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Changes in soil carbon, nitrogen and phosphorus due to land-use changes in Brazil

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Abstract

In this paper soil carbon, nitrogen and phosphorus concentrations and related elemental ratios, as well as and nitrogen and phosphorus stocks were investigated in 17 paired sites and in a regional survey encompassing more than 100 pasture soils in the Cerrado, Atlantic Forest, and Pampa, the three important biomes of Brazil. In the paired sites, elemental soil concentrations and stocks were determined in native vegetation, pastures and crop-livestock systems (CPS). Overall, there were significant differences in soil element concentrations and ratios between different land uses, especially in the surface soil layers. Carbon and nitrogen contents were lower, while phosphorus contents were higher in the pasture and CPS soils than in forest soils. Additionally, soil stoichiometry has changed with changes in land use. The soil C : N ratio was lower in the forest than in the pasture and CPS soils; and the carbon and nitrogen to available phosphorus ratio (P_{ME}) decreased from the forest to the pasture to the CPS soils. The average native vegetation soil nitrogen stocks at 0–10, 0–30 and 0–60 cm soil depth layers were equal to approximately 2.3, 5.2, 7.3 Mg ha^{-1}, respectively. In the paired sites, nitrogen loss in the CPS systems and pasture soils were similar and equal to 0.6, 1.3 and 1.5 Mg ha^{-1} at 0–10, 0–30 and 0–60 cm soil depths, respectively. In the regional pasture soil survey, nitrogen soil stocks at 0–10 and 0–30 soil layers were equal to 1.6 and 3.9 Mg ha^{-1}, respectively, and lower than the stocks found in the native vegetation of paired sites. On the other hand, the soil phosphorus stocks were higher in the CPS and pasture of the paired sites than in the soil of the original vegetation. The original vegetation soil phosphorus stocks were equal to 11, 22, and 43 kg ha^{-1} in the three soil depths, respectively. The soil phosphorus stocks increased in the CPS systems to 30, 50, and 63 kg ha^{-1}, respectively, and in the pasture pair sites to 22, 47, and 68 kg ha^{-1}, respectively. In the regional pasture survey, the soil phosphorus stocks were lower than in the native vegetation, and equal to 9 and 15 kg ha^{-1} at 0–10 and 0–30 depth layer. The findings of this paper illustrate that land-use changes that are currently common in Brazil alter soil concentrations, stocks

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and elemental ratios of carbon, nitrogen and phosphorus. These changes could have an impact on the subsequent vegetation, decreasing soil carbon, increasing nitrogen limitation, but alleviating soil phosphorus deficiency.

1 Introduction

Based on a regional scale analysis of several paired sites in Brazil, Assad et al. (2013) recently showed a decrease in soil carbon stocks of pasture-livestock systems compared with carbon stocks of the native vegetation of the area. This finding is supported by several other studies that showed a decrease of soil carbon stocks with cultivation (Davidson and Ackerman, 1993; Amundson, 2001; Guo and Gifford, 2002; Ogle et al., 2005; Baker et al., 2007; Don et al., 2011; Eclisa et al., 2012; Mello et al., 2014). On the other hand, there is also a rich body of literature showing that cultivated soil carbon stocks become neutral or may increase compared to the soil stocks under original vegetation (Guo and Gifford, 2002; Ogle et al., 2005; Zinn et al., 2005; Braz et al., 2012; Mello et al., 2014). The carbon gain with cultivation seems to be faster and higher when agricultural practices like no till, green manure, crop rotation and crop-livestock systems are adopted (Sá et al., 2001; Ogle et al., 2005; Zinn et al., 2005; Bayer et al., 2006; Baker et al., 2007).

On the other hand, there are few global or regional studies considering how land-use changes affect nitrogen and phosphorus soil contents. Plot-level studies have reported a decrease in soil nitrogen stocks with cultivation in several N-fertilized areas of Brazil and under different cropping systems (Lima et al., 2011; Fracetto et al., 2012; Barros et al., 2013; Sacramento et al., 2013; Cardoso et al., 2010; Silva et al., 2011; Guareschi et al., 2012; Sisti et al., 2004; Santana et al., 2013; Sá et al., 2013). The same trend has been observed in Chernozen soils in Russia and in prairie soils of Wisconsin in the US (Mikhailova et al., 2000; Kucharik et al., 2001).

In unfertilized pasture soils of Brazil, nitrogen availability decreased as the age of pastures increased. In theses soils, there was an inversion in relation to forest soils, and an ammonium dominance over nitrate was observed, followed by lower mineralization and nitrification rates that in turn were followed by lower emissions of N_2O (Davidson et al., 2000; Erickson et al., 2001; Wick et al., 2005; Neill et al., 2005; Cerri et al., 2006; Carmo et al., 2012).

Therefore, it seems that receiving N-fertilizer inputs or not, agro-ecosystem nitrogen losses via leaching, gaseous forms and harvesting exports are higher than N-inputs resulting in decreased soil nitrogen stocks.

Phosphorus is particularly important in the tropics due to the ability of acidic tropical soils to fix phosphorus on oxides and clay minerals rendering them unavailable to plants (Uehara and Gillman, 1981; Sanchez et al., 1982; Oberson et al., 2001; Numata et al., 2007; Gama-Rodriguez et al., 2014). As a consequence, tropical wild plants develop a series of strategies to cope with soil acidity and the low phosphorus concentration (Fujii, 2014). This widespread lack of phosphorus in tropical soils also affects crops, consequently there is a rich body of literature on phosphorus dynamics in tropical soils and how land-use changes result in different phosphorus fractions (e.g. Garcia-Montiel et al., 2000; Oberson et al., 2001; Townsend et al., 2002; Numata et al., 2007; Pavinatto et al., 2009; Fonte et al., 2014; Fujii, 2014), but there have been considerably fewer studies on changes in soil stocks of phosphorus with cultivation.

The P-adsorption by the clay fraction in tropical soils (Oberson et al., 2001), as well as the fact that phosphorus does not have a gaseous phase like nitrogen, renders phosphorous less mobile in the soil-plant-atmosphere system than nitrogen. One consequence of this lower phosphorus mobility throughout the soil profile is that when P-fertilizers are applied, they tend to increase soil phosphorus concentration on the soil surface, but also make phosphorus available by loss through the soil erosion process and surface runoff (Messiga et al., 2013). The use of agricultural practices like no-till may further increases phosphorus concentration in the surface soil due to the non-movement of the soil layer (Pavinatto et al., 2009; Messiga et al., 2010, 2013). Soil phosphorus is also affected by physical characteristics of the soil, such as how the size of soil aggregates influences the extent of soil phosphorus availability to plants (Fonte

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et al., 2014). Therefore, agricultural practices have the potential to alter soil phosphorus concentration and consequently soil phosphorus stocks (Aguiar et al., 2013).

Besides concentrations and stocks, land-use changes are also capable of altering the ratios between carbon, nitrogen and phosphorus (C : N : P) (Ding et al., 2013; Jiao et al., 2013; Schrumpf et al., 2014). In turn, changes in C : N : P ratios may affect several aspects of ecosystem functioning, including carbon sequestration, and, consequently ecosystem responses to climate change (Hessen et al., 2004; Cleveland and Liptzin, 2007; Allison et al., 2010). For instance, soil microorganisms adjusting their stoichiometry with that of the substrate may release or immobilize nitrogen depending on the substrate C : N ratio (Mooshammer et al., 2014a). In turn, litter decomposition also depends on the stoichiometry of the litter, especially on the C : N ratios (Hättenschwiler et al., 2011). In agricultural lands that receive inputs of nitrogen and phosphorus as mineral fertilizer, changes in C : N : P ratios could be significant, and these changes have the ability to trigger changes in entire ecosystem functions (Tischer et al., 2014). However, most studies of soil stoichiometry have been conducted on the surface soil layer (0–10 cm), and fewer on deep soil layers (Tian et al., 2010). Changes at deeper levels could be important and distinct from the surface layers, since most of the applied fertilizer tends to be concentrated on the surface (Sartori et al., 2007).

