6781

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Ocean acidification affects iron speciation in seawater

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Abstract

Rising atmospheric CO_2 is acidifying the surface ocean, a process which is expected to greatly influence the chemistry and biology of the future ocean. Following the development of iron-replete phytoplankton blooms in a coastal mesocosm experiment at 350,

⁵ 700, and 1050 μ atm *p*CO₂, we observed significant increases in dissolved iron concentrations, Fe(II) concentrations, and Fe(II) half-life times during and after the peak of blooms in response to CO₂ enrichment, suggesting increased iron bioavailability. If applicable to the open ocean this may provide a negative feedback mechanism to the rising atmospheric CO₂ by stimulating marine primary production.

10 **1** Introduction

Paleoclimate data indicate significant effects from the deposition of iron in aeolian dust on ocean biogeochemistry with feedbacks on global climate (Watson et al., 2000). Studies of artificial and natural iron input have demonstrated iron control of phytoplankton productivity and CO₂ drawdown over vast oceanic regions (Boyd et al., 2007; Blain et al., 2007; Pollard et al., 2009) and in coastal upwelling regions (Bruland et al., 2001; Hutchins and Bruland, 1998). Temporal control of iron on phytoplankton productivity was also observed in a Norwegian fjord system (Öztürk et al., 2002). The Pelagic ecosystem CO₂ enrichment study (PeECE III) studied natural phytoplankton blooms under atmospheric CO₂ scenarios of 350, 700, and 1050 μ atm *p*CO₂ in a coastal mesocosm experiment (Schulz et al., 2008). We used this unique opportunity to investigate the combined effects of phytoplankton bloom development and ocean acidification on Fe chemistry.

Iron solubility in surface seawater is low and the speciation is largely controlled by organic complexation and photochemical processes (Kuma et al., 1996; Sunda and
 Huntsman, 2003; Kuma et al., 1992). In addition to organic compounds generally present in coastal seawater, phytoplankton blooms can affect Fe(III)-complexation,

BGD

6, 6781-6802, 2009

Ocean acidification affects iron speciation in seawater





which affects the dissolved Fe fraction and photoreactivity (Croot et al., 2001; Kuma et al., 1992; Rue and Bruland, 1995). Ligand production is generally facilitated via zooplankton and protist grazing, microbial production, and potentially also by phytoplankton cell degradation (Heldal et al., 1996; Barbeau et al., 1996; Hutchins and 5 Bruland, 1994). In return, iron bioavailability differs among types of organic iron complexation and prokaryotic or eukaryotic phytoplankton (Hutchins et al., 1999). Fe(III)complexation is interlinked with Fe(II) production in marine water, mainly via supplying the substrate for photoreduction in sunlit surface waters (Kuma et al., 1992; Oztürk et al., 2004; Barbeau et al., 2001; Barbeau et al., 2003). Fe(II) is generally considered bioavailable, but rapid reoxidation to Fe(III) in temperate waters result in limited con-10 centrations. Half-life times range in the order of minutes and are largely dependent on temperature, oxygen and hydrogen peroxide concentrations, and pH (Santana-Casiano et al., 2005; Millero et al., 1987; Millero and Sotolongo, 1989), with presently limited knowledge about the role of organic Fe(II) complexation in natural seawater. However, the contribution of Fe(II) to phytoplankton nutrition may be significant in the open ocean 15

(Roy et al., 2008) as well as in estuarine waters (Breitbarth et al., 2009). Despite our growing knowledge of iron biogeochemistry in seawater, we currently have little information on the effects of ocean acidification thereon. Seawater pH affects phytoplankton physiology (Fu et al., 2008; Hare et al., 2008; Riebesell, 2004) and

- thus indirect effects via phytoplankton exudates that complex iron may also alter biological influences on iron solubility and cycling. Further, the photoreduction of organic Feligand complexes is the main pathway for Fe(II) production in the euphotic zone (King et al., 1993) and moreover Fe(II) oxidation rates are strongly pH dependent (e.g. Santana-Casiano et al., 2005). For unabated CO₂ emissions, oceanic uptake of anthropogenic
- ²⁵ CO₂ will lower surface ocean pH from a pre-industrial 8.25 to an estimated 7.85 within this century, and further by up to 0.7 units until 2300 (Caldeira and Wickett, 2003; Jacobson, 2005). This change far exceeds any glacial-interglacial differences (Caldeira and Wickett, 2003) and thus may have profound effects on the biogeochemistry of iron in seawater. The PeECE III experiment offered a unique opportunity to study iron

BGD

6, 6781-6802, 2009

Ocean acidification affects iron speciation in seawater





chemistry in phytoplankton blooms grown simultaneously at different pCO_2 concentrations. We particularly evaluate dissolved iron concentrations, as well as Fe(II) levels and oxidation rates over the course of the PeECE III mesocosm study and elucidate if ocean acidification may affect iron speciation in seawater.

