

Thank you very much to the reviewers and editors for their time on this manuscript. The authors have reviewed the changes suggested in the technical corrections, and have highlighted the changes here. A tracked changes version of the manuscript is included.

Reviewer 1

Line numbers corresponding to the document still including the changes.

Line 93: ...annual flux....

Changed

Line 396: 't' needs to be in italics.

Changed

Line 397: ...as well as a response to the mesocosm wall cleaning which took place on t16...this part of the sentence does not fit with the beginning of the sentence. Needs to be reworded.

The section about wall cleaning has been put into a second sentence. Now reads 'These higher DMS concentrations were likely connected to a peak in dissolved organic carbon (DOC) on t15, as well as increasing bacterial abundance during Phase II (Hornick *et al.*, 2016). It is also likely that DMS concentrations increased as a response to the mesocosm wall cleaning which took place on t16.'

Line 441: ...vary...Typo

changed

Line 544: ...coastal environment. However, ...

changed

Line 599: ...t17 or t17 as before and just one line below?

CO2 did increase between t16 and t17, but the largest increase was seen after t17. However, as the first increase was seen after t16, and for ease of discussion, this has been changed to t16, as it corresponds with the change in DMS.

Line 651: ...hard... I don't think that 'hard' is the appropriate word to use here.

Suggestion: ...therefore these upwelling events cannot be considered as natural high CO2 analogues. But the beginning of this sentence needs to be reworded also.

The section has been reworded to 'The upwelling event occurring mid-way through our experiment allowed comparison of the mesocosm findings with a natural analogue of the system, as well as showing the extent to which the system perturbation can occur (up to 800  $\mu\text{atm}$ ). This event was a fortuitous occurrence during this mesocosm experiment, but as the scale and timing of these upwelling events is difficult to determine, and therefore these upwelling events are extremely challenging to study as natural high  $\text{CO}_2$  analogues.'

Reviewer 2

The data in this manuscript are important information.

The results show a lack of response of gas concentrations to the experimental design and no linkage to the external conditions due to the outside undergoing its own “experiment”, i.e., upwelling. No rates were measured, concentrations (mostly) remained the same, and the need to have rates of production, consumption, (photo/chem)-degradation or even “net” rates to include in prognostic and predictive models has been emphasized.

I recommend publishing pending changes. For me personally it would be important to have a clearer number agreement for the mean CO<sub>2</sub> values.

Could the reviewer clarify what is meant by this, as I do not understand?

Also I think the firm and clearly stated replies to the editor comments under Point 1 and 2 would strengthen the discussion, as they nicely address the criticism that could still emerge.

The section on DMSP has been edited to include the following (TC version)

L 235 This method had been used during a previous mesocosm experiment (Bergen, Norway) and the results correlated well with those measured immediately on a similar GC-FPD system (Webb *et al.* 2015).

L 255 The DMSP acidification method is used worldwide as a simple and effective method of DMSP storage. The findings here, alongside those of del Valle *et al.* (2011) question the applicability of this method in other marine environments, and suggests significant testing prior to reliance on this method as a sole means of DMSP storage.

The section on the Baltic Sea has been amended to include comments from Section 2 of the replies to editors. Specifically (TC version):

L 578 The changes in biological parameters and trace gas concentrations are therefore discussed here separately from the concentrations measured in the mesocosms.

L 581 Given the separation of the waters within the mesocosms, and the movement of water masses within the Baltic Sea, it is expected that phytoplankton population structure could be significantly different inside the mesocosms compared to the external waters.

It could be worth to have a (short, taken from the comments) discussion section on methodological caveats/challenges and reasoning for the way the data is presented.

In terms of this, I believe this comment has been covered in the individual sections mentioned, and to include this recommended section would cause repetition within the manuscript. It is fair to say that the layout of the sections has caused a lot of discussion during the preparation of this manuscript.

L 607 In addition, the community demands of sulphur are likely to be very different in the Baltic Sea compared to the mesocosms, due to differences in community composition and sulphur availability, and therefore direct comparisons with mesocosm concentrations are inappropriate.

L634 Macroalgal production in the Baltic Sea is likely the predominant iodocarbon source, compared to the mesocosms where macroalgae are excluded.

1 **Effect of ocean acidification and elevated  $f\text{CO}_2$  on trace gas**  
2 **production by a Baltic Sea summer phytoplankton community**

3 **A.L. Webb<sup>1,2</sup>, E. Leedham-Elvidge<sup>1</sup>, C. Hughes<sup>3</sup>, F.E. Hopkins<sup>4</sup>, G. Malin<sup>1</sup>, L.T. Bach<sup>5</sup>,**  
4 **K. Schulz<sup>6</sup>, K. Crawford<sup>7</sup>, C.P.D. Brussaard<sup>7,8</sup>, A. Stühr<sup>5</sup>, U. Riebesell<sup>5</sup>, and P.S. Liss<sup>1</sup>.**

5 [1] {Centre for Ocean and Atmospheric Sciences, School of Environmental Science, University of  
6 East Anglia, Norwich, UK, NR4 7TJ}

7 [2] {Groningen Institute for Evolutionary Life Sciences, University of Groningen, 9700 CC  
8 Groningen, The Netherlands}

9 [3] {Environmental Department, University of York, York, UK, YO10 5DD}

10 [4] {Plymouth Marine Laboratory, Plymouth, UK, PL1 3DH}

11 [5] {GEOMAR Helmholtz Centre for Ocean Research Kiel, Düsternbrooker Weg 20, 24148 Kiel,  
12 Germany.}

13 [6] {Centre for Coastal Biogeochemistry, School of Environment, Science and Engineering,  
14 Southern Cross University, Lismore, NSW 2480, Australia.}

15 [7] {Department of Biological Oceanography, NIOZ – Royal Netherlands Institute for Sea  
16 Research, PO Box 59, 1790 AB Den Burg, Texel, The Netherlands}

17 [8] {Aquatic Microbiology, Institute for Biodiversity and Ecosystem Dynamics, University of  
18 Amsterdam, P.O. Box 94248, 1090 GE, Amsterdam, The Netherlands}

19  
20 Correspondence to: Alison Webb ([a.l.webb@rug.nl](mailto:a.l.webb@rug.nl))

21  
22 **Abstract**

23 **The Baltic Sea is a unique environment as the largest body of brackish water in the world.**  
24 **Acidification of the surface oceans due to absorption of anthropogenic  $\text{CO}_2$  emissions is an**  
25 **additional stressor facing the pelagic community of the already challenging Baltic Sea. To**  
26 **investigate its impact on trace gas biogeochemistry, a large-scale mesocosm experiment was**  
27 **performed off Tvärminne Research Station, Finland in summer 2012. During the second half**  
28 **of the experiment, dimethylsulphide (DMS) concentrations in the highest  $f\text{CO}_2$  mesocosms**

29 (1075 - 1333  $\mu\text{atm}$ ) were 34% lower than at ambient  $\text{CO}_2$  (350  $\mu\text{atm}$ ). However, the net  
30 production (as measured by concentration change) of seven halocarbons analysed was not  
31 significantly affected by even the highest  $\text{CO}_2$  levels after 5 weeks exposure. Methyl iodide  
32 ( $\text{CH}_3\text{I}$ ) and diiodomethane ( $\text{CH}_2\text{I}_2$ ) showed 15% and 57% increases in mean mesocosm  
33 concentration ( $3.8 \pm 0.6 \text{ pmol L}^{-1}$  increasing to  $4.3 \pm 0.4 \text{ pmol L}^{-1}$  and  $87.4 \pm 14.9 \text{ pmol L}^{-1}$   
34 increasing to  $134.4 \pm 24.1 \text{ pmol L}^{-1}$  respectively) during Phase II of the experiment, which  
35 were unrelated to  $\text{CO}_2$  and corresponded to 30% lower Chl-*a* concentrations compared to  
36 Phase I. No other iodocarbons increased or showed a peak, with mean chloriodomethane  
37 ( $\text{CH}_2\text{ClI}$ ) concentrations measured at  $5.3 (\pm 0.9) \text{ pmol L}^{-1}$  and iodoethane ( $\text{C}_2\text{H}_5\text{I}$ ) at  $0.5 (\pm 0.1)$   
38  $\text{pmol L}^{-1}$ . Of the concentrations of bromoform ( $\text{CHBr}_3$ ; mean  $88.1 \pm 13.2 \text{ pmol L}^{-1}$ ),  
39 dibromomethane ( $\text{CH}_2\text{Br}_2$ ; mean  $5.3 \pm 0.8 \text{ pmol L}^{-1}$ ) and dibromochloromethane ( $\text{CHBr}_2\text{Cl}$ ,  
40 mean  $3.0 \pm 0.5 \text{ pmol L}^{-1}$ ), only  $\text{CH}_2\text{Br}_2$  showed a decrease of 17% between Phases I and II,  
41 with  $\text{CHBr}_3$  and  $\text{CHBr}_2\text{Cl}$  showing similar mean concentrations in both Phases. Outside the  
42 mesocosms, an upwelling event was responsible for bringing colder, high  $\text{CO}_2$ , low pH water  
43 to the surface starting on day *t*16 of the experiment; this variable  $\text{CO}_2$  system with frequent  
44 upwelling events implies the community of the Baltic Sea is acclimated to regular significant  
45 declines in pH caused by up to 800  $\mu\text{atm } f\text{CO}_2$ . After this upwelling, DMS concentrations  
46 declined, but halocarbon concentrations remained similar or increased compared to  
47 measurements prior to the change in conditions. Based on our findings, with future  
48 acidification of Baltic Sea waters, biogenic halocarbon emissions are likely to remain at  
49 similar values to today, however emissions of biogenic sulphur could significantly decrease  
50 from this region.

51

## 52 1 Introduction

53 Anthropogenic activity has increased the fugacity of atmospheric carbon dioxide ( $f\text{CO}_2$ ) from 280  
54  $\mu\text{atm}$  (pre-Industrial Revolution) to over 400  $\mu\text{atm}$  today (Hartmann *et al.*, 2013). The IPCC AR5  
55 long-term projections for atmospheric  $p\text{CO}_2$  and associated changes to the climate have been  
56 established for a variety of scenarios of anthropogenic activity until the year 2300. As the largest  
57 global sink for atmospheric  $\text{CO}_2$ , the global ocean has absorbed an estimated 30% of excess  $\text{CO}_2$   
58 produced (Canadell *et al.*, 2007). With atmospheric  $p\text{CO}_2$  projected to possibly exceed 2000  $\mu\text{atm}$   
59 by the year 2300 (Collins *et al.*, 2013; Cubasch *et al.*, 2013), the ocean will take up increasing  
60 amounts of  $\text{CO}_2$ , with a potential lowering of surface ocean pH by over 0.8 units (Raven *et al.*,  
61 2005). The overall effect of acidification on the biogeochemistry of surface ocean ecosystems is

62 unknown and currently unquantifiable, with a wide range of potential positive and negative impacts  
63 (Doney *et al.*, 2009; Hofmann *et al.*, 2010; Ross *et al.*, 2011).

64 A number of volatile organic compounds are produced by marine phytoplankton (Liss *et al.*, 2014),  
65 including the climatically important trace gas dimethylsulphide (DMS, C<sub>2</sub>H<sub>6</sub>S) and a number of  
66 halogen-containing organic compounds (halocarbons) including methyl iodide (CH<sub>3</sub>I) and  
67 bromoform (CHBr<sub>3</sub>). These trace gases are a source of sulphate particles and halide radicals when  
68 oxidised in the atmosphere, and have important roles as ozone catalysts in the troposphere and  
69 stratosphere (O'Dowd *et al.*, 2002; Solomon *et al.*, 1994) and as cloud condensation nuclei (CCNs;  
70 Charlson *et al.*, 1987).

