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Larval development and settling of *Macoma balthica* in a large-scale mesocosm experiment at different fCO_2 levels

A. Jansson^{1,2}, S. Lischka³, T. Boxhammer³, K. G. Schulz^{3,4}, and J. Norkko²

Received: 30 November 2015 – Accepted: 4 December 2015 – Published: 21 December 2015

Correspondence to: A. Jansson (jansson.anna.e@gmail.com)

Published by Copernicus Publications on behalf of the European Geosciences Union.

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¹Environmental and Marine Biology, Faculty of Science and Engineering, Åbo Akademi University, Åbo, Finland

²Tvärminne Zoological Station, University of Helsinki, Hanko, Finland

³GEOMAR Helmholtz Centre for Ocean Research Kiel, Düsternbrooker Weg 20, 24105 Kiel, Germany

⁴Centre for Coastal Biogeochemistry, School of Environment, Science and Engineering, Southern Cross University, P.O. Box 157, Lismore, NSW, Australia

Anthropogenic carbon dioxide (CO₂) emissions are causing severe changes in the global inorganic carbon balance of the oceans. Associated ocean acidification is expected to impose a major threat to marine ecosystems worldwide, and it is also expected to be amplified in the Baltic Sea where the system is already at present exposed to relatively large natural seasonal and diel pH fluctuations. The response of organisms to future ocean acidification has primarily been studied in single-species experiments, whereas the knowledge of community-wide responses is still limited. To study responses of the Baltic Sea pelagic community to a range of future CO₂-scenarios, six ~ 55 m³ pelagic mesocosms were deployed in the northern Baltic Sea in June 2012. In this specific study we focused on the tolerance, development and subsequent settlement process of the larvae of the benthic key-species Macoma balthica when exposed to different levels of future CO2. We found that the settling of M. balthica was delayed along the increasing CO₂ gradient of the mesocosms. Also, when exposed to increasing CO₂ levels larvae settled at a larger size, indicating a developmental delay. With ongoing climate change, both the frequency and extent of regularly occurring high CO₂ conditions is likely to increase, and a permanent pH decrease will likely occur. The strong impact of increasing CO2 levels on early-stage bivalves is alarming as these stages are crucial for sustaining viable populations, and a failure in their recruitment would ultimately lead to negative effects on the population.

Introduction

Anthropogenic CO₂-emissions are causing severe changes in the oceans (Feely et al., 2004). Future ocean acidification (OA), which includes changes in the inorganic carbon balance of the seawater coupled with a decrease in pH, is occurring at a rate faster than experienced before in the geologic past (Hönisch et al., 2012), and is expected to impose a major threat to marine ecosystems worldwide (Fabry et al., 2008;

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Orr et al., 2005). The sea surface pH is estimated to decrease by 0.4 units in the global open oceans by the year 2100 (Caldeira and Wickett, 2003), whereas many coastal areas already at present experience large pH fluctuations reaching to considerably lower pH levels than predicted for the near future (Blackford and Gilbert, 2007; Johnson et al., 2013). The multiple environmental stressors impacting coastal areas and the local processes that impact watersheds make the precise modelling of future pH levels exceedingly challenging for these areas (Borges and Gypens, 2010; Duarte et al., 2013).

The majority of studies investigating the biological effects of future CO₂ levels have focused on its impacts on calcifying species and on pelagic primary producers. Pelagic calcifiers such as bivalve early life-stages are generally considered susceptible to increasing CO₂ level (Kurihara, 2008; Dupont and Thorndyke, 2009), with a range of observed (mostly negative) impacts on development, survival and growth of larval stages as consequences of the CO₂ increase (Gazeau et al., 2013). Also the settling and survival of post-larvae is impacted by the changes in the water chemistry (Green et al., 2004, 2009; Clements and Hunt, 2014). The response of organisms to future CO₂ levels has traditionally been studied in experiments focusing on single species, and the community-wide responses are still not well known. However, to understand complex, system-wide responses that take into account ecological processes such as competition, predation and the effect of/on different trophic levels, several species interactions need to be tested simultaneously. In mesocosms, the natural community can be maintained to a high degree, and organismal performance can be measured in near-natural surroundings (Riebesell et al., 2010). Mesocosm studies have the additional advantages of allowing experimental manipulation of environmental factors such as CO2, possibility for replication, and repeated sampling of the closed study systems over long experimental duration. The disadvantages of limited ecosystem realism that arise from the exclusion of factors such as currents and large predators, which impact the natural succession and dispersion patterns of the species, nevertheless have to be accounted for when interpreting the results.

