

1 Seasonal characterization of CDOM for lakes in semi-arid regions of
2 Northeast China using excitation-emission matrices fluorescence and
3 parallel factor analysis (EEM-PARAFAC)

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12
13 **Abstract.** The seasonal characteristics of fluorescent components in CDOM for lakes
14 in the semi-arid region of Northeast China were examined by excitation-emission
15 matrix (EEM) spectra and parallel factor analysis (PARAFAC). Two humic-like (C1
16 and C2) and the protein-like (C3 and C4) components were identified using
17 PARAFAC. The average fluorescence intensity of the four components differed under
18 seasonal variation from June and August 2013 to February and April 2014.
19 Components 1 and 2 exhibited strong linear correlation ($R^2 = 0.633$). Significantly
20 positive linear relationships between CDOM absorption coefficients $a(254)$ ($R^2 = 0.72$,
21 0.46 , $p < 0.01$), $a(280)$ ($R^2 = 0.77$, 0.47 , $p < 0.01$), $a(350)$ ($R^2 = 0.76$, 0.78 , $p < 0.01$)
22 and F_{max} for two humic-like components (C1 and C2) were exhibited, respectively. A

23 significant relationship ($R^2 = 0.931$) was found between salinity and DOC. However,
24 almost no obvious correlation was found between salinity and EEM-PARAFAC
25 extracted components except for C3 ($R^2 = 0.469$). Results from this investigation
26 demonstrate that the EEM-PARAFAC technique can be used to evaluate the seasonal
27 dynamics of CDOM fluorescent components for inland waters in the semi-arid
28 regions of Northeast China, and to quantify CDOM components for other waters with
29 similar environmental conditions.

30 **Keywords:** CDOM, fluorescent components, EEMs, PARAFAC, DOC, Salinity

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32

33 **1 Introduction**

34 Dissolved organic matter (DOM), a heterogeneous mixture of humic acids,
35 proteins and carbohydrates, plays important roles in aquatic ecosystems (Zhang et al.,
36 2010). Chromophoric dissolved organic matter (CDOM), the colored fraction of
37 DOM, absorbs light energy in the ultraviolet (UV) and visible region of the spectrum
38 and inhibits the propagation of UV radiation. CDOM in waters also affects the
39 transport and bio-availability of materials such as trace metals and other pollutants
40 (Song et al., 2013), so it can be used as a proxy of water quality. In natural water
41 bodies, CDOM originates from the degradation of plant materials and other organisms
42 and terrestrially imported substances, which varies in time and space and is controlled
43 by its structure and composition (Stedmon et al., 2003). CDOM is compositionally
44 complex, making it difficult to isolate hydrophobic from hydrophilic acids using XAD
45 ion-exchange resins (Aiken et al., 1992; Spencer et al., 2010). Nonetheless, some

46 optically active components of CDOM can emit fluorescence after absorbing light at
47 certain wavelengths (Zhang et al., 2010) so that the fluorescence spectroscopic
48 techniques can be used to provide detailed information about the source and
49 concentration of CDOM. The traditional fluorescence techniques including
50 fluorescence emission spectrometry and synchronous fluorescence scanning applied
51 to examine CDOM components have the drawback that the output was restricted to a
52 linear scan (Hudson et al., 2007).

53 Recently, excitation-emission matrix fluorescence spectroscopy (EEM) has been
54 applied to identify CDOM components because of its ability of producing
55 synchronous scan spectra in the form of contours (Stedmon et al., 2003; Zhang et al.,
56 2010). The EEM spectroscopy is considered the most effective technique for studying
57 the composition of fluorophores given its high selectivity and sensitivity to CDOM in
58 water columns (Zhang et al., 2010). In recent years, EEM spectroscopy has been
59 widely used to investigate the dynamics of marine, freshwaters and ice-water
60 ecosystems as well as snow melting water (Barker et al., 2006, 2009, 2010, 2013;
61 Coble, 2007; Fellman et al., 2010; Guo et al., 2010; Hudson et al., 2007; Stedmon et
62 al., 2007). Moreover, the EEM spectroscopy can also be used to distinguish
63 allochthonous and autochthonous CDOM source in aquatic environment (Coble et al.,
64 1998; Mayer et al., 1999; Yamashita et al., 2008, 2010; Zhang et al., 2013). Based on
65 the peak positions in EEMs, two main fluorescent components, i.e., humic-like and
66 protein-like substances, have been identified and investigated (DelCastillo et al., 1999;
67 Jaffe' et al., 2004). However, overlapped fluorophores of CDOM EEMs could make

68 this traditional ‘peak-picking’ method unreliable to evaluate CDOM dynamics in
69 aquatic ecosystems (Coble, 1996; Stedmon et al., 2003). Recently, the combined
70 EEMs-PARAFAC (parallel factor analysis) technique has been shown to effectively
71 decompose EEM of CDOM into independent fluorescent components and assess the
72 source of CDOM and relationships with other water quality parameters. A number of
73 investigators have used EEMs-PARAFAC to characterize DOM in freshwater and
74 marine aquatic environments (Broisover et al., 2009; Cory et al., 2005; Guo et al.,
75 2010; Stedmon et al., 2003; Stedmon and Markager, 2005; Yamashita, 2008; Zhang et
76 al., 2010, 2011, 2013). Stedmon et al. (2003) introduced PARAFAC and identified
77 five distinct DOM components for a Danish estuary and its catchment. In coastal
78 environments, Yamashita et al. (2008) reported on seven components using the
79 combined EEMs-PARAFAC technique and assess the dynamic of individual
80 fluorophores and relationship with salinity in Ise Bay. Zhang et al. (2011) also found
81 three different components by PARAFAC modeling and analyzed the correlations
82 between the fluorescent components and absorption coefficients of CDOM for Lake
83 Tianmu and its catchment.

84 The Songnen Plain is a fluvial plain with semi-arid climate, in which many fresh
85 and brackish waters are distributed according to its geomorphological characteristics
86 (Song et al., 2013). Dissolved organic carbon (DOC) characteristics of these fresh
87 and brackish waters across the Songnen Plain have been studied by Song et al. (2013);
88 the results indicated that a huge amount of DOC were stored in these waters. In
89 particular, brackish waters would exhibit high average DOC concentration and

90 significantly contributed the carbon budget to inland waters (Duarte et al., 2008; Song
91 et al., 2013; Tranvik et al., 2009). However, little study has been made on the detailed
92 information of DOC sources for these waters in the Songnen Plain. Therefore, it
93 motivated us to investigate the components in CDOM for both fresh and brackish
94 waters in the semi-arid region. In the present study, the absorption and fluorescence of
95 CDOM were determined for the water samples collected from seven lakes in the
96 western part of Jilin province, which varied at different season. The specific
97 objectives of this study are to: 1) characterize CDOM components contained in these
98 lakes using EEMs and their origins through the EEM-PARAFAC method; 2) assess
99 the dynamic of individual fluorescent component of CDOM under seasonal variation;
100 and most importantly 3) link CDOM fluorescence intensities, absorption coefficients,
101 DOC concentrations and salinity to each other, in order to establish proxies for
102 CDOM bioavailability and photoreactivity in waters.

