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Relationships between burned area, forest cover loss and land use change in the Brazilian Amazon based on satellite data

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Abstract

Fires are used as a tool in the deforestation process. Yet, the relationship between fire and deforestation may vary temporally and spatially depending on the type of deforestation and climatic conditions. This study evaluates spatiotemporal dynamics of deforestation and fire represented by burned area over the 2002–2012 period in the Brazilian Legal Amazon. As a first step, we compared newly available Landsat-based maps of gross forest cover loss from the Global Forest Change (GFC) project with maps of deforestation extent from the Amazon Deforestation Monitoring Project (PRODES) produced by the Brazilian National Institute for Space Research (INPE). As a second step, we rescaled the Landsat-based data to the 500 m resolution of the Moderate Resolution Imaging Spectroradiometer (MODIS) burned area data (MCD64A1) and stratified this using MODIS land cover data to study the role of burned area in forest cover loss and deforestation. We found that while GFC forest cover loss and PRODES deforestation generally agreed on spatial and temporal dynamics, there were several key differences between the datasets. Both showed a decrease in the extent of forest cover loss or deforestation after 2004, but the drop was larger and more continuous in PRODES than in GFC. The observed decrease in forest cover loss or deforestation rates over our study period was mainly due to lower clearing rates in the evergreen broadleaf forests in the states of Mato Grosso, Pará and Rondônia. GFC indicated anomalous high forest cover loss in the years 2007 and 2010 not reported by PRODES. The burned area data showed that this was predominantly related to increased fire activity occurring outside of the tropical forest area during these dry years, mainly in Pará. This indicates that fire and forest loss dynamics in woodlands or secondary forests may be equally important as deforestation in regulating atmospheric CO_2 concentrations. In addition to the decrease in forest cover loss rates, we also found that post-deforestation fire use declined; burned area within 5 years after forest cover loss decreased from 54 to 39 % during our study period.

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1 Introduction

In the tropics, fires are regularly used as a cheap means to dispose of vegetative debris on deforested land for the purpose of agriculture or pasture, and to maintain an open landscape. Fire represents a major source of carbon emissions in the Amazon, especially in the southern fringe of the Amazon known as the arc of deforestation (Fearnside et al., 2009). Various factors contribute to relatively high rates of deforestation in this region, including available infrastructure and capital and relatively long dry seasons (Kirby et al., 2006). Fire susceptibility is influenced by factors such as deforestation, forest management, cultivation and reforestation (Osborne, 2000). Development of human infrastructure including road construction, and the spread of residential settlements in general increase deforestation and fire activity in more remote regions (Nepstad et al., 2001; Malhi et al., 2009). In 2005, approximately 90 % of the observed burning in the southeast of Rondônia and the northwest of Mato Grosso occurred in previously deforested areas, thereby demonstrating the use of fire in and after land use and land cover change (LULCC) processes (Lima et al., 2012).

The objectives of landowners partly govern the degree to which fire is used in the deforestation process. In general, three stages can be distinguished: logging, clearing and maintenance. In tropical logging, by only harvesting selected trees, bulldozers create networks of forest roads, killing non-harvested trees and fragmenting the forest and its canopy (Johns et al., 1996). This practice increases the amount of dead wood in the understory and the likelihood of fires. The damaged canopy allows light and wind to penetrate, increasing desiccation (Holdsworth and Uhl, 1997; Cochrane and Laurance, 2008). If fires occur they can affect forest composition and structure. Recovery is diminished due to the impact of surface fires on seed availability (Van Nieuwstadt et al., 2001). For clearing, slash-and-burn is a very efficient technique to clear land intentionally to become pasture or agricultural land (Tinker et al., 1996). To prepare the field, vegetation is cut and left on the ground to dry. At the end of the dry season this is

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subsequently burned. If biomass is not fully consumed after the first fire, the remains are piled up and burned again, often until all biomass is removed. Once everything is cleared, planting can be done directly in the ashes. Forest edge biodiversity is impoverished by grass invasion due to repetitive burning in rapid agricultural expansion (Balch et al., 2009). Maintenance fires are used regularly to prevent trees from invading pastures after deforestation, for nutrient recycling and to remove crop residues (van der Werf et al., 2009).

The vast scales of fire and deforestation have become increasingly evident as a result of developments in remote sensing and research that has used the continuously increasing data streams associated with this. Often these deforestation-focused studies used NASA's Landsat data (Tucker et al., 2004), but coarser scale Moderate Resolution Imaging Spectroradiometer (MODIS) data is also used for rapid identification of deforestation location areas and trends in dynamics (Morton et al., 2005). Landsat generated deforestation datasets allow accurate quantification of the total area of deforestation (Broich et al., 2009). Now spanning a time frame of approximately 25 years, the Brazilian National Institute for Space Research (INPE) has created the Amazon Deforestation Monitoring Project (PRODES) to quantify deforestation rates in the Brazilian Legal Amazon (BLA) (INPE, 2014). Initially only available in a tabular format, digital maps were created using a new methodology based on the work of Shimabukuro et al. (1998) from 1997 onwards (Hansen et al., 2008). PRODES has been used in many studies to evaluate the effects of deforestation (e.g., Morton et al., 2006; Broadbent et al., 2008). Also using Landsat satellite imagery, the Global Forest Change project (GFC) has compiled a global forest cover loss dataset at the same spatial resolution as PRODES with an annual time step for the period 2001–2013 (Hansen et al., 2013). The objectives of these two datasets diverge as PRODES only focuses on primary forest loss, while GFC includes forest changes in every type of vegetation taller than 5 m in height and is a global data set while PRODES only focuses on the BLA.

