Biogeosciences Discuss., 12, 18883–18911, 2015
www.biogeosciences-discuss.net/12/18883/2015/
doi:10.5194/bgd-12-18883-2015

Biogeosciences
Discussions
Open Access

Net global warming potential and greenhouse gas intensity in rice agriculture driven by high yields and nitrogen use efficiency: a 5 year field study

X. Zhang[1], Z. Zhou[1], Y. Liu[1], X. Xu[1], J. Wang[1,2], H. Zhang[1], and Z. Xiong[1]

[1]Jiangsu Key Laboratory of Low Carbon Agriculture and GHGs Mitigation, College of Resources and Environmental Sciences, Nanjing Agricultural University, Nanjing, 210095, China
[2]State Key Laboratory of Soil and Sustainable Agriculture, Institute of Soil Science, Chinese Academy of Sciences, Nanjing 210008, China

Received: 15 September 2015 – Accepted: 9 November 2015 – Published: 26 November 2015

Correspondence to: Z. Xiong (zqxiong@njau.edu.cn)

Published by Copernicus Publications on behalf of the European Geosciences Union.

Abstract

Our understanding of how net global warming potential (NGWP) and greenhouse gas intensity (GHGI) is affected by management practices aimed at food security with respect to rice agriculture remains limited. In the present study, a 5 year field experiment was conducted in China to evaluate the effects of integrated soil-crop system management (ISSM) on NGWP and GHGI after accounting for carbon dioxide (CO_2) emissions from all sources (methane, CH_4, and nitrous oxide, N_2O, emissions, agrochemical inputs, Ei, and farm operations, Eo) and sinks (i.e., soil organic carbon, SOC, sequestration). For the improvement of rice yield and agronomic nitrogen use efficiency (NUE), four ISSM scenarios consisting of different nitrogen (N) fertilization rates relative to the local farmers' practice (FP) rate were carried out, namely, N1 (25 % reduction), N2 (10 % reduction), N3 (FP rate) and N4 (25 % increase). The results showed that compared with the FP, the four ISSM scenarios, i.e., N1, N2, N3 and N4, significantly increased the rice yields by 10, 16, 28 and 41 % and the agronomic NUE by 75, 67, 86 and 82 %, respectively. In addition, compared with the FP, the N1 and N2 scenarios significantly reduced the GHGI by 14 and 18 %, respectively, despite similar NGWPs. The N3 and N4 scenarios remarkably increased the NGWP and GHGI by an average of 67 and 36 %, respectively. In conclusion, the ISSM strategies are promising for both food security and environmental protection, and the ISSM scenario of N2 is the optimal strategy to realize high yields and high NUE together with low environmental impacts for this agricultural rice field.

1 Introduction

Rapid population growth and economic development place a growing pressure on increasing food production (Barrett, 2010). An increase in global food production of 100 % is the most appropriate way to sustain the increase in human population and the consumption of animal protein (Tilman et al., 2011). Within a limited land area, the intensive agricultural regions of China are facing serious environmental problems due to large inputs of chemical fertilizers and low nitrogen use efficiency (NUE) (Ju et al., 2009; Makino, 2011). Thus, integrated soil-crop system management (ISSM), which redesigns the whole production system based on the local environment and draws on appropriate fertilizer varieties and application ratios, crop densities and advanced water regime management, has been advocated and developed to simultaneously increase crop productivity and NUE with low carbon (C) costs in China (Chen et al., 2014).

Carbon dioxide (CO_2), methane (CH_4) and nitrous oxide (N_2O) are the most important greenhouse gases (GHGs) that greatly contribute to global warming (IPCC, 2013). Although agriculture releases significant amounts of CH_4 and N_2O into the atmosphere, the net emission of CO_2 equivalents from farming activities can be partly offset by changing agricultural management to increase the soil organic matter content and/or decrease the emissions of CH_4 and N_2O (Mosier et al., 2006; Smith et al., 2008). If global agricultural techniques are improved, the mitigation potential of agriculture (excluding fossil fuel offsets from biomass) is estimated to be approximately 5.5–6.0 Pg CO_2 eq. yr^{-1} by 2030 (Smith et al., 2008). However, the release of CO_2 during the manufacturing and application of N fertilizer to crops and from fuel used in machines for farm operations can counteract these mitigation efforts (West and Marland, 2002). This indicates that agricultural ecosystems are not only a very important source of GHG emissions but also present substantial opportunities for mitigation. Therefore, when determining the net global warming potential (NGWP) of GHG (CO_2, CH_4 and N_2O) emissions from agroecosystems, there is a need to account for all sources (gas emissions, agrochemical inputs (Ei) and farm operations (Eo)) and sinks (e.g., soil organic carbon (SOC) sequestration) of the C cost or CO_2 equivalents (Sainju et al., 2014).

Information on the effects of ISSM scenarios on NGWP and greenhouse gas intensity (GHGI) is limited in China (Ma et al., 2013; Liu et al., 2015). The annual rotation of summer rice-upland crop is a dominant cropping system in China. Previous studies mainly investigated the initial influences of ISSM practices but did not account for the contributions of CO_2 emissions from Ei and Eo (Ma et al., 2013; Zhang et al., 2014). In this study, we evaluated NGWP and GHGI by taking CO_2 equivalents from all sources and sinks into account for 5 years. We hypothesized that the ISSM strategies would reduce the overall NGWP and GHGI compared with local farmers' practices (FP); thus, specific ISSM scenarios can be adopted by policy makers based on specific targets, such as high yield, high NUE and GHG mitigation. The specific objectives of this study were to (i) evaluate the effects of different ISSM scenarios on NGWP and GHGI; (ii) determine the main sources of NGWP and GHGI in a rice-wheat cropping system; and (iii) elucidate the overall performance for each ISSM scenario for different targets to increase grain yields and NUE and reduce NGWP and GHGI.

2 Materials and methods

2.1 Experimental site

A 5 year field experiment was conducted at the Changshu agro-ecological experimental station (31°32′93″ N, 120°41′88″ E) in Jiangsu Province, China. This is a typical, intensively managed agricultural area where the cropping regime is dominated by a flooding rice (*Oryza sativa* L.)-drained wheat (*Triticum aestivum* L.) rotation system. The site is characterized by a subtropical humid monsoon climate with a mean annual air temperature of 15.5 °C and precipitation of 1038 mm. The soil of the field is classified as an *Anthrosol*, which developed from lacustrine sediment. The major properties of the soil at 0–20 cm can be described as follows: bulk density, 1.11 g cm^{-3}; pH, 7.35; organic matter content, 35.0 g kg^{-1}; and total N, 2.1 g kg^{-1}. The daily mean air temperatures and precipitation during the study period from 15 June 2011 to 15 June 2014, are shown in Fig. 1.

2.2 Experimental design and management

A completely randomized design was established in 2009 with four replicates of six treatments, including no nitrogen (NN) and FP as controls, and four ISSM scenarios at different N application rates relative to the local FP rate, namely N1 (25 % reduction), N2 (10 % reduction), N3 (FP rate) and N4 (25 % increase). The designed ISSM including a redesign of a split N fertilizer application, a balanced fertilizer application, additional phosphorus and potassium application, and transplanting density, used as the main techniques for improving rice yield and agronomic NUE (calculated as the difference in grain yield between the plots that received N application and the NN plot, divided by the N fertilizer rate). The details of the fertilizer applications, irrigation, and field management practices of the six different treatments are presented in Table 1. Further detailed information was described previously (Zhang et al., 2014). Each plot was 6 m × 7 m in size with an independent drainage/irrigation system.

One midseason drainage (about one week) and final drainage before harvest were used during the rice-growing season, whereas the plots only received precipitation during the wheat-growing season. P, Si, Zn and rapeseed cake manure were applied as basal fertilizers for both crops. K was added as a split (1 : 1) application to the rice crop and used as basal fertilizer for the wheat crop. The basal fertilization occurred at the time of rice transplanting and wheat seeding.

