

1 **Modeling spatiotemporal dynamics of global wetlands:**  
2 **Comprehensive evaluation of a new sub-grid**  
3 **TOPMODEL parameterization and uncertainties**  
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17  
18 **Abstract:** Simulations of the spatiotemporal dynamics of wetlands are key to  
19 understanding the role of wetland biogeochemistry under past and future  
20 climate. Hydrologic inundation models, such as TOPography-based hydrological  
21 model (TOPMODEL), are based on a fundamental parameter known as the  
22 *compound topographic index (CTI)* and offer a computationally cost-efficient  
23 approach to simulate wetland dynamics at global scales. However, there remains  
24 large discrepancy in the implementations of TOPMODEL in land-surface models  
25 (LSMs) and thus their performance against observations. This study describes  
26 new improvements to TOPMODEL implementation and estimates of global  
27 wetland dynamics using the LPJ-wsl (“Lund-Potsdam-Jena WSL version”)  
28 Dynamic Global Vegetation Model (DGVM), and quantifies uncertainties by  
29 comparing three digital elevation model (DEM) products (HYDRO1k, GMTED,  
30 and HydroSHEDS) at different spatial resolution and accuracy on simulated  
31 inundation dynamics. In addition, we found that calibrating TOPMODEL with a  
32 benchmark wetland dataset can help to successfully delineate the seasonal and  
33 interannual variation of wetlands, as well as improve the spatial distribution of  
34 wetlands to be consistent with inventories. The HydroSHEDS DEM, using a river-  
35 basin scheme for aggregating the CTI, shows best accuracy for capturing the  
36 spatiotemporal dynamics of wetlands among the three DEM products. The  
37 estimate of global wetland potential/maximum is  $\sim 10.3 \text{ Mkm}^2$  ( $10^6 \text{ km}^2$ ), with a  
38 mean annual maximum of  $\sim 5.17 \text{ Mkm}^2$  for 1980-2010. When integrated with  
39 wetland methane emission submodule, [the uncertainty of global annual CH<sub>4</sub>](#)  
40 [emissions from topography inputs is estimated to be 29.0 Tg yr<sup>-1</sup>.](#) This study

41 demonstrates the feasibility of TOPMODEL to capture spatial heterogeneity of  
42 inundation at large scale and highlights the significance of correcting maximum  
43 wetland extent to improve modeling of interannual variations in wetland area. It  
44 additionally highlights the importance of an adequate investigation of  
45 topographic indices for simulating global wetlands and shows the opportunity to  
46 converge wetland estimates across LSMs by identifying the uncertainty  
47 associated with existing wetland products.

48

49 **Keywords:** Seasonal wetland dynamics, DGVM, LPJ, methane emission,  
50 Topographic index, Compound topography index (CTI)

51

## 52 **Introduction**

53 For their ability to emit the greenhouse gas methane (CH<sub>4</sub>), wetland ecosystems  
54 play a disproportionately important role in affecting the global climate system  
55 through biogeochemical feedbacks (Fisher et al., 2011; Seneviratne et al., 2010).  
56 Wetlands are thought to be the largest natural source of CH<sub>4</sub> emission by  
57 contributing 20-40% of the total annual emissions to atmosphere, which adds a  
58 strong radiative forcing from CH<sub>4</sub> (Bousquet et al., 2006; IPCC, 2013). The  
59 seasonal and interannual distribution of wetland area remains one of the largest  
60 uncertainties in the global CH<sub>4</sub> budget (Kirschke et al., 2013), in particular for the  
61 roughly 60% of wetlands that are not inundated permanently (Petrescu et al.,  
62 2010). Changes in the spatial extent of seasonally inundated wetlands was most  
63 likely a major driver for CH<sub>4</sub> variations during last glacial period (Kaplan, 2002)  
64 and are considered as an important driver of the strong atmospheric CH<sub>4</sub> growth  
65 rate resumed in 2007 (Nisbet et al., 2014) and in future climate change scenarios  
66 (Stocker et al., 2013).

67

68 Improving our understanding of the role of wetlands in global greenhouse-gas  
69 (GHG) budgets requires a representation of wetlands and their biogeochemical  
70 processes in land surface models (LSM) to both hindcast observed past  
71 variations (Singarayer et al., 2011) and to predict future trajectories in  
72 atmospheric CH<sub>4</sub> and terrestrial C balance (Ito and Inatomi, 2012; Meng et al.,  
73 2012; Spahni et al., 2011; Stocker et al., 2014; Zürcher et al., 2013). Dynamic  
74 wetland schemes in LSMs were based on conceptual theories and physical  
75 processes describing surface water processes (e.g., infiltration and  
76 evapotranspiration) and water movement in the soil column using probability  
77 distributions derived from subgrid topographic information (Beven and Kirkby,  
78 1979), or using analytical functional parametric forms with fixed parameters  
79 (Liang et al., 1994). Currently, the most common approach for global wetland  
80 modelling is to use a runoff simulation scheme such as TOPMODEL  
81 (TOPOgraphy-based hydrological MODEL) (Beven and Kirkby, 1979; Kleinen et  
82 al., 2012; Ringeval et al., 2012; Zhu et al., 2014), which includes the assumption  
83 that lateral soil water transport driven by topography follows the same

84 exponential decline as the vertical decrease in hydraulic conductivity within soil  
85 profiles in a basin (Sivapalan et al., 1987).

86

87 TOPMODEL-based implementations have proven successful at capturing the  
88 broad geographic distribution of wetlands and their seasonal variability (Gedney  
89 and Cox, 2003; Ringeval et al., 2012; Stocker et al., 2014; Zhu et al., 2014), but  
90 have consistently overestimated both the extent of wetlands and duration of  
91 inundation at global and regional scale when compared with existing current  
92 surveys (Junk et al., 2011; Prigent et al., 2007; Quiquet et al., 2015). For instance,  
93 simulations using the Earth system model HadGEM2 predict much larger  
94 persistent Amazonian wetlands than an inventory (Collins et al., 2011). In  
95 general, independently determined wetland area using hydrologic modules of  
96 LSMs in The Wetland and Wetland CH<sub>4</sub> Inter-comparison of Models Projects  
97 (WETCHIMP) experiment simulated larger global wetland extent than those  
98 informed by remotely sensed product and inventories (Melton et al., 2013). This  
99 large disagreement also exists across specific regions (Ringeval et al., 2014). For  
100 example, Bohn et al. (2015) carried out a model inter-comparison of wetland  
101 extent on the West Siberian Lowland, one of the major wetland regions in high  
102 latitudes, and highlighted similar uncertainties of wetland extent simulation in  
103 the LSMs participating in the WETCHIMP experiment and using TOPMODEL.

104

105 Meanwhile, uncertainties in wetland area estimation partly come from a paucity  
106 of observational datasets and different definitions of wetland (Matthews and  
107 Fung, 1987). Remotely sensed datasets have difficulties in capturing small or  
108 isolated water in saturated soils that are not flooded on the surface (Prigent et  
109 al., 2007), as well capturing the forested wetlands that obscure detection of  
110 inundation because of dense forest canopies (Bohn et al., 2015). In addition,  
111 ground-based survey or inventories that determine wetlands usually limited as  
112 static distribution that cannot provide temporal patterns for inundated area,  
113 making it hard to evaluate with simulated results. On the other hand, the  
114 definition of wetland for regional- or global-scale modelling assumes the land  
115 surface has both inundated and saturated conditions, which is not necessarily  
116 the same as inundated area measured by satellite observations (Melton et al.,  
117 2013).

118

119 While prognostic wetland dynamics schemes are promising to resolve these  
120 observational issues, the configuration parameters for TOPMODEL are a  
121 potential source of uncertainty in estimating wetland dynamics (Marthews et al.,  
122 2015). Among all parameters in TOPMODEL, the Compound Topographic Index  
123 (CTI) is of critical importance for determining inundated areas in terrain-related  
124 hydrological applications (Ward and Robinson, 2000; Wilson and Gallant, 2000).  
125 It measures the tendency of soils to become saturated (Beven and Cloke, 2012)  
126 and consequently it drives the accuracy of wetland area scaled to the larger grid

127 cell (Ducharne, 2009; Mulligan and Wainwright, 2013). Although the importance  
128 of CTI has been highlighted, only few studies have so far evaluated the effect of  
129 CTI on modelling the spatial and temporal patterns of global wetland dynamics.  
130 This is due to a limited availability of global CTI products. During the last decade,  
131 the first CTI product at 1km resolution from HYDRO1k global dataset released by  
132 U.S. Geological Survey (USGS) in 2000 has become the most commonly applied  
133 global dataset for large-scale applications (Kleinen et al., 2012; Lei et al., 2014;  
134 Ringeval et al., 2012; Wania et al., 2013). However, HYDRO1k has been proven to  
135 potentially overestimate inundation extent due to a lack in quality of the  
136 underlying digital elevation model (DEM) (Grabs et al., 2009; Lin et al., 2010; Lin  
137 et al., 2013; Sørensen and Seibert, 2007;). With recent advances in the  
138 development of DEMs (Danielson and Gesch, 2011; Lehner et al., 2008), there is a  
139 both a requirement and an opportunity to investigate uncertainties caused by  
140 CTI parameter.

141

142 The primary goal of our study is to improve the modeling of dynamically varying  
143 wetland extents with i) a parameter constraint to match integrated satellite and  
144 inventory observations, and with ii) a better parameterizations of CTI values for  
145 determining wetland seasonal cycles using new topographic data and  
146 aggregation schemes (i.e., grid versus catchment). To this end, we develop a new  
147 version of Dynamic Global Vegetation Model (DGVM) LPJ-wsl (“Lund-Potsdam-  
148 Jena WSL version”) that includes the TOPMODEL approach for wetland extent  
149 modelling by also accounting for soil thermal dynamics and high-latitude soil-  
150 water freeze and thaw cycles, and by incorporating the necessary physical  
151 processes (e.g. snow aging) that constrain global wetland dynamics. We utilize  
152 three commonly used global DEM products to evaluate the effects of sub-grid  
153 parameterizations on simulated global wetland extent uncertainties. We perform  
154 six global simulations resulting from the combination of three DEM products and  
155 two aggregation schemes under the same common experimental protocol. The  
156 specific aims are: (1) to improve the performance of estimated wetland extent  
157 based on TOPMODEL for the purpose of large-scale modelling, (2) to develop a  
158 new parameterization scheme using inventory in combination with satellite-  
159 based retrievals, and (3) to evaluate the uncertainties and the spatial and  
160 temporal differences of CTI from three major DEM products in model behavior.

161

## 162 **2 Model Descriptions and Experimental Design**

163 The model LPJ-wsl is a process-based dynamic global vegetation model  
164 developed for carbon cycle applications based on development of the LPJ-DGVM  
165 (Sitch et al., 2003). LPJ-wsl includes land surface processes, such as water,  
166 carbon fluxes, and vegetation dynamics that are intimately represented by plant  
167 functional types (PFTs) (Poulter et al., 2011). The distribution of PFTs is  
168 simulated based on a set of bioclimatic limits and by plant-specific parameters  
169 that govern the competition for resources. The soil hydrology is modeled using

170 semi-empirical approach, with the soil treated as bucket consisting of two layers  
171 each with fixed thickness (Gerten et al., 2004). The LPJ-wsl CH<sub>4</sub> model used in  
172 this study is the same as presented in Hodson et al., (2011) and is a function of  
173 two scaling factors ( $r_{CH_4:C}$  and  $f_{ecosys}$ ), soil temperature, soil-moisture-dependent  
174 fraction of heterotrophic respiration, and wetland extent according to the  
175 following equation:

$$176 \quad E(x, t) = r_{CH_4:C} \cdot f_{ecosys}(x) \cdot A(x, t) \cdot R_h(x, t) \quad (1)$$

177 where  $E(x, t)$  is wetland CH<sub>4</sub> flux,  $A(x, t)$  is wetland extent,  $R_h(x, t)$  is  
178 heterotrophic respiration,  $f_{ecosys}$  is a scaling factor representing different wetland  
179 ecosystems,  $r_{CH_4:C}$  is the ratio C to CH<sub>4</sub> fluxes.

180

181 LPJ-wsl has been evaluated in previous studies using global inventory datasets  
182 and satellite observations and has been one of the participating models in the  
183 WETCHIMP study (Melton et al., 2013). Modifications made here to the original  
184 LPJ-wsl model and a detailed description of changes are summarized below:

185 - A permafrost module that simulate soil freeze and thaw processes, is  
186 implemented and modified following the Wania et al. (2009) study (see  
187 description in Sect. 2.1).

188 - The snow module from Wania et al. (2009) was included and modified to  
189 include some of the effects of snow ageing on snow thermal properties. We use  
190 an updated parameterization of soil thermal properties both for the permafrost  
191 and the snow module, which is calibrated by satellite observations specifically  
192 for global application.

193 - A new parameterization of soil texture was formulated based on the  
194 Harmonized World Soil Database (HWSD), which combines the recently  
195 collected extensive volumes of regional and national updates of soil parameter  
196 information (Nachtergaele et al., 2008). The new soil texture in LPJ-wsl follows  
197 the U.S. Department of Agriculture soil classification with 14 soil types grouped  
198 according to a particular range of particle-size fractions (e.g. sand, clay, loam,  
199 etc.), instead of using the original Food and Agriculture Organization of the U.N.  
200 classification with 9 soil types (Sitch et al., 2003). Thus, the volumetric water  
201 holding capacity, also defined as potential maximum soil water content (SWC), is  
202 assumed to vary spatially, calculated as a function of the surface soil texture  
203 using pedotransfer functions from Cosby et al., 1984. Wilting point, porosity,  
204 mineral soil content and organic soil content for each soil type are derived from  
205 a look-up table available from the Air Force Weather Agency (2002) as listed in  
206 Table 1.

207 The modified LPJ-wsl version is thus the starting point upon which the  
208 TOPMODEL-based wetland and permafrost modules are included (Sect. 2.2).

209

## 210 **2.1 Permafrost Model**

211 In order to consider the functional wetland area extension during the spring  
212 thaw and their shrinking or disappearances during autumn freeze, we added to

213 LPJ-wsl a soil temperature scheme and freeze-thaw processes, as in Wania et al.  
 214 (2009). The modified version considers the soil heat capacity and its thermal  
 215 conductivity, which are both affected by the volumetric fractions of the soil  
 216 physical components, such as water-ice fraction, mineral soil, or peat. The  
 217 thermal scheme of LPJ-wsl is discretized vertically using 8-layers of variable  
 218 thickness, while the water-balance scheme is kept the same as the original LPJ-  
 219 DGVM, which means the daily changes in water content are allocated to the “old”  
 220 upper and lower layer of LPJ while considering percolation between these two  
 221 layers and baseflow from the lower layer. Fractional water and ice content in  
 222 each of the 8-layers is calculated on a daily time step. Soil temperature is  
 223 updated in the thermal routine and then passed to the hydrological routine to  
 224 determine the water-ice phase change in permafrost routine.

225

## 226 **2.2 Dynamic Wetland Model**

227 To represent the grid cell fraction covered by wetlands, we have implemented an  
 228 approach based on the TOPMODEL hydrological framework (Beven and Kirkby,  
 229 1979). TOPMODEL was initially developed to operate at the scale of large  
 230 watersheds using the channel network topography and dynamics contributing  
 231 areas for runoff generation, and was later extended to perform over areas that  
 232 are much larger than a typical river catchment (Gedney and Cox, 2003). The  
 233 fundamental information to determine the area fraction with soil water  
 234 saturation is derived from knowledge of the mean watershed water table depth  
 235 and a probability density function (PDF) of combined topographic and soil  
 236 properties (Sivapalan et al., 1987). The CTI, which provides the sub-grid scale  
 237 topographic information in TOPMODEL, determines the likelihood of a grid box  
 238 to be inundated. It is defined as:

$$239 \quad \lambda_l = \ln\left(\frac{\alpha_l}{\tan\beta_l}\right) \quad (2)$$

240 where  $\lambda_l$  represents local CTI value,  $\alpha_l$  represent the contributing area per unit  
 241 contour,  $\tan\beta_l$ , the local topographic slope, approximates the local hydraulic  
 242 gradient where  $\beta$  is the local surface slope. The CTI distribution can be generated  
 243 from digital elevation models and near global datasets are readily available, e.g.  
 244 HYDRO1k dataset from USGS.

