



¹ Controls of longitudinal variation in δ^{13} C-DIC in rivers: A

- 2 global meta-analysis
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4 K. A. Roach¹, M. A. Rodríguez¹, Y. Paradis², and G. Cabana¹

- 5 [1] {Université du Québec à Trois-Rivières, Département des sciences de l'environnement, C.P.
- 6 500, Trois-Rivières, Québec G9A 5H7, Canada}
- 7 [2] {Ministère des Forêts, de la Faune et des Parcs, 880, Chemin Ste-Foy, Québec City, Québec
- 8 G1S 4X4, Canada}
- 9 Correspondence to: K. A. Roach (roackat@gmail.com)
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11 Abstract

- 12 We conducted a literature survey to investigate controls and spatial and temporal patterns of
- 13 δ^{13} C-DIC and deviations between δ^{13} C-DIC and the δ^{13} C signature of DIC at isotopic
- 14 equilibrium with the atmosphere ($\Delta \delta^{13}$ C-DIC) in streams and rivers throughout the world. We
- 15 used generalized additive mixed models to relate $\Delta \delta^{13}$ C-DIC and δ^{13} C-DIC in lotic ecosystems to
- 16 ecological variables including elevation, Strahler order, and partial pressure of dissolved CO₂
- 17 (pCO₂), and to examine seasonal shifts in $\Delta \delta^{13}$ C-DIC and δ^{13} C-DIC over a range in latitude.
- 18 Elevation, Strahler order, and DIC concentrations explained a large fraction of the variation in
- 19 $\Delta \delta^{13}$ C-DIC, and these variables plus pH and pCO₂ explained much of the variation in δ^{13} C-DIC.
- 20 Seasonal fluctuations in δ^{13} C-DIC were most apparent in rivers located in temperate regions with
- seasonal snow cover. Small streams tended to have lower δ^{13} C-DIC values than large rivers.
- 22 Overall, our analysis indicates that processes that add CO₂ to the water column, including
- 23 groundwater inputs, decomposition, and respiration, should have a greater influence on δ^{13} C-DIC





- than processes that remove CO₂. Both physical (gas exchange with the atmosphere, weathering,
- 25 ice cover) and biological (respiration in regions with high C₄ grass abundance, photosynthesis by
- 26 cyanobacteria) processes appear to control δ^{13} C-DIC in streams and rivers, but the relative
- 27 importance of these processes shifts from upstream to downstream.
- 28
- 29 Key words: biogeochemistry, carbon, generalized additive mixed model, river, stable isotope
- 30

31 **1 Introduction**

Stable isotope analysis of carbon (δ^{13} C) is widely used to investigate biogeochemical 32 cycling and food web dynamics in aquatic ecosystems. Geological, atmospheric, and biological 33 sources of dissolved inorganic carbon (DIC) often have unique isotopic signatures (Boutton 34 1991); consequently, δ^{13} C is a valuable tracer of the origin of DIC. The δ^{13} C signature of DIC 35 $(\delta^{13}$ C-DIC) in the water column can be altered by the addition of DIC with a distinctive δ^{13} C 36 signature and by processes that affect the relative abundance of ${}^{13}C$: ${}^{12}C$ (fractionation). The 37 resulting spatial variation in δ^{13} C-DIC yields important information about the carbon cycle. In 38 addition to being of interest to biogeochemists, large-scale spatial gradients of δ^{13} C have been 39 used by ecologists in trophic applications. Agreement in δ^{13} C between consumer tissues and DIC 40 along a fluvial isotope gradient has been used to estimate the spatial scale of consumer feeding 41 movements (Rasmussen et al., 2009; Bertrand et al., 2011). The δ^{13} C signature of algae often 42 tracks the longitudinal gradient in δ^{13} C-DIC, whereas δ^{13} C of terrestrial-based detritus is 43 relatively constant over space (Gray et al., 2011). Recent stable isotope mixing models use the 44 upstream-downstream gradient in δ^{13} C-DIC to estimate contributions of algal versus terrestrial 45 production sources to consumer biomass by assuming that consumer δ^{13} C is a weighted mixture 46





- of algal and terrestrial signature gradients (Rasmussen, 2010). A better understanding of controls and spatial and temporal patterns of δ^{13} C-DIC in rivers would facilitate the use of δ^{13} C as a
- 49 natural tracer.

Gas exchange with the atmosphere is one mechanism that might strongly influence δ^{13} C-50 DIC in rivers. Exchange of CO_2 between the water and the atmosphere causes fractionation 51 between different carbonate species that is dependent on pH and temperature (Zhang et al., 52 1995). At isotopic equilibrium with the atmosphere, $CO_{2 (aq)}$ has a lower $\delta^{13}C$ signature relative 53 to HCO_3^{-1} and CO_3^{-2} . In streams with neutral to basic pH (dominant in HCO_3^{-1}), this process yields 54 $DIC_{(aq)}$ with a relatively high $\delta^{13}C$ signature. Other physical processes influencing $\delta^{13}C$ -DIC in 55 rivers include carbonate mineral weathering and mixing of different water bodies. Carbonate 56 weathering, the formation of HCO_3^{-1} via the dissolution of carbonate minerals, produces DIC with 57 a relatively high δ^{13} C signature, reflecting δ^{13} C of carbonate rocks (Kendall and Doctor, 2003). 58 Influx of a water body also can influence δ^{13} C-DIC in the recipient system. For example, 59 groundwater is supersaturated in CO_{2 (aq)} from soil respiration and decomposition of organic 60 matter (Wanninkhof et al., 1990), and its input lowers δ^{13} C-DIC. 61 Biological processes, including heterotrophic respiration, decomposition of organic 62 matter, and algal primary production also cause changes in δ^{13} C-DIC. Whereas DIC derived 63 from heterotrophic respiration and decomposition has a low δ^{13} C signature, algae preferentially 64 use ¹²C over ¹³C during photosynthesis, yielding DIC with a high δ^{13} C signature (McKenzie, 65 1985). Other studies have indicated that in productive rivers, high demand for CO_2 by algae 66 promotes invasion of atmospheric CO₂ into the water column (Finlay, 2003; Roussel et al., 67 2013). Thus, productive rivers tend to have low pCO₂ values due to algae uptake and δ^{13} C-DIC 68 values that are near isotopic equilibrium with the atmosphere (δ^{13} C-DIC_{equilibrium}) (Finlay, 2003). 69