2.3 Gas sampling and measurements

We measured the CH_4 emissions and N_2O fluxes in each plot of the field experiment over five annual cycles from the 2009 rice-growing season to the 2014 wheat-growing season. The initial 2 year measurements during the 2009–2011 rice-wheat rotational systems were described in our previous study (Ma et al., 2013). Emissions were measured manually using the static-opaque chamber method. Each replicate plot was equipped with a chamber with a size of 50 cm × 50 cm × 50 cm or 50 cm × 50 cm × 110 cm, depending on the crop growth and plant height. The chamber was placed on a fixed

PVC frame in each plot and wrapped with a layer of sponge and aluminum foil to minimize air temperature changes inside the chamber during the period of sampling.

The gas samples were analyzed for CH_4 and N_2O concentrations using a gas chromatograph (Agilent 7890A, Shanghai, China) equipped with a hydrogen flame ionization detector and an SS-2 m × 2 mm Porapak Q (80/100 mesh) column. The oven temperature remained at 50 °C, and the detector was maintained at 300 °C. The carrier gas was purified N_2 with a flow rate of $35\,\mathrm{mL\,min^{-1}}$.

2.4 Topsoil organic carbon sequestration measurements

To measure the organic carbon content of the topsoil as described by Zhang et al. (2014), soil samples were collected after the wheat harvest in 2009 and 2014 from all experimental plots at a plowing depth of 0–20 cm. The soil organic carbon sequestration rates (SOCSR) were calculated as follows:

$$\mathrm{SOCSR}\left(\mathrm{t\,C\,ha^{-1}\,yr^{-1}}\right) = (\mathrm{SOC_t} - \mathrm{SOC_0})/T \times \gamma \times \left(1 - \delta_{2\,\mathrm{mm}}/100\right) \times 20 \times 10^{-1} \qquad (1)$$

In Eq. (1), $\mathrm{SOC_t}$ and $\mathrm{SOC_0}$ are the SOC contents measured in the soils sampled after the wheat was harvested in 2014 and 2009, respectively. T refers to the experimental period (yr). γ and $\delta_{2\,\mathrm{mm}}$ are the average bulk density and the gravel content (> 2 mm) of the topsoil (0–20 cm), respectively.

2.5 NGWP and GHGI measurements

In addition to the CH_4 emissions and N_2O fluxes, we considered some “hidden” CO_2 equivalent emissions, including agrochemical inputs (Ei), such as the manufacture and transportation of the N, P and K fertilizers (Snyder et al., 2009), and farm operations (Eo), such as the water used for irrigation (Zhang et al., 2013) and diesel fuel (Huang et al., 2013a). The CO_2 equivalent emissions of N fertilizer were calculated as the mean value of the C emissions of 1.3 kg C eq. $\mathrm{kg^{-1}}$ (Lal, 2004). Similarly, the CO_2

equivalent for irrigation was calculated from the total amount of water used during the rice-growing season; the coefficient for the C cost was 5.16 ($\mathrm{kg\,C\ eq.\,cm^{-1}\,ha^{-1}}$) (Lal, 2004). The CO_2 equivalents of other Ei (P and K fertilization, manure, herbicide, pesticide and fungicide applications) and Eo (tillage, planting, harvest, and farm machinery production) were recorded and estimated according to the methods provided by Lal (2004).

The NGWP of the cropland ecosystem equals the total CO_2 emission equivalents minus the SOC change. Thus, the NGWP and GHGI were calculated using the following equations (IPCC, 2013):

$$\mathrm{NGWP}\left(\mathrm{kg\,CO_2\,eq.\,ha^{-1}yr^{-1}}\right) = \mathrm{GWP\,(CH_4 + N_2O + Ei + Eo) - GWP\ (SOCSR)} \qquad (2)$$

$$\mathrm{GHGI}\left(\mathrm{kg\,CO_2\,eq.\,kg^{-1}grain\ yield\,yr^{-1}}\right) = \mathrm{NGWP/grain\ yield} \qquad (3)$$

In Eq. (2), Ei, Eo and SOCSR represent the agrochemical inputs, farm operations and soil organic carbon sequestration rate, respectively. The grain yield is expressed as the air-dried grain yield.

2.6 Statistical analysis

One-way analysis of variance was conducted to determine the emissions of CH_4 and N_2O, and the grain yield among the different treatments using JMP, ver. 7.0 (SAS Institute, USA, 2007). Tukey's HSD test was used to determine whether significant differences occurred between the treatments at a significance level of $P < 0.05$. The results are presented as the means and standard deviation (mean ± SD, $n = 4$). Two-way analysis of variance (ANOVA) and linear relationships were determined using JMP 7.0. The F test was applied to determine whether there were significant effects of the practices, years and their interaction at $P < 0.05$.

3 Results

3.1 Crop production and agronomic NUE

During the three cropping rotations from 2011 to 2014, the rice and wheat yields varied significantly among these cultivation patterns; these results are shown in Table 2. The grain yields ranged from 5.83 to 12.11 $\mathrm{t\,ha^{-1}}$ for rice and 1.75 to 6.14 $\mathrm{t\,ha^{-1}}$ for wheat (Fig. 2a). On average over the three cycles, the annual rice yield of the FP was significantly lower than that of the ISSM scenarios of N1, N2, N3 and N4. Compared with the FP, rice grain yields increased by 10 and 16 % for the N1 and N2 scenarios, respectively, i.e., with the lower N input, by 28 % for the N3 scenario with the same N input and by 41 % for the N4 scenario with the highest N input. However, we did not observe any increases in the wheat-grain yields compared with the FP except for the N4 scenario.

The agronomic NUE for the rice and wheat of the fertilized plots ranged from 9.2 to 17.1 and 19.5 to 24.7 kg grain $\mathrm{kg^{-1}}$ N, respectively (Fig. 2b). The higher NUE in the wheat season was mainly due to the reduced N fertilizer (40 %) during this season. As expected, the rice agronomic NUE significantly increased by 75, 67, 86 and 82 % for the N1, N2, N3 and N4 scenarios, respectively, compared with the FP (Fig. 2b). For the wheat crop, the agronomic NUE merely increased by 12 and 14 % for the N1 and N2 scenarios, respectively, and decreased to some extent for the N3 and N4 scenarios compared with the FP, mainly because the current ISSM strategy was only designed for rice and not wheat production.

3.2 CH_4 and N_2O emissions

All plots showed similar CH_4 emission patterns, being a source in the rice season and negligible in the wheat season. During the three annual rice-wheat rotations from 2011 to 2014, the CH_4 fluxes ranged from −3.89 to 99.67 $\mathrm{mg\,C\,m^{-2}\,h^{-1}}$ (Fig. 3). The seasonal CH_4 emissions varied significantly among the treatments during the rice-growing season (Table 3, Fig. 3). No significant difference was found between the FP, N1 and N2 plots. Averaged across years, the CH_4 emission was greater in the N3 and N4 plots than in the NN, FP, N1 and N2 plots (Table 2, $P < 0.05$). However, compared with the NN plots, the FP, N1 and N2 plots with inorganic fertilizer application resulted in increased CH_4 emission rates of 59.9, 41.9 and 43.0 %, respectively, averaged over the rice-growing seasons. The CH_4 emission rates were further enhanced by 198.5 % in the N3 plots and by 246.7 % in the N4 plots due to the combined application of inorganic and organic fertilizers.

The annual N_2O fluxes varied from −33.1 to 647.5 µg N_2O-N m^{-2} h^{-1} (Fig. 4). With respect to the N application effect, the annual cumulative N_2O emissions for all four ISSM scenarios were significantly higher than in NN ($P < 0.05$). Relative to the FP plot, the N1 and N2 scenarios decreased the annual N_2O emissions by an average of 41 and 22 %, respectively, while the N4 scenario significantly increased it by 46 % ($P < 0.05$) although there was no significant difference between the N3 and FP plots. Correlations between seasonal cumulative N_2O emissions and fertilizer N application rates were also calculated and the seasonal cumulative N_2O emissions increased exponentially with an increase in the N application rate (Fig. 2c).