Fire has been used as part of the clearing process at least since the inauguration of the Transamazon Highway in 1970 (Fearnside, 2005). Different remotely sensed datasets allow the quantity and intensity of fires to be estimated. Active fire algorithms (e.g., Giglio et al., 2003) detect a fire at the time of satellite overpass, while burned area algorithms (e.g., Roy et al., 2008) use satellite time series of vegetation indices and other information to classify whether a pixel was burned or not. Burned area algorithms are less sensitive to cloud cover and smoke and are therefore more reliable for spatiotemporal studies allowing better analysis of spatial association between landscape attributes and burning events, while active fires are better for identifying smaller fires and yield near-real time information (Roy et al., 2008).

After the first fire following deforestation there is generally still too much biomass on the ground to enable agriculture or cattle grazing. Usually, repetitive burning is carried out to remove remaining biomass. The number of fires is thought to depend on the post-deforestation land use with agriculture requiring more complete combustion than pastures. For example, in a previous modeling exercise pasture combustion completeness was set between 50 and 90 %, while conversion to cropland may lead to 100 % combustion completeness over a period of 1–3 years (Morton et al., 2008). The connection between these fires and deforestation is well established. For example, according to a study focusing on one Landsat 5 TM tile (231/67 in the WR-2 indexing system) at the border of Rondônia and Mato Grosso, 31 % of the area burned in 2005 was deforested in one of the preceding 3 years (Lima et al., 2012).

Deforestation and forest cover loss rates are impacted by numerous socio-economic factors, and via the fire–deforestation link drought also plays a role in explaining spatio-temporal variability (van der Werf et al., 2008). In general, drier than normal dry seasons see increased fire activity (Chen et al., 2013) which can inhibit rainfall, creating a feedback mechanism (Nobre et al., 1991; Nepstad et al., 2001). In addition, the vulnerability of these forests may increase due to climate change (Aragão et al., 2008) with increased temperatures and stronger droughts (IPCC, 2007). In 2005, 2007, and 2010, three strong droughts occurred in the Amazon. Standardized anomalies

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of dry-season rainfall showed that in 2005 37 %, and in 2010 57 %, of the total area was experiencing lower than average rainfall, when defined as more than 1 standard deviation (SD) below the long-term mean (Lewis et al., 2011). The 2005 event increased the number of fires observed by satellite-derived monthly and annual time series by 33 % compared to the 1999–2005 mean (Aragão et al., 2007). Droughts are associated with the El Niño–Southern Oscillation, but are also impacted by tropical Atlantic sea surface temperature (SST) anomalies related to the Atlantic Multidecadal Oscillation (AMO) (Marengo et al., 2008). Eastern Amazonia is more vulnerable to fire during El Niño whereas the south and southwest is more linked to AMO (Chen et al., 2011). Southern and Eastern states in the Amazon basin are relatively dry with lower annual rainfall and stronger dry seasons, while the interior of the Amazon is more humid (Nepstad et al., 1994).

In this study we focus on the relationship between fires and forest cover loss using newly available data. We discuss and highlight the patterns seen in the forest cover loss, deforestation, and burned area datasets. Our main objective is to identify the role that fires play in these differences, and how droughts influence the fire–deforestation–degradation dynamics. We also focus on the spatial and temporal variability in the use of fire as a tool for deforestation for different states and land cover types.

2 Datasets and methods

Our study area is the Brazilian Legal Amazon or BLA, which consists of 9 Brazilian states in the northern part of Brazil (Fig. 1). Its geographic area is approximately 5 million km^2. 62 % of the Amazon forest lies within Brazilian boarders. It consists predominantly of closed tropical forest but also of cerrado (wooded savanna), agriculture, and pasture land (Carreiras et al., 2006). Deforestation is concentrated in the arc of deforestation along the southern and eastern edges of the BLA (Fearnside, 2005). We used four satellite-derived datasets characterizing forest dynamics and fire: GFC forest cover loss, PRODES deforestation, MODIS burned area, and MODIS land

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cover. After describing these datasets in more detail, we explain what steps we have taken to study relationships between the various datasets.

The PRODES deforestation dataset is derived from Landsat data using algorithms described by Shimabukuro et al. (1998). Landsat TM bands 3–5 are used to generate the fraction of vegetation, soil, and shade within each pixel using the constrained least squares method. The image fraction is then classified using image segmentation followed by unsupervised image classification and image editing to determine primary forest deforestation (Valeriano et al., 2004). This methodology has been used since 1997 and once an area is detected as deforested, it is masked out and not considered anymore in following years.