Agricultural land in Brazil has increased dramatically over recent decades and land-use changes and not agricultural practices have become the most important emitter of greenhouse gases (Lapola et al., 2014). Particularly important is the area covered with pasture that includes approximately 200 million hectares encompassing degraded areas with well-managed pasture (Martinelli et al., 2010). Arable land comprises almost 70 million hectares, with approximately 30 million hectare under no-till cultivation (Boddey et al., 2010), with crop-livestock systems being especially important in the southern region of the country.

Most studies in Brazil on the effects of land-use changes on soil properties deal with soil carbon stocks due to its importance for a low-carbon agriculture (Sá et al.,

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2001; Bayer et al., 2006; Marchão et al., 2009; Maia et al., 2009; Braz et al., 2012; Assad et al., 2013; Mello et al., 2014). On the other hand, there are fewer studies on land-use change affecting soil nitrogen concentration, and especially stocks, and even fewer studies on changes in soil phosphorus stocks. Based on this, this paper aims to investigate effects of land-use changes on carbon concentration, and nitrogen and phosphorus soil concentration and stocks, and on the soil stoichiometry (C : N : P ratio) in several Brazilian regions, using the same study sites and methodology used by Assad et al. (2013) who evaluated changes in soil carbon stocks due to different land uses. Two sampling approaches were used in Assad et al. (2013), one, at the plot level, addressed 17 paired sites comparing soil stocks among native vegetation, pasture and crop-livestock systems, and the second was a regional survey of pasture soils in more than 100 sites.

2 Material and methods

2.1 Study area

A full description of the study area can be found in Assad et al. (2013). Briefly, we conducted two types of surveys: one at the regional level, exclusively in pasture soils, and a second, in which seventeen paired sites were sampled encompassing soils of pastures, crop-livestock systems (CPS) and native vegetation. The regional pasture survey was conducted in November and December of 2010, and 115 pastures located between 6.58 and 31.53° S were selected based first on satellite images in an attempt to broadly encompass three major Brazilian biomes: Cerrado, Atlantic Forest and Pampa, and, secondly, sites were also selected based on their ability to be accessed by roads (Fig. 1). A bias in this scheme is that sampling sites were not randomly selected. A second bias is that, although all pastures were in use at the time they were sampled, it was difficult to visually assess their grazing conditions or stocking rates, which may affect the soil nutrient stocks (Maia et al., 2009; Braz et al., 2012; Assad et al., 2013).

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Paired sites were selected by the EMBRAPA (Empresa Brasileira de Pesquisa Agropecuária) regional offices and sampled between November and December 2011. In these sites, there was an attempt to sample areas of native vegetation, pasture and sites that encompass crop rotation integrated with livestock (CPS). A detailed description of crop rotation and sites that combine crops and livestock management is shown in Table 1. Native vegetation is composed of wood 150 vegetation either in the Atlantic Forest and Cerrado biomes. In sites located in the southern region of the country (Arroio dos Ratos, Tuparecetã, Bagé, and Capão do Leão) the original vegetation is grassy temperate savanna locally referred to as *Campos*, which belongs to the Pampas biome (Table 1). For the sake of simplicity, forests and *Campos* soils were grouped under the category named "original vegetation". Pasture was composed mostly of C_4 grass species of the genus *Brachiaria*; exceptions were in sites located in the southern region of the country where a C_3 grass (*Lolium perenne*) were cultivated. Land-use history is always difficult to obtain with accuracy in Brazil, but Assad et al. (2013) using $\delta^{13}C$ values of soil organic matter showed that most pastures have been in this condition for a long time, and most of the native vegetation seems to have been in this state also for a long time.

Integrated crop-livestock or crop-livestock-forest, and agroforestry systems (CPS) are not a new idea. However, these systems have only been consolidated in recent decades (Machado et al., 2011). The aim of the system is to combine environmental health, as well as increase production and economic viability of farming.

The system consists of diversifying and integrating crops, livestock and forestry systems, within the same area, in intercropping, in succession or rotation. The system can provide environmental benefits such soil conservation, build up soil carbon, reduce environmental externalities and ultimately increase productivity. CPSs include but are not restricted to: no till, the use of cover crops, elimination of agricultural fires (slash-and-burn), and restoration of vast areas of degraded pastures (Hou et al., 2008; Machado et al., 2011; Bustamante et al., 2012; Lapola et al., 2014).

Additionally, the Brazilian law (Law no. 12187 of 29 December 2009), encourages the adoption of good agricultural practices to promote low carbon emission (Low Carbon Emission Program – ABC Program), and stipulates that mitigation should be conducted by adopting: (i) recovery of degraded pastures, (ii) a no-tillage system, (iii) integrated livestock-crop-forest systems, and (iv) re-forestation, in order to reduce approximately 35 to 40 % of Brazil's projected greenhouse gas emissions by 2020 (Assad et al., 2013).

2.2 Precipitation and temperature

The precipitation and temperatures were obtained using the Prediction of Worldwide Energy Resource (POWER) Project (http://power.larc.nasa.gov).

2.3 Sample collection and analysis

Soil sampling is described in detail in Assad et al. (2013). Briefly, in each site, a trench of 60 cm by 60 cm, yielding an area of approximately 360 cm^2 was excavated. For the regional pasture survey, the depth of the trench was approximately 30 cm, and in the paired sites, the depth was approximately 60 cm. Trenches were excavated according to interval depth samples for bulk density were collected first, and after this approximately 500 g of soil was collected for chemical analysis.

Air-dried soil samples were separated from plant material and stones, and then homogenized. The samples were then run through sieves for chemical and physical analysis (2.0 mm sieve diameter) and analysis of soil carbon (0.15 mm sieve diameter).

The concentration of soil nitrogen and carbon was determined by using the elemental analyzer at the Laboratory of Isotopic Ecology Center for Nuclear Energy in Agriculture, University of São Paulo (CENA-USP) in Piracicaba, Brazil.

Phosphorus concentration was determined by extracting soil phosphorus using the Mehlich-3 method of extraction (Mehlich, 1984), and phosphorus concentration was quantified by the colorimetric blue method. Accordingly, the C : P and N : P ratios shown

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here did not use total phosphorus concentration, but available inorganic phosphorus concentration (P_{ME}).

2.4 Soil nitrogen and phosphorus stocks

Carbon stocks were reported in Assad et al. (2013). In this paper, besides carbon concentrations, nitrogen stocks expressed in Mg ha^{-1} and phosphorus stocks expressed in kg ha^{-1} were calculated for the soil depth intervals 0–10, 0–30, and 0–60 cm for the paired sites and 0–10, and 0–30 cm for the pasture regional survey by sum stocks obtained in each sampling intervals (0–5, 5–10, 10–20, 20–30, 30–40, 40–60 cm). Soil nitrogen and phosphorus stocks were estimated based on a fixed mass in order to correct differences caused by land-use changes in soil density (Wendt and Hauser, 2013) using the methodology proposed by Ellert et al. (2008), for details of this correction see Assad et al. (2013).

The cumulative soil nitrogen and phosphorus stocks for fixed depths were calculated by the following equations:

$$S = [X] \cdot \rho \cdot z \quad (1)$$

where S is the cumulative soil nitrogen or phosphorus stock for fixed depths and $[X]$ is the soil nitrogen or phosphorus concentration at the designated depth (z), and ρ is the bulk soil density.