Deviations between δ^{13} C-DIC and the δ^{13} C signature of DIC at isotopic equilibrium with the atmosphere ($\Delta\delta^{13}$ C-DIC) should reveal biological control of δ^{13} C-DIC in streams and rivers. Sites from rivers with high algal primary production and low pCO₂ would be expected to be closest to δ^{13} C-DIC_{equilibrium}.

A number of studies in lotic ecosystems have shown that δ^{13} C-DIC undergoes a 74 predictable shift from upstream to downstream, with headwaters having lower δ^{13} C-DIC values 75 than further downstream (Yang et al., 1996; Telmer and Veizer, 1999; Finlay, 2003). In a study 76 of the Ottawa River Basin in Canada, Telmer and Veizer (1999) found that δ^{13} C-DIC in upland 77 tributaries dominated by silicate lithology was -16‰, and δ^{13} C-DIC in the lowland main channel 78 dominated by carbonate lithology was -8‰. Similarly, in a survey of streams and rivers in 79 Northern California, USA, Finlay (2003) found that δ^{13} C-DIC and CO_{2 (aq)} increased with 80 discharge. DIC dynamics in rivers may respond to shifts along the upstream-downstream 81 gradient in environmental factors. For example, Telmer and Veizer (1999) concluded that soil 82 respiration and carbonate weathering produced much of the longitudinal variation in δ^{13} C-DIC. 83 and that CO₂ exchange between the water and atmosphere and instream primary production were 84 relatively unimportant. Finlay (2003) suggested that groundwater inputs, outgassing (evasion) of 85 excess CO₂, and microbial respiration primarily control δ^{13} C-DIC in small streams, whereas 86 algal photosynthesis and CO_2 exchange with the atmosphere become more important in large 87 rivers. An increase in carbonate weathering, a decline in groundwater inputs, CO_2 loss to the 88 atmosphere, and increasing rates of algal primary production are all associated with a decline in 89 CO_{2 (aq)} concentrations with increasing river size (Jones and Mulholland, 1998; Dawson et al., 90 2001), making the relative influence of these processes on δ^{13} C-DIC difficult to parse out. 91 In addition to having high spatial heterogeneity, rivers are temporally dynamic. 92





Floodplain rivers in particular show strong seasonality resulting from changes in river-floodplain 93 connectivity. Flood stage has a strong impact on several major processes influencing δ^{13} C-DIC. 94 including ecosystem metabolism, the fraction of groundwater relative to total discharge volume, 95 and decomposition of organic matter. Several studies have identified seasonal shifts in δ^{13} C-DIC 96 that are comparable to longitudinal variation in δ^{13} C-DIC. In lowland reaches of the Okavango 97 River in Botswana, decomposition of terrestrial organic matter contributes to δ^{13} C-DIC values 98 that are 4‰ lower during the annual period of floodplain inundation (Akoko et al., 2013). In the 99 St. Lawrence River in Canada, seasonality in algal primary production and respiration in soils 100 and groundwater causes δ^{13} C-DIC values to shift from -6.8% in spring to -1.0% in autumn 101 (Hélie et al., 2002). 102 We conducted a literature survey of $\Delta \delta^{13}$ C-DIC and δ^{13} C-DIC in lotic ecosystems 103 throughout the world, with the aim of contrasting the physical and biological processes 104 controlling $\Delta \delta^{13}$ C-DIC and δ^{13} C-DIC in streams and in large rivers. We used generalized 105 additive mixed models (GAMMs; Wood, 2006) to relate $\Delta \delta^{13}$ C-DIC and δ^{13} C-DIC in lotic 106 ecosystems to ecological variables including elevation, Strahler order, and the partial pressure of 107 dissolved CO₂ (pCO₂), and to examine seasonal shifts in $\Delta\delta^{13}$ C-DIC and δ^{13} C-DIC over a range 108 in latitude. We originally expected that if δ^{13} C-DIC was mainly under biotic control, we would 109 find a negative relationship between pCO₂ and $\Delta \delta^{13}$ C-DIC. 110

111

112 2 Methods

113 2.1 Literature survey

We collated data from the scientific literature using the search engines Google Scholar and Web of Science and using the search terms "river", "stable isotope", and "dissolved inorganic