5 2 Methods

The experimental work was carried out from 15 May through 09 June 2005 at the National Mesocosm Facility located at the Espeland Marine Biological Station at the Raunefjord (60.3° N, 5.2° E), University of Bergen, Norway. Nine polyethylene (PE) enclosures (2 m diameter, 10 m deep, hereafter called mesocosms), were moored to a raft about 200 m from shore. The mesocosms were capped with gastight and light 10 transparent (95% UV permeability) tents. Atmospheric CO₂ concentrations within three mesocosms each were adjusted to 350, 700, and $1050 \,\mu \text{atm} \, p\text{CO}_2$. Phytoplankton blooms were initiated with NaNO₃ and Na₂HPO₄ additions. The pH values are calculated based on daily measurements of total alkalinity (TA) and dissolved inorganic carbon (DIC) in the mesocosm upper mixed layer and are expressed on the total pH 15 scale. TA was measured using the classical Gran electrotitration method (Gran, 1952) (precision $\pm 4\mu$ mol kg⁻¹). DIC was measured by coulometric titration (Johnson et al., 1987) with a precision of $2 \mu \text{mol kg}^{-1}$. Chl-*a* concentrations were determined using HPLC (Barlow et al., 1997) and particulate organic carbon (POC) was measured on an elemental analyzer (EuroEA 3000, EuroVector) (Ehrhard and Koeve, 1999). Please 20

see Schulz et al. (2008) for more details about the experimental set-up. Total (tFe) and dissolved iron (dFe, $0.2 \mu m$ filtered) measurements were conducted using chemiluminescence flow injection analysis (CL-FIA, Waterville Analytical) (Bowie et al., 1998) and focused on the period of strong bloom development and pH shifts (days 7, 12) and a past bloom measurement (day 22). Fo(II) was determined based

(days 7–13) and a past bloom measurement (day 23). Fe(II) was determined based on Croot and Laan (2002) using the same CL-FIA instrument, which was installed on the mesocosm raft for this purpose, during days 20 and 22 of the experiment. Samples

BGD 6,6781-6802,2009 Ocean acidification affects iron speciation in seawater E. Breitbarth et al. **Title Page** Abstract Introduction Conclusions References **Figures Tables** Back Close Full Screen / Esc **Printer-friendly Version** Interactive Discussion

were obtained from 2.5 m depth via syringe pumping through PTFE tubing, immediately followed by $0.2 \,\mu$ m filtration and injection into CL-FIA system. Additionally, Fe(II) oxidation rates were determined on $0.2 \,\mu$ m filtered samples form day 25 of the experiment, which were stored in the dark for 24 h to oxidize any Fe(II) present. Measurements were

- ⁵ performed while having the samples, all reagents, and the sample loop of the flow injection analyzer immerged in a temperature controlled (10°C) water bath, in order to minimize analytical temperature effects during the measurements. The temperature was chosen as it is representative of the temperature inside the mesocosms during the study. The naperian log transformed chemiluminescent signal over time of 0.5, 0.75,
- ¹⁰ and 1 nmol L⁻¹ Fe(II) standard additions yield a linear signal decrease and the slope represents the Fe(II) oxidation rate constant ($k_{ox} s^{-1}$). Predicted Fe(II) oxidation rates were further calculated based on Millero et al. (1987) after calculating the [OH⁻] concentration of the water using the CO₂SYS program (Lewis and Wallace, 1998), taking seawater carbonate chemistry measurements and oxygen concentrations into account.

15 3 Results and discussion

CO₂ perturbation and phytoplankton bloom development result in pH shifts from 7.67– 7.97, 7.82–8.06, and 8.13–8.26 at 1050, 700, and 350 μ atm *p*CO₂, respectively during days 3–12 and relatively constant levels thereafter (Figs. 1 and 2). The peak of the bloom is marked by Chl-*a* concentrations at days 9 (low CO₂) and 10 (mid and high CO₂) and results in particulate organic carbon (POC) concentrations peaking at days 10 (low CO₂, 91.9 μ mol L⁻¹) and 11 (mid and high CO₂, 98.9 and 97.4 μ mol L⁻¹ respectively) (Figs. 1 and 2). Please see Schulz et al. (2008) for more details about the bloom development.