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In the Baltic Sea a drop in pH of 0.5 units is estimated for the surface waters within this century (Hjalmarsson et al., 2008; Omstedt et al., 2012). Similar to coastal and estuarine areas (Duarte et al., 2013), however, the natural pH variability in the Baltic Sea is large and regularly exceeds the estimates made for the near-future (Omstedt et al., 5 2009; Jansson et al., 2013; Melzner et al., 2012). For example, during the summer season pH changes of nearly one unit per day driven by changes in primary production and respiration are common in the shallow coastal areas of the northern Baltic Proper (pers. obs.). Yet, ocean acidification is likely to increase the pH fluctuations, making the occasionally experienced extreme pH levels even more pronounced, further expanding the pH range which the Baltic species are exposed to (Thomas and Schneider, 1999; Melzner et al., 2012; Omstedt et al., 2010). A key species in the Baltic Sea softbottom communities, the bivalve Macoma balthica (L.), is experiencing variable conditions throughout its life-cycle. During the larval phase, it is exposed to large pelagic diel pH-fluctuations (Jansson et al., 2013; Almen et al., 2014) followed by the harsh reducing conditions of the sedimentary system when settling into the benthic environment (Woodin et al., 1998). The tolerance of M. balthica to low pH conditions has so far been studied in aquarium experiments of different types and durations (Jansson et al., 2013; van Colen et al., 2012), which have shown negative effects on the early-stage bivalves. In such experiments, however, the potential impact of future environmental changes on e.g. the settlement process is challenging to study.

The aim of this large-scale pelagic mesocosm experiment was to study the responses of the Baltic Sea pelagic community to different future fCO₂-scenarios. In this specific study we wanted to explicitly shed light on (1) the tolerance and development of M. balthica larvae and (2) the subsequent settling of the post-larvae, when exposed to different levels of future CO₂ in their natural surroundings. Based on the results of our previous experiments (Jansson et al., 2013; van Colen et al., 2015), we predicted the growth of the larvae to decrease along the increasing fCO_2 gradient and the survival and settling to be negatively impacted by the fCO_2 increase.

The study species

The infaunal bivalve M. balthica is abundant throughout the Baltic Sea, often dominating biomass in soft sediments from organic mud to sandy bottoms from the very shallow down to 190 m depth (Bonsdorff, 2006; Elmgren et al., 1986; Segerstråle, 1960). The spawning of M. balthica occurs when water temperature has reached approximately 7°C (Caddy, 1967). The planktonic life stage (ca. 6 weeks) ends when the individual has reached a sufficient size and developmental stage (including increased mobility of the foot) to metamorphose and settle to the seafloor (Caddy, 1969). A majority of the very newly settled bivalves encountered in the Baltic Sea have a size of 250-300 µm (Ankar, 1980; Elmgren et al., 1986; Olafsson, 1989). Peak settling in the northern parts of the Baltic Sea typically occurs in July. During the pelagic larval phase, abundances of up to 12 000 larvae m⁻³ are measured in the Baltic Sea, with a settling population of around 30 000 m⁻² each year, at peak settling even up to 300 000 m⁻² (Ankar, 1980; Bonsdorff et al., 1995; Elmgren et al., 1986). M. balthica is an important prey organism, and has a central role in sediment reworking and bioturbation, contributing to the overall health and functioning of the benthic ecosystem (Michaud et al., 2006). In the species-poor northern Baltic Sea, this species is essential to the functioning of the benthic ecosystem through these key processes (Norkko et al., 2013; Villnäs et al., 2012).

