

1 Representing northern peatland microtopography and
2 hydrology within the Community Land Model

3 Xiaoying Shi^{1*}, Peter E. Thornton^{1*}, Daniel M. Ricciuto¹, Paul J. Hanson¹, Jiafu Mao,
4 Stephen D. Sebestyen², Natalie A. Griffiths¹, and Gautam Bisht³

5 ¹ Climate Change Science Institute and Environmental Sciences Division, Oak Ridge
6 National Laboratory, Oak Ridge, TN 37831, USA

7 ² Northern Research Station, USDA Forest Service, Grand Rapids, Minnesota 55744,
8 USA

9 ³ Earth Sciences Division, Lawrence Berkeley National Laboratory, Berkeley, CA
10 94720, USA

11 * To whom correspondence should be addressed

12 Corresponding authors' e-mail: shix@ornl.gov, thorntonpe@ornl.gov

13 Fax: 865-574-2232

14

15 This manuscript has been authored by UT-Battelle, LLC, under Contract No. DE-

16 AC05-00OR22725 with the U.S. Department of Energy. The United States

17 Government retains and the publisher, by accepting the article for publication,

18 acknowledges that the United States Government retains a non-exclusive, paid-up,

19 irrevocable, world-wide license to publish or reproduce the published form of this

20 manuscript, or allow others to do so, for United States Government purposes.

21

22

23

24

25 **Abstract**

26 Predictive understanding of northern peatland hydrology is a necessary
27 precursor to understanding the fate of massive carbon stores in these systems
28 under the influence of present and future climate change. Current models have
29 begun to address microtopographic controls on peatland hydrology, but none have
30 included a prognostic calculation of peatland water table depth for a vegetated
31 wetland, independent of prescribed regional water tables. We introduce here a new
32 configuration of the Community Land Model (CLM) which includes a fully prognostic
33 water table calculation for a vegetated peatland. Our structural and process changes
34 to CLM focus on modifications needed to represent the hydrologic cycle of bogs
35 environment with perched water tables, as well as distinct hydrologic dynamics and
36 vegetation communities of the raised hummock and sunken hollow
37 microtopography characteristic of peatland bogs. The modified model was
38 parameterized and independently evaluated against observations from an
39 ombrotrophic raised-dome bog in northern Minnesota (S1-Bog), the site for the
40 Spruce and Peatland Responses Under Climatic and Environmental Change
41 experiment (SPRUCE). Simulated water table levels compared well with site-level
42 observations. The new model predicts hydrologic changes in response to planned
43 warming at the SPRUCE site. At present, standing water is commonly observed in
44 bog hollows after large rainfall events during the growing season, but simulations
45 suggest a sharp decrease in water table levels due to increased evapotranspiration
46 under the most extreme warming level, nearly eliminating the occurrence of
47 standing water in the growing season. Simulated soil energy balance was strongly

48 influenced by reduced winter snowpack under warming simulations, with the
49 warming influence on soil temperature partly offset by the loss of insulating
50 snowpack in early and late winter. The new model provides improved predictive
51 capacity for seasonal hydrological dynamics in northern peatlands, and provides a
52 useful foundation for investigation of northern peatland carbon exchange.

53

54

55 **1. Introduction**

56 Peatlands contain about 30% of global soil carbon, despite covering only 3%
57 of the Earth's land surface (Gorham, 1991; Bridgham et al., 2006; Tarnocai, 2009).
58 Northern peatlands play an important role in global carbon balance due to their
59 capacity to store carbon and to exchange both CO₂ and methane with the
60 atmosphere. Carbon accumulated over thousands of years in these systems is
61 projected to be vulnerable to climate warming (Wu et al., 2013). Manipulative
62 experiments and process-resolving models are needed to make defensible
63 projections of the net carbon balance of the northern peatlands in the face of a
64 warming global environment.

65 In this paper we focus on hydrological dynamics of peatlands, as the water
66 balance of peatlands plays a critical role in their carbon balance (Lafleur et al.,
67 2003). Kettridge et al. (2013) showed the importance of peatland hydrology as a
68 regulatory control on carbon dynamics and the future stability of peatland carbon
69 stocks and regional water dynamics. Seasonal and interannual fluctuations in water
70 table elevation can affect peatland net CO₂ exchange through complex effects on soil
71 processes (Mezbahuddin et al., 2013). Modeling by Grant et al. (2012) suggested
72 that the productivity of wetlands is strongly affected by changes in water table level,
73 and that the effects are complex and site-specific. Hydrologic dynamics can affect
74 tree growth, and modify the density, size, and species distribution in peatlands
75 (MacDonald and Yin, 1999; Robreck et al., 2009). Wu et al. (2013) showed that an
76 increase of water table level by 15cm could decrease net ecosystem production by
77 up to 200% and switch peatlands from a net sink to a net source of carbon. As

78 hydrology and biogeochemical cycling are tightly linked in peatlands (Waddington
79 et al., 2001; Silvola et al., 1996; Dise et al., 2011), the accuracy of predicted peatland
80 water table dynamics is likely to affect the accuracy of the predicted peatland
81 carbon exchange. To the extent they are used to evaluate carbon cycle – climate
82 system feedbacks, a reasonable requirement for land surface models operating
83 within global climate models should therefore be that they make reliable
84 predictions of peatland hydrology and hydrologic processes.

85 Peatland surfaces are often characterized by distinct micro-topography
86 (hollows and hummocks) (Nungesser, 2003). The existence of hummock-hollow
87 microtopography has important impacts on hydrological dynamics (Lindholm and
88 Markkula 1984; Verry et al., 2011b), nutrient availability (Chapin et al., 1979;
89 Damman, 1978), plant species distribution and productivity (Andrus et al. 1983;
90 Moore 1989), and decomposition rates (Johnson and Damman 1991). Many
91 wetland ecosystem models drive biogeochemical simulations using observed water
92 table depth as an input variable (St-Hilaire et al., 2010; Froking et al., 2002; Hilbert
93 et al., 2000). Even though such models include water table effects, the models have
94 not simulated observed variation for hummock/hollow microtopography common
95 to raised-dome bog peatlands. The absence of this important detail may limit the
96 predictive capabilities of existing peatland models. Other ecohydrological models
97 couple hydrology and carbon cycles in peatlands, but differ with respect to their
98 hydrological schemes and the way they treat (or ignore) topography (Dimitrov et al,
99 2011). Some models, such as Biome-BGC (Bond-Lamberty et al., 2007), and
100 Wetland-DNDC (Zhang et al., 2002) only simulate vertical soil water flow, neglecting

101 lateral flow components (Dimitrov et al, 2011) within peatlands. Wania et al.
102 (2010) describe a model of wetland hydrology and biogeochemistry (LPJ-WHyMe),
103 but do not include consideration of microtopography or lateral flows. Others, such
104 as BEPS (Chen et al., 2005, 2007) and InTEC v3.0 (Ju et al., 2006) include
105 sophisticated ecohydrological and biogeochemical sub-models capable of simulating
106 three-dimensional hydrology (for large scale topography) coupled to peatland
107 carbon dynamics. Sonnentag et al. (2008) further adapted BEPS to model the effects
108 of mesoscale (site level) topography on hydrology, and hence on CO₂ exchange at
109 Mer Bleue bog. Some advanced theoretical wetland models have included
110 ecohydrological feedbacks for the patterning on peatlands (Frolking et al., 2010; Mirris,
111 2013). Additionally, some cellular landscape models described by Swanson and Grigal
112 (1988), Couwenberg and Joosten (2005), Eppinga et al. (2009) and Morris et al. (2013),
113 dealing explicitly with fine-scale variability of peatland hydrology, have also been
114 applied to explore peat development. The model presented by Bohn et al. (2013) includes
115 fractional area representations for ridge and hollow in a wetland, but does not consider
116 explicit lateral fluxes between these microtopographic units. To the best of our
117 knowledge, only one ecosystem model currently includes representation of
118 microtopographic variability (hummock-hollow topography) with lateral
119 connection, that being the “ecosys” model (Grant et al., 2012). Ecosys tracks
120 horizontal exchange between hummock and hollow elements, but its prediction of
121 water table dynamics is constrained by specifying a regional water table at a fixed
122 height and a fixed distance from the site of interest (mainly applied for a fen
123 environment). Here we explore an extension of that approach, with lateral

124 connections between hummock and hollow elements, and with a more mechanistic
125 simulation of water table dynamics. Rather than specifying an external water table
126 height, we predict bog water table dynamics in part as a function of bog geometry,
127 including height of the bog's raised-dome center relative to a bog-scale drainage
128 element (lagg), relative surface height differences between hummock and hollow,
129 and fractional area contributions from hummocks and hollows. We implement this
130 new capability within the Community Land Model (CLM), with the aim of expanding
131 our simulations to large-scale bog simulations in subsequent studies.

132 The Community Land Model (CLM) (Oleson et al., 2013), the land component
133 of the Community Earth System Model (CESM), couples water, carbon, nitrogen, and
134 energy cycles together for the study of ecosystems. CLM does not currently
135 represent vegetated peatlands (or vegetated wetlands of any type), nor does it
136 represent lateral flow pathways common to surficial peats (Verry et al., 2011a, b).
137 To realistically represent the hydrological dynamics of raised-dome bog
138 microtopography in CLM, we incorporated structural and process changes
139 characteristic of vegetated peatlands. CLM without and with our new modifications
140 is hereafter referred to as CLM_Default and CLM_SPRUCE, respectively. A key
141 objective for this effort was to produce an enhanced CLM_SPRUCE capable of being
142 used for accurate simulations of high-carbon wetland hydrologic and carbon cycle
143 responses for application to plausible future climate conditions. SPRUCE, the
144 Spruce and Peatland Responses Under Climatic and Environmental Change
145 experiment, is a 10-year warming by elevated CO₂ manipulation of a high-carbon
146 forested peatland in northern Minnesota designed to provide information on

147 ecosystem changes under unique future warming and atmospheric conditions
148 (<http://mnspruce.ornl.gov>). The modified CLM model is parameterized from, and
149 independently evaluated against, observations from pre-treatment data sets for the
150 SPRUCE experiment and long-term peatland hydrology studies on the Marcell
151 Experimental Forest (MEF). The model improvements reported here represent the
152 first time that the isolated hydrologic cycle of an ombrotrophic bog, with its
153 characteristic raised hummocks and sunken hollows, has been represented in the
154 land surface component of an Earth system model. Hummock-hollow functionality
155 within CLM_SPRUCE allows for the simulation of defensible estimates of peatland
156 water table dynamics, necessary to predict dynamic CO₂ and CH₄ flux components
157 for peatland carbon cycle predictions.

158 **2. Site description and measurement**

159 Our study focuses on an ombrotrophic bog (a raised-dome peat bog in which
160 water and nutrient inputs originate from atmospheric sources). The specific study
161 site is a high-carbon, boreal peatland, which is located approximately 40 km north
162 of Grand Rapids, Minnesota, USA (N 47° 30.476'; W 93° 27.162' and 412 m above
163 mean sea level). The site is designated the S1-Bog and is situated within the S1
164 watershed (Fig. 1). The S1-Bog and watershed have been part of a long-term
165 research program of the USDA Forest Service Northern Research Station at the MEF
166 for over 50 years (Verry et al., 2011c).

167 The S1-Bog is an 8.1-ha *Picea-Sphagnum* bog that was harvested in two
168 successive strip cuts 5 years apart (1969 and 1974) (Sebestyen et al., 2011a). The
169 bog surface has a hummock/hollow microtopography with a typical relief of 10 to

170 30 cm between the tops of the hummocks and the bottoms of the hollows (Nichols
171 1998). The elevation of the hollows is fairly consistent throughout the S1-Bog, but
172 increases along a gentle slope to the highest point of the raised-dome near the
173 center of the bog (Verry, 1984; Richardson et al., 2010). The vegetation, climate,
174 hydrology, long-term monitoring, and post-European settlement site history are
175 described in Sebestyen et al. (2011a). Briefly, vegetation within the S1-Bog is
176 dominated by the tree species *Picea mariana* (Mill.) B.S.P and *Larix laricina* (Du Roi)
177 K. Koch, a variety of ericaceous shrubs, and *Sphagnum* sp. moss. Mean annual air
178 temperature is 3.4°C, and the average annual precipitation is 780 mm (Verry et al.,
179 2011d), with 75% of the precipitation occurring in the snow-free period from mid-
180 April to early November. Mean annual air temperatures have increased about 0.4°C
181 per decade over the last 40 years.