- 116 carbon". We recorded data directly from tables or interpolated data from figures and maps using
- 117 data extraction software (GraphClick v.3.0.2) and Google Earth. For each site we recorded
- 118 geographic coordinates, δ^{13} C-DIC, temperature, pH, alkalinity, pCO₂, DIC concentrations, and
- 119 day-of-the-year of sample collection. If pCO₂ was not provided, we calculated it from pH,
- 120 alkalinity, and temperature (Stumm and Morgan, 2012). For the studies that only provided the
- 121 month or season of sampling, we used the midpoint of month or season as an estimate of day-of-
- 122 the-year. We were mainly interested in investigating controls of DIC cycling in the main channel
- 123 of rivers and thus eliminated sites from natural lakes. Our search yielded a total of 1,530 δ^{13} C-
- 124 DIC values from 26 publications (Appendix 1). Our dataset has a nested structure, with sites (n =
- 125 801) nested within rivers (n = 302), which are nested within watersheds (n = 31). Watersheds
- 126 were located in Africa (4), Asia (4), Australia (1), Europe (6), North America (7), and South
- 127 America (9) (Fig. 1).

The δ^{13} C-DIC signature at isotopic equilibrium with the atmosphere, δ^{13} C-DIC_{equilibrium}, 128 was calculated from Zhang et al. (1995) using the equations $\varepsilon CO_{2(aq)} - CO_{2(gas)} = -0.0049 \times T$ -129 1.31, εHCO_3^- - CO_{2(gas)} = -0.141 × T + 10.78, and εCO_3^- - CO_{2(gas)} = -0.052 × T + 7.22, where T is 130 temperature in °C. We assumed a δ^{13} C signature of -7.8 for atmospheric CO₂ (Levin et al. 1987). 131 The proportion of each DIC species was then weighted using the equation $\delta^{13}C_{DIC} = \sum f_i \varepsilon_i$, where 132 f is the proportion of each species (i) and ε is the permit fractionation. We used the proportion of 133 DIC species estimated from pH and temperature (Stumm and Morgan, 2012) to determine δ^{13} C-134 135 DIC_{equilibrium}.

We used a geographic information system to determine elevation (masl), Strahler order,
and channel distance from river mouth for each site. We characterized Strahler order using the
15-sec river network data from HydroSHEDS, a hydrographic database derived from elevation





- 139 data (Lehner et al., 2008). The majority of sites were at low elevation (i.e., 82% were at < 500
- 140 masl), but the highest-elevation sites (> 3500 masl) were from second or third order streams.
- 141 Rivers with Strahler order > 4 accounted for 44% of the sites in our literature survey. In the
- 142 northern hemisphere, < 1% of sites were collected from rivers in polar regions (latitude > 66.5°),
- 143 72% were from temperate regions (latitude 23.5° to 66.5°), and 7% were from tropical regions
- 144 (latitude 0° to 23.5°) (Fig. 1). In the southern hemisphere, no sites were collected from polar
- regions, 5% were from temperate regions, and 16% were from tropical regions.
- 146

147 2.2 Generalized additive mixed models

Sites were spatially nested within rivers and watersheds, requiring us to account for intragroup correlations in our statistical analyses. By adding random effects to the additive predictor, GAMMs enable modeling of overdispersed and correlated data that frequently arise in spatial analyses (Wood, 2006). Before constructing GAMMs, we examined correlations among the explanatory variables using matrix scatterplots and transformed strongly skewed variables to reduce the influence of extreme values.

We constructed separate GAMMs relating $\Delta \delta^{13}$ C-DIC and δ^{13} C-DIC to all the explanatory 154 variables. We then iteratively removed variables if 95% confidence intervals for the smooth 155 function included zero throughout the range of measured values, and re-fitted the model until the 156 zero line was not entirely contained by the confidence interval for all variables. The GAMMs for 157 $\Delta\delta^{13}$ C-DIC and δ^{13} C-DIC both had δ^{13} C-DIC as the dependent variable, but models for $\Delta\delta^{13}$ C-158 DIC additionally included δ^{13} C-DIC_{equilibrium} as an offset. We allowed for different seasonal 159 patterns in $\Delta\delta^{13}$ C-DIC and δ^{13} C-DIC in the Northern and Southern hemispheres by including the 160 interaction between day of the year and latitude in the GAMMs. We modeled seasonality and 161