Experimental set-up

Six pelagic mesocosms (KOSMOS, Riebesell et al., 2013a) of ~ 55 m³ were deployed in the western Gulf of Finland (59°51.5' N, 23°15.5' E) on 12 June 2012 to study responses of the Baltic Sea plankton community to increased fugacity of carbon dioxide (fCO₂). The mesocosm bags were lowered down to a depth of 17 m to enclose the natural plankton community, excluding organisms larger than 3 mm by a mesh installed at Discussion

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the top and bottom of the cylindrical bags. With the bags fully submerged below the sea surface, water and organisms inside the bags could exchange with the surrounding water mass for five days before closing the mesocosms on 17 June (day -5, 5 days before CO₂ manipulation). To seal the bottom of each mesocosm, a two meter long sediment trap funnel collecting settling particles and organisms was installed by divers to replace the 3 mm mesh. The top end of the bags was simultaneously pulled above the sea surface to fully isolate the enclosed water bodies. Bubbling the systems with compressed air for three and a half minutes right after closure destroyed the halocline present inside the bags. The mesocosms were manipulated with filtered (50 µm), CO₂-saturated seawater as described by Riebesell et al. (2013) on four consecutive days (day 0-3) to establish a range of four fCO₂ target treatments (600–1650 µatm) and two ambient blind manipulated mesocosms (Table 1). On day 15 fCO₂ was readjusted inside the treated mesocosms to counteract outgassing of CO2. For a more detailed description of the experimental set-up, manipulations and maintenance of the mesocosms please see Paul et al. (2015).

Sampling the mesocosms

2.3.1 Water parameters

CTD profiles were measured daily with a handheld self-logging CTD60M probe (Sea and Sun Technology) from 0.3 down to 18 m (mesocosms) and to 30 m (surrounding bayc) with sensors for salinity, temperature, dissolved oxygen, PAR (photosynthetic active radiation) and pH. Details on the sensors and their accuracy are described in Schulz and Riebesell (2013). Depth-integrated water samples (IWS, HYDRO-BIOS Kiel) were collected regularly (daily to every other day, see Paul et al., 2015) from all mesocosms and the surrounding water body to measure e.g. total pH (pHT), total alkalinity (TA) and dissolved inorganic carbon (DIC) for determining the inorganic carbon components, and chlorophyll a to follow the development of the phytoplankton bloom. pHT was determined by analyzing samples with a Cary 100 (Varian) spectrophotome-

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ter (Dickson et al., 2007). The details of the procedure (fCO₂ was calculated from measured DIC and pHT) are described in Paul et al. (2015). CTD pH measurements were corrected to pH on the total scale by linear correlations of mean water column potentiometric pH measurements to spectrophotometric pHT measurements. Exact details of 5 all sampling procedures, equipment used and sample analyses are described in Paul et al. (2015), Riebesell et al. (2013a) and Schulz et al. (2013).

2.3.2 Water column: mesozooplankton sampling and quantification of M. balthica larvae

Mesozooplankton samples from the six mesocosms were taken with an Apstein net of 17 cm diameter and 100 µm mesh size by towing the net vertically from 17 m depth to the sea surface. Net hauls were taken from the mesocosms on eleven sampling days: prior to the first CO₂ addition (days -3, -2, -1), at the day of the first CO₂ addition (day 0), and after the first CO₂ addition in a seven day rhythm (days 3, 10, 17, 24, 31, 38, 45). Mesozooplankton samples were preserved in 70% ethanol. The larvae of M. balthica were counted in the whole sample under a stereo microscope (WILD M3B). For size range determination, on average 70 individuals were measured from each mesocosm on days 0 and 10. The individuals were photographed using a dissecting microscope connected to a Nikon DS-Fi2 camera system, and sizes were determined by measuring shell lengths using the Nikon DS camera interface. Zooplankton abundance was calculated as individuals per cubic meter, assuming 100 % filtering efficiency of the net. For more details on mesozooplankton sampling and processing see Lischka et al. (2015).