For the paired sites, changes in nutrient stocks between current land use and native vegetation were obtained by comparing differences between the two stocks. The absolute difference (ΔN_{abs} or ΔP_{abs}) was expressed in Mg ha^{-1} for nitrogen or kg ha^{-1} for phosphorus and the relative difference compared to the native vegetation was expressed in percentage (ΔN_{rel} or ΔP_{rel}).

Due to time and financial constraints, we were unable to sample soil from native vegetation near each pasture site in the regional survey. This poses a challenge because it is important to compare changes in the soil nitrogen and phosphorus stocks

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with the native vegetation as done in the paired study sites. In order to overcome the lack of original nutrient soil stocks, we used estimates of native vegetation obtained in the paired sites. Another difficulty is the lack of reliable information on the land-use history; we cannot guarantee that differences among land uses already existed or were due to the replacement of the native vegetation (Braz et al., 2012; Assad et al., 2013). In addition, we only have a point-in-time measurement; we did not follow temporal changes in nitrogen and phosphorus soil stocks. Therefore, it is not possible to know if the soil organic matter achieved a new steady-state equilibrium; as a consequence our results should be interpreted with caution (Sanderman and Baldock, 2010).

2.5 Statistical analysis

In order to test for differences in element concentrations and their respective ratios, we grouped element contents by land use (forest, pasture, CPS) and soil depth (0–5, 5–10, 10–20, 20–30, 30–40, 40–60 cm). Carbon, nitrogen and phosphorus concentration, and soil nitrogen and phosphorus stocks must be transformed using Box–Cox techniques because they did not follow a normal distribution. Accordingly, statistical tests were performed using transformed values, but non-transformed values were used to report average values. The element ratio was expressed as molar ratios and ratios followed a normal distribution and were not transformed.

For the paired sites, differences between land uses (native vegetation, CPS and pasture) were tested with ANCOVA, with the dependent variables being transformed nutrient concentrations at the soil depth intervals described above, and stocks at the soil layers of 0–10, 0–30, and 0–60 cm; the independent variables were land-use type. As mean annual temperature (MAT), mean annual precipitation (MAP), and soil texture may influence soil nutrient concentration, ratios, and stocks, these variables were also included in the model as co-variables. The *post-hoc* Tukey Honest Test for unequal variance was used to test for differences among nutrient stocks of different land uses. In order to determine whether changes in soil nutrient stocks between current land use

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and native vegetation were statistically significant, we used a one-sample *t* test, where the null hypothesis was that the population mean was equal to zero.

All tests were reported as significant at a level of 10 %. Statistical tests were performed using a STATISTICA12 package.

3 Results

3.1 Paired study sites

3.1.1 Soil carbon, nitrogen, and phosphorus concentrations and related ratios

As expected, carbon, nitrogen, and phosphorus concentrations decreased with soil depth (Fig. 2). The average carbon concentration was higher in the topsoil (0–5 and 5–10 cm) of native vegetation soils compared with pasture and CPS soils ($p = 0.05$). However, in deeper soil layers, there was no statistically significant difference between native vegetation, pasture and CPS soils (Fig. 2a). The average soil nitrogen concentration followed the same pattern as carbon (Fig. 2b). However, differences between forest, and pasture and CPS soils were significant down to the 10–20 cm soil layer. The phosphorus concentrations in the soil profiles showed a different pattern than carbon and nitrogen. Phosphorus concentrations were higher in the CPS and pasture soils than in forest soils in the topsoil and also in the soil depth layer of 10–20 cm (Fig. 2c). The C : N ratios of pasture and CPS soils were higher than the native vegetation soils in all soil depths; however, this difference was not statistically significant for any particular depth (Fig. 3a). There was a difference in the $C:P_{ME}$ ratio between forest, pasture and CPS soils, this ratio was higher in the forest soils, intermediate in the pasture, and lower in the CPS soils (Fig. 3b). Due to the wide variability of the data, differences were only significant in the first three soil depth intervals: 0–5 cm ($p < 0.01$); 5–10 cm ($p < 0.01$); and 10–20 cm ($p = 0.03$). Finally, the $N:P_{ME}$ showed a similar trend than $C:P_{ME}$, with higher ratios in native vegetation soils, decreasing in

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the pasture and reaching the lowest values in the CPS soils (Fig. 3c). Again, values were only different at the same soil depth intervals observed for C : P_{ME}, with all of them at a probability ratio lower than 0.01.

3.1.2 Soil nitrogen and phosphorus stocks

The average nitrogen stock of the native vegetation soils in the topsoil was 2.27 Mg ha^{-1} decreasing significantly to 1.72 Mg ha^{-1} in the CPS ($p = 0.05$) and to 1.54 Mg ha^{-1} in pasture soils ($p < 0.01$) (Table 2). In the next soil layer (0–30 cm), the same tendency was observed. The average nitrogen stock was equal to 5.12 Mg ha^{-1}, decreasing significantly to 3.94 Mg ha^{-1} in the CPS ($p = 0.04$), and to 3.84 Mg ha^{-1} in pasture soils ($p = 0.03$) (Table 2). On the other hand, differences in soil nitrogen stocks among different land uses were not significant at the 0–60 cm of the soil layer; the nitrogen soil stock was 7.30 Mg ha^{-1} in the native vegetation, and 5.93 and 6.16 Mg ha^{-1} in the CPS and pasture soils, respectively (Table 2).

In general, there was a net loss of nitrogen stocks between native vegetation and current land uses in the soil (Table 2). In the forest-CPS pairs for the topsoil the $\Delta N_{abs} = -0.64$ Mg ha^{-1}, and a $\Delta N_{rel} = -22$ %, both differences were significant at 1 % level (Table 2). The same pattern was observed for the 0–30 cm soil interval, where $\Delta N_{abs} = -1.28$ Mg ha^{-1}, and the $\Delta N_{rel} = -20$ % (Table 2). In the forest-pasture paired sites, the $\Delta N_{abs} = -0.63$ Mg ha^{-1}, and the $\Delta N_{rel} = -28$ % found in the topsoil were both statistically significant at 1 % (Table 2). The same was true for the 0–30 cm soil layer, where the $\Delta N_{abs} = -1.10$ Mg ha^{-1}, which was equivalent to a loss of −22 ‰ (Table 2).

On the other hand, a net gain of phosphorus was observed between native vegetation and current land uses in the soil. The phosphorus soil stock in the topsoil of native vegetation areas was equal to 11.27 kg ha^{-1}, increasing significantly to 30.06 kg ha^{-1} ($p < 0.01$) in the CPS soil and to 21.6 kg ha^{-1} ($p < 0.01$) in the pasture soils (Table 3). Considering the 0–30 cm soil layer, the phosphorus stock in the native vegetation soils was 21.74 kg ha^{-1}, also significantly increasing in the CPS soils to 49.50 kg ha^{-1} ($p = 0.02$), and to 47.60 kg ha^{-1} in the pasture soils (Table 3). Finally, in the 0–60 cm soil layer, the phosphorus stock in the native vegetation soils was 42.70 kg ha^{-1}, which was not significantly lower than the phosphorus soil stock in the CPS soils, which was equal to 62.90 kg ha^{-1}. On the other hand, the soil phosphorus stock in the pasture soils was 68.33 kg ha^{-1}, which is significantly different ($p = 0.02$) than the soil phosphorus stock of the native vegetation soils (Table 3).