103

104 **2 Materials and Methods**

105 **2.1 Lakes and water sampling**

106 The water bodies investigated in this study were located in the western part of Jilin
107 Province, which belongs to the semi-arid part of the Songnen Plain (Song et al., 2013).
108 Two groups of lakes were investigated, i.e., the Chagan lake group and the Yuelianghu
109 lake group. The Chagan lake group is made up of Lake Chagan (CGL), Xinmiaopao
110 (XMP), Xindianpao (XDP) and Kulipao (KLP). The Yuelianghu lake group mainly
111 includes Lake Yueliang (YLL), Talahong (TLH) and Xinhuangpao (XHP) (Fig. 1).

112 The two groups are about 60 km away from each other, of which each includes both
113 fresh and brackish waters. The primary economic value for these lakes is fishery,
114 agricultural irrigation and recreation. The average annual precipitation is about 391
115 mm, but the average evaporation is up to 1790 mm, resulting in water scarcity. Due to
116 the area dominated by saline-alkali soil, the rainfall flush and agricultural catchment
117 land use can result in an increase of lake salinities. These seven lakes are endowed
118 with similar geological, hydrological and climatic settings, thus we presume that
119 similar process may control the CDOM components. In order to characterize the
120 CDOM fluorescent components under seasonal variation using EEMs-PARAFAC, 67
121 water samples were collected from the surface of the seven lakes in 1-liter
122 acid-cleaned plastic bottles during four field campaigns in June and August 2013 as
123 well as in February and April 2014, respectively. These samples were collected during
124 the ice covering period using an ice drilling auger. The under-ice surface water was
125 coming up when the ice layer was drilled a hole by the auger. The ice shavings were
126 collected in plastic bags and the under-ice surface water was collected in plastic
127 bottles. The collected samples were held on ice and immediately transported to the
128 laboratory in the Changchun City of Jilin province within 3-5 hours. In the laboratory,
129 these samples were filtered within 24h and then kept at 4 °C until analysis within two
130 days. Latitude and longitude of each sample location were recorded *in situ* using a
131 Trimble Global Positioning System (GPS).

132

133 **Figure 1. Locations of the water sampling sites for 7 lakes in the western part of Jilin**
134 **province, Northeast China.**

135

136 **2.2 Analytical procedures**

137 To characterize the basic parameters of water quality, salinity was measured through a
138 DDS-307 electrical conductivity (EC) meter in laboratory. Salinity was expressed in
139 the basis of the UNESCO practical salinity unit (PSU 1978). The pH was measured
140 using a PHS-3C pH meter at room temperature (20 ± 2 °C) in laboratory. Water
141 turbidity was determined using the Shimadzu UV-2600PC UV-Vis dual beam
142 spectrophotometer with matching 3 cm quartz cells at room temperature (20 ± 2 °C)
143 with Milli-Q water as the reference (UV talk letter vol. 10,
144 <https://shimadzu.com.au/uv-talk-letter-volume-10>). To determine DOC concentrations,
145 water samples were filtered through 0.45 µm filters and then measured using a
146 Shimadzu TOC-5000 Analyzer and a 1.2 % Pt on silica catalyst at 680 °C. Potassium
147 hydrogen phthalate was used as standard. The reproducibility of the analytical
148 procedure was within 2-3 % for the current study (APHA, 1998; Song et al., 2011).

149

150 **2.3 Absorption measurement**

151 In laboratory, all the samples were filtered at low pressure, first through a
152 pre-combusted Whatman GF/F filter (0.7 µm), and then through a pre-rinsed 25 mm
153 Millipore membrane cellulose filter (0.22 µm) into glass bottles. Absorption spectra of
154 the samples were measured between 200 and 800 nm at 1 nm increments using the
155 Shimadzu UV-2600PC UV-Vis dual beam spectrophotometer with a 1 cm quartz
156 cuvette and Milli-Q water as reference. The absorption coefficient a_{CDOM} was

157 calculated from the measured optical density (OD) of the sample using Eq. (1):

$$158 \quad a_{CDOM}(\lambda) = 2.303 \left[OD_{(\lambda)} - OD_{(null)} \right] / \gamma \quad (1)$$

159 where γ is the cuvette path length (0.01 m) and the factor 2.303 converts from base 10
160 to base natural logarithm transformation. Some fine particles possibly remained in the
161 filtered solution (Babin et al., 2003; Bricaud et al., 1995), therefore it was necessary
162 to correct for scattering by fine particles and in this case, $OD_{(null)}$ is the average optical
163 density over 740-750 nm where the absorbance of CDOM can be assumed to be zero.

164 A CDOM absorption spectrum ($a_{CDOM}(\lambda)$) can be expressed as an exponential
165 function (Babin et al., 2003; Bricaud et al., 1995):

$$166 \quad a_{CDOM}(\lambda_i) = a_{CDOM}(\lambda_r) \exp[-S(\lambda_i - \lambda_r)] \quad (2)$$

167 where $a_{CDOM}(\lambda_i)$ is the CDOM absorption at a given wavelength λ_i , $a_{CDOM}(\lambda_r)$ is the
168 absorption estimate at the reference wavelength λ_r (440 nm), and S is the spectral
169 slope of the CDOM absorption. According to Helms et al. (2008), S is calculated by
170 fitting a linear model to the data over a wavelength range of 275 to 295 nm ($S1$) or
171 350 to 400 nm ($S2$). To eliminate the inter-laboratory variability, the slope ratio $S_R =$
172 $S1/S2$ is defined to indicate the molecular weight and photo-bleaching of CDOM
173 (Helms et al., 2008; Zhang et al., 2010).

174

175 **2.4 Three-dimensional fluorescence measurement**

176 The EEMs analysis of CDOM were conducted using a Hitachi F-7000 fluorescence
177 spectrometer (Hitachi High-Technologies, Tokyo, Japan) with a 700-voltage xenon
178 lamp. The scanning ranges were 200–450 nm for excitation, and 250–500 nm for

179 emission. Readings were collected in the ratio mode at 5 nm intervals for excitation,
180 and at 1 nm intervals for emission, using a scanning speed of 2400 nm min⁻¹. The
181 band-passes were 5 nm for both excitation and emission. A Milli-Q water blank of the
182 EEMs was subtracted to eliminate the water Raman scatter peaks (McKnight et al.,
183 2001; Stedmon et al., 2003; Zhang et al., 2010, 2011).

184 The inner-filter effect, which results from reabsorption and excitation of the
185 fluorescence itself, can reduce the fluorescence intensity by 5% (Larsson et al., 2007;
186 McKnight et al., 2001). In order to eliminate the inner-filter effect, the EEMs were
187 corrected for absorbance by multiplying each value in the EEMs with a correction
188 factor based on the assumption that the average path length of absorption of the
189 excitation and emission light is one-half length of the cuvette (McKnight et al., 2001;
190 Zhang et al., 2010). The correction function is expressed as follows:

$$191 \quad F_{corr} = F_{obs} \times 10^{(A_{ex}+A_{em})/2} \quad (3)$$

192

193 where F_{Corr} and F_{obs} are the corrected and uncorrected fluorescence intensities and A_{ex}
194 and A_{em} are the absorbance values at the respective excitation and emission
195 wavelengths.