2.3.3 Sediment traps: collection of material, subsampling and quantification of settling M. balthica

The sediment traps were emptied every second day using a gentle vacuum to pump the samples through a silicon tube into sampling flasks at the sea surface (for more details

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see Boxhammer et al., 2015). Subsamples of 20 mL were taken with a pipette of the homogeneously mixed samples (on average 2.5 L) and preserved in 4 % buffered formalin for quantification and size determination of settling bivalves. Abundance and size range determinations of settled bivalves were made on 3 replicates of 1 mL subsamples. *M. balthica* collected in the sediment traps included settling individuals as well as individuals that died in the water column or in the sediment trap after settling. However, the gaping shells of individuals that were dead at the time of sampling were identified in the preserved samples and such individuals were not counted. Individuals that were assessed to be living at the time of sampling were counted and photographed using a dissecting microscope connected to a Nikon DS-Fi2 camera system. During the main settling period (days 11, 13, 15 and 17) on average 35 individuals were measured from each mesocosm. Sizes were determined by measuring shell lengths using the Nikon DS camera interface.

2.4 Numerical analysis

The abundance of bivalve larvae in the water column of each mesocosm over time was compared by calculating a rate of change between each sampling day and comparing the timing of decreasing abundances. This was done by calculating Spearman correlation ranks for each time point. To analyse the differences in post-larval settling between the mesocosms, we performed a chi-square test to compare the cumulative abundances of settling individuals on days 9, 11, 13, 15, 17 and 19. Graphical post-hoc tests were performed to identify differences between mesocosms.

The sizes of both the larvae in the water column and the post-larvae in the settling traps in the different fCO_2 levels were compared by a linear regression model. To standardize the comparisons, they were conducted on average sizes of a batch of individuals measured in each mesocosm. The residuals of the regressions adhered to the assumption of normality. All analyses were performed in the software R (version 3.0.2; R Core Team 2012). The differences were considered significant at p < 0.05 for all tests. Data are presented as means \pm SD.

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3.1 Abjotic conditions in the mesocosms

Water temperature varied from 8 to 16 °C during the experiment, following the natural conditions in the bay. Salinity was on average 5.7 and measured total alkalinity on average 1550 mmol kg⁻¹ at the closing of the mesocosms. Both parameters remained fairly constant during the experiment in all mesocosms (Paul et al., 2015, this issue). Initial pHT after closing of the mesocosms and before the CO₂-manipulations was ca. 8.2 in the mesocosms and the bay. Average pHT levels and other parameters of each mesocosm over the course of the experiment are shown in Table 1.

3.2 Larval abundance

After the closing of the mesocosms (day -3 to -2), some unexplained variation was found in the abundance of bivalve larvae (Fig. 1). On day 0, however, the abundances in the water column were relatively similar within the mesocosms (5522–5936 ind. m⁻³), except in the ambient mesocosm M1 where the abundance decreased earliest and with a steep slope. During the first week after the CO_2 -manipulation, by day 10, the larval abundance had decreased strongest in the ambient mesocosms M1 and M5, with > 80% decrease in abundance in comparison to the 35–50% decrease in the two highest fCO_2 mesocosms M3 and M8 (Spearman r = -0.83, p < 0.05). Consequently, on day 10 the highest abundance was measured in the highest fCO_2 mesocosm M8 (3194 ind. m⁻³) and the lowest abundances in the ambient mesocosms M1 and M5 (545 resp. 1064 ind. m⁻³). A strong decrease in abundance (> 85%) occurred a week later (day 10 to 17) in all the high fCO_2 mesocosms M3, M8, M6 and M7, with up to a 93% decrease found in M8 (Spearman r = 0.94, p < 0.05). From day 17 onwards, the abundances were low in all of the mesocosms (Fig. 1).