In relative terms, in the topsoil, for the native vegetation-CPS paired sites an overall phosphorus gain was observed, the $\Delta P_{abs} = 20.56$ kg ha^{-1}, and the $\Delta P_{rel} = 325$ %, both significant at 1 % level (Table 3). The same pattern was observed at the 0–30 cm soil layer, where the $\Delta P_{abs} = 27.03$ kg ha^{-1}, and the $\Delta P_{rel} = 205$ %, and at the 0–60 cm soil layer, where the $\Delta P_{abs} = 25.64$ kg ha^{-1}, and the $\Delta P_{rel} = 145$ % (Table 3). In the native vegetation-pasture pair sites, the same increase in phosphorus stocks was also observed in the pasture soils. In the topsoil, the $\Delta P_{abs} = 10.06$ kg ha^{-1} ($p < 0.01$), and the $\Delta P_{rel} = 52$ % ($p < 0.01$) were statistically significant (Table 3). The same was true for the 0–30 cm soil layer, in this case the $\Delta P_{abs} = 25.70$ kg ha^{-1} ($p < 0.01$) and the $\Delta P_{rel} = 220$ % ($p < 0.01$); and for the 0–60 cm soil layer, where the $\Delta P_{abs} = 25.42$ kg ha^{-1} ($p < 0.01$), and the $\Delta P_{rel} = 172$ % ($p < 0.01$) (Table 3).

3.2 Regional survey of pasture soils

3.2.1 Soil carbon, nitrogen, and phosphorus concentrations and related ratios

We compared element concentrations and ratios of the regional survey pasture soils with the native vegetation soil site of the paired sites (Figs. 2 and 3). Carbon, nitrogen and phosphorus concentrations decreased with soil depth, and were significantly lower ($p < 0.01$) in the pasture soils than in the native vegetation soils (Fig. 2). The C : N ratio of the regional pasture survey was higher than the native vegetation soil (Fig. 3). The $C:P_{ME}$ and $N:P_{ME}$ ratios were much higher in the pasture soils of the regional survey compared with forest soils, and in these cases, there was a sharp increase with soil depth (Fig. 3).

3.2.2 Soil nitrogen and phosphorus stocks

At the 0–10 cm soil layer the average total soil nitrogen stock was equal to $1.66 \pm 0.87\,\mathrm{Mg\,ha^{-1}}$ (Table 4), and at 0–30 cm the average soil stock was $3.91 \pm 1.90\,\mathrm{Mg\,ha^{-1}}$. At the 0–10 and 0–30 cm soil layers the average phosphorus stock was 8.50, and 14.71 $\mathrm{kg\,ha^{-1}}$, respectively (Table 4). The average nitrogen stock in the pasture soils of the regional survey at both depth layers (0–10 and 0–30 cm) was very similar to the stocks found in the pasture and CPS of the paired sites survey, and, therefore, also lower than the soil stocks found in the native vegetation areas (Table 4). On the other hand, the average phosphorus stock in the pasture soils of the regional survey was much lower than the soil stocks of pasture and CPS of the paired sites surveys, being even smaller than the soil stocks of native vegetation areas (Table 4).

4 Discussion

4.1 Land-use changes alter C : N : P soil stoichiometry

In this section we will focus our discussion on changes in soil stoichiometry, because changes in element concentrations will be discussed in the next section that deals with changes in nitrogen and phosphorus stocks. We observed important changes not only in concentrations, but also in soil stoichiometry (Figs. 2 and 3).

Overall, the C : N ratio was lower in the native vegetation soils compared with pasture and CPS soils (Fig. 3a), yet despite such differences, it was only statistically different at the soil surface. Such differences are probably explained by a nitrogen loss and not a carbon gain, since soil carbon stocks in pasture and CPS soils were lower than in native vegetation soils (Assad et al., 2013). The reasons for preferential nitrogen loss in these systems in relation to the forest soil are discussed in the next section.

Different soil C : N ratios as observed in the native vegetation, and pasture and CPS systems could influence nitrogen dynamics, favoring faster organic matter

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decomposition and nitrogen mineralization in native vegetation soils due to lower soil C : N ratios (Mooshammer et al., 2014b). However, it is difficult to conclude whether a small difference between native vegetation soils and the others would be enough to trigger such changes. According to Mooshammer et al. (2014a), the threshold value of the C : N ratio required to change the status of nitrogen to be mineralized or immobilized by the soil biota is around 20. Soil C : N ratios, even in the pasture and CPS soils are well below this value (Fig. 3a).

Another important trend was the lower depth variability of C : N ratios compared with the carbon and nitrogen variability with depth (Fig. 2a and b). This trend is consistent with the initial hypothesis of Tian et al. (2010) who hypothesized that the C : N ratio would not vary with depth because of the coupling of carbon and nitrogen in the soil. According to Tischer et al. (2014) such constancy is a consequence of similar inputs of organic matter by primary producers to the soils.

On the other hand, it is expected that soil C : P and N : P decreases with soil depth mainly because the most important source of phosphorus to the soil is from weathering (Tian et al., 2010). Although vegetation extracts phosphorus from deep soil layers and allocates its phosphorus on the soil surface through litterfall and decomposition, weathering appears to be more important, causing a decrease of the element : P ratios with soil depth. As already mentioned, we do not have total P, but only available inorganic P (P_{ME}). As available P generally decreases with soil depth, we expected an increase of C : P_{ME} and N : P_{ME} with soil depth. In fact we observed a decrease of these ratios, but only between the surface down to 40 cm, in the deepest soil layer (40–60 cm), both ratios decreased again (Fig. 3b and c). Without having total P contents, it is difficult to further speculate about the reasons of such trends.

Among different land uses, the elements: P_{ME} were also distinct (Fig. 3b and c). As the carbon concentration and stocks, especially, decreased in pasture and CPS soils compared to native vegetation soils (Assad et al., 2013), it is clear that the C : P_{ME} decreased in the pasture soils and further in the CPS soils because there was an increase in available phosphorus caused by the use of P-fertilizers (Fig. 2c). The same

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trend was observed with $N:P_{ME}$, but in this case there is probably a combination of N loss and P enrichment in pasture and CPS soils compared with native vegetation soils as discussed below.

4.2 Land-use changes alter nitrogen and phosphorus stocks

In most of the plot-level paired sites and in most of the regional soil survey, we found a loss of nitrogen compared to the native vegetation. It seems that this is a common pattern observed for different crops and different types of land management in several regions of Brazil; like in the Northeast (Lima et al., 2011; Fracetto et al., 2012; Barros et al., 2013; Sacramento et al., 2013); in Central Brazil (Cardoso et al., 2010; Silva et al., 2011; Guareschi et al., 2012) and in the South (Sisti et al., 2004; Sá et al., 2013; Santana et al., 2013). Sá et al. (2013) found lower soil nitrogen stocks in several farms located in southern Brazil (Paraná State) that have adopted no-till and crop rotation systems for at least ten years compared with the native vegetation of the region. On the other hand, the adoption of no-till systems tends to increase soil nitrogen stocks compared to conventional tillage (Sisti et al., 2004; Sá et al., 2013). In this respect, it is interesting to note that the only three sites (SL, PG, AP) where the soil nitrogen stocks were higher in the agriculture field than in the native vegetation, were CPS sites, where no-till was practiced and there was a system of crop rotation, with soybean in the summer, and oat or wheat in the winter (Table 1).