196 The measured fluorescence intensity is dependent on the concentration of the
197 dissolved fluorophores in water bodies. Finally, the fluorescence intensities of all
198 sample's EEMs were normalized to the area under the Milli-Q water Raman peak
199 (λ_{ex} =350 nm, λ_{em} =371-428 nm) measured daily (Lawaetz and Stedmon, 2009). The
200 contour figures of the EEMs were plotted using the Matlab 10.0 software package

201 (Math Works, Natick Massachusetts, America).

202

203 **2.5 The PARAFAC modeling**

204 PARAFAC, a three-way method, is applied to decompose the CDOM fluorescence

205 into separate fluorescent signals (Andersen and Bro, 2003; Stedmon and Bro, 2008).

206 According to Stedmon and Bro (2008), a similar PARAFAC analysis is carried out in

207 the present study using the DOMFluor toolbox in MATLAB with the “N-way toolbox

208 for MATLAB” (Andersson et al., 2000). Before PARAFAC modeling, the excitation

209 wavelengths from 200 to 220 nm and the emission wavelengths from 250 to 300 nm

210 were deleted because of their poor quality. In order to remove the effect of Rayleigh

211 scatter on PARAFAC modeling, the missing values (NaN-Not a number) were

212 inserted in the regions ($Ex-20 \leq Em \leq Ex+20$ and $2Ex-20 \leq Em \leq 2Ex+20$; unit: nm)

213 which are significantly influenced by the first and second order scattering from the

214 measured spectroscopic data (Hua et al., 2007; Stedmon and Bro, 2008).

215 To determine the appropriate number of PARAFAC components, the split-half

216 validation procedure was executed to verify whether the model was valid by

217 comparing the emission and excitation loadings from each half (Stedmon and Bro,

218 2008). Split-half analysis is the most effective method for implementing the

219 PARAFAC models, in which the EEMs are randomly divided into four groups of

220 equal size, and then analyzed for two half splits (1-2 and 3-4 half) respectively. If the

221 correct number of components is chosen, the excitation and emission loadings from

222 the two groups should show the same shape and size (Bro, 1997, 1999). The

223 fluorescence intensity of every component was represented by F_{\max} (Raman unit: nm^{-1})
224 (Stedmon and Markager, 2005).

225

226 **2.6 Statistical analysis**

227 Statistical analysis was conducted using the SPSS 16.0 software package (Statistical
228 Program for Social Sciences). Regression and correlation analysis was used to
229 describe the relationship between CDOM absorption coefficient, DOC concentration,
230 salinity and F_{\max} . A model II-ANOVA was performed to determine seasonal
231 variability is higher than between-lake variability. The difference is considered to be
232 statistically significant when p -values are less or equal to 0.05.

233

234 **3 Results and discussion**

235 **3.1 Water quality conditions**

236 The water quality parameters, i.e., pH, salinity, turbidity for the 67 water samples
237 collected from June 2013 to April 2014 in the western part of Jilin province are
238 displayed in Table 1. When the set of samples from various field trips was pooled
239 together, the waters had high pH values and high salt contents. The highest salinity
240 was present when the lakes were frozen in February 2014, whereas relatively constant
241 values (around 0.40 PSU) were exhibited in other three seasons. Also the water bodies
242 were highly turbid. The highest turbidity was present in June 2013, and then reduced
243 in August 2013, and the lowest value was recorded in February 2014. Compared with
244 February 2014, the turbidity had almost no change in April 2014 (Table 1).

245

246 **Table 1. Mean value of water quality parameters from June 2013 to April 2014.**

247

248 **3.2 EEMs characterization of CDOM**

249 Based on the EEMs ‘peak picking’ technique, the key fluorescence peaks can be
250 observed in 67 water samples: two humic-like and two protein-like substances (Coble,
251 1996; Stedmon et al., 2003). The humic-like components are the mixture of aromatic
252 and aliphatic compounds-humic-like acids from terrestrial substances, and aquatic
253 humic-like substances of phytoplankton origin. With respect to the protein-like
254 components, they, i.e., tyrosine-like and tryptophan-like substance, mainly consist of
255 dissolved amino acids. As an example, Fig. S1 displays the EEMs of samples from
256 lake Xindianpao at different seasons. The peaks comprise two humic-like
257 fluorescence peaks: one in the ultraviolet range (Ex/Em = 220-240/410-430 nm) and
258 the other in the visible range (Ex/Em = 300-340/410-450 nm) and the protein-like
259 fluorescence peaks: tyrosine-like (Ex/Em = 210-230, 270-280/310-330 nm) and
260 tryptophan-like (Ex/Em = 220-230, 280-300/350-370 nm).

261

262 In our study, four separate fluorescent components (Fig. 2a-d) and the
263 excitation and emission loadings (e-h) of the four components identified by
264 EEM-PARAFAC are summarized in Fig. 2 and Table 2. The first fluorescent
265 component (C1) was a biological degradation humic-like component comparable to
266 humic-like peaks (M and N) in marine and in phytoplankton degradation experiments

267 for inland waters (Coble, 1996; Zhang et al., 2009). Component 2 was consistent with
268 the humic-like peaks (A and C) defined by Coble (1996). Component 3 resembles the
269 tryptophan-like (T) component as found by Baker et al. (2004) and Hudson et al.
270 (2007). For component 4, it is likely related to tyrosine-like component (B) (Hudson
271 et al., 2007). Components 3 and 4 represent autochthonous semi-labile CDOM
272 associated with bacteria activity and phytoplankton degradation (Borisover et al.,
273 2009; Stedmon et al., 2003). Particularly, there was a shoulder at the excitation
274 wavelength 310-330 nm in component 3 and 330-340 nm in component 4, which may
275 be due to the residual Raman peaks in some water samples (Fig. 2c-d). In this study,
276 not all of the four components were present in all of the samples.