162	latitudinal effects as cosinor functions of day-of-year and latitude (Barnett and Dobson, 2010)
163	comprising sine and cosine terms because measurements on a linear scale do not adequately
164	represent the circular nature of these variables (e.g., the equal spacing between days-of-year 365,
165	1, and 2 is not captured by the linear scale).
166	All GAMMs were fit with an identity link function, a normal distribution for errors, and a
167	penalized regression spline (Wood, 2006, 2011). We incorporated Strahler order as a simple
168	linear effect and river and watershed as random effects. Distance from river mouth (river km)
169	was included as a covariate to account for smaller-scale spatial correlation. We used restricted
170	maximum likelihood estimation (REML) to fit GAMMs. Potential explanatory variables for
171	$\Delta\delta^{13}$ C-DIC included elevation, Strahler order, pCO ₂ , DIC, and the interaction between latitude
172	and day of the year. Potential explanatory variables for δ^{13} C-DIC included elevation, Strahler
173	order, temperature, pH, pCO ₂ , DIC, and the interaction between latitude and day of the year. The
	final models ware:
174	mai models were.
174 175	mai models were.
174 175 176	δ^{13} C-DIC = log _e (DIC) + β_0 + β_1 Strahler order + $f(log_e(Elevation + 8)) + f(sin(\pi \times Day-of-$
174 175 176 177	$\delta^{13}\text{C-DIC} = \log_e(\text{DIC}) + \beta_0 + \beta_1 \text{ Strahler order} + f(\log_e(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}), \sin(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365}), \cos(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365})) + f(\cos(\pi \times $
 174 175 176 177 178 	$\delta^{13}\text{C-DIC} = \log_{e}(\text{DIC}) + \beta_{0} + \beta_{1} \text{ Strahler order} + f(\log_{e}(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}), \sin(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365}), \cos(\pi \times \text{Latitude/180})) + f(\log_{e}(\text{Distance from river mouth} + 1)) + f(\text{Watershed}) + f(\text{River}) + \varepsilon$
 174 175 176 177 178 179 	$\delta^{13}\text{C-DIC} = \log_{e}(\text{DIC}) + \beta_{0} + \beta_{1} \text{ Strahler order} + f(\log_{e}(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}), \sin(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365}), \cos(\pi \times \text{Latitude/180})) + f(\log_{e}(\text{Distance from river mouth} + 1)) + f(\text{Watershed}) + f(\text{River}) + \varepsilon$
 174 175 176 177 178 179 180 	hild models were. $\delta^{13}\text{C-DIC} = \log_e(\text{DIC}) + \beta_0 + \beta_1 \text{ Strahler order} + f(\log_e(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}), \sin(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365}), \cos(\pi \times \text{Latitude/180})) + f(\log_e(\text{Distance from river mouth} + 1)) + f(\text{Watershed}) + f(\text{River}) + \varepsilon$ for Δδ ¹³ C-DIC, and
 174 175 176 177 178 179 180 181 	har models were. $\delta^{13}\text{C-DIC} = \log_{e}(\text{DIC}) + \beta_{0} + \beta_{1} \text{ Strahler order} + f(\log_{e}(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}), \sin(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365}), \cos(\pi \times \text{Latitude/180})) + f(\log_{e}(\text{Distance from river mouth} + 1)) + f(\text{Watershed}) + f(\text{River}) + \varepsilon$ for $\Delta\delta^{13}\text{C-DIC}$, and
 174 175 176 177 178 179 180 181 182 	Thial models were. $\delta^{13}\text{C-DIC} = \log_{e}(\text{DIC}) + \beta_{0} + \beta_{1} \text{ Strahler order} + f(\log_{e}(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}), \sin(\pi \times \text{Latitude/180})) + f(\cos(\pi \times \text{Day-of-year/365}), \cos(\pi \times \text{Latitude/180})) + f(\log_{e}(\text{Distance from river mouth} + 1)) + f(\text{Watershed}) + f(\text{River}) + \varepsilon$ for $\Delta\delta^{13}\text{C-DIC}$, and $\delta^{13}\text{C-DIC} = \beta_{0} + \beta_{1} \text{ Strahler order} + f(\text{pH}) + f(\log_{e}(\text{pCO}_{2})) + f(\log_{e}(\text{Elevation} + 8)) + f(\sin(\pi \times \text{Day-of-year/365}))$

 $f(\log_e(\text{Distance from river mouth} + 1)) + f(\text{Watershed}) + f(\text{River}) + \varepsilon$





185	
186	for δ^{13} C-DIC.
187	
188	In the models, $log_e(DIC)$ is an offset, β are regression coefficients, $f(.)$ are smooth
189	functions, $f(Watershed)$ and $f(River)$ represent random effects, and ε is normally distributed
190	error. We used R software (R Development Core Team, 2015) for all statistical analyses and the
191	mgcv package to generate GAMMs (Wood 2011).
192	
193	3 Results
194	In the literature survey, δ^{13} C-DIC ranged from -28.1‰ to 0.7‰. δ^{13} C-DIC _{equilibrium} ranged
195	from -10.6‰ to 2.9‰, and thus did not capture the full variation in δ^{13} C-DIC (Fig. 2). Many of
196	the sites from lowland rivers, including the Indus, Negro (Argentina), Okavango, Santa Cruz,
197	and St. Lawrence rivers, had δ^{13} C-DIC values that were near δ^{13} C-DIC _{equilibrium} (Fig. 2).
198	However, other sites from lowland rivers, including the Slave and Madeira rivers, had δ^{13} C-DIC
199	that was considerably lower than δ^{13} C-DIC _{equilibrium} .
200	The GAMM of $\Delta \delta^{13}$ C-DIC modeled deviations between δ^{13} C-DIC and δ^{13} C-
201	DIC _{equilibrium} . Explanatory variables retained in this GAMM included elevation, Strahler order,
202	DIC, and the interaction between latitude and day of the year (Table 1). The final GAMM for
203	$\Delta \delta^{13}$ C-DIC included 889 data points from the literature survey; the model explained a total of
204	85% of the deviance and showed good agreement between fitted and observed values (Fig. 3).
205	The GAMM of $\Delta \delta^{13}$ C-DIC indicated that sites at the lowest and highest elevations were
206	closer to δ^{13} C-DIC _{equilibrium} (Fig. 4). $\Delta\delta^{13}$ C-DIC values were positively related to river size, as
207	measured by Strahler order, indicating that streams tended to have δ^{13} C-DIC that was far from