We found a positive and significant ($p < 0.01$) correlation between soil carbon stock losses found by Assad et al. (2013) and the soil nitrogen stock losses found in this study. Such correlations were especially significant in the CPS systems, where more than 70 % of the variance in the nitrogen losses were explained by carbon losses (Figs. 4 and 5). These correlations are an indication that whatever mechanisms are leading to such losses, they are simultaneously affecting carbon and nitrogen. There are several studies at the plot level showing that changes in soil properties is one of the leading causes affecting losses of organic matter with soil cultivation (e.g. Mikhailova et al., 2000; Kucharik et al., 2001). In addition, several regional and global surveys

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also pointed in the same direction (e.g. Davidson and Ackerman, 1993; Amundson, 2001; Guo and Gifford, 2002; Zinn et al., 2005; Ogle et al., 2005; Don et al., 2011; Eclesia et al., 2012). It seems that a combination of decreasing organic matter inputs, in the case where crops replaced native forests, with an increase in soil organic matter decomposition leads to a decrease in the long run. This decrease seems to especially be fostered in annual crops by exposing bare soil between harvests, leading to higher temperatures (Baker et al., 2007; Coutinho et al., 2010; Salimon et al., 2004), which in turn leads to higher decomposition rates (e.g. Davidson and Janssens, 2006; Dorrepaal et al., 2009). For instance, Carmo et al. (2012) found higher soil temperature and high CO_2 emissions in pasture soil compared with the forest soil nearby, with both sites located in the southeast region of Brazil (State of São Paulo).

Nitrogen dynamics is regulated by a balance between inputs, losses and transformations between different forms of nitrogen (Drinkwater et al., 2000). Generally, land-use changes tend to disrupt the nitrogen cycle of the native vegetation. The main natural nitrogen input is via biological nitrogen fixation (BNF), and the main anthropogenic addition is via N-mineral fertilizer inputs. In tropical forests, BNF is considered one of the main inputs of nitrogen (Vitousek et al., 2002). In crops like soybean, BNF is also important as a source of new nitrogen to the system, especially in Brazil where soybean may fix higher amounts of nitrogen (Alves et al., 2003). Several of the CPS systems evaluated in this study involve the use of soybean under crop rotation systems (Table 1). However, decreases of soil nitrogen stocks of these CPS were observed, compared with soils of the native vegetation (Fig. 6a and b). The same was observed by Boddey et al. (2010) comparing soil carbon and nitrogen stocks of no-till and conventional tillage systems involving a crop rotation with soybean in farms located in the State of Rio Grande do Sul (southern Brazil). According to these authors, the nitrogen export by grain harvesting is high enough to prevent a build-up of this nutrient in the soil (Boddey et al., 2010).

Most pastures in Brazil are not fertilized, so over time, a decrease in nitrogen inputs coupled with an increase of nitrogen outputs is generally observed, leading to lower

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mineralization and nitrification rates (Verchot et al., 1999; Melillo et al., 2002; Garcia-Montiel et al., 2000; Wick al., 2005; Neill et al., 2005; Carmo et al., 2012). According to Boddey et al. (2004), not even the return of nitrogen to soil pasture via urine and dung is sufficient to compensate for other nitrogen losses. As a consequence the continuous use of unfertilized pastures leads to overall N-impoverishment in the system, leading to lower soil nitrogen stocks, as observed in this study.

On the other hand, we observed a general increase in soil phosphorus stocks of pasture and CPS-paired sites compared with soil stocks of the native vegetation (Fig. 7a and b). The higher soil phosphorus stocks in the CPS could be explained by the addition of phosphorus fertilizer to the fields (Aguiar et al., 2013; Messsiga et al., 2013). Generally, an increase of soil phosphorus is observed after use of P-fertilizers in the topsoil due to the low mobility of phosphorus, especially in no-till systems (Pavinatto et al., 2009; Messiga et al., 2010). In several of the CPS sites, there are crop rotations between maize, rice and soybean, and all these crops are fertilized with phosphorus, especially soybean, because phosphorus is an important nutrient in the biological nitrogen fixation process (Divito and Sadras, 2014). The variation of phosphorus concentration with soil depth provides indirect support for this hypothesis. In the majority of the CPS sites and even pasture soils of the paired sites there is a gradient in phosphorus concentration with much higher concentrations near the soil surface (Fig. 2c).

The soil phosphorus stocks of pastures located in the paired sites were higher than soil phosphorus stocks of the regional pasture survey. For instance, at the 0–10 cm soil layer, the average P_{stock} of pasture soil at the paired sites was equal to 22 kg ha^{-1} (Table 3), which is significantly higher than the average P_{stock} of pasture soil sampled in the regional level survey (9 kg ha^{-1}, Table 4). This latter average is similar to the average P_{stock} of the native vegetation sampled in the paired study sites, which was equal to 12 kg ha^{-1} (Table 3). As we mentioned earlier, we do not have accurate information on pasture management and grazing conditions. However, as the pasture-paired sites were located in research stations and well-managed farms, we believe that overall, the pasture in these areas is in better condition compared with pasture included in the regional survey. As already mentioned, in some pasture of the paired sites, a steep decrease in phosphorus content with soil depth was observed, being indirect evidence that these pastures received some kind of phosphorus amendment or lime application that raised the pH and made phosphorus available to plants (Uehara and Gillman, 1981). If this is the case, these differences in pasture management will probably explain differences observed in soil phosphorus stocks between pastures of the paired sites and regional survey. This is because Fonte et al. (2014) found that soils of well-managed pastures located on poor tropical soils had great differences in soil aggregation, which in turn influence the soil phosphorus level, favoring a higher phosphorus content in well-managed pastures compared to degraded pastures. On the other hand, Garcia-Montiel et al. (2000) and Hamer et al. (2013) found an increase in soil phosphorus stocks for several years after the conversion of Amazonian forests to unfertilized pastures. The main cause of this increase seems to be soil fertilization promoted by ash of forest fires, coupled with root decomposition of the original vegetation. However, it seems that with pasture aging, there is a decrease in available phosphorus mainly in strongly weathered tropical soils (Townsend et al., 2002; Numata et al., 2007).

In an earlier paper Assad et al. (2013) have shown a decrease in soil carbon stock in relation to the original vegetation either for pasture and CPS soils. In this paper we found that nitrogen stocks also decrease considerably with land-use changes, even in well managed CPS systems, and especially in pastures of the regional survey that reflect better the reality of pasture management in Brazil. These findings have important policy implications because Brazil recently implemented a program (Low Carbon Agriculture) devoted to increasing carbon and nitrogen concentration in soils by a series of techniques, especially no-till, crop-livestock systems (CPS), and improvement of degraded pastures. Therefore, the findings of this paper set a baseline of soil nutrients stocks and stoichiometry for future comparisons.

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Table 1. Characterization of sampled sites: native vegetation (NV), pastures (P), crop-livestock systems (CPS).