The Chl-*a* and POC biomasses in the three pCO_2 treatments are not markedly different. Nevertheless, our experiments show significantly higher dissolved iron (<0.2 μ m, dFe) concentrations for high CO₂ treatments in comparison to the mid- and low CO₂ scenarios (e.g. 4.42 vs. 1.92 and 2.73 nmol L⁻¹ on day 9, Fig. 3). During the bloom dFe



decreases and reflects iron uptake in all treatments. Remineralization during bloom decline increases dFe levels again and dFe was maintained at significantly higher levels in the future scenario compared to the mid and low CO₂ treatments. Distinctions in dFe towards the end of the bloom (day 23, 2.8–4.2 nmol L^{-1} at low and mid CO₂, 6.0– $_{5}$ 8.2 nmol L⁻¹ under high CO₂, Fig. 3) suggest differences in Fe remineralization in the treatments. Fe concentrations in the mesocosms were higher than in nearby fjord water (6.2 nmol L^{-1} total Fe (tFe) and 3.0 nmol L^{-1} dFe on day 13). Total Fe ranged from 23.7 to 96.1 nmol L⁻¹ and varied between enclosures and over time. However, dissolved Fe values do not correlate with tFe and thus tFe concentrations are not responsible for the systematically increased dFe values in the high CO₂ treatments (Fig. 4). The total Fe data suggest an input of relatively unreactive particulate Fe during the filling process of the enclosures. Clearly, the mesocosms were Fe-replete and while we lack iron measurements from the start of the experiment, applying an Fe:C ratio of 65 μ mol:mol (Sarthou et al., 2005) yields an iron demand of 2-6 nmol L⁻¹ during the bloom, which approximates the iron detected in fjord water. 15

Any Fe(III)-hydroxide solubility change over the observed pH ranges during the iron measurements (7.77–8.21 up to 7.94–8.26, Fig. 2) is significantly lower than the observed differences in dFe suggesting biological iron-ligand production and colloid formation to be responsible for maintaining elevated dFe in the high CO_2 mesocosms (Kump et al., 1000). Willers, 1000)

- (Kuma et al., 1996; Millero, 1998). Croot et al. (2001) show a ~2-fold Fe-ligand concentration increase resulting in rising dFe concentrations after 12–13 days iron induced phytoplankton bloom development. This was paralleled by a ~5-fold increase in chlorophyll-a biomass, similar to our study. Likewise, Croot et al. (2004) suggest a biological source of iron ligands in association with chlorophyll maxima in the water col-
- ²⁵ umn. Further supporting the inclination that the increased dFe concentrations during the high CO₂ treatments are biologically controlled are associations of colloidal and organically complexed Fe fractions with phytoplankton blooms in Norwegian coastal waters (Öztürk et al., 2002, 2003), which may further be controlled by bacterial production of extracellular matrixes (Heldal et al., 1996).

BGD 6,6781-6802,2009 Ocean acidification affects iron speciation in seawater E. Breitbarth et al. **Title Page** Abstract Introduction Conclusions References **Figures Tables |**◀ Back Close Full Screen / Esc **Printer-friendly Version** Interactive Discussion



The pH decrease may affect iron-ligand complex stabilities, resulting in altered photolability of Fe(III)-ligand complexes (Lewis et al., 1995; Sunda and Huntsman, 2003). The future ocean scenarios showed higher Fe(II) values compared to the lower CO₂ treatments. Values between 52 and 411 pmol L^{-1} were detected, depending on treatment and time of day (Fig. 5). Due to fast moving cloud cover though, a clear diel 5 cycle could not be determined. The solubility of Fe(II) is significantly greater than for Fe(III) and though while at these pH levels the oxidation of Fe(II) is still rapid (Santana-Casiano et al., 2006) it points towards an increase in the photolability of the iron organic complexes at lower pH (Sunda and Huntsman, 2003). Predictions of the Fe(II) half-life's over the course of the experiment reflect the expected pH dependence of Fe(II) oxidation rates (Fig. 6). The Fe(II) speciation shifts towards the Fe²⁺ ion below pH 8, while the oxidation rate is still largely controlled by Fe(OH)₂ (Santana-Casiano et al., 2006; Millero et al., 1995). Bloom dynamics cause pH shifts in the treatments that result in a much broader range of predicted Fe(II) half-life's in the high CO₂ treatments (3.4-20.6 min and 2.2–9.8 min, high and mid CO₂ respectively) vs. the low CO₂ treatment 15 (1.1–2.6 min, Fig. 6). Especially during the early phase of the bloom, the relatively slow Fe(II) oxidation rates at low pH may aid in the bioavailability of iron via allowing for a larger standing stock of Fe(II). Dropping oxygen concentrations with bloom decline and organic matter remineralization after day 12 (by ~10% saturation overall) do not fully counteract acceleration effects on Fe(II) oxidation of the increasing temperature 20 by ~0.5°C each on days 15 and 17 (Fig. 7). The product of both effects is visible in the parallel shift of the oxidation rate and resulting half-life times towards the lower end of the pH range of each treatment (Fig. 6). It should be noted that the contribution of hydrogen peroxide is not considered here, but may have implications when concentrations exceed 200 nmol L⁻¹, which is possible in coastal waters (Santana-Casiano et 25 al., 2005). Temperature controlled Fe(II) oxidation rate measurements conducted at the end of the study reveal a decrease in oxidation rate from -0.81 to -1.02 min^{-1} (log K_{ox}) and result in an increase of Fe(II) half lives from 4.4–7.3 min in the pH range from 8.21-7.95 respectively. In comparison, estimates of the sole effect of pH on the Fe(II)