 δ^{13} C-DIC_{equilibrium} and large rivers tended to have δ^{13} C-DIC that was near δ^{13} C-DIC_{equilibrium} (Fig. 208 5). Sites with high DIC concentrations also tended to be near δ^{13} C-DIC_{equilibrium}, as shown by 209 high $\Delta\delta^{13}$ C-DIC values (Fig. 4). The relationship between DIC concentration and $\Delta\delta^{13}$ C-DIC 210 was not linear. There was a positive relationship between DIC and $\Delta \delta^{13}$ C-DIC at low 211 concentrations (0.05 to 0.20 mmol/L), and no relationship between DIC and $\Delta \delta^{13}$ C-DIC for 212 concentrations > 0.20 mmol/L. Seasonal variation in $\Delta \delta^{13}$ C-DIC was apparent in rivers located in 213 temperate regions from 40° to 60°N and -60° to -40°S (Fig. 6). In the northern hemisphere from 214 40° to 60°, $\Delta\delta^{13}$ C-DIC values were lowest during the winter and early spring (i.e., from January 215 to April). In the southern hemisphere from 40 to 60°, $\Delta \delta^{13}$ C-DIC values were lowest from 216 January to August and highest during the spring and early summer (i.e., from September to 217 December). Tropical rivers tended to have lower $\Delta \delta^{13}$ C-DIC values than temperate rivers. The 218 highest $\Delta \delta^{13}$ C-DIC fitted values were for rivers located from 40° to 60°N during September to 219 December. 220 Explanatory variables retained in the model for δ^{13} C-DIC included elevation, Strahler 221

Explanatory variables retained in the model for δ^{13} C-DIC included elevation, Strahler order, pCO₂, DIC, pH, and the interaction between latitude and day of the year (Table 2). The final GAMM for δ^{13} C-DIC included 890 data points from the literature survey; the model explained a total of 91% of the deviance and showed good agreement between fitted and observed values (Fig. 3).

The GAMM of δ^{13} C-DIC showed that sites at the lowest and highest elevations tended to have high δ^{13} C-DIC values (Fig. 7). δ^{13} C-DIC values were positively related to river size, as measured by Strahler order (Fig. 5). δ^{13} C-DIC was related negatively to pCO₂ and positively to DIC (Fig. 7). Again, the relationship between DIC concentrations and δ^{13} C-DIC was nonlinear. There was a positive relationship between DIC and δ^{13} C-DIC at low concentrations (0.05 to 0.20





- 231 mmol/L), and a weakly positive relationship between DIC and δ^{13} C-DIC at higher concentrations
- 232 (0.20 to 7.5 mmol/L). The relationship between pH and δ^{13} C-DIC also was nonlinear. Sites with
- 233 lowest and highest pH tended to have low δ^{13} C-DIC values (Fig. 7). Seasonal patterns in δ^{13} C-
- 234 DIC and $\Delta \delta^{13}$ C-DIC were very similar (Fig. 6).
- 235

236 4 Discussion

230	
237	Our main objective was to investigate controls and spatial and temporal patterns of δ^{13} C-
238	DIC in rivers throughout the world. pCO2 explained a large fraction of the deviance in the
239	GAMM of δ^{13} C-DIC, but was not retained in the GAMM of $\Delta\delta^{13}$ C-DIC, suggesting that algal
240	primary production has limited control over δ^{13} C-DIC. However, as we will discuss below, our
241	results provide evidence that other biological processes, including photosynthesis by
242	cyanobacteria and respiration in regions with high C4 grass abundance are important
243	determinants of δ^{13} C-DIC in rivers (Fig. 8). Our results revealed changes in the dominant
244	processes influencing δ^{13} C-DIC from upstream to downstream. Furthermore, we found that δ^{13} C-
245	DIC changed seasonally in rivers in temperate regions with seasonal snow cover.
246	$\Delta \delta^{13}$ C-DIC values were low (δ^{13} C-DIC was near δ^{13} C-DIC _{equilibrium}) in high-elevation
247	streams and large rivers, indicating that exchange of CO ₂ between surface water and the
248	atmosphere occurs via invasion or evasion in these systems. Most studies show that lotic
249	ecosystems are supersaturated in pCO ₂ (Richey et al., 2002; Jones et al., 2003; Cole et al., 2007)
250	and are sources, not sinks, of CO ₂ to the atmosphere (Mulholland et al., 2001; Battin et al.,
251	2008). Thus, in most low-order streams, high $CO_{2 (aq)}$ concentrations originating from inputs of
252	groundwater and terrestrial soil water from sediments (e.g., Jones and Mulholland, 1998) result
253	in high $\Delta \delta^{13}$ C-DIC values, indicating that δ^{13} C-DIC is far from δ^{13} C-DIC _{equilibrium} . In high-





elevation streams, low $\Delta \delta^{13}$ C-DIC values have been attributed to the high gradient and low 254 surface:volume ratio, which increases water turbulence and promotes CO₂ outgassing (Rebsdorf 255 et al., 1991; Dawson et al., 2004). For example, a recent estimate of gas transfer velocities in 256 streams and rivers throughout the United States found that CO₂ outgassing is highest in 257 headwater streams originating in areas with steep topography (Butman and Raymond, 2011). In 258 small streams, groundwater consists of a large fraction of stream discharge, but its volume 259 relative to the total volume of discharge decreases downstream (e.g., Devol et al., 1987; Johnson 260 et al., 2006). $\Delta \delta^{13}$ C-DIC values were likely low in large rivers because groundwater has less of 261 262 an influence on water chemistry and because surface water has been exposed to the atmosphere for a relatively long period of time, promoting evasion of $CO_{2 \text{ (aq)}}$. $\Delta \delta^{13}$ C-DIC values are 263 particularly low in lacustrine rivers because the prolonged residence time of water in large lakes 264 allows CO_{2 (aq)} to equilibrate with the atmosphere (Yang et al., 1996). $\Delta \delta^{13}$ C-DIC values also 265 were low in rivers with high DIC concentrations (> 0.2 mmol/L). In our dataset, DIC 266 concentrations were strongly correlated with alkalinity (r = 0.98) and CO_{2 (aq)} concentrations did 267 not have a strong effect on total DIC. $\Delta \delta^{13}$ C-DIC values were probably low in rivers with high 268 DIC concentrations because the carbon was derived from carbonate dissolution, which has a high 269 δ^{13} C-DIC signature (Kendall and Doctor, 2003). 270 pCO₂ did not explain a large fraction of the total variation in $\Delta \delta^{13}$ C-DIC, providing 271