3.3 Annual NGWP and GHGI

Based on the perspective of the carbon footprint, we included the GHG emissions associated with all of the inputs (Ei and Eo), and SOC sequestration was expressed as kg CO_2 eq. ha^{-1} yr^{-1}. The emission of CO_2 equivalents for Ei and Eo are classified in Table 4. While irrigation was a large proportion of farm operations, these were much less significant than chemical inputs. The CO_2 equivalents rates from N fertilizer dominated not only the chemical input section (68–76 % of Ei) but also the total CO_2 equivalents from agricultural management (46–51 % of the sum of the Ei and Eo). The NGWP ranged from 7871 to 20 847 kg CO_2 eq. ha^{-1} yr^{-1} for the NN and the N4 plots, respectively (Table 5). Although N fertilizer increased the annual CH_4 and N_2O emissions, it also increased the SOC sequestration in this cropping system. Of the main field GHGs that were directly emitted, CH_4 accounted for 56–75 % of the NGWP in all plots. An increase in the annual SOC content led to a significant decrease in the NGWP (contributed to 5–10 % of the NGWP except in the NN plot). The CO_2 equivalents from machinery used for Ei (2449–4192 CO_2 eq. ha^{-1} yr^{-1}) were higher than those for Eo (1285–1697 CO_2 eq. ha^{-1} yr^{-1}) in the fertilized plots. There was no significant difference in the annual NGWP observed between the FP, N1 and N2 plots (Table 5). Across the three years, N1 and N2 slightly reduced the NGWP by 12 and 10 %, respectively; however, N3 and N4 significantly increased the NGWP by an average of 52 and 81 %, respectively, in comparison with the FP. Consequently, the lowest NGWP was achieved under the N1 scenario for the ISSM.

The GHGI was used to express the relationship between NGWP and grain yield. The GHGIs in this study ranged from 664 to 1143 kg CO_2 eq. t^{-1} (Table 5). A significant difference in the annual GHGI was found between the FP and the ISSM strategies. Compared with the FP, N1 and N2 significantly reduced the GHGI by 14 and 18 %, respectively, mainly due to the increased grain yield and SOC sequestration as well as reduced GHG emissions for the ISSM strategies. Although N fertilizer or organic/inorganic combination fertilizer application reduced the SOC losses caused by crop cultivation and increased the grain yields, the GHGIs were generally higher for the N3 and N4 scenarios than the N1 and N2 scenarios due to further increases in CH_4 and N_2O emissions.

4 Discussion

4.1 Grain yield and agronomic NUE as affected by ISSM strategies

Grain yields are directly related to fertilizer management. The two-way ANOVA results indicated that the rice and wheat grain yields were significantly affected by the cultivation strategies (Table 3, $P < 0.001$), which is in agreement with previous results (Chen et al., 2011; Zhang et al., 2011). Compared with the FP plot, the rice yields were remarkably increased by all four ISSM scenarios (Table 2). However, the wheat grain yield decreased significantly when the N fertilizer rate was reduced by 25 % (N1 scenario). It has been reported in previous studies that ISSM strategies can effectively improve the rice grain yield (Ma et al., 2013; Liu et al., 2015). In the present study, N1 and N2 significantly increased annual rice production by 10 and 16 %, respectively, in comparison with the FP (Table 2). The finding is consistent with the result of Peng et al. (2006), who reported that a 30 % reduction in the total N rate during the early vegetative stage did not reduce the yield but slightly increased it when combined with the modified farmers' fertilizer practice. As expected, when the total N rate was at the FP rate and increased by 25 % and applied with rapeseed cake manure, the rice yield in these N3 and N4 plots remarkably increased by 28 and 41 %, respectively. Based on a long-term fertilizer experiment, Shang et al. (2011) reported that organic fertilizer incorporation significantly increased the early rice grain yield. This may have resulted from the organic fertilizer applied in combination with adequate nutrients, which improved the rice yield and the efficient control of pests and diseases (Slaton et al., 2005; Wang et al., 2005). The higher yields in N3 and N4 scenarios would have also attributed to the addition of 15 kg Zn ha^{-1} for better seedling establishment and vigor according to Slaton et al. (2005) and the application of silicon to reduce both biotic and abiotic stress according to Wang et al. (2005).

It has been suggested that N losses vary depending on the timing, rate, and method of N application, as well as the source of N fertilizer (Zhu, 1997). In general, rapid N losses (via ammonia volatilization, denitrification, surface runoff, and leaching), the high application rate of fertilizer N and improper timing of N application are three important factors that cause low agronomic NUE of irrigated rice in China (Peng et al., 2006). Compared with the FP plot, the rice agronomic NUE was significantly increased by 75, 67, 86 and 82 % under the N1, N2, N3 and N4 scenarios, respectively (Fig. 2b). The higher rice agronomic NUE in our study over the experimental period was primarily due to the greatly reduced N losses by leaching and volatilization as well as the improvement of N bioavailability in the rice crop season. Organic/inorganic combination fertilizer application also increases uptake by crops compared with the traditional farmers' practice (Peng et al., 2006). These findings suggest that the ISSM strategy is an effective method for improving grain yield and agronomic NUE for future sustainable rice agriculture in China.

4.2 CH_4 and N_2O emissions as affected by ISSM strategies

During the three years, the annual cumulative CH_4 emissions, on average, varied from 133 to 469 kg C ha^{-1} yr^{-1} (Table 2), and these values fell within the range of 4.1 to 1015.6 kg CH_4 ha^{-1} observed previously in a rice field (Huang et al., 2004). The two-way ANOVA results indicated that obvious effects of cultivation patterns on CH_4 emissions were found during the rice-wheat rotations (Table 3, $P < 0.001$). The CH_4 emissions were not significantly affected by the cycles but affected by crop season (Table 5, Fig. 3). In this study, no significant difference in CH_4 emission was observed between the FP, N1 and N2 plots. However, compared with the FP plot, the N3 and N4 scenarios emitted 87 and 118 % more CH_4 emissions, respectively (Table 5), which is probably due to the incorporation of the organic rapeseed cake manure. Previous reports support the observations that CH_4 emissions were significantly increased with the application of organic amendments (Ma et al., 2009; Zou et al., 2005). Moreover, rice growth was found to be significantly increased under the N3 and N4 scenarios. In this case, the organic matter inputs such as root litter and rhizodeposits in the N3 and N4 scenarios were probably also higher than in the other plots, and thus soil C input, which served as an additional source of substrates for the methanogens in the rice paddies, likely contributed to the increase in CH_4 emissions (Ma et al., 2009). Finally, because the rice plants acted as the main pathway for CH_4 transports from the soil to the atmosphere, the higher biomass the more CH_4 emissions (Yan et al., 2005). The results obtained in the present study revealed that both inorganic and organic fertilizer application significantly increased the CH_4 emissions in the rice season (Table 2), which was probably associated with the increase in the SOC content and crop biomass (Ma et al., 2013).

Denitrification and nitrification are the main processes that produce N_2O in the soil. Changes in the soil water content strongly affected the soil N_2O emissions and resulted in negligible N_2O emissions when the rice field was flooded (Fig. 4), which is consistent with previous reports (Akiyama et al., 2005; Murdiyarso et al., 2010). A relatively high N_2O peak was observed in the first two weeks of the wheat-growing season (Fig. 4), possibly because soil changes from flooded to drained condition may have enhanced N_2O release. Alternation of drainage and flooding may induce large amounts of N_2O emissions, particularly in fertilized systems; this has commonly been proved in earlier studies (Wang et al., 2013; Xiong et al., 2007; Zou et al., 2005). The seasonal and annual rates of N_2O emission were significantly affected by the cultivation practice patterns (Table 3). Our results showed that the seasonal cumulative N_2O emissions increased exponentially with an increase in the N application rate (Fig. 2c). Compared with the FP plot, the N2 scenario greatly decreased the seasonal N_2O emissions in this study, which may have resulted from a reduction in the N fertilizer rate (Tables 1 and 2). The total N_2O emissions decreased by 7–38 % and 26–42 % in the rice and wheat seasons, respectively, when the conventional N management (300 kg N ha^{-1} for rice and 180 kg N ha^{-1} per crop for wheat) changed to optimum N management (225–270 kg N ha^{-1} for rice and 135–162 kg N ha^{-1} per crop for wheat). It is likely that more N_2O was emitted (Mosier et al., 2006) as a result of the additional N made available to the soil microbes through N fertilizer application, which probably increased the CH_4 emissions (Banger et al., 2013). Strategies that can reduce N fertilization rates without influencing crop yields can inevitably lower GHG emissions (Mosier et al., 2006).