277

278 **Figure 2. The PARAFAC modeling output shows the contour plots of the four PARAFAC**
279 **fluorescent components (a-d) and excitation (black) and emission (red) loadings (e-h) of each**
280 **component. Fluorescence is in Raman units: nm⁻¹.**

281

282 **Table 2. Positions of the fluorescence maximum peaks of the four components identified by**
283 **PARAFAC modeling in the present study compared with those previously identified.**
284 **Secondary excitation maxim is given in brackets.**

285

286 **3.3 Temporal distribution of PARAFAC components**

287 These fresh and brackish water in Jilin province in northeast China are endowed with
288 similar geological, hydrological and climatic settings, thus it is presumed that similar
289 process may control the CDOM components. When a model II-ANOVA using season
290 and lake as random effect factors was performed, it shows that the seasonal variability

291 ($F > F_{crit}$, $p < 0.05$) is higher than between-lake variability. Therefore, the water
292 samples from different lakes for every season were pooled together in order to study
293 the seasonal variation of the fluorescent components. As shown in Fig. 3a, the average
294 fluorescence intensity of the four components had seasonal variation. When all the
295 water samples at different seasons were pooled together, the average value of total
296 fluorescence intensity was $2.05 \pm 0.93 \text{ nm}^{-1}$, corresponding to the intensities of $0.71 \pm$
297 0.32 (C1), 0.33 ± 0.11 (C2), 0.50 ± 0.24 (C3), and 0.51 ± 0.26 (C4) nm^{-1} for different
298 components. These results can demonstrate that the fluorescence intensity was
299 dominated by C1, implying most part of the CDOM for the seven inland lakes was
300 originated from the degradation of phytoplankton and microorganisms. The
301 protein-like components (C3 and C4), related to bioavailability and microbial activity
302 of CDOM, had almost the same magnitude. At all four seasons, the fluorescent
303 component C2, which was terrestrially imported to water bodies, contributed less to
304 total fluorescence than the other three. The total fluorescence intensity differed under
305 seasonal variation, varying from $2.54 \pm 0.68 \text{ nm}^{-1}$ in June to $1.93 \pm 0.70 \text{ nm}^{-1}$ in
306 August 2013, and then increased to $2.34 \pm 0.92 \text{ nm}^{-1}$ in February and reduced to the
307 lowest $1.57 \pm 0.55 \text{ nm}^{-1}$ in April 2014 (Fig. 3c). The intensities of four fluorescent
308 components (i.e., 0.75 ± 0.17 (C1), 0.32 ± 0.06 (C2), 0.69 ± 0.24 (C3), and $0.77 \pm$
309 0.20 (C4) nm^{-1}) (Fig. 3d) from the samples collected in June 2013 exhibited similar
310 trends to that for the pooled data set. These values were higher than the seasonal
311 average except C2 ($0.32 \pm 0.06 \text{ nm}^{-1}$). This can be explained by enhanced activities
312 from plant degradation and microbial activities, but less terrestrial substances were

313 imported to the water bodies in June and therefore the fluorescence intensity of C2
314 was lower than the seasonal average. Compared to the fluorescence intensity in June,
315 the three fluorescence intensities (0.65 ± 0.14 (C1), 0.33 ± 0.16 (C3), 0.52 ± 0.36 (C4)
316 nm^{-1}) from the samples collected in August 2013 reduced, but an increased value was
317 recorded for C2 ($0.42 \pm 0.05 \text{ nm}^{-1}$) (Fig. 3d). Especially, the fluorescence intensities
318 of two protein-like components showed an obvious difference. This can be attributed
319 to substantially increased precipitation up to 180 mm in July from June to August
320 2013 (Fig. 3b) so that flood occurred when rainfall continued to increase in August.
321 Gradually, DOM contained in terrestrial CDOM was flushed by rainfall to the lakes so
322 that the C2 ($0.42 \pm 0.05 \text{ nm}^{-1}$) fluorescence intensity became higher. In accordance
323 with Cheng et al. 2010, the rainwater CDOM for this study was largely characterized
324 by protein-like components (Cheng et al., 2010). The fluorescence intensity of the
325 rainwater CDOM was very weak, and also the rainwater CDOM contained much
326 lower humic-like concentration (Fig. S2b). The intensities of the other three
327 components decreased because of dilution resulting from heavy rain and relatively
328 weak microbial decomposition of plants.

329 The highest C1 ($1.02 \pm 0.38 \text{ nm}^{-1}$) presented in February 2014 and the C2 (0.39
330 $\pm 0.12 \text{ nm}^{-1}$) intensity remained almost the same as that in August 2013. However, the
331 protein-like components indicated that the C3 ($0.57 \pm 0.25 \text{ nm}^{-1}$) intensity was higher
332 than the C4 ($0.35 \pm 0.17 \text{ nm}^{-1}$) intensity, which was opposite to the results from other
333 months (Fig. 3d). In cold winter, the surface waters formed a thick layer of ice
334 covering the lake waters. Because the ice cover reduced light penetration and

335 restricted gas exchange between the underlying water and atmosphere, vigorous
336 biological activities in the lakes would be reduced at low temperature and low light
337 level (Thomas K., 1983; Uusikiv et al., 2010; Wharton, et al., 1993). Although the
338 biological activities was very weak, there still be a bit of production of C1 and C3 in
339 lake water. Also, dissolved materials were left in the underlying surface waters and
340 little terrestrial matters were imported to the lakes once covered by ice (Stedmon et al.,
341 2007). Therefore, the C1 and C3 in the water of the lakes beneath the ice layers would
342 be produced and cumulated simultaneously, whereas, the C2 remained the same.
343 Obviously, the fluorescence intensity of component 1 reached the highest value for
344 the winter samples. As shown in Fig. S2a, another striking feature for the winter
345 samples was that the fluorescence of CDOM in the ice was dominated by the
346 tyrosine-like C4 component, which is consistent with the findings of Barker et al.
347 (2009, 2013) and Stedmon et al. (2007). It showed that the C4 component was left in
348 the ice-cover when the lakes were frozen. Therefore, it is not surprising that the
349 intensity of component C4 for water beneath ice layers was reduced and the
350 concentrated C3 showed a much higher fluorescence intensity. In April 2014, the
351 intensities of four fluorescent components (0.47 ± 0.17 (C1), 0.25 ± 0.08 (C2), $0.40 \pm$
352 0.16 (C3), and 0.45 ± 0.13 (C4) nm^{-1}) (Fig. 3d) exhibited similar seasonal trends
353 though these values were much lower than the average. Our interpretation is that the
354 ice CDOM was characterized by tyrosine-like component (C4) (Fig. S2a), and the
355 fluorescence intensity of C4 contributed by the ice-melt water was very weak.
356 However, the underlying lake CDOM included both humic-like (C1 and C2) and

357 protein-like (C3 or C4) components. When the ice in the lakes melt into water with
358 warming weather and biological degradation and human activity was weak, the lake
359 CDOM was diluted by the ice-melted water and the fluorescence intensity would
360 reach to the lowest value in early spring.

361

362 **Figure 3. a) Seasonal average of F_{max} for EEM-PARAFAC components (C1, C2, C3 and C4)**
363 **for lakes in the western part of Jilin province; b) Monthly variation of rainfall for the lakes**
364 **in western part of Jilin province from April 2013 to February 2014; c) Seasonal variation of**
365 **the total fluorescence intensity at different seasons; d) Seasonal variation of the four**
366 **EEM-PARAFAC components at different seasons. The error bar represents standard**
367 **deviations.**

368

369 **3.4 CDOM versus EEM-PARAFAC extracted components**

370 The concentration of DOC, CDOM absorption coefficients and the slope ratio S_R are
371 shown in Table 3. The DOC concentrations ranged from 10.03 to 56.60 mg L⁻¹ with
372 an average value of 31.66 mg L⁻¹ during the study period from April to August 2013,
373 demonstrating a seasonal dynamics that can be attributed to hydrological, climatic and
374 landscape variations (Song et al., 2013). It is because the prolonged sunshine duration
375 can result in an evapo-condensed DOC concentration in the Songnen Plain. The
376 highest averaged DOC concentration (55.04 ± 20.00 mg L⁻¹) was present in February
377 2014; whereas, relatively constant values of approximate 30 mg L⁻¹ were observed in
378 the other three seasons. This can be attributed to the accumulated DOC when the
379 lakes freeze in winter, which leaves DOC in the liquid phase, resulting in a higher
380 DOC concentration in the underlying water (unpublished material). Generally, the