evidence that algal primary production does not control exchange of CO₂ between surface water

and the atmosphere in most rivers. Calculation of pCO₂ from pH, alkalinity, and temperature

- results in values that are overestimated in acidic waters rich in dissolved organic carbon (Abril et
- al., 2015). In our meta-analysis, pCO₂ was calculated from these physicochemical variables in
- 276 90% of the studies. Direct measurements of pCO₂ in streams and rivers would allow for a more





298	Similar to the spatial patterns in $\Delta \delta^{13}$ C-DIC values, high-elevation streams and large
297	CO ₂ .
296	water column should generally have a greater influence on $\delta^{13}\mbox{C-DIC}$ than processes that remove
295	$\delta^{13}C$ values relative to equilibrium with the atmosphere. Therefore, processes that add CO_2 to the
294	Aufdenkampe et al., 2011). Most lotic systems are supersaturated in CO ₂ and typically have low
293	decomposition of organic matter becoming more important in large rivers (Johnson et al., 2008;
292	is believed to be groundwater and terrestrial soil water, with in situ respiration and
291	include CO ₂ evasion and algal primary production. In small streams, the major source of CO _{2 (aq)}
290	column should also influence $\delta^{13}\text{C-DIC}.$ Major processes that remove CO_2 in lotic systems
289	on the biogeochemistry of δ^{13} C-DIC; thus, processes that remove or add CO ₂ to the water
288	2011), similar to our results for latitudinal variation. The cycling of CO_2 has a strong influence
287	rivers typically have higher concentrations of $CO_{2 (aq)}$ than temperate rivers (Aufdenkampe et al.,
286	relationship between δ^{13} C-DIC and Strahler order revealed by our analysis. Furthermore, tropical
285	(Raymond et al., 1997; Teodoru et al., 2009; Butman and Raymond, 2011), similar to the
284	have been observed elsewhere. For example, pCO_2 in rivers tends to decline downstream
283	δ^{13} C-DIC indicated by our analysis correspond to patterns in pCO ₂ in streams and rivers that
282	Our results showed that pCO_2 is negatively related to $\delta^{13}C$ -DIC. The spatial patterns in
281	DIC might have been caused by preferential assimilation of 12 C over 13 C by algae.
280	the water and atmosphere. Some of the unexplained variation in the GAMM analysis of $\delta^{13}\text{C-}$
279	may have been sufficient to increase δ^{13} C-DIC values without increasing mixing of DIC between
278	production apparently did not cause invasion of atmospheric CO ₂ , low rates of photosynthesis
277	accurate assessment of the relationship between pCO ₂ and $\Delta \delta^{13}$ C-DIC. Although algal primary

298 Similar to the spatial patterns in $\Delta\delta^{13}$ C-DIC values, high-elevation streams and large 299 rivers had DIC with a high δ^{13} C signature. The congruent spatial trends between $\Delta\delta^{13}$ C-DIC and





300	δ^{13} C-DIC suggest that gas exchange influenced δ^{13} C-DIC in these systems. Our results also show
301	that DIC had a high δ^{13} C signature in rivers with high DIC concentrations. Large rivers likely
302	have high δ^{13} C-DIC values because their waters have higher concentrations of carbonate
303	minerals, and this higher buffering capacity inhibits decrease in δ^{13} C-DIC. Within a river
304	network, concentrations of dissolved ions are frequently heterogeneous among headwater
305	reaches, but tend to average out and increase downstream (Livingstone, 1963). Our results
306	indicated that sites with low and high pH had DIC with a low $\delta^{13}C$ signature. Many of the sites
307	with the lowest surface water pH (3.8 to 6.5) also tended to have low DIC concentrations (range
308	= < 0.1 to 2.5 mmol/L, average = 0.6 mmol/L). δ^{13} C-DIC values may have been low in rivers
309	with acidic surface water because these were blackwater rivers draining forested watersheds
310	dominated by igneous rock. In addition to having low pH, blackwater rivers have high
311	concentrations of dissolved organic matter that increase rates of microbial respiration (Meyer
312	1990), lowering δ^{13} C-DIC values. Sites with high pH and low δ^{13} C-DIC values also had low
313	pCO_2 (sites with $pH > 9$ had a maximum pCO_2 of 115 ppmv). When pH is high and pCO_2 is low,
314	intense photosynthesis by cyanobacteria can increase the rate of CO ₂ invasion from the
315	atmosphere, producing fractionation that results in low δ^{13} C-DIC values (Herczog and Fairbanks,
316	1987). Following the depletion of free CO ₂ that occurs in highly alkaline waters, cyanobacteria
317	can float on the water surface and use atmospheric CO ₂ during carbon fixation (Paerl and Ustach,
318	1982). Lakes also exhibit a positive trend between pH and δ^{13} C-DIC until pH values of
319	approximately 8-9, at which point δ^{13} C-DIC decreases (Bade et al., 2004).
320	Rivers in the northern hemisphere between 40° and 60° exhibited seasonal cycles in
321	$\Delta\delta^{13}$ C-DIC and δ^{13} C-DIC. In these rivers, $\Delta\delta^{13}$ C-DIC and δ^{13} C-DIC values were lowest from
322	January to April, corresponding with the formation of ice cover. Studies of lakes have shown that