4.3 NGWP and GHGI as affected by ISSM strategies

The NGWP in our study (10 104–20 847 kg CO_2 eq. ha^{-1}) with the ISSM strategies was higher than that in a double-cropping cereal rotation (1346–4684 kg CO_2 eq. ha^{-1}) and a rice-wheat annual rotation (290–4580 kg CO_2 eq. ha^{-1}) reported by Huang et al. (2013b) and Yang et al. (2015), respectively. Dominant CH_4 emissions as well as additional CO_2 emitted by the machinery/equipment used for irrigation and farm operations under the ISSM strategies may increase the NGWP more than in other cropping systems. However, the current NGWP was still much lower than that of a double-rice cropping system (13 407–26 066 kg CO_2 eq. ha^{-1}) (Shang et al., 2011). The GHGIs, which ranged from 0.66 to 1.14 kg CO_2 eq. kg^{-1} grain in this study, were slightly higher than previous estimates of 0.24–0.74 kg CO_2 eq. kg^{-1} grain from rice paddies with mid-season drainage and organic manure incorporation (Qin et al., 2010; Li et al., 2006) but were lower than the DNDC model estimates for continuous waterlogged paddies (3.22 kg CO_2 eq. kg^{-1} grain) (Li et al., 2006). Differences in NGWP or GHGI were found in the cultivation patterns over the three rice-wheat rotations (Table 5). Although there were no significant differences among the FP, N1 and N2 plots, the N1 and N2 scenarios with optimized ISSM strategies led to a lower NGWP than the FP (Table 5). Compared with the FP, the N1 and N2 scenarios dramatically reduced the GHGI. In spite of the considerable NGWP compared with the FP plot, the lowest GHGI (0.66 kg CO_2 eq. kg^{-1} grain) was obtained under the N2 scenario. This finding is consistent with the suggestion made by Burney et al. (2010), i.e., that the net effect of higher yields offsets emissions. It is well known that CH_4 emissions dominate the NGWP in rice paddies (Ma et al., 2013; Shang et al., 2011). In comparison to the NGWP (11 545 kg CO_2 eq. ha^{-1} yr^{-1}) and GHGI (0.81 kg CO_2 eq. kg^{-1} grain) of the FP, the N3 and N4 scenarios increased both the NGWP and GHGI, mainly because these scenarios notably increased the CH_4 emissions compared with the FP, which resulted in relatively higher NGWP (Table 5).

Agricultural management practices that change one type of GWP source/sink may also impact other sources/sinks and therefore change the NGWP and GHGI (Mosier et al., 2006; Shang et al., 2011). Although the N-fertilizer plots, especially those with the incorporation of organic fertilizer, increased the annual CH_4 and N_2O emissions, they increased the SOC sequestration in this cropping system, which is agreement with previous reports (Huang and Sun, 2006). In the present study, the N2 scenario with ISSM decreased the CH_4 and N_2O emissions as well as the energy consumption related to irrigation and the manufacture and transport of N fertilizer (depending on coal combustion), ultimately leading to a decrease in the NGWP relative to the FP plot. Moreover, despite the lower N fertilizer input, the grain yield did not decline and the GHGI of the N2 scenario was thus lower than of the FP plot, indicating less consumption of CO_2 equivalents per unit grain produced. We demonstrate that high yield and agronomic NUE, together with low GWP, are not conflicting goals by optimizing ISSM strategies.

4.4 Main components of NGWP and GHGI and implementation significance for the ISSM strategies

Determining the main components of the NGWP and GHGI in specific cropping systems is very important for mitigating GHG emissions in the future because the benefits of C sequestration would be negated by CH_4 and N_2O emissions and the CO_2 equivalents released with the use of high N fertilizer application rates (Schlesinger, 2010). In the current study, the five main components of the CO_2 equivalents for the NGWP were ranked in decreasing order of importance as follows: CH_4 emissions > agrochemical inputs of N fertilizer > farm operations related to irrigation > SOC sequestration > N_2O emissions (Table 5).

Although N fertilizer application increased SOC sequestration when it was applied with rapeseed cake manure, this benefit was consistently overshadowed, on a CO_2 equivalent basis, by the increases in CH_4 and N_2O emissions (Table 5). Similar results have been reported, i.e., that GHG emissions substantially offset SOC increases (Six et al., 2004). It is possible that the realization of reducing the NGWP and GHGI in China should focus on increasing the SOC and simultaneously decreasing the CO_2 equivalents from CH_4 emissions and N fertilizer inputs. Several studies reported possible methods for these types of mitigation strategies, such as optimizing the chemical fertilizer application amount and rate (Ju et al., 2011), the amount of water used for irrigation (Gao et al., 2015), and the timing and rate of N using the in-season N management approach, as well as improving the N fertilizer manufacturing technologies (Zhang et al., 2013), and using nitrification inhibitors or polymer-coated controlled-release fertilizers (Hu et al., 2013).

China is a rapidly developing country that faces the dual challenge of substantially increasing grain yields at the same time as reducing the very substantial environmental impacts of intensive agriculture (Chen et al., 2011). We used the ISSM strategies to develop a rice production system that achieved mean yields of 10.63 t ha^{-1} (an increment of almost 24 %) and an agronomic NUE of 16.33 kg grain kg^{-1} N (an approximate doubling) in long-term field experiments compared with current farmers' practices. The ISSM redesigned the whole production system only for the rice crop based on the local environment and drawing on appropriate fertilizer varieties and application ratios, crop densities and an advanced water regime management. If the ISSM strategies were also developed for the rotated wheat crop, the overall performance of the whole rice-wheat system would be much improved, with further increases in yield and reductions in the NGWP and GHGI. We conclude that the ISSM strategies are promising, particularly the ISSM-N2 scenario, which is the most favorable to realize higher yields with lower environmental impact. The proposed ISSM strategies can provide substantial benefits to intensive agricultural systems and can be applied feasibly using current technologies.

5 Conclusions

Reasonable agricultural management practices are the key to reducing GHG emissions from agricultural ecosystems. This study provided an insight into the complete GHG emission accounting of the NGWP and GHGI affected by different ISSM scenarios. After a five-year field experiment, we found that the CH_4 emissions, production of N fertilizer, irrigation, SOC sequestration and N_2O fluxes were the main components of the NGWP in a typical rice-wheat rotation system. In contrast with the FP, N1 and N2 significantly reduced the GHGI, though they resulted in similar NGWPs, and N3 and N4 remarkably increased the NGWP and GHGI, indicating that further research is required for mitigating GHG emissions when aiming to increasing rice yield and NUE. By adopting the ISSM strategy, the conventional N application rate was reduced by 10 % while the rice yield was significantly increased by 16 %, the NUE was improved by 67 % and the GHGI was lowered. ISSM scenarios could be adopted for both food security and environmental protection with specific targets. We propose that the ISSM-N2 scenario is the most appropriate management strategy for realizing higher yields and NUE, together with some potential to reduce GHGI by integrated soil-crop management.

Acknowledgements. This work was jointly supported by the National Science Foundation of China (41171238, 41471192), the Special Fund for Agro-Scientific Research in the Public Interest (201503106) and the Ministry of Science and Technology (2013BAD11B01).

References

Akiyama, H., Yagi, K. and Yan, X. Y.: Direct N_2O emissions from rice paddyfields: summary of available data, Global Biogeochem. Cy., 19, GB002378, doi:10.1029/2004GB002378, 2005.