245

246 Following the central equations of TOPMODEL, the relationship between local  
 247 water table depth  $z_l$  and the grid mean water table depth  $z_m$  can be given as:

$$248 \quad \lambda_l - \lambda_m = f\{z_l - z_m\} \quad (3)$$

249 where  $\lambda_m$  is the mean CTI averaged over the grid box,  $f$  is the saturated  
 250 hydraulic conductivity decay factor with depth for each soil type. This equation  
 251 is valuable in that it relates the local moisture status to the grid box mean  
 252 moisture status based on the subgrid-scale variations in topography. Higher CTI  
 253 values than average are indicative of areas with higher water table depth than  
 254 average water table, and vice versa. We therefore calculate the inundated areas

255 ( $F_{wet}$ ) of all the sub-grid points within a grid cell that have a local water table  
 256 depth  $z_l \geq 0$ :

$$257 \quad F_{wet} = \int_{z_l}^{z_{max}} pdf(\lambda) d\lambda \quad (4)$$

258 where instead of using the CTI values themselves, we followed a common up-  
 259 scaling approach to approximate the distribution of CTI values within a grid cell  
 260 in order to reduce computation costs. Here, the discrete distribution of the CTI  
 261 for lowland pixels (i.e.  $\lambda_l \geq \lambda_m$ ) has been represented as an exponential function,  
 262 not as a three-parameter gamma distribution as applied in recent applications  
 263 for modeling wetland extent (Kleinen et al., 2012; Ringeval et al., 2012). As  
 264 shown in Figure 1, the new exponential function agrees well with the three-  
 265 parameter gamma distribution function when the CTI is larger than the mean  
 266 CTI  $\lambda_m$ . This change allows linking the inundated fraction directly to water table  
 267 depth, thus improving the parameterization by providing physical meaning and  
 268 fewer calibration parameters. This change also improves the parameterization of  
 269 fractional saturated area, especially in mountainous regions (Niu et al., 2005).

270

271 Finally, the wetland area fraction ( $F_{wet}$ ) is represented as:

$$272 \quad F_{wet} = F_{max} e^{-C_s f (\lambda_l - \lambda_m)} \quad (5)$$

273 Where  $C_s$  is a coefficient representing the topographic information generated by  
 274 fitting the exponential function to the discrete cumulative distribution function  
 275 (CDF) of the CTI.  $F_{max}$  is the maximum wetland fraction of a grid cell. Because of  
 276 the uncertainties involved in determining the water table depth, the hydraulic  
 277 factor  $f$ , and the coarse resolution DEMs, the maximum soil saturated fraction  
 278 calculated from discrete CDF are prone to large uncertainties and thus  
 279 complicate the comparison of the saturated fraction with existing observations  
 280 (Ducharne, 2009; Ringeval et al., 2012). Here, we introduce a calibration of  
 281 maximum wetland fractions  $F_{max}$ . We used the inventory-calibrated satellite  
 282 observations (see description in 3.3) combining with inventory dataset to  
 283 calculate representative long-term maximum wetland extents within each grid  
 284 box ( $0.5^\circ$ ), i.e. the parameter  $F_{max}$  for each grid cell i:

$$285 \quad F_{max_i} = \max(A_{GLWD_i}, \max(A_{SWAMP-GLWD_i})) \quad (6)$$

286  $A_{GLWD}$  represents wetland estimate from GLWD, and  $A_{SWAMP-GLWD}$  represents long-  
 287 term wetland estimate from SWAMPS-GLWD. The reason for combining these  
 288 two datasets is to take the advantage of satellite-based observations at capturing  
 289 temporal wetlands and inventory-based datasets at estimating forested wetlands  
 290 and small wetlands ignored by remote sensing. This calibration is also based on  
 291 the assumption that water is stagnant within local grids at large scale, in  
 292 particular for model using simple 'bucket' concept to calculate grid-mean water  
 293 table depth.

294 In addition, we used nonlinear least squares (nls) estimates to fit the discrete  
 295 CDF curve of CTI for lowlands ( $\lambda_l < \lambda_m$ ) to calculate parameter  $C_s$ , the parameter

296 that determines varying trend of wetland extent. By this, the parameters  
 297  $F_{max}$ ,  $\lambda_m$  and  $C_s$  for determining inundated areas are derived (Figure 2).

298

299 To account for the permafrost effects on soil infiltration properties, we followed  
 300 Fan and Miguez-Macho (2011) and Kleinen et al. (2012) who modified  $f$  by a  
 301 function  $k$  depending on January temperature  $T_{jan}$ . Since LPJ-wsl uses two soil  
 302 layers from the HWSD soil texture database (Nachtergaele et al., 2008) to  
 303 represent the different texture characteristics, the modification depends on the  
 304 combination of a look-up table (Table 1) from soil types and water table depth:

$$305 \quad k = \begin{cases} 1 & \forall T_{jan} > -5^{\circ}\text{C} \\ 1.075 + 0.015T_{jan} & -25^{\circ}\text{C} < \forall T_{jan} < -5^{\circ}\text{C} \\ 0.75 & \forall T_{jan} < -5^{\circ}\text{C} \end{cases} \quad (7)$$

306 Since the observed CH<sub>4</sub> emission during winter are mainly attributed to physical  
 307 processes during soil freezing effects (Whalen and Reeburgh, 1992), for the  
 308 partially frozen wetland in high latitudes, we introduced an effective fraction of  
 309 wetland area ( $F_{wet}^{eff}$ ) defined by:

$$310 \quad F_{wet}^{eff} = \left( \frac{\omega_{liq}}{\omega_{liq} + \omega_{froz}} \right)_{50\text{ cm}} \cdot F_{wet} \quad (8)$$

311

312 where  $\omega_{liq}$  and  $\omega_{froz}$  are the fraction of liquid and frozen soil water content in  
 313 the upper soil (0-0.5 m) respectively. Since the liquid water content in the lower  
 314 soil layer gets trapped and cannot contribute to CH<sub>4</sub> emission when upper soil is  
 315 frozen, we didn't consider the lower layer for surface wetland calculations.

316

### 317 **3 Experimental set-up and datasets**

#### 318 **3.1 Topographic information**

319 In this study we used three DEMs of varying spatial resolution, HYDRO1k at 30  
 320 arc-second (USGS, 2000; <http://lat.cr.usgs.gov/HYDRO1K>), Global Multi-  
 321 resolution Terrain Elevation Data 2010 (GMTED) at 15 arc-second (Danielson  
 322 and Gesch, 2011), and HydroSHEDS at 15 arc-second (Lehner et al., 2008) to  
 323 compare the effect of sub-grid topographic attributes on simulated seasonal and  
 324 interannual variability of wetlands. HYDRO1k, developed from the USGS released  
 325 30 arc-second digital elevation model of the world (GTOPO30), is the first  
 326 product that allowed spatially explicit hydrological routines applied in large-  
 327 scale applications (USGS, 2000). HydroSHEDS, developed from satellite-based  
 328 global mapping by the Shuttle Radar Topography Mission (SRTM), is a significant  
 329 improvement in the availability of high- resolution DEMs covering all land areas  
 330 south of 60°N (the limit of SRTM). For the areas at higher latitudes we used  
 331 HYDRO1k by aggregating the GTOPO30 DEM to provide global grids. GMTED was  
 332 produced using seven data sources including SRTM, global Digital Terrain  
 333 Elevation Data (DTED), Canadian elevation data, Spot 5 Reference3D data, and  
 334 data from the Ice, Cloud, and land Elevation Satellite (ICESat), covering nearly all  
 335 global terrain.



336

337 To avoid mismatch of CTI value inherent in computing CTI with different CTI  
338 algorithms, we generated three global CTI maps based on the three DEM  
339 products, instead of relying on existing CTI products (e.g. HYDRO1k CTI,  
340 HydroSHEDS CTI product from Centre for Ecology and Hydrology) (Marthews et  
341 al., 2015). Since studies show that multiple flow direction algorithms for  
342 calculating CTI give better accuracy compared with single-flow algorithms in flat  
343 areas (Kopecký and Čížková, 2010; Pan et al., 2004), thus we selected an  
344 algorithm from R library 'topmodel' (Buytaert, 2011), which applies the multiple  
345 flow routing algorithm of Quinn et al. (1995) to calculate the global CTI maps.  
346 The DEMs from HYDRO1k and HydroSHEDS had been previously processed for  
347 hydrological-correction, meaning that the DEMs were processed to remove  
348 elevation depressions that would cause local hydrologic 'sinks'. To include a  
349 comparison of (hydrologically) corrected and uncorrected DEMs in our analyses  
350 as some studies have been done previously (Stocker et al., 2014), the GMTED  
351 DEM was applied without hydrological correction.

352

### 353 **3.2 Description of the simulation**

354 For running LPJ-wsl with permafrost and TOPMODEL, we used global  
355 meteorological forcing (temperature, cloud cover, precipitation and wet days) as  
356 provided by the Climatic Research Unit (CRU TS 3.22) at 0.5° resolution (Harris  
357 et al., 2014). To spin up the LPJ-wsl model using the CRU climatology, climate  
358 data for 12-months were randomly selected from 1901-1930 and repeated for  
359 1000 years with a fixed pre-industrial atmospheric CO<sub>2</sub> concentration. The first  
360 spinup simulation started from initial soil temperature derived from LPJ-wsl  
361 simulated results on January 1901 and continued with a land use spin-up  
362 simulation. These procedures ensure that carbon stocks and permafrost are in  
363 equilibrium before performing transient simulations. The transient simulations,  
364 with observed climate and CO<sub>2</sub> were performed with monthly climate  
365 disaggregated to daily time steps over the 1901-2013 period. The 1993-2013  
366 years were used for evaluation against satellite data and inventories.

367 One of key assumptions in TOPMODEL is that the water table is recharged at a  
368 spatially uniform and steady rate with respect to the flow response timescale of  
369 the catchment (Stieglitz et al., 1997). Given the fact that we consider the water to  
370 be stagnant within each grid, the mean CTI parameter was estimated with two  
371 alternative schemes: (1) a regular 'grid-based' or gridded approach, i.e., the  
372 subgrid CTI values were averaged per 0.5° grids, and (2) an irregular 'basin-  
373 based' approach, where mean CTI were calculated over the entire catchment  
374 area in which the respective pixel is located. For generating a global catchment  
375 map at 0.5° resolution, we applied a majority algorithm in the case of multi-  
376 catchments in a grid with consideration of avoiding isolated pixels for specific  
377 river basin. There are two catchment area products applied in this study,  
378 HYDRO1k (2013) and HydroSHEDS. Similarly, the parameter C<sub>s</sub> was generated

379 using nonlinear least squares estimates from both of these two different CTI  
380 calculation strategies. Two sets of model experiments were carried out to  
381 compare the wetland dynamics under basin and grid-based TOPMODEL  
382 parameterizations respectively (Table 2).

383

### 384 **3.3 Evaluation and benchmarking data**

385 Since the soil freeze-thaw cycles are a key component for determining seasonal  
386 cycles of wetlands in cold regions, in this study we benchmarked the general  
387 pattern of permafrost locations by comparing the model output against satellite  
388 observations of freeze and thaw status and inventories of permafrost extent.  
389 Since soil depth in LPJ-wsl is held at 2.0 m for the permafrost module, the  
390 permafrost extent in this study is defined as the lower soil (0.5-2 m) that is  
391 always at or below the freezing point of water 0°C for multiple years. The  
392 permafrost extent map at 0.5-degree resolution from National Snow and Ice Data  
393 Center (NSIDC) is adopted for benchmarking (Brown et al., 2001). The global  
394 dataset of Freeze/Thaw (FT) status is derived from satellite microwave remote  
395 sensing provided by the Numerical Terradynamic Simulation Group (NTSG) at  
396 University of Montana and is based on daily maps over a 34-year record (1979-  
397 2012). It represents the FT status of the composite landscape vegetation-snow-  
398 soil medium to constrain surface water mobility and land-atmosphere carbon  
399 fluxes (Kim et al., 2012).

400

401 Two global inundation products derived from satellite observations were  
402 additionally used for evaluation purposes: the Global Inundation Extent from  
403 Multi-Satellites (GIEMS), derived from visible (AVHRR) and active (SSM/I) and  
404 passive (ERS) microwave sensors over the period 1993-2007; the Surface Water  
405 Microwave Product Series (SWAMPS), derived from active (SeaWinds-on-  
406 QuikSCAT, ERS, and ASCAT) and passive (SSM/I, SSMI/S, AMSR-E) microwave  
407 sensors over the period 1992-2013. This new SWAMPS global dataset, hereby  
408 denoted as SWAMPS-GLWD, was first developed at NASA JPL (Schroeder et al., In  
409 preparation). We re-scaled this dataset with the Global Lake and Wetland  
410 Database (GLWD) (Lehner and Döll, 2004), a well-established global inventory of  
411 water bodies at high resolution to match SWAMPS-GLWD with the inventory  
412 estimates. This post-processed SWAMPS product covers the required regions for  
413 forested wetlands, which are not readily observable by passive or active  
414 microwave measurements (Poulter, et al., in preparation). For evaluating  
415 regional wetland patterns, we selected two study areas (the largest peatland  
416 West Siberian Lowland (WSL); the largest floodplain, Amazon River Basin).  
417 Three wetland map products over the WSL from (Sheng et al., 2004), (Pregon et  
418 al., 2008) and (Tarnocai et al., 2009) (denoted by “Sheng2004”, “Pregon2008”,  
419 Tarnocai2009 respectively) and one up-date high resolution dual-season  
420 inundated area inventory for lowland Amazon basin from Japanese Earth  
421 Resources Satellite (JERS-1) were applied (Hess et al., 2015) (denoted by

422 “Hess2015”). We aggregated all above-mentioned datasets from the native 25  
423 km to a 0.5-degree spatial resolution and from daily to monthly temporal  
424 resolution for comparison with model outputs (Table A1).

425

## 426 **4 Results**

### 427 **4.1 Evaluation against observations**

428 We first evaluated the permafrost module that constrains the seasonal cycles of  
429 wetland area in cold regions with respect to inventory and remote sensing  
430 observations. Figure 3a compares the spatial distribution of permafrost extent  
431 from inventory and the modeled permafrost extent over the period 1980-2000.  
432 Figure 3b gives the spatial distribution of spearman rank correlation between  
433 the simulated and observed number of monthly frozen-days. The modeled  
434 permafrost extent shows high agreement with benchmarking dataset, with a  
435 slightly higher coverage of permafrost regions in North-Western Eurasia. The  
436 model successfully captures the seasonally frozen soil, which is closely linked to  
437 surface wetland formation and seasonal variation of wetland in cold regions.  
438 Most of the regions reveal a temporal correlation  $> 0.9$ , while Eastern Siberia and  
439 the Southern permafrost distribution edge is generally around 0.5. The lower  
440 correlation in East Siberia probably originates from two issues: high snow depth  
441 in LPJ-wsl that insulates soil temperature and consequent delay of soil  
442 temperature to reach complete freezing; and the relatively large uncertainty of  
443 FT-ESDR derived soil frozen status in those regions (Kim et al., 2012). This  
444 difference can be partly explained by the different representation of frozen  
445 status between simulated results and satellite retrievals. Remotely sensed maps  
446 reflect the mixed condition of the upper vegetation canopy, snow layer and  
447 surface soil, while the simulated frozen days only represent the frozen state of  
448 topsoil.

449

450 Figure 4 illustrates the model evaluation at the regional scale over the West  
451 Siberian Lowland (Figure 4). The model generally captures the spatial extent of  
452 the seasonal maximum wetland area fraction across the whole WSL for the JJA  
453 season successfully. However, the TOPMODEL approach without calibration  
454 (denoted as ‘Original’) shows large areas with relatively low wetland proportion  
455 and cannot capture high values. This suggests poor model performance in  
456 simulating wetland areas without  $F_{max}$  calibration. The calibrated model  
457 generally exhibits good agreement with inventories and satellite retrievals. It is  
458 especially successful at capturing the spatial heterogeneity of wetland areal  
459 extent over the whole WSL regions. LPJ-wsl simulated results reveal additional  
460 wetland area in the northeast, where wetlands entirely lacked in the GLWD map,  
461 although captured in other datasets. Meanwhile, LPJ-wsl captured the higher  
462 wetland area in region between 61 and 66°N and 70 and 80°E regions compared  
463 with GLWD, where mire/bog/fen was dominated across that region. LPJ-wsl also  
464 maintained well the spatial pattern of wetlands in forested region south of 60°N,

465 which was captured by inventories (Sheng2004, Peregon2008, and GLWD), but  
466 was missed by two satellite products (SWAMPS-GLWD, GIEMS) due to the  
467 limitation of remotely sensed datasets in detecting water under vegetative  
468 canopy and/or due to reduced sensitivity.

469 As illustrated in Figure 5, LPJ-wsl captured the spatial pattern of simulated  
470 wetlands well with lower estimates of the total wetland area in low-water season  
471 compared to the JERS-1 observed maps. Differences between Hess2015 and LPJ-  
472 wsl maps were primarily in two regions, Marañon-Ucayali region of Peru (MUP,  
473 3-7°S, 73-77°W) and Llanos de Moxos in Bolivia (LMB, 11-17°S, 60-68°W). LPJ-  
474 wsl shows higher wetland coverage in MUP while Hess2015 indicates high  
475 wetland fraction in LMB in high-water season. Global satellite products largely  
476 ignore the LMB region that was partly captured in LPJ-wsl, indicating that LPJ-  
477 wsl using hybrid TOPMODEL approach can yield estimates closer to those of  
478 fine-resolution mapping, while large-scale satellite products are likely to  
479 underestimate Amazon wetland extent because of their coarse spatial resolution  
480 that limit the ability to detect inundation outside of large wetlands and river  
481 floodplains (Hess et al., 2015).