GFC defines forest cover loss as a stand-replacement disturbance or the complete removal of tree cover canopy for the period 2001–2013. GFC uses pre-processed cloud-free Landsat images to calculate time series spectral metrics. Over the study interval three groups of per-band metrics were employed: reflectance values, mean reflectance and slope of linear regression of band reflectance value vs. image date. Using these time-series metrics, a decision tree was used to find forest cover loss (Hansen et al., 2013). A key difference between GFC and PRODES is that PRODES only quantifies primary forest loss while GFC includes changes in "all vegetation taller than 5 m in height". This may include the clearing of secondary forest or certain types of cultures such as palm oil production, and in general one expects that GFC flags more area than PRODES. PRODES only focuses on primary forest loss, or deforestation, while using GFC, it is potentially possible to differentiate between deforestation and degradation. In the Reducing Emissions from Deforestation and forest Degradation (REDD) context, deforestation has been defined as a "measurable sustained decrease in crown cover" below a 10–30 % threshold (UNFCCC, 2006), while degradation is a loss of biomass density without a change in the area of forest cover (Olander et al., 2008).

MODIS burned area (MCD64A1) is based on 500 m MODIS imagery coupled with 1 km MODIS active fire observations. It produces composite imagery to detect persistent changes and then uses spatial and temporal active-fire information to create probabilistic thresholds to classify each pixel as burned or unburned (Giglio et al., 2009). To identify land cover changes, we used the MODIS MCD12Q1 Land cover type product version 051 (Friedl et al., 2010) with the International Global Biosphere Program (IGBP) global land cover classification scheme (Loveland and Belward, 1997).

To combine deforestation, burned area and land cover datasets we used ArcMap10 to rescale PRODES and GFC to the 500 m MODIS grid using the most common Z value. The rescaled dataset slightly overestimated the area of forest loss in the BLA by approximately 4 %. However, this bias was uniform over time and space, and did not change the general trend and patterns in which we are interested here. We performed all analyses at 500 m resolution and rescaled the results to 0.1° to facilitate visual interpretation of the figures.

Because of the annual time step in the GFC and PRODES data, we have also merged the monthly burned area data into annual files to compare that layer with deforestation and forest cover loss. We have simplified the land cover classification scheme to three classes with the dominant vegetation types in the BLA: Evergreen Broadleaf forest (EB), Low Tree Cover (LT, which includes shrublands and savannas) and Cropland (C, including croplands interspersed with natural vegetation). Other land cover type, such as grassland, needle leaf forest or permanent wetland, have been merged (Other) but excluded in our analyses as they only represent a very small portion of BLA land cover.

Finally, we aimed to not only better understand how burned area and forest cover loss are related, but also how frequently burned area is observed after land cover transition and whether this varies between post-deforestation land use. For each evergreen broadleaf pixel where GFC indicated forest cover loss, we summed burned area occurrence during the 5 years following this event. Because our time series ended in 2012 we could only do this for 2002 through 2007. In the process, we kept track of the post-deforestation land use defined by the land cover type 5 years after deforestation. For each year of evergreen broadleaf forest cover loss, we calculated

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the fraction burned in the subsequent 5 years, and associated them with the initial year of deforestation. Finally, we stratified these summed fractions by post-deforestation land cover type to evaluate the trend of conversion.

3 Results

During our study period, within the BLA, both deforestation and burned areas were concentrated along the arc of deforestation (Fig. 2). The largest difference between GFC and PRODES on one hand and burned area on the other hand is that forest loss and deforestation are concentrated along the edge of the Amazon forest while most fire activity is observed further away from the deforestation frontier. In addition to the arc of deforestation, forest cover loss and deforestation occurs along the Amazon River and close to roads. A state-level breakdown indicated that Mato Grosso (both GFC and PRODES indicate 33 % of the total BLA loss), Pará (32 % for GFC, 36 % for PRODES) and Rondônia (12 and 13 %) were the states with most forest cover loss over our full study period. There is spatial overlap between forest cover loss/deforestation and burned area, mainly in the Southeast of Pará and in Eastern Mato Grosso. However, regions with more shrubland or savannas, further away from the arc of deforestation (e.g. Tocantins and Maranhão) have the highest burned area.

Over our study period, PRODES and GFC annual forest cover loss increased up to the year 2004 and decreased afterwards (Fig. 3). While both datasets demonstrated this post-2004 decline, the extent differed. PRODES showed a continuous decrease except for 2008, while GFC plateaued towards the end of the study period at a higher level. In both datasets, all states showed a decrease between 2004 (highest year of deforestation during our study) and 2012. Mato Grosso, Pará and Rondônia showed the largest reduction between 2004 and 2012 in both datasets, respectively 63, 50 and 56 % in GFC and, 94, 80 and 80 % with PRODES. While all other states had a decrease of at least 40 % according to PRODES, GFC does not show any other state having more than a 30 % decrease. In addition, GFC has annual peaks in forest cover loss

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during dry years, especially in 2010, while PRODES does not show this variability in deforestation.