Banger, K., Tian, H. and Lu, C.: Do nitrogen fertilizers stimulate or inhibit methane emissions from rice fields?, Glob. Change Biol., 18, 3259–3267, 2013.

Barrett, C. B.: Measuring food insecurity, Science, 327, 825–828, 2010.

Burney, J. A., Davis, S. J. and Lobell, D. B.: Greenhouse gas mitigation by agricultural intensification, Proc. Natl. Acad. Sci. USA, 107, 12052–12057, 2010.

Chen, X., Cui, Z., Vitousek, P. M., Cassman, K. G., Matson, P. A., Bai, J., Meng, Q., Hou, P., Yue, S. and Römheld, V.: Integrated soil-crop system management for food security, P. Natl. Acad. Sci. USA, 108, 6399–6404, 2011.

Chen, X., Cui, Z., Fan, M., Vitousek, P., Zhao, M., Ma, W., Wang, Z., Zhang, W., Yan, X. and Yang, J.: Producing more grain with lower environmental costs, Nature, 514, 486–489, 2014.

Gao, B., Ju, X., Meng, Q., Cui, Z., Christie, P., Chen, X. and Zhang, F.: The impact of alternative cropping systems on global warming potential, grain yield and groundwater use, Agr. Ecosyst. Environ., 203, 46–54, 2015.

Hu, X., Su, F., Ju, X., Gao, B., Oenema, O., Christie, P., Huang, B., Jiang, R. and Zhang, F.: Greenhouse gas emissions from a wheat-maize double cropping system with different nitrogen fertilization regimes, Environ. Pollut., 176, 198–207, 2013.

Huang, J., Chen, Y., Sui, P. and Gao, W.: Estimation of net greenhouse gas balance using crop-and soil-based approaches: two case studies, Sci. Total Environ., 456, 299–306, 2013.

Huang, T., Gao, B., Christie, P., and Ju, X.: Net global warming potential and greenhouse gas intensity in a double-cropping cereal rotation as affected by nitrogen and straw management, Biogeosciences, 10, 7897–7911, doi:10.5194/bg-10-7897-2013, 2013.

Huang, Y. and Sun, W.: Changes in topsoil organic carbon of croplands in mainland China over the last two decades, Chinese Sci. Bull., 51, 1785–1803, 2006.

Huang, Y., Zhang, W., Zheng, X. H., Li, J. and Yu, Y. Q.: Modeling methane emission from rice paddies with various agricultural practices. J. Geophys. Res.-Atmos., 109, D08, doi:10.1029/2003JD004401, 2004.

IPCC: Climate Change 2013: The Physical Science Basis: working group I contribution to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change, Cambridge University Press, Stockholm, 2013.

Ju, X., Xing, G., Chen, X., Zhang, S., Zhang, L., Liu, X., Cui, Z., Yin, B., Christie, P. and Zhu, Z.: Reducing environmental risk by improving N management in intensive Chinese Agricultural Systems, P. Natl. Acad. Sci. USA, 106, 3041–3046, 2009.

Ju, X., Lu, X., Gao, Z., Chen, X., Su, F., Kogge, M., Römheld, V., Christie, P. and Zhang, F.: Processes and factors controlling N_2O production in an intensively managed low carbon calcareous soil under sub-humid monsoon conditions, Environ. Pollut., 159, 1007–1016, 2011.

Lal, R.: Carbon emission from farm operations, Environ. Int., 30, 981–990, 2004.

Li, C. S., Salas, W., DeAngelo, B. and Rose, S.: Assessing alternatives formitigating net greenhouse gas emissions and increasing yields from rice production in China over the next twenty years, J. Environ. Qual., 35, 1554–1565, 2006.

Liu, Y. L., Zhou, Z., Zhang, X., Xu, X., Chen, H. and Xiong, Z.: Net global warming potential and greenhouse gas intensity from the double rice system with integrated soil-crop system management: a three-year field study, Atmos. Environ., 116, 92–101, 2015.

Ma, J., Ma, E., Xu, H., Yagi, K. and Cai, Z.: Wheat straw management affects CH_4 and N_2O emissions from rice fields, Soil Biol. Biochem., 41, 1022–1028, 2009.

Ma, Y., Kong, X., Yang, B., Zhang, X., Yan, X., Yang, J. and Xiong, Z.: Net global warming potential and greenhouse gas intensity of annual rice–wheat rotations with integrated soil–crop system management, Agr. Ecosyst. Environ., 164, 209–219, 2013.

Makino, A.: Photosynthesis, grain yield and nitrogen utilization in rice and wheat, Plant Physiol., 155, 125–129, 2011.

Mosier, A. R., Halvorson, A. D., Reule, C. A. and Liu, X. J.: Net global warming potential and greenhouse gas intensity in irrigated cropping systems in northeastern Colorado, J. Environ. Qual., 35, 1584–1598, 2006.

Murdiyarso, D., Hergoualc'h, K. and Verchot, L. V.: Opportunities for reducing greenhouse gas emissions in tropical peatlands, P. Natl. Acad. Sci. USA, 107, 19655–19660, 2010.

Peng, S., Buresh, R. J., Huang, J., Yang, J., Zou, Y., Zhong, X., Wang, G. and Zhang, F.: Strategies for overcoming low agronomic nitrogen use efficiency in irrigated rice systems in China, Field Crop. Res., 96, 37–47, 2006.

Qin, Y. M., Liu, S. W., Guo, Y. Q., Liu, Q., Zou and J. W.: Methane and nitrous oxide emissions from organic and conventional rice cropping systems in Southeast China, Biol. Fertil. Soils., 46, 825–834, 2010.

Sainju, U. M., Stevens, W. B., Caesar-TonThat, T., Liebig, M. A. and Wang, J.: Net global warming potential and greenhouse gas intensity influenced by irrigation, tillage, crop rotation and nitrogen fertilization, J. Environ. Qual., 43, 777–788, 2014.

Schlesinger, W. H.: On fertilizer-induced soil carbon sequestration in China's croplands, Glob. Change Biol., 16, 849–850, 2010.

Shang, Q., Yang, X., Gao, C., Wu, P., Liu, J., Xu, Y., Shen, Q., Zou, J. and Guo, S.: Net annual global warming potential and greenhouse gas intensity in Chinese double rice-cropping systems: a 3 year field measurement in long-term fertilizer experiments, Glob. Change Biol., 17, 2196–2210, 2011.

Six, J., Ogle, S. M., Conant, R. T., Mosier, A. R. and Paustian, K.: The potential to mitigate global warming with no-tillage management is only realized when practised in the long term, Glob. Change Biol., 10, 155–160, 2004.

Slaton, N. A., Norman, R. J. and Wilson, C. E.: Effect of zinc source and application time on zinc uptake and grain yield of flood-irrigated rice, Agron. J., 97, 272–278, 2005.

Smith, P., Martino, D., Cai, Z., Gwary, D., Janzen, H., Kumar, P., McCarl, B., Ogle, S., O'Mara, F. and Rice, C.: Greenhouse gas mitigation in agriculture, Philos. T. R. Soc. B, 363, 789–813, 2008.

Snyder, C., Bruulsema, T., Jensen, T. and Fixen, P.: Review of greenhouse gas emissions from crop production systems and fertilizer management effects, Agr. Ecosyst. Environ., 133, 247–266, 2009.

Tilman, D., Balzer, C., Hill, J., Befort, B. L. and Affiliations, A.: Global food demand and the sustainable intensification of agriculture, P. Natl. Acad. Sci. USA, 108, 20260–20264, 2011.

Wang, L. J., Nie, Q., Li, M., Zhang, F. S., Zhuang, J. Q., Yang, W. S., Li, T. J. and Wang, Y. H.: Biosilicified structures for cooling plant leaves: a mechanism of highlyefficient midinfrared thermal emission, Appl. Phys. Lett., 87, 194105, doi:10.1063/1.2126115, 2005.