City (Code) – Region	Point	Latitude	Longitude	Land-use system	Established	Biome
Sete Lagoas (SL) – Southeast	1	19°29′57″	44°11′03″	Pasture	–	Cerrado
	2	19°29′24″	44°10′48″	CPS (1 year of pasture followed by 2 years of corn)	–	Cerrado
	3	19°29′11″	44°11′19″	CPS (corn, pasture and eucalyptus)	2009	Cerrado
	4	19°29′37″	44°11′09″	Forest	–	Cerrado
	5	19°29′28″	44°11′08″	CPS (1 year of pasture followed by 2 years of soybean)	–	Cerrado
Coronel Xavier (CX) – Southeast	6	21°01′06″	44°12′53″	Native Vegetation	–	Atlantic Forest
	7	21°01′13″	44°12′56″	Pasture	–	Atlantic Forest
	8	21°01′12″	44°12′53″	CPS (corn, pasture and eucalyptus)	2009	Atlantic Forest
	9	20°59′35″	44°10′18″	Pasture	–	Atlantic Forest
	10	20°59′36″	44°10′18″	Forest	–	Atlantic Forest
	11	20°59′40″	44°10′20″	CPS (corn, pasture and eucalyptus)	2009	Atlantic Forest
São Carlos (SC) – Southeast	15	21°58′49″	47°51′10″	Pasture	–	Cerrado
	16	21°58′27″	47°51′10″	CPS (pasture and eucalyptus)	2010	Cerrado
	17	21°58′38″	47°51′17″	Forest	–	Cerrado
	18	21°57′47″	47°51′00″	CPS (pasture and eucalyptus)	2007	Cerrado
Cafeara (CS) – Southeast	19	22°50′38″	51°42′28″	CPS (pasture and soybean)	2003	Atlantic Forest
	20	22°50′02″	51°42′52″	Forest	–	Atlantic Forest
	21	22°52′12″	51°43′37″	Pasture	–	Atlantic Forest
Iporã (IP) – Southeast	22	24°00′26″	53°45′01″	CPS (1 year of pasture and 3 years of soybean)	–	Atlantic Forest
	23	24°00′06″	53°45′32″	Pasture	–	Atlantic Forest
	24	24°01′20″	53°45′38″	Forest	–	Atlantic Forest
Xambrê (XA) – Southeast	25	23°47′34″	53°36′20″	Pasture	–	Atlantic Forest
	26	23°47′14″	53°36′10″	CPS (pasture and soybean)	2000	Atlantic Forest
	27	23°47′23″	53°36′31″	CPS (soybean and eucalyptus)	2010	Atlantic Forest
	28	23°48′29″	53°35′25″	Forest	–	Atlantic Forest
Campo Mourão (CM) – Southeast	29	24°06′25″	52°21′40″	Pasture	–	Atlantic Forest
	30	24°06′21″	52°21′34″	CPS (corn and pasture)	2001	Atlantic Forest
	31	24°06′18″	52°21′34″	Forest	–	Atlantic Forest
Juranda (JU) – Southeast	32	24°18′21″	52°42′17″	CPS (rotation soybean or corn and pasture)	2006	Atlantic Forest
	33	24°18′34″	52°42′16″	Pasture	–	Atlantic Forest
	34	24°18′10″	52°42′18″	Forest	–	Atlantic Forest

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Table 1. Continued.

City (Code) – Region	Point	Latitude	Longitude	Land-use system	Established	Biome
Ponta Grossa (PG) – Southeast	35	25°06′37″	50°03′04″	CPS (soybean, pasture and eucalyptus)	2006	Atlantic Forest
	36	25°06′32″	50°03′26″	CPS (soy in summer and oats in winter)	2010	Atlantic Forest
	37	25°06′43″	50°03′49″	Forest	–	Atlantic Forest
	38	25°06′54″	50°03′49″	Pasture	–	Atlantic Forest
Arroio dos Ratos (AR) – South	39	30°06′14″	51°41′32″	CPS (soy in summer and *L. multiflorum* in winter)	2002	Pampa
	40	30°06′12″	51°41′33″	CPS (corn or soy in summer and *L. multiflorum* in the winter)	2002	Pampa
	41	30°06′06″	51°41′58″	*Campos*	–	Pampa
	42	30°06′06″	51°41′31″	Pasture	–	Pampa
Tuparecetã (TU) – South	43	28°56′34″	54°21′35″	CPS (soy in summer and *L. multiflorum* in winter)	2001	Pampa
	44	28°56′11″	54°21′25″	CPS (soy in summer and *L. multiflorum* in winter)	2001	Pampa
	45	28°56′31″	54°20′02″	Pasture	–	Pampa
	46	28°55′48″	54°20′29″	*Campos*	–	Pampa
Nova Esperança do Sul (NS) – South	47	29°27′12″	54°48′40″	CPS (sorghum, pasture and eucalyptus)	2007	Atlantic Forest
	48	29°27′33″	54°49′17″	Pasture	–	Atlantic Forest
	49	29°27′31″	54°49′18″	Forest	–	Atlantic Forest
Bagé (BA) – South	**50**	31°22′11″	54°00′11″	CPS (rice in summer and *L. multiflorum* in winter)	2007	Pampa
	51	31°22′01″	54°00′28″	*Campos*	–	Pampa
	52	31°28′30″	53°58′15″	CPS (sorghum, pasture and eucalyptus)	2005	Pampa
	53	31°19′17″	54°00′12″	CPS (soy in summer and *L. multiflorum* in winter)	2007	Pampa
Capão do Leão (CL) – South	54	31°49′57″	52°28′28″	*Campos*	–	Pampa
	55	31°49′19″	52°28′40″	CPS (soy in summer and *L. multiflorum* in winter)	2007	Pampa
	56	31°49′19″	52°28′11″	CPS (soy or rice in summer and *L. multiflorum* in winter)	2007	Pampa
Passo Fundo (PF) – South	57	28°13′32″	52°24′30″	CPS (soy or corn in summer and *L. multiflorum* or oats in the winter)	1996	Atlantic Forest
	58	28°13′31″	52°24′28″	CPS (soy or corn in summer and *L. multiflorum* or oats in winter)	1996	Atlantic Forest
	59	28°13′30″	52°24′24″	Forest	–	Atlantic Forest

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Table 2. Mean, standard-deviation (SD), minimum and maximum of soil nitrogen stocks (N_{stock}) at 0–10, 0–30, and 0–60 cm soil depth layer for forest, crop-livestock systems and pasture soils at the paired study sites. ΔN_{abs} is the difference between the soil nitrogen stock of native vegetation and crop livestock systems and pasture soils obtained in the paired study sites. ΔN_{rel} is the same difference expressed as percentage. Nitrogen losses are indicated by a minus sign (−).

			Native vegetation (0–10 cm)		
	N	Mean	SD	Minimum	Maximum
N_{stock} (Mg ha^{-1})	16	2.27	1.04	0.97	4.64
			CPS (0–10 cm)		
	N	Mean	SD	Minimum	Maximum
N_{stock} (Mg ha^{-1})	27	1.72	0.72	0.52	2.80
ΔN_{abs} (Mg ha^{-1})	27	−0.64	0.76	−2.54	0.52
Δ_{rel} (%)	27	−21.81	30.63	−71.37	42.93
			Pasture (0–10 cm)		
	N	Mean	SD	Minimum	Maximum
N_{stock} (Mg ha^{-1})	13	1.54	0.89	0.55	2.82
ΔN_{abs} (Mg ha^{-1})	13	−0.63	0.70	−2.02	0.43
Δ_{rel} (%)	13	−27.89	27.53	−70.77	18.71
			Native vegetation (0–30 cm)		
	N	Mean	SD	Minimum	Maximum
N_{stock} (Mg ha^{-1})	16	5.12	2.12	2.20	9.01
			CPS (0–30 cm)		
	N	Mean	SD	Minimum	Maximum
N_{stock} (Mg ha^{-1})	27	3.94	1.65	1.45	7.65
ΔN_{abs} (Mg ha^{-1})	27	−1.28	1.70	−4.89	1.60
Δ_{rel} (%)	27	−19.81	29.19	−65.14	45.81
			Pasture (0–30 cm)		
	N	Mean	SD	Minimum	Maximum
N_{stock} (Mg ha^{-1})	13	3.84	1.85	1.52	6.49
ΔN_{abs} (Mg ha^{-1})	13	−1.10	1.14	−3.20	0.80
Δ_{rel} (%)	13	−21.84	18.95	−63.63	14.06
			Native vegetation (0–60 cm)		
	N	Mean	SD	Minimum	Maximum
SN_{stock} (Mg ha^{-1})	16	7.30	3.28	2.68	12.00
			CPS (0–60 cm)		
	N	Mean	SD	Minimum	Maximum
SN_{stock} (Mg ha^{-1})	27	5.93	2.51	2.12	11.68
ΔSN_{abs} (Mg ha^{-1})	27	−1.48	2.37	−5.12	2.82
ΔSN_{rel} (%)	27	−13.41	31.47	−59.97	41.42
			Pasture (0–60 cm)		
	N	Mean	SD	Minimum	Maximum
SN_{stock} (Mg ha^{-1})	13	6.16	2.79	2.80	10.19
ΔSN_{abs} (Mg ha^{-1})	13	−1.54	1.47	−3.89	1.05
ΔSN_{rel} (%)	13	−17.67	20.20	−47.21	20.62

