,
914 272, 5382-5387, 1997.

915 Goldman, J.C., and Denner, M.R.: Growth of marine bacteria in batch and continuous culture
916 under carbon and nitrogen limitation, Limnol. Oceanogr., 45, 789-800, 2000.

917 Hall, E.K., Neuhauser, C., and Cotner, J.B.: Toward a mechanistic understanding of how natural
918 bacterial communities respond to changes in temperature in aquatic ecosystems. ISME J, 2, 471-
919 481, 2008.

920 Hanson, P.J., Edwards, N.T., Garten, C.T., and Andrews, J.A.: Separating root and microbial
921 contributions to soil respiration: a review of methods and observations. Biogeochem., 48, 115-
922 146, 2000.

923 Hedges, J.I., Eglinton, G., Hatcher, P.G., Kirchman, D.L., Arnosti, C., Derenne, S., Evershed,
924 R.P., Kögel-Knabner, I., de Leeuw, J.W., Littke, R., Michaelis, W., Rullkötter, J.: The
925 molecularly-uncharacterized component of nonliving organic matter in natural environments,
926 Org. Geochem., 31, 945-958, 2000.

927 Hobbie, J.E., and Hobbie, E.A.: Microbes in nature are limited by carbon and energy: the
928 starving-survival lifestyle in soil and consequences for estimating microbial rates, Front.
929 Microbiol., 4, doi: 10.3389/fmicb.2013.00324, 2013.

930 Kayser, A., Weber, J., Hecht, V., and Rinas, U.: Metabolic flux analysis of Escherichia coli in
931 glucose-limited continuous culture. I. Growth-rate dependent metabolic efficiency at steady-
932 state, Microbiol., 151, 693-706, 2005.

933 Kirschbaum, M.U.F.: The temperature-dependence of organic-matter decomposition – still a
934 topic of debate, *Soil Biol. Biochem.*, 38, 2510-2518, 2006.

935 Klumpp, K., Schäufele, R., Lötscher, M., Lattanzi, F.A., Feneis, W., and Schnyder, H.: C-isotope
936 composition of CO₂ respired by shoots and roots: fractionation during dark respiration?, *Plant*
937 *Cell Environ.*, 28, 241-250, 2005.

938 [Kucera, C.L., and Kirkham, D.R.: Soil respiration studies in tallgrass prairie in Missouri,](#)
939 [Ecology, 52, 912-915, 1971.](#)

940 Lehmeier, C.A., Min, K., Niehues, N.D., Ballantyne, F.IV., and Billings, S.A.: Temperature-
941 mediated changes of exoenzyme-substrate reaction rates and their consequences for the carbon to
942 nitrogen flow ration of liberated resources, *Soil Biol. Biochem.*, 57, 374-382, 2013.

943 Manzoni, S., Taylor, P., Richter, A., Porporato, A., and Ågren, G.I.: Environmental and
944 stoichiometric controls on microbial carbon-use efficiency in soils, *New Phytol.*, 196, 79-91,
945 2012.

946 Min, K., Lehmeier, C.A., Ballantyne, F., Tatarko, A., and Billings, S.A.: Differential effects of
947 pH on temperature sensitivity of organic carbon and nitrogen decay, *Soil Biol. Biochem.*, 76,
948 193-200, 2014.

949 Mook, W.G., Bommerson, J.C., and Staverman, W.H.: Carbon isotope fractionation between
950 dissolved bicarbonate and gaseous carbon dioxide, *Earth Planet. Sci. Lett.*, 22, 169-176, 1974.

951 Nanchen, A., Schicker A., and Sauer U.: Nonlinear dependency of intracellular fluxes on growth
952 rate in miniaturized continuous cultures of *Escherichia coli*, *Appl. Environ. Microb.*, 73, 1164-
953 1172, 2006.

954 Nickerson, N., Egan, J., and Risk, D.: Iso-FD: A novel method for measuring the isotopic
955 signature of surface flux, *Soil Biol. Biochem.*, 62, 99-106, 2013.