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The abundances of settling individuals differed significantly between mesocosms and sampling days of the main settling period (days 9–17, chi-square χ^2 = 1168.588, df = 25, p < 0.001). The graphical post-hoc tests showed three distinct settling peaks of M. balthica. In the ambient and near-ambient (< 500 μ atm) fCO $_2$ mesocosms M1, M5 and M7, a large increase in the abundance of settling individuals was found between days 9–13, with 71, 74 and 54% of all the individuals having settled by day 13. In comparison, 39 and 47% of the individuals had settled during that time period in the two highest (1000–1231 μ atm) fCO $_2$ mesocosms M8 and M3 (Fig. 2a and b). In the 800–1000 μ atm fCO $_2$ mesocosm M6 and M3, a smaller settling event occurred on days 11–15 and in the highest fCO $_2$ mesocosm M8 the settling peaked on day 17, where after the number of settling individuals soon ceased in all mesocosms. On average 6130 \pm 240 individuals settled in the mesocosms during the course of the experiment, with the exception of M3 where only ca. 4850 individuals settled (Fig. 2b).

3.4 Larval sizes in the water column

On day 0, larval size in the water column was on average $287\pm23\,\mu\text{m}$ with no difference found between the mesocosms. After 10 days of exposure to different $f\text{CO}_2$ levels, the average size of the larvae in the water column $(0-17\,\text{m})$ varied from 286 to 313 μm , increasing significantly along the increasing $f\text{CO}_2$ gradient ($R^2=0.78$, F=14.47, p=0.019, Fig. 3) with ca. 10 % larger larvae still in the water column in the highest $f\text{CO}_2$ mesocosms M3 and M8.

3.5 The sizes of settling individuals

On average > 80% of the individuals settled in the mesocosms during days 11 to 17. No significant differences were found in the sizes of the settling individuals in the different fCO_2 levels at any of these investigated time points (Fig. 4). On days 11 and

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13 the average size within the mesocosms varied between 285 to 303 μ m, and on days 15 and 17 the average size varied between 293 to 317 μ m.

4 Discussion

In this study we investigated the effects of different future CO_2 scenarios on the larval development and settling of a Baltic Sea benthic key-species M. balthica in a large-scale mesocosm setting. We found that M. balthica settled later along the increasing fCO_2 gradient of the mesocosms. Moreover, an indication that M. balthica post-larvae settled at a larger size in the high fCO_2 treatments was also observed, possibly indicating that at increasing fCO_2 a sufficient mass for settling is not reached until a larger shell length has been attained.

During the week after first CO_2 manipulation (day 3 to day 10) settling of M. balthica occurred faster in the ambient and middle fCO_2 mesocosms M1, M5 and M7 (365–497 μ atm) than in the higher fCO_2 mesocosms. Consequently, the main settling peak occurred ca. 6 days earlier in these mesocosms (< 500 μ atm). When comparing the sizes of the larvae, we found that the ones remaining in the water column on day 10 had an average size of 290 μ m in the ambient mesocosms M1 and M5, whereas in the other mesocosms M7, M6, M3 and M8 ($fCO_2 > 400 \mu$ atm), the sizes of the remaining larvae were 300–315 μ m. We hypothesise that in the ambient fCO_2 the bivalves settled at the expected size (< 300 μ m), and thus only the smaller larvae remained in the upper water column when the settling was reaching its peak. In the high fCO_2 treatments the development of the M. balthica larvae might have been compromised and/or delayed as on day 10, despite being relatively large (> 300 μ m), a large part of the bivalves remained in the upper water column without initiating settlement.

The observed inconsistency between the growth and settling of the early-stage bivalves can be explained by proximate factors that regulate settling. For successful metamorphosis and settling from the planktonic phase to the benthos, the individuals need to reach a sufficient size or weight and developmental stage, including increased **BGD**

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mobility/appearance of the foot (Caddy, 1969; Drent, 2002). Shell growth alone, the growth measure used in our experiment as in many other studies, is not automatically reflecting the overall biomass production and developmental stage of the organism (Lewis and Cerrato, 1997; Wood et al., 2008). In undersaturated conditions, calcification of the shell might be compromised so that even though shell length reaches its typical size for settling, shell thickness is reduced. This could be a factor that restricts the gaining of necessary mass to settle to the sea floor (Waldbusser et al., 2010). During the entire experiment, undersaturation with respect to aragonite occurred in all mesocosms apart from the two ambient mesocosms, and the three highest fCO₂ treatments were also undersaturated with respect to calcite (Table 1). It is also likely that at decreased pHT levels shell growth was occurring at the cost of tissue development and biomass increase. Unfortunately we were not able to measure soft tissue weight of collected larvae due to the very small size. Larvae that stay longer in the water column, e.g. due to slower growth or delayed development, face a higher risk of predation. The population dynamics of a bivalve species is largely dependent on successful settlement and recruitment of the post-larvae, and dispersal of larval and post-larval stages (Pedersen et al., 2008; Pineda et al., 2009; Valanko et al., 2010), and a reduced performance of the early-life stages, as found in the present study, is thus alarming. Some other important factors that impact the settlement process, but cannot be mimicked in this mesocosm setup include, e.g., sediment type and quality, cues from adult conspecifics and water movements that can prevent or facilitate the settlement process (Woodin et al., 1998).