West, T. O. and Marland, G.: A synthesis of carbon sequestration, carbon emissions and net carbon flux in agriculture: comparing tillage practices in the United States, Agr. Ecosyst. Environ., 91, 217–232, 2002.

Xiong, Z. Q., Xing, G. X. and Zhu, Z. L.: Nitrous oxide and methane emissions asaffected by water, soil and nitrogen, Pedosphere, 17, 146–155, 2007.

Yan, X. Y., Yagi, K., Akiyama, H. and Akimoto, H.: Statistical analysis of the majorvariables controlling methane emission from rice fields, Glob. Change Biol., 11, 1131–1141, 2005.

Yang, B., Xiong, Z., Wang, J., Xu, X., Huang, Q. and Shen, Q.: Mitigating net global warming potential and greenhouse gas intensities by substituting chemical nitrogen fertilizers with organic fertilization strategies in rice-wheat annual rotation systems in China: a 3 year field experiment, Ecol. Eng., 81, 289–297, 2015.

Zhang, F., Cui, Z., Fan, M., Zhang, W., Chen, X. and Jiang, R.: Integrated soil-crop system management: reducing environmental risk while increasing crop productivity and improving nutrient use efficiency in China, J. Environ. Qual., 40, 1051–1057, 2011.

Zhang, W., Dou, Z., He, P., Ju, X., Powlson, D., Chadwick, D., Norse, D., Lu, Y., Zhang, Y. and Wu, L.: New technologies reduce greenhouse gas emissions from nitrogenous fertilizer in China, P. Natl. Acad. Sci. USA, 110, 8375–8380, 2013.

Zhang, X., Fan, C., Ma, Y., Liu, Y., Li, L., Zhou, Q. and Xiong, Z.: Two approaches for net ecosystem carbon budgets and soil carbon sequestration in a rice-wheat rotation system in China, Nutr. Cycl. Agroecosys., 100, 301–313, 2014.

Zhu, Z.: Fate and Management of Fertilizer Nitrogen in Agro-Ecosystems. Nitrogen in Soils of China, Springer Netherlands, 239–279, 1997.

Zou, J. W., Huang, Y., Jiang, J. Y., Zheng, X. H. and Sass, R. L.: A 3 year field measurement of methane and nitrous oxide emissions from rice paddies in China: effects of water regime, crop residue, and fertilizer application, Global Biogeochem. Cy., 19, GB002401, doi:10.1029/2004GB002401, 2005.

Table 1. The establishment of different treatments for the annual rice-wheat rotations during the 2011–2014 cycle.

Scenario	NN	FP	ISSM-N1	ISSM-N2	ISSM-N3	ISSM-N4
				Rice-growing season		
Chemical fertilizer application rate ($N:P_2O_5:K_2O:Na_2SiO_3:ZnSO_4$, kg ha^{-1})	0 : 90 : 120 : 0 : 0	300 : 90 : 120 : 0 : 0	225 : 90 : 120 : 0 : 0	270 : 90 : 120 : 0 : 0	300 : 108 : 144 : 225 : 15	375 : 126 : 180 : 225 : 15
Split N application ratio		6 : 2 : 0 : 2	5 : 1 : 2 : 2	5 : 1 : 2 : 2	5 : 1 : 2 : 2	5 : 1 : 2 : 2
Rapeseed cake manure (t ha^{-1})	0	0	0	0	2.25	2.25
Water regime	F-D-F-M	F-D-F-M	F-D-F-M	F-D-F-M	F-D-F-M	F-D-F-M
Planting density (cm)	20 × 20	20 × 20	20 × 15	20 × 15	20 × 15	20 × 20
				Wheat-growing season		
Chemical fertilizer application rate ($N:P_2O_5:K_2O$, kg ha^{-1})	0 : 90 : 180	180 : 90 : 180	135 : 90 : 180	162 : 90 : 180	180 : 108 : 216	225 : 126 : 270
Split N application ratio		6 : 1 : 3	6 : 1 : 3	6 : 1 : 3	6 : 1 : 3	6 : 1 : 3
Seed sowing density (kg ha^{-1})	180	180	180	180	180	180

NN, no N application; FP, farmers' practice; the four integrated soil-crop system management (ISSM) practices at different nitrogen application rates relative to the FP rate of 300 kg N ha^{-1} for the rice crop and 180 kg N ha^{-1} for the wheat crop, namely, N1 (25 % reduction), N2 (10 % reduction), N3 (FP rate) and N4 (25 % increase). Urea, calcium biphosphate and potassium chloride were used as N, P and K fertilizer respectively.
F-D-F-M, flooding-midseason drainage-re-flooding-moist irrigation.

Table 2. Seasonal CH_4 and N_2O emissions, and rice and wheat grain yields during the rice- and wheat-growing seasons in the three cycles of 2011–2014.

Treatment	Rice season CH_4 (kg C ha^{-1})	N_2O (kg N ha^{-1})	Yield (t ha^{-1})	Wheat season CH_4 (kg C ha^{-1})	N_2O (kg N ha^{-1})	Yield (t ha^{-1})
2011						
NN	153.15 ± 10.76	0.03 ± 0.05	5.85 ± 0.08	−0.48 ± 0.63	0.45 ± 0.09	1.74 ± 0.18
FP	265.78 ± 25.28	0.11 ± 0.08	8.38 ± 0.35	−0.48 ± 1.86	1.43 ± 0.19	5.67 ± 0.20
ISSM-N1	212.18 ± 30.26	0.08 ± 0.03	9.27 ± 0.26	0.78 ± 0.97	0.65 ± 0.11	5.05 ± 0.16
ISSM-N2	219.88 ± 32.50	0.17 ± 0.11	9.79 ± 0.44	2.25 ± 2.07	0.80 ± 0.06	5.71 ± 0.18
ISSM-N3	518.02 ± 58.94	0.38 ± 0.15	10.81 ± 0.26	0.04 ± 3.23	1.40 ± 0.10	5.31 ± 0.26
ISSM-N4	560.75 ± 50.88	0.37 ± 0.07	11.76 ± 0.24	−0.09 ± 1.40	1.93 ± 0.09	6.15 ± 0.15
2012						
NN	149.27 ± 25.78	0.13 ± 0.10	5.80 ± 0.22	−4.32 ± 7.29	0.65 ± 0.09	1.73 ± 0.11
FP	239.32 ± 34.54	0.33 ± 0.11	8.72 ± 0.62	4.85 ± 10.30	2.13 ± 0.43	5.64 ± 0.34
ISSM-N1	225.82 ± 30.42	0.27 ± 0.07	9.43 ± 0.34	1.46 ± 6.38	1.39 ± 0.14	4.94 ± 0.38
ISSM-N2	228.29 ± 32.61	0.38 ± 0.29	9.99 ± 0.50	−1.02 ± 0.84	1.77 ± 0.38	5.78 ± 0.59
ISSM-N3	431.46 ± 26.79	0.52 ± 0.16	10.92 ± 0.61	2.45 ± 8.35	2.19 ± 0.24	5.39 ± 0.39
ISSM-N4	535.55 ± 58.69	0.78 ± 0.13	12.24 ± 0.60	5.91 ± 6.18	2.61 ± 0.42	6.10 ± 0.49
2013						
NN	101.45 ± 39.24	0.16 ± 0.09	5.84 ± 0.15	−1.45 ± 1.34	0.35 ± 0.06	1.80 ± 0.03
FP	140.54 ± 25.20	0.43 ± 0.39	8.67 ± 0.26	−3.70 ± 1.76	0.80 ± 0.20	5.70 ± 0.30
ISSM-N1	135.07 ± 15.68	0.19 ± 0.16	9.66 ± 0.29	−1.00 ± 1.61	0.49 ± 0.16	5.15 ± 0.20
ISSM-N2	129.31 ± 32.24	0.26 ± 0.13	10.15 ± 0.07	−0.79 ± 1.60	0.69 ± 0.24	5.80 ± 0.18
ISSM-N3	256.25 ± 45.61	0.59 ± 0.42	11.14 ± 0.10	−0.62 ± 1.14	0.71 ± 0.10	5.51 ± 0.33
ISSM-N4	304.06 ± 22.27	0.74 ± 0.40	12.34 ± 0.16	0.55 ± 1.68	1.02 ± 0.11	6.19 ± 0.63
Average 2011–2013[a]						
NN[b]	134.63 ± 19.64 d	0.11 ± 0.05 c	5.83 ± 0.04 f	−2.08 ± 1.89 a	0.48 ± 0.07 d	1.75 ± 0.04 d
FP[b]	215.21 ± 19.87 c	0.29 ± 0.13 bc	8.59 ± 0.25 e	0.22 ± 3.96 a	1.45 ± 0.24 b	5.67 ± 0.16 b
ISSM-N1[b]	191.02 ± 19.18 c	0.18 ± 0.06 c	9.45 ± 0.18 d	0.42 ± 2.77 a	0.84 ± 0.08 c	5.04 ± 0.08 c
ISSM-N2[b]	192.49 ± 11.62 c	0.27 ± 0.12 bc	9.98 ± 0.25 c	0.15 ± 0.58 a	1.08 ± 0.12 c	5.76 ± 0.22 ab
ISSM-N3[b]	401.91 ± 23.78 b	0.50 ± 0.16 ab	10.95 ± 0.13 b	0.63 ± 3.51 a	1.43 ± 0.05 b	5.40 ± 0.16 bc
ISSM-N4[b]	466.79 ± 39.18 a	0.68 ± 0.15 a	12.11 ± 0.28 a	2.12 ± 2.57 a	1.85 ± 0.16 a	6.14 ± 0.35 a