182 Peatlands at the MEF formed as ice-block depressions infilled over the past
183 11,000 years (Verry et al., 2011d). The peatlands are surrounded by gently sloping
184 upland mineral soils that drain toward the peatland. The peat deposit in the S1-Bog
185 is generally 2 to 4 m deep with maximum depths of 11 m (Parsekian et al., 2012). In
186 a typical year, the peatland water table fluctuates within the top 30-cm of peat
187 (Sebestyen et al., 2011b), which corresponds to peats that are least decomposed and
188 have the highest hydraulic conductivities (Verry et al., 2011a). As such, water flows
189 laterally through these highly conductive peats when water tables are near the peat
190 surface and the peatland water table is above the elevation of the peatland outlet.
191 The peatland has two hydrologically and vegetationally distinct zones: the bog and
192 the surrounding lag zone (Verry et al, 2011b). The central raised-dome bog radially

193 drains to the peatland perimeter (the lagg zone) when water tables are near the
194 peat surface (Fig. 2). Water flows into the peatland lagg from both the upland and
195 bog soils and the lagg coalesces into an outlet stream (Fig. 2). Streamflow is
196 intermittent, with flow occurring during snowmelt and after large rainfall events.
197 Some water does exit the peatland through lateral subsurface flow through a sand
198 berm that forms the southern boundary of the peatland, and through the bottom of
199 the ancient lake bed. The broadly-domed surface of the bog is characterized by a
200 microtopography of raised hummocks and sunken hollows. The mean height of the
201 bog surface above the level of the lagg is estimated as 0.7m to the hummock
202 surfaces, and 0.4m to the hollow surfaces, which is typical of raised-dome bog
203 structure in general.

204 Evapotranspiration (ET) is ~65% of annual precipitation in peatlands at the
205 MEF (Brooks et al., 2011). As a part of SPRUCE pretreatment measurement
206 protocols, water table levels have been measured every 30-minutes at the
207 meteorological stations (EM1 and EM2) that are approximately 3 m apart in the S1-
208 Bog (data and metadata are available at
209 ftp://mnspruce.ornl.gov/SPRUCE_EM_DATA_2010_2011/). Water levels were
210 recorded from TruTrack WT-VO water level sensors ($\pm\sim 2\text{mm}$ resolution,
211 <http://www.trustrack.com/WT-VO.html>) using Campbell Scientific CR1000
212 dataloggers. The two water level sensors were placed in hollows at EM1 and EM2.
213 Water table levels have been recorded since 2011 with the exception of periods of
214 frozen peat when the sensors are nonfunctional. In this study, water table height is
215 referenced to zero at the hollow surface. Positive values indicate standing water in

216 the hollows, and negative values indicate that water table is below hollow surfaces.

217 While measurements of ET are not available for the S1-Bog, annual ET
218 estimated on the basis of water budget measurements is available for a 21-year
219 period of record (1979-1999) at the nearby S2-Bog (Nichols and Verry, 2001). The
220 physical setting, vegetation types and water table dynamics of the S1 and S2 Bogs
221 are similar, except the S2-Bog did not undergo the strip cuts.

222 3. Model description and experiment design

223 3.1 Model description

224 We used the Community Land Model version 4.5 (CLM4.5) as the starting point
225 for our model development and evaluation. The new features of CLM4.5 (Oleson et
226 al., 2013) compared to its predecessor CLM4 include improved canopy processes,
227 soil hydrology updates, a new lake model, a vertically resolved soil biogeochemistry
228 scheme (Koven et al. 2013), a new fire model, a methane production, oxidation, and
229 emissions model (Riley et al. 2011) and an optional runoff generation scheme (Li et
230 al. 2011). Hydrology improvements in CLM4.5 include introduction of an ice
231 impedance function, surface water and other corrections that increase the
232 consistency between soil water saturated state and water table position, allowing
233 for the maintenance of a perched water table above permafrost layers (Swenson et
234 al., 2012).

235 The default CLM4.5 hydrology parameterizes interception, throughfall, canopy
236 drip, snow accumulation and melt, water transfer between snow layers, infiltration,

237 evaporation, surface runoff, sub surface drainage, vertical transport through the
238 vadose zone, and groundwater discharge and recharge (Fig. 7.1 in Oleson et. al,
239 2013). CLM4.5 also includes hydrologic and thermal properties for organic-rich
240 peat in addition to mineral soils. CLM4.5 does not include interactions between
241 horizontally variable soil columns, so no lateral flows are represented. The default
242 CLM4.5 parameterization for subsurface drainage produces an unrealistically deep
243 water table relative to observations in wetlands (Oleson et al., 2008). For this study
244 we use the thermal and hydraulic properties of peat as defined globally in CLM 4.5
245 (Lawrence and Slater, 2007) with the exception of the maximum subsurface
246 drainage rate, which is calibrated for the site ($q_{drain,0}$, see next section).

247 **3.2 New formulation for raised-dome bog hydrology**

248 Microtopography is simulated in CLM_SPRUCE by two interconnected soil
249 profiles representing the hummock and hollow areas, with the hollow surface set
250 0.3 m lower than that of the hummock, and with otherwise identical physical
251 properties with depth. The bog area is assumed to be 75% hummock and 25%
252 hollow, an approximation based on site measurements. We added several new
253 structure and process representations to CLM4.5 to improve correspondence with
254 observed features of the S1-Bog (Fig. 2, inset). Modifications include 1)
255 reformulation of the subsurface drainage term to represent horizontal subsurface
256 flow from the bog to the lagg which then drains to the outlet stream; 2) introduction
257 of a two-column structure to represent hummock/hollow microtopography; 3)
258 addition of lateral transport to represent saturated equilibration between the
259 hummock and hollow columns; 4) introduction of surface runoff from hummocks to

260 hollows; and 5) drainage of surface water in the hollows to the lagg. Our intent with
261 this particular set of modifications was to capture the physical and hydrological
262 processes that distinguish an ombrotrophic bog from the more general soil
263 hydrology representation in the original model.

264 The original formulation for subsurface runoff uses a TOPMODEL-based
265 approach for surface and subsurface runoff (Niu et al., 2005). Subsurface drainage
266 rate q_{drai} exponentially decays with water table depth:

$$q_{drai} = q_{drai,0} \exp(-f_{drai} (z_w)) \quad (1)$$

267 Here, $q_{drai,0}$ is the maximum subsurface drainage rate ($\text{Kg m}^{-2} \text{s}^{-1}$), which occurs
268 when the water table is at the surface. f_{drai} is the exponential decay factor (m^{-1}), and
269 z_w is the depth of the water table below the surface (m). For our new model we use
270 the default global value $f_{drai} = 2.5 \text{ m}^{-1}$ from CLM4.5 (Oleson et al. 2013), but we
271 modify equation (1) such that the subsurface drainage term becomes zero when the
272 water table depth drops to the local depth of the lagg z_{lagg} (0.4m relative to the
273 hollow surface and 0.7m relative to the hummock surface as a mean value for the
274 S1-Bog):

$$q_{drai} = q_{drai,0} \left(\exp(-f_{drai} (z_w)) - \exp(-f_{drai} (z_{lagg})) \right) , z_w < z_{lagg}$$
$$q_{drai} = 0, \text{ otherwise} \quad (2)$$

275 This model parameterization represents an assumed barrier to lateral and vertical
276 drainage imposed by the glacial till layer when the water table is below the level of
277 the lagg. At the S1-Bog, there is an observed small but persistent “deep seepage”

278 term representing vertical subsurface drainage from the perched bog water table
279 (included in Fig. 2, but set to zero for simulations here, due to lack of adequate data
280 for parameterization). For this study, the maximum subsurface drainage rate $q_{\text{drai},0}$
281 is calibrated against site water table observations.

282 For modification (3), the simulated lateral transport of water, $q_{\text{lat},\text{aqu}}$ (mms⁻
283 ¹), is a function of the difference in simulated water table level between the
284 hummock and hollow columns, the specific yield of the soil layer in question, the
285 average hydraulic conductivity and the average horizontal distance between
286 columns:

$$q_{\text{lat},\text{aqu}} = \frac{\left(\frac{k_{\text{hum}} + k_{\text{hol}}}{2} (z_{\text{w},\text{hum}} - z_{\text{w},\text{hol}}^*) \right)}{\Delta x} \quad (3)$$

287 k_{hum} (mm s⁻¹) is the weighted mean saturated hydraulic conductivity of the
288 hummock layers containing the aquifer, and k_{hol} (mm s⁻¹) is the weighted mean
289 saturated hydraulic conductivity hollow layers containing the aquifer. We used
290 CLM4.5 default peatland saturated hydraulic conductivity values, which decrease as
291 a function of depth (Oleson et al. 2013) and fall in the observed range for bogs in the
292 region (Verry et al. 2011a). Δx is the horizontal separation between the hummock
293 and hollow columns, which is assumed to be 1 meter. Variables $z_{\text{w},\text{hum}}$ and $z_{\text{w},\text{hol}}^*$
294 represent the hummock and adjusted hollow water table depths (meters) relative to
295 the hollow surface. The adjusted hollow water table depth $z_{\text{w},\text{hol}}^*$ reflects a reduction
296 in water table depth by the height of surface water that is present on the hollow
297 surface. To transport water laterally between hummock and hollow, we first use

298 CLM's calculation of specific yield for the soil layer containing the water table and
299 determine the difference in water table height resulting from the lateral flux. If the
300 magnitude of the lateral flux is larger than the capacity of that layer, the water table
301 may move into higher or lower layers using the same relationship between specific
302 yield of those layers and water table height. Transport from hollow surface water
303 into the hummock soil column may also occur.

304 Modification (4) involves directing the surface runoff term calculated on the
305 hummock as an input term to the hollow surface. Because of the large infiltration
306 capacity of peat, this term is most relevant when the upper peat layers of the
307 hummock are frozen.

308 The implementation of surface water storage and runoff in CLM 4.5
309 considers microtopography across an entire grid cell rather than within the wetland
310 portion of a grid cell, and does not account for the effects of peatland
311 microtopography (Oleson et al., 2013). Here we assume that the hollows are
312 interconnected, and the surface water runoff from the hollows is determined by the
313 slope of the raised dome bog and the surface water height. Therefore for
314 modification (5), we replace the formulation of surface water runoff using the
315 formulation for wetland flow by Kadlec and Knight (2009) that includes a vertical
316 stem density gradient and a bottom slope, modified by Kazezyilmaz-Alhan et al.
317 (2007):

$$q_{h2o,sfc} = r_{h2osfc} z_{h2osfc}^2 \quad (4)$$

318 Here, $q_{h2o,sfc}$ is the surface water drainage rate ($\text{kg m}^{-2} \text{s}^{-1}$) and z_{h2osfc} is the surface
319 water height in the hollow (m). The parameter r_{h2osfc} is an aggregated coefficient
320 that includes both vegetation-induced friction and the bottom slope of the hollows
321 in the raised dome bog. This parameter is calibrated against observations to
322 improve model performance at the S1-Bog (results in Table 1).

323 Our current implementation of CLM_SPRUCE does not include a unique
324 biophysical parameterization for Sphagnum moss, which is a recognized
325 shortcoming. Other efforts are underway to quantify the unique hydraulic and
326 physiological properties of moss, including field studies in the S1-Bog and
327 laboratory studies based on S1-Bog samples. Introduction of lateral connectivity and
328 bog geometry and microtopography are first-order steps toward improved
329 representation of peatland hydrology. We intend to include new parameterizations
330 emerging from observational and experimental efforts in subsequent work with
331 CLM_SPRUCE.

332 **3.3 Simulation experiment setup**

333 Picea, Larix, and shrubs are represented by the corresponding CLM plant
334 functional types. Because Sphagnum moss physiology is not represented in CLM, we
335 use the C3 grass plant functional type to represent both sedges and Sphagnum moss.
336 Both hummock and hollows include the same vegetation distributions. Simulations
337 were conducted using CLM_Default and CLM_SPRUCE with prescribed vegetation
338 canopy phenology. To capture site evapotranspiration from vegetation, the
339 maximum leaf area indices were based on site observations. Several model

340 parameters were set to match site observations, including leaf C:N ratios, rooting
341 depth profiles and specific leaf area (Table 1). Since this study focuses on the site
342 hydrology, biogeochemistry is turned off in the model to avoid computationally
343 costly carbon pool spinups, and carbon fluxes have not been tracked for these
344 annual hydrologic simulations.

345 Half-hour SPRUCE environmental driver data are being collected and are
346 available since 2011, but a longer data sequence was needed for model simulations.
347 The model is driven by 35-year (1979-2013) environmental reanalysis data from
348 NCEP2 (Kanamitsu et al., 2002) including temperature, precipitation, specific
349 humidity, solar radiation, wind speed, pressure and long wave radiation at a 6-hour
350 time step and extracted for the gridcell containing the S1-Bog. NCEP precipitation
351 was rescaled using daily precipitation data from a recording rain gage in the nearby
352 South Meteorological Station at the MEF (Sebestyen et al., 2011a).