71 DMS is found globally in surface waters originating from the algal-produced precursor  
72 dimethylsulphoniopropionate (DMSP, C<sub>5</sub>H<sub>10</sub>O<sub>2</sub>S). Both DMS and DMSP ~~are~~ provide the basis for  
73 major routes of sulphur and carbon flux through the marine microbial food web, and can provide up  
74 to 100% of the bacterial (Simó *et al.*, 2009) and phytoplanktonic (Vila-Costa *et al.*, 2006a) sulphur  
75 demand (Simó *et al.*, 2009; Vila-Costa *et al.*, 2006a). DMS is also a volatile compound which  
76 readily passes through the marine boundary layer to the troposphere, where oxidation results in a  
77 number of sulphur-containing particles important for atmospheric climate feedbacks (Charlson *et al.*,  
78 1987; Quinn and Bates, 2011); for this reason, any change in the production of DMS may have  
79 significant implications for climate regulation. Several previous acidification experiments have  
80 shown differing responses of both compounds (e.g. Avgoustidi *et al.*, 2012; Hopkins *et al.*, 2010;  
81 Webb *et al.*, 2015), while others have shown delayed or more rapid responses as a direct effect of  
82 CO<sub>2</sub> (e.g. Archer *et al.*, 2013; Vogt *et al.*, 2008). Further, some laboratory incubations of coastal  
83 microbial communities showed increased DMS production with increased *f*CO<sub>2</sub> (Hopkins and  
84 Archer, 2014), but lower DMSP production. The combined picture arising from existing studies is  
85 that the response of communities to *f*CO<sub>2</sub> perturbation is not predictable and requires further study.  
86 Previous studies measuring DMS in the Baltic Sea measured concentrations up to 100 nmol L<sup>-1</sup>  
87 during the summer bloom, making the Baltic Sea a significant source of DMS (Orlikowska and  
88 Schulz-Bull, 2009).

89 In surface waters, halocarbons such as methyl iodide (CH<sub>3</sub>I), chloriodomethane (CH<sub>2</sub>ClI) and  
90 bromoform (CHBr<sub>3</sub>) are produced by biological and photochemical processes: many marine  
91 microbes (for example cyanobacteria; Hughes *et al.*, 2011, diatoms; Manley and De La Cuesta,  
92 1997 and haptophytes; Scarratt and Moore, 1998) and macroalgae (e.g. brown-algal *Fucus* species;  
93 Chance *et al.*, 2009 and red algae; Leedham *et al.*, 2013) utilise halides from seawater and emit a  
94 range of organic and inorganic halogenated compounds. This production can lead to significant

95 annual flux to the marine boundary layer in the order of 10 Tg iodine-containing compounds  
96 ('iodocarbons'; O'Dowd *et al.*, 2002) and 1 Tg bromine-containing compounds ('bromocarbons';  
97 Goodwin *et al.*, 1997) into the atmosphere. The effect of acidification on halocarbon concentrations  
98 has received limited attention, but two acidification experiments measured lower concentrations of  
99 several iodocarbons while bromocarbons were unaffected by  $f\text{CO}_2$  up to 3000  $\mu\text{atm}$  (Hopkins *et al.*,  
100 2010; Webb, 2015), whereas an additional mesocosm study did not elicit significant differences  
101 from any compound up to 1400  $\mu\text{atm}$   $f\text{CO}_2$  (Hopkins *et al.*, 2013).

102 Measurements of the trace gases within the Baltic Sea are limited, with no prior study of DMSP  
103 concentrations in the region. The Baltic Sea is the largest body of brackish water in the world, and  
104 salinity ranges from 1 to 15. Furthermore, seasonal temperature variations of over 20 °C are  
105 common. A permanent halocline at 50-80 m separates  $\text{CO}_2$ -rich, bottom waters from fresher, lower  
106  $\text{CO}_2$  surface waters, and a summer thermocline at 20 m separates warmer surface waters from those  
107 below 4 °C (Janssen *et al.*, 1999). Upwelling of bottom waters from below the summer thermocline  
108 is a common summer occurrence, replenishing the surface nutrients while simultaneously lowering  
109 surface temperature and pH (Brutemark *et al.*, 2011). Baltic organisms are required to adapt to  
110 significant variations in environmental conditions. The species assemblage in the Baltic Sea is  
111 different to those studied during previous mesocosm experiments in the Arctic, North Sea and  
112 Korea (Brussaard *et al.*, 2013; Engel *et al.*, 2008; Kim *et al.*, 2010), and are largely unstudied in  
113 terms of their community trace gas production during the summer bloom. Following the spring  
114 bloom (July-August), a low dissolved inorganic nitrogen (DIN) to dissolved inorganic phosphorous  
115 (DIP) ratio combines with high temperatures and light intensities to encourage the growth of  
116 heterocystous cyanobacteria, (Niemisto *et al.*, 1989; Raateoja *et al.*, 2011), in preference to nitrate-  
117 dependent groups.

118 Here we report the concentrations of DMS, DMSP and halocarbons from the 2012 summer post-  
119 bloom season mesocosm experiment aimed to assess the impact of elevated  $f\text{CO}_2$  on the microbial  
120 community and trace gas production in the Baltic Sea. Our objective was to assess how changes in  
121 the microbial community driven by changes in  $f\text{CO}_2$  impacted DMS and halocarbon concentrations.  
122 It is anticipated that any effect of  $\text{CO}_2$  on the growth of different groups within the phytoplankton  
123 assemblage will result in an associated change in trace gas concentrations measured in the  
124 mesocosms as  $f\text{CO}_2$  increases, which can potentially be used to predict future halocarbon and  
125 sulphur emissions from the Baltic Sea region.

126



## 127 2 Methods

### 128 2.1 Mesocosm design and deployment

129 Nine mesocosms were deployed on the 10th June 2012 (day  $t-10$ ; days are numbered negative prior  
130 to CO<sub>2</sub> addition and positive afterward) and moored near Tvärminne Zoological Station (59° 51.5'  
131 N, 23° 15.5' E) in Tvärminne Storfjärden in the Baltic Sea. Each mesocosm comprised a  
132 thermoplastic polyurethane (TPU) enclosure of 17 m depth, containing approximately 54,000 L of  
133 seawater, supported by an 8m tall floating frame capped with a polyvinyl hood. For full technical  
134 details of the mesocosms see Czerny *et al.* (2013) and Riebesell *et al.* (2013). The mesocosm bags  
135 were filled by lowering through the stratified water column until fully submerged, with the opening  
136 at both ends covered by 3 mm mesh to exclude organisms larger than 3 mm such as fish and large  
137 zooplankton. The mesocosms were then left for 3 days ( $t-10$  to  $t-7$ ) with the mesh in position to  
138 allow exchange with the external water masses and ensure the mesocosm contents were  
139 representative of the phytoplankton community in the Storfjärden. On  $t-7$  the bottom of the  
140 mesocosm was sealed with a sediment trap and the upper opening was raised to approximately 1.5  
141 m above the water surface. Stratification within the mesocosm bags was broken up on  $t-5$  by the use  
142 of compressed air for three and a half minutes to homogenise the water column and ensure an even  
143 distribution of inorganic nutrients at all depths. Unlike in previous experiments, there was no  
144 addition of inorganic nutrients to the mesocosms at any time during the experiment; mean inorganic  
145 nitrate, inorganic phosphate and ammonium concentrations measured across all mesocosms at the  
146 start of the experiment were 37.2 ( $\pm$  18.8 s.d.) nmol L<sup>-1</sup>, 323.9 ( $\pm$  19.4 s.d.) nmol L<sup>-1</sup> and 413.8 ( $\pm$   
147 319.5 s.d.) nmol L<sup>-1</sup> respectively.

148 To obtain mesocosms with different  $f\text{CO}_2$ , the carbonate chemistry of the mesocosms was altered  
149 by the addition of different volumes of 50  $\mu\text{m}$  filtered, CO<sub>2</sub>-enriched Baltic Sea water (sourced from  
150 outside the mesocosms), to each mesocosm over a four-day period, with the first day of addition  
151 being defined as day  $t_0$ . Addition of the enriched CO<sub>2</sub> water was by the use of a bespoke dispersal  
152 apparatus ('Spider') lowered through the bags to ensure even distribution throughout the water  
153 column (further details are in Riebesell *et al.* 2013). Measurements of salinity in the mesocosms  
154 throughout the experiment determined that three of the mesocosms were not fully sealed, and had  
155 undergone unquantifiable water exchange with the surrounding waters. These three mesocosms  
156 (M2, M4 and M9) were excluded from the analysis. Two mesocosms were designated as controls  
157 (M1 and M5) and received only filtered seawater via the Spider; four mesocosms received addition  
158 of CO<sub>2</sub>-enriched waters, with the range of target  $f\text{CO}_2$  levels between 600 and 1650  $\mu\text{atm}$  (M7, 600

159  $\mu\text{atm}$ ; M6, 950  $\mu\text{atm}$ ; M3, 1300  $\mu\text{atm}$ ; M8 1650  $\mu\text{atm}$ ). Mesocosms were randomly allocated a  
160 target  $f\text{CO}_2$ ; a noticeable decrease in  $f\text{CO}_2$  was identified in the three highest  $f\text{CO}_2$  mesocosms (M6,  
161 M3 and M8) over the first half of the experiment, which required the addition of more  $\text{CO}_2$  enriched  
162 water on  $t15$  to bring the  $f\text{CO}_2$  back up to maximum concentrations (Fig. 1a; Paul *et al.*, 2015). A  
163 summary of the  $f\text{CO}_2$  in the mesocosms can be seen in Table 1. At the same time as this further  $\text{CO}_2$   
164 addition on  $t15$ , the walls of the mesocosms were cleaned using a bespoke wiper apparatus (See  
165 Riebesell *et al.*, 2013 for more information), followed by weekly cleaning to remove aggregations  
166 on the film which would block incoming light. Light measurements showed that over 95% of the  
167 photosynthetically active radiation (PAR) was transmitted by the clean TPU and PVC materials  
168 with 100% absorbance of UV light (Riebesell *et al.*, 2013). Samples for most parameters were  
169 collected from the mesocosms at the same time every morning from  $t-3$ , and analysed daily or every  
170 other day.

## 171 2.2 Trace gas extraction and analysis

### 172 2.2.1 DMS and halocarbons

173 A depth-integrated water sampler (IWS, HYDRO-BIOS, Kiel, Germany) was used to sample the  
174 entire 17 m water column daily or alternative daily. As analysis of Chlorophyll-*a* (Chl-*a*) showed it  
175 to be predominantly produced in the first 10 m of the water column, trace gas analysis was  
176 conducted on only integrated samples collected from the surface 10 m, with all corresponding  
177 community parameter analyses with the exception of pigment analysis performed also to this depth.  
178 Water samples for trace gas analysis were taken from the first IWS from each mesocosm to  
179 minimise the disturbance and bubble entrainment from taking multiple samples in the surface  
180 waters. As in Hughes *et al.* (2009), samples were collected in 250 mL amber glass bottles in a  
181 laminar flow with minimal disturbance to the water sample, using Tygon tubing from the outlet of  
182 the IWS. Bottles were rinsed twice before being carefully filled from the bottom with minimal  
183 stirring, and allowed to overflow the volume of the bottle approximately three times before sealing  
184 with a glass stopper to prevent bubble formation and atmospheric contact. Samples were stored  
185 below 10°C in the dark for 2 hours prior to analysis. Each day, a single sample was taken from each  
186 mesocosm, with two additional samples taken from one randomly selected mesocosm to evaluate  
187 the precision of the analysis ( $<4\%$ , no further data shown).

188 On return to the laboratory, 40 mL of water was injected into a purge and cryotrap system (Chuck *et al.*  
189 *et al.*, 2005), filtered through a 25 mm Whatman glass fibre filter (GF/F; GE Healthcare Life Sciences,  
190 Little Chalfont, England) and purged with oxygen-free nitrogen (OFN) at 80 mL  $\text{min}^{-1}$  for 10

191 minutes. Each gas sample passed through a glass wool trap to remove particles and aerosols, before  
192 a dual nafion counterflow drier (180 mL min<sup>-1</sup> OFN) removed water vapour from the gas stream.  
193 The gas sample was trapped in a stainless steel loop held at -150 °C in the headspace of a liquid  
194 nitrogen-filled dewar. The sample was injected by immersion of the sample loop in boiling water  
195 into an Agilent 6890 gas chromatograph equipped with a 60 m DB-VRX capillary column (0.32  
196 mm ID, 1.8 µm film thickness, Agilent J&W Ltd) according to the programme outlined by Hopkins  
197 *et al.* (2010). Analysis was performed by an Agilent 5973 quadrupole mass spectrometer operated  
198 in electron ionisation, single ion mode. Liquid standards of CH<sub>3</sub>I, diiodomethane (CH<sub>2</sub>I<sub>2</sub>), CH<sub>2</sub>ClI,  
199 iodoethane (C<sub>2</sub>H<sub>5</sub>I), iodopropane (C<sub>3</sub>H<sub>7</sub>I), CHBr<sub>3</sub>, dibromoethane (CH<sub>2</sub>Br<sub>2</sub>), dibromochloromethane  
200 (CHBr<sub>2</sub>Cl), bromiodomethane (CH<sub>2</sub>BrI) and DMS (Standards supplied by Sigma Aldrich Ltd, UK)  
201 were gravimetrically prepared by dilution in HPLC-grade methanol (Table 2) and used for  
202 calibration. The relative standard error was expressed as a percentage of the mean for the sample  
203 analysis, calculated for each compound using triplicate analysis each day from a single mesocosm,  
204 and was <7% for all compounds. GC-MS instrument drift was corrected by the use of a surrogate  
205 analyte standard in every sample, comprising deuterated DMS (D<sub>6</sub>-DMS), deuterated methyl iodide  
206 (CD<sub>3</sub>I) and <sup>13</sup>C dibromoethane (<sup>13</sup>C<sub>2</sub>H<sub>4</sub>Br<sub>2</sub>) via the method described in Hughes *et al.* (2006) and  
207 Martino *et al.* (2005). Five-point calibrations were performed weekly for each compound with the  
208 addition of the surrogate analyte, with a single standard analysed daily to check for instrument drift;  
209 linear regression from calibrations typically produced r<sup>2</sup>>0.98. All samples measured within the  
210 mesocosms were within the concentration ranges of the calibrations (Table 2).