Partitioning GFC forest cover loss by land cover, we found that until 2006 most loss occurred in evergreen broadleaf forests (Fig. 4a). From 2002 to 2009, the amount of annual evergreen broadleaf forest loss dropped from approximately 16 000 to almost 5000 km^2 and stayed relatively constant in the following years. The most significant decrease came from Mato Grosso and Rondônia with an absolute forest cover loss reduction in evergreen broadleaf forests of 93 and 81 % respectively (Fig. 4b). The decrease in forest cover loss started earlier in Mato Grosso (in 2004) than in Rondônia (in 2005) probably related to the strong 2005 drought, which was most prominent in the western BLA. The state with on average the second highest amount of forest cover loss, Pará, showed a slower diminution in evergreen broadleaf forest loss (60 %). Total loss in other types of land cover remained relatively stable during our study period (Fig. 4c and d). Mato Grosso is the only state with a significant decrease in clearing in the low tree cover biome, halving the amount of shrubland and savannas cleared between 2004 and 2012.

In contrast to these trends in deforestation and forest cover loss, total burned area remained relatively stable over our study period except in drier years, such as 2007 and 2010, when burned area increased markedly, especially in low tree vegetation. These peaks mainly occurred in southern and eastern states such as Mato Grosso, Pará, Maranhão and Tocantins (Fig. 5a and b). Focusing on burned area coinciding with GFC forest cover loss (so observed in the same grid cell in the same year), we found that they occur mainly in evergreen broadleaf forest and that this is only a small fraction of the total burned area in the BLA (Fig. 5c and d). The general trend is similar to total GFC although the decrease starts after 2005 due to a drought in the southwest of the BLA, mainly in Acre and Amazonas. Looking into more detail at 4 years (2004, 2005, 2007 and 2010) with more deforestation and/or fire activity, we found that the spatial distribution of coinciding forest cover loss and burning has changed (Fig. 6). Especially in 2004 but also in 2005, most of this was observed in northern Mato Grosso

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and Rondônia. These regions featured less prominently in latter part of this analysis during the dry years of 2007 and 2010. In contrast, Pará had very low fire activity in 2004, but started to have a major hotspot in the southeast in 2005 and in following dry years. The central region of Pará had also more activity from 2005 onwards. Maranhão and Tocantins have little remaining primary forest but the use of fire in the process of clearing other vegetation above 5 m of height has increased during our study period.

Using burned area and GFC to evaluate the time interval between forest cover loss and fire we found that a substantial part (46 % when averaged over the full study period) of cleared evergreen broadleaf forest was burned within 5 years after forest cover loss, with the use of fire decreasing over time. 24.5 % was converted to shrubland or savannas, 18 % was converted to cropland and the remaining 3.5 % had no permanent conversion (Fig. 7). We also found that 10 % of the land converted to shrubland or savanna was burned the same year as forest cover loss was detected in these grid cells (we could not identify whether it was the cause of conversion), while conversions to cropland were burned in the first year in about 7.5 % of the grid cells. Grid cells identified as forest cover loss but remaining evergreen broadleaf forest had little burned area, between 0–2 %.

Strikingly, when looking at the temporal trend of burned area after evergreen broadleaf GFC loss, we found that the use of fire after conversions to shrubland, savannas or cropland decreased over our study period (Fig. 8). The fraction that burned after or during conversion to shrubland or savannas decreased from over 31 % between 2002 and 2007 to 17 % between 2007 and 2012. Combining all conversion types, post-deforestation fire use was most frequent for the grid cells that had forest loss in 2003 where 60 % burned in the following 5 years, decreasing to 29 % for 2006, but increasing again to 39 % in 2007. This trend came mainly from converted lands, in the grid cells without permanent conversion the use of fire actually increased somewhat (Fig. 8).

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4 Discussion

Multiple lines of evidence demonstrate the successful decrease of deforestation within the BLA under the 2004 Action Plan for the Prevention and Control of Deforestation in the Legal Amazon. This plan introduced innovative procedures for monitoring, environmental control and territorial management (Assunção et al., 2015). In both datasets used in this study, the post-2004 drop in deforestation can be clearly identified. According to GFC, the resulting downward trend is mainly occurring in states within the arc of deforestation. In 2006, the soy moratorium was implemented to prohibit the trade of soybean produced in deforested areas (Meyfroidt and Lambin, 2011). After 5 year of application, Rudorff et al. (2011) concluded that although it was premature to attribute the decrease in deforestation in the Amazon biome to this moratorium, it had undoubtedly created an inhibitory effect on the expansion of soybean in this biome. After this moratorium, 90 % of the soybean plantation expansion occurred on land previously used as rangeland (Lapola et al., 2010). In the future, the Brazilian government is willing to significantly increase the production of biofuel (Bergmann et al., 2013). One consequence could be the migration of cattle ranchers to other regions and thereby increasing deforestation (Nepstad et al., 2006). According to GFC, forest cover loss ceased to decrease shortly after the implementation of this moratorium. This study cannot connect these two facts, but one hypothesis could be that more non-primary forest has been cleared to allow the soybean extension or the resettlement of cattle ranchers without breaching the moratorium agreements.