956 O'Leary, M.H.: Carbon isotope fractionation in plants, *Phytochem.*, 20, 553-567, 1981.

957 Park, R., and Epstein, S.: Metabolic fractionation of C^{13} & C^{12} in plants, *Plant Physiol.*, 36, 133-
958 138, 1961.

959 Pataki, D.E., Ehleringer, J.E., Flanagan, L.B., Yakir, D., Bowling, D.R., Still, C.J., Buchmann,
960 N., Kaplan, J.O., and Berry, J.A.: The application and interpretation of Keeling plots in terrestrial
961 carbon cycle research, *Glob. Biogeochem. Cycles*, 17, doi:10.1029/2001GB001850, 2003.

962 [Pomeroy, L.R., and Wiebe, W.J.: Temperature and substrate as interactive limiting factors for](#)
963 [marine heterotrophic bacteria, *Aquat. Microb. Ecol.*, 23, 187-204, 2001.](#)

964 [Rivkin, R.B., and Legendre, L.: Biogenic carbon cycling in the upper ocean: effects of microbial](#)
965 [respiration, *Science*, 291, 2398-2400, 2001.](#)

966 Rossmann, A., Butzenlechner, M., and Schmidt, H.L.: Evidence for a nonstatistical carbon
967 isotope distribution in natural glucose, *Plant Physiol.*, 96, 609-614, 1991.

968 Šantrůčková, H., Bird, M.I., and Lloyd, J.: Microbial processes and carbon-isotope fractionation
969 in tropical and temperate grassland soils, *Funct. Ecol.*, 14, 108-114, 2000.

970 Schimel, D.S.: Terrestrial ecosystems and the carbon-cycle, *Glob. Change Biol.*, 1, 77-91, 1995.

- 971 | [Smith, H.L., and Waltman, P.E.: The theory of the chemostat: dynamics of microbial](#)
972 | [competition, Cambridge University Press, New York, 1995.](#)
- 973 | Stumm, W., and Morgan, J.J.: Aquatic Chemistry: An introduction emphasizing chemical
974 | equilibria in natural waters, John Wiley & Sons, New York, 1981.
- 975 | Subke, J.-A., Inghima, I., and Cotrufo, M.F.: Trends and methodological impacts in soil CO₂ flux
976 | partitioning: A metaanalytical review, Glob. Change Biol., 12, 921-943, 2006.
- 977 | Szaran, J.: Achievement of carbon isotope equilibrium in the system HCO₃⁻ (solution) – CO₂
978 | (gas). Chem. Geol., 142, 79-86, 1997.
- 979 | Tcherkez, G., Mahé, A., and Hodges, M.: ¹²C/¹³C fractionations in plant primary metabolism,
980 | Trends Plant Sci., 16, 499-506, 2012.
- 981 | [Tempest, D.W., and Neijssel, O.M.: Eco-physiological aspects of microbial growth in aerobic](#)
982 | [nutrient-limited environments, Adv. Microb. Ecol., 2, 105-153, 1978.](#)
- 983 | Trumbore, S.: Carbon respired by terrestrial ecosystems – recent progress and challenges, Glob.
984 | Change Biol., 12, 141-153, 2006.
- 985 | Tucker, C.L., Bell, J., Pendall, E., and Ogle, K.: Does declining carbon-use efficiency explain
986 | thermal acclimation of soil respiration with warming?, Glob. Change Biol., 19, 252-263, 2013.
- 987 | Vogel, J.C., Grootes, P.M., and Mook, W.G.: Isotopic fractionation between gaseous and
988 | dissolved carbon dioxide, Z. Physik, 230, 225-238, 1970.

989 Werner, C., and Gessler, A.: Diel variations in the carbon isotope composition of respired CO₂
990 and associated carbon sources: a review of dynamics and mechanisms, *Biogeosci.*, 8, 2437-2459,
991 2011.