482 To evaluate the effect of  $F_{max}$  calibration on  $CH_4$  emission estimates, two  
483 estimates of  $CH_4$  (with and w/o calibration) over the WSL regions were  
484 compared with observation-based estimate from Glagolev et al. (2011) (Figure  
485 6). The 3-year mean annual total emission from original version is  $6.29 \pm 0.51$  Tg  
486  $CH_4$   $yr^{-1}$ , falling into the upper part of range from land surface models and  
487 inversions (Bohn et al., 2015), whereas the calibrated version is close to the  
488 estimate of Glagolev et al. (2011) ( $3.91 \pm 1.29$  Tg  $CH_4$   $yr^{-1}$ ) with  $4.6 \pm 0.45$  Tg  $CH_4$   
489  $yr^{-1}$ . In addition, the spatial pattern of  $CH_4$  emission with  $F_{max}$  calibration shows  
490 better agreement with observation than non-calibration one with relatively  
491 larger emissions in Taiga forests and central region (55-65°N, 65-85°E). We also  
492 compared our estimates with recent airborne campaign observations for Alaska  
493 during 2012 growing season. Estimates with  $F_{max}$  calibration also falls well into  
494 the range of recent estimate ( $2.1 \pm 0.5$  Tg  $CH_4$   $yr^{-1}$ ) for Alaska based on airborne  
495 observations (Chang et al., 2014) with a total of 1.7 Tg  $CH_4$   $yr^{-1}$  during 2012  
496 growing season ( $3.1$  Tg  $CH_4$   $yr^{-1}$  from non-calibrated estimate), indicating the  
497 necessity to calibrate  $F_{max}$  to accurately capture annual  $CH_4$  emission and spatial  
498 variability for boreal wetlands.

#### 499 **4.2 Spatial distribution**

500 Several observations applicable to evaluate the difference among sub-grid  
501 parameterizations of TOPMODEL are available for the WSL region. Figure 7 lists  
502 the spatial patterns of simulated JJA (June-July-August) wetland area over WSL  
503 regions to illustrate differences among wetland maps. The general patterns of  
504 wetland extent are substantially similar, because they both used the same  
505 calibrated  $F_{max}$  map. Both of these datasets show wetlands distributed across  
506 most of the WSL, with extensive wetlands in the central region (55-65°N, 60-  
507 90°E). However, the detailed pattern is differing between the approaches and

508 DEMs used, which indicate the uncertainty of parameterizations on wetland  
509 distribution. The basin-based parameterization can capture the higher wetland  
510 areas in regions with bog, mire, or fen vegetation in the central east (63-67°N,  
511 85-90°E) as was found in the GLWD benchmark map. The grid-based  
512 parameterizations fail to reproduce this pattern. It seems that the grid-based  
513 parameterizations are less sensible in capturing the spatial heterogeneity  
514 throughout most of the WSL. The difference in parameterization derived from  
515 DEM datasets also affects the simulated regional pattern. Both of HydroSHEDS-  
516 based results successfully reproduce the high wetland fractions in the southern-  
517 forested regions (55-60°N, 65-80°E), while HYDRO1k and GMTED both cannot  
518 capture this feature. Note that GMTED is derived from the same DEM product  
519 SRTM as HydroSHEDS but without hydro-correction, indicating the importance  
520 of hydro-correction in simulating spatial patterns of wetlands.

521

522 The comparison of simulated mean annual minimum, maximum, and amplitude  
523 of wetland extent with observational datasets (Table 3) reveals that the  
524 simulated wetland area for 1980-2010 falls within the range of  $4.37 \pm 0.99$  Mkm<sup>2</sup>  
525 (Mkm<sup>2</sup>=10<sup>6</sup> km<sup>2</sup>). This number is close to GIEMS (5.66 Mkm<sup>2</sup>) (Prigent et al.,  
526 2012) and inventory-based estimates (6.2 Mkm<sup>2</sup>) (Bergamaschi et al., 2007)  
527 after exclusion of other water bodies like lakes, rivers, and rice paddies (Leff et  
528 al., 2004). Considering potential underestimation of satellite-based observation  
529 in forested regions, the realistic estimate could possibly be in the upper part of  
530 our range. Note that one must be careful when directly comparing model results  
531 with the observational datasets based on inventories or digitized maps, because  
532 these datasets might represent the long-term maximal area as wetland potential.  
533 The higher seasonal wetland extent in GIEMS compared with LPJ-wsl could be  
534 partly due to permanent wetlands that are difficult to detect by GIEMS. Lastly,  
535 the definition of wetland is another possible source of discrepancy. Remotely  
536 sensed inundation datasets emphasizes on open water while wetland area in our  
537 study is specifically defined from inventories following the National Wetlands  
538 Working Group (1988) classification that include peatlands, mineral wetlands,  
539 and seasonally inundated shallow waters.

540

#### 541 **4.3 Seasonal cycle**

542 The shapes of the seasonal patterns in wetland area are generally similar in  
543 model simulation compared to satellite observations, despite disagreement in  
544 the timing of the seasonal cycle of wetland area in some boreal regions (Figure  
545 8). The modeled results show slightly larger wetland areas in the SON (Sept-Nov)  
546 months than satellite-based observations. The larger seasonal wetland areas  
547 during SON may originate from the longer periods of unfrozen and relatively  
548 water saturated soil in the model data. It thus seems realistic that the satellite-  
549 based inundation product AMSR-E observed a similar trend of seasonal  
550 inundation patterns for North America and Boreal Eurasia (Jennifer et al., 2014).

551 This is also supported by field studies in boreal regions, indicating that water  
552 table depth during the SON months is still in a high level and soil temperature is  
553 above freezing status (Rinne et al., 2007; Turetsky et al., 2014). In contrast, the  
554 modeled seasonal cycle of wetland in tropical and temperate regions show a  
555 good agreement with GIEMS and SWAMPS-GLWD. Given the difficulties of  
556 satellite-based observations in detecting wetlands in forested regions and the  
557 reduced sensitivity where open water fraction is low (<10%) (Prigent et al.,  
558 2007), the inundation numbers by GIEMS might slightly underestimate the area  
559 compared with the simulated results.

560

561 Figure 8 reveals that the six data sets of monthly wetland extent for 1993-2007  
562 based on different TOPMODEL parameterization show the same general  
563 behavior in the different regions. The six data sets are highly correlated, with  
564 largest differences at the maximal wetland extents during growing seasons,  
565 especially in the boreal regions. In addition, the differences in seasonal cycle  
566 among the six model experiments are relatively small, mostly below 5%  
567 regardless of the month. This indicates that the averaged total wetland area is  
568 not dependent on the introduction of the new sub-grid parameterizations at the  
569 global scale. Among the DEM datasets, HYDRO1k shows the largest difference  
570 between basin and grid-based estimates with annual mean wetland area of  
571 89,663 km<sup>2</sup> in boreal regions, while HydroSHEDS has a lowest difference of 6550  
572 km<sup>2</sup> between the two versions. Examining the seasonal amplitude for basin-  
573 based schemes, HydroSHEDS shows a better agreement with satellite-based  
574 observations than the other two datasets.

575

#### 576 **4.4 Interannual variability**

577 For evaluating the performance of all the sub-grid parameterizations, we  
578 calculated the Pearson's correlation coefficient ( $r$ ) between modeled and  
579 satellite-based results (Table 4). Generally, the comparison demonstrates that  
580 simulated interannual variability shows a good agreement with GIEMS and  
581 SWAMPS-GLWD in most regions as defined in Fig. 2. For boreal and tropical  
582 regions, all correlation coefficients are ranging from 0.7-0.8. The comparison of  
583 the inter-annual trends (Figure A1) indicates that absolute values of simulated  
584 interannual variations are close to satellite-based observation with good  
585 agreement in shape and timing in these regions. This demonstrates the ability of  
586 TOPMODEL to capture the large-scale variations in wetland/inundation. Highest  
587 disagreements are found in temperate regions that are strongly affected by  
588 human activities (likely strong global anthropogenic effect on continental surface  
589 freshwater), which is indicated by GIEMS (Prigent et al., 2012) but not by  
590 modeled results.

591

592 The interannual variability originating from six different sub-grid DEM  
593 parameterizations is very similar between these schemes with Spearman rank

594 correlation coefficient  $r > 90\%$ . Among the six schemes, the parameters  
595 calculated from HydroSHEDS using basin-based statistics result in better  
596 agreement between simulated and measured wetland area than the other  
597 schemes. In most regions, the SWAMPS-GLWD and GIEMS are consistent in their  
598 observed wetland area patterns, except for temperate regions (e.g. Temperate  
599 South America, Temperate North America, Europe). This confirms that the  
600 differences in surface water extent detection between GIEMS and SWAMPS-  
601 GLWD, which might be caused by observational behaviors from different satellite  
602 instruments and algorithms. In addition, parameters estimation based on river  
603 basins are slightly better than grid-based results.

604

## 605 **5. Discussion**

### 606 **5.1 Wetland modelling based on TOPMODEL concept**

607 The coupling between LPJ-wsl and TOPMODEL with calibrated parameters as  
608 described in this study, improves the dynamic simulation of wetlands, in  
609 particular their geographic location and extent. This is based on the recent  
610 discussions of the suitability of TOPMODEL applications to simulate wetland  
611 variations at large spatial scale (Ringeval et al., 2012), and intercomparisons of  
612 the wetland-area-driven model bias in  $\text{CH}_4$  emission at regional scale (Bohn et  
613 al., 2015). The large discrepancies of wetland area among LSMs so far have  
614 shown extensive disagreement with inventories and remotely sensed inundation  
615 datasets (Melton et al., 2013), which is partly due to large varieties of schemes  
616 used for representing hydrological processes, or due to the inappropriate  
617 parameterizations for simulating inundations. Our results suggest that  
618 benchmarking  $F_{\max}$  is necessary for global wetland modelling.

619

620 The simulation of hydrological dynamics within LSMs remains relatively simple  
621 because the physical processes described in LSMs occur at much finer spatial  
622 scales (Ducharne, 2009; Mulligan and Wainwright, 2013). The coupling of  
623 TOPMODEL with process-based LSMs allows for retrieving of the maximum  
624 saturated fraction ( $F_{\max}$ ), which is defined by the pixels with no water deficit  
625 estimated from the partial integration of the spatial distribution of CTI in a  
626 catchment. The estimated distribution of  $F_{\max}$  is much larger than that obtained  
627 from the satellite-based observations (Papa et al., 2010). As a key parameter for  
628 determining the soil saturated area, the calculation of  $F_{\max}$  at large scale is prone  
629 to large uncertainties, in particular linked to uncertainties in topographic  
630 information, as well as the hydrological processes implemented in large-scale  
631 LSMs. Ringeval et al. (2012) pointed to the difficulty of two-layer bucket  
632 hydrological model in estimating the mean deficit to the saturation over each  
633 grid-cell. This can lead to nonrealistic absolute values of the contributing area in  
634 a watershed. We constructed several strategies for optimizing  $F_{\max}$  by correcting  
635 topographic information to match the wetland inventories (Gedney and Cox,  
636 2003; Kleinen et al., 2012). This is one possible solution for global wetland

637 modeling as it assumes that wetland area can be considered constant at coarse  
638 spatial resolution (e.g. 0.5° or 1°), following the classical approach of Beven and  
639 Kirkby (1979). However, due to the uncertainties from topographic information  
640 used in global applications and due to limitations in model parameterizations,  
641 this approximation cannot capture the fine scale wetland extent, which  
642 complicates the comparison with inventories.

643

644 The integration of satellite-based and inventory-based observations to calibrate  
645  $F_{\max}$  is highlighted in this study. Combining SWAMPS and GLWD led to simulated  
646 wetland area consistent with detailed regional distribution (Poulter et al., in  
647 preparation). Our estimation of global wetland potential/maximum is  $\sim 10.3$   
648  $\text{Mkm}^2$ , and in agreement with the deduction ( $10.4 \text{ Mkm}^2$ ) from recent estimates  
649 at finer resolution for total open water ( $\sim 17.3 \text{ Mkm}^2$ ) (Fluet-Chouinard et al.,  
650 2015), lakes ( $\sim 5 \text{ Mkm}^2$ ) (Verpoorter et al., 2014), and rice paddies ( $1.9 \text{ Mkm}^2$ )  
651 (Leff et al., 2004). The calibration of  $F_{\max}$  allows for simulating the dynamics of  
652 wetland on decade-to-century long time scales. As shown in Figure 9, the  
653 wetland potential for permafrost and arid/semi-arid regions is high. Even in  
654 tropical regions, there is  $\sim 20\text{-}30\%$  of potential for areas to be inundated.

655

656 According to our evaluation using satellite-based observations and inventories,  
657 the spatial distribution of the wetland areas and its temporal variability are  
658 generally well captured by our model, both at regional and global scales. In  
659 addition, the modeled wetland areas and interannual variability compare well  
660 with inventories and satellite-based observations respectively. Unfortunately,  
661 the wide disagreement in simulated wetland dynamics among estimates from  
662 WETCHIMP hampers our ability to assess model performance (Bohn et al.,  
663 2015). Narrowing down the uncertainty of wetland areas by existing maps could  
664 minimize the controversial use of the definition between wetlands and  
665 inundations. Wetlands have considerable variations in hydrologic conditions,  
666 size, locations that make difficult to reconcile a single definitions of wetlands. In  
667 current parameterizations, the connectivity of wetlands cannot be represented  
668 since wetlands are considered invariant within grid cells.

669

## 670 **5.2 CTI parameterizations**

671 As shown in this study, global wetland simulations can benefit from improved  
672 spatial resolution of topographic maps, thus creating a more realistic  
673 representation of processes at sub-grid resolution, and correspondingly better  
674 inundation simulations. This is supporting the ideas of Wood et al. (2011) who  
675 claimed that higher-resolution modeling leads to better spatial representation of  
676 saturated and nonsaturated areas, even though limitations in up-scaling  
677 parameterizations may potentially outrun this advantage. The comparison  
678 between HydroSHEDS and GMTED also indicated that, for capturing inundated  
679 areas under the same spatial resolution, the parameter maps derived from DEM



680 without hydrological corrections have less accuracy compared to corrected ones  
681 (Lehner and Grill, 2013). Without hydrological corrections, valleys would appear  
682 as closed depressions in the DEM, leading to an underestimation of inundated  
683 areas (Marthews et al., 2015). It could be foreseen that if DEMs in process-based  
684 models are being applied at higher resolution, this drawback could be amplified.  
685 The comparison between basin- and grid-based parameterizations suggests that  
686 grid-based calculations are not appropriate and consequently underestimates  
687 wetland areas even when assuming invariant inundated areas at large scale.

688

689 The algorithm to calculate CTI is another potential source of error for modelling  
690 inundations. The method we applied here is based on calculating a CTI  
691 distribution map using a simple algorithm in the R package 'topmodel' instead of  
692 using an existing CTI product with improved contributing area. The algorithm  
693 we applied using the multi-flow direction algorithm that allows for multiple in-  
694 flow and out-flow of water among neighboring pixels when generating  
695 topographic values. This could potentially overestimate the contributing areas  
696 (Pan et al., 2004). As a result, it might underestimate the wetland areas within  
697 each grid cell, and slightly underestimate the temporal pattern of saturated areas  
698 because of improper estimates of parameter  $C_s$  (Güntner et al., 2004). One  
699 limitation of HydroSHEDS is that its projection is not equal-area like HYDRO1k  
700 (Marthews et al., 2015), and will cause a potential bias in slope calculation along  
701 east-west directions at high latitudes. However, since there is no common  
702 method to calculate slope or flow direction, we believe that our calculations  
703 provide a reasonable approximation for global applications.

704

705 In addition, variability in TOPMODEL parameterizations have considerable  
706 influence on simulated  $\text{CH}_4$  fluxes, so that the uncertainty of mean annual  $\text{CH}_4$   
707 emissions from variable topography inputs is estimated to be  $29.0 \text{ Tg yr}^{-1}$  (Table  
708 5). Nevertheless, all of the model estimates generally fall within the value range  
709 of inversion estimates. The differences of  $\text{CH}_4$  emissions among the model  
710 experiments is related to simulated magnitude of wetland extents because the  
711 fraction of  $\text{CH}_4$  emissions from the tropics ( $\sim 63\%$ ) and Extratropics ( $\sim 27\%$ )  
712 remain constant due to the same parameters  $r_{\text{C:CH}_4}$  and  $f_{\text{ecosys}}$ . The importance of  
713 hydrological correction is highlighted by results based on GMTED, suggesting  
714 that applying DEMs without hydro-correction may potentially underestimate  
715  $\text{CH}_4$  fluxes due to lower hydrological connectivity, which dampen the generation  
716 of inundation. In addition, fine-scale topography data like HydroSHEDS reveals  
717 higher  $\text{CH}_4$  fluxes than HYDRO1k, highlighting the importance to capturing small  
718 wetlands/inundated areas that may be ignored by coarse-resolution products.