During our study period, the BLA faced severe drought years in 2005, 2007 and 2010. According to PRODES, these events did not influence the rate of deforestation, except in 2008 when there was a small increase possibly related to the 2007 drought. GFC shows more sensitivity to droughts. This is most evident in 2010 when there was an increase in forest cover loss compared to the previous year, the first time since 2004. But also 2005 and 2007 showed less of a decline compared to the previous year in GFC than in PRODES. These dry years had also a significant effect on fire in

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southern and western states. Maranhão and Tocantins had a very strong burning peak, but this did not influence forest cover loss, probably because savannas are by far the dominant land cover here. On the other hand, Mato Grosso and Pará have also seen elevated burning activity increasing the burning of evergreen broadleaf forest, which could explain part of GFC plateau towards the end of our study period.

Pará ranked second with regard to forest cover loss and deforestation in 2002, behind Mato Grosso. Mato Grosso drastically reduced its tree clearing rates, and so did Pará but to a lesser extent. Mato Grosso had only 44 % of its state covered by evergreen broadleaf forest in 2002 compared to 80 % for Pará. The availability of land accessible to clear may explain part of the difference in absolute trend strength. We observed that some regions in Pará that were not substantially impacted by clearing fires early in our study period became hotspots towards the end. For example, in the region of Tucumã in central Pará, fire and clearing were not collocated in 2004, but became really active in 2005, 2007 and 2010. A highly fragmented forest region may potentially explain this, as a close access to the Rio Xingu, tributary of the Amazon River, facilitates tree removal. Further north in Pará, we noted that strong deforestation was occurring without burned area being observed. This could be due to higher annual rainfall, which can limit the use of fire in forest clearing.

Previous work has shown a strong correlation between deforestation and fire occurrences (Aragão et al., 2008; Chen et al., 2013). Although deforestation and forest cover loss have declined sharply since 2004, burned area did not follow this trend because only a small fraction of burned area is associated with forest cover loss and because droughts have a strong impact on the inter-annual variability in burned area. Although the effects of droughts are not uniform between all states, general impacts are evident. Burned area always exceeded forest cover loss area due to the fact that fires in grassland and on agricultural grounds are not part of deforestation or forest cover loss. States with a relatively small fraction of evergreen broadleaf forest but a larger portion of cerrado (low tree cover), such as Mato Grosso, Tocantins and Maranhão

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have in general more fire activity in the conversion to agriculture or grazing process (Mistry, 1998).

Our results on time intervals between forest cover loss and burned area show that although large fire years occurred throughout our study period, the overall rate of burned area per unit deforested land is decreasing. Over our study period the post-conversion use of fire decreased. About 54 % of the evergreen broadleaf forest that was cleared in 2002 burned in the following 5 years, dropping to 29 % for the evergreen broadleaf forest cleared in 2006. 2007 was a dry year, the second highest fire year during our study period. We have shown that most fires occur in the same year as deforestation. Longer time series will enable us to see whether the increase in post-forest loss fire use in 2007 was a temporary result because of the drought or that the downward trend is increasing again. Lima et al. (2012) showed in one Landsat 5 TM tile located between Rondônia and Mato Grosso that 53 % of the 2005 fires occurred in land deforested within the prior 5 years. Aragão and Shimabukuro (2010), however, found that between 2000 and 2007, 59 % of areas experiencing reduction in deforestation rates had an increase in fire events. Using 0.25° spatial resolution, they explain that this reverse trend could come from secondary forest clearing, not detected by PRODES. We demonstrated that within the same time period, more than half of the cleared vegetation taller than 5 m in height, was burned within a few years, dropping below 40 % after 2004.

By combining the three different datasets we have aimed not only to understand the differences between them, but also to highlight uncertainties associated with relying on one dataset. For example, large scale carbon budget studies often use deforestation rates in their assessments of sources and sinks of carbon (Le Quéré et al., 2013). Our study shows that, at least for the Brazilian Amazon, degradation and fire dynamics outside the tropical forest domain may be equally important in regulating atmospheric CO_2 concentrations. These types of fires are also important in other regions outside the BLA (e.g. Chen et al., 2014) and expanding the current focus on deforestation regions to also include woodland dynamics may be an important step in better understanding biosphere–atmosphere interactions and biodiversity issues.

One key uncertainty we have not addressed here is the use of burned area to study fire extent. The burned area algorithms are often geared towards detecting relatively large fires and not all fires in deforestation zones may be flagged as burned by them. The use of active fires and ancillary data may mitigate some of these issues (Randerson et al., 2012) but the footprint of active fire detection is four times larger than that of burned area and active fires have their own limitations related to overpass times and cloud cover. In the future, higher resolution observations from for example VIIRS (Schroeder et al., 2014) may enable simultaneous use of burned area and active fires in native resolution applications like ours.

5 Conclusions

We used satellite observation of deforestation (PRODES), forest cover loss (GFC), burned area (MCD64A1) and land cover (MCD12Q1) between 2002 and 2012 in the Brazilian Legal Amazon (BLA) to study relationships between the different datasets. The PRODES dataset indicated that deforestation has sharply declined since 2004, something also seen in the GFC forest cover loss data but to a smaller extent. The states with the strongest decline were in the south of the BLA. Most deforestation occurred on the edge of the Amazon forest while more burned area was located on the outskirt of the arc of deforestation. Looking at coinciding forest cover loss and burned area, we found variability in the spatial distribution during our study period. This was most predominantly found in the state of Pará, where burned area and forest cover loss were not linked in 2004, which changed later when large patches of burned area coinciding with forest cover loss were observed in the southwest and the centre of the state during the dry years of 2007 and 2010. This indicates that droughts not only enhance burned area, but also increase forest cover loss in areas not monitored by PRODES and call for including woodland dynamics in carbon cycle studies.