381 absorption coefficient $a(350)$ is used as a proxy for characterizing CDOM
382 concentration (Guo et al., 2010; Zhang et al., 2011). $a(280)$ is related to DOC
383 biodegradation (McDowell et al., 2006). $a(254)$ can be used to character the optical
384 properties of DOC aromaticity (Jaffe' et al., 2004; Weishaar et al., 2003). The highest
385 averaged CDOM absorption coefficients $a(350)$, $a(280)$, $a(254)$ were also present in
386 February 2014, corresponding to the highest DOC concentration. The S_R values of the
387 two wavelength ranges (275-295 nm over 350-400 nm) were used to represent DOM
388 molecular weight (Helms et al., 2008). The lowest mean of S_R was present in August
389 2013 suggesting the relatively weak microbial decomposition of plants and lots of
390 terrestrially imported substances through rainwash resulted in the higher average
391 molecular weight of DOC.

392

393 **Table 3. Mean values of DOC concentration and CDOM absorption coefficients groups at**
394 **different seasons. S_R : the slope ratio of $S_{275-295nm} : S_{350-400nm}$.**

395

396 When the whole data set ($N = 67$) was pooled together, there were significantly
397 positive linear relationships between $a(254)$, $a(280)$, $a(350)$ and F_{max} for two
398 humic-like components (C1 and C2), respectively, but mostly such correlations were
399 not observed for the protein-like components (Fig. 4a and b, Table 3). These results
400 were in accordance with previous investigations (Zhang et al., 2010, 2011).
401 Components 1 and 2 were strongly linearly correlated with each other ($R^2 = 0.633$)
402 (Fig. 4c), indicating that the concentrations of the two humic-like components were
403 controlled by common sources (Baker and Spencer, 2004). There was a weak

404 relationship ($R^2 = 0.051$) between the protein-like components (C3 and C4) because of
405 a complex origin of CDOM such as rainfall in summer, ice in winter and organic
406 pollutants derived from domestic, agricultural and industrial sewerage, which
407 represent the complex origins of CDOM. However, there was almost no correlation
408 between the humic-like and protein-like components. The linkage of a fluorescence
409 signal to DOC was very complicated because of the seasonal impacts, i.e., increased
410 rainfall, algal blooms and ice-cover, which affect the DOC concentration. Due to both
411 steady and labile CDOM fluorescent components in DOC, the fluorescent signal
412 would change with the ratio of fluorescent and non-fluorescent CDOM components
413 (Henderson et al., 2009). A weak relationship ($R^2 = 0.42$) (Fig. 4d) was found between
414 DOC and component 3 from the decay of plants through microbial activity or the
415 pollution from human and animal wastes.

416 Different from the findings by Yamashita et al. (2008) for ocean water, this study
417 did not find obvious correlation between salinity and EEM-PARAFAC extracted
418 components with the exception of C3 ($R^2 = 0.469$) (Table 4 and Fig. 4f). The most
419 important finding for the water samples collected at different seasons from the
420 Songnen Plain is a significant relationship ($R^2 = 0.931$) between salinity and DOC
421 (Fig. 4e). This is because DOC is evapo-condensed from spring to autumn and
422 freeze-accumulated in winter in the semi-arid region. A prolonged sunshine duration
423 can result in an evapo-condensed DOC concentration from April to August 2013. On
424 the other hand, the DOC is accumulated when the lakes freeze in winter leaving DOC
425 in the liquid phase.

426

427 **Table 4. Correlation coefficients (R) and significance levels (p) of the linear relationships**
428 **between CDOM absorption, DOC, salinity and fluorescent components.**

429

430 **Figure 4. Relationships between CDOM absorption coefficient a(350) with a) $F_{max}(C1)$, b)**
431 **with $F_{max}(C2)$, c) peak $F_{max}(C1)$ versus $F_{max}(C2)$, d) peak $F_{max}(C3)$ versus DOC, e) Salinity**
432 **versus DOC, f) Salinity versus $F_{max}(C3)$.**

433

434 **4 Conclusions**

435 A model II-ANOVA using season and lake as random effect factors shows that the
436 seasonal variability ($F > F_{crit}$, $p < 0.05$) is higher than between-lake variability. In this
437 study, the application of EEM-PARAFAC to characterize four fluorescent
438 components under seasonal variation in CDOM was presented with 67 water samples
439 collected from June 2013 to April 2014 in the semi-arid region of the Songnen Plain.

440 Two humic-like and the protein-like components were identified using PARAFAC
441 model. The average fluorescence intensity of the four components differed under
442 seasonal variation from June 2013 to April 2014. The highest C1 1.02 nm^{-1} was
443 presented in February 2014 due to the condensed CDOM caused by ice formation in
444 winter. Especially in summer when quantities of rainfall take place and in winter
445 when water is frozen, the fluorescence intensity is dominated by tyrosine-like
446 component in rain and ice-melt water. Component 1 and 2 exhibited strong linear
447 correlation ($R^2 = 0.633$). There were significantly positive linear relationships
448 between F_{max} and CDOM absorption coefficient a(254) ($R^2 = 0.72$, 0.46 , $p < 0.01$),
449 a(280) ($R^2 = 0.77$, 0.47 , $p < 0.01$), a(350) ($R^2 = 0.76$, 0.78 , $p < 0.01$) for two

450 humic-like components (C1 and C2), respectively. A weak relationship ($R^2 = 0.42$)
451 was found between DOC and component 3 from the decay of plants through
452 microbial activity or the pollution from human and animal wastes. However, almost
453 no obvious correlation was found between salinity and EEM-PARAFAC extracted
454 components except C3 ($R^2 = 0.469$), though the correlation was not as strong as with
455 DOC concentration. Most importantly, a significant relationship ($R^2 = 0.931$) was
456 found between salinity and DOC. In order to understand the biogeochemical effects
457 on the aquatic ecosystem, further study should be required to identify CDOM source
458 and assess physical/chemical, bioavailable and photoreactive transformation in
459 various lakes with larger saline gradients in the semi-arid region, Northeast China.

460

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465 their persistent assistance with both field sampling and laboratory analysis.

466

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646 Table 1. Mean value of water quality parameters from June 2013 to April 2014. Turb

647 denotes water turbidity; N denotes sampling numbers.

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Sampling season	pH	Salinity (PSU)	Turb (NTU)	N
Jun.2013	8.54	0.40	166.20±108.73	15
Aug.2013	8.63	0.37	63.13±31.21	13
Feb.2014	8.35	0.70	21.33±15.87	17
Apr.2014	8.67	0.43	22.24±16.42	22
All	8.55	0.48	62.18±79.07	67

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671 Table 2. Positions of the fluorescence maximum peaks of the four components
672 identified by PARAFAC modeling in the present study compared with those
673 previously identified. Secondary excitation maxima is given in brackets.