719

720

### 721 **5.3 Future needs for global wetland modelling**

722 Substantial progress has been made in the development of wetland modeling,  
723 but the wide disagreement among estimates from LSMs still exists (Bohn et al.,  
724 2015; Melton et al., 2013). Considering that spatiotemporal variation of wetland

725 area can largely influence CH<sub>4</sub> emissions, the selection of appropriate maps  
726 needs to be done with care. The parameterization and evaluation of multi-  
727 resolution topographic products presented in this study would enhance global  
728 wetland modeling if progress could be made in four areas particularly:

- 729 • *Improved parameters of TOPMODEL for large-scale application.* Our results  
730 demonstrate that model simulation after calibrating TOPMODEL are  
731 comparable in absolute value with inventories and satellite-based  
732 observations at coarser resolution. This supports the ideas of (Beven and  
733 Cloke, 2012) that an appropriate scale-dependent subgrid  
734 parameterization is the main challenge, regardless of whether it is carried  
735 out at global modeling scales or landscape scales. The saturated soil water  
736 content is the decisive unit that determines wetland distributions and  
737 reasonable estimates of global wetland areas. Hydraulic parameters,  
738 which describe soil characteristics for water movement, are critical for  
739 modelling wetland seasonal cycles (Marthews et al., 2014). Assessing the  
740 uncertainties introduced by aggregating sub-pixel to pixel areas also need  
741 to be evaluated.
- 742 • *Implementing human impact within wetland modeling.* There are  
743 evidences from long-term satellite-based observations detecting a  
744 significant effect of human activities on wetland drainage at continental  
745 scale (Prigent et al., 2012). At finer scale, the variability of wetland extent  
746 has also been affected by land-use change (e.g. wetland restoration,  
747 deforestation, drainage for forestry, agriculture, or peat mining) and  
748 consequently influences spatiotemporal patterns of CH<sub>4</sub> emission  
749 (Petrescu et al., 2015; Zona et al., 2009). Land-use change may therefore  
750 feedback water available to wetlands through altering water balance  
751 between land surface and atmosphere (Woodward et al., 2014). An  
752 implementation of human impacts within LSMs at large scale may be  
753 important for accurate estimation of interannual variations of wetlands.
- 754 • *Improved modelling of soil moisture.* The quality of soil moisture  
755 simulation using LSMs depends largely on the accuracy of the  
756 meteorological forcing data, surface-atmosphere interaction schemes, and  
757 a wide range of parameters (Zhang et al., 2013) (e.g. CO<sub>2</sub> concentration,  
758 albedo, minimum stomatal resistance, and soil hydraulic properties). As  
759 the fundamental variable for determining water table depth at global  
760 scale (Fan et al., 2013), soil moisture plays a key role in simulating the  
761 spatiotemporal variability of wetland dynamics. Since it is impossible to  
762 produce accurate large-scale estimates of soil moisture from in situ  
763 measurement networks (Bindlish et al., 2008; Dorigo et al., 2011),  
764 simulation combined with long-term surface and root zone remotely  
765 sensed estimates (de Rosnay et al., 2013; Kerr et al., 2010) via data  
766 assimilation technology, represents a strategy to improve the capturing of  
767 global wetland variability. Future hydrology-oriented satellite missions

768 such as Soil Moisture Active Passive (SMAP) (Entekhabi et al., 2010), and  
769 Surface Water and Ocean Topography (SWOT) mission (Durand et al.,  
770 2010) are expected to provide soil moisture and will improve the capacity  
771 of global soil moisture simulations.

772 • *Improved satellite benchmark observations.* Current satellite-based  
773 estimates of wetland area remain generally uncertain, despite being  
774 important for monitoring global wetland variability. Remotely sensed  
775 global inundation is prone to underestimate small wetlands, as well as  
776 covered with dense vegetation canopies (Papa et al., 2010). Moreover,  
777 estimated coastal areas show large bias due to interference with the  
778 ocean surface (Prigent et al., 2007). This raises the need for benchmark  
779 dataset useful to generate accurate products with lower uncertainties.  
780 Downscaling methodology has been made to refine existing satellite-  
781 based inundation estimates by coupling the mapping process with  
782 reliable inventories (Fluet-Chouinard et al., 2015). This may improve  
783 global inundation products, as well as the TOPMODEL parameter  
784 estimation in the future.

785

## 786 **Conclusion**

787 The new LPJ-wsl version incorporates a TOPMODEL approach and a permafrost  
788 module representing soil freeze-thaw processes to simulate global wetland  
789 dynamics. Once the  $F_{max}$  parameter in TOPMODEL was calibrated against a  
790 benchmark dataset, the model successfully mapped regional spatial pattern of  
791 wetlands in West Siberian Lowland and lowland Amazon basin, and captured  
792 well the spatiotemporal variations of global wetlands. The parameterization of  
793 TOPMODEL based on three DEM products, HYDRO1k, GMTED, and HydroSHEDS  
794 revealed that HydroSHEDS performed best in capturing the spatial heterogeneity  
795 and interannual variability of inundated areas compared to inventories. River-  
796 basin based parameterization schemes using HYDRO1k and GMTED marginally  
797 but significantly improve wetland area estimates. The estimates of global  
798 wetland potential/maximum is  $\sim 10.3$  Mkm<sup>2</sup>, with a mean annual maximum of  $\sim$   
799 5.17 Mkm<sup>2</sup> for 1980-2010. This development of the wetland modeling method  
800 reduces the uncertainties in modeling global wetland area and opens up new  
801 opportunities for studying the spatiotemporal variability of wetlands in LSMs  
802 that are directly comparable with inventories and satellite datasets.

803

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811

812

813 **References**

814 Air Force Weather Agency: Data Format Handbook for AGRMET, technical  
815 report, Air Force Weather Agency's (AFWA) Agricultural Meteorological  
816 modeling system (AGRMET), Air Force Weather Agency, available at:  
817 [http://www2.mmm.ucar.edu/mm5/documents/DATA\\_FORMAT\\_HANDBOOK.pdf](http://www2.mmm.ucar.edu/mm5/documents/DATA_FORMAT_HANDBOOK.pdf)  
818 f (last access: July 2014), 2002.

819 Bergamaschi, P., Frankenberg, C., Meirink, J. F., Krol, M., Dentener, F., Wagner, T.,  
820 Platt, U., Kaplan, J. O., Körner, S., Heimann, M., Dlugokencky, E. J., and Goede, A.:  
821 Satellite cartography of atmospheric methane from SCIAMACHY on board  
822 ENVISAT: 2. Evaluation based on inverse model simulations, *Journal of*  
823 *Geophysical Research: Atmospheres*, 112, D02304, 2007.

824 Beven, K. J. and Cloke, H. L.: Comment on "Hyperresolution global land surface  
825 modeling: Meeting a grand challenge for monitoring Earth's terrestrial water" by  
826 Eric F. Wood et al, *Water Resources Research*, 48, W01801, 2012.

827 Beven, K. J. and Kirkby, M. J.: A physically based, variable contributing area model  
828 of basin hydrology / Un modèle à base physique de zone d'appel variable de  
829 l'hydrologie du bassin versant, *Hydrological Sciences Bulletin*, 24, 43-69, 1979.

830 Bindlish, R., Jackson, T. J., Gasiewski, A., Stankov, B., Klein, M., Cosh, M. H.,  
831 Mladenova, I., Watts, C., Vivoni, E., Lakshmi, V., Bolten, J., and Keefer, T.: Aircraft  
832 based soil moisture retrievals under mixed vegetation and topographic  
833 conditions, *Remote Sensing of Environment*, 112, 375-390, 2008.

834 [Bloom, A. A., Palmer, P. I., Fraser, A., and Reay, D. S.: Seasonal variability of  
835 tropical wetland CH<sub>4</sub> emissions: the role of the methanogen-available carbon  
836 pool, \*Biogeosciences\*, 9, 2821-2830, 2012.](#)

837 [Bloom, A. A., Palmer, P. I., Fraser, A., Reay, D. S., and Frankenberg, C.: Large-Scale  
838 Controls of Methanogenesis Inferred from Methane and Gravity Spaceborne  
839 Data, \*Science\*, 327, 322-325, 2010.](#)

840 Bohn, T. J., Melton, J. R., Ito, A., Kleinen, T., Spahni, R., Stocker, B. D., Zhang, B., Zhu,  
841 X., Schroeder, R., Glagolev, M. V., Maksyutov, S., Brovkin, V., Chen, G., Denisov, S.  
842 N., Eliseev, A. V., Gallego-Sala, A., McDonald, K. C., Rawlins, M. A., Riley, W. J.,  
843 Subin, Z. M., Tian, H., Zhuang, Q., and Kaplan, J. O.: WETCHIMP-WSL:  
844 intercomparison of wetland methane emissions models over West Siberia,  
845 *Biogeosciences*, 12, 3321-3349, 2015.

846 Bousquet, P., Ciais, P., Miller, J. B., Dlugokencky, E. J., Hauglustaine, D. A., Prigent,  
847 C., Van der Werf, G. R., Peylin, P., Brunke, E. G., Carouge, C., Langenfelds, R. L.,  
848 Lathiere, J., Papa, F., Ramonet, M., Schmidt, M., Steele, L. P., Tyler, S. C., and White,  
849 J.: Contribution of anthropogenic and natural sources to atmospheric methane  
850 variability, *Nature*, 443, 439-443, 2006.

851 [Bousquet, P., Ringeval, B., Pison, I., Dlugokencky, E. J., Brunke, E. G., Carouge, C.,  
852 Chevallier, F., Fortems-Cheiney, A., Frankenberg, C., Hauglustaine, D. A.,  
853 Krummel, P. B., Langenfelds, R. L., Ramonet, M., Schmidt, M., Steele, L. P., Szopa,  
854 S., Yver, C., Viovy, N., and Ciais, P.: Source attribution of the changes in  
855 atmospheric methane for 2006–2008, \*Atmos. Chem. Phys.\*, 11, 3689-3700, 2011.](#)

856 Brown, J., Ferrians, O. J., Jr., Heginbottom, J. A., and Melnikov, E. S.: Circum-arctic  
857 map of permafrost and ground ice conditions, edited by National Snow and Ice  
858 Data Center, Boulder, CO, USA, available at:

859 [http://nsidc.org/data/docs/fgdc/ggd318\\_map\\_circumarctic/](http://nsidc.org/data/docs/fgdc/ggd318_map_circumarctic/) (last access:  
860 December 2014), 2001.

861 Buytaert, W.: TOPMODEL, available at: [http://cran.r-](http://cran.r-project.org/web/packages/topmodel/index.html)  
862 [project.org/web/packages/topmodel/index.html](http://cran.r-project.org/web/packages/topmodel/index.html), (last access: February 2015)  
863 2015.

864 Chang, R. Y.-W., Miller, C. E., Dinardo, S. J., Karion, A., Sweeney, C., Daube, B. C.,  
865 Henderson, J. M., Mountain, M. E., Eluszkiewicz, J., Miller, J. B., Bruhwiler, L. M. P.,  
866 and Wofsy, S. C.: Methane emissions from Alaska in 2012 from CARVE airborne  
867 observations, *Proceedings of the National Academy of Sciences*, doi:  
868 10.1073/pnas.1412953111, 2014. 2014.

869 [Chen, X., Bohn, T. J., and Lettenmaier, D. P.: Model estimates of climate controls](#)  
870 [on pan-Arctic wetland methane emissions, \*Biogeosciences\*, 12, 6259-6277, 2015.](#)

871 [Glagolev, M., Kleptsova, I., Filippov, I., Maksyutov, S., and Machida, T.: Regional](#)  
872 [methane emission from West Siberia mire landscapes, \*Environmental Research\*](#)  
873 [Letters, 6, 045214, 2011.](#)

874 Collins, W. J., Bellouin, N., Doutriaux-Boucher, M., Gedney, N., Halloran, P., Hinton,  
875 T., Hughes, J., Jones, C. D., Joshi, M., Liddicoat, S., Martin, G., O'Connor, F., Rae, J.,  
876 Senior, C., Sitch, S., Totterdell, I., Wiltshire, A., and Woodward, S.: Development  
877 and evaluation of an Earth-System model – HadGEM2, *Geosci. Model Dev.*, 4,  
878 1051-1075, 2011.

879 Cosby, B. J., Hornberger, G. M., Clapp, R. B., and Ginn, T. R.: A Statistical  
880 Exploration of the Relationships of Soil Moisture Characteristics to the Physical  
881 Properties of Soils, *Water Resources Research*, 20, 682-690, 1984.

882 Danielson, J. J. and Gesch, D. B.: Global Multi-resolution Terrain Elevation Data  
883 2010, available at [http://topotools.cr.usgs.gov/gmted\\_viewer/](http://topotools.cr.usgs.gov/gmted_viewer/) (last access:  
884 September 2014), 2011.

885 de Rosnay, P., Drusch, M., Vasiljevic, D., Balsamo, G., Albergel, C., and Isaksen, L.: A  
886 simplified Extended Kalman Filter for the global operational soil moisture  
887 analysis at ECMWF, *Quarterly Journal of the Royal Meteorological Society*, 139,  
888 1199-1213, 2013.

889 Dorigo, W. A., Wagner, W., Hohensinn, R., Hahn, S., Paulik, C., Xaver, A., Gruber, A.,  
890 Drusch, M., Mecklenburg, S., van Oevelen, P., Robock, A., and Jackson, T.: The  
891 International Soil Moisture Network: a data hosting facility for global in situ soil  
892 moisture measurements, *Hydrol. Earth Syst. Sci.*, 15, 1675-1698, 2011.

893 Ducharne, A.: Reducing scale dependence in TOPMODEL using a dimensionless  
894 topographic index, *Hydrol. Earth Syst. Sci.*, 13, 2399-2412, 2009.

895 Durand, M., Lee-Lueng, F., Lettenmaier, D. P., Alsdorf, D. E., Rodriguez, E., and  
896 Esteban-Fernandez, D.: The Surface Water and Ocean Topography Mission:  
897 Observing Terrestrial Surface Water and Oceanic Submesoscale Eddies,  
898 *Proceedings of the IEEE*, 98, 766-779, 2010.

899 Entekhabi, D., Njoku, E. G., O'Neill, P. E., Kellogg, K. H., Crow, W. T., Edelstein, W.  
900 N., Entin, J. K., Goodman, S. D., Jackson, T. J., Johnson, J., Kimball, J., Piepmeier, J. R.,  
901 Koster, R. D., Martin, N., McDonald, K. C., Moghaddam, M., Moran, S., Reichle, R.,  
902 Shi, J. C., Spencer, M. W., Thurman, S. W., Leung, T., and Van Zyl, J.: The Soil  
903 Moisture Active Passive (SMAP) Mission, *Proceedings of the IEEE*, 98, 704-716,  
904 2010.

905 Fan, Y., Li, H., and Miguez-Macho, G.: Global Patterns of Groundwater Table  
906 Depth, *Science*, 339, 940-943, 2013.

907 Fan, Y. and Miguez-Macho, G.: A simple hydrologic framework for simulating  
908 wetlands in climate and earth system models, *Clim Dyn*, 37, 253-278, 2011.

909 Fisher, R. E., Sriskantharajah, S., Lowry, D., Lanoisellé, M., Fowler, C. M. R., James,  
910 R. H., Hermansen, O., Lund Myhre, C., Stohl, A., Greinert, J., Nisbet-Jones, P. B. R.,  
911 Mienert, J., and Nisbet, E. G.: Arctic methane sources: Isotopic evidence for  
912 atmospheric inputs, *Geophysical Research Letters*, 38, L21803, 2011.

913 Fluet-Chouinard, E., Lehner, B., Rebelo, L.-M., Papa, F., and Hamilton, S. K.:  
914 Development of a global inundation map at high spatial resolution from  
915 topographic downscaling of coarse-scale remote sensing data, *Remote Sensing of*  
916 *Environment*, 158, 348-361, 2015.

917 Gedney, N. and Cox, P. M.: The Sensitivity of Global Climate Model Simulations to  
918 the Representation of Soil Moisture Heterogeneity, *Journal of Hydrometeorology*,  
919 4, 1265-1275, 2003.

920 Gerten, D., Schaphoff, S., Haberlandt, U., Lucht, W., and Sitch, S.: Terrestrial  
921 vegetation and water balance—hydrological evaluation of a dynamic global  
922 vegetation model, *Journal of Hydrology*, 286, 249-270, 2004.

923 Glagolev, M., Kleptsova, I., Filippov, I., Maksyutov, S., and Machida, T.: Regional  
924 methane emission from West Siberia mire landscapes, *Environmental Research*  
925 *Letters*, 6, 045214, 2011.

926 Grabs, T., Seibert, J., Bishop, K., and Laudon, H.: Modeling spatial patterns of  
927 saturated areas: A comparison of the topographic wetness index and a dynamic  
928 distributed model, *Journal of Hydrology*, 373, 15-23, 2009.

929 Gurney, K. R., Law, R. M., Denning, A. S., Rayner, P. J., Baker, D., Bousquet, P.,  
930 Bruhwiler, L., Chen, Y.-H., Ciais, P., Fan, S., Fung, I. Y., Gloor, M., Heimann, M.,  
931 Higuchi, K. A. Z., John, J., Kowalczyk, E. V. A., Maki, T., Maksyutov, S., Peylin, P.,  
932 Prather, M., Pak, B. C., Sarmiento, J., Taguchi, S., Takahashi, T., and Yuen, C.-W.:  
933 TransCom 3 CO<sub>2</sub> inversion intercomparison: 1. Annual mean control results and  
934 sensitivity to transport and prior flux information, *Tellus B*, 55, 555-579, 2003.