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Table 3. Mean, standard-deviation (SD), minimum and maximum of soil phosphorus stocks (P_{stock}) at 0–10, 0–30, and 0–60 cm soil depth layer for forest, crop-livestock systems and pasture soils at the paired study sites. ΔP_{abs} is the difference between the soil phosphorus stock of native vegetation and crop livestock systems and pasture soils obtained in the paired study sites. ΔP_{rel} is the same difference expressed as a percentage. Phosphorus losses are indicated by a minus sign (–).

			Native vegetation (0–10 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	16	11.27	14.26	0.80	60.50
			CPS (0–10 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	27	30.06	25.63	1.60	95.50
ΔP_{abs} (kg ha^{-1})	27	20.56	23.91	−14.50	78.50
Δ_{rel} (%)	27	324.96	381.11	−23.97	1650.11
			Pasture (0–10 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	13	21.63	22.35	0.60	78.10
ΔP_{abs} (kg ha^{-1})	13	10.06	26.78	−50.50	62.05
Δ_{rel} (%)	13	52.14	813.43	−83.47	2818.72
			Native vegetation (0–30 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	16	21.74	24.49	3.10	105.50
			CPS (0–30 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	27	49.50	37.11	3.20	137.50
ΔP_{abs} (kg ha^{-1})	27	27.03	41.48	−79.01	102.50
Δ_{rel} (%)	27	205.05	245.34	−74.18	900.08
			Pasture (0–30 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	13	47.60	60.77	2.30	218.00
ΔP_{abs} (kg ha^{-1})	13	25.70	64.17	−83.51	191.35
Δ_{rel} (%)	13	218.59	324.31	−79.16	937.76
			Native vegetation (0–60 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	16	42.70	53.92	6.40	216.50
			CPS (0–60 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	27	62.90	39.75	6.90	155.49
ΔP_{abs} (kg ha^{-1})	27	25.64	62.51	−175.00	107.49
Δ_{rel} (%)	27	145.54	178.00	−100.00	535.23
			Pasture (0–60 cm)		
	N	Mean	SD	Minimum	Maximum
P_{stock} (kg ha^{-1})	13	68.33	72.12	11.90	241.40
ΔP_{abs} (kg ha^{-1})	13	25.42	89.37	−184.52	201.16
Δ_{rel} (%)	13	171.92	285.12	−100.00	850.26

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Table 4. Mean, standard-deviation (SD), median minimum, and maximum, standard-deviation (SD) Soil nitrogen (N_{stocks}) and phosphorus (P_{stocks}) at 0–10 and 0–30 cm soil depth layers for pasture soils included in the regional survey.

	Depth (cm)	N	Mean	SD	Median	Minimum	Maximum
N_{stocks} (Mg ha^{-1})	10	115	1.66	0.87	1.49	0.40	4.20
N_{stocks} (Mg ha^{-1})	30	115	3.91	1.90	3.61	1.01	8.90
P_{stocks} (kg ha^{-1})	10	115	8.50	14.60	3.08	0.50	89.50
P_{stocks} (kg ha^{-1})	30	115	14.71	26.90	5.72	1.01	179.50

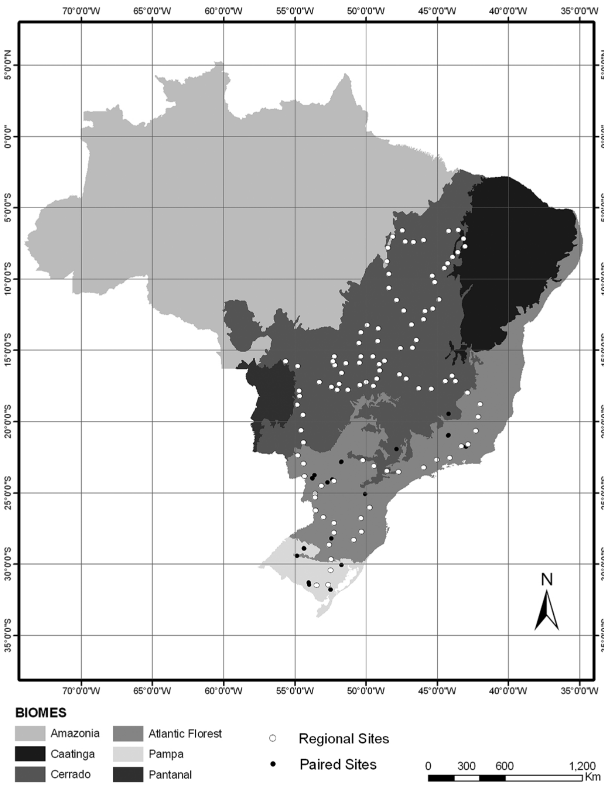

Figure 1. Sampling sites located throughout Brazil. White circles indicate pasture sites of the regional survey; black circles indicate paired study sites, and various shaded areas indicate Brazilian biomes.

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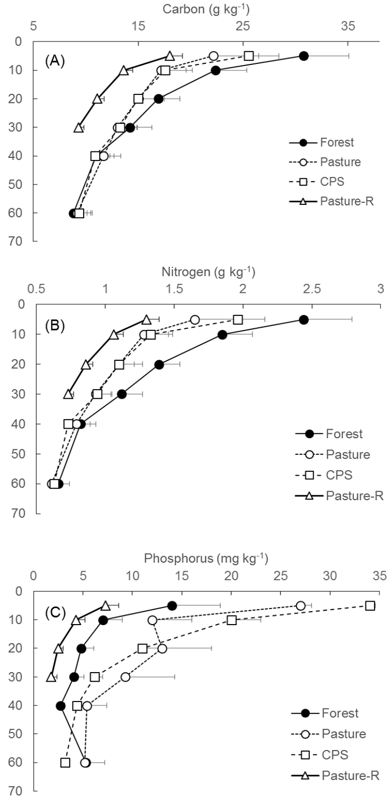

Figure 2. Soil depth variability of **(a)** carbon, **(b)** nitrogen and **(c)** phosphorus in forest, pasture and CPS soils. The horizontal bars are standard errors.

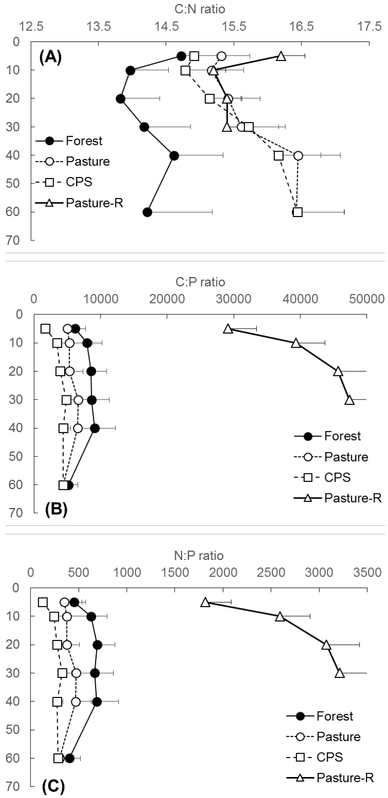

Figure 3. Soil depth variability of **(a)** C : N ratios, **(b)** C : P_{ME} and **(c)** N : P_{ME} in forest, pasture and CPS soils. The horizontal bars are standard errors.