353 The 10-year long SPRUCE climate change field experiment at the S1-Bog will
354 consist of combined manipulations of temperature (various differentials up to 9K
355 above ambient) and CO₂ concentration (ambient and 800-900 ppm). To investigate
356 how the bog water table levels in hummock/hollow microtopography may respond
357 to different warming scenarios, we performed 8 simulations from the same starting
358 point in year 2000, designed to reflect the warming treatments being implemented
359 in the field. The model simulations include a control simulation (CTL), and six
360 simulations with increasing air temperature (+3K, +6K, and +9K above ambient,
361 respectively) under two humidity conditions. Three of these six simulations used
362 the same specific humidity (Q) as CTL, which will be referred to here as 'warming

363 and constant Q' . The other three simulations used the same relative humidity (RH)
364 as CTL, (and hence, due to warmer air temperatures, higher Q), denoted here as
365 'warming and constant RH'. The final simulation increased air temperature by +9K
366 and specific humidity by 30%, which is lower than the constant RH scenario. This
367 final humidity setting is based on the projection of CESM under RCP8.5 scenario at
368 the end of 21st century (Moore et al., 2013). We note that the warming and
369 constant Q scenario most closely represents the planned experimental manipulation
370 at SPRUCE, since there will be no water vapor additions. The treatments for the
371 SPRUCE field experiment will include belowground soil warming achieved with
372 vertical heating elements (Hanson et al., 2011). The purpose of the belowground
373 heating is to compensate for subsurface heat loss around the edges of the
374 aboveground enclosures. Since CLM_SPRUCE does not account for lateral heat flow,
375 the planned SPRUCE active belowground heat manipulations are not included in the
376 current simulations. To estimate incoming longwave radiation under the warming
377 scenarios, we use clear-sky assumptions about atmospheric temperature, humidity,
378 and emissivity (Idso et al. 1981)

379 Parameter calibrations for q_{drai} and r_{h2osfc} are performed jointly using a
380 genetic algorithm (Thomas and Yao, 2000) requiring 1000 simulations, and
381 optimizes the model against the daily observed water table level from 2011 and
382 2012. Observations from the year 2013 are used for evaluation. The calibrated
383 model with our new modifications is then compared with the observations and used
384 to predict future scenarios.

385 As an independent evaluation of the modeled relationship between annual

386 evapotranspiration (ET) and annual air temperature (T), we compared model
387 results for the 21-year period 1979-1999 with observations of ET from the nearby
388 S2-Bog. We further explored the ET vs. T relationship over the range of warming
389 treatments, in an effort to place some confidence bounds on our model findings
390 regarding changes in bog hydrology under experimental warming.

391

392 **4. Results**

393 **4.1. Simulated water table level**

394 Simulations with (CLM_SPRUCE) and without (CLM_Default) our new
395 hydrological treatment are used to test the influence of new model representations
396 of hydrological processes at the microtopographic level of peatland hummocks and
397 hollows. CLM_Default produces a water table depth of 3-4 m (Fig. 3a), which can be
398 considered representative of the regional water table in the upland and below the
399 bog (Verry et al. 2011b), but is not realistic of the perched water table in the bog
400 itself. Reformulating lateral drainage flow from the bog to the lagg as a function of
401 the height difference between the simulated bog water table and the lagg outlet,
402 CLM_SPRUCE simulates a water table depth of <1m (a perched water table, Fig. 3a).
403 CLM_SPRUCE simulates independent water tables for the hummock and hollow bog
404 elements, but by parameterizing near-surface and sub-surface hydraulic
405 connectivity between hummock and hollow, the water tables in these two elements
406 track each other on short time scales (Fig. 3b). The small differences between
407 hummock and hollow water tables occur during large precipitation events.
408 CLM_SPRUCE simulates standing water in the hollow during snowmelt and after

409 large precipitation events, with drying of the hollows due to drainage to the lagg and
410 evapotranspiration. In the summer of 2012 a prolonged period of low precipitation
411 resulted in a simulated water table decline to approximately 30 cm below the
412 surface of the hollow. Mean annual water budget predicted by the model has ET as
413 57.48% of annual precipitation, in reasonable agreement with the observed value of
414 65% (Fig.8).

415 Time series observations of water table height from two sensors (EM1 and
416 EM2) located within the S1-Bog and separated by ~3 m were available for parts of
417 calendar years 2011, 2012, and 2013. The water table depth data from these
418 sensors are in good agreement with water table data from 10 additional sensors
419 distributed across the S1-Bog in 2014. Data from 2011 and 2012 were used to
420 parameterize the new lateral drainage terms in CLM_SPRUCE (Table 1), with
421 observations from 2013 used for evaluation. The model CLM_SPRUCE captures the
422 timing and magnitude of observed water table dynamics in 2011 and 2012, with
423 some periods of underestimation in 2011 and overestimation in 2012, but no clear
424 indication of a consistent prediction bias. Water table height predictions for the
425 evaluation year, 2013, are in good agreement with observations ($R^2 = 0.51$) for both
426 timing and magnitude (Fig. 4).

427 **4.2 Simulated hydrologic response to climate warming**

428 **4.2.1 Influence of warming and humidity changes on simulated water table** 429 **heights**

430 Simulated warming of the bog through an imposed increase in near-surface
431 air temperature results in model prediction of drying and a deeper water table (Fig.
432 5). The magnitude of warming effects on water table height is influenced strongly by
433 assumptions regarding changes in humidity. For the case where absolute humidity
434 (Q , kg H₂O/kg dry air) is unchanged (in comparison to control), all warming
435 treatments (+3, +6, and +9 K) cause a deepening of water table level (Table 2), with
436 deepening of ~15cm year-round for the +9K scenario (Figure 5a). Under this
437 scenario the system shifts from frequent periods of standing water in the hollows in
438 spring and following large precipitation events (CTL), to an almost complete
439 absence of standing water periods (+9K). The mean state of the water table in
440 summer months for the +9K case with constant Q is lower than the deepest water
441 table exhibited in the control scenario under dry conditions in the summer of 2012.
442 Under the alternative assumption for humidity changes, in which relative humidity
443 is maintained as in the control (requiring progressive increases in absolute
444 humidity under +3, +6, and +9K warming scenarios), water table height is lowered
445 only on average by ~5cm, with some evidence of slower recovery from deeper
446 water table at the end of the summer 2012 dry period (Fig. 5b). The planned
447 experimental manipulations for the SPRUCE chambers will consist of increased air
448 temperature but no additions of water vapor (due to cost constraint), so the
449 eventual experimental conditions will be close to the assumptions shown here for
450 the constant Q case (Fig. 5a). Earth system model predictions for future climate
451 change actually fall somewhere between the two end-point cases illustrated in
452 Figure 6. Based on results from the CESM for a future radiative forcing of 8.5 W/m²,

453 which generates a regional near-surface air temperature increase of almost 9K by
454 year 2100 and so seems a reasonable candidate global simulation for this purpose,
455 the regional specific humidity increased by 30%, corresponding to a 14% decrease
456 in relative humidity. Evaluating CLM_SPRUCE results when forced with this example
457 climate model projection, we find that the projected response for the end of the
458 century falls between the two endpoint simulations already shown, and is generally
459 closer to our constant Q case than to our constant RH case (Fig. 6).

460 **4.2.2 Influence of warming on simulated evapotranspiration**

461 Since the constant Q scenarios most closely follow the planned experimental
462 treatment, we explore evapotranspiration (ET) and its components response to the
463 warming with only these simulations. ET in CLM_SPRUCE is the sum of three
464 components: transpiration (TR), canopy evaporation (E_c) and soil evaporation (E_s).
465 These modeled water budget terms include tree, shrub, and herbaceous vegetation.
466 ET and its components increase with air temperature for both hummock and
467 hollow under air temperature increased by +3K, +6K and +9K warming scenarios,
468 whose magnitudes scale with the increases of temperature (Table 2). ET is
469 predicted to increase by 53.24%, 76.7% and 87.61% for the hummock under the
470 three warming scenarios (by 61.25%, 91.5% and 116.35% for hollow), respectively.
471 Soil evaporation shows the biggest percentage increase with warming, especially
472 soil evaporation from the hollows. For example, evaporation from hollows increased
473 by about 132%, 198%, and 256% when the air temperature increased by +3K, +6K
474 and +9K, respectively. Canopy evaporation shows the smallest changes with the
475 three different increases of air temperature (Table 2). The seasonal pattern of

476 transpiration shows that warming causes higher simulated transpiration
477 throughout the growing season, with the largest absolute increases in mid-summer
478 (Fig. 7, top row). Three year averaged time evolution of canopy evaporation
479 demonstrates that E_c is little affected by warming in these simulations, indicating
480 that temperatures and incident radiation are adequate to evaporate most of the
481 canopy intercepted precipitation even in the control simulation (Fig. 7, middle row).
482 Evaporation from the bog surface (E_s) in these simulations is increased under the
483 warming treatments throughout the year, with the largest increases in late winter
484 and spring (Fig. 7, bottom row). At the highest levels of warming, simulated E_s is
485 sometimes reduced compared to moderate warming in the late summer, due to
486 reduced hydraulic conductivity for the dried upper layers of the soil. While
487 observations of ET are not available for the S1-Bog site, a 21-year record of ET
488 based on water budget observations at the nearby S2-Bog provides a valuable basis
489 for evaluation of our simulation results. Comparing the predicted and observed
490 relationship of interannual ET to air temperature for a pre-warming simulation, we
491 find that the simulated and observed slopes are nearly identical for the period 1979-
492 1999, although the mean ET predicted at the S1-bog is about 15% lower than the
493 mean observed at the S2-Bog (Fig. 8). Examining the same relationship for a 14-year
494 record of simulated warming at multiple warming levels, we find that the slope of
495 the interannual relationship is consistent across the warming treatments and
496 similar to the control period, although the strength of the interannual relationship
497 weakens with higher levels of warming. Taken together, the control and warming
498 experiments define a broad and approximately linear relationship between ET and

499 T with a slope that does not depart dramatically from the observed or modeled
500 interannual relationships (Fig. 8). The current model does not consider Sphagnum
501 physiology, which may help to explain the underestimation of ET since Sphagnum
502 lack stomata while the model includes stomatal regulation. The S2 bog estimates of
503 ET also include some contribution from upland vegetation, which could further
504 contribute to bias in the model-data comparison. The model predicted summertime
505 (June-August) ET rate is 2.4 mm/d, well within the range of values reported from
506 other peatlands in the northern latitudes (Moore et al., 1994; Lafleur et al., 1997;
507 Wu et al., 2010)

508 **4.3 Influence of warming on simulated snow dynamics and soil temperature**

509 The S1-Bog (and surrounding region) is subject to snowpack accumulation,
510 with a persistent snowpack commonly observed for the period November to April.
511 Our control simulation reproduces this observed behavior (results not shown).
512 Since snow is a good thermal insulator (e.g. Ge and Gong, 2010), and since a thick
513 snowpack occurs during the coldest part of the year, the observed average soil
514 temperature in the bog is warmer than average air temperature, a pattern also
515 reproduced by our control simulation.

516 In our warming simulations, higher air temperatures lead to a reduced
517 snowfall amount (some precipitation that falls as snow in the control simulation
518 falls as rain in the +9K warming simulation) and increased snow melt, both of which
519 contribute to a reduced snowpack depth, with the effect concentrated during the
520 period of typical highest snowpack accumulation in the late winter and spring (Fig.
521 9). The simulated influence on near-surface soil temperature of this modification of

522 snowpack is dramatic, with very little difference between control and +9K
523 treatment for the period January – February, warming effects increasing to a
524 maximum over the period March – April, then declining to an intermediate level of
525 warming which persists through the summer and into fall (Fig. 9). Reduced
526 insulating effect of the thinner and more intermittent snowpack in the +9K
527 simulation allows more cooling of the soil during very cold periods, negating the
528 effect of increased air temperature. The influence of warming on deeper soil
529 temperature (shown in Fig. 9 for a layer averaging 3.0 m deep) is much more
530 consistent through the seasons: it is this deep soil temperature offset which sets the
531 thermal baseline toward which shallower soil layers are drawn in the snow-free
532 season, resulting in summer and fall near-surface soil temperatures that are less
533 than the imposed warming of air temperature. The loss of insulation also results in
534 more variability and more extremes in soil temperature.

535 **5. Discussion**

536 The current study moves us closer to our long-term goal by improving the
537 prediction of peatland water table depth in CLM, and by advancing the state of
538 peatland ecosystem modeling within land surface models by introducing a
539 formulation for the prediction of bog water table depth that does not depend on an
540 externally forced regional water table. Our laterally-coupled two-column hydrology
541 scheme is a first-order approximation of the real bog's undulating hummock-hollow
542 microtopography, and provides a basis for evaluating differences in vegetation
543 distribution or function and differences in sub-surface biogeochemical processes as
544 they exist pre-treatment and as they may evolve under experimental manipulation.

545 Simulations presented here suggest that the hydrologic cycle within the S1-
546 bog will respond to increased air temperature under the planned warming
547 experiment and expected under projected climate change. Specifically, water table
548 levels are expected to drop with increased air temperature as a result of increased
549 evapotranspiration. However, the predicted reduction in water table level depends
550 strongly on the level of warming and on the details of humidity modification. The
551 warming influence on water table depth is expected to be larger for the anticipated
552 experimental manipulation (close to constant Q) than would be the case if the
553 experimental manipulation included injection of water vapor with heating of near-
554 surface air (maintaining constant RH).