### 211 2.2.2 DMSP

212 Samples for total DMSP (DMSP<sub>T</sub>) were collected and stored for later analysis by the acidification  
213 method of Curran *et al.* (1998). A 7 mL sub-sample was collected from the amber glass bottle into  
214 an 8 mL glass sample vial (Labhut, Churcham, UK), into which 0.35 µL of 50% H<sub>2</sub>SO<sub>4</sub> was added,  
215 before storage at ambient temperature. Particulate DMSP (DMSP<sub>P</sub>) samples were prepared by the  
216 gravity filtration of 20 mL of sample through a 47 mm GF/F in a glass filter unit, before careful  
217 removal and folding of the GF/F into a 7 mL sample vial filled with 7 mL of Milli-Q water and 0.35  
218 µL of H<sub>2</sub>SO<sub>4</sub> before storage at ambient temperature. Samples were stored for approximately 8  
219 weeks prior to analysis. DMSP samples (total and particulate) were analysed on a PTFE purge and  
220 cryotrap system using 2 mL of the sample purged with 1 mL of 10M NaOH for 5 minutes at 80 mL  
221 min<sup>-1</sup>. The sample gas stream passed through a glass wool trap and Nafion counterflow (Permapure)  
222 drier before being trapped in a PTFE sample loop kept at -150 °C by suspension in the headspace of  
223 a liquid nitrogen-filled dewar and controlled by feedback from a thermocouple. Immersion in

224 boiling water rapidly re-volatilised the sample for injection into a Shimadzu GC2010 gas  
225 chromatograph with a Varian Chrompack CP-Sil-5CB column (30 m, 0.53 mm ID) and flame  
226 photometric detector (FPD). The GC oven was operated isothermally at 60 °C which resulted in  
227 DMS eluting at 2.1 minutes. Liquid DMSP standards were prepared and purged in the same manner  
228 as the sample to provide weekly calibrations of the entire analytical system. Involvement in the  
229 2013 AQA 12-23 international DMS analysis proficiency test (National Measurement Institute of  
230 Australia, 2013) in February 2013 demonstrated excellent agreement between our method of DMSP  
231 analysis and the mean from thirteen laboratories measuring DMS using different methods, with a  
232 measurement error of 5%.

233 DMSP was not detected in any of the samples (total or particulate) collected and stored during the  
234 experiment, and it was considered likely that this was due to an unresolved issue regarding  
235 acidifying ~~the Baltic Sea~~ samples for later DMSP analysis. This method had been used during a  
236 previous mesocosm experiment (Bergen, Norway) and the results correlated well with those  
237 measured immediately on a similar GC-FPD system (Webb *et al.* 2015). ~~It~~ was considered unlikely  
238 that rates of bacterial DMSP turnover through demethylation rather than through cleavage to  
239 produce DMS (Curson *et al.*, 2011) were sufficiently high in the Baltic Sea to remove all detectable  
240 DMSP, yet still produce measureable DMS concentrations. Also, rapid turnover of dissolved  
241 DMSP<sub>D</sub> in surface waters being the cause of low DMSP<sub>T</sub> concentrations does not explain the lack  
242 of intracellular particulate-phase DMSP. Although production of DMS is possible from alternate  
243 sources, it is highly unlikely that there was a total absence of DMSP-producing phytoplankton  
244 within the mesocosms or Baltic Sea surface waters around Tvärminne; DMSP has been measured in  
245 surface waters of the Southern Baltic Sea at 22.2 nmol L<sup>-1</sup> in 2012, indicating that DMSP-producing  
246 species are present within the Baltic Sea (Cathleen Zindler, GEOMAR, Pers. Comm.).

247 A previous study by del Valle *et al.* (2011) highlighted up to 94% loss of DMSP<sub>T</sub> from acidified  
248 samples of colonial *Phaeocystis globosa* culture, and field samples dominated by colonial  
249 *Phaeocystis antarctica*. Despite filamentous, colonial cyanobacteria in the samples from Tvärminne  
250 mesocosms potentially undergoing the same process, these species did not dominate the community  
251 at only 6.6% of the total Chl-*a*, implying that the acidification method for DMSP fixation also failed  
252 for unicellular phytoplankton species. The findings of this mesocosm study is ~~is~~ suggests that the  
253 acidification method is unreliable in the Baltic Sea, and should be considered inadequate as the sole  
254 method of DMSP fixation in future experiments in the region. The DMSP acidification method is  
255 used worldwide as a simple and effective method of DMSP storage. The findings here, alongside  
256 those of del Valle *et al.* (2011) The question of its ~~the~~ applicability of this method in other marine

Formatted: Font: Italic

Formatted: Font: Italic

257 ~~waters environments, and suggests significant testing prior to reliance on this method as a sole~~  
258 ~~means of DMSP storage. also needs further investigation.~~

### 260 **2.3 Measurement of carbonate chemistry and community dynamics**

261 Water samples were collected from the 10~~b~~<sub>m</sub> and 17 m IWS on a daily basis and analysed for  
262 carbonate chemistry, fluorometric Chl-*a*, phytoplankton pigments (17 m IWS only) and cell  
263 abundance to analyse the community structure and dynamics during the experiment. The carbonate  
264 system was analysed through a suite of measurements (Paul *et al.*, 2015), including potentiometric  
265 titration for total alkalinity (TA), infrared absorption for dissolved inorganic carbon (DIC) and  
266 spectrophotometric determination for pH. For Chl-*a* analysis and pigment determination, 500 mL  
267 sub-samples were filtered through a GF/F and stored frozen (-20 °C for two hours for Chl-*a* and -80  
268 °C for up to 6 months for pigments), before homogenisation in 90 % acetone with glass beads. After  
269 centrifuging (10 minutes at 800 x g at 4 °C) the Chl-*a* concentrations were determined using a  
270 Turner AU-10 fluorometer by the methods of Welschmeyer (1994), and the phytoplankton pigment  
271 concentrations by reverse phase high performance liquid chromatography (WATERS HPLC with a  
272 Varian Microsorb-MV 100-3 C8 column) as described by Barlow *et al.* (1997). Phytoplankton  
273 community composition was determined by the use of the CHEMTAX algorithm to convert the  
274 concentrations of marker pigments to Chl-*a* equivalents (Mackey *et al.*, 1996; Schulz *et al.*, 2013).  
275 Microbes were enumerated using a Becton Dickinson FACSCalibur flow cytometer (FCM)  
276 equipped with a 488 nm argon laser (Crawford *et al.*, 2016) and counts of phytoplankton cells >20  
277 µm were made on concentrated (50 mL) sample water, fixed with acidic Lugol's iodine solution  
278 with an inverted microscope. Filamentous cyanobacteria were counted in 50 µm length units.

### 279 **2.4 Statistical Analysis**

280 All statistical analysis was performed using Minitab V16. In analysis of the measurements between  
281 mesocosms, one-way ANOVA was used with Tukey's post-hoc analysis test to determine the effect  
282 of different  $f\text{CO}_2$  on concentrations measured in the mesocosms and the Baltic Sea ( $H_0$  assumes no  
283 significant difference in the mean concentrations of trace gases measured through the duration of  
284 the experiment). Spearman's Rank Correlation Coefficients were calculated to compare the  
285 relationships between trace gas concentrations,  $f\text{CO}_2$ , and a number of biological parameters, and  
286 the resulting  $\rho$ -values for each correlation are given in Supplementary table S1 for the mesocosms  
287 and S2 for the Baltic Sea data.

288

## 289 **3 Results and Discussion**

### 290 **3.1 Biogeochemical changes within the mesocosms**

291 The mesocosm experiment was split into three phases based on the temporal variation in Chl-*a* (Fig.  
292 2; Paul *et al.*, 2015) evaluated after the experiment was completed:

- 293 • Phase 0 (days *t*-5 to *t*0) – pre-CO<sub>2</sub> addition
- 294 • Phase I (days *t*1 to *t*16) – ‘productive phase’
- 295 • Phase II (days *t*17 to *t*30) – temperature induced autotrophic decline.

#### 296 **3.1.1 Physical Parameters**

297 *f*CO<sub>2</sub> decreased over Phase I in the three highest *f*CO<sub>2</sub> mesocosms, mainly through air-sea gas  
298 exchange and carbon fixation by phytoplankton (Fig. 1a). All mesocosms still showed distinct  
299 differences in *f*CO<sub>2</sub> levels throughout the experiment (Table 1), and there was no overlap of  
300 mesocosm *f*CO<sub>2</sub> values on any given day, save for the two controls (M1 and M5). The control  
301 mesocosm *f*CO<sub>2</sub> increased through Phase I of the experiment, likely as a result of undersaturation of  
302 the water column encouraging dissolution of atmospheric CO<sub>2</sub> (Paul *et al.*, 2015). Salinity in the  
303 mesocosms remained constant throughout the experiment at  $5.70 \pm 0.004$ , and showed no variation  
304 with depth ([data not shown but available in Paul \*et al.\* 2015](#)). It remained similar to salinity in the  
305 Baltic Sea surrounding the mesocosms, which was  $5.74 \pm 0.14$ . Water temperature varied from a  
306 low of  $8.6 \pm 0.4$  °C during Phase 0 to a high of  $15.9 \pm 2.2$  °C measured on day *t*16, before  
307 decreasing once again (Fig. 1b).

308 Summertime upwelling events are common and well described (Gidhagen, 1987; Lehmann and  
309 Myrberg, 2008), and induce a significant temperature decrease in surface waters; such an event  
310 appears to have commenced around *t*16, as indicated by significantly decreasing temperatures  
311 inside and out of the mesocosms (Fig. 1b) and increased salinity in the Baltic Sea from 5.5 to 6.1  
312 over the following 15 days to the end of the experiment. Due to the enclosed nature of the  
313 mesocosms, the upwelling affected only the temperature and not pH, *f*CO<sub>2</sub> or the microbial  
314 community. However, the temperature decrease after *t*16 was likely to have had a significant effect  
315 on phytoplankton growth ([and biogenic gas production](#)), explaining the lower Chl-*a* in Phase II.

### 316 3.1.2 Community Dynamics

317 Mixing of the mesocosms and redistribution of the nutrients throughout the water column after  
318 closure (prior to  $t-3$ ) did not trigger a notable increase in total Chl-*a* in Phase 0 as was identified in  
319 previous mesocosm experiments; in previous mesocosm experiments, mixing redistributed nutrients  
320 from the deeper stratified layers throughout the water column. During Phase I, light availability,  
321 combined with increasing water temperatures favoured the growth of phytoplankton in all  
322 mesocosms (Paul *et al.* 2015), and was unlikely to be a direct result of the CO<sub>2</sub> enrichment, as no  
323 difference was identified between enriched mesocosms and controls. Mean Chl-*a* during Phase I  
324 was 1.98 ( $\pm$  0.29)  $\mu\text{g L}^{-1}$  from all mesocosms, decreasing to 1.44 ( $\pm$  0.46)  $\mu\text{g L}^{-1}$  in Phase II; this  
325 decrease was attributed to a temperature induced decrease in phytoplankton growth rates and  
326 higher grazing rates as a result of higher zooplankton reproduction rates during Phase I (Lischka *et*  
327 *al.*, 2015; Paul *et al.*, 2015). Mesocosm Chl-*a* decreased until the end of the experiment on  $t31$ .