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Component No	Ex _{max} (nm)	Em _{max} (nm)	Description and source	Components (Coble) and (Zhang)	Components (Stedmon and Markager)
C1	230 (300)	425	Marine humic-like (phytoplankton degradation)	M	6
C2	255 (350)	460	Terrstrial humic-like	A and C	1 and 4
C3	225 (290)	360	Autochthonous tryptophan-like	T	
C4	220 (275)	320	Autochthonous tyrosine-like	B	8

675 Fluorescence peaks were named as Components (Coble) and (Zhang) by Coble et al. (1996, 1998) and Zhang et al.
676 (2010, 2011), while as Componets (Stedmon and Markager) by Stedmon and Markager (2005).

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694 Table 3. Mean values of DOC concentration and CDOM absorption coefficients

695 groups at different seasons. S_R : the slope ratio of $S_{275-295nm} : S_{350-400nm}$.

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Sampling season	a(254) m ⁻¹	a(280) m ⁻¹	a(350) m ⁻¹	S_R	DOC mg L ⁻¹	N
Jun.2013	38.39±9.23	25.98±6.38	5.73±1.68	1.29±0.16	31.84±14.67	15
Aug.2013	29.71±4.73	19.36±2.91	5.82±0.81	0.96±0.22	32.83±14.78	13
Feb.2014	52.88±18.13	34.62±11.54	6.36±2.17	1.18±0.11	55.04±20.00	17
Apr.2014	34.43±11.38	22.45±7.36	4.17±1.49	1.32±0.13	30.86±10.91	22
All	39.08±14.73	25.73±9.58	5.40±1.84	1.21±0.20	37.60±18.05	67

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716 Table 4. Correlation coefficients (R) and significance levels (*p*) of the linear
717 relationships between CDOM absorption, DOC, salinity and fluorescent components.

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	a(254)	a(280)	a(350)	DOC	Salinity	C1	C2	C3	C4
DOC	0.711**	0.646**	0.294*	1.000**					
Salinity	0.650**	0.579**	0.159	0.965**	1.000**				
C1	0.850**	0.875**	0.873**	0.496**	0.383**	1.000**			
C2	0.677**	0.686**	0.885**	0.414**	0.270*	0.796**	1.000**		
C3	0.452**	0.417**	0.134	0.648**	0.685**	0.267*	0.103	1.000**	
C4	-0.040	-0.016	0.078	-0.101	0.135	0.084	0.069	0.225	1.000**

719 ***p*< 0.01 level ; **p*<0.05 level.

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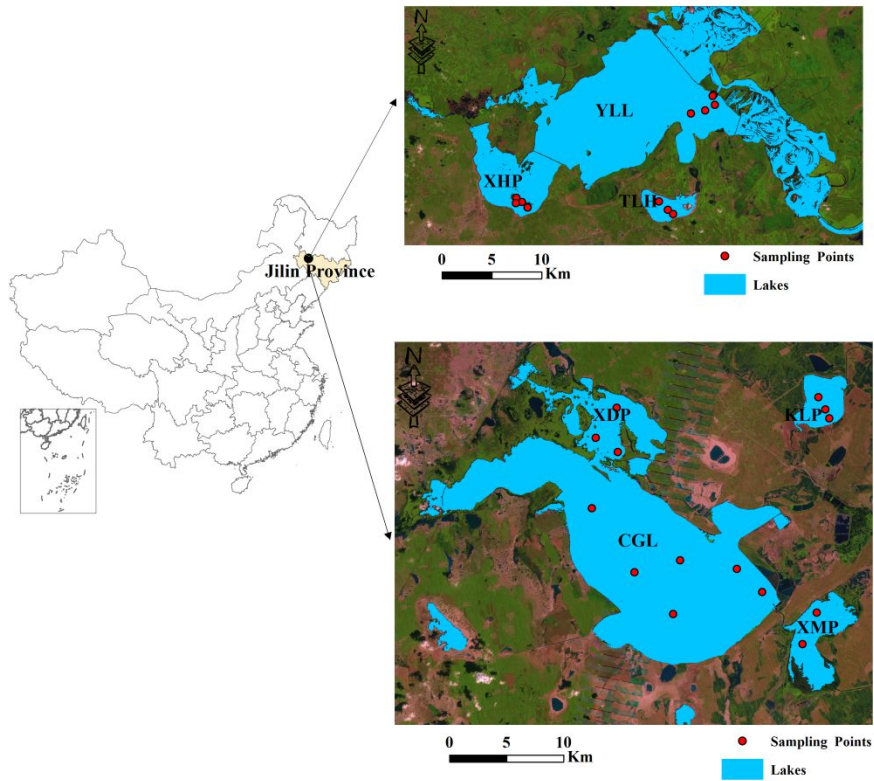
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732 Figure 1. Locations of the water sampling sites for 7 lakes in the western part of Jilin

733 province, Northeast China.