555 The predicted influence of warming on ET under as assumption of constant
556 specific humidity is quite dramatic at the higher warming levels, producing a
557 significant drop in the simulated water table. Our evaluation of predicted ET and its
558 sensitivity to air temperature indicates that the model produces a very realistic
559 response of ET to temperature variation on interannual to decadal timescales. While
560 we do not yet have any observations from the experimental warming treatments,
561 we are able to show that the simulated response under those warming treatments
562 follows an approximately linear extension of the response in the control period.
563 Based on these preliminary evaluations, we do not have any particular reason to
564 suspect that the simulated response of ET to warming is departing in an unrealistic
565 way from the behavior of the real system. It is remarkable to note that at the highest
566 warming levels nearly all of the annual precipitation is being evaporated, with only a
567 few percent leaving as runoff. This suggests a fundamental shift in the character of

568 the bog under levels of warming approximating “business-as-usual” climate change
569 scenarios. Perched water tables would likely decline under long-term exposure to
570 these environmental conditions, if our model predictions are correct. Evaluations
571 against observations at other bog sites, in particular for sites instrumented for eddy
572 covariance estimates of ET, will be an important next step in our model evaluation
573 efforts.

574 The interactions of air warming with snowpack and soil temperature
575 simulated by CLM_SPRUCE raise some interesting challenges for the eventual
576 interpretation of results from the SPRUCE warming experiment. Based on results
577 presented here, we expect soil warming to be less than near-surface air warming in
578 systems with consistent over-winter snowpack, under a scenario of radiatively-
579 forced climate change. Since the experimental protocol for warming at the SPRUCE
580 field site includes active below-ground heating elements and the maintenance of
581 differential set points for below-ground temperature that match the air-warming
582 differentials, the differences between soil warming and near-surface air warming
583 expected in nature will be attenuated in the experimental plots. Our modeling
584 results suggest that extra energy will be added by the belowground control system
585 to offset the effect of reduced thermal insulation due to smaller and shorter duration
586 snowpack. This energy source belowground could drive additional interactions with
587 snowpack and other aspects of hydrologic cycle in the heated plots.

588 In addition to simulations aimed at improved understanding of bog response
589 to experimental manipulations at the plot-scale, we are also pursuing model
590 implementations at larger spatial scales. By extending our simulation framework to

591 include the entire bog domain, we will be able to evaluate large-scale hydrology
592 against streamflow measurements from S1 and nearby bogs. We are already
593 exploring the use of high-resolution gridded domains with explicit vertical and
594 lateral flows as a foundation for more highly parameterized simulations that could
595 allow us to estimate water, energy, and greenhouse gas fluxes for large landscapes
596 in which peatland bogs are an important component. High-resolution elevation and
597 remote sensing information could be incorporated into these simulations to derive
598 model parameters associated with microtopography, surface runoff and subsurface
599 drainage such as lagg elevation. Since the CLM framework is already well suited to
600 simulations in the upland regions of these domains, our current progress on
601 simulating bog hydrology places this large-scale simulation goal within reach.

602 The success demonstrated here with a two-column parameterization of
603 vertical and lateral hydrologic connectivity in a landscape characterized by small-
604 scale topographic variation suggests that similar approaches may be useful in
605 simulations of other similar landscapes. For example, our team is actively exploring
606 multi-scale simulation approaches for representing geophysical, hydrological,
607 biogeochemical, and botanical dynamics in Arctic polygonal tundra underlain by
608 permafrost (e.g. Painter et al., 2013). With a modest increase in complexity we
609 expect to be able to represent water table dynamics over regions of hundreds of m²
610 by tracking the vertical and lateral connectivity of a small number of columns
611 representing, for example, polygon edges and centers. Parameterization for such
612 coarse representations is being developed through very fine-scale process-resolving
613 thermal-hydrology simulation (e.g. Painter and Karra, 2014).

614 Simulating hydrological dynamics for microtopography of hummocks and
615 hollows in the raised-dome bog environment is a necessary step toward more
616 complete process representation that connects hydrology with vegetation dynamics,
617 soil biogeochemistry, and estimation of greenhouse gas fluxes under changing
618 climate conditions. Our initial efforts have focused on improvements in the
619 modeling of peatland hydrology, but our ultimate goal is to integrate new modeling
620 tools with observed ecosystem characteristics and results from experimental
621 manipulations to understand the interactions of climate, hydrology, vegetation
622 physiology, and biogeochemical cycling in these carbon-rich systems. Warming
623 temperatures and shifting precipitation patterns have the potential to alter all
624 aspects of these interactions, including the possibility of shifting peatland systems
625 from net sinks to net sources of carbon (Limpens et al., 2008; Ise et al., 2008).

626 It has been suggested that the *Sphagnum* layer contributes significantly to
627 total ecosystem CO₂ flux (Oechel and Van Cleve, 1986), and thus plays an important
628 role in the functioning of peatland ecosystem. Our current model work does not
629 include a moss plant functional type and instead uses C3 grass as a proxy, which
630 introduces potential biases. Mosses lack stomata, and the conductance to CO₂
631 diffusion is controlled by a passively variable water layer (Silvola, 1990; Williams
632 and Flanagan, 1996). Work is underway to introduce a new moss plant functional
633 type in CLM_SPRUCE, and we will use observations being gathered from the S1-Bog
634 to parameterize the influence of water content on *Sphagnum* photosynthesis, and to
635 better understand the influence of moss on hydrological and biogeochemical
636 conditions in peatland bogs. Previous efforts at synthesizing and modeling moss

637 physiology and physical properties are informing our progress in this area (St-
638 Hilaire et al., 2010; Turetsky et al., 2012)

639

640 **6. Conclusions**

641 In this study, the CLM model (Oleson et al., 2013) was modified to explicitly
642 simulate hydrological dynamics for the microtopography of hummocks and hollows
643 in a raised bog environment (CLM_SPRUCE). The model was evaluated against half-
644 hourly measurements of daily water table levels for 3 years. CLM_SPRUCE captures
645 annual mean and seasonal dynamics in water table levels reasonably well, and
646 eliminates deep water table and exaggerated seasonal dynamics biases associated
647 with the default version of CLM4.5. The model reproduces the relationship between
648 interannual ET and air temperature as observed at a nearby site. We used the new
649 model CLM_SPRUCE to investigate the hydrological responses to different warming
650 and humidity scenarios. Based on those simulations, we predict a deepening of the
651 bog water table for the highest warming treatment (+9K) planned in the SPRUCE
652 experiment, greatly reducing the occurrence of standing water in the hollows. We
653 estimate that the observed relationship between ET and air temperature will hold
654 under conditions of experimental warming at levels out to +9K. We also predict a
655 strong interaction between the air heating treatment and the thickness and duration
656 of snowpack, with consequences for subsurface temperatures that depend on
657 snowfall amounts and mean winter temperature. These modeling results have
658 helped raise awareness of the influence of operational decisions regarding over-
659 winter heating of the sub-surface in the experimental design.

660

661 **Acknowledgements**

662 This material is based upon work supported by the U.S. Department of Energy,
663 Office of Science, Office of Biological and Environmental Research. Oak Ridge
664 National Laboratory is managed by UT-Battelle, LLC for the U.S. Department of
665 Energy under Contract No. DE-AC05-00OR22725. The long-term research program
666 at the MEF is funded by the Northern Research Station of the USDA Forest Service

667

668 **References**

669

670 Andrus, R., Wagner, D. J., and Titus, J. E.: Vertical distribution of Sphagnum mosses along
671 hummock-hollow gradient, Canadian Journal of Botany, 61, 3128-3139, 1983.

672

673 Bridgham, S. D., Megonigal, J. P., Keller, J. K., Bliss, N. B., and Trettin, C.: The carbon balance
674 of North American wetlands, Wetlands, 26, 889-916, 2006.

675

676 Bond - Lamberty, B., Gower, S. T., and Ahl, D. E.: Improved simulation of poorly drained
forests using Biome - BGC, Tree Physiol., 27, 703-715, 2007.

677

678 Brooks, K. N., Verma, S. B., Kim, J., and Verry, E. S.: Scaling up evapotranspiration estimates
679 from process studies to watersheds, in Peatland biogeochemistry and watershed hydrology at
680 the Marcell Experimental Forest, edited by: Kolka, R. K. Sebestyen, S. D., Verry, E. S., and
681 Brooks, K.N., CRC Press, New York, 177-192, 2011.

682

683 Bohn, T. J., Podest, E., Schroeder, R., Pinto, N., McDonald, K.C., Glagolev, M., Filippov, I.,
684 Maksyutov, S., Heimann, M., Chen, X., and Lettenmaier, D.P.: Modeling the large- scale
685 effects of surface moisture heterogeneity on wetland carbon fluxes in the West Siberian
686 Lowland, Biogeosciences, 10, 6559-6576, doi: 10.5194/bg-10-6559-2013.

687

688 Chapin, F.S., Van Cleve, K., Chapin, M. C.: Soil temperature and nutrient cycling in the tussock
689 growth form of Eriophorum vaginatum, Journal of Ecology, 67, 169-189, 1979.

690

691 Couwenberg, J., and Joosten, H.: Self-organization in raised bog patterning: the origin of
microtope zonation and mesotope diversity. Journal of Ecology, 93, 1238-1248, 2005.

692

693 Damman, A. W. H.: Distribution and movement of elements in ombrotrophic peat bogs, Oikos,
30, 480-495, 1978.

694

695 Dimitrov, D. D., Grant, R. F., Lafleur, P. M., and Humphreys, E. R.: Modeling the effects of
696 hydrology on gross primary productivity and net ecosystem productivity at Mer Bleue bog, J.
Geophys. Res., 116, G04010, doi:10.1029/2010JG001586, 2011.

697

698 Dise, N., Shurpali, N. J., Weishampel, P., Verma, S. B., Verry, E. S., Gorham, E., Crill, P. M.,
699 Harriss, R.C., Kelley, C.A., Yavitt, J.B., Smemo, K.A., Kolka, R.K., Smith, K., Kim, J.,
700 Clement, R.J., Arkebauer, T.J., Bartlett, K.B., Billesbach, D.P., Bridgham, S.D., Elling, A.E.,
701 Flebbe, P.A., King, J.Y., Martens, C.S., Sebacher, D.I., Williams, C.J., Wieder, R.K.: Carbon
702 emissions from peatlands, in Peatland biogeochemistry and watershed hydrology at the
703 Marcell Experimental Forest, edited by: Kolka, R. K. . Sebestyen, S. D, Verry, E. S., and
704 Brooks, K.N, CRC Press, New York, 297-347, 2011.

705

706 Eppinga, M.B., de Ruyter, P.C., Wassen, M.J., Rietkerk, M.: Nutrients and hydrology indicate the
707 driving mechanisms of peatland surface patterning. The American Natu- ralist, 173, 803-818,
708 2009

709

710 Froelking, S., Roulet, N. T., Moore, T. R., Lafleur, P. M., Bubier, J. L., and Crill, P. M.: Modelling
711 the seasonal to annual carbon balance of Mer Bleue bog, Ontario, Canada, Global Biogeochem.
C., 16, doi:10.1029/2001GB001457, 2002.