328 The largest contributors to Chl-*a* in the mesocosms during the summer of 2012 were the  
329 chlorophytes and cryptophytes, with up to 40% and 21% contributions to the Chl-*a* respectively  
330 (Table 3; Paul *et al.*, 2015). Significant long-term differences in abundance between mesocosms  
331 developed as a result of elevated  $f\text{CO}_2$  in only two groups: picoeukaryotes I showed higher  
332 abundance at high  $f\text{CO}_2$  ( $F=8.2$ ,  $p<0.01$ ; Crawford *et al.*, 2016 and Supplementary Fig. S2), as seen  
333 in previous mesocosm experiments (Brussaard *et al.*, 2013; Newbold *et al.*, 2012) and  
334 picoeukaryotes III the opposite trend ( $F=19.6$ ,  $p<0.01$ ; ~~Crawford *et al.*, 2016~~ Crawford *et al.* ~~this~~  
335 ~~issue~~). Temporal variation in phytoplankton abundance was similar between all mesocosms  
336 (Supplementary Fig. S1 and S2).

337 Diazotrophic, filamentous cyanobacterial blooms in the Baltic Sea are an annual event in summer  
338 (Finni *et al.*, 2001), and single-celled cyanobacteria have been found to comprise as much as 80%  
339 of the cyanobacterial biomass and 50% of the total primary production during the summer in the  
340 Baltic Sea (Stal *et al.*, 2003). However, CHEMTAX analysis identified cyanobacteria as  
341 contributing less than 10% of the total Chl-*a* in the mesocosms (Crawford *et al.*, 2016; Paul *et al.*,  
342 2015). These observations were backed up by satellite observations showing reduced cyanobacterial  
343 abundance throughout the Baltic Sea in 2012 compared to previous and later years (Oberg, 2013). It  
344 was proposed that ~~environmental conditions of limited~~ light availability and ~~lower~~ surface water  
345 temperatures during the summer of 2012 were sub-optimal for triggering a filamentous  
346 cyanobacteria bloom (Wasmund, 1997).

## 347 3.2 DMS and DMSP

### 348 3.2.1 Mesocosm DMS

349 A significant 34% reduction in DMS concentrations was detected in the high  $f\text{CO}_2$  treatments  
350 during Phase II compared to the ambient  $f\text{CO}_2$  mesocosms ( $F=31.7$ ,  $p<0.01$ ). Mean DMS  
351 concentrations of  $5.0 (\pm 0.8)$ ; range  $3.5 - 6.8$   $\text{nmol L}^{-1}$  in the ambient treatments compared to  $3.3 (\pm$   
352  $0.3)$ ; range  $2.9 - 3.9$   $\text{nmol L}^{-1}$  in the 1333 and 1075  $\mu\text{atm}$  mesocosms (Fig. 23a). The primary  
353 differences identified were apparent from the start of Phase II on  $t17$ , after which maximum  
354 concentrations were observed in the ambient mesocosms on  $t21$ . The relationship between DMS  
355 and increasing  $f\text{CO}_2$  during Phase II was found to be linear (Fig. 32b), a finding also identified in  
356 previous mesocosm experiments (Archer *et al.*, 2013; Webb *et al.*, 2015). Furthermore, increases in  
357 DMS concentrations under high  $f\text{CO}_2$  were delayed by three days relative to the ambient and  
358 medium  $f\text{CO}_2$  treatments, a situation which has been observed in a previous mesocosm experiment.  
359 This was attributed to small-scale shifts in community composition and succession which could not  
360 be identified with only a once-daily measurement regime (Vogt *et al.*, 2008). DMS measured in all  
361 mesocosms fell within the range 2.7 to 6.8  $\text{nmol L}^{-1}$  across the course of the experiment. During  
362 Phase I, no difference was identified in DMS concentrations between  $f\text{CO}_2$  treatments with the  
363 mean of all mesocosms  $3.1 (\pm 0.2)$   $\text{nmol L}^{-1}$ . Concentrations in all mesocosms gradually declined  
364 from  $t21$  until the end of DMS measurements on  $t31$ . DMS concentrations measured in the  
365 mesocosms and Baltic Sea were comparable to those measured in temperate coastal conditions in  
366 the North Sea (Turner *et al.*, 1988), the Mauritanian upwelling (Franklin *et al.*, 2009; Zindler *et al.*,  
367 2012) and South Pacific (Lee *et al.*, 2010).

368 ~~Although~~ The majority of DMS production is presumed to be from DMSP. However, an alternative  
369 production route for DMS is available through the methylation of methanethiol (Drotar *et al.*, 1987;  
370 Kiene and Hines, 1995; Stets *et al.*, 2004) predominantly identified in anaerobic environments such  
371 as freshwater lake sediments (Lomans *et al.*, 1997), saltmarsh sediments (Kiene and Visscher,  
372 1987) and microbial mats (Visscher *et al.*, 2003; Zinder *et al.*, 1977). ~~However,~~ Recent studies  
373 have also identified this pathway of DMS production from *Pseudomonas deceptionensis* in an  
374 aerobic environment (Carrión *et al.*, 2015), where *P. deceptionensis* was unable to synthesise or  
375 catabolise DMSP, but was able to enzymatically mediate DMS production from methanethiol  
376 (MeSH). The same enzyme has also been identified in a wide range of other bacterial taxa,  
377 including the cyanobacterial *Pseudanabaena*, which was identified in the Baltic Sea during this and  
378 previous investigations (Stuhr, pers. comm.; Kangro *et al.*, 2007; Nausch *et al.*, 2009). Correlations



379 between DMS and the cyanobacterial equivalent Chl-*a* ( $\rho=0.42$ ,  $p<0.01$ ; [Supplementary Figure](#)  
380 [S1g](#)) and DMS and single-celled cyanobacteria ( $\rho=0.58$ ,  $p<0.01$ ; [Supplementary Figure S2a](#))  
381 ~~indicate-suggest~~ that the methylation pathway may be a potential source of DMS within the Baltic  
382 Sea community. In addition to the methylation pathway, DMS production has been identified from  
383 S-methylmethionine (Bentley and Chasteen, 2004), as well as from the reduction of  
384 dimethylsulphoxide (DMSO) in both surface and deep waters by bacterial metabolism (Hatton *et al.*, 2004). As these compounds were not measured in the mesocosms, it is impossible to determine  
385 if they were significant sources of DMS.  
386

387

388

### 389 3.2.2 DMS and Community Interactions

390 Throughout Phase I, DMS showed no correlation with any measured variables of biological activity  
391 or cell abundance, and was unaffected by elevated  $f\text{CO}_2$ , indicating ~~measured DMS net-production~~  
392 ~~concentrations was were~~ not directly related to the perturbation of the system and associated cellular  
393 stress (Sunda *et al.*, 2002). ~~Of the studied phytoplankton groupings, neither the cryptophytes or~~  
394 ~~chlorophytes as the largest contributors of Chl-*a* were identified as significant producers of DMSP.~~  
395 During Phase II, DMS was negatively correlated with Chl-*a* in the ambient and medium  $f\text{CO}_2$   
396 mesocosms ( $\rho=-0.60$ ,  $p<0.01$ ). During Phase II, a significant correlation was seen between DMS  
397 and single-celled cyanobacteria identified ~~predominantly~~ as *Synechococcus* ( $\rho=0.53$ ,  $p<0.01$ ;  
398 Crawford *et al.* 2016 and supplementary table S1) and picoeukaryotes III ( $\rho=0.75$ ,  $p<0.01$ ). The  
399 peak in DMS concentrations ~~on  $t_{21}$~~  is unlikely to be a delayed response to the increased Chl-*a* on  
400  ~~$t_{16}$  due to the time lag of 7 days. These higher DMS concentrations were likely connected to a peak~~  
401 ~~in dissolved organic carbon (DOC) on  $t_{15}$ , as well as increasing bacterial abundance during Phase II~~  
402 ~~(Hornick *et al.*, 2016). It is also likely that DMS concentrations increased as a response to the~~  
403 ~~mesocosm wall cleaning which took place on  $t_{16}$ . The variation in inorganic nutrient concentrations~~  
404 ~~between mesocosms at the start of the experiment did not have an effect on DMS concentrations~~  
405 ~~during Phase I, and by the start of Phase II the variation between mesocosms had decreased.~~

406 In previous mesocosm experiments (Archer *et al.*, 2013; Hopkins *et al.*, 2010; Webb *et al.*, 2015),  
407 DMS has shown poor correlations with many of the indicators of primary production and  
408 phytoplankton abundance, as well as showing the same trend of decreased concentrations in high  
409  $f\text{CO}_2$  mesocosms compared to ambient. DMS production is often uncoupled from measurements of  
410 primary production in open waters (Lana *et al.*, 2012), and also often from production of its

Formatted: Font: Italic

Formatted: Font: Italic

411 precursor DMSP (Archer *et al.*, 2009). DMS and DMSP are important sources of sulphur and  
412 carbon in the microbial food web for both bacteria and algae (Simó *et al.*, 2002, 2009), and since  
413 microbial turnover of DMSP and DMS play a significant role in net DMS production, it is  
414 unsurprising that DMS concentrations have shown poor correlation with DMSP-producing  
415 phytoplankton groups in past experiments and open waters.

416 DMS concentrations have been reported lower under conditions of elevated  $f\text{CO}_2$  compared to  
417 ambient controls, in both mesocosm experiments (Table 4) and phytoplankton monocultures  
418 (Arnold *et al.*, 2013; Avgoustidi *et al.*, 2012). However, the varying response of the community  
419 within each experiment se-experiments limit our ability to generalise the response of algal  
420 production of DMS and DMSP in all situations due to the characteristic community dynamics of  
421 each experiment in specific geographical areas and temporal periods. Previous experiments in the  
422 temperate Raunefjord of Bergen, Norway, showed lower abundance of DMSP-producing algal  
423 species, and subsequently DMSP-dependent DMS concentrations (Avgoustidi *et al.*, 2012; Hopkins  
424 *et al.*, 2010; Vogt *et al.*, 2008; Webb *et al.*, 2015). In contrast mesocosm experiments in the Arctic  
425 and Korea have shown increased abundance of DMSP producers (Archer *et al.*, 2013; Kim *et al.*,  
426 2010) but lower DMS concentrations, while incubation experiments by Hopkins and Archer (2014)  
427 showed lower DMSP production but higher DMS concentrations at high  $f\text{CO}_2$ . However, in all  
428 previous experiments with DMSP as the primary precursor of DMS, elevated  $f\text{CO}_2$  had a less  
429 marked effect on measured DMSP concentrations than on measured DMS concentrations. Hopkins  
430 *et al.* (2010) suggested that 'the perturbation of the system has a greater effect on the processes that  
431 control the conversion of DMSP to DMS rather than the initial production of DMSP itself'. ~~This is~~  
432 ~~relevant even for the current experiment, where DMSP was not identified, since processes~~  
433 ~~controlling DMS concentrations were likely more affected by the change in  $f\text{CO}_2$  than the~~  
434 ~~production of precursors.~~

435 Previous mesocosm experiments have suggested significant links between increased bacterial  
436 production through greater availability of organic substrates at high  $f\text{CO}_2$  (Engel *et al.*, 2013;  
437 Piontek *et al.*, 2013). Further, Endres *et al.* (2014) identified significant enhanced enzymatic  
438 hydrolysis of organic matter with increasing  $f\text{CO}_2$ , with higher bacterial abundance. Higher  
439 bacterial abundance will likely result in greater bacterial demand for sulphur, and therefore greater  
440 consumption of DMS and conversion to DMSO. This was suggested as a significant sink for DMS  
441 in a previous experiment (Webb *et al.*, 2015), but during the present experiment, both bacterial  
442 abundance and bacterial production were lower at high  $f\text{CO}_2$  (Hornick *et al.*, 2016). However, as it  
443 has been proposed that only specialist bacterial groups are DMS consumers (Vila-Costa *et al.*,