Early in our study period, more than 50 % of cleared evergreen broadleaf forest was burned throughout the subsequent 5 years, reducing to below 40 % during the last time segment studied (2007–2012). In other words, besides the well-documented decrease in deforestation rates we also showed that fire intensity in deforested area decreased over our study period. More research is required to better understand these dynamics; one potential reason could be that with lower deforestation rates the deforested lands are becoming more intensively managed. Combined, our results indicate a decoupling of burned area and evergreen broadleaf forest deforestation but a stronger link between burned area and degradation.

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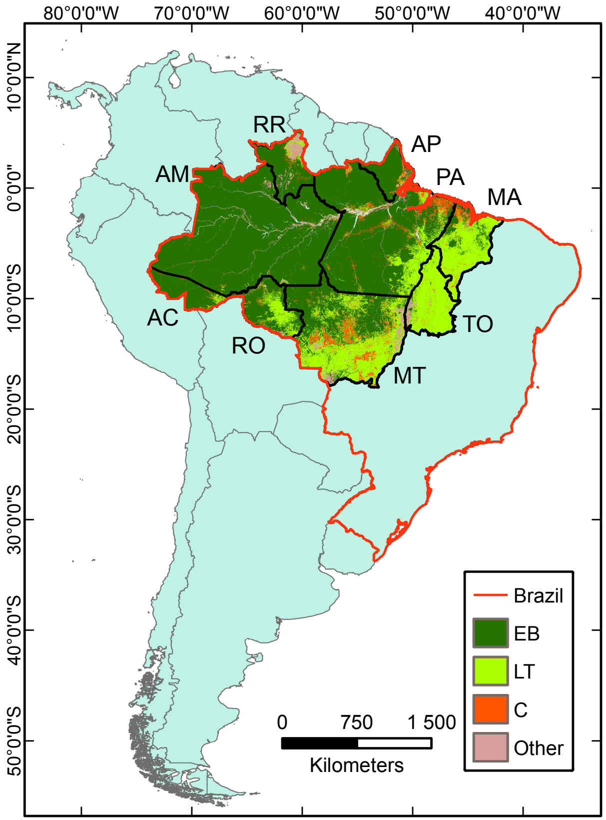

Figure 1. MODIS MCD12Q1 (v051) Land cover in the Brazilian Legal Amazon (BLA). EB = Evergreen Broadleaf forest, LT = Low Tree cover (shrubland and savannas) and C = Cropland and Cropland/Natural Vegetation Mosaic. Other includes all remaining land cover such as grassland, needle leaf forest, permanent wetlands or urban areas. Letter codes for the provinces in the BLA are clockwise: AP, Amapá; PA, Pará; MA, Maranhão; TO, Tocantins; MT, Mato Grosso; RO, Rondônia; AC, Acre; AM, Amazonas; RR, Roraima.

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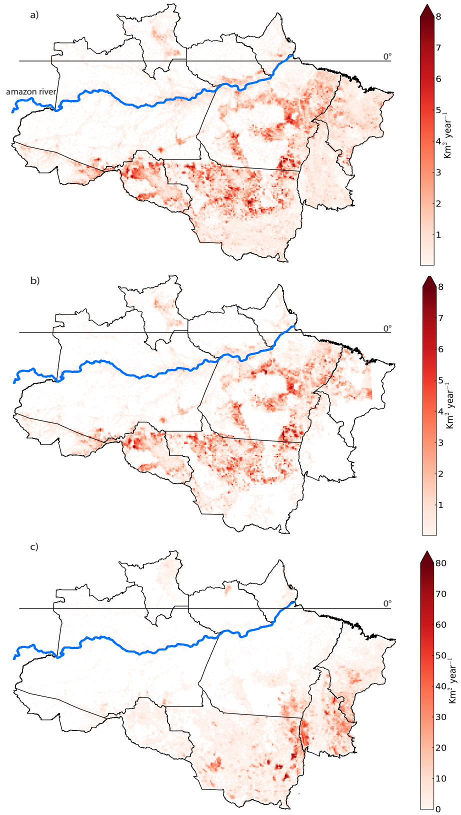

Figure 2. 2002–2012 mean forest loss (**a**, based on GFC), deforestation (**b**, based on PRODES), and burned area (**c**, based on MCD64A1). All data sets are regridded to 0.1° resolution. Note the order of magnitude difference between forest loss and deforestation one hand and burned area on the other hand.