992 Werner, R.A., Buchmann, N., Siegwolf, R.T.W., Kornexl, B.E., and Gessler, A.: Metabolic
993 fluxes, carbon isotope fractionation and respiration – lessons to be learned from plant
994 biochemistry, *New Phytol.*, 191, 10-15, 2011.

995 Werth, M., and Kuzyakov, Y.: ¹³C fractionation at the root-microorganisms-soil interface: a
996 review and outlook for partitioning studies, *Soil Biol. Biochem.*, 42, 1372-1384, 2010.

997 Wieder, W.R., Bonan, G.B., and Allison, S.D.: Global soil carbon projections are improved by
998 modelling microbial processes, *Nat. Clim. Change*, 3, 909-912, 2013.

999 Wittmann, C., Weber, J., Betiku, E., Krömer, J., Böhm, D., and Rinas, U.: Response of fluxome
1000 and metabolome to temperature-induced recombinant protein synthesis in *Escherichia coli*, *J.*
1001 *Biotechnol.*, 132, 375-384, 2007.

1002

1003

1004

1005

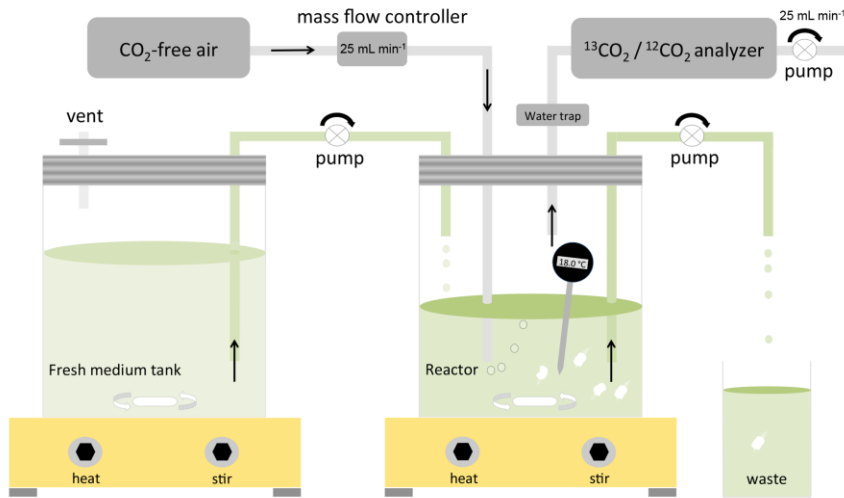
1006

1007

1008 **Figure legends:**

1009 **Fig. 1.** Chemostat system comprised of *P. fluorescens* growing on cellobiose. Seven independent
1010 experiments were conducted, with reactor temperatures of 13, 14.5, 16, 18, 21, 23.5 and 26.5 °C;
1011 all other conditions were identical. During continuous flow, dilution rate of the reactor
1012 (mean=0.137±0.01 h⁻¹ across all experiments) equals microbial growth rate. A peristaltic pump
1013 supplied fresh nutrient medium from a reservoir tank to the reactor and removed reactor medium
1014 (including biomass) at a constant rate. Headspace volume was flushed with CO₂-free air,
1015 bubbling through reactor medium and supplying microorganisms with O₂. A ¹³CO₂/¹²CO₂
1016 analyzer continuously sampled reactor headspace and measured the concentration and δ¹³C of
1017 respired CO₂.