- 712 Frolking, S., Roulet, N.T., Tuittila, E., Bubier, J.L., Quillet, A., Talbot, J., Richard, P.J.H.: A new
713 model of Holocene peatland net primary production, decomposition, water balance, and peat
714 accumulation. *Earth System Dynamics*, 1, 1-21.
- 715 Gorham, E.: Northern peatlands: Role in the carbon cycle and probable responses to climatic
716 warming, *Ecol. Appl.*, 1, 182-195, 1991.
- 717 Ge, Y., and Gong, G.: Land surface insulation response to snow depth variability, *J. Geophys.*
718 *Res.*, 115, D08107, doi:10.1029/2009JD012798, 2010.
- 719 Grant, R. F., Desai, A. R., and Sulman, B. N.: Modelling contrasting responses of wetland
720 productivity to changes in water table depth, *Biogeosciences*, 9, 4215-4231, 2012.
- 721 Hilbert, D.W, Roulet, N., Moore, T.R.: Modelling and analysis of peatlands as dynamical
722 systems. *Journal of Ecology*, 88, 230-242, 2000.
- 723 Hanson, Paul J., Kenneth W. Childs, Stan D. Wullschleger, Jeffery S. Riggs, Warren K. Thomas,
724 Donald E. Todd, and Jeffrey M. Warren (2011), A method for experimental heating of intact
725 soil profiles for application to climate change experiments, *Global Change Biol.*, 17(2), 1083-
726 1096.
- 727 Idso, S. B.: A Set of Equations for Full Spectrum and 8-Mu-M to 14-Mu-M and 10.5-Mu-M to
728 12.5-Mu-M Thermal-Radiation from Cloudless Skies, *Water Resour Res.*, 17(2), 295-304,
729 1981.
- 730 IPCC: The physical science basis. Contribution of working group I to the Forth Assessment
731 Report of the Intergovernmental Panel on Climate Change, 996 pp., Cambridge University
732 Press, Cambridge, , 2007.
- 733 Ise, T., Dunn, A. L., Wofsy, S. C., and Moorcroft, P. R.: High sensitivity of peat decomposition
734 to climate change through water-table feedback, *Nat. Geosci.*, 1(11), 763-766, 2008.
- 735 Johnson, L. C., and Damman, A. W. H.: Species-controlled Sphagnum decay on a south Swedish
736 raised bog, *Oikos*, 61, 234-242, 1991.
- 737 Kanamitsu M., Ebisuzaki, W., Woollen, J., Yang, S., Hnilo, J. J., Fiorino, M., Potter, G. L.:
738 NCEP-DOE AMIP-II Reanalysis (R-2), *Bull.Amer. Meteor. Soc.*, 83 1631-43, 2002.
739
- 740 Kadlec, R. H. and Knight, R.L.: *Treatment Wetlands (second edition)*, CRC Press, Boca Raton,
741 Fla., 21-57, 2009.
742
- 743 Kazezyilmaz-Alhan, C. M., Medina Jr., M.A, and Richardson C. J.: A wetland hydrology and
744 water quality model incorporating surface water/groundwater interactions, *Water Resour.*
745 *Res.*, 43, W04434, doi:10.1029/2006WR005003, 2007.
746
- 747 Kolka, R.K., Nater, E. A., Grigal, D.F., and Verry, E. S.: Atmospheric inputs of mercury and
748 organic carbon into a forested upland/bog watershed, *Water Air and Soil Pollution*, 113, 273-
749 294, 1999.
750
- 751 Koven, C. D., Riley, W. J., Subin, Z. M., Tang, J., Torn, M. S., Collins, W. D. Bonan, G. B.,
752 Lawrence, D. M., and Swenson, S. C.: The effect of vertically resolved soil biogeochemistry

753 and alternate soil C and N models on C dynamics of CLM4, *Biogeosciences*, 10, 7109-7131,
754 doi:10.5194/bg-10-7109-2013, 2013.
755

756 Lindholm, T., and Markkula, I.: Moisture conditions in hummocks and hollows in virgin and
757 drained sites on the raised bog Laaviosuo, Southern Finland, *Ann. Bot. fenn.*, 21, 241-255,
758 1984.
759

760 L’vovich, M. I., Whit, G. F.: Use and transformation of terrestrial water systems, in *The Earth as*
761 *Transformed by Human Action*, edited by: Turner, B., Clark, W., Kates, R., Richards, J.
762 Mathews, J. and Meyer, W., Cambridge University Press, Cambridge, 235-52, 1990.
763

764 Lafleur, P. M., McCaughey, J. H., Joiner, D. W., Bartlett, P. A., and Jelinski, D. E.: Seasonal
765 trends in energy, water, and carbon dioxide fluxes at a northern boreal wetland, *J. Geophys.*
766 *Res.*, 102, 29,009–29,020, doi:10.1029/96JD03326, 1997.

767

768 Lafleur, P. M., Roulet, N. T., Bubier, J. L., Frolking, S., and Moore, T.: Interannual variability in
769 the peatland-atmosphere carbon dioxide exchange at an ombrotrophic bog, *Global*
770 *Biogeochem. Cy.*, 17(2), 1036, doi:10.1029/2002GB001983, 2003.

771

772 Lawrence, D. M., and A. G. Slater (2007), Incorporating organic soil into a global climate model,
773 *Clim. Dyn.*, doi:10.1007/s00382-007-0278-1.

774

775 Lettenmaier, D. P. and Famiglietti, J. S.: Hydrology: Water from on high, *Nature*, 444, 562-63,
776 doi:10.1038/444562a, 2006.

777

778 Limpens, J., F. Berendse, Blodau, C., Canadell, J. G., Freeman, C., Holden, J., Roulet, N., Rydin,
779 H., and Schaepman-Strub, G.: Peatlands and the carbon cycle: from local processes to global
780 implications-a synthesis, *Biogeosciences*, 5, 1475-1491, 2008.

781

782 Li, H., Huang, M., Wigmosta, M. S., Ke, Y., Coleman, A. M., Leung, L. R., Wang, A., and
783 Ricciuto, D. M.: Evaluating runoff simulations from the community Land Model 4.0 using
784 observations from flux towers and a mountainous watershed, *J. Geophys. Res.*, 116, D24120.
785 DOI: 10.1029/2011JD016276, 2011.

786

787 Moore, T. R., and Roulet, N. T.: A comparison of dynamic and static chambers for methane
788 emission measurements from subarctic fens, *Atmosphere-Ocean*, 29, 102-109, 1991.

789

790 Moore, K. E., Fitzjarrald, D. R., Wofsy, S. C., Daube, B. C., Munger, J. W., Bakwin, P. S., and
791 Crill, P.: A season of heat, water vapor, total hydrocarbon, and ozone fluxes at a subarctic fen,
792 *J. Geophys. Res.*, 99(D1), 1937–1952, doi:10.1029/93JD01442, 1994.

793

794 MacDonal, S. E., and Yin, F.: Factors influencing size inequality in peatland black spruce and
795 tamarack evidence from post drainage release growth, *J. Eco.*, 87, 404-412, 1999.

796

797 Moore, J. K., Lindsay, K., Doney, S. C., Long, M. C., and Misumi, K.: Marine Ecosystem
798 Dynamics and Biogeochemical Cycling in the Community Earth System Model
799 [CESM1(BGC)]: Comparison of the 1990s with the 2090s under the RCP4.5 and RCP8.5
800 Scenarios. *J. Climate*, 26, 9291–9312, 2013.

- 798 Morris, P.J., Belyea, L.R., Baird, A.J.: Ecohydrological feedbacks in peatland development: a
799 theoretical modelling study. *Journal of Ecology*, 99, 1190-1201, 2011.
- 800 Morris, P.J., Baird, A.J., Belyea, L.R.: The role of hydrological transience in peatland pattern
801 formation. *Earth Surface Dynamics*, 1, 29-43, 2013.
- 802
803 Nichols, D.S.: Temperature of upland and peatland soils in a north central Minnesota forest,
804 *Canadian Journal of Soil Science*, 78(3), 493-509, 1998.
- 805 Nungesser, M. K.: Modelling microtopography in boreal peatlands: hummocks and hollows,
806 *Ecological Modelling*, 165, 175-207, 2003.
- 807 Nichols, D.S., and Verry, E. S.: Stream flow and ground water recharge from small forested
808 watersheds in north central Minnesota, *Journal of Hydrology*, 245, 89-103, 2001.
- 809 Niu, G., Yang, Z., Dickinson, R.E., and Gulden, L.E.: A simple TOPMODEL-based runoff
810 parameterization (SIMTOP) for use in global climate model, *J. Geophys. Res.* 110D, D21106,
811 doi:10.1029/2005JD00611, 2005.
- 812 Oechel, W.C., and Van Cleve, K.: The role of bryophytes in nutrient cycling in the taiga, in
813 *Ecological Studies*, Vol. 57: Forest Ecosystems in the Alaskan Taiga, edited by: Van Cleve,
814 K., Chapin III, F. S., Flanagan, P.W., Viereck, L.A., and Dyrness, C.T., Springer-Verlag, New
815 York, 121-137, 1986.
- 816 Oki, T., and Kanae, S.: Global hydrological cycles and world water resources, *Science*, 313,
817 1068-1072, 2006.
- 818 Oleson, K. W., Niu, G., Yang, Z., Lawrence, D. W., Thornton, P. E., Lawrence, P. J., Sto ¨ckli, R.,
819 Dickinson, R. E., Bonan, G. B., Levis, S., Dai, A., and Qian, T.: Improvements to the
820 Community Land Model and their impact on the hydrological cycle. *J. Geophys. Res.*, 113,
821 G01021, doi:10.1029/2007JG000563, 2008.
- 822 Oleson, K. W., Lawrence, D. W., Bonan, G. B., Drewniak, B., Huang, M., Koven, C. D., Levis,
823 S., Li, F., Riley, W. J., Subin, Z. M., Swenson, S. C., Thornton, P. E., Bozbiyik, A., Fisher, R.,
824 Heald, C. L., Kluzek, E., Lamarque, J., Lawrence, P. J., Leung, L. R., Lipscomb, W., Muszala,
825 S., Ricciuto, D. M., Sacks, W., Sun, Y., Tang, J., and Yang, Z.: Technical description of
826 version 4.5 of the Community Land Model (CLM), NCAR/TN-503+STR, NCAR Technical
827 Note, 2013.
- 828 Painter, S.L., and Karra, S.: Constitutive model for unfrozen water content in subfreezing
829 unsaturated soils. *Vadose Zone Journal*, 13, 1-8, DOI: 10.2136/vzj2013.04.0071, 2014.
830
- 831 Painter, S.L., Moulton, J.D., and Wilson, C.J.: Modeling challenges for predicting hydrologic
832 response to degrading permafrost. *Hydrogeology Journal*, 21, 221-224. DOI: 10.1007/s10040-
833 012-0917-4. 2013.
834
- 835 Parsekian, A. D., Slater, L., Ntarlagiannis, D., Nolan, J., S. Sebestyen, D., Kolka, R. K., and
836 Hanson, P. J.: Uncertainty in peat volume and soil carbon estimated using ground-penetrating
837 radar and probing. *Soil Sci. Soc. Am. J.*, 76, 1911-1918. DOI: 10.2136/sssaj2012.0040, 2012.

838
839 Richardson, M. C., C. Mitchell, P. J., Branfireun, B. A., and Kolka, R. K.: Analysis of airborne
840 LiDAR surveys to quantify the characteristic morphologies of northern forested wetlands, *J.*
841 *Geophys. Res.*, 115, G03005. DOI: 10.1029/2009jg000972, 2010.
842
843 Riley, W. J., Subin, Z. M., Lawrence, D. M., Swenson, S. C., Torn, M. S., Meng, J., Mahowald,
844 N., and Hess, P.: Barriers to predicting global terrestrial methane fluxes: Analyses using a
845 methane biogeochemistry model integrated in CESM, *Biogeosciences*, 8, 1925-1953. DOI:
846 10.5194/bg-8-1925-2011, 2011.
847
848 Robroek, B. J. M., Schouten, M. G. C., Limpens, J., Berendse, F., and Poorter, H.: Interactive
849 effects of water table and precipitation on net CO₂ assimilation of three co-occurring
850 *Sphagnum* mosses differing in distribution above the water table, *Global Change Biol.*, 15(3),
851 680-691, 2009.

852 Sachs, W.J., Cook, B. I., Buening, N., Levis, S., and Helkowski, J. H.: Effects of global
853 irrigation on the near-surface climate, *Clim. Dyn.*, 33,159-175, 2009.

854 Sebestyen, S. D., Dorrance, C., Olson, D. M., Verry, E. S., Kolka, R. K., Elling, A. E., and
855 Kyllander, R. (2011a), Long-term monitoring sites and trends at the Marcell Experimental
856 Forest, in *Peatland biogeochemistry and watershed hydrology at the Marcell Experimental*
857 *Forest*, edited by: Kolka, R. K. Sebestyen, S. D., Verry, E. S., and Brooks, CRC Press, New
858 YorK, 15-71, 2011a.

859
860 Sebestyen, S.D., Verry, E. S., and Brooks, K. N. (2011b), Hydrological responses to forest cover
861 changes on uplands and peatlands, in *Peatland biogeochemistry and watershed hydrology at*
862 *the Marcell Experimental Forest*, edited by: Kolka, R. K. Sebestyen, S. D., Verry, E. S., and
863 Brooks,, CRC Press, New YorK, 433-458, 2011b.
864
865 Silvola, J.: Combined effects of varying water content and CO₂ concentration on photosynthesis
866 in *Sphagnum fuscum*, *Holarctic Ecology* 13, 224-228, 1990.

867 Silvola, J., Alm, J., Ahlholm, U., Nykanen, H., and Martikainen, P. J., CO₂ fluxes from peat in
868 boreal mires under varying temperature and moisture conditions, *Journal Ecology*, 84, 219-228,
869 1996.

870 Sonnentag, O., Chen, J. M., Roulet, N. T., Ju, W. and Govind, A.: Spatially explicit simulation of
871 peatland hydrology and carbon dioxide exchange: Influence of mesoscale topography, *J.*
872 *Geophys. Res.*, 113, G02005, doi:10.1029/2007JG000605, 2008.

873 St-Hilaire, F., Wu, J., Roulet, N. T., Frolking, S., Lafleur, P. M., Humphreys, E. R., and Arora,
874 V.: McGill wetland model: evaluation of a peatland carbon simulator developed for global
875 assessments, *Biogeosciences*, 7, 3517–3530, 2010.