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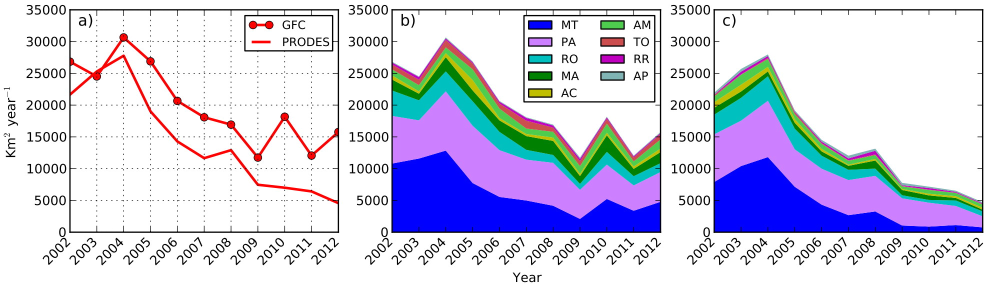

Figure 3. Annual forest loss estimates from the Global Forest Change (GFC) project and deforestation estimates from the Amazon Deforestation Monitoring Project (PRODES) for 2002–2012 in all of the BLA **(a)**, and partitioned on state level for GFC **(b)** and PRODES **(c)**.

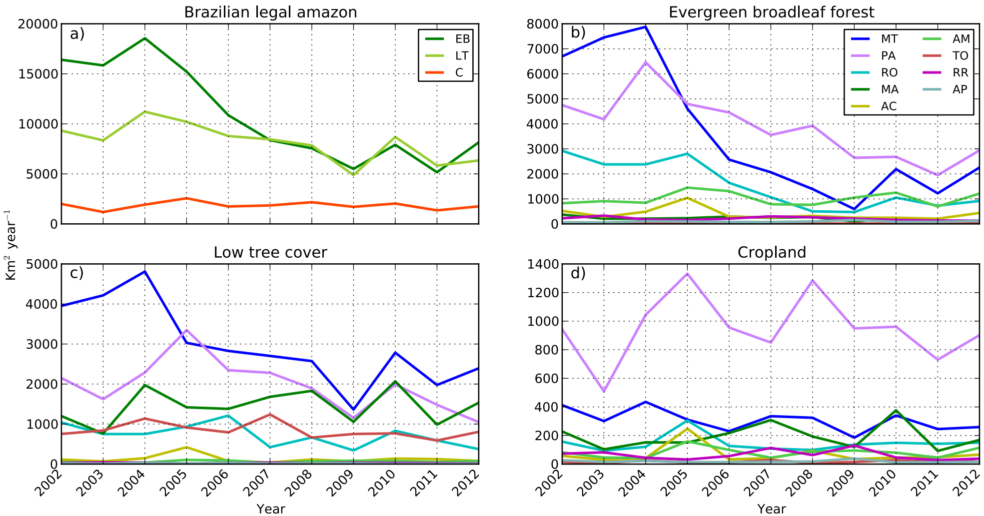

Figure 4. Annual state-level GFC forest loss for the Brazilian legal amazon and three land cover classes defined at the year of forest loss.

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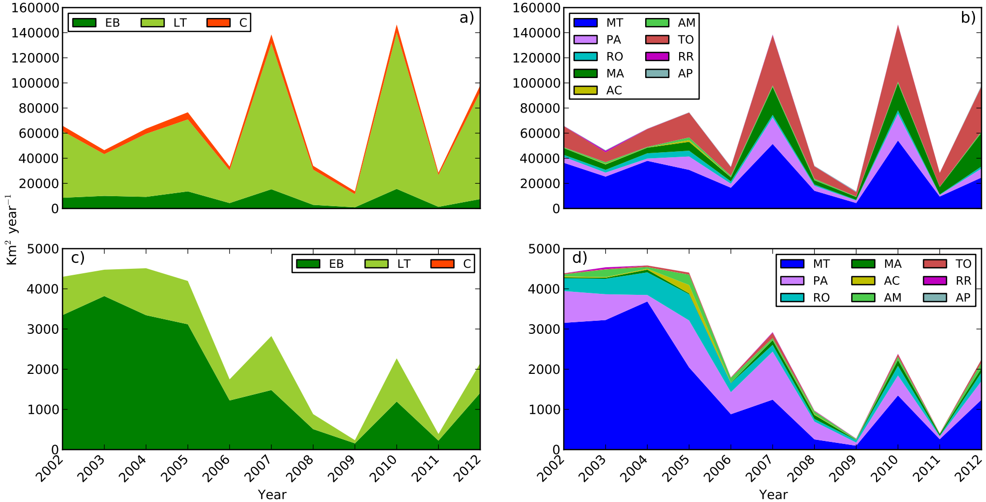

Figure 5. Burned area stratified by land cover **(a, c)** and by state **(b, d)**. Top panels (**a** and **b**) indicate total burned area; bottom panels show burned area coinciding with GFC forest loss.