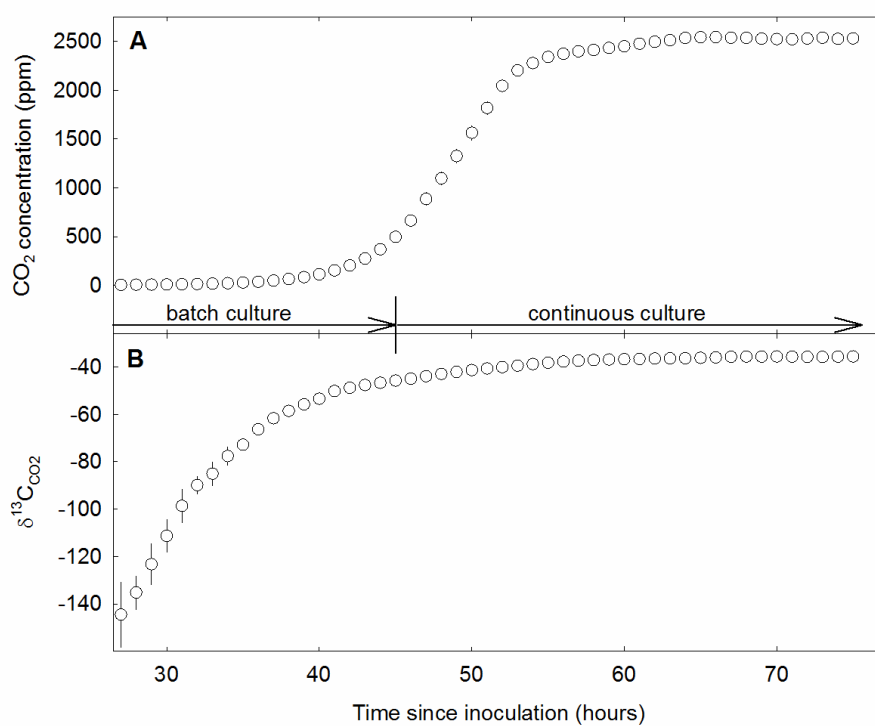
1018



1019

1020

1021 **Fig. 2.** Example time course of the evolution of reactor headspace CO₂ concentration (A) and
1022 $\delta^{13}\text{C}$ of the CO₂ (B) of the chemostat run at 23.5 °C in hours since inoculation of the reactor with
1023 pre-cultured *P. fluorescens*. Data points are hourly means. Error bars (where visible) denote ± 1
1024 SD. The reactor was shifted from batch to continuous culture mode 45 h after inoculation.
1025 Microbial respiration rate and the $\delta^{13}\text{C}$ of respired CO₂ were measured between 70 and 74 h after
1026 inoculation when the culture reached steady-state.