2006b), and there is no determination of the DMS consumption characteristics of the bacterial community in the Baltic Sea, ~~it is not known if this loss pathway is stimulated at this is still a potential stimulated DMS loss pathway at high  $f\text{CO}_2$ . As microbial DMS yields can vary between 5-40% depending on the sulphur and carbon demand (Kiene and Linn, 2000), a change in the bacterial sulphur requirements could change DMS turnover despite lower abundance. *Synechococcus* has been identified as a DMS consumer in the open ocean, but abundance of this group was negatively correlated with  $f\text{CO}_2$ , implying that DMS consumption by this group would have been lower as  $f\text{CO}_2$  increased.~~

### 3.3 Iodocarbons in the mesocosms and relationships with community composition

Elevated  $f\text{CO}_2$  did not affect the concentration of iodocarbons in the mesocosms significantly at any time during the experiment, which is in agreement with the findings of Hopkins *et al.* (2013) in the Arctic, but in contrast to Hopkins *et al.* (2010) and Webb (2015), where iodocarbons were measured significantly lower under elevated  $f\text{CO}_2$  (Table 4). Concentrations of all iodocarbons measured in the mesocosms and the Baltic Sea fall within the range of those measured previously in the region (Table 5). Mesocosm concentrations of  $\text{CH}_3\text{I}$  (Fig. 34a) and  $\text{C}_2\text{H}_5\text{I}$  (Fig. 34b) showed concentration ranges of 2.91 to 6.25 and 0.23 to 0.76  $\text{pmol L}^{-1}$  respectively.  $\text{CH}_3\text{I}$  showed a slight increase in all mesocosms during Phase I, peaking on *t*16 which corresponded with higher Chl-*a* concentrations, and correlated throughout the entire experiment with picoeukaryote groups II ( $\rho=0.59$ ,  $p<0.01$ ) and III ( $\rho=0.23$ ,  $p<0.01$ ; Crawford *et al.* [this issue 2016](#)) and nanoeukaryotes I ( $\rho=0.37$ ,  $p<0.01$ ). Significant differences identified between mesocosms for  $\text{CH}_3\text{I}$  were unrelated to elevated  $f\text{CO}_2$  ( $F=3.1$ ,  $p<0.05$ ), but concentrations were on average 15% higher in Phase II than Phase I.  $\text{C}_2\text{H}_5\text{I}$  decreased slightly during Phases I and II, although concentrations of this halocarbon were close to its detection limit (0.2  $\text{pmol L}^{-1}$ ), remaining below 1  $\text{pmol L}^{-1}$  at all times. As this compound showed no significant effect of elevated  $f\text{CO}_2$ , and was identified by Orlikowska and Schulz-Bull (2009) as having extremely low concentrations in the Baltic Sea (Table 5), it will not be discussed further.

No correlation was found between  $\text{CH}_3\text{I}$  and Chl-*a* at any phase, and the only correlation of any phytoplankton grouping was with nanoeukaryotes II ( $\rho=0.88$ ,  $p<0.01$ ; Crawford *et al.*, 2015). These  $\text{CH}_3\text{I}$  concentrations compare well to the 7.5  $\text{pmol L}^{-1}$  measured by Karlsson *et al.* (2008) during a cyanobacterial bloom in the Baltic Sea (Table 5), and the summer maximum of 16  $\text{pmol L}^{-1}$  identified by Orlikowska and Schulz-Bull (2009).

475 Karlsson *et al.* (2008) showed Baltic Sea halocarbon production occurring predominately during  
476 daylight hours, with concentrations at night decreasing by 70% compared to late afternoon. Light  
477 dependent production of CH<sub>3</sub>I has been shown to take place through abiotic processes, including  
478 radical recombination of CH<sub>3</sub> and I (Moore and Zafiriou, 1994). However, since samples were  
479 integrated over the surface 10m of the water column, it was impossible to determine if  
480 photochemistry was affecting iodocarbon concentrations near the surface where some UV light was  
481 able to pass between the top of the mesocosm film material and the cover. For the same reason,  
482 photodegradation of halocarbons (Zika *et al.*, 1984) within the mesocosms was also likely to have  
483 been significantly restricted. Thus, as photochemical production was expected to be minimal,  
484 biogenic production was likely to have been the dominant source of these compounds. Karlsson *et al.*  
485 (2008) identified *Pseudanabaena* as a key producer of CH<sub>3</sub>I in the Baltic Sea. However, the  
486 abundance of *Pseudanabaena* was highest during Phase I of the experiment (A. Stuhr, Pers.  
487 Comm.) when CH<sub>3</sub>I concentrations were lower, and as discussed previously, the abundance of these  
488 species constituted only a very small proportion of the community. Previous investigations in the  
489 laboratory have identified diatoms as significant producers of CH<sub>3</sub>I (Hughes *et al.*, 2013; Manley  
490 and De La Cuesta, 1997), and the low, steady-state abundance of the diatom populations in the  
491 mesocosms could have produced the same relatively steady-state trends in the iodocarbon  
492 concentrations.

493 Measured in the range 57.2 – 202.2 pmol L<sup>-1</sup> in the mesocosms, CH<sub>2</sub>I<sub>2</sub> (Fig. 34c) showed the  
494 clearest increase in concentration during Phase II, when it peaked on *t*21 in all mesocosms, with a  
495 maximum of 202.2 pmol L<sup>-1</sup> in M5 (348 µatm). During Phase II, concentrations of CH<sub>2</sub>I<sub>2</sub> were 57%  
496 higher than Phase I, and were therefore negatively correlated with Chl-*a*. The peak on *t*21  
497 corresponds with the peak identified in DMS on *t*21, and concentrations through all three phases  
498 correlate with picoeukaryotes II ( $\rho=0.62$ ,  $p<0.01$ ) and III ( $\rho=0.47$ ,  $p<0.01$ ) and nanoeukaryotes I  
499 ( $\rho=0.88$ ,  $p<0.01$ ; Crawford *et al.*, 2015). CH<sub>2</sub>CII (Fig. 34d) showed no peaks during either Phase I  
500 or Phase II, remaining within the range 3.81 to 8.03 pmol L<sup>-1</sup>, and again correlated with  
501 picoeukaryotes groups II ( $\rho=0.34$ ,  $p<0.01$ ) and III ( $\rho=0.38$ ,  $p<0.01$ ). These results may suggest that  
502 these groups possessed halo-peroxidase enzymes able to oxidise I, most likely as an anti-oxidant  
503 mechanism within the cell to remove H<sub>2</sub>O<sub>2</sub> (Butler and Carter-Franklin, 2004; Pedersen *et al.*, 1996;  
504 Theiler *et al.*, 1978). However, given the lack of response of these compounds to elevated *f*CO<sub>2</sub>  
505 ( $F=1.7$ ,  $p<0.01$ ), it is unlikely that production was increased in relation to elevated *f*CO<sub>2</sub>. Production  
506 of all iodocarbons increased during Phase II when total Chl-*a* decreased, particularly after the walls  
507 of the mesocosms were cleaned for the first time, releasing significant volumes of organic

508 aggregates into the water column. Aggregates have been suggested as a source of CH<sub>3</sub>I and C<sub>2</sub>H<sub>5</sub>I  
509 (Hughes *et al.*, 2008), likely through the alkylation of inorganic iodide (Urhahn and Ballschmiter,  
510 1998) or through the breakdown of organic matter by microbial activity to supply the precursors  
511 required for iodocarbon production (Smith *et al.*, 1992). Hughes *et al.* (2008) did not identify this  
512 route as a pathway for CH<sub>2</sub>I<sub>2</sub> or CH<sub>2</sub>ClI production, but Carpenter *et al.* (2005) suggested a  
513 production pathway for these compounds through the reaction of HOI with aggregated organic  
514 materials.

### 515 3.4 Bromocarbons in the mesocosms and the relationships with community 516 composition

517 No effect of elevated *f*CO<sub>2</sub> was identified for any of the three bromocarbons, which compared with  
518 the findings from previous mesocosms where bromocarbons were studied (Hopkins *et al.*, 2010,  
519 2013; Webb, 2015; Table 4). Measured concentrations were comparable to those of Orlikowska and  
520 Schulz-Bull (2009) and Karlsson *et al.* (2008) measured in the Southern part of the Baltic Sea  
521 (Table 3). The concentrations of CHBr<sub>3</sub>, CH<sub>2</sub>Br<sub>2</sub> and CHBr<sub>2</sub>Cl showed no major peaks of  
522 production in the mesocosms. CHBr<sub>3</sub> (Fig. 45a) decreased rapidly in all mesocosms over Phase 0  
523 from a maximum measured concentration of 147.5 pmol L<sup>-1</sup> in M1 (mean of 138.3 pmol L<sup>-1</sup> in all  
524 mesocosms) to a mean of 85.7 (±8.2 s.d.) pmol L<sup>-1</sup> in all mesocosms for the period *t*0 to *t*31 (Phases  
525 I and II). The steady-state CHBr<sub>3</sub> concentrations indicated a production source, however there was  
526 no clear correlation with any measured algal groups. CH<sub>2</sub>Br<sub>2</sub> concentrations (Fig. 5b4b) decreased  
527 steadily in all mesocosms from *t*-3 through to *t*31, over the range 4.0 to 7.7 pmol L<sup>-1</sup>, and CHBr<sub>2</sub>Cl  
528 followed a similar trend in the range 1.7 to 4.7 pmol L<sup>-1</sup> (Fig. 5e4c). Of the three bromocarbons,  
529 only CH<sub>2</sub>Br<sub>2</sub> showed correlation with total Chl-*a* ( $\rho=0.52$ ,  $p<0.01$ ), and with cryptophyte ( $\rho=0.86$ ,  
530  $p<0.01$ ) and dinoflagellate ( $\rho=0.65$ ,  $p<0.01$ ) derived Chl-*a*. Concentrations of CH<sub>2</sub>BrI were below  
531 detection limit for the entire experiment.

532 CH<sub>2</sub>Br<sub>2</sub> showed positive correlation with Chl-*a* ( $\rho=0.52$ ,  $p<0.01$ ), nanoeukaryotes II ( $\rho=0.34$ ,  
533  $p<0.01$ ) and cryptophytes ( $\rho=0.86$ ,  $p<0.01$ ; see supplementary material), whereas CHBr<sub>3</sub> and  
534 CHBr<sub>2</sub>Cl showed very weak or no correlation with any indicators of **algal biomass primary**  
535 **production**. Schall *et al.* (1997) have proposed that CHBr<sub>2</sub>Cl is produced in seawater by the  
536 nucleophilic substitution of bromide by chloride in CHBr<sub>3</sub>, which given the steady-state  
537 concentrations of CHBr<sub>3</sub> would explain the similar distribution of CHBr<sub>2</sub>Cl concentrations.  
538 Production of all three bromocarbons was identified from large-size cyanobacteria such as  
539 *Aphanizomenon flos-aquae* by Karlsson *et al.* (2008), and in addition, significant correlations were

Formatted: Subscript

540 found in the Arabian Sea between the abundance of the cyanobacterium *Trichodesmium* and several  
541 bromocarbons (Roy *et al.*, 2011), and the low abundance of such bacteria in the mesocosms would  
542 explain the low variation in bromocarbon concentrations through the experiment.

543 Halocarbon loss processes such as nucleophilic substitution (Moore, 2006), hydrolysis (Elliott and  
544 Rowland, 1995), sea-air exchange and microbial degradation are suggested as of greater importance  
545 than production of these compounds by specific algal groups, particularly given the relatively low  
546 growth rates and low net increase in total Chl-*a*. Hughes *et al.* (2013) identified bacterial inhibition  
547 of CHBr<sub>3</sub> production in laboratory cultures of *Thalassiosira* diatoms, but that it was not subject to  
548 bacterial breakdown; which could explain the relative steady state of CHBr<sub>3</sub> concentrations in the  
549 mesocosms. In contrast, significant bacterial degradation of CH<sub>2</sub>Br<sub>2</sub> in the same experiments could  
550 explain the steady decrease in CH<sub>2</sub>Br<sub>2</sub> concentrations seen in the mesocosms. Bacterial oxidation  
551 was also identified by Goodwin *et al.* (1998) as a significant sink for CH<sub>2</sub>Br<sub>2</sub>. As discussed for the  
552 iodocarbons, photolysis was unlikely due to the UV absorption of the mesocosm film, and limited  
553 UV exposure of the surface waters within the mesocosm due to the mesocosm cover. The ratio of  
554 CH<sub>2</sub>Br<sub>2</sub> to CHBr<sub>3</sub> was also unaffected by increased *f*CO<sub>2</sub>, staying within the range 0.04 to 0.08. This  
555 range in ratios is consistent with that calculated by Hughes *et al.* (2009) in the surface waters of an  
556 Antarctic depth profile, and attributed to higher sea-air flux of CHBr<sub>3</sub> than CH<sub>2</sub>Br<sub>2</sub> due to a greater  
557 concentrations gradient, despite the similar transfer velocities of the two compounds (Quack *et al.*,  
558 2007). Using cluster analysis in a time-series in the Baltic Sea, Orlikowska and Schulz-Bull (2009)  
559 identified both these compounds as originating from different sources and different pathways of  
560 production.