In a previous experiment conducted with newly hatched larvae (ca. 150 µm) from the same bay (Jansson et al., 2013), both the growth and survival of the larvae were found to be negatively impacted by decreasing pH. Other typical consequences of pH decrease for early-stage bivalves are e.g. delayed and/or abnormal development (Talmage and Gobler, 2010; Crim et al., 2011; Kurihara et al., 2008), reduced calcification (Miller et al., 2009) and higher mortality (Crim et al., 2011; Talmage and Gobler, 2009; van Colen et al., 2012). The settling of post-larvae to the seafloor may be im-

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pacted by the changes in the water chemistry created by CO₂ increase (Green et al., 2004; Clements and Hunt, 2014; Cigliano et al., 2010). The major part of ocean acidification research has been conducted by studying the response of single species, with a few studies focusing on the interactions between a small number of species, whereas studies on intact communities have so far only rarely been conducted (but see e.g. work done at CO₂ vents by Hall-Spencer et al., 2008 or Kroeker et al., 2011 and previous/other mesocosm studies by Christen et al., 2013; Riebesell et al., 2013b). For species such as M. balthica, a mesocosm setting provides an excellent platform to study the development and succession of pelagic early-life stages resulting in recruitment into the benthic system, which cannot be studied in a simple, small-scale aquarium experiment. The direct and indirect factors that essentially impact the early life success of a bivalve, e.g. natural food quality and quantity, can be incorporated in a mesocosm setting in a more comprehensive way. In the case of future ocean acidification, potential changes in phytoplankton dynamics due to increased CO2 levels are likely to have consequences for the other trophic levels. The growth of nanoplankton and diatom species (< 20 µm), which are the main food particles of larval bivalves (Bos et al., 2006), has been shown to benefit from changing CO₂ conditions (e.g. Engel et al., 2007; Feng et al., 2009; Meakin and Wyman, 2011; but see also e.g. Tortell et al., 2002), potentially impacting the capacity of the larvae to survive in a changing environment via consequences in their energy balance. In this study, no significant changes were detected in the phytoplankton abundance or the total chlorophyll a concentration within the mesocosms during the main occurrence of M. balthica larvae in the water column (until days 10 and 17). An increase in the abundance of phytoplankton and Chl a concentration in the highest fCO₂ mesocosms was, however, found later on during the experiment (day 16 onwards; Crawfurd et al., 2015; Paul et al., 2015).

The Baltic Sea is a unique system to study future ocean acidification. Already at present, large pH fluctuations that occur seasonally in the northern Baltic Sea in the shallow coastal areas, primarily due to changes in productivity (Schneider et al., 2003; Thomas and Schneider, 1999), result in high pH values of up to 8.4 during daytime and

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low pH values such as 7.4 during respiration at night (pers. obs.). For areas such as this, accurate modelling of the future pH change is generally challenging. Yet, future ocean acidification is predicted to permanently decrease the pH and thus shift the pH range the organisms are exposed to towards lower values (Omstedt et al., 2010). In our study we found negative effects of increasing CO₂ levels on the settling and early development of M. balthica. The strong impact on the success of these earlystage bivalves is alarming as these stages are crucial for sustaining viable populations. A failure in their recruitment would ultimately lead to negative effects on the population, and considering the key role M. balthica has in the Baltic Sea, also for the functioning and resilience of the benthic ecosystem