935 Güntner, A., Seibert, J., and Uhlenbrook, S.: Modeling spatial patterns of saturated  
936 areas: An evaluation of different terrain indices, *Water Resources Research*, 40,  
937 W05114, 2004.

938 Harris, I., Jones, P. D., Osborn, T. J., and Lister, D. H.: Updated high-resolution  
939 grids of monthly climatic observations – the CRU TS3.10 Dataset, *International*  
940 *Journal of Climatology*, 34, 623-642, 2014.

941 Hess, L., Melack, J., Affonso, A., Barbosa, C., Gastil-Buhl, M., and Novo, E. L. M.:  
942 Wetlands of the Lowland Amazon Basin: Extent, Vegetative Cover, and Dual-  
943 season Inundated Area as Mapped with JERS-1 Synthetic Aperture Radar,  
944 *Wetlands*, doi: 10.1007/s13157-015-0666-y, 2015. 1-12, 2015.

945 Hodson, E. L., Poulter, B., Zimmermann, N. E., Prigent, C., and Kaplan, J. O.: The El  
946 Niño–Southern Oscillation and wetland methane interannual variability,  
947 *Geophysical Research Letters*, 38, L08810, 2011.

948 IPCC: *Climate Change 2013: The Physical Science Basis. Contribution of Working*  
949 *Group I to the Fifth Assessment Report of the Intergovernmental Panel on*  
950 *Climate Change*, Cambridge University Press, Cambridge, United Kingdom and  
951 New York, NY, USA, 2013.

952 Ito, A. and Inatomi, M.: Use of a process-based model for assessing the methane  
953 budgets of global terrestrial ecosystems and evaluation of uncertainty,  
954 *Biogeosciences*, 9, 759-773, 2012.

955 Jennifer, D. W., John, S. K., Annett, B., and Kyle, C. M.: Surface water inundation in  
956 the boreal-Arctic: potential impacts on regional methane emissions,  
957 Environmental Research Letters, 9, 075001, 2014.

958 Junk, W., Piedade, M., Schöngart, J., Cohn-Haft, M., Adeney, J. M., and Wittmann, F.:  
959 A Classification of Major Naturally-Occurring Amazonian Lowland Wetlands,  
960 Wetlands, 31, 623-640, 2011.

961 Kaplan, J. O.: Wetlands at the Last Glacial Maximum: Distribution and methane  
962 emissions, Geophysical Research Letters, 29, 3-1-3-4, 2002.

963 Kerr, Y. H., Waldteufel, P., Wigneron, J. P., Delwart, S., Cabot, F., Boutin, J.,  
964 Escorihuela, M. J., Font, J., Reul, N., Gruhier, C., Juglea, S. E., Drinkwater, M. R.,  
965 Hahne, A., Martin-Neira, M., and Mecklenburg, S.: The SMOS Mission: New Tool  
966 for Monitoring Key Elements of the Global Water Cycle, Proceedings of the IEEE,  
967 98, 666-687, 2010.

968 Kim, Y., Kimball, J. S., Zhang, K., and McDonald, K. C.: Satellite detection of  
969 increasing Northern Hemisphere non-frozen seasons from 1979 to 2008:  
970 Implications for regional vegetation growth, Remote Sensing of Environment,  
971 121, 472-487, 2012.

972 Kirschke, S., Bousquet, P., Ciais, P., Saunois, M., Canadell, J. G., Dlugokencky, E. J.,  
973 Bergamaschi, P., Bergmann, D., Blake, D. R., Bruhwiler, L., Cameron-Smith, P.,  
974 Castaldi, S., Chevallier, F., Feng, L., Fraser, A., Heimann, M., Hodson, E. L.,  
975 Houweling, S., Josse, B., Fraser, P. J., Krummel, P. B., Lamarque, J.-F., Langenfelds,  
976 R. L., Le Quere, C., Naik, V., O'Doherty, S., Palmer, P. I., Pison, I., Plummer, D.,  
977 Poulter, B., Prinn, R. G., Rigby, M., Ringeval, B., Santini, M., Schmidt, M., Shindell,  
978 D. T., Simpson, I. J., Spahni, R., Steele, L. P., Strode, S. A., Sudo, K., Szopa, S., van der  
979 Werf, G. R., Voulgarakis, A., van Weele, M., Weiss, R. F., Williams, J. E., and Zeng,  
980 G.: Three decades of global methane sources and sinks, Nature Geosci, 6, 813-  
981 823, 2013.

982 Kleinen, T., Brovkin, V., and Schuldt, R. J.: A dynamic model of wetland extent and  
983 peat accumulation: results for the Holocene, Biogeosciences, 9, 235-248, 2012.

984 Kopecký, M. and Čížková, Š.: Using topographic wetness index in vegetation  
985 ecology: does the algorithm matter?, Applied Vegetation Science, 13, 450-459,  
986 2010.

987 Kopecký, M. and Čížková, Š.: Using topographic wetness index in vegetation  
988 ecology: does the algorithm matter?, Applied Vegetation Science, 13, 450-459,  
989 2010.

990 Leff, B., Ramankutty, N., and Foley, J. A.: Geographic distribution of major crops  
991 across the world, Global Biogeochemical Cycles, 18, GB1009, 2004.

992 Lehner, B. and Döll, P.: Development and validation of a global database of lakes,  
993 reservoirs and wetlands, Journal of Hydrology, 296, 1-22, 2004.

994 Lehner, B. and Grill, G.: Global river hydrography and network routing: baseline  
995 data and new approaches to study the world's large river systems, Hydrological  
996 Processes, 27, 2171-2186, 2013.

997 Lehner, B., Verdin, K., and Jarvis, A.: New Global Hydrography Derived From  
998 Spaceborne Elevation Data, Eos, Transactions American Geophysical Union, 89,  
999 93-94, 2008.

1000 Lei, H., Huang, M., Leung, L. R., Yang, D., Shi, X., Mao, J., Hayes, D. J., Schwalm, C. R.,  
1001 Wei, Y., and Liu, S.: Sensitivity of global terrestrial gross primary production to  
1002 hydrologic states simulated by the Community Land Model using two runoff

1003 parameterizations, *Journal of Advances in Modeling Earth Systems*, 6, 658-679,  
1004 2014.

1005 Liang, X., Lettenmaier, D. P., Wood, E. F., and Burges, S. J.: A simple hydrologically  
1006 based model of land surface water and energy fluxes for general circulation  
1007 models, *Journal of Geophysical Research: Atmospheres*, 99, 14415-14428, 1994.

1008 Lin, K., Zhang, Q., and Chen, X.: An evaluation of impacts of DEM resolution and  
1009 parameter correlation on TOPMODEL modeling uncertainty, *Journal of*  
1010 *Hydrology*, 394, 370-383, 2010.

1011 Lin, S., Jing, C., Coles, N., Chaplot, V., Moore, N., and Wu, J.: Evaluating DEM source  
1012 and resolution uncertainties in the Soil and Water Assessment Tool, *Stoch*  
1013 *Environ Res Risk Assess*, 27, 209-221, 2013.

1014 Marthews, T. R., Dadson, S. J., Lehner, B., Abele, S., and Gedney, N.: High-  
1015 resolution global topographic index values for use in large-scale hydrological  
1016 modelling, *Hydrol. Earth Syst. Sci.*, 19, 91-104, 2015.

1017 Marthews, T. R., Quesada, C. A., Galbraith, D. R., Malhi, Y., Mullins, C. E., Hodnett,  
1018 M. G., and Dharssi, I.: High-resolution hydraulic parameter maps for surface soils  
1019 in tropical South America, *Geosci. Model Dev.*, 7, 711-723, 2014.

1020 Matthews, E. and Fung, I.: Methane emission from natural wetlands: Global  
1021 distribution, area, and environmental characteristics of sources, *Global*  
1022 *Biogeochemical Cycles*, 1, 61-86, 1987.

1023 [Melack, J. M., Hess, L. L., Gastil, M., Forsberg, B. R., Hamilton, S. K., Lima, I. B. T.,  
1024 and Novo, E. M. L. M.: Regionalization of methane emissions in the Amazon Basin  
1025 with microwave remote sensing, \*Global Change Biology\*, 10, 530-544, 2004.](#)

1026 Melton, J. R., Wania, R., Hodson, E. L., Poulter, B., Ringeval, B., Spahni, R., Bohn, T.,  
1027 Avis, C. A., Beerling, D. J., Chen, G., Eliseev, A. V., Denisov, S. N., Hopcroft, P. O.,  
1028 Lettenmaier, D. P., Riley, W. J., Singarayer, J. S., Subin, Z. M., Tian, H., Zürcher, S.,  
1029 Brovkin, V., van Bodegom, P. M., Kleinen, T., Yu, Z. C., and Kaplan, J. O.: Present  
1030 state of global wetland extent and wetland methane modelling: conclusions from  
1031 a model inter-comparison project (WETCHIMP), *Biogeosciences*, 10, 753-788,  
1032 2013.

1033 Meng, L., Hess, P. G. M., Mahowald, N. M., Yavitt, J. B., Riley, W. J., Subin, Z. M.,  
1034 Lawrence, D. M., Swenson, S. C., Jauhiainen, J., and Fuka, D. R.: Sensitivity of  
1035 wetland methane emissions to model assumptions: application and model  
1036 testing against site observations, *Biogeosciences*, 9, 2793-2819, 2012.

1037 Mulligan, M. and Wainwright, J.: *Modelling and Model Building*, in:  
1038 *Environmental Modelling*, John Wiley & Sons, Ltd, Chichester, UK, 7-26, 2013.

1039 Nachtergaele, F., Van Velthuisen, H., Verelst, L., Batjes, N., Dijkshoorn, K., Van  
1040 Engelen, V., Fischer, G., Jones, A., Montanarella, L., and Petri, M.: *Harmonized*  
1041 *world soil database*, FAO, Rome, Italy and IIASA, Laxenburg, Austria, 2008.

1042 Nisbet, E. G., Dlugokencky, E. J., and Bousquet, P.: Methane on the Rise—Again,  
1043 *Science*, 343, 493-495, 2014.

1044 National Wetland Working Group: *Wetlands of Canada. Ecological Land*  
1045 *Classification Series*, No., 24. Canada Committee on Ecological Land  
1046 *Classification*, Sustainable Development Branch, Environment Canada and  
1047 *Polyscience Publications Inc*, Montreal, Quebec, Canada.

1048 Niu, G.-Y., Yang, Z.-L., Dickinson, R. E., and Gulden, L. E.: A simple TOPMODEL-  
1049 based runoff parameterization (SIMTOP) for use in global climate models,  
1050 *Journal of Geophysical Research: Atmospheres*, 110, D21106, 2005.



1051 Pan, F., Peters-Lidard, C. D., Sale, M. J., and King, A. W.: A comparison of  
1052 geographical information systems-based algorithms for computing the  
1053 TOPMODEL topographic index, *Water Resources Research*, 40, W06303, 2004.

1054 Papa, F., Prigent, C., Aires, F., Jimenez, C., Rossow, W. B., and Matthews, E.:  
1055 Interannual variability of surface water extent at the global scale, 1993–2004,  
1056 *Journal of Geophysical Research: Atmospheres*, 115, D12111, 2010.

1057 Peregon, A., Maksyutov, S., Kosykh, N. P., and Mironycheva-Tokareva, N. P.: Map-  
1058 based inventory of wetland biomass and net primary production in western  
1059 Siberia, *Journal of Geophysical Research: Biogeosciences*, 113, G01007, 2008.

1060 Petrescu, A. M. R., Lohila, A., Tuovinen, J.-P., Baldocchi, D. D., Desai, A. R., Roulet,  
1061 N. T., Vesala, T., Dolman, A. J., Oechel, W. C., Marcolla, B., Friborg, T., Rinne, J.,  
1062 Matthes, J. H., Merbold, L., Meijide, A., Kiely, G., Sottocornola, M., Sachs, T., Zona,  
1063 D., Varlagin, A., Lai, D. Y. F., Veenendaal, E., Parmentier, F.-J. W., Skiba, U., Lund,  
1064 M., Hensen, A., van Huissteden, J., Flanagan, L. B., Shurpali, N. J., Grünwald, T.,  
1065 Humphreys, E. R., Jackowicz-Korczyński, M., Aurela, M. A., Laurila, T., Grüning, C.,  
1066 Corradi, C. A. R., Schrier-Uijl, A. P., Christensen, T. R., Tamstorf, M. P., Mastepanov,  
1067 M., Martikainen, P. J., Verma, S. B., Bernhofer, C., and Cescatti, A.: The uncertain  
1068 climate footprint of wetlands under human pressure, *Proceedings of the National  
1069 Academy of Sciences*, 112, 4594-4599, 2015.

1070 Petrescu, A. M. R., van Beek, L. P. H., van Huissteden, J., Prigent, C., Sachs, T.,  
1071 Corradi, C. A. R., Parmentier, F. J. W., and Dolman, A. J.: Modeling regional to  
1072 global CH<sub>4</sub> emissions of boreal and arctic wetlands, *Global Biogeochemical  
1073 Cycles*, 24, GB4009, 2010.

1074 Poulter, B., Ciais, P., Hodson, E., Lischke, H., Maignan, F., Plummer, S., and  
1075 Zimmermann, N. E.: Plant functional type mapping for earth system models,  
1076 *Geosci. Model Dev.*, 4, 993-1010, 2011.

1077 Poulter, B., Bousquet, P., Canadell, P., Ciais, P., Peregon, A., Arora, V., Beerling, D.,  
1078 Brovkin, V., Hopcroft, P., Jones, C., Joos, F., Gedney, N., Ito, A., Kleinen, T., Koven,  
1079 C., MacDonald, K., Melton, J., Peng, C., Peng, S., Schroder, R., Prigent, C., Riley, B.,  
1080 Saito, M., Spahni, R., Tian, H., Taylor, L., Viovy, N., Wilton, D., Wiltshire, A., Xu, X.,  
1081 and Zhang, Z.: Global wetland contribution to increasing atmospheric methane  
1082 concentrations (2000-2012), in preparation, 2015.

1083 Prigent, C., Papa, F., Aires, F., Jimenez, C., Rossow, W. B., and Matthews, E.:  
1084 Changes in land surface water dynamics since the 1990s and relation to  
1085 population pressure, *Geophysical Research Letters*, 39, L08403, 2012.

1086 Prigent, C., Papa, F., Aires, F., Rossow, W. B., and Matthews, E.: Global inundation  
1087 dynamics inferred from multiple satellite observations, 1993–2000, *Journal of  
1088 Geophysical Research: Atmospheres*, 112, D12107, 2007.

1089 Quinn, P. F., Beven, K. J., and Lamb, R.: The  $\ln(a/\tan\beta)$  index: How to calculate it  
1090 and how to use it within the topmodel framework, *Hydrological Processes*, 9,  
1091 161-182, 1995.

1092 Quiquet, A., Archibald, A. T., Friend, A. D., Chappellaz, J., Levine, J. G., Stone, E. J.,  
1093 Telford, P. J., and Pyle, J. A.: The relative importance of methane sources and  
1094 sinks over the Last Interglacial period and into the last glaciation, *Quaternary  
1095 Science Reviews*, 112, 1-16, 2015.

1096 Ringeval, B., Decharme, B., Piao, S. L., Ciais, P., Papa, F., de Noblet-Ducoudré, N.,  
1097 Prigent, C., Friedlingstein, P., Gouttevin, I., Koven, C., and Ducharne, A.: Modelling  
1098 sub-grid wetland in the ORCHIDEE global land surface model: evaluation against  
1099 river discharges and remotely sensed data, *Geosci. Model Dev.*, 5, 941-962, 2012.

1100 Ringeval, B., Houweling, S., van Bodegom, P. M., Spahni, R., van Beek, R., Joos, F.,  
1101 and Röckmann, T.: Methane emissions from floodplains in the Amazon Basin:  
1102 challenges in developing a process-based model for global applications,  
1103 *Biogeosciences*, 11, 1519-1558, 2014.

1104 Rinne, J., Riutta, T., Pihlatie, M., Aurela, M., Haapanala, S., Tuovinen, J.-P., Tuittila,  
1105 E.-S., and Vesala, T.: Annual cycle of methane emission from a boreal fen  
1106 measured by the eddy covariance technique, *Tellus B*, 59, 449-457, 2007.

1107 Seneviratne, S. I., Corti, T., Davin, E. L., Hirschi, M., Jaeger, E. B., Lehner, I.,  
1108 Orlowsky, B., and Teuling, A. J.: Investigating soil moisture–climate interactions  
1109 in a changing climate: A review, *Earth-Science Reviews*, 99, 125-161, 2010.

1110 Sheng, Y., Smith, L. C., MacDonald, G. M., Kremenetski, K. V., Frey, K. E., Velichko,  
1111 A. A., Lee, M., Beilman, D. W., and Dubinin, P.: A high-resolution GIS-based  
1112 inventory of the west Siberian peat carbon pool, *Global Biogeochemical Cycles*,  
1113 18, GB3004, 2004.

1114 Singarayer, J. S., Valdes, P. J., Friedlingstein, P., Nelson, S., and Beerling, D. J.: Late  
1115 Holocene methane rise caused by orbitally controlled increase in tropical  
1116 sources, *Nature*, 470, 82-85, 2011.