561 Macroalgal production would not have influenced the mesocosm concentrations after the bags were  
562 sealed due to the isolation from the coastal environment. ~~However~~ macroalgal production into the  
563 water column prior to mesocosm installation the higher bromocarbon concentrations identified in  
564 the mesocosms during Phase 0 may have originated from macroalgal sources (Klick, 1992;  
565 Leedham *et al.*, 2013; Moore and Tokarczyk, 1993) ~~prior to mesocosm closure, could account for~~  
566 the high initial concentrations with concentrations decreasing through the duration of the  
567 experiment via turnover and transfer to the atmosphere.

568

569 **3.5 Natural variations in Baltic Sea  $f\text{CO}_2$  and the effect on biogenic trace gases**

570 **3.5.1 Physical variation and community dynamics**

571 Baltic Sea deep waters have high  $f\text{CO}_2$  and subsequently lower pH (Schneider *et al.*, 2002), and the  
572 influx to the surface waters surrounding the mesocosms resulted in  $f\text{CO}_2$  increasing to 725  $\mu\text{atm}$  on  
573  $t_{31}$ , close to the average  $f\text{CO}_2$  of the third highest mesocosm (M6: 868  $\mu\text{atm}$ ). The input of  
574 upwelled water into the region mid-way through the experiment significantly altered the  
575 biogeochemical properties of the waters surrounding the mesocosms, and as a result it is  
576 inappropriate to directly compare the community structure and trace gas production of the Baltic  
577 Sea and the mesocosms. These conditions imply that pelagic communities in the Baltic Sea are  
578 regularly exposed to rapid changes in  $f\text{CO}_2$  and the associated pH, as well as having communities  
579 associated with the elevated  $f\text{CO}_2$  conditions. The changes in biological parameters and trace gas  
580 concentrations are therefore discussed here separately from the concentrations measured in the  
581 mesocosms.

582 Given the separation of the waters within the mesocosms, and the movement of water masses within  
583 the Baltic Sea, it is expected that phytoplankton population structure could be significantly different  
584 inside the mesocosms compared to the external waters. Chl-*a* followed the pattern of the  
585 mesocosms until  $t_4$ , after which concentrations were significantly higher than any mesocosm,  
586 peaking at 6.48  $\mu\text{g L}^{-1}$  on  $t_{16}$ , corresponding to the maximum Chl-*a* peak in the mesocosms and the  
587 maximum peak of temperature. As upwelled water intruded into the surface waters, the surface Chl-  
588 *a* was diluted with low Chl-*a* deep water: Chl-*a* in the surface 10m decreased from around  $t_{16}$  at the  
589 start of the upwelling until  $t_{31}$  when concentrations were once again equivalent to those found in  
590 the mesocosms at 1.30  $\mu\text{g L}^{-1}$ . In addition, there was potential introduction of different algal groups  
591 to the surface, but chlorophytes and cryptophytes were the major contributors to the Chl-*a* in the  
592 Baltic Sea, as in the mesocosms. Cyanobacteria contributed less than 2% of the total Chl-*a* in the  
593 Baltic Sea (Crawford *et al.*, 2016; Paul *et al.*, 2015).

594 Temporal community dynamics in the Baltic Sea were very different to that in the mesocosms  
595 across the experiment, with euglenophytes, chlorophytes, diatoms and prasinophytes all showing  
596 distinct peaks at the start of Phase II, with these same peaks identified in the nanoeukaryotes I and  
597 II, and picoeukaryotes II (Crawford *et al.*, 2016; Paul *et al.*, 2015; Supplementary Figs. S1 and S2).  
598 The decrease in abundance of many groups during Phase II was attributed to the decrease in  
599 temperature and dilution with low-abundance deep waters.

### 600 3.5.2 DMS in the Baltic Sea

601 The Baltic Sea samples gave a mean DMS concentration of  $4.6 \pm 2.6 \text{ nmol L}^{-1}$  but peaked at  $11.2$   
602  $\text{nmol L}^{-1}$  on *t*16, and were within the range of previous measurements for the region (Table 5).  
603 Strong correlations were seen between DMS and Chl-*a* ( $\rho=0.84$ ,  $p<0.01$ ), with the ratio of DMS:  
604 Chl-*a* at  $1.6 (\pm 0.3) \text{ nmol } \mu\text{g}^{-1}$ . Other strong correlations were seen with euglenophytes ( $\rho=0.89$ ,  
605  $p<0.01$ ), dinoflagellates ( $\rho=0.61$ ,  $p<0.05$ ) and nanoeukaryotes II ( $\rho=0.88$ ,  $p<0.01$ ), but no  
606 correlation was found between DMS and cyanobacterial abundance, or with picoeukaryotes III  
607 which was identified in the mesocosms, suggesting that DMS had a different origin in the Baltic  
608 Sea community than in the mesocosms. In addition, the community demands of sulphur are likely to  
609 be very different in the Baltic Sea compared to the mesocosms, due to differences in community  
610 composition and sulphur availability, and therefore direct comparisons with mesocosm  
611 concentrations are inappropriate. Once again, there was no DMSP detected in the samples.

612 As  $\text{CO}_2$  levels increased after *t*16 during Phase II, the DMS concentration measured in the Baltic  
613 Sea decreased, from the peak on *t*16 to the lowest recorded sample of the entire experiment at  $1.85$   
614  $\text{nmol L}^{-1}$  on *t*31. As with Chl-*a*, DMS concentrations in the surface of the Baltic Sea may have  
615 been diluted with low-DMS deep water, ~~however, the inverse relationship of DMS with  $\text{CO}_2$  shown~~  
616 ~~in the mesocosms may suggest that this decrease in DMS is attributed to the increase in  $\text{CO}_2$  levels.~~  
617 ~~Bacterial abundance was similar in the Baltic Sea as in the mesocosms (Hornick *et al.*, 2015),~~  
618 ~~however the injection of high  $\text{CO}_2$  water may have stimulated bacterial consumption of DMS~~  
619 ~~during the upwelling, which combined with the dilution of DMS rich surface water could have~~  
620 ~~resulted in the rapid decrease in DMS concentrations. As no discernible decrease in total bacterial~~  
621 ~~abundance was identified during the upwelling, it is also possible that the upwelled water contained~~  
622 ~~a different microbial community, and may potentially have introduced a higher abundance of DMS-~~  
623 ~~consuming microbes. No breakdown of bacterial distributions was available with which to test this~~  
624 ~~hypothesis.~~

Formatted: Font: Italic

### 625 3.5.3 Halocarbon concentrations in the Baltic Sea

626 Outside the mesocosms in the Baltic Sea,  $\text{CH}_3\text{I}$  was measured at a maximum concentration of  $8.65$   
627  $\text{pmol L}^{-1}$ , during Phase II, and showed limited effect of the upwelling event. Both  $\text{CH}_2\text{I}_2$  and  
628  $\text{CH}_2\text{CII}$  showed higher concentrations in the Baltic Sea samples than the mesocosms ( $\text{CH}_2\text{I}_2$ :  $373.9$   
629  $\text{pmol L}^{-1}$  and  $\text{CH}_2\text{CII}$ :  $18.1 \text{ pmol L}^{-1}$ ), and were correlated with the euglenophytes ( $\text{CH}_2\text{I}_2$ ;  $\rho=0.63$ ,  
630  $p<0.05$  and  $\text{CH}_2\text{CII}$ ;  $\rho=0.68$ ,  $p<0.01$ ) and nanoeukaryotes II ( $\text{CH}_2\text{I}_2$ ;  $\rho=0.53$ ,  $p<0.01$  and  $\text{CH}_2\text{CII}$ ;  
631  $\rho=0.58$ ,  $p<0.01$ ), but no correlation with Chl-*a*. Both polyhalogenated~~iodinated~~ compounds showed



632 correlation with picoeukaryote groups II and III, indicating that production was probably not limited  
633 to a single source. These concentrations of CH<sub>2</sub>I<sub>2</sub> and CH<sub>2</sub>ClI compared well to those measured  
634 over a macroalgal bed in the higher saline waters of the Kattegat by Klick and Abrahamsson (1992),  
635 suggesting that macroalgae were a significant iodocarbon source in the Baltic Sea. Macroalgal  
636 production in the Baltic Sea is likely the predominant iodocarbon source, compared to the  
637 mesocosms where macroalgae are excluded.

638 As with the iodocarbons, the Baltic Sea showed significantly higher concentrations of CHBr<sub>3</sub>  
639 (F=28.1, p<0.01), CH<sub>2</sub>Br<sub>2</sub> (F=208.8, p<0.01) and CHBr<sub>2</sub>Cl (F=23.5, p<0.01) than the mesocosms,  
640 with maximum concentrations 191.6 pmol L<sup>-1</sup>, 10.0 pmol L<sup>-1</sup> and 5.0 pmol L<sup>-1</sup> respectively. In the  
641 Baltic Sea, only CHBr<sub>3</sub> was correlated with Chl-*a* ( $\rho=0.65$ , p<0.05), cyanobacteria ( $\rho=0.61$ , p<0.01;  
642 Paul *et al.*, 2015) and nanoeukaryotes II ( $\rho=0.56$ , p<0.01; Crawford *et al.*, 2016), with the other  
643 two bromocarbons showing little to no correlations with any parameter of community activity.  
644 Production of bromocarbons from macroalgal sources (Laternus *et al.*, 2000; Leedham *et al.*, 2013;  
645 Manley *et al.*, 1992) was likely a significant contributor to the concentrations detected in the Baltic  
646 Sea; over the macroalgal beds in the Kattegat, Klick (1992) measured concentrations an order of  
647 magnitude higher than seen in this experiment for CH<sub>2</sub>Br<sub>2</sub> and CHBr<sub>2</sub>Cl. There was only a slight  
648 increase in bromocarbon concentrations as a result of the upwelling, indicating that the upwelled  
649 water had similar concentrations to the surface waters. These data from the Baltic Sea are presented  
650 as an important time-series of halocarbon measurements during the summer of 2012, which are  
651 expected to add to existing Baltic Sea trace gas datasets.