Acknowledgements. The authors would like to thank the whole KOSMOS Team for deployment and maintenance of the KOSMOS infrastructure, in particular, we would like to thank Ulf Riebesell, Andrea Ludwig and Jan Czerny. Alf Norkko, Martin Seltmann and Judi Hewitt contributed to the manuscript by providing valuable comments. We would also like to sincerely thank the Tvärminne Zoological Station for the excellent working facilities and warm hospitality. We also gratefully acknowledge the captain and crew of R/V ALKOR (AL394 and AL397) for their work in transporting, deploying and recovering the mesocosms. This collaborative project was funded by BMBF projects BIOACID II (FKZ 03F06550) and SOPRAN Phase II (FKZ 03F0611).

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Table 1. Carbonate system parameters in the mesocosms during the experiment (average values on days 0-43, except for aragonite and calcite saturation states where the values given are averages of days 0-17).

	M1	M5	M7	M6	МЗ	M8	Bay
Target fCO ₂ (µatm)	ambient/control	ambient/control	600	950	1300	1650	ambient
fCO_2 (µatm)	365	368	497	821	1007	1231	417
pHT	7.89	7.89	7.77	7.59	7.52	7.44	7.88
Ω aragonite	1.07	1.06	0.77	0.47	0.39	0.33	1.19
Ω calcite	1.92	1.91	1.39	0.84	0.71	0.59	2.14

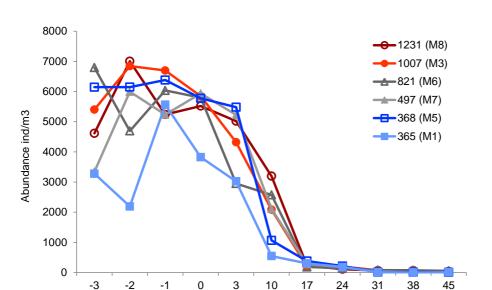


Figure 1. Larval abundance in the water column of the individual mesocosms over time.

Day

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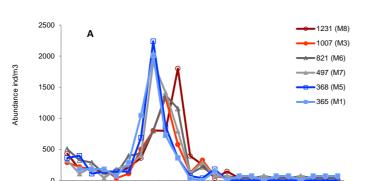
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11 15 19 23 27 31 35 39

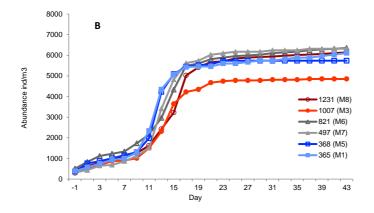


Figure 2. (a) The abundance of settling individuals per cubic meter water mass enclosed in the different mesocosms over the course of the experiment. **(b)** The cumulative abundance of settled *M. balthica* per cubic meter of individual mescosom volume.

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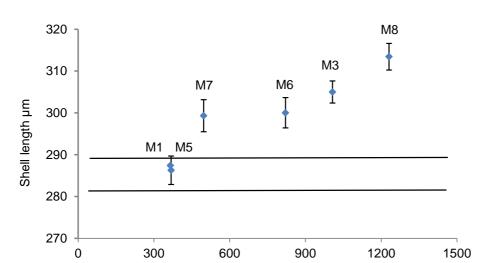


Figure 3. Larval sizes in different fCO_2 levels at day 10. Data is presented as means \pm SE, n = ca. 70 individuals. The horizontal lines indicate the range of average larval sizes on day 0.

fCO2

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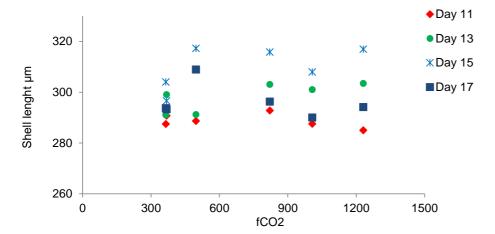


Figure 4. Sizes of the settling individuals exposed to different fCO_2 levels on days 11, 13, 15 and 17. Data is presented as means, n = ca. 35 at each data point. For clarity, SD are not shown.

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