[a] Mean ± SD, different lower case letters within the same column for each item indicate significant differences at $P < 0.05$ according to Tukey's multiple range test.

[b] See Table 1 for treatment codes.

Table 3. Two-way ANOVA for the effects of cultivation patterns (P) and cropping year (Y) on CH_4 and N_2O emissions, and rice and wheat grain yields for the three annual rice-wheat rotations of 2011–2014.

Crop season	Factors	df	CH_4 (kg C ha^{-1}) SS	F	P	N_2O (kg N ha^{-1}) SS	F	P	Yield (t ha^{-1}) SS	F	P
Rice	P	5	598 297	94.33	< 0.001	0.77	3.57	< 0.01	85.24	135.48	< 0.001
	Y	2	6633	2.61	0.08	0.03	0.39	0.68	0.01	0.03	0.97
	PY	10	95 740	7.55	< 0.001	0.12	0.28	0.98	0.56	0.44	0.92
	Model	17	1 462 862	67.84	< 0.001	3.39	4.63	< 0.001	283.54	132.54	< 0.001
	Error	54	68 500			2.33			6.80		
Wheat	P	5	22	0.23	0.95	6.47	28.21	< 0.001	51.95	98.58	< 0.001
	Y	2	32	0.84	0.44	0.19	2.09	0.13	0.01	0.06	0.94
	PY	10	239	1.25	0.28	2.75	5.99	< 0.001	0.09	0.09	1.00
	Model	17	438	1.35	0.20	32.31	41.42	< 0.001	156.57	87.39	< 0.001
	Error	54	1033			2.48			5.69		
Rice-Wheat	P	5	596 596	94.05	< 0.001	11.02	22.19	< 0.001	251.28	182.42	< 0.001
	Y	2	6471	2.55	0.09	0.21	1.08	0.35	0.03	0.05	0.95
	PY	10	95 673	7.54	< 0.001	2.63	2.65	< 0.05	0.66	0.24	0.99
	Model	17	1 489 015	69.04	< 0.001	47.83	28.33	< 0.001	806.40	172.19	< 0.001
	Error	54	68 506			5.36			14.88		

df – degrees of freedom; SS – sum of squares; *F*– ratio of mean squares of two independent samples; *P*– index of differences between the control group and the experimental group. $P < 0.05$, $P < 0.01$ and $P < 0.001$ represent significant at the 0.05, 0.01 and 0.001 probability level, respectively.

Table 4. Agricultural management practices for chemical inputs and farm operations and contributions to carbon dioxide equivalents (kg CO_2 eq. ha^{-1} yr^{-1}) in the annual rice-wheat rotations from 2011 to 2014 (chemical inputs and farm operations used in each year were similar except for irrigation water).

Treatment	Chemical inputs (kg ha^{-1})[a]						Farm operations (kg ha^{-1})[c]							
	N fertilizer	P fertilizer	K fertilizer	Herbicide	Insecticide	Fungicide	Irrigation (cm)[b]			Tillage and raking	Crop planting	Farm manure	Crop harvest	Farm machinery production
							2011	2012	2013					
NN	0	180	300	2	18	4	75	80	80	37	1	0	11	135
FP	480	180	300	2	20	4.4	75	80	80	37	1	0	11	147
ISSM-N1	360	180	300	2	20	4.4	50	65	55	37	1	0	11	139
ISSM-N2	432	180	300	2	20	4.4	50	65	55	37	1	0	11	177
ISSM-N3	480	216	360	2	27	6	50	65	55	37	1	2250	11	177
ISSM-N4	600	252	450	2	41	9	50	65	55	37	1	2250	11	275
	Chemical inputs (Ei)						Farm operations (Eo)							
NN	0	132	165	2	338	53	1419	1514	1514	127	12	0	37	36
FP	2288	132	165	2	375	59	1419	1514	1514	127	12	0	37	39
ISSM-N1	1716	132	165	2	375	59	946	1230	1041	127	12	0	37	37
ISSM-N2	2059	132	165	2	375	59	946	1230	1041	127	12	0	37	47
ISSM-N3	2288	158	198	2	506	79	946	1230	1041	127	12	62	37	47
ISSM-N4	2860	185	248	2	768	129	946	1230	1041	127	12	62	37	73

[a] The carbon emission coefficients were 1.3, 0.2 ,0.15, 0.3, 5.1 and 3.9 C cost (kg C eq. kg^{-1} active ingredient) per applied nitrogen fertilizer, phosphorus, potassium, herbicide, insecticide and fungicide, respectively, as referred to in Lal (2004).

[b] The carbon emission coefficient for irrigation was 5.16 C cost (kg C eq. cm^{-1} ha^{-1}) as referred to in Lal (2004).

[c] The carbon emission coefficients were 0.94, 3.2, 0.0075, 0.94 and 0.0725 C cost (kg C eq. kg^{-1} active ingredient) for tillage and raking, crop planting, per farm manure application, harvesting, spraying and threshing, respectively, as referred to in Lal (2004).

See Table 1 for treatment codes.

Table 5. Mean net global warming potentials (NGWPs) and greenhouse gas intensity (GHGI) over the three annual cycles of the 2011 rice season–2014 wheat season.

Treatment	CH_4	N_2O	Ei	Eo	SOCSR	NGWP[a]	Grain yield	GHGI[b]
			kg CO_2 eq. ha^{-1} yr^{-1}				t ha^{-1} yr^{-1}	kg CO_2 eq. t^{-1} grain
NN	4418 ± 628 d[c]	276 ± 29 d	690	1694	−792 ± 327 c	7871 ± 646 d	7.58 ± 0.04 d	1038 ± 85 b
FP	7181 ± 766 c	816 ± 55 b	3021	1697	1170 ± 396 ab	11 545 ± 505 c	14.26 ± 0.36 c	810 ± 23 c
ISSM-N1	6381 ± 633 c	479 ± 62 c	2449	1285	491 ± 435 b	10 104 ± 930 c	14.50 ± 0.14 c	697 ± 63 d
ISSM-N2	6421 ± 379 c	633 ± 97 c	2792	1295	709 ± 193 ab	10 433 ± 516 c	15.74 ± 0.44 b	664 ± 49 d
ISSM-N3	13 418 ± 744 b	906 ± 87 b	3231	1357	1383 ± 503 a	17 529 ± 688 b	16.36 ± 0.18 b	1071 ± 33 ab
ISSM-N4	15 630 ± 1246 a	1188 ± 65 a	4192	1383	1545 ± 348 a	20 847 ± 1289 a	18.26 ± 0.46 a	1143 ± 84 a

[a] NGWP (kg CO_2 eq. ha^{-1} yr^{-1}) = GWP (CH_4 + N_2O + Ei + Eo) – GWP (SOCSR), Ei (agrochemical inputs), Eo (farm operations), SOCSR (SOC sequestration rate)

[b] GHGI (kg CO_2 eq. t^{-1} grain) = NGWP/grain yields.