323	duration of ice cover is an important control of seasonal patterns in pCO ₂ and δ^{13} C-DIC. Ice
324	insulates the lake from mixing by wind and gas exchange, which causes pCO ₂ to increase in the
325	water column and lowers δ^{13} C-DIC values (Striegl et al., 2001; Karlsson et al., 2008). The same
326	phenomenon has been observed in terrestrial ecosystems, with snow cover increasing
327	accumulation of CO ₂ in the soil and lowering δ^{13} C-DIC values in the winter (Aravena et al.,
328	1992). Ice cover also has been documented to increase pCO ₂ in the water column of rivers (e.g.,
329	Raymond et al., 1997), and should be responsible for the seasonal shifts in $\Delta\delta^{13}$ C-DIC and δ^{13} C-
330	DIC values in rivers at high latitudes in the northern hemisphere. Sample size was low in
331	temperate regions of the southern hemisphere. A greater number of sites sampled from these
332	rivers may result in a seasonal trend that is similar to the pattern observed in the northern
333	hemisphere. Our results also revealed that DIC has a lower $\delta^{13}C$ signature in many tropical rivers
334	than in temperate rivers. An analysis of ¹³ C and ¹⁴ C of DIC, dissolved organic carbon, and
335	multiple particulate organic carbon fractions in Amazonian rivers concluded that high pCO ₂ was
336	sustained by in situ respiration of terrestrial C4 grasses (Mayorga et al., 2005). Terrestrial C4
337	grasses decompose more rapidly than terrestrial C ₃ plants (Wynn and Bird, 2007), and the high
338	relative abundance of this group of macrophytes in the tropics promotes microbial respiration,
339	lowering δ^{13} C-DIC values even during low-water periods.
340	Our analysis indicated a contrast between mechanisms influencing δ^{13} C-DIC in streams
341	and in floodplain rivers (Fig. 8). Physical and biological processes control δ^{13} C-DIC throughout
342	the upstream-downstream gradient. However, whereas gas exchange with the atmosphere and
343	groundwater inputs appear to be dominant controls of δ^{13} C-DIC in small streams, carbonate
344	weathering and photosynthesis by cyanobacteria are particularly important in higher-order rivers.
345	To fully evaluate physical and biological controls of δ^{13} C-DIC in streams and rivers,





346	measurements of rates of algal primary production and respiration, gas exchange between the air
347	and water, and groundwater inputs would be needed. Our results also have implications for
348	future studies using $\delta^{13}C$ to investigate food web structure in rivers. Studies in temperate regions
349	with seasonal snow cover should take temporal shifts in $\delta^{13}\mbox{C-DIC}$ into account. Failure to
350	consider temporal changes in δ^{13} C-DIC can bias inferences about consumer-resource dynamics
351	(Woodland et al., 2012). Finally, food web models that incorporate patterns of spatial variation in
352	δ^{13} C-DIC (Rasmussen et al., 2009; Rasmussen, 2010) are dependent upon an underlying isotope
353	gradient. Thus, the applicability of these models in many aquatic ecosystems has been
354	questioned (Layman et al., 2012). Our GAMM analysis showed a positive relationship between
355	Strahler order and δ^{13} C-DIC, indicating that headwaters have low δ^{13} C-DIC values relative to
356	downstream reaches in most river systems. The rivers with the most pronounced upstream-
357	downstream gradients in $\delta^{13}\text{C-DIC}$ will be those in which headwaters have high levels of $p\text{CO}_2$
358	(Fig. 8). Incorporation of spatial variation in δ^{13} C-DIC in future lotic food web studies will thus
359	provide insights about the ecology of aquatic organisms, including movements and assimilation
360	of organic matter from alternative production sources.
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Watershed	Location	Citation
Amazon	South America	Mayorga et al., 2005
Brazos	Texas, USA	Zeng et al., 2011
Chico	Argentina	Brunet et al., 2005
Chubut	Argentina	Brunet et al., 2005
Colorado	Argentina	Brunet et al., 2005
Congo	Africa	Bouillon et al., 2012; Bouillon
		et al., 2014
Coyle	Argentina	Brunet et al., 2005
Danube	Europe	Kanduč et al., 2007
Deseado	Argentina	Brunet et al., 2005
Ems	Germany	Stögbauer et al., 2008
Fraser	British Columbia, Canada	Cameron et al., 1995; Spence
		and Telmer, 2005
Gallegos	Argentina	Brunet et al., 2005
Ganges-Brahmaputra	Asia	Galy and France-Lanord, 1999
Han	Korea	Lee et al., 2007
Indus	Asia	Karim and Veizer, 2000
Khwai	Africa	Akoko et al., 2013
Lagan	Ireland	Barth et al., 1998
Mackenzie	Canada	Hitchon and Krouse, 1972;
		Reeder et al., 1972
Murray	Australia	Cartwright, 2010
Nass	British Columbia, Canada	Spence and Telmer, 2005
Negro	Argentina	Brunet et al., 2005
Okavango	Africa	Akoko et al., 2013
Pearl	Asia	Zhang et al., 2009
Rhine	Europe	Flintrop et al., 1996; Stögbauer
		et al., 2008
Rhône	Europe	Aucour et al., 1999
Santa Cruz	Argentina	Brunet et al., 2005
Skeena	British Columbia, Canada	Spence and Telmer, 2005
Squamish	British Columbia, Canada	Spence and Telmer, 2005
St. Lawrence	Canada and USA	Yang et al., 1996; Telmer and
		Veizer, 1999; Hélie et al., 2002
Tana	Kenya	Bouillon et al., 2009; Tamooh
		et al., 2013
Vistula	Poland	Wachniew, 2006