#### 653 4 The Baltic Sea as a natural analogue to future ocean acidification?

654 Mesocosm experiments are a highly valuable tool in assessing the potential impacts of elevated CO<sub>2</sub>  
655 on complex marine communities, however they are limited in that the rapid change in *f*CO<sub>2</sub>  
656 experienced by the community may not be representative of changes in the future ocean (Passow  
657 and Riebesell, 2005). This inherent problem with mesocosm experiments can be overcome through  
658 using naturally low pH/ high CO<sub>2</sub> areas such as upwelling regions or vent sites (Hall-Spencer *et al.*,  
659 2008), which can give an insight into populations already living and adapted-acclimated to high  
660 CO<sub>2</sub> regimes by exposure over timescales measured in years. This mesocosm experiment was  
661 performed at such a location with a relatively low-high *f*CO<sub>2</sub> excursion, however still low compared  
662 to some sites (800  $\mu$ atm compared to >2000  $\mu$ atm; Hall-Spencer *et al.*, 2008), and it was clear  
663 through the minimal variation in Chl-*a* between all mesocosms that the community was relatively

664 unaffected by elevated  $f\text{CO}_2$ , although variation could be identified in some phytoplankton groups  
665 and some shifts in community composition. The upwelling event occurring mid-way through our  
666 experiment allowed comparison of the mesocosm findings with a natural analogue of the system, as  
667 well as showing the extent to which the system perturbation can occur (up to 800  $\mu\text{atm}$ ). ~~This event  
668 was a fortuitous occurrence during this mesocosm experiment, but as the scale and timing of these  
669 upwelling events is difficult to determine. However, it is very difficult to determine where and  
670 when an upwelling will occur, and therefore these upwelling events are extremely challenging to  
671 study as natural high  $\text{CO}_2$  analogues. hard to utilise these events as natural high  $\text{CO}_2$  analogues.~~

Formatted: Subscript

672 In this paper, we described the temporal changes in concentrations of DMS and halocarbons in  
673 natural Baltic phytoplankton communities exposed to elevated  $f\text{CO}_2$  treatments. In contrast to the  
674 halocarbons, concentrations of DMS were significantly lower in the highest  $f\text{CO}_2$  treatments  
675 compared to the control. Despite very different physicochemical and biological characteristics of  
676 the Baltic Sea (e.g. salinity, community composition and nutrient concentrations), this is a very  
677 similar outcome to that seen in several other high  $f\text{CO}_2$  experiments. The Baltic Sea trace gas  
678 samples give a good record of trace gas ~~production-cycling~~ during the injection of high  $f\text{CO}_2$  deep  
679 water into the surface community during upwelling events. For the concentrations of halocarbons,  
680 ~~no response~~ the measured concentrations did not change during the upwelling event in  
681 the Baltic Sea, which may indicate that emissions of organic iodine and bromine are unlikely to  
682 change with future acidification of the Baltic Sea without significant alteration to the  
683 meteorological conditions. Further studies of these compounds are important to determine rates of  
684 production and consumption to include in prognostic and predictive models. However, net  
685 production of organic sulphur within the Baltic Sea region is likely to decrease with an acidified  
686 future ocean scenario, despite the possible acclimation of the microbial community to elevated  
687  $f\text{CO}_2$ . This will potentially impact the flux of DMS to the atmosphere over Northern Europe, and  
688 could have significant impacts on the local climate through the reduction of atmospheric sulphur  
689 aerosols. Data from a previous mesocosm experiment has been used to estimate future global  
690 changes in DMS production, and predicted that global warming would be amplified (Six *et al.*,  
691 2013); utilising the data from this experiment combined with those of other mesocosm, field and  
692 laboratory experiments and associated modelling provide the basis for a better understanding of the  
693 future changes in global DMS production and their climatic impacts.

694  
695

696 **Acknowledgements**

697 The Tvärminne 2012 mesocosm experiment was part of the SOPRAN II (Surface Ocean Processes  
698 in the Anthropocene) Programme (FKZ 03F0611) and BIOACID II (Biological Impacts of Ocean  
699 Acidification) project (FKZ 03F06550), funded by the German Ministry for Education and  
700 Research (BMBF) and led by the GEOMAR Helmholtz Centre for Ocean Research Kiel, Germany.

701 The authors thank all participants in the SOPRAN Tvärminne experiment for their assistance,  
702 including A. Ludwig for logistical support, the diving team, and the staff of Tvärminne Zoological  
703 Research Station for hosting the experiment. We also acknowledge the captain and crew of RV  
704 *ALKOR (AL394 and AL397)* for their work transporting, deploying and recovering the mesocosms.

705 This work was funded by a UK Natural Environment Research Council Directed Research  
706 Studentship (NE/H025588/1) through the UK Ocean Acidification Research Programme, with  
707 CASE funding from Plymouth Marine Laboratory. Additional funding was supplied by the EU  
708 Seventh Framework Program (FP7/2007-2013) MESOAQUA (EC Contract No. 228224).

709

710 Archer, S. D., Cummings, D., Llewellyn, C. and Fishwick, J.: Phytoplankton taxa, irradiance and nutrient availability  
711 determine the seasonal cycle of DMSP in temperate shelf seas, *Mar. Ecol. Prog. Ser.*, 394, 111–124,  
712 doi:10.3354/meps08284, 2009.

713 Archer, S. D., Kimmance, S. A., Stephens, J. A., Hopkins, F. E., Bellerby, R. G. J., Schulz, K. G., Piontek, J. and Engel,  
714 A.: Contrasting responses of DMS and DMSP to ocean acidification in Arctic waters, *Biogeosciences*, 10(3), 1893–  
715 1908, doi:10.5194/bg-10-1893-2013, 2013.

716 Arnold, H. E., Kerrison, P. and Steinke, M.: Interacting effects of ocean acidification and warming on growth and  
717 DMS-production in the haptophyte coccolithophore *Emiliania huxleyi*., *Glob. Chang. Biol.*, 19(4), 1007–16,  
718 doi:10.1111/gcb.12105, 2013.

719 Avgoustidi, V., Nightingale, P. D., Joint, I., Steinke, M., Turner, S. M., Hopkins, F. E. and Liss, P. S.: Decreased  
720 marine dimethyl sulfide production under elevated CO<sub>2</sub> levels in mesocosm and in vitro studies, *Environ. Chem.*, 9(4),  
721 399–404, doi:10.1071/EN11125, 2012.

722 Barlow, R. G., Cummings, D. G. and Gibb, S. W.: Improved resolution of mono- and divinyl chlorophylls a and b and  
723 zeaxanthin and lutein in phytoplankton extracts using reverse phase C-8 HPLC, *Mar. Ecol. Prog. Ser.*, 161, 303–307,  
724 1997.

725 Bentley, R. and Chasteen, T. G.: Environmental VOSCs—formation and degradation of dimethyl sulfide, methanethiol  
726 and related materials, *Chemosphere*, 55(3), 291–317, doi:10.1016/j.chemosphere.2003.12.017, 2004.

727 Brussaard, C. P. D., Noordeloos, A. A. M., Witte, H., Collenteur, M. C. J., Schulz, K., Ludwig, A. and Riebesell, U.:  
728 Arctic microbial community dynamics influenced by elevated CO<sub>2</sub> levels, *Biogeosciences*, 10(2), 719–731,  
729 doi:10.5194/bg-10-719-2013, 2013.

730 Brutemark, A., Engström-Öst, J. and Vehmaa, A.: Long-term monitoring data reveal pH dynamics, trends and  
731 variability in the western Gulf of Finland, *Oceanol. Hydrobiol. Stud.*, 40(3), 91–94, doi:10.2478/s13545-011-0034-3,  
732 2011.

733 Butler, A. and Carter-Franklin, J. N.: The role of vanadium bromoperoxidase in the biosynthesis of halogenated marine  
734 natural products, *Nat. Prod. Rep.*, 21(1), 180–188, doi:10.1039/b302337k, 2004.

735 Canadell, J. G., Le Quééré, C., Raupach, M. R., Field, C. B., Buitenhuis, E. T., Ciais, P., Conway, T. J., Gillett, N. P.,  
736 Houghton, R. A. and Marland, G.: Contributions to accelerating atmospheric CO<sub>2</sub> growth from economic activity,  
737 carbon intensity, and efficiency of natural sinks., *Proc. Natl. Acad. Sci. U. S. A.*, 104(47), 18866–18870,  
738 doi:10.1073/pnas.0702737104, 2007.

739 Carpenter, L. J., Hopkins, J. R., Jones, C. E., Lewis, A. C., Parthipan, R., Wevill, D. J., Poissant, L., Pilote, M. and  
740 Constant, P.: Abiotic source of reactive organic halogens in the sub-arctic atmosphere?, *Environ. Sci. Technol.*, 39(22),  
741 8812–8816 [online] Available from: <http://www.ncbi.nlm.nih.gov/pubmed/16323781>, 2005.

742 Carrión, O., Curson, A. R. J., Kumaresan, D., Fu, Y., Lang, A. S., Mercadé, E. and Todd, J. D.: A novel pathway  
743 producing dimethylsulphide in bacteria is widespread in soil environments, *Nat. Commun.*, 6, 6579,  
744 doi:10.1038/ncomms7579, 2015.

745 Chance, R., Baker, A. R., Küpper, F. C., Hughes, C., Kloareg, B. and Malin, G.: Release and transformations of  
746 inorganic iodine by marine macroalgae, *Estuar. Coast. Shelf Sci.*, 82, 406–414, doi:10.1016/j.ecss.2009.02.004, 2009.

747 Charlson, R. J., Lovelock, J. E., Andreae, M. O. and Warren, S. G.: Oceanic phytoplankton, atmospheric sulphur, cloud  
748 albedo and climate, *Nature*, 326(6114), 655–661 [online] Available from:  
749 [http://www.atmos.washington.edu/~sgw/PAPERS/1987\\_CLAW.pdf](http://www.atmos.washington.edu/~sgw/PAPERS/1987_CLAW.pdf) (Accessed 15 July 2011), 1987.

750 Chuck, A. L., Turner, S. M. and Liss, P. S.: Oceanic distributions and air-sea fluxes of biogenic halocarbons in the open  
751 ocean, *J. Geophys. Res.*, 110(C10022), doi:10.1029/2004JC002741, 2005.

752 Collins, M., Knutti, R., Arblaster, J., Dufresne, J.-L., Fichet, T., Frielingstein, P., Gao, X., Gutowski, W. J., Johns, T.,  
753 Krinner, G., Shongwe, M., Tebaldi, C., Weaver, A. J. and Wehner, M.: Long-term climate change: projections,  
754 commitments and irreversibility, in *Climate Change 2013: The Physical Science Basis. Contribution of Working Group  
755 1 to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, edited by T. Stocker, D. Qin, G.-  
756 K. Plattner, M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, and P. M. Midgley, Cambridge University  
757 Press, Cambridge, UK. [online] Available from: <http://dial.academielouvain.be/handle/boreal:140396> (Accessed 12  
758 June 2014), 2013.

759 Crawford, K., Brussaard, C. P. D. and Riebesell, U.: Shifts in the microbial community in the Baltic Sea with increasing  
760 CO<sub>2</sub>, *Biogeosciences*, In Press, 2016.

761 Cubasch, U., Wuebbles, D., Chen, D., Facchini, M. C., Frame, D., Mahowald, N. and Winther, J.-G.: Introduction, in  
762 Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of  
763 the Intergovernmental Panel on Climate Change, edited by T. F. Stocker, D. Qin, G.-K. Plattner, M. Tignor, S. K. Allen,  
764 J. Boschung, A. Nauels, Y. Xia, V. Bex, and P. M. Midgley, Cambridge University Press, Cambridge, UK., 2013.

765 Curran, M. A. J., Jones, G. B. and Burton, H.: Spatial distribution of dimethylsulfide and dimethylsulfoniopropionate in  
766 the Australasian sector of the Southern Ocean, *J. Geophys. Res.*, 103(D13), 16677 – 16689, 1998.

767 Curson, A. R. J., Todd, J. D., Sullivan, M. J. and Johnston, A. W. B.: Catabolism of dimethylsulphoniopropionate:  
768 microorganisms, enzymes and genes, *Nat. Rev. Microbiol.*, 9(12), 849–859, doi:10.1038/nrmicro2653, 2011.

769 Czerny, J., Schulz, K. G., Krug, S. A., Ludwig, A. and Riebesell, U.: Technical Note: The determination of enclosed  
770 water volume in large flexible-wall mesocosms “KOSMOS,” *Biogeosciences*, 10, 1937–1941, doi:10.5194/bg-10-1937-  
771 2013, 2013.

772 Doney, S. C., Fabry, V. J., Feely, R. A. and Kleypas, J. A.: Ocean acidification: the other CO<sub>2</sub> problem., *Ann. Rev.*  
773 *Mar. Sci.*, 1, 169–192, doi:10.1146/annurev.marine.010908.163834, 2009.

774 Drotar, A., Burton, G. A., Tavernier, J. E. and Fall, R.: Widespread occurrence of bacterial thiol methyltransferases and  
775 the biogenic emission of methylated sulfur gases, *Appl. Environ. Microbiol.*, 53(7), 1626–1631 [online] Available from:  
776 <http://aem.asm.org/content/53/7/1626.short> (Accessed 25 March 2014), 1987.

777 Elliott, S. and Rowland, F. S.: Methyl halide hydrolysis rates in natural waters, *J. Atmos. Chem.*, 20, 229–236, 1995.

778 Endres, S., Galgani, L., Riebesell, U., Schulz, K.-G. and Engel, A.: Stimulated bacterial growth under elevated pCO<sub>2</sub>:  
779 results from an off-shore mesocosm study., *PLoS One*, 9(6), e99228, doi:10.1371/journal.pone.0099228, 2014.

780 Engel, A., Schulz, K. G., Riebesell, U., Bellerby, R. G. J., Delille, B. and Schartau, M.: Effects of CO<sub>2</sub> on particle size  
781 distribution and phytoplankton abundance during a mesocosm bloom experiment (PeECE II), *Biogeosciences*, 5(2),  
782 509–521, doi:10.5194/bg-5-509-2008, 2008.