876 Swanson, D.K., Grigal, D.F.: A simulation model of mire patterning, *Oikos*, 53, 309– 314, 1988.

877 Swenson, S.C., Lawrence, D. M., and lee, H.: Improved simulation of the terrestrial hydrological
878 cycle in permafrost regions by the community Land Model, *J. Adv. Model. Earth syst.*,
879 doi:10.1029/2012MS000165, 2012.

- 880 Thomas, P. R., and Yao, X.: Stochastic Ranking for Constrained Evolutionary Optimization.
881 IEEE Transactions on Evolutionary Computation, 4(3), 274-283, 2000.
- 882 Tarnocai, C.: The impact of climate change on Canadian Peatlands, Canadian Water Resources
883 Journal 34(4), 453-466, 2009.
- 884 Turetsky, M. R., Bond-Lamberty, B., Euskirchen, E., Talbot, J., Frohking, S., McGuire, A.D. and
885 Tuittila, E.S.: The resilience and functional role of moss in boreal and arctic ecosystems. New
886 Phytologist 196, 49-67. DOI: 10.1111/j.1469-8137.2012.04254.x. 2012
- 887 Verry, E.S.: Water table and streamflow changes after stripcutting and clearcutting an undrained
888 black spruce bog. In proceeding of the Sixth International Peat Congress, August 17-23, 1980,
889 Duluth, MN, Eveleth, MN: W.A. Fisher Company, 493-498, 1981.
- 890 Verry, E. S., Brooks, K. N., and barten, P. L.: Streamflow response from an ombrotrophic mire.
891 In Proceedings, International Symposium on the Hydrology of Wetlands in Temperate and
892 Cold Regions. Helsinki, finland: Internatinal Peat Society/The Academy of Finland, 52-59,
893 1988.
- 894
895 Verry, E. S., Boelter, D. H., Päivänen, J., Nichols, D.S., Malterer, T. J., Gafni, A.: Physical
896 properties of organic soils, in Peatland biogeochemistry and watershed hydrology at the
897 Marcell Experimental Forest, edited by: Kolka, R. K. Sebestyen, S. D., Verry, E. S., and
898 Brooks, CRC Press, New YorK, 135-176, 2011, 2011a.
899
- 900 Verry, E. S., Brooks, K. N., Nichols, D. S., Ferris, D. R., and Sebestyen, S. D.: Watershed
901 hydrology, in Peatland biogeochemistry and watershed hydrology at the Marcell Experimental
902 Forest, edited by Kolka RK, Sebestyen SD, Verry ES, Brooks KN, pp. 193-212, CRC Press,
903 New York, 2011b.
904
- 905 Verry, E. S., Bay, R.R, Boelter, D. H.: Establishing the Marcell Experimental Forest: Threads in
906 time, in Peatland biogeochemistry and watershed hydrology at the Marcell Experimental
907 Forest, edited by: Kolka, R. K. Sebestyen, S. D., Verry, E. S., and Brooks, CRC Press, New
908 YorK, 1-13, 2011c.
909
910
- 911 Verry, E. S., and Jansenns, J. (2011d), Geology, vegetation, and hydrology of the S2 bog at the
912 MEF: 12,000 years in northern Minnesota, in Peatland biogeochemistry and watershed
913 hydrology at the Marcell Experimental Forest, edited by R. K. Kolka S. D. Sebestyen, E. S.
914 Verry, K.N. Brooks, CRC Press, New YorK, 93-134, 2011d.
915
- 916 Verry, E. S.: Concrete frost in peatlands and mineral soils: northern Minnesota, In Proceedings
917 of the International Peat Symposium: Peat and peatlands: the resource and its utilization,
918 edited by D. N Grubich, T. J Malterer, International Peat Society, 121-141, 1991.
919
- 920 Verry, E. S.: Microtopography and water table fluctuation in a Sphagnum mire, In Proceedings
921 of the 7th International Peat Congress, Dublin, Ireland, The Irish National Peat Committee /
922 The International Peat Society, pp., 11-31, 1984.
923
- 924 Williams, T.G., and Flanagan, L. B.: Effect of changes in water content on photosynthesis,

- 925 transpiration and discrimination against $^{13}\text{CO}_2$ and $\text{C}_{18}\text{O}_{16}\text{O}$ in *Pleurozium* and *Sphagnum*,
926 *Oecologia*, 108, 38-46, 1996.
- 927 Wu, J. B., Kutzbach, L., Jager, D., Wille, C., and Wilmking, M.: Evapotranspiration dynamics in
928 a boreal peatland and its impact on the water and energy balance, *J. Geophys. Res.*, 115,
929 G04038, doi:10.1029/2009JG001075, 2010.
- 930 Waddington, J. M., Rotenberg, P. A., and Warren, F. J.: Peat CO_2 production in a natural and
931 cutover peatland: Implications for restoration, *Biogeochemistry*, 54, 115-130, 2001.
- 932 Wu, J., Roulet, N. T., Sagerfors, J., Nilsson, M. B.: Simulation of six years of carbon fluxes for a
933 sedge-dominated oligotrophic minerogenic peatland in Northern Sweden using the McGill
934 Wetland Model (MWM), *J. Geophys. Res. -Biogeosciences*, DOI: 10.1002/jgrg.20045, 2013.
- 935 Wania, R., Ross, I., and Prentice, I. C.: Implementation and evaluation of a new methane model
936 within a dynamic global vegetation model: LPJ-WHyMe v1.3.1, *Geosci. Model Dev.*, 3,564-
937 584, 2010.
- 938 Zhang, Y., Li, C., Trettin, C. C., Li, H., and Sun, G.: An integrated model of soil, hydrology, and
939 vegetation for carbon dynamics in wetland ecosystems, *Global Biogeochem. C.*, 16(4), 1061,
940 doi:10.1029/ 2001GB001838, 2002.
- 941

942 Table 1. Model parameter values which were modified from default values to reflect site-level measurements or optimized parameters. All measured
 943 parameters were used to define vegetation physiology, and were applied separately to each of the dominant species or plant functional types (PFTs)
 944 included in the simulations. Optimized parameters were generated at the level of the soil column. For the case of the vegetation physiological parameter
 945 (mp), this means that the same value was applied for all vegetation types. Optimized hydrologic parameters were applied to both hummock and hollow
 946 columns. ENFB = evergreen needleleaf forest – boreal; DNFB = deciduous needleleaf forest – boreal. Units: dim = dimensionless.
 947

parameter	description	units	Black Spruce (ENFB)	Larch (DNFB)	Shrub	Grass
<i>Measured parameter</i>						
pct_pft	Percentage of PFT in gridcell	%	35	15	20	30
leafcn	leaf carbon:nitrogen ratio	gC/gN	67	24	40	38
SLAtop	specific leaf area at canopy top	m ² /gC	0.0075	0.022	0.012	0.03
<i>Optimized parameter</i>			Column-level			
mp	Ball-Berry stomatal conductance slope	dim	6.4			
r _{h2osfc}	surface water runoff	kg m ⁻⁴ s ⁻¹	8.40E-08			
q _{drai,0}	maximum subsurface drainage rate	kg m ⁻² s ⁻¹	9.20E-06			

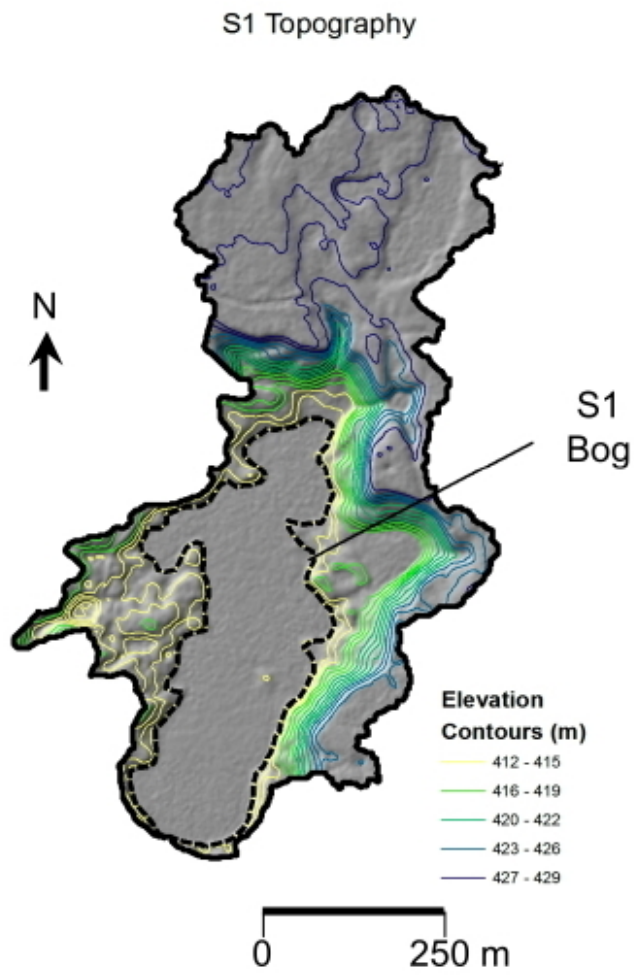
948 Table 2. The relative changes (as compared to the control) in annual ET and its components, and water table levels (WT) (averaged over the period of
 949 2011 to 2013) in the hummock (Hum) and hollow (Hol) under different warming scenarios with constant Q experiments (TR, E_c and E_s are canopy
 950 transpiration, canopy evaporation and soil evaporation, respectively).
 951

Warming scenario	Effects on ET (%)		Effects on TR(%)		Effects on E_c (%)		Effects on E_s (%)		Effects on WT(cm)	
	Hum	Hol	Hum	Hol	Hum	Hol	Hum	Hol	Hum	Hol
+3K	53.24	61.25	30.6	32.29	45.11	39.86	97.89	131.83	-7.32	-7.12
+6K	76.7	91.51	48.45	51.30	50.22	47.60	137.82	197.97	-12.84	-12.61
+9K	87.61	116.35	65.42	66.75	51.76	48.87	147.21	255.58	-14.60	-14.32

952

953

954
955

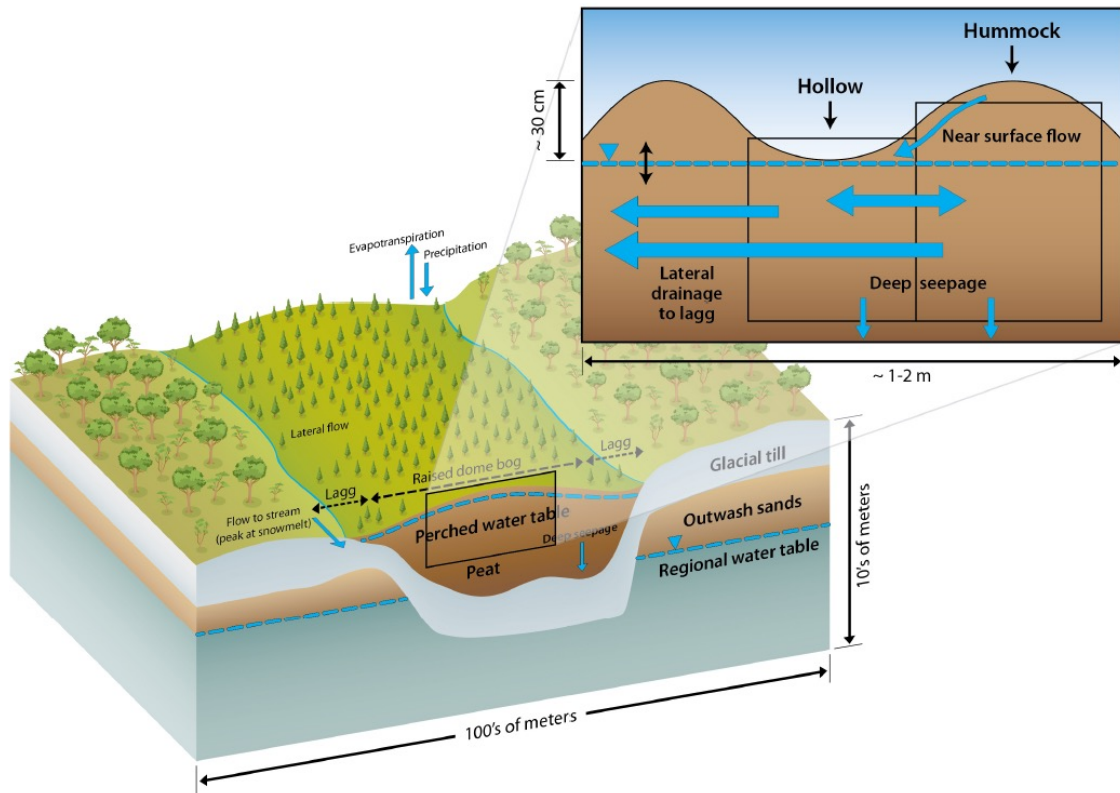


956
957
958
959
960

961 Figure 1. The topography of the S1 Watershed on the Marcell Experimental Fore
962 which contains the S1-Bog peatland.