1117 Sitch, S., Smith, B., Prentice, I. C., Arneth, A., Bondeau, A., Cramer, W., Kaplan, J. O.,  
1118 Levis, S., Lucht, W., Sykes, M. T., Thonicke, K., and Venevsky, S.: Evaluation of  
1119 ecosystem dynamics, plant geography and terrestrial carbon cycling in the LPJ  
1120 dynamic global vegetation model, *Global Change Biology*, 9, 161-185, 2003.

1121 Sivapalan, M., Beven, K., and Wood, E. F.: On hydrologic similarity: 2. A scaled  
1122 model of storm runoff production, *Water Resources Research*, 23, 2266-2278,  
1123 1987.

1124 Sørensen, R. and Seibert, J.: Effects of DEM resolution on the calculation of  
1125 topographical indices: TWI and its components, *Journal of Hydrology*, 347, 79-  
1126 89, 2007.

1127 Spahni, R., Wania, R., Neef, L., van Weele, M., Pison, I., Bousquet, P., Frankenberg,  
1128 C., Foster, P. N., Joos, F., Prentice, I. C., and van Velthoven, P.: Constraining global  
1129 methane emissions and uptake by ecosystems, *Biogeosciences*, 8, 1643-1665,  
1130 2011.

1131 Stieglitz, M., Rind, D., Famiglietti, J., and Rosenzweig, C.: An Efficient Approach to  
1132 Modeling the Topographic Control of Surface Hydrology for Regional and Global  
1133 Climate Modeling, *Journal of Climate*, 10, 118-137, 1997.

1134 Stocker, B. D., Roth, R., Joos, F., Spahni, R., Steinacher, M., Zaehle, S., Bouwman, L.,  
1135 Xu, R., and Prentice, I. C.: Multiple greenhouse-gas feedbacks from the land  
1136 biosphere under future climate change scenarios, *Nature Clim. Change*, 3, 666-  
1137 672, 2013.

1138 Stocker, B. D., Spahni, R., and Joos, F.: DYPTOP: a cost-efficient TOPMODEL  
1139 implementation to simulate sub-grid spatio-temporal dynamics of global  
1140 wetlands and peatlands, *Geosci. Model Dev.*, 7, 3089-3110, 2014.

1141 Tarnocai, C., Canadell, J. G., Schuur, E. A. G., Kuhry, P., Mazhitova, G., and Zimov, S.:  
1142 Soil organic carbon pools in the northern circumpolar permafrost region, *Global  
1143 Biogeochemical Cycles*, 23, GB2023, 2009.

1144 Turetsky, M. R., Kotowska, A., Bubier, J., Dise, N. B., Crill, P., Hornibrook, E. R. C.,  
1145 Minkinen, K., Moore, T. R., Myers-Smith, I. H., Nykänen, H., Olefeldt, D., Rinne, J.,  
1146 Saarnio, S., Shurpali, N., Tuittila, E.-S., Waddington, J. M., White, J. R., Wickland, K.  
1147 P., and Wilmking, M.: A synthesis of methane emissions from 71 northern,

1148 temperate, and subtropical wetlands, *Global Change Biology*, 20, 2183-2197,  
1149 2014.

1150 USGS: HYDRO1k Elevation Derivative Database, US Geological Survey Earth  
1151 Resources Observation and Science (EROS) Data Center (EDC), Sioux Falls, South  
1152 Dakota, USA, 1996.

1153 Verpoorter, C., Kutser, T., Seekell, D. A., and Tranvik, L. J.: A global inventory of  
1154 lakes based on high-resolution satellite imagery, *Geophysical Research Letters*,  
1155 41, 6396-6402, 2014.

1156 Wania, R., Melton, J. R., Hodson, E. L., Poulter, B., Ringeval, B., Spahni, R., Bohn, T.,  
1157 Avis, C. A., Chen, G., Eliseev, A. V., Hopcroft, P. O., Riley, W. J., Subin, Z. M., Tian, H.,  
1158 van Bodegom, P. M., Kleinen, T., Yu, Z. C., Singarayer, J. S., Zürcher, S.,  
1159 Lettenmaier, D. P., Beerling, D. J., Denisov, S. N., Prigent, C., Papa, F., and Kaplan, J.  
1160 O.: Present state of global wetland extent and wetland methane modelling:  
1161 methodology of a model inter-comparison project (WETCHIMP), *Geosci. Model*  
1162 *Dev.*, 6, 617-641, 2013.

1163 Wania, R., Ross, I., and Prentice, I. C.: Integrating peatlands and permafrost into a  
1164 dynamic global vegetation model: 1. Evaluation and sensitivity of physical land  
1165 surface processes, *Global Biogeochemical Cycles*, 23, GB3014, 2009.

1166 Ward, R. C. and Robinson, M.: *Principles of Hydrology*, 4th Edn, McGraw-Hill,  
1167 Maidenhead, UK, 2000.

1168 Whalen, S. C. and Reeburgh, W. S.: Interannual variations in tundra methane  
1169 emission: A 4-year time series at fixed sites, *Global Biogeochemical Cycles*, 6,  
1170 139-159, 1992.

1171 Wilson, J. P. and Gallant, J. C.: *Terrain analysis: principles and applications*, John  
1172 Wiley & Sons, New York, NY, USA, 2000.

1173 Wood, E. F., Roundy, J. K., Troy, T. J., van Beek, L. P. H., Bierkens, M. F. P., Blyth, E.,  
1174 de Roo, A., Döll, P., Ek, M., Famiglietti, J., Gochis, D., van de Giesen, N., Houser, P.,  
1175 Jaffé, P. R., Kollet, S., Lehner, B., Lettenmaier, D. P., Peters-Lidard, C., Sivapalan, M.,  
1176 Sheffield, J., Wade, A., and Whitehead, P.: Hyperresolution global land surface  
1177 modeling: Meeting a grand challenge for monitoring Earth's terrestrial water,  
1178 *Water Resources Research*, 47, W05301, 2011.

1179 Woodward, C., Shulmeister, J., Larsen, J., Jacobsen, G. E., and Zawadzki, A.: The  
1180 hydrological legacy of deforestation on global wetlands, *Science*, 346, 844-847,  
1181 2014.

1182 Zhang, Z., Jiang, H., Liu, J., Ju, W., and Zhang, X.: Effect of heterogeneous  
1183 atmospheric CO<sub>2</sub> on simulated global carbon budget, *Global and Planetary*  
1184 *Change*, 101, 33-51, 2013.

1185 Zhu, X., Zhuang, Q., Lu, X., and Song, L.: Spatial scale-dependent land-atmospheric  
1186 methane exchanges in the northern high latitudes from 1993 to 2004,  
1187 *Biogeosciences*, 11, 1693-1704, 2014.

1188 [Zhuang, Q., Melillo, J. M., Kicklighter, D. W., Prinn, R. G., McGuire, A. D., Steudler, P.  
1189 A., Felzer, B. S., and Hu, S.: Methane fluxes between terrestrial ecosystems and  
1190 the atmosphere at northern high latitudes during the past century: A  
1191 retrospective analysis with a process-based biogeochemistry model, \*Global\*  
1192 \*Biogeochemical Cycles\*, 18, GB3010, 2004.](#)

1193 Zona, D., Oechel, W. C., Kochendorfer, J., Paw U, K. T., Salyuk, A. N., Olivas, P. C.,  
1194 Oberbauer, S. F., and Lipson, D. A.: Methane fluxes during the initiation of a large-  
1195 scale water table manipulation experiment in the Alaskan Arctic tundra, *Global*  
1196 *Biogeochemical Cycles*, 23, GB2013, 2009.

1197 Zürcher, S., Spahni, R., Joos, F., Steinacher, M., and Fischer, H.: Impact of an abrupt  
1198 cooling event on interglacial methane emissions in northern peatlands,  
1199 *Biogeosciences*, 10, 1963-1981, 2013.

## Tables

Table 1. Soil parameters for LPJ-wsl soil classes.  $f$  is a parameter describing the exponential decline of transmissivity with depth for each soil type.

Soil type	$f$	Mineral Content (%)	Organic Content (%)	Wilting Point (%)	Porosity (%)
Clay Heavy	3.2	0.508	0.01	0.138	0.138
Silty Clay	3.1	0.531	0.01	0.126	0.468
Clay	2.8	0.531	0.01	0.138	0.468
Silty Clay Loam	2.9	0.534	0.01	0.120	0.464
Clay Loam	2.7	0.595	0.01	0.103	0.465
Silt	3.4	0.593	0.01	0.084	0.476
Silt Loam	2.6	0.593	0.01	0.084	0.476
Sandy Clay	2.5	0.535	0.01	0.100	0.406
Loam	2.5	0.535	0.01	0.066	0.439
Sandy Clay Loam	2.4	0.565	0.01	0.067	0.404
Sandy Loam	2.3	0.565	0.01	0.047	0.434
Loamy Sand	2.2	0.578	0.01	0.028	0.421
Sand	2.1	0.578	0.01	0.010	0.339
Organic	2.5	0.01	0.20	0.066	0.439

Table 2 Model experiments for different parameterization schemes and corresponding DEM products applied in this study.

Model Experiment	DEM	DEM source	Resolution (arc seconds)	Coverage	River Basin	Aggregation type	Hydro-corrected
HYDRO1k_BASIN	Hydro1k	GTOPO30	30	Global*	HYDRO1K	Catchment	Yes
HYDRO1k_GRID	Hydro1k	GTOPO30	30	Global*	HYDRO1K	Grid	Yes
GMTED_BASIN	GMTED	SRTM&others	15	Global	HYDRO1K	Catchment	No
GMTED_GRID	GMTED	SRTM&others	15	Global	HYDRO1K	Grid	No
SHEDS_BASIN	HydroSHEDS	SRTM	15	<60°N	HydroSHEDS	Catchment	Yes
SHEDS_GRID	HydroSHEDS	SRTM	15	<60°N	HydroSHEDS	Grid	Yes

Table 3 Summary of simulated and observed mean annual minimum (MIN), maximum (MAX), and amplitude (AMP) of wetland extent for 1980-2010. All units are Mkm<sup>2</sup> (106 km<sup>2</sup>) ± 1σ, where standard deviation represents the inter-annual variation in model estimates except for the row Average, which represents uncertainties of estimates from each model experiment.

Model	Lowland Amazon Basin			West Siberian Lowland			Global		
	MIN	MAX	AMP	MIN	MAX	AMP	MIN	MAX	AMP
SHEDS_BASIN	0.27±0.02	0.38±0.01	0.11±0.01	0±0	0.45±0.05	0.45±0.05	2.96±0.06	5.17±0.11	2.23±0.10
SHEDS_GRID	0.32±0.01	0.40±0.01	0.08±0.01	0±0	0.45±0.05	0.45±0.05	3.56±0.06	5.93±0.11	2.38±0.10
GMTED_BASIN	0.21±0.02	0.35±0.01	0.14±0.02	0±0	0.39±0.06	0.39±0.06	2.09±0.05	3.75±0.12	1.66±0.12
GMTED_GRID	0.19±0.02	0.34±0.01	0.15±0.02	0±0	0.38±0.06	0.38±0.06	1.80±0.05	3.32±0.13	1.52±0.13
HYDRO1k_BASIN	0.25±0.02	0.37±0.01	0.12±0.01	0±0	0.39±0.06	0.39±0.06	2.44±0.05	4.32±0.11	1.89±0.11
HYDRO1k_GRID	0.22±0.02	0.36±0.01	0.14±0.02	0±0	0.36±0.07	0.36±0.07	2.12±0.05	3.73±0.13	1.61±0.13
<b>Average</b>	<b>0.27±0.04</b>	<b>0.38±0.02</b>	<b>0.11±0.01</b>	<b>0±0</b>	<b>0.40±0.04</b>	<b>0.40±0.04</b>	<b>2.49±0.65</b>	<b>4.37±0.99</b>	<b>1.88±0.35</b>
Observations									
Hess2015	0.23	0.58							
GIEMS	0.12±0.01	0.25±0.03	0.14±0.04	0±0	0.24±0.05	0.25±0.05	1.38±0.09	4.47±0.20	3.09±0.19
SWAMPS-GLWD	0.22±0.03	0.34±0.01	0.12±0.03	0±0	0.50±0.03	0.51±0.03	3.03±0.13	6.62±0.18	3.63±0.14

Table 4 Spearman correlations between satellite-based vs. modeled interannual anomalies of the grid-cells contained in each region defined in Fig. 2f at global scale. Values out and in parentheses are correlation efficient with SWAMPS-GLWD and GIEMS respectively. The two highest value within one column is in bold.

Regions	SHDES BASIN	SHDES GRID	GMTED BASIN	GMTED GRID	HYDRO1K BASIN	HYDRO1k GRID
Boreal North America	<b>0.770</b> ( <b>0.378</b> )	<b>0.768</b> (0.376)	0.751 (0.354)	0.745 (0.341)	0.765 ( <b>0.378</b> )	0.748 (0.343)
Boreal Eurasia	<b>0.785</b> ( <b>0.513</b> )	<b>0.782</b> ( <b>0.511</b> )	0.763 (0.487)	0.764 (0.487)	0.763 (0.493)	0.760 (0.484)
Europe	<b>0.604</b> (0.091)	<b>0.595</b> (0.079)	0.313 (-0.198)	0.211 ( <b>-0.278</b> )	0.588 (0.076)	0.218 ( <b>-0.272</b> )
Tropical South America	0.723 ( <b>0.838</b> )	<b>0.725</b> (0.831)	0.724 (0.835)	0.666 (0.825)	0.708 ( <b>0.836</b> )	<b>0.726</b> (0.835)
South Africa	0.082 ( <b>0.736</b> )	0.044 (0.725)	<b>0.084</b> (0.735)	0.076 (0.734)	0.040 (0.717)	<b>0.088</b> ( <b>0.740</b> )
Tropical Asia	<b>0.689</b> ( <b>0.674</b> )	<b>0.681</b> (0.673)	0.705 ( <b>0.682</b> )	0.677 (0.625)	0.670 (0.660)	0.648 (0.632)
Temperate North America	0.359 (0.139)	0.380 (0.155)	0.406 (0.262)	0.347 (0.229)	<b>0.518</b> ( <b>0.288</b> )	<b>0.479</b> ( <b>0.305</b> )
Temperate South America	<b>-0.193</b> ( <b>0.633</b> )	<b>-0.205</b> (0.597)	-0.153 (0.622)	-0.162 ( <b>0.641</b> )	-0.178 (0.627)	-0.166 (0.627)
Temperate Eurasia	<b>0.742</b> ( <b>0.645</b> )	<b>0.760</b> ( <b>0.660</b> )	0.735 (0.642)	0.721 (0.643)	0.732 (0.642)	0.716 (0.642)

Table 5. List of global and regional wetland CH<sub>4</sub> estimates from our model experiments (see Table 2) over the period 1980-2000. All units are Tg CH<sub>4</sub> yr<sup>-1</sup>±1σ, where standard deviation represents the interannual variation in the model estimates. Note that estimates from some reference studies are not for the same period.