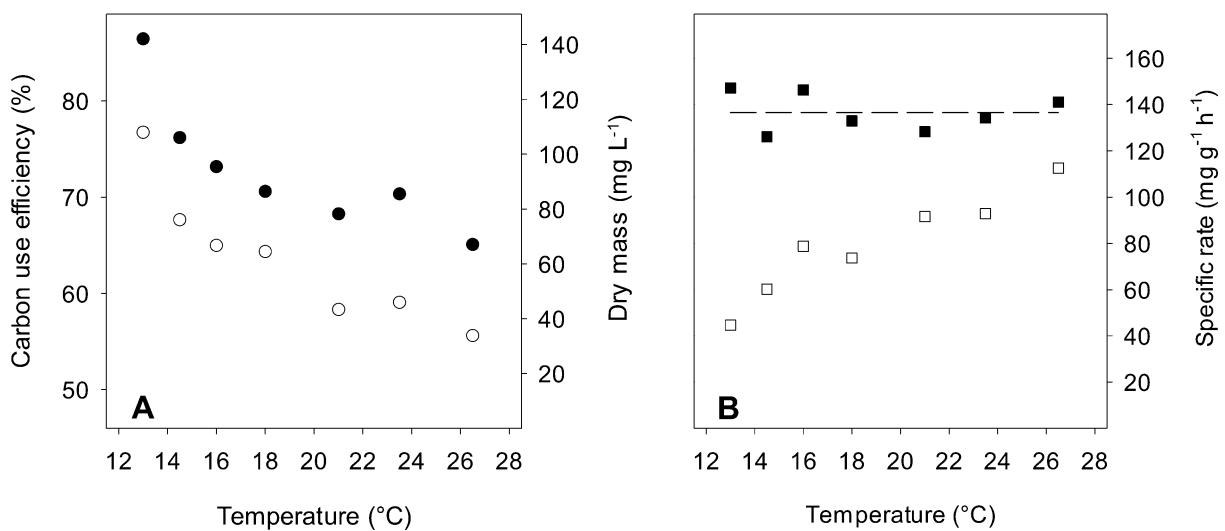


1027

1028

1029

1030 **Fig. 3.** Steady-state process variables of *P. fluorescens* growing in chemostats at specified
1031 temperatures. Microbial carbon use efficiency (○; A), dry microbial biomass (●; A), specific C
1032 growth rate (■; B), and specific C respiration rate (□; B), expressed per unit of microbial
1033 biomass-C. The dashed line denotes the average of the seven specific growth rates (137 mg g^{-1}
1034 h^{-1}).



1035

1036

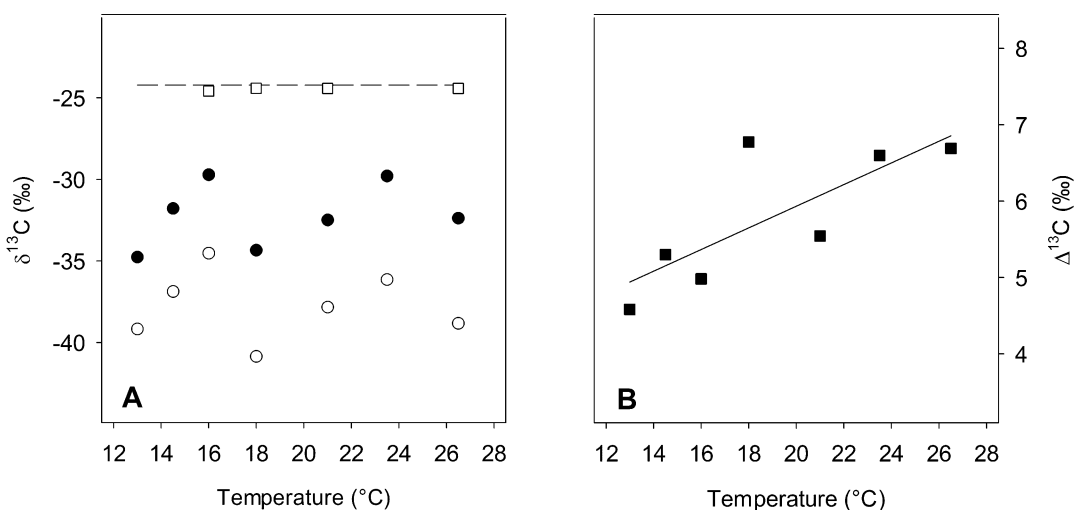
1037

1038

1039

1040

1041 **Fig. 4.** Steady-state $\delta^{13}\text{C}$ of microbial biomass (\bullet ; A) and of respired CO_2 (\circ ; A), and C isotope
 1042 discrimination during respiration ($\Delta^{13}\text{C}$; B) of *P. fluorescens* growing in chemostats at specified
 1043 temperatures. In panel A, the dashed line denotes the $\delta^{13}\text{C}$ of the substrate cellobiose (-24.2‰),
 1044 and $\delta^{13}\text{C}$ of reactor filtrate is shown as open squares; standard errors, derived from multiple
 1045 measurements across time during steady-state, are smaller than the size of the symbols. $\Delta^{13}\text{C}$ is
 1046 calculated as $\Delta^{13}\text{C} = (\delta^{13}\text{C}_{\text{biomass}} - \delta^{13}\text{C}_{\text{respired CO}_2}) / (1 + (\delta^{13}\text{C}_{\text{respired CO}_2} / 1000))$. The solid line
 1047 denotes linear regression of $\Delta^{13}\text{C}$ vs. temperature ($y = 0.14x + 3.1$; $R^2 = 0.61$; $P=0.04$).