Appendix A: Sources of data for the literature review 369

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380 References

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- 568 Table 1. Quantitative output of the generalized additive mixed models (GAMM) of $\Delta \delta^{13}$ C-DIC,
- 569 including information on units and model effects. Estimated degrees of freedom = edf. The
- 570 GAMM explained a total of 85.2% of the deviance.

Model term	Unit	Effect	Estimate	Std. Error	edf	Standard deviation	p-value
Intercept			-10.07	0.63			< 0.001
Strahler order		Linear	0.27	0.09			< 0.01
Dissolved inorganic carbon (DIC)	mmol/L	Smooth			6.4		<0.001
Elevation	masl	Smooth			5.0		< 0.001
Day-of-year × Latitude		Smooth			19.9		< 0.001
Day-of-year × Latitude		Smooth			14.1		< 0.001
Distance from river mouth	km	Smooth			7.1		< 0.001
Watershed		Random			11.7	1.7	< 0.001
River		Random			97.1	1.5	< 0.001





- 581 Table 2. Quantitative output of the generalized additive mixed models (GAMM) of δ^{13} C-DIC,
- including information on units and model effects. Estimated degrees of freedom = edf. The
- 583 GAMM explained a total of 90.6% of the deviance.

Model term	Unit	Effect	Estimate	Std. Error	edf	Standard deviation	p-value
Intercept			-10.76	0.58			< 0.001
Strahler order		Linear	0.27	0.09			< 0.01
pН		Smooth			5.9		< 0.001
Dissolved	mmol/L	Smooth			7.1		< 0.001
inorganic carbon (DIC)							
Partial pressure	ppmv	Smooth			3.8		< 0.001
of dissolved							
CO_2 (p CO_2)							
Elevation	masl	Smooth			5.1		< 0.001
Day-of-year ×		Smooth			21.0		< 0.001
Latitude							
Day-of-year \times		Smooth			15.2		< 0.001
Latitude							
Distance from river mouth	km	Smooth			6.4		< 0.001
Watershed		Random			9.9	1.4	< 0.001
River		Random			104.5	1.6	< 0.001







- 595 Figure 1. Map of the study sites included in the literature survey. Major world rivers are shown
- 596 in light grey.







Figure 2. The relationship between δ^{13} C-DIC_{equilibrium} (the δ^{13} C signature of DIC at isotopic equilibrium with the atmosphere) and δ^{13} C-DIC in rivers. Grey symbols are from rivers with Strahler order = 8 and black symbols are from rivers with Strahler order < 8. The line indicates the 1:1 relationship.







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639 Figure 3. Plots of fitted versus observed values for $\Delta\delta^{13}$ C-DIC (A) and δ^{13} C-DIC (B).







Figure 4. Smooth function (curve) derived from the generalized additive mixed model (GAMM) fit for relationships between $\Delta\delta^{13}$ C-DIC and elevation (meters above sea level, A) and dissolved inorganic carbon (DIC) concentrations (mmol/L, B). Dashed curves represent 95% confidence intervals for the smooth. Partial residuals (points) are also shown. Higher $\Delta\delta^{13}$ C-DIC values indicate that δ^{13} C-DIC is closer to isotopic equilibrium with the atmosphere. Note that the smooth function y-axis is centered to zero mean.

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Figure 5. Relationships between Strahler order and $\Delta\delta^{13}$ C-DIC (A) and δ^{13} C-DIC (B). The regression lines determined by the coefficients of the generalized additive mixed models (GAMMs) are shown.







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Figure 6. Contour plots showing the seasonal variation in fitted $\Delta \delta^{13}$ C-DIC (top) and δ^{13} C-DIC (bottom) at different latitudes. Lighter colors indicate higher $\Delta \delta^{13}$ C-DIC values (values closer to atmospheric equilibrium) for the top plot and higher δ^{13} C-DIC values for the bottom plot.







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Figure 7. Smooth function (curve) derived from the generalized additive mixed model (GAMM) fit for relationships between rescaled δ^{13} C-DIC and elevation (masl, A), pCO₂ (ppmv, B), dissolved inorganic carbon (DIC) concentrations (mmol/L, C), and pH (D). Dashed curves represent 95% confidence intervals for the smooth. Partial residuals (points) are also shown. Note that the smooth function y-axis is centered to zero mean.

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Figure 8. Schematic diagram showing dominant physical and biological controls of δ^{13} C-DIC in rivers throughout the world. The grey figures represent the longitudinal river gradient. The black outlines indicate the relative importance of various processes influencing δ^{13} C-DIC along the upstream-downstream gradient. Patterns in pCO₂ along the river gradient also are shown. The dotted outline indicates that floodplain rivers are near isotopic equilibrium with the atmosphere, but this may be because carbonate weathering also is associated with high δ^{13} C-DIC values.