BGD 6,6781-6802,2009 Ocean acidification affects iron speciation in seawater E. Breitbarth et al. **Title Page** Abstract Introduction Conclusions References **Figures Tables |**◀ Back Close Full Screen / Esc **Printer-friendly Version** Interactive Discussion

half life in the same samples ranges from 1.5–5.1 min (Table 1). Similarly, predicted oxidation rates during the measurements in the mesocosms were faster than actual measured values (Fig. 5, Table 2). While the pH effect on inorganic Fe(II) speciation and thus on Fe(II) oxidation rates is clearly evident, the additional influence of phy-

- toplankton bloom derivates on the actual half-life times are obvious. Hydrocarboxylic acids, such as glucaric acid, affect photoreduction of Fe(III) and may be released from phytoplankton (Kuma et al., 1992; Öztürk et al., 2004) while their direct effect on Fe(II) oxidation remains to be shown. The data suggest two possible mechanisms tying into Fe(II) cycling at different pH. Differences in oxidation rates and their deviation from
- the predicted rates indicate organic Fe(II)-complexation, which additionally may affect the Fe(II) half life in the low pH treatments stronger than in the high pH mesocosms. More so, biologically mediated Fe(III)-chelates supply the major pool of iron for photoreduction and this main Fe(II) production pathway (Boyd et al., 2000; Kuma et al., 1992) appears to operate more effectively at high CO₂ allowing for the elevated Fe(II)
 ¹⁵ concentrations detected in the future ocean treatments. Effects thereof on biological production and remineralization may be profound.

Our study indicates that ocean acidification may lead to enhanced Fe-bioavailability due to an increased fraction of dFe and elevated Fe(II) concentrations in coastal systems due to pH induced changes in organic complexation and Fe(II) oxidation rates.

- ²⁰ Overall this will result in increased residence times for Fe in surface seawater leading ultimately to an enhancement of iron bioavailability since equilibrium partitioning eventually restores the biolabile Fe pools that have been depleted by biological uptake. These processes may further fuel increased phytoplankton carbon acquisition and export at future atmospheric CO₂ levels (Riebesell et al., 2007). Provided that
- ²⁵ the observed CO₂ sensitivity of iron chemistry represents a general phenomenon operating also in phytoplankton blooms of oceanic areas, it could have a profound effect on productivity in the future ocean (Blain et al., 2007; Boyd et al., 2007). Our results support the notion that changes in iron speciation and the resulting potential negative feedback mechanism of phytoplankton productivity on atmospheric CO₂ need to be

BGD

6, 6781-6802, 2009

Ocean acidification affects iron speciation in seawater





considered when assessing the ecological effects of ocean acidification.

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BGD 6,6781-6802,2009 Ocean acidification affects iron speciation in seawater E. Breitbarth et al. **Title Page** Introduction Abstract Conclusions References **Tables Figures |**◀ Back Close Full Screen / Esc **Printer-friendly Version** Interactive Discussion



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BGD				
6, 6781–6802, 2009				
Ocean acidification affects iron speciation in seawater				
E. Breitbarth et al.				
Title	Page			
Abstract	Introduction			
Conclusions	References			
Tables	Figures			
14	۶I			
•				
Back	Close			
Full Screen / Esc				
Printer-friendly Version				
Interactive Discussion				



Hare, C. E., Leblanc, K., DiTullio, G. R., Kudela, R. M., Hutchins, D. A., Zhang, Y., Lee, P. A., and Riseman, S.: Consequences of increased temperature and CO₂ for phytoplankton community structure in the Bering Sea, Mar. Ecol.-Prog. Ser., 352, 9–16, 2008.