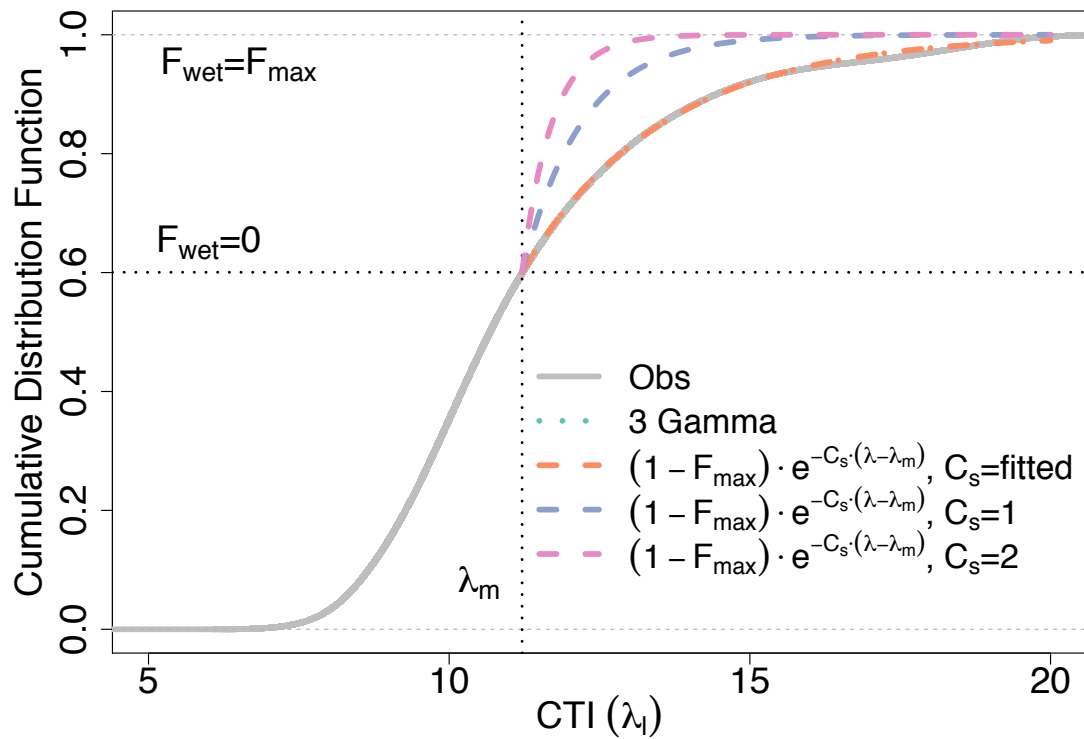
Estimates	Global	Regions				Hotspot		
		Tropics (20N-30S)	Temperate (20-45N, 30S-50S)	Northern (>45N)	Central Amazon <sup>b</sup>	WSL	Hudson Bay	Alaska
SHEDS_BASIN	171.9	109.3±2.3	26.4±1.0	36.1±1.8	10.9±0.3	5.4±0.9	6.5±0.5	1.7±0.3
SHEDS_GRID	193.0	123.7±2.2	31.4±1.0	38.7±1.9	11.4±0.3	5.5±0.9	7.1±0.6	1.5±0.3
GMTED_BASIN	130.1	85.5±2.3	19.0±0.9	26.3±1.4	9.5±0.4	4.5±0.9	4.4±0.6	1.6±0.3
GMTED_GRID	117.2	76.7±2.3	16.4±0.9	24.2±1.4	9.2±0.4	4.1±0.9	4.2±0.6	1.4±0.3
HYDRO1K_BASIN	148.3	96.4±2.3	21.5±0.9	30.3±1.6	10.4±0.3	4.4±0.9	5.8±0.6	1.7±0.3
HYDRO1K_GRID	128.8	85.0±2.3	17.8±0.9	26.0±1.4	10.0±0.4	3.9±0.9	4.8±0.6	1.5±0.3
Melton et al. (2013) <sup>a</sup>	190±39						5.4±3.2	
Zhu et al. (2015)	209-245			38.1-55.4				
Chen et al. (2015)				35			3.11±0.45	
Zhu et al. (2014)				34-58			3.1± 0.5	
Ringeval et al. (2012)	193.8	102	51	40.8				
Glagolev et al. (2011)						3.91±1.3		
Melack et al. (2004)					9.1			
Zhuang et al. (2004)				57.3				
Chang et al. (2014)								2.1±0.5
Bloom et al. (2012)		111.1						
Bousquet et al. (2011)	151±10	91±11						
Bloom et al. (2010)	165±50	91±28					4.9±1.4	

<sup>a</sup> WETCHIMP estimates for 1993-2004

<sup>b</sup> Central Amazon (54-72°W,0-8°S)



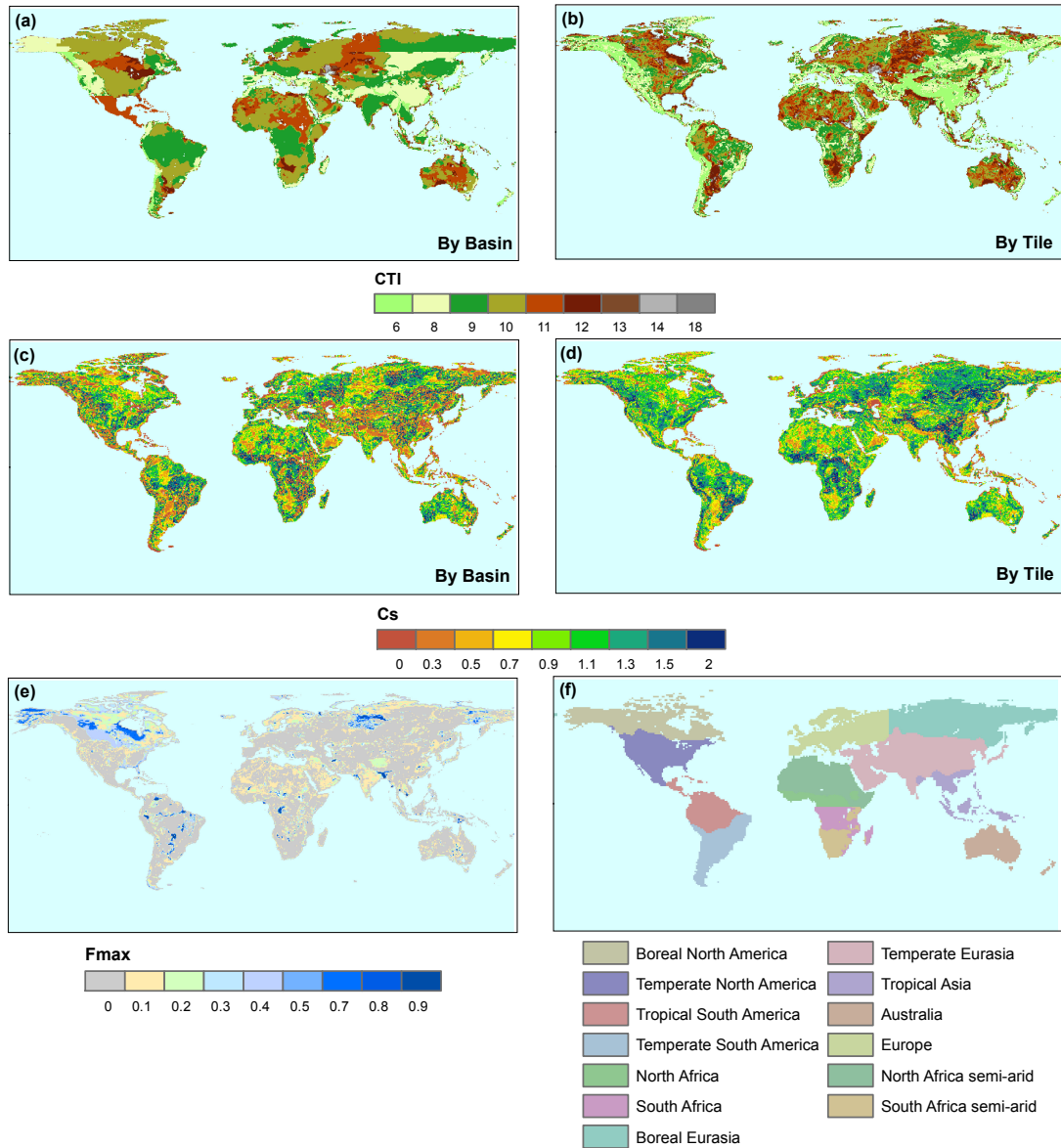
1 **Figures**



2

3 Figure 1. Cumulative distribution function (CDF) of the fitted exponential curve  
 4 (blue line) as a function of compound topographic index (CTI) in comparison  
 5 with the three-parameter gamma function (red line), as well as the observations  
 6 (grey line) with in a sample grid box.

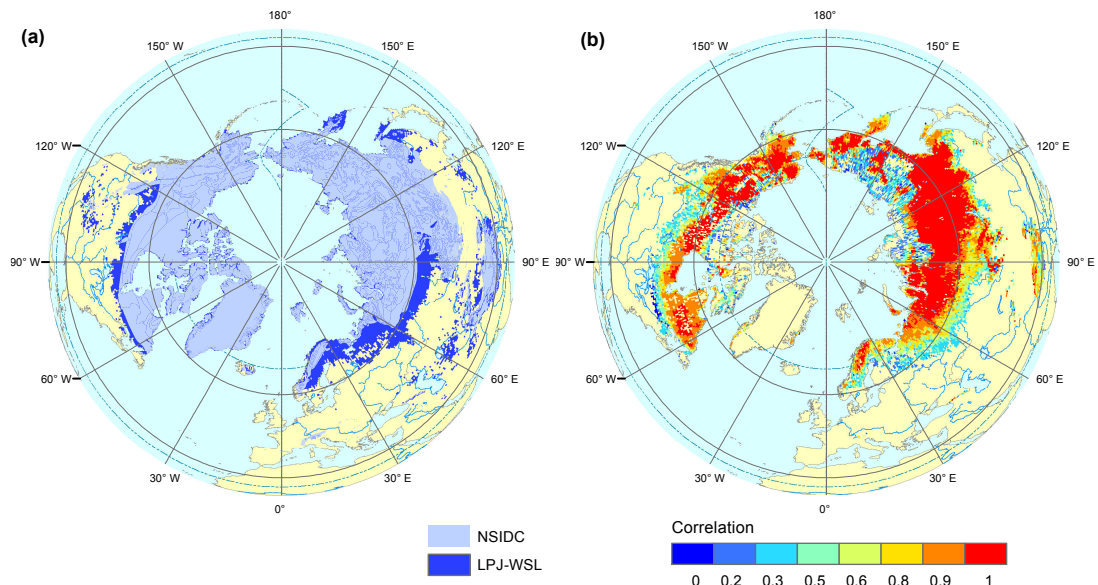
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9 Figure 2. TOPMODEL parameter maps in model experiments. Mean CTI (a, b) and  
 10  $C_s$  (c, d) aggregated by river basin (denoted as “By Basin”) and grid cell (denoted  
 11 as “By Tile”) schemes from HydroSHEDS were listed.  $F_{max}$  (e) for calibration was  
 12 generated using SWAMPS-GLWD and GLWD. Map of regions (f) was used to  
 13 partition globe into boreal, temperate, tropical biomes (Gurney et al. 2003).

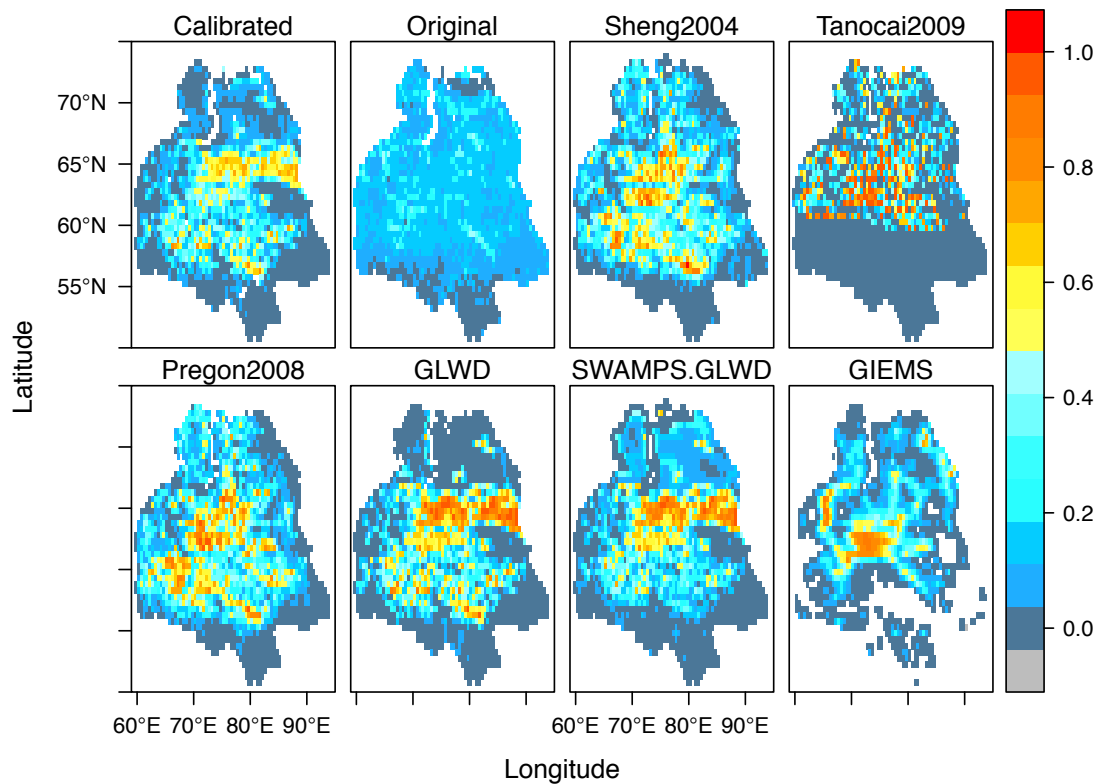
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16 Figure 3 Evaluation of permafrost simulation in LPJ-wsl. (a) Inventory-based  
 17 (light blue) and simulated (dark blue) permafrost extent from NSIDC and LPJ-wsl  
 18 respectively. The inventory contains discontinuous, sporadic or isolated  
 19 permafrost boundaries, as well as the location of subsea and relict permafrost.  
 20 We only compare the distribution of all permafrost against model outputs  
 21 without distinguishing each permafrost types. (b) Spatial distribution of  
 22 Spearman correlation between simulated monthly frozen-days from LPJ-wsl  
 23 over 2002-2011 and satellite retrievals of FT status from AMSR-E.

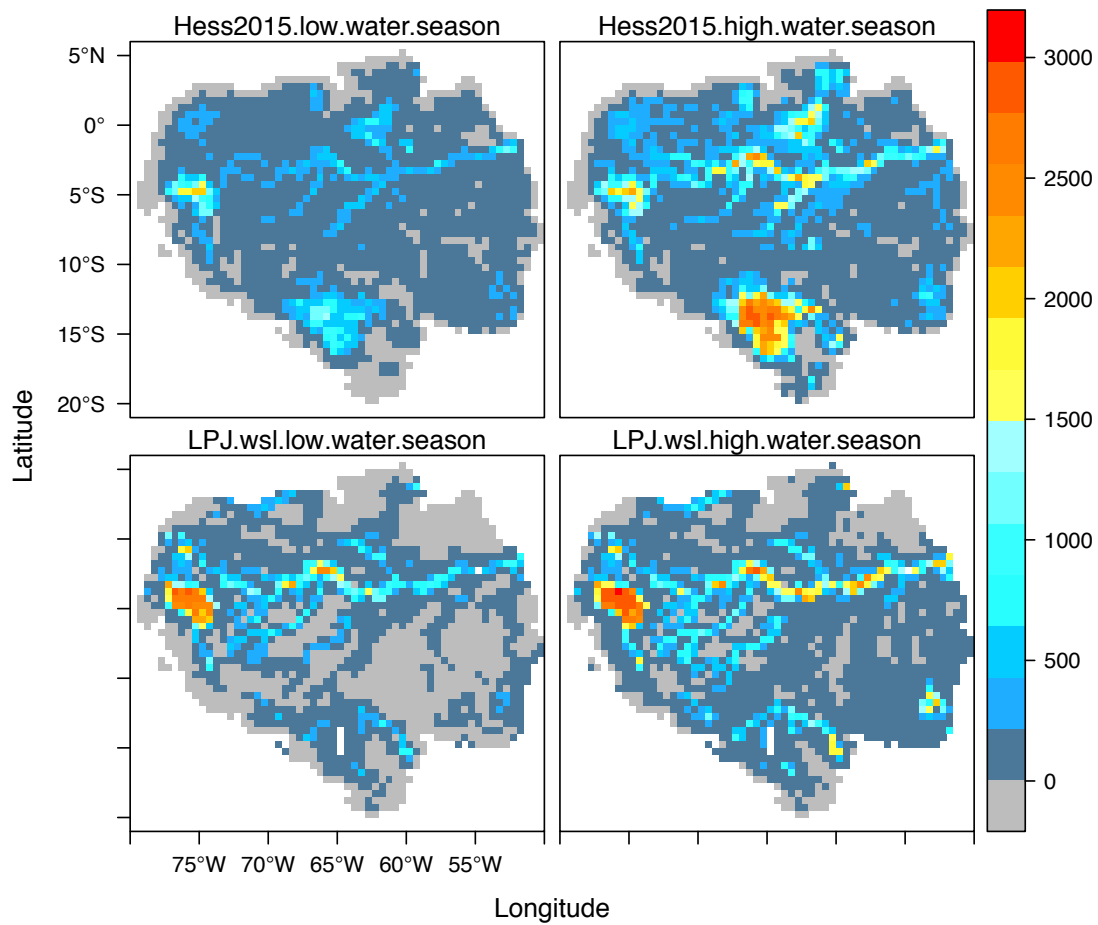
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25

26 Figure 4 Comparison of TOPMODEL-based wetland areas and Observational  
 27 datasets over the region West Siberian Lowland (WSL) for June-July-August (JJA)  
 28 average over the period 1993-2012. 'Calibrated' and 'Original' represent  
 29 simulated wetland areas with and without  $F_{max}$  calibration respectively. For  
 30 Sheng2004, Tanocai, Pregon2008, and GLWD, it represents maximum wetland  
 31 extent per  $0.5^\circ$  cell as derived from static inventory maps. For SWAMPS-GLWD  
 32 and GIEMS, areas shown are averaged for JJA over the period 1993-2007 and  
 33 2000-2012 respectively.