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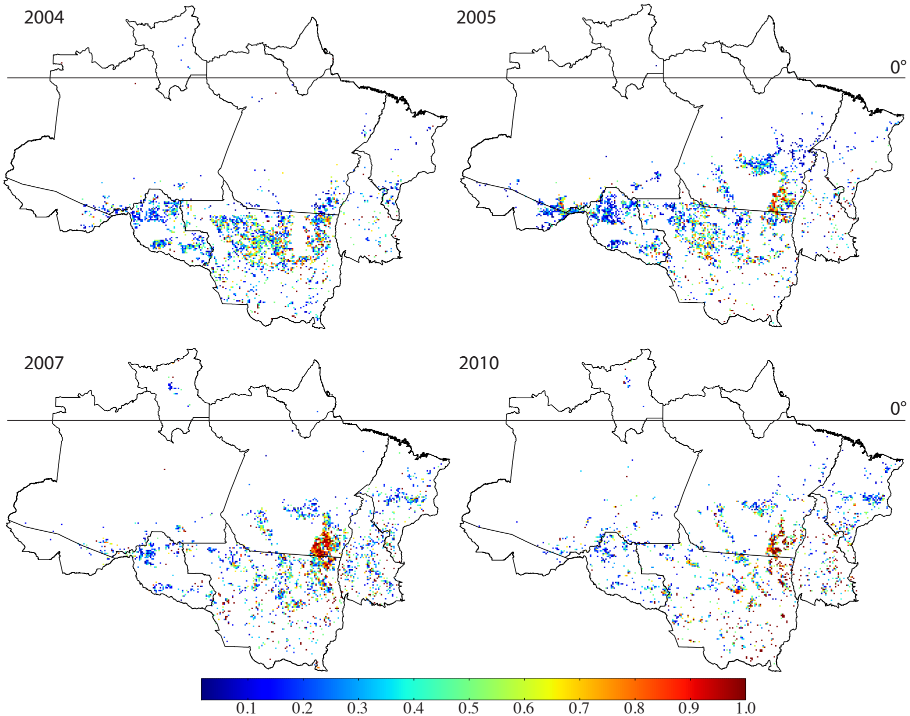

Figure 6. Fraction of GFC forest loss coinciding with burned area for the 4 highest fire and/or tree cover loss years in our study period, averaged on a 0.1° resolution.

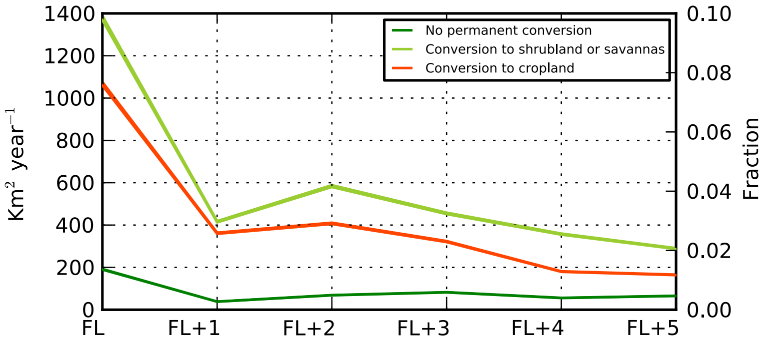

Figure 7. The degree to which fire is used in the deforestation process for different post-forest loss land uses, shown for the year of forest loss (FL) and 5 years afterwards. Data is averaged for the 6 year periods starting between 2002 and 2007. The right *y* axis indicates the burned area as a percentage of 2002–2007 averaged evergreen broadleaf forest loss (14 211 km^2 year^{-1}) and in total sums up to 46 % of total forest loss.

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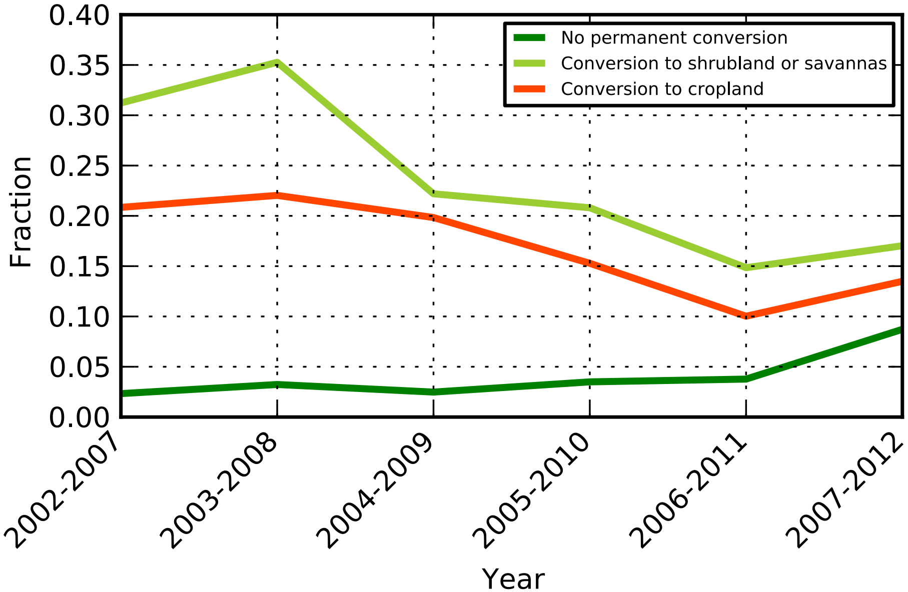

Figure 8. Fraction of total evergreen broadleaf forest loss grid cells where burned area is observed in the 5 years following forest loss. Results are shown for three different post-forest loss land cover types defined as the land cover type observed 5 years after the forest loss event.

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