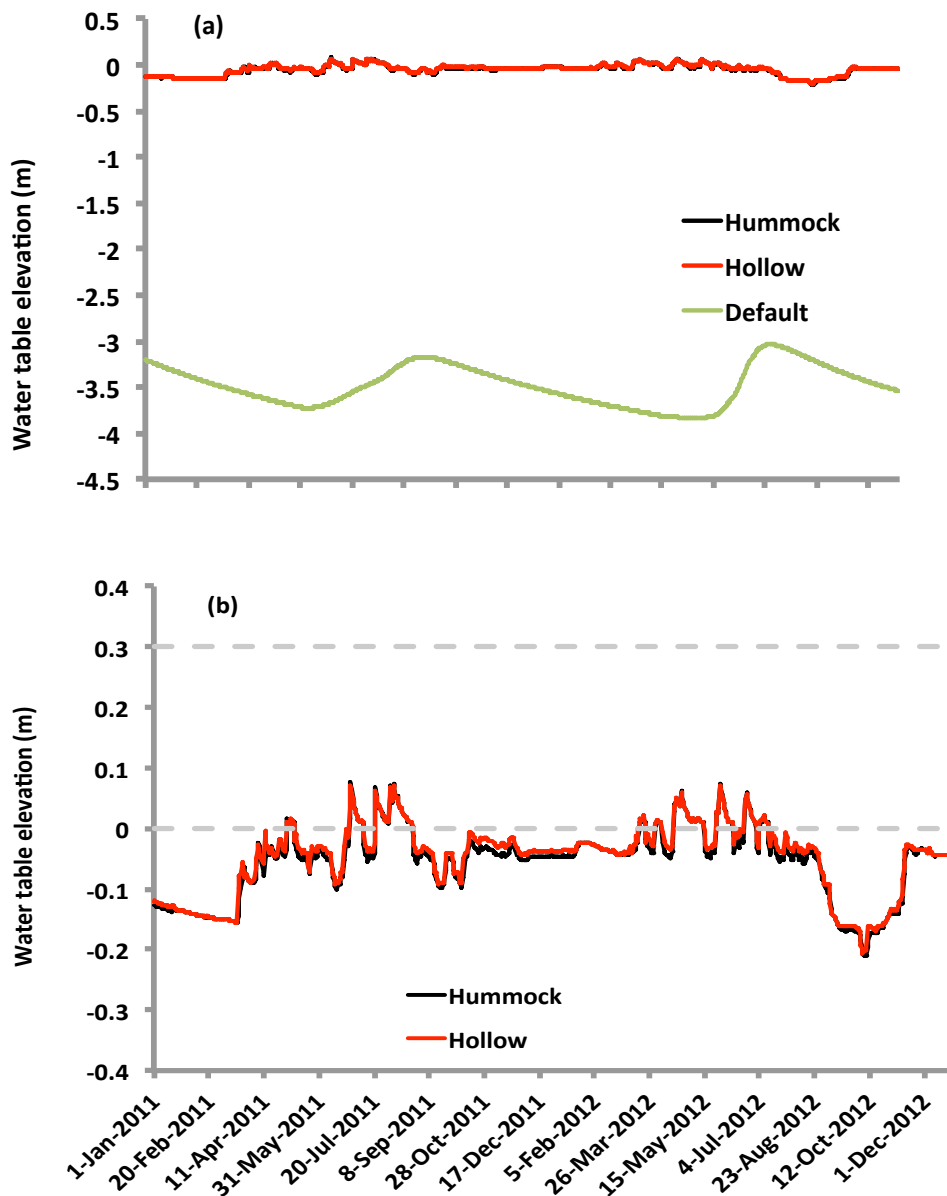
963
964

965



966

967 Figure2. Conceptualization of hummock/hollow microtopography within raised-
968 dome bog ecosystem. The broad view shows the bog water table perched above
969 regional water table, due to hydrologic isolation of the bog by underlying glacial till.
970 The inset shows an idealized cross-section view of microtopography, with model
971 representation of hummock and hollow columns (black outlines) and water fluxes
972 (solid blue arrows) between columns and from columns to the bog-scale drainage
973 feature (lagg). Bog water table is shown as dashed blue line.

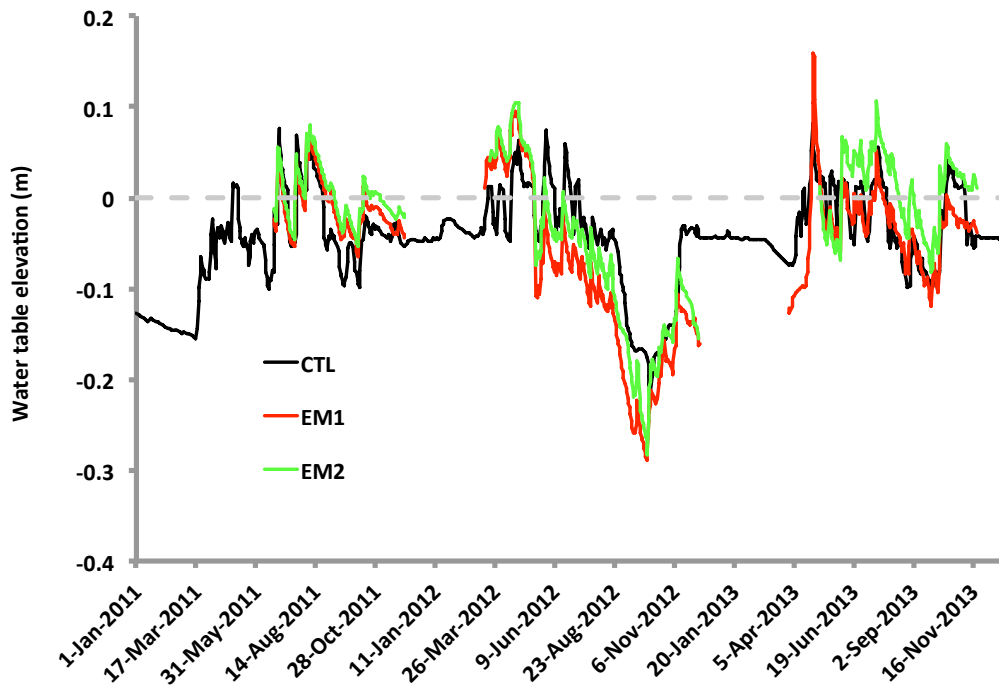


974

975

976 Figure 3. CLM_Default and CLM_SPRUCE simulated hummock and hollow water
 977 table levels (a), and CLM_SPRUCE predicted water table dynamics (b) for years 2011
 978 and 2012. Dashed lines show the height of the surface of the hummock (0.3m) and
 979 hollow (0m).

980



981

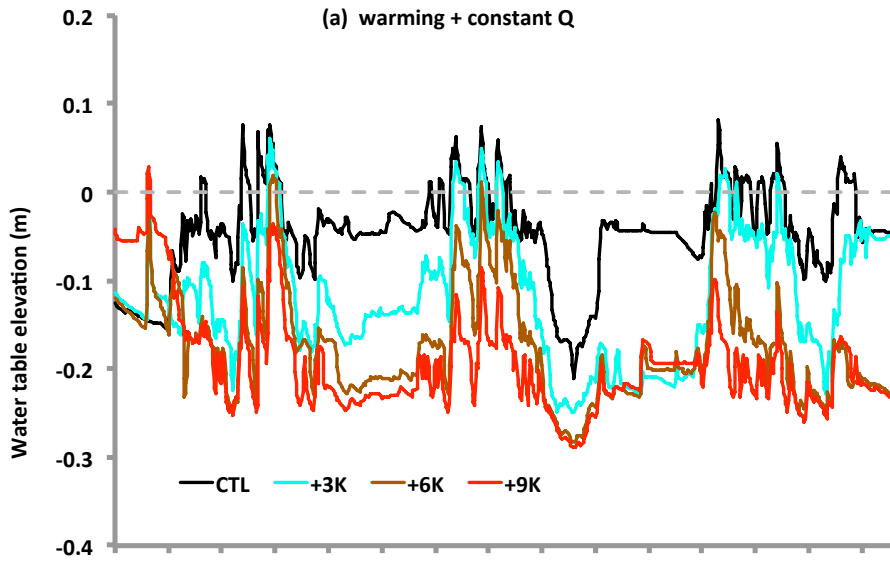
982 Figure 4. The comparison of CLM_SPRUCE simulated (CTL) and observed water
 983 table levels (EM1 and EM2) for hummocks for the years 2011 to 2013. Zero line
 984 indicates the surface of the hollow. For clarity, model results are shown only for the
 985 simulated hummock: simulated water table heights are nearly identical for
 986 hummock and hollow (Data are missing from EM1 and EM2 during winter when the
 987 bog surface is frozen and the water table sensors are not collecting data).