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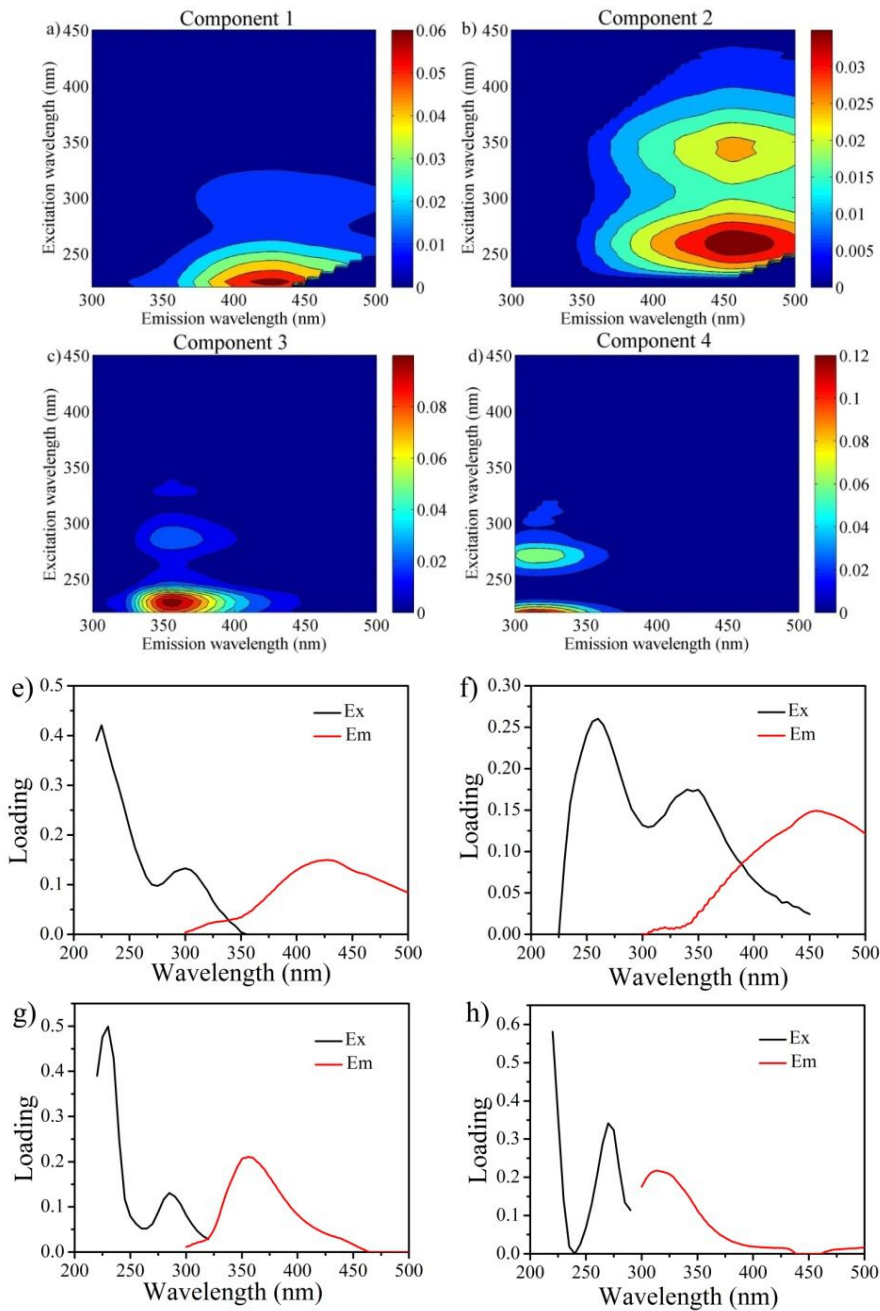
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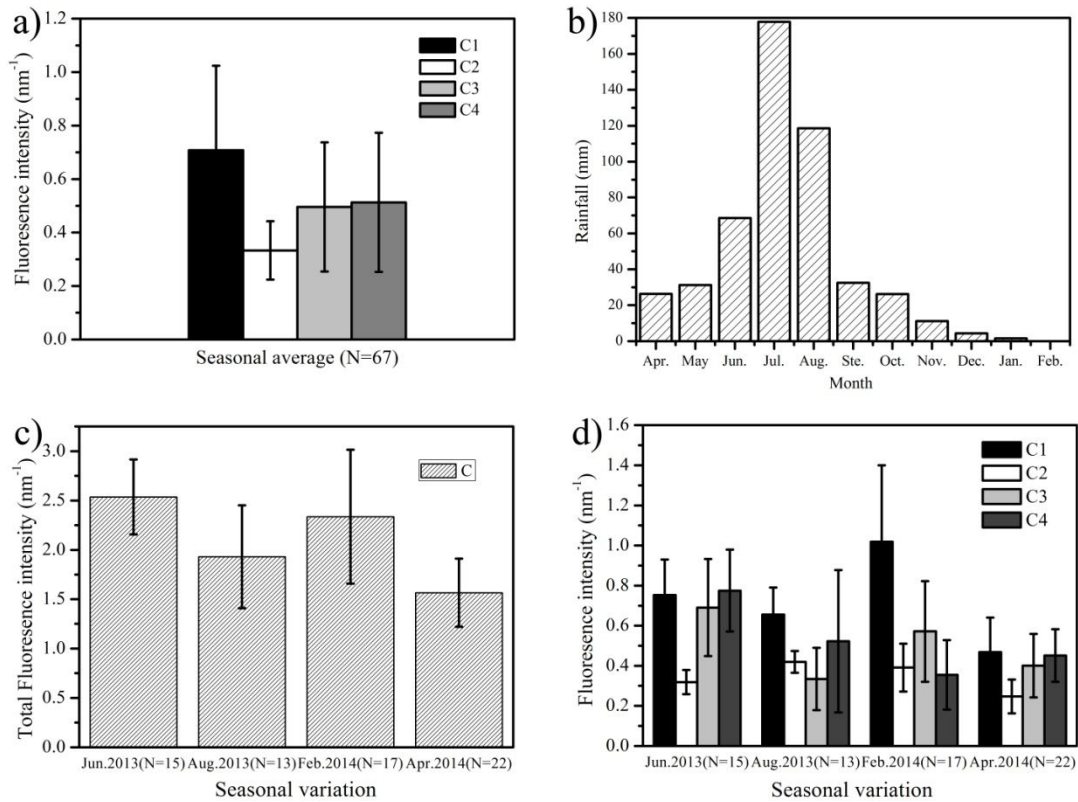
760 Figure 2. The PARAFAC model output shows the contour plots of the four

761 PARAFAC fluorescent components (a-d) and excitation (black) and emission (red)

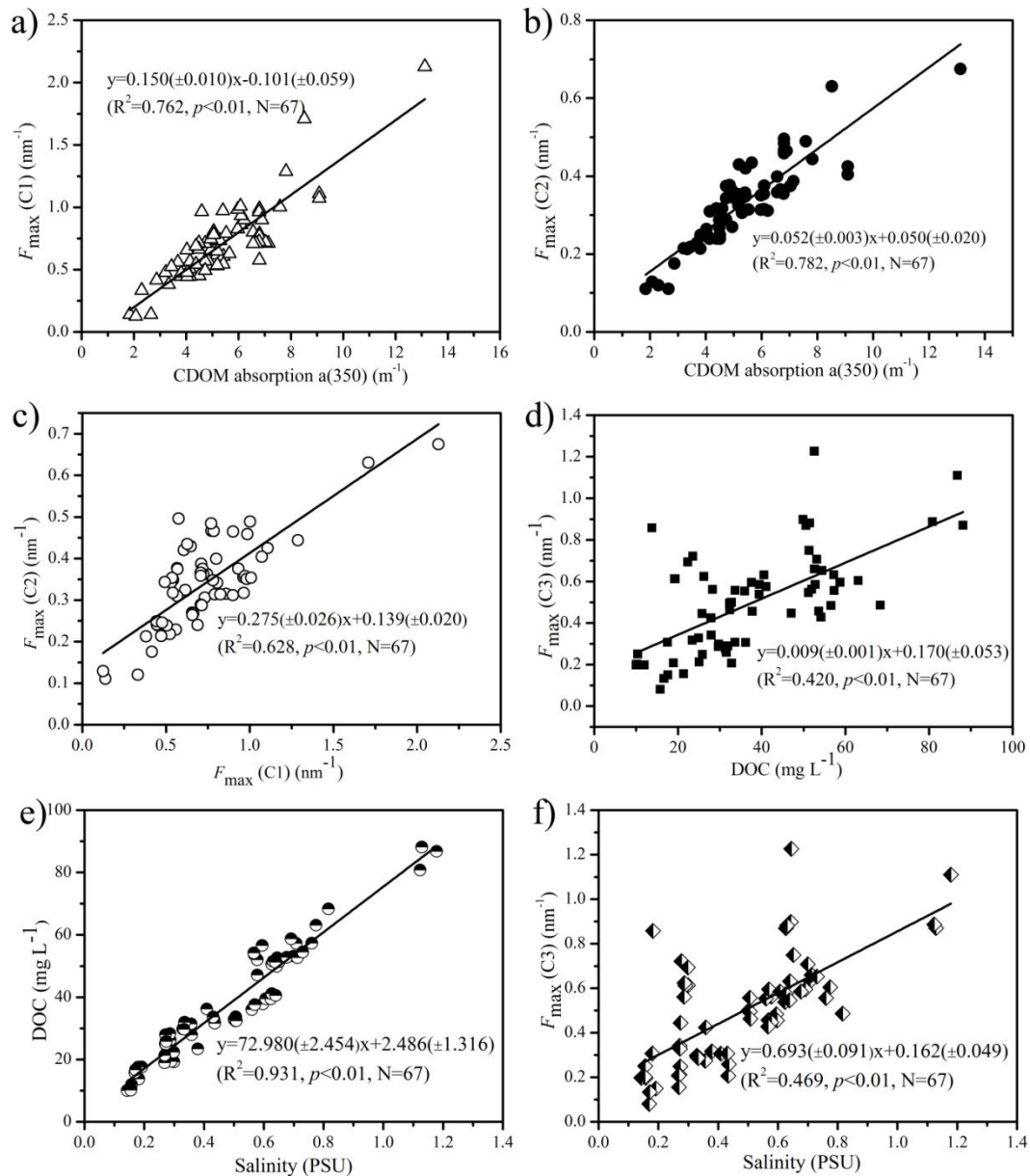
762 loadings (e-h) of each component. Fluorescence is in Raman units: nm^{-1} .