1048

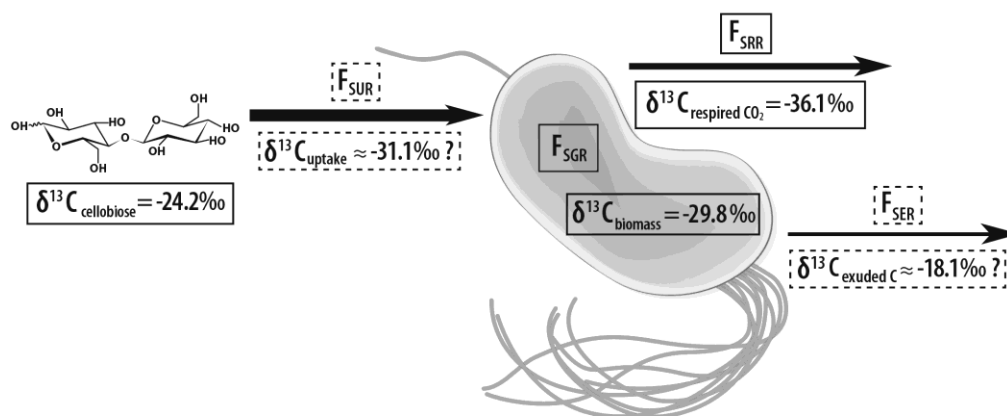
1049

1050

1051

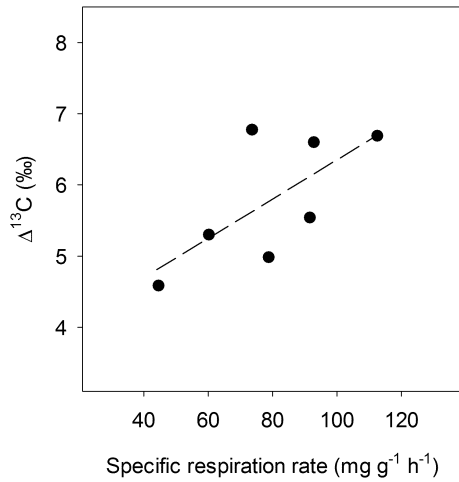
1052

1053 **Fig. 5.** Schematic of an individual *P. fluorescens* cell, representing ~~a~~the sample population
 1054 growing at 23.5 °C with one available substrate at a constant relative growth rate of 0.13 h⁻¹ at
 1055 steady-state (Fig. 2), with measured (solid boxed) and ~~unknown~~estimated (dashed boxed)
 1056 magnitudes of C and ¹³C fluxes into and out of the population. Designated fluxes include
 1057 specific uptake rate of cellobiose (F_{SUR}), specific growth rate (F_{SGR}), specific respiration rate
 1058 (F_{SRR}) and specific excretion rate (F_{SER}), in relation to steady-state biomass-C in the chemostat,
 1059 where F_{SUR} = F_{SGR} + F_{SRR} + F_{SER}. The estimate of δ¹³C_{uptake} is based on the assumption that
 1060 δ¹³C_{exuded C} is 11.7‰ less negative than δ¹³C_{biomass} (Blair et al., 1985) and that F_{SER} is 10% of the
 1061 sum of F_{SGR} and F_{SRR} (El-Mansi and Holms, 1989). In contrast to this experimental system, in
 1062 soils/natural environments, measurements of boxed pools and fluxes ~~are~~can be confounded by
 1063 the presence of dormant microorganisms, unknown microbial growth rates, diverse available
 1064 substrates, and a lack of steady-state CO₂ fluxes.



1065
 1066
 1067
 1068

1069 **Fig. 6.** Correlation between the specific respiration rate of *P. fluorescens* growing in continuous
1070 chemostat culture at temperatures ranging from 13 °C to 26.5 °C and the carbon isotope
1071 discrimination during respiration. The dashed line denotes a linear regression of the form $y =$
1072 $0.03x + 3.6$; $R^2 = 0.48$; $P=0.08$.



1073