988

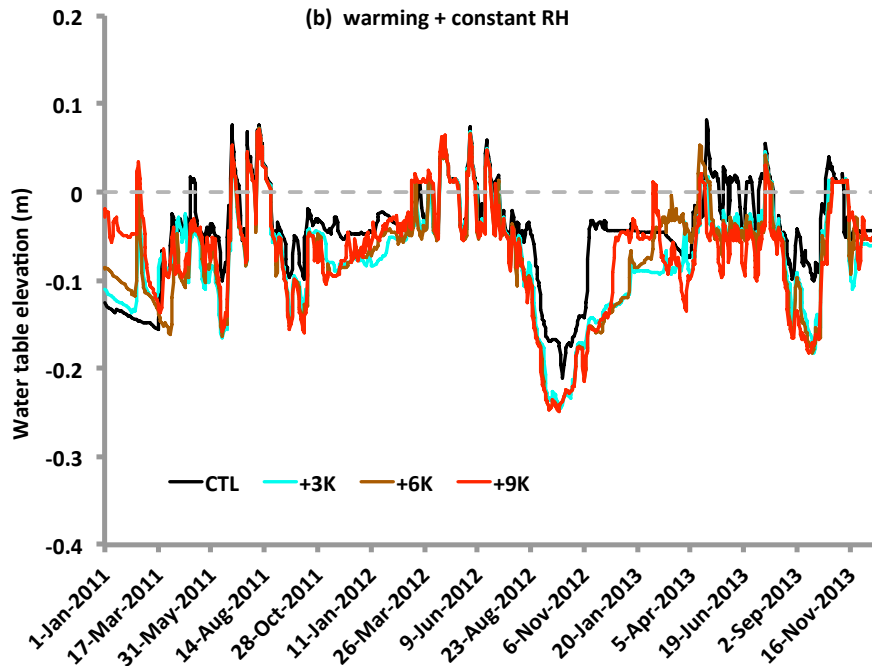
989

990

991

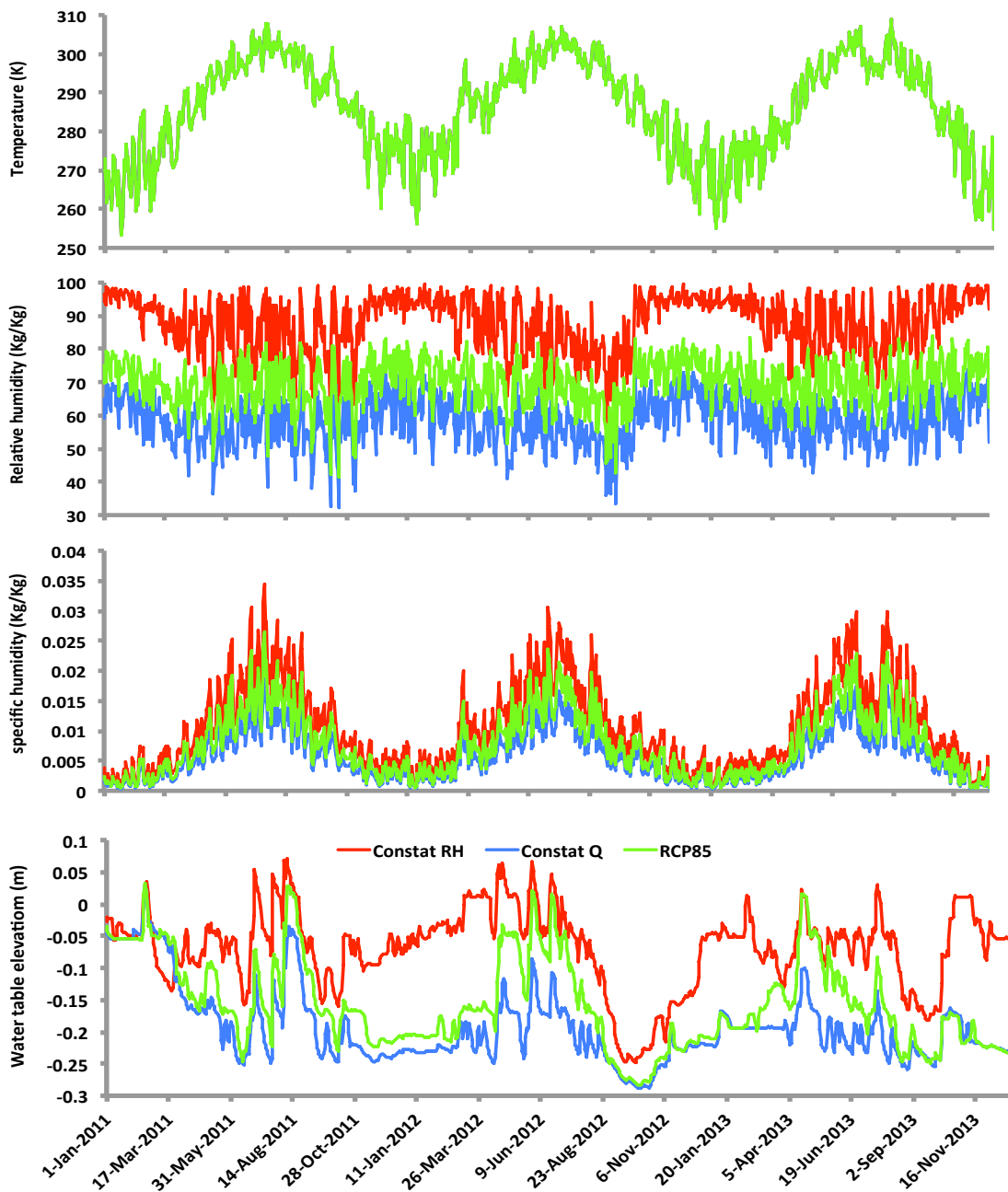


992
993



994
995

996 Figure 5. The simulated water table levels for years 2011-2013 for control
 997 (ambient) and warming scenarios (+3K, +6K, +9K above ambient) based on two
 998 humidity conditions: a) the same specific humidity for all 4 simulations (constant
 999 Q); and b) the same relative humidity for all 4 simulations (constant RH). Zero line
 1000 indicates the surface of the hollow



1001

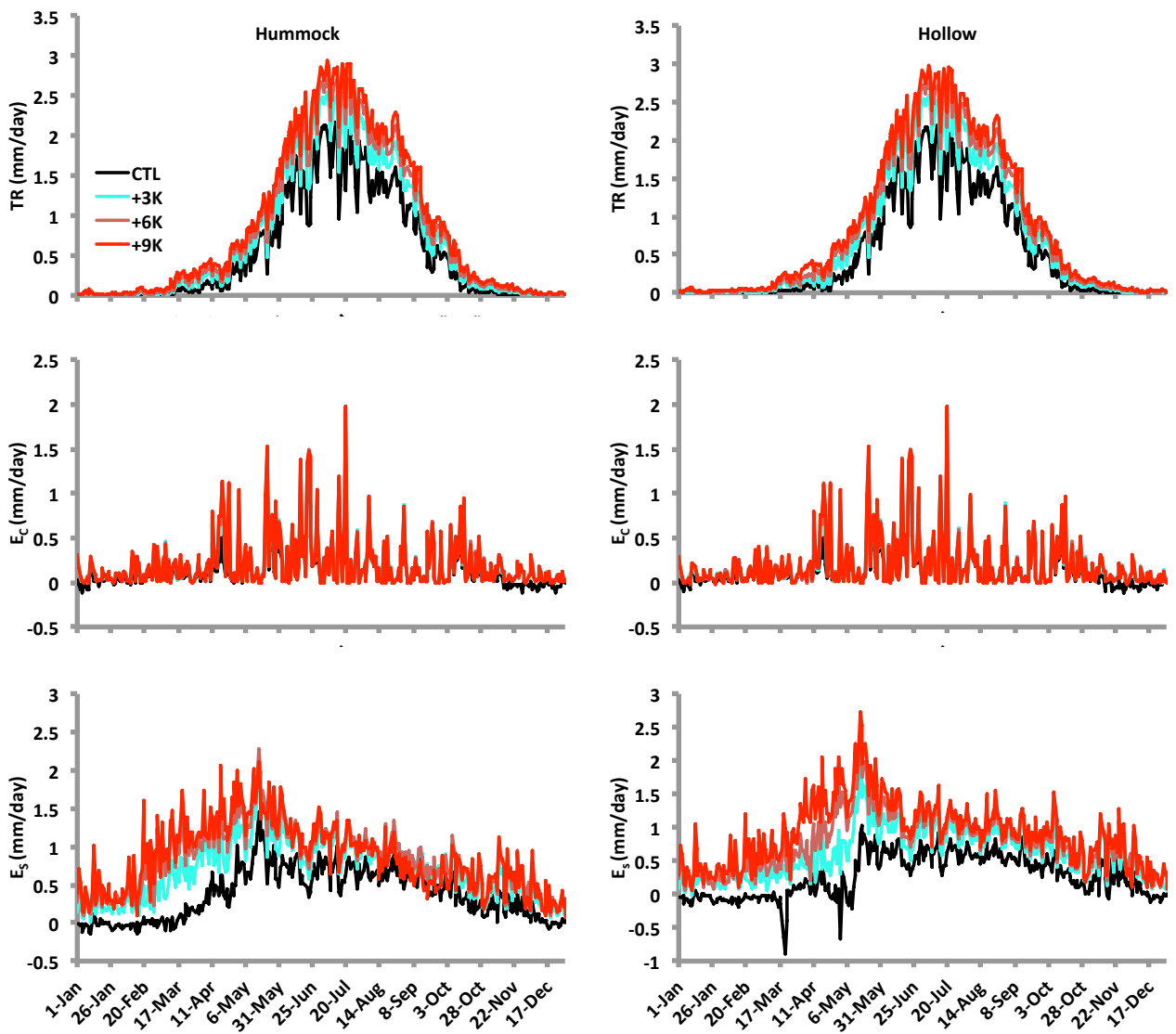
1002 Figure 6. The evolution of temperature, relative humidity (RH), specific humidity

1003 (Q), and water table levels for warming (+9K) scenarios with three different

1004 humidity conditions: red lines designate constant RH; blue lines designate constant

1005 Q; green lines are for increasing Q 30% and decreasing RH 14% (RCP8.5 scenario)

1006 for years 2011-2013.



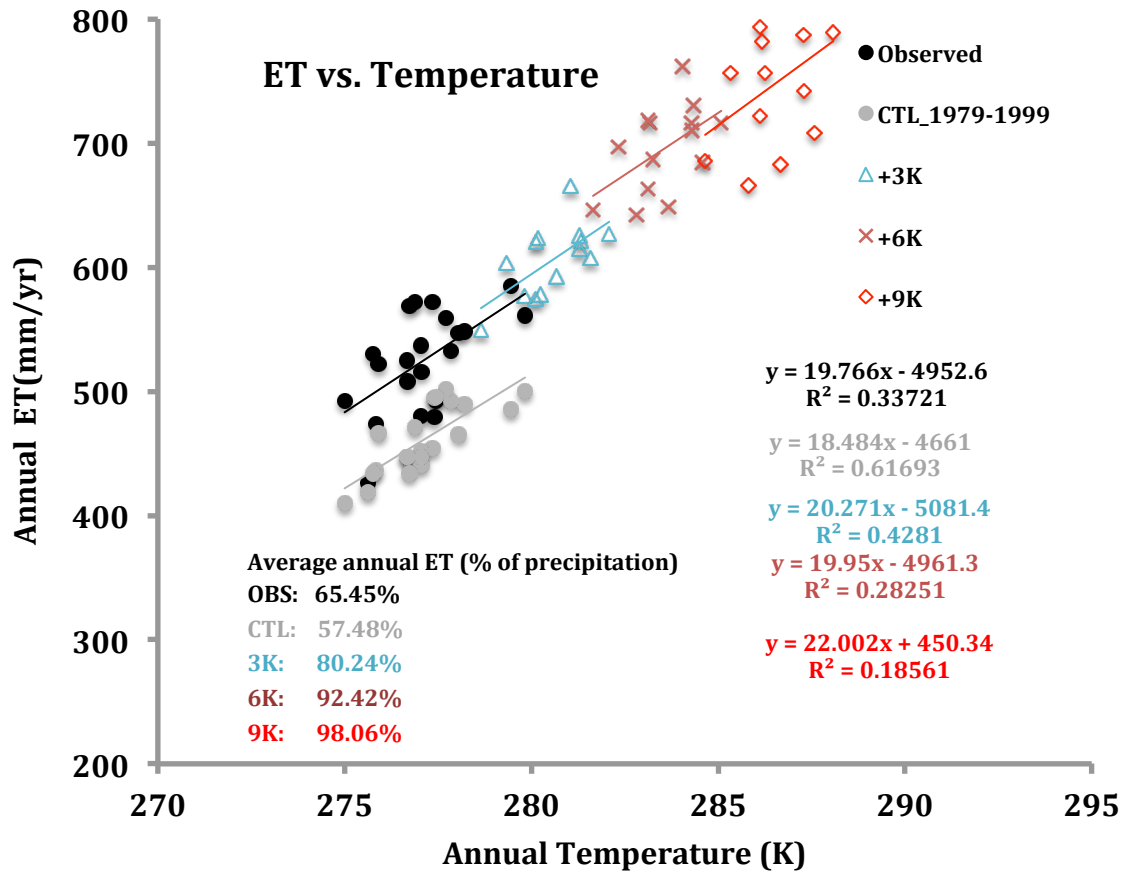
1007

1008 Figure 7. Seasonal dynamics (averaged over the 2011-2013 period) of three
 1009 components of evapotranspiration, for control and 3 warming scenarios under the
 1010 constant Q assumption for humidity. TR, E_c and E_s are canopy transpiration, canopy
 1011 evaporation and soil evaporation, respectively.

1012

1013

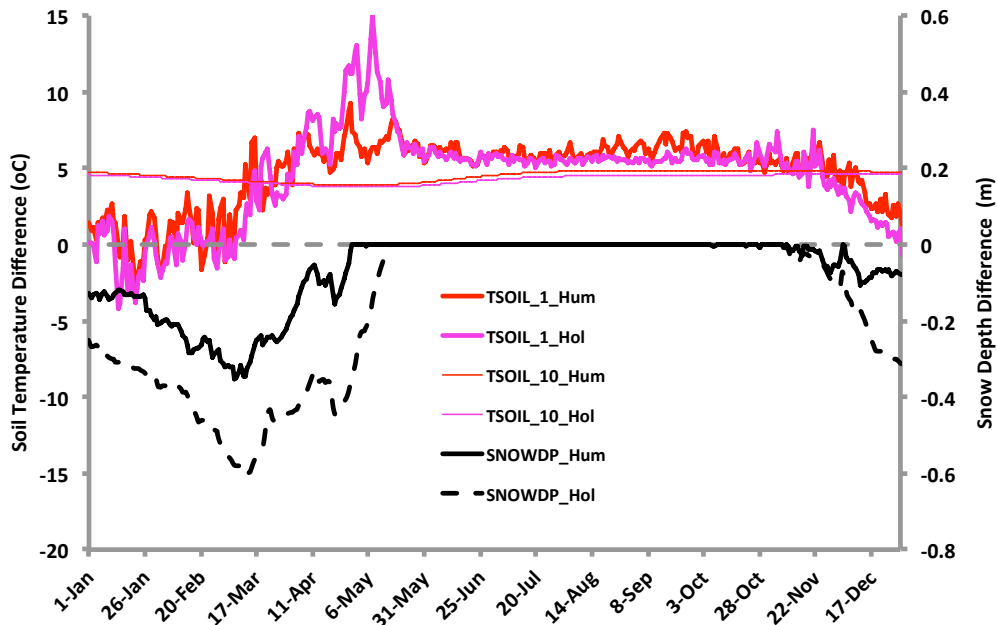
1014



1015
1016

1017 Figure 8. Annual total ET vs. annual mean air temperature. Observed values from
 1018 the period 1979-1999 at the S2-Bog. Model values for pre-treatment control
 1019 simulation using weather data for the period 1979-1999, and for three levels of
 1020 imposed warming (+3K, +6K, and +9K) based on weather data for the period 2000-
 1021 2013. Regression lines are shown for interannual variation within each of these five
 1022 categories.

1023



1024

1025

1026 Figure 9. Differences (9K -control) in soil temperature for first (TSOIL_1) and tenth
 1027 (TSOIL_10) soil layers as predicted by CLM_SPRUCE under constant Q assumptions
 1028 for humidity. Also shown (right x-axis) is the difference in snowdepth (SNOWDP)
 1029 over the hummock (Hum) and hollow (Hol) for the same pair of experiments. Model
 1030 results are shown as the average seasonal cycle over the 2011 to 2013 period.

1031

1032