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767 Figure 3. a) Seasonal average of F_{max} for EEM-PARAFAC components (C1, C2, C3
 768 and C4) for lakes in the western part of Jilin province; b) Monthly variation of rainfall
 769 for the lakes in western part of Jilin province from April 2013 to February 2014; c)
 770 Seasonal variation of the total fluorescence intensity at different seasons; d) Seasonal
 771 variation of the four EEM-PARAFAC components at different seasons; The error bar
 772 represents standard deviations.



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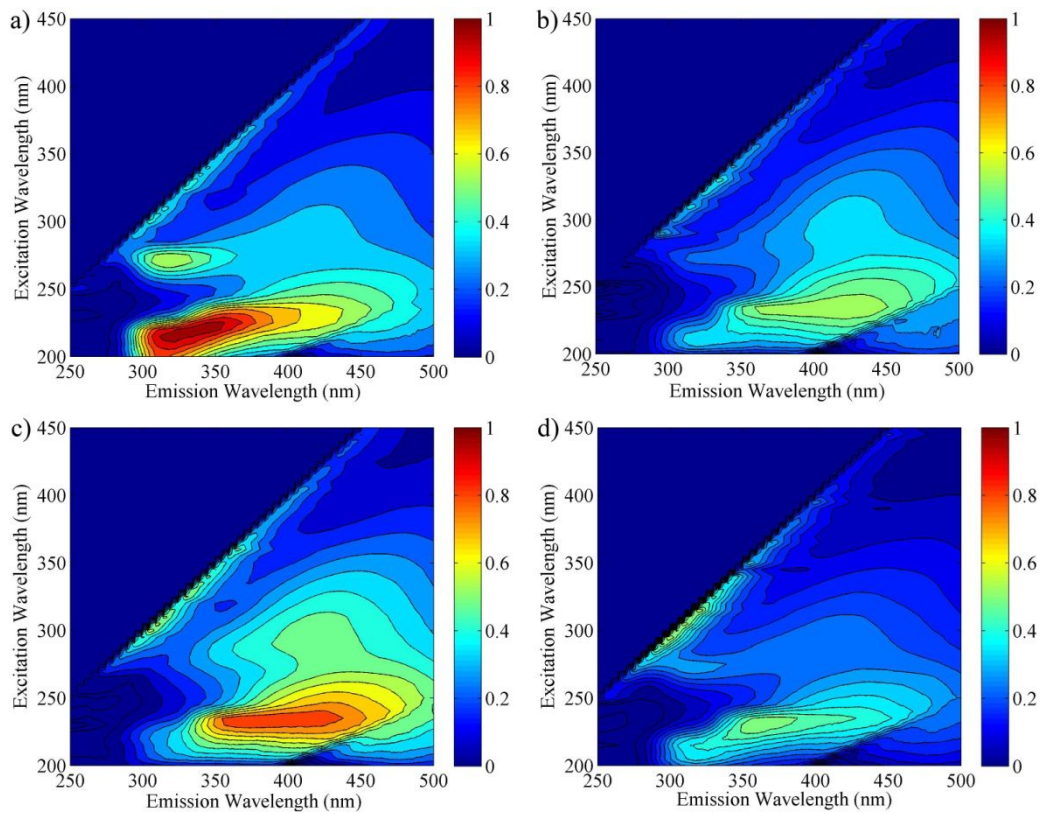
779 Figure 4. Relationships between CDOM absorption coefficient $a(350)$ with a)780 $F_{max}(C1)$, b) with $F_{max}(C2)$, c) $F_{max}(C1)$ versus $F_{max}(C2)$, d) $F_{max}(C3)$ versus DOC, e)781 Salinity versus DOC, f) Salinity versus $F_{max}(C3)$.

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787 Figure S1. Examples of EEMs for one water sample from Xindianpao Lake in the

788 western part of Jilin province at different seasons a) June 2013; b) August 2013; c)

789 February 2014; d) April 2014 (Fluorescence is in Raman unit: nm^{-1}).

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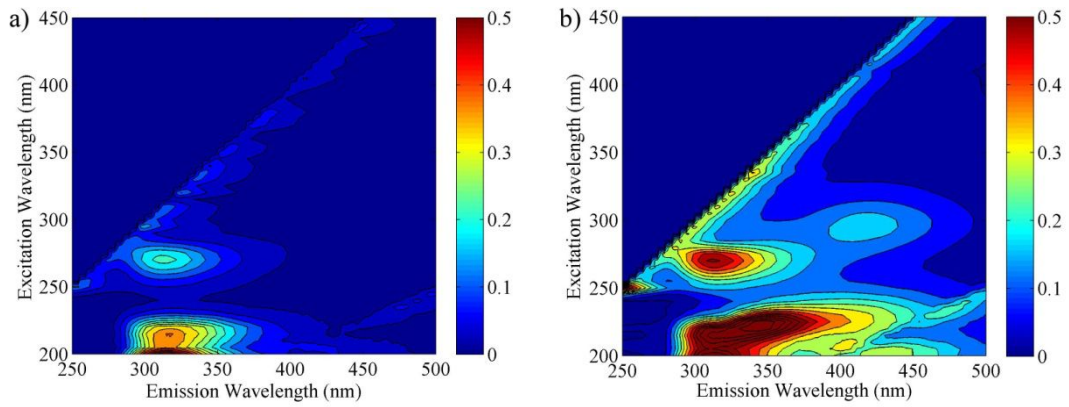
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800 Figure S2. Representative examples of EEMs for a) lake ice-melt water sample, and b)

801 rainwater CDOM in the western part of Jilin province (Raman: nm^{-1}).

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