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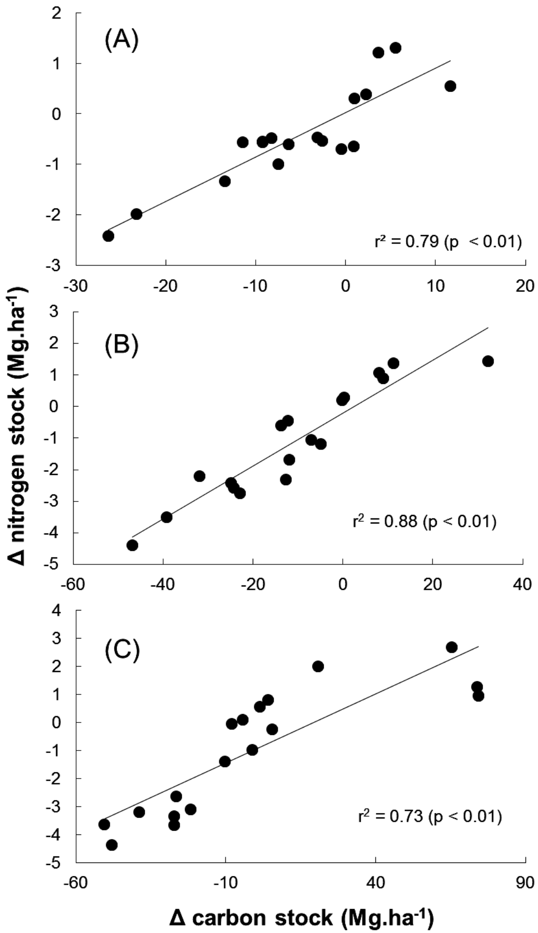

Figure 4. Scatter plot of soil carbon stock losses from the paper of Assad et al. (2013) and soil nitrogen stock losses found in this study between CPS and native vegetation in the paired-study sites **(a)** 0–10 cm **(b)** 0–30 cm **(c)** 0–60 cm depth intervals.

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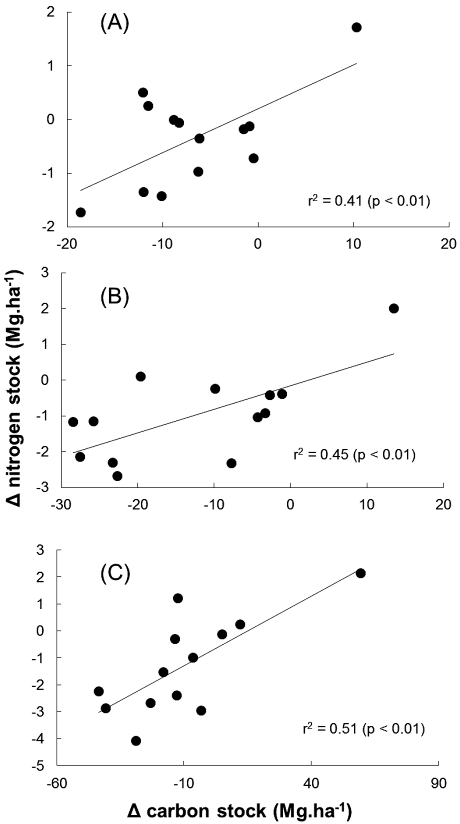

Figure 5. Scatter plot of soil carbon stock losses from Assad et al. (2013) and soil nitrogen stock losses found in our study between pasture and native vegetation in the paired-study sites **(a)** 0–10 cm **(b)** 0–30 cm **(c)** 0–60 cm depth intervals.

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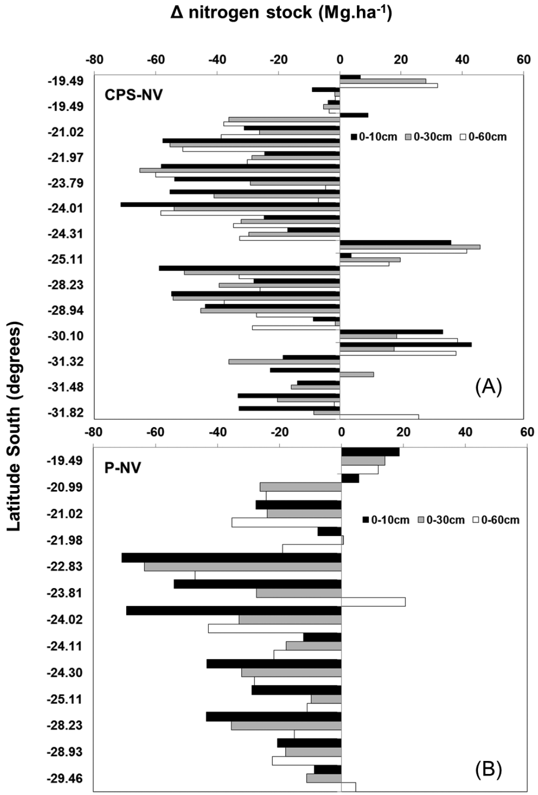

Figure 6. Absolute difference of soil nitrogen stocks between different depth intervals: **(a)** crop-livestock systems (CPS) and native vegetation (NV); and **(b)** pasture (P) and native vegetation (NV) at different paired study sites. Each paired-site study area is indicated by its latitude. Losses are indicated by a minus sign (–).

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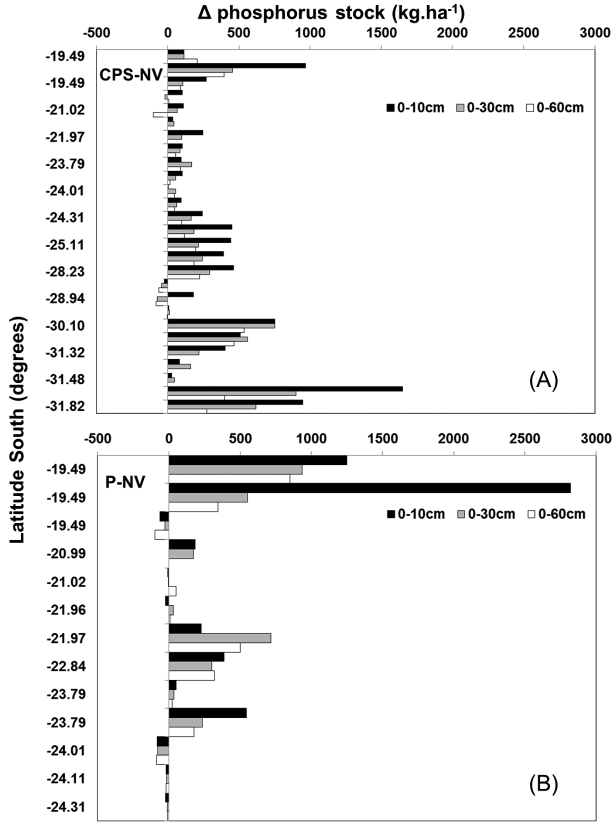

Figure 7. Absolute difference of soil phosphorus stocks between different depth intervals: **(a)** crop-livestock systems (CPS) and native vegetation (NV); and **(b)** pasture (P) and native vegetation (NV) at different paired study sites. Each paired-site study area is indicated by its latitude. Losses are indicated by a minus sign (−).

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