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BGD

6, 6781–6802, 2009

Ocean acidification affects iron speciation in seawater

Title Page					
Abstract	Introduction				
Conclusions	References				
Tables	Figures				
14	N				
•	•				
Back	Close				
Full Scre	Full Screen / Esc				
Printer-friendly Version					
Interactive Discussion					



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BGD				
6, 6781–6802, 2009				
Ocean acidification affects iron speciation in seawater				
E. Breitbarth et al.				
Title Page				
Abstract	Introduction			
Conclusions	References			
Tables	Figures			
I	۶I			
•	•			
Back	Close			
Full Screen / Esc				
Printer-friendly Version				
Interactive Discussion				

¹⁵

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BGD

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Ocean acidification affects iron speciation in seawater

Title Page				
Abstract	Introduction			
Conclusions	References			
Tables	Figures			
14	۶I			
•	•			
Back	Close			
Full Screen / Esc				
Printer-friendly Version				
Interactive Discussion				

BGD

6, 6781-6802, 2009

Ocean acidification affects iron speciation in seawater

E. Breitbarth et al.

Title Page				
Abstract	Introduction			
Conclusions	References			
Tables	Figures			
I.	۶I			
•	•			
Back	Close			
Full Screen / Esc				
Printer-friendly Version				
Interactive Discussion				

Table 1. FE oxidation rates and half lifes at 10°C for oxygen saturated seawater after incubation the mesocosms for 25 days compared with FE oxidation rates estimated based on Millero et al. (1987).

measured				predicted		
pH_{total}	log k _{ox} (min ⁻¹)	s. d.	t 1/2 (min)	log k _{ox} (min ⁻¹)	t <u>1</u> (min)	
8.214	-0.81	0.12	4.4	-0.35	1.5	
8.015	-0.90	0.07	5.4	-0.73	3.8	
7.950	-1.02	0.05	7.3	-0.86	5.1	

Table 2. Fe(II) oxidation rates and half lifes in the mesocosms during mid-day in synchronization with carbonate system measurements on day 11 and day 13 compared with Fe(II) oxidation rates estimated based on Millero et al. (1987). See also Fig. 5 for the respective Fe(II) concentrations measured.

measured					predic	cted	
day	рΗ	Т	O ₂	log k _{ox}	t <u>1</u>	log k _{ox}	t <u>1</u>
	total	(°C)	$(\mu mol L^{-1})$	(min ⁻¹)	(min)	(min ⁻¹)	(min)
20	8.20	11.5	287.6	-0.90	5.5	-0.26	1.3
20	8.00	11.6	297.1	-1.21	11.2	-0.64	3.0
20	7.97	11.6	311.3	-1.24	12.0	-0.69	3.4
22	8.23	10.5	291.2	-0.96	6.3	-0.31	1.4
22	8.03	10.7	297.2	-1.18	10.4	-0.67	3.3
22	7.98	10.6	305.4	-1.21	11.2	-0.78	4.2

BGD

6, 6781-6802, 2009

Ocean acidification affects iron speciation in seawater











Fig. 2. Development of pH (symbols) and Chl-a (lines) over the course of the mesocosm experiment shown as the mean values of the high (red), mid (grey), and low CO_2 (green) treatments. See Schulz et al. (2008) for the complete data set.

BGD

6, 6781-6802, 2009

Ocean acidification affects iron speciation in seawater E. Breitbarth et al. Title Page Abstract Introduction Conclusions References Tables **Figures** .∎◄ Close Back Full Screen / Esc **Printer-friendly Version** Interactive Discussion













6, 6781-6802, 2009 Ocean acidification affects iron speciation in seawater E. Breitbarth et al. **Title Page** Abstract Introduction Conclusions References Tables **Figures** .∎◄ Back Close Full Screen / Esc **Printer-friendly Version** Interactive Discussion

BGD

BGD

6, 6781-6802, 2009



Fig. 5. Fe^{2+} at high (red), mid (grey), low CO₂ (green), and in fjord water (black) and light intensity (dotted line) during days 20 (left) and 22 (right).











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Fig. 7. Oxygen saturation and temperature as a function of time during the mesocosm experiment. For oxygen saturation: high CO_2 =red, mid CO_2 =grey, low CO_2 =green. The relatively high oxygen saturation values in the mid CO_2 treatment (day 5–12) originate from enclosure # 4, which was not used for the discrete Fe(II) samples discussed in this work. The overall mean temperatures of all nine enclosures are shown as a black line (error bars are standard deviations).