[c] Different lower case letters within the same column for each item indicate significant differences at $P < 0.05$ based on Tukey's multiple range tests.

See Table 1 for treatment codes.

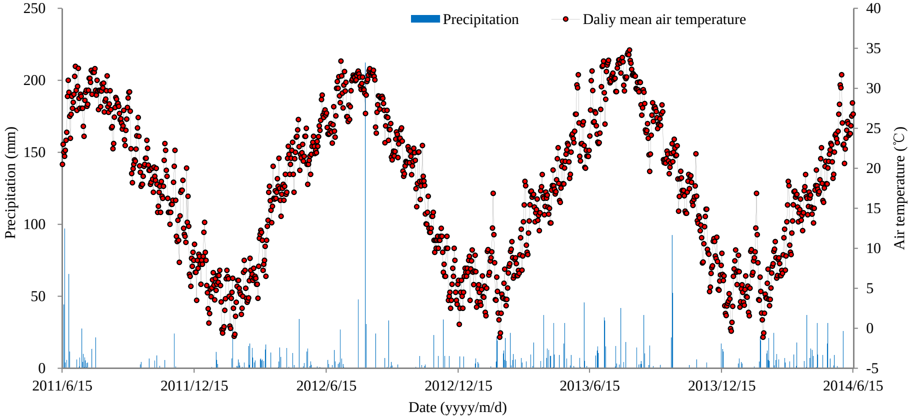

Figure 1. Daily mean air temperature and precipitation during the rice-wheat rotation in 2011–2014 in Changshu, China.

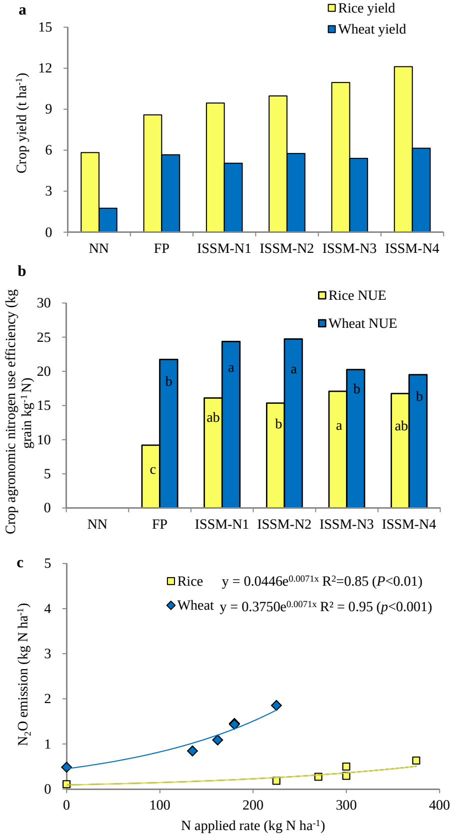

Figure 2. (a) Rice and wheat grain yield, **(b)** agronomic nitrogen use efficiency (NUE) and **(c)** relationship between nitrous oxide (N_2O) emissions and the total N rate for rice and wheat crops in 2011–2014 in Changshu, China. Different letters indicate a significant difference between treatments ($p < 0.05$). See Table 1 for treatment codes.

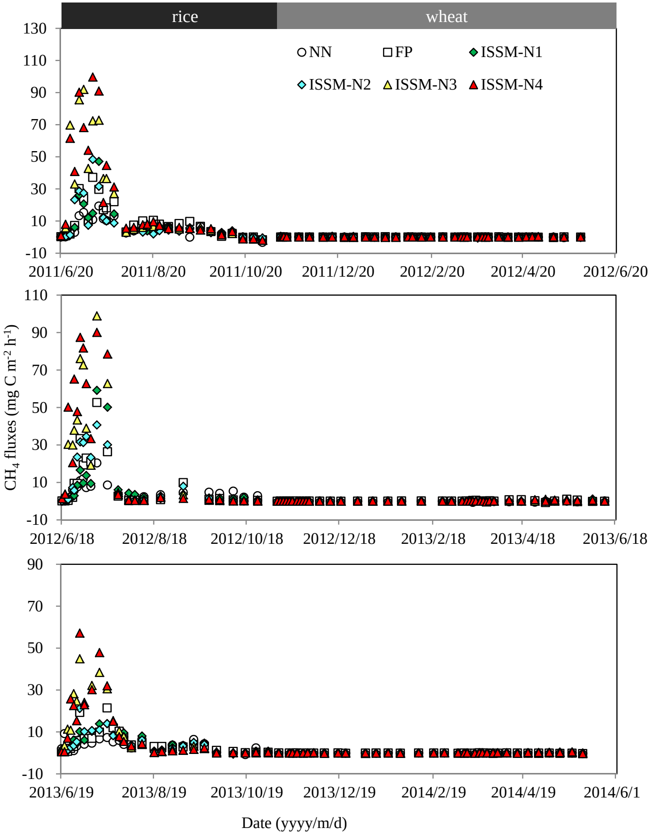

Figure 3. Seasonal variation of methane (CH_4) fluxes from the rice-wheat rotation cropping systems from 2011 to 2014. The black and gray part in figure separates different grain growth periods. See Table 1 for treatment codes.

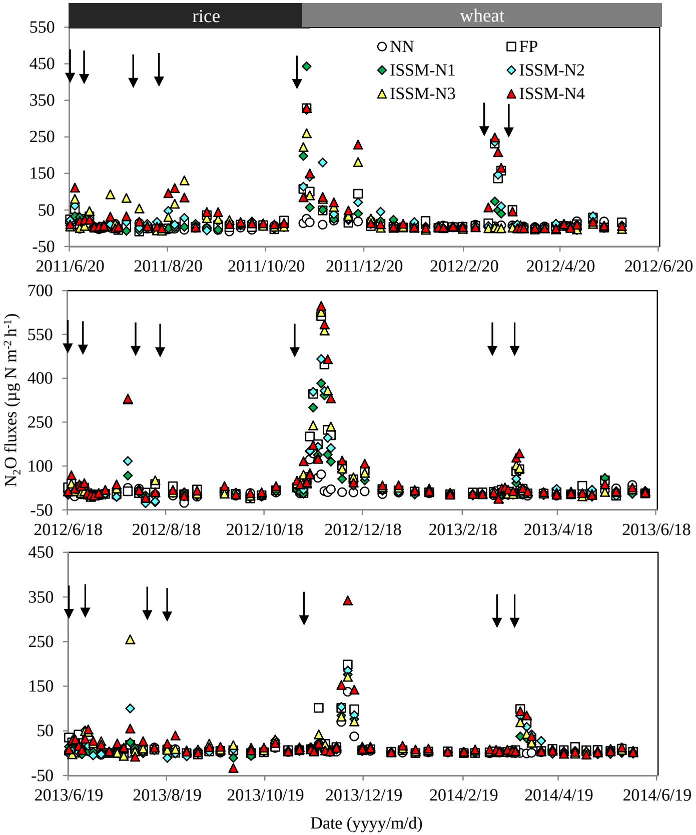

Figure 4. Seasonal variation of nitrous oxide (N_2O) fluxes from rice-wheat rotation cropping systems in three annual cycles over the period 2011–2014. The black and gray part in the figure separates different growth periods. See Table 1 for treatment codes. The solid arrows indicate fertilization.

BGD
12, 18883–18911, 2015

Title Page
Abstract | Introduction
Conclusions | References
Tables | Figures
◀| |▶
◀ ▶
Back | Close
Full Screen / Esc
Printer-friendly Version
Interactive Discussion