34

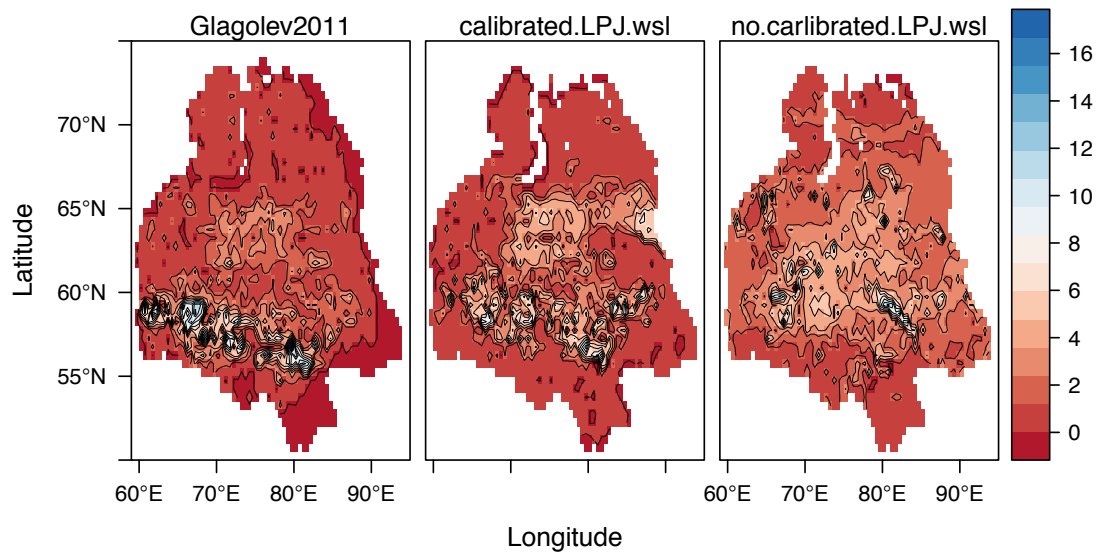


35

36 Figure 5. Comparison of wetland areas (km<sup>2</sup>) between LPJ-wsl simulated results  
 37 (SHEDS\_basin version) and JERS-1 satellite observation over Lowland Amazon  
 38 Basin for low-water season and high-water season. The low water season and  
 39 high-water season in LPJ was calculated by mean annual minimum and  
 40 maximum respectively during 1993-2013.

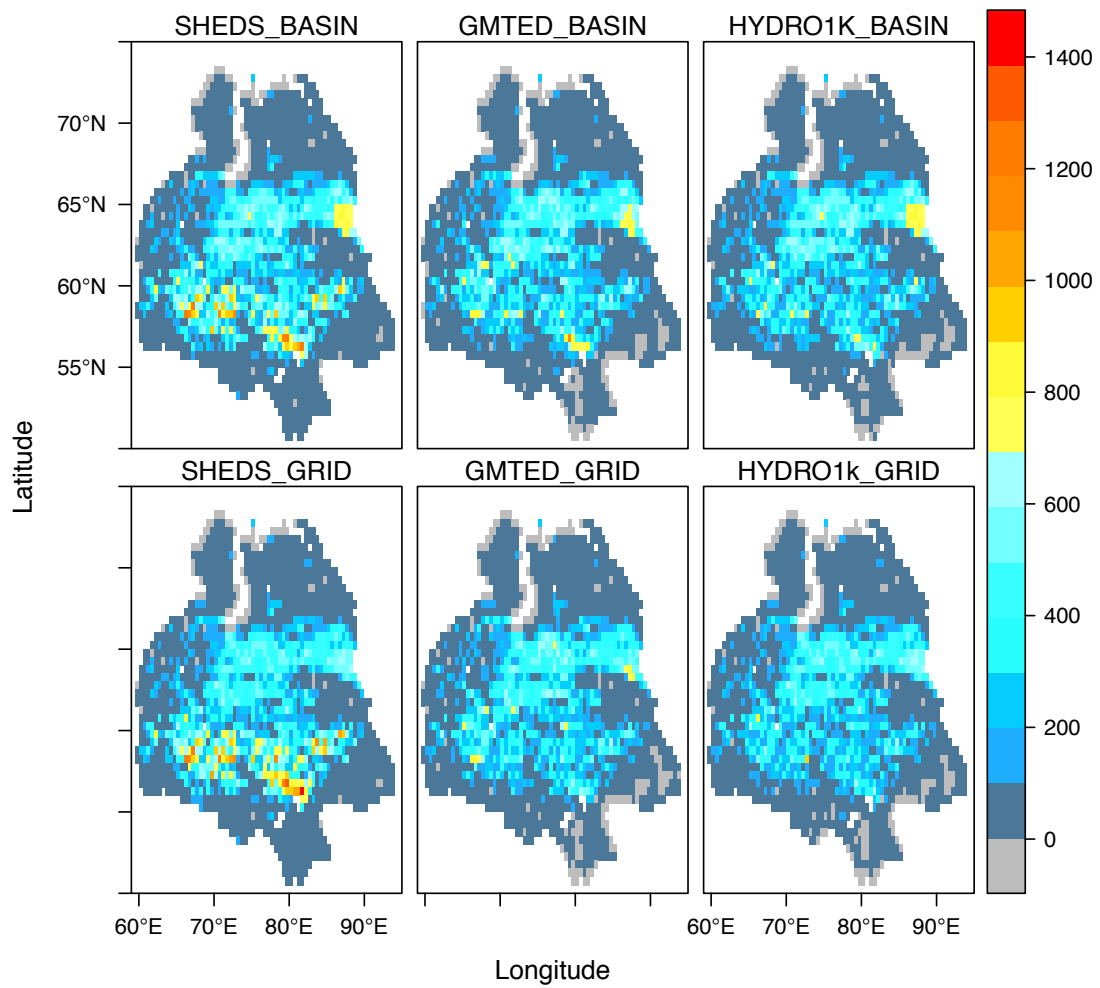
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44 Figure 6. Observation-based estimate from Glagolev et al., 2011 and two LPJ-wsl  
 45 estimates using Hydro-SHEDS (calibrated  $F_{\max}$  and non-calibrated  $F_{\max}$ ) for  
 46 annual  $\text{CH}_4$  emission ( $\text{g CH}_4 \text{ m}^{-2} \text{ yr}^{-1}$  of grid cell area). Averages from LPJ-wsl are  
 47 over the time period 2007-2010.  
 48



49

50 Figure 7. Spatial distributions of average June-July-August (JJA) wetland area  
 51 (km<sup>2</sup>) over the West Siberian Lowland (WSL) area from model experiments (see  
 52 Table 2).

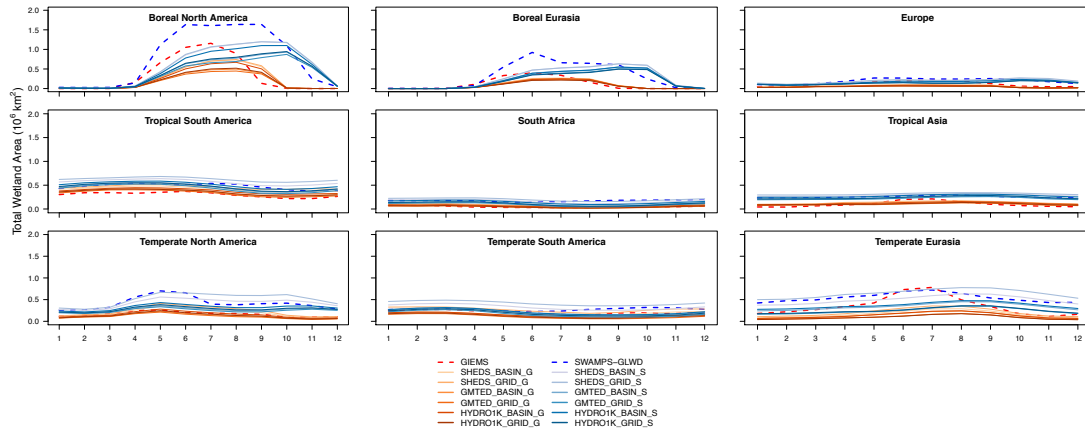


Figure 8. Average seasonal variability of observed and simulated monthly total wetland area for Transcom regions (see Fig. 2). For consistent comparison, two sets of simulated results were generated by masking out pixels for which GIEMS (red, dashed) or SWAMPS-GLWD (blue, dashed) don't have observations (denoted as 'G' and 'S', respectively).

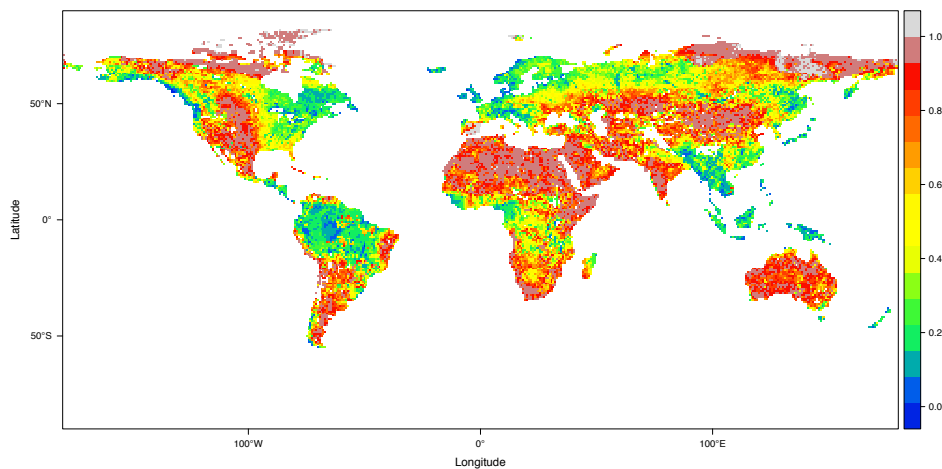
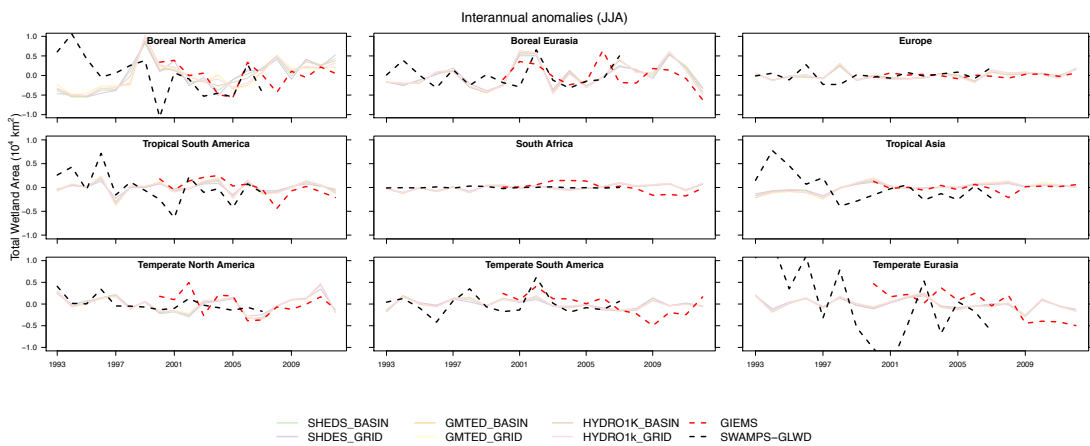
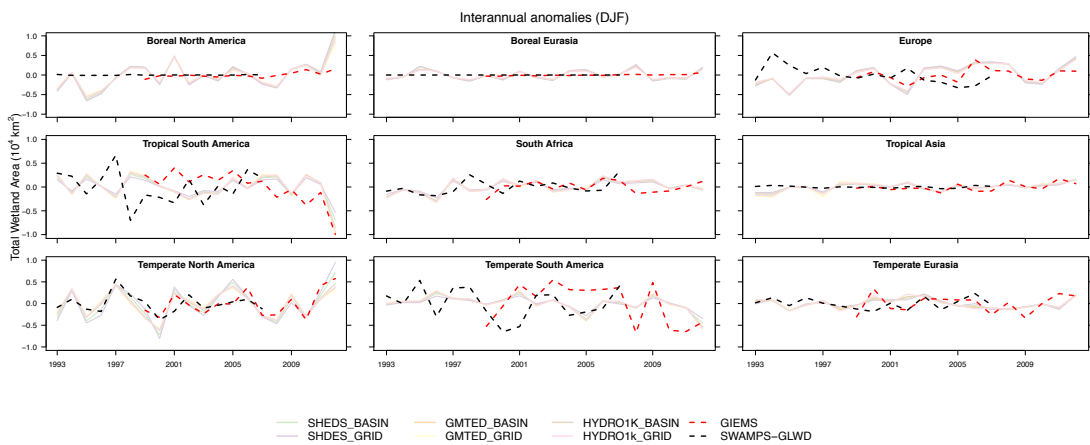
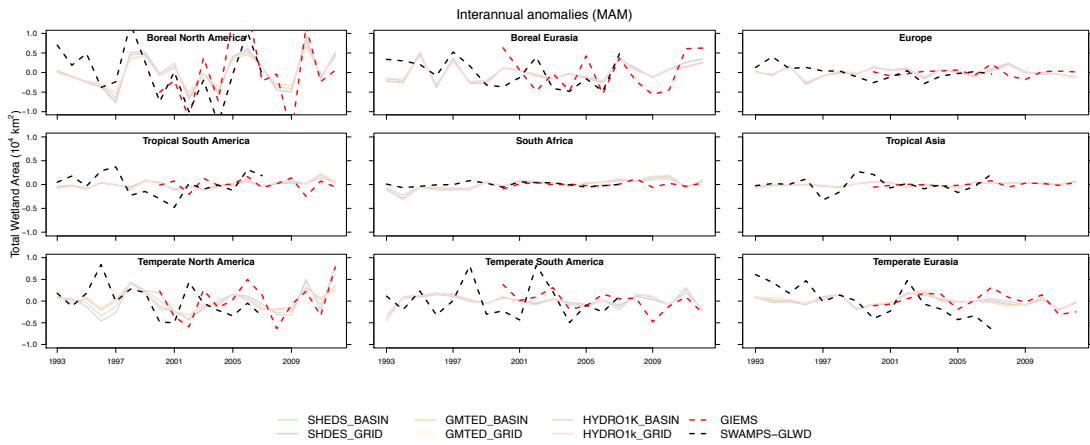


Figure 9. Global wetland potential map, which is calculated by the ratio of the mean annual maximum wetland extent averaged for the time period 1980-2010 and the long-term potential maximum wetland area ( $F_{\max}^{\text{wet}}$ ). Higher value represents higher availability for sub-grids to be inundated.



# Appendix A



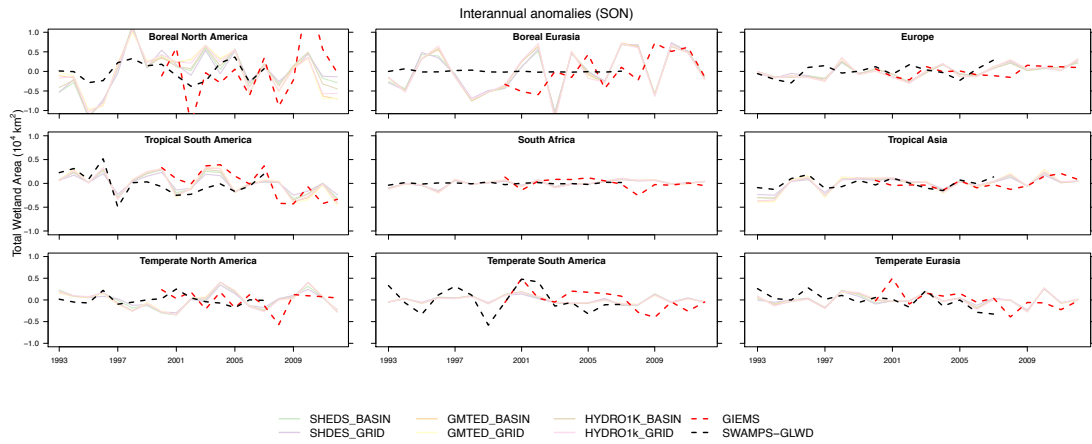


Figure A1. Interannual variations of seasonal wetland area anomalies from LPJ-wsl and satellite-derived observations for the period 1993-2012.

Table A1. Reclassification table for aggregating JERS-1 lowland Amazon basin to 0.5° cell. Code NA, 0, 1, and 2 represent Not-Available, Not Wetlands, wetland only exist in low-water season and wetland exist in high-water season.

DN	Cover at low-water stage	Cover at higher-water stage	Flag for minimum/maximum wetlands
0	Land outside Amazon Basin	Land outside Amazon Basin	NA
1	Non-wetland within Amazon Basin	Non-wetland within Amazon Basin	0
11	Open water	Open water	0
13	Open water	Aquatic macrophyte	0
21	Bare soil or herbaceous, non-flooded	Open water	2
23	Bare soil or herbaceous, non-flooded	Aquatic macrophyte	2
33	Aquatic macrophyte	Aquatic macrophyte	1
41	Shrub, non-flooded	Open water	2
44	Shrub, non-flooded	Shrub, non-flooded	0
45	Shrub, non-flooded	Shrub, flooded	2
51	Shrub, flooded	Open water	1
55	Shrub, flooded	Shrub, flooded	1
66	Woodland, non-flooded	Woodland, non-flooded	0
67	Woodland, non-flooded	Woodland, flooded	2

77	Woodland, flooded	Woodland, flooded	1
88	Forest, non-flooded	Forest, non-flooded	0
89	Forest, non-flooded	Forest, flooded	2
99	Forest, flooded	Forest, flooded	1
200	Elevation $\geq$ 500m, in Basin	Elevation $\geq$ 500, in Basin	NA
255	Ocean	Ocean	NA