783 Engel, A., Borchard, C., Piontek, J., Schulz, K. G., Riebesell, U. and Bellerby, R. G. J.: CO<sub>2</sub> increases 14C primary  
784 production in an Arctic plankton community, *Biogeosciences*, 10(3), 1291–1308, doi:10.5194/bg-10-1291-2013, 2013.

785 Finni, T., Kononen, K., Olsonen, R. and Wallström, K.: The History of Cyanobacterial Blooms in the Baltic Sea,  
786 *AMBIO A J. Hum. Environ.*, 30(4), 172–178, doi:10.1579/0044-7447-30.4.172, 2001.

787 Franklin, D. J., Poulton, A. J., Steinke, M., Young, J., Peeken, I. and Malin, G.: Dimethylsulphide, DMSP-lyase activity  
788 and microplankton community structure inside and outside of the Mauritanian upwelling, *Prog. Oceanogr.*, 83(1-4),  
789 134–142, doi:10.1016/j.pocean.2009.07.011, 2009.

790 Gidhagen, L.: Coastal upwelling in the Baltic Sea—Satellite and in situ measurements of sea-surface temperatures  
791 indicating coastal upwelling, *Estuar. Coast. Shelf Sci.*, 24, 449–462 [online] Available from:  
792 <http://www.sciencedirect.com/science/article/pii/0272771487901272> (Accessed 16 August 2014), 1987.

793 Goodwin, K., Schaefer, J. K. and Oremland, R. S.: Bacterial oxidation of dibromomethane and methyl bromide in  
794 natural waters and enrichment cultures, *Appl. Environ. Microbiol.*, 64(12), 4629 –4636 [online] Available from:  
795 <http://aem.asm.org/content/64/12/4629.short> (Accessed 30 July 2014), 1998.

796 Goodwin, K. D., North, W. J. and Lidstrom, M. E.: Production of bromoform and dibromomethane by giant kelp:  
797 factors affecting release and comparison to anthropogenic bromine sources, *Limnol. Oceanogr.*, 42(8), 1725–1734,  
798 doi:10.4319/lo.1997.42.8.1725, 1997.

799 Hall-Spencer, J. M., Rodolfo-Metalpa, R., Martin, S., Ransome, E., Fine, M., Turner, S. M., Rowley, S. J., Tedesco, D.  
800 and Buia, M.-C.: Volcanic carbon dioxide vents show ecosystem effects of ocean acidification, *Nature*, 454(7200), 96–  
801 99, doi:10.1038/nature07051, 2008.

802 Hartmann, D. L., Klein Tank, A. M. G., Rusticucci, M., Alexander, L. V., Bronnimann, S., Charabi, Y., Dentener, F. J.,  
803 Dlugokencky, E. J., Easterling, D. R., Kaplan, A., Soden, B. J., Thorne, P. W., Wild, M. and Zhai, P. M.: Observations:  
804 Atmosphere and Surface, in Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the  
805 Fifth Assessment Report of the Intergovernmental Panel on Climate Change, edited by T. F. Stocker, D. Qin, G.-K.  
806 Plattner, M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, and P. M. Midgley, Cambridge University  
807 Press, Cambridge, Cambridge, UK., 2013.

808 Hatton, A. D., Darroch, L. and Malin, G.: The role of dimethylsulphoxide in the marine biogeochemical cycle of  
809 dimethylsulphide, *Oceanogr. Mar. Biol. Annu. Rev.*, 42, 29–56, 2004.

810 Hofmann, G. E., Barry, J. P., Edmunds, P. J., Gates, R. D., Hutchins, D. A., Klinger, T. and Sewell, M. A.: The effect of  
811 ocean acidification on calcifying organisms in marine ecosystems: an organism-to-ecosystem perspective, *Annu. Rev.*  
812 *Ecol. Evol. Syst.*, 41(1), 127–147, doi:10.1146/annurev.ecolsys.110308.120227, 2010.

813 Hopkins, F. E. and Archer, S. D.: Consistent increase in dimethyl sulphide (DMS) in response to high CO<sub>2</sub> in five  
814 shipboard bioassays from contrasting NW European waters, *Biogeosciences*, 11(2), 4925 – 4940, doi:10.5194/bg-11-  
815 2267-2014, 2014.

816 Hopkins, F. E., Turner, S. M., Nightingale, P. D., Steinke, M., Bakker, D. and Liss, P. S.: Ocean acidification and  
817 marine trace gas emissions., *Proc. Natl. Acad. Sci. U. S. A.*, 107(2), 760–765, doi:10.1073/pnas.0907163107, 2010.

818 Hopkins, F. E., Kimmance, S. A., Stephens, J. A., Bellerby, R. G. J., Brussaard, C. P. D., Czerny, J., Schulz, K. G. and  
819 Archer, S. D.: Response of halocarbons to ocean acidification in the Arctic, *Biogeosciences*, 10(4), 2331–2345,  
820 doi:10.5194/bg-10-2331-2013, 2013.

821 Hornick, T., Bach, L. T., Crawford, K. J., Spilling, K., Achterberg, E. P., Brussaard, C. P. D., Riebesell, U. and  
822 Grossart, H.-P.: Ocean acidification indirectly alters trophic interaction of heterotrophic bacteria at low nutrient  
823 conditions, *Biogeosciences Discuss.*, (March), 1–37, doi:10.5194/bg-2016-61, 2016.

824 Hughes, C., Malin, G., Nightingale, P. D. and Liss, P. S.: The effect of light stress on the release of volatile iodocarbons  
825 by three species of marine microalgae, *Limnol. Oceanogr.*, 51(6), 2849–2854 [online] Available from:  
826 <http://cat.inist.fr/?aModele=afficheN&cpsid=18312251> (Accessed 26 July 2013), 2006.

827 Hughes, C., Malin, G., Turley, C. M., Keely, B. J., Nightingale, P. D. and Liss, P. S.: The production of volatile  
828 iodocarbons by biogenic marine aggregates, *Limnol. Oceanogr.*, 53(2), 867–872, 2008.

829 Hughes, C., Chuck, A. L., Rossetti, H., Mann, P. J., Turner, S. M., Clarke, A., Chance, R. and Liss, P. S.: Seasonal  
830 cycle of seawater bromoform and dibromomethane concentrations in a coastal bay on the western Antarctic Peninsula,  
831 *Global Biogeochem. Cycles*, 23, doi:10.1029/2008GB003268, 2009.

832 Hughes, C., Franklin, D. J. and Malin, G.: Iodomethane production by two important marine cyanobacteria:  
833 *Prochlorococcus marinus* (CCMP 2389) and *Synechococcus* sp. (CCMP 2370), *Mar. Chem.*, 125(1-4), 19–25,  
834 doi:10.1016/j.marchem.2011.01.007, 2011.

835 Hughes, C., Johnson, M., Utting, R., Turner, S., Malin, G., Clarke, a. and Liss, P. S.: Microbial control of bromocarbon  
836 concentrations in coastal waters of the western Antarctic Peninsula, *Mar. Chem.*, 151, 35–46,  
837 doi:10.1016/j.marchem.2013.01.007, 2013.

838 Janssen, F., Schrum, C. and Backhaus, J.: A climatological data set of temperature and salinity for the Baltic Sea and  
839 the North Sea, *Dtsch. Hydrogr. Zeitschrift, Supplement* [online] Available from:  
840 <http://link.springer.com/article/10.1007/BF02933676> (Accessed 16 August 2014), 1999.

841 Kangro, K., Olli, K., Tamminen, T. and Lignell, R.: Species-specific responses of a cyanobacteria-dominated  
842 phytoplankton community to artificial nutrient limitation in the Baltic Sea, *Mar. Ecol. Prog. Ser.*, 336, 15–27

843 Karlsson, A., Auer, N., Schulz-Bull, D. and Abrahamsson, K.: Cyanobacterial blooms in the Baltic — A source of  
844 halocarbons, *Mar. Chem.*, 110, 129–139, doi:10.1016/j.marchem.2008.04.010, 2008.

845 Kiene, R. P. and Hines, M. E.: Microbial formation of dimethyl sulfide in anoxic sphagnum peat, *Appl. Environ.*  
846 *Microbiol.*, 61(7), 2720–2726 [online] Available from: <http://aem.asm.org/content/61/7/2720.short> (Accessed 6  
847 February 2014), 1995.

848 Kiene, R. P. and Linn, L. J.: Distribution and turnover of dissolved DMSP and its relationship with bacterial production  
849 and dimethylsulfide in the Gulf of Mexico, *Limnol. Oceanogr.*, 45(4), 849–861, 2000.

850 Kiene, R. P. and Visscher, P. T.: Production and fate of methylated sulfur compounds from methionine and  
851 dimethylsulfoniopropionate in anoxic salt marsh sediments., *Appl. Environ. Microbiol.*, 53(10), 2426–2434, 1987.

852 Kim, J.-M., Lee, K., Yang, E. J., Shin, K., Noh, J. H., Park, K.-T., Hyun, B., Jeong, H.-J., Kim, J.-H., Kim, K. Y., Kim,  
853 M., Kim, H.-C., Jang, P.-G. and Jang, M.-C.: Enhanced production of oceanic dimethylsulfide resulting from CO<sub>2</sub>-  
854 induced grazing activity in a high CO<sub>2</sub> world., *Environ. Sci. Technol.*, 44(21), 8140–8143, doi:10.1021/es102028k,  
855 2010.

856 Klick, S.: Seasonal variations of biogenic and anthropogenic halocarbons in seawater from a coastal site, *Limnol.*  
857 *Oceanogr.*, 37(7), 1579–1585 [online] Available from: <http://cat.inist.fr/?aModele=afficheN&cpsid=4788349>  
858 (Accessed 4 August 2014), 1992.

- 859 Klick, S. and Abrahamsson, K.: Biogenic volatile iodated hydrocarbons in the ocean, *J. Geophys. Res.*, 97(C8), 12683–  
860 12687 [online] Available from: <http://www.agu.org/journals/ABS/1992/92JC00948.shtml> (Accessed 26 July 2013),  
861 1992.
- 862 Lana, A., Simó, R., Vallina, S. M. and Dachs, J.: Re-examination of global emerging patterns of ocean DMS  
863 concentration, *Biogeochemistry*, 110, 173–182, doi:10.1007/s10533-011-9677-9, 2012.
- 864 Laturnus, F., Giese, B., Wiencke, C. and Adams, F. C.: Low-molecular-weight organoiodine and organobromine  
865 compounds released by polar macroalgae—the influence of abiotic factors, *Fresenius. J. Anal. Chem.*, 368, 297–302  
866 [online] Available from: <http://www.ncbi.nlm.nih.gov/pubmed/11220596>, 2000.
- 867 Leck, C. and Rodhe, H.: Emissions of marine biogenic sulfur to the atmosphere of northern Europe, *J. Atmos. Chem.*,  
868 12, 63–86 [online] Available from: <http://www.springerlink.com/index/H5472PG48150N025.pdf> (Accessed 26 July  
869 2013), 1991.
- 870 Leck, C., Larsson, U., Bågander, L. E., Johansson, S. and Hajdu, S.: Dimethyl sulfide in the Baltic Sea: annual  
871 variability in relation to biological activity, *J. Geophys. Res.*, 95(C3), 3353–3363, doi:10.1029/JC095iC03p03353,  
872 1990.
- 873 Lee, G., Park, J., Jang, Y., Lee, M., Kim, K. R., Oh, J. R., Kim, D., Yi, H. II and Kim, T. Y.: Vertical variability of  
874 seawater DMS in the South Pacific Ocean and its implication for atmospheric and surface seawater DMS,  
875 *Chemosphere*, 78(8), 1063–1070, doi:10.1016/j.chemosphere.2009.10.054, 2010.
- 876 Leedham, E. C., Hughes, C., Keng, F. S. L., Phang, S.-M., Malin, G. and Sturges, W. T.: Emission of atmospherically  
877 significant halocarbons by naturally occurring and farmed tropical macroalgae, *Biogeosciences*, 10(6), 3615–3633,  
878 doi:10.5194/bg-10-3615-2013, 2013.
- 879 Lehmann, A. and Myrberg, K.: Upwelling in the Baltic Sea — A review, *J. Mar. Syst.*, 74, S3–S12,  
880 doi:10.1016/j.jmarsys.2008.02.010, 2008.
- 881 Lischka, S., Riebesell, U., Stühr, A. and Bermudez, J. R.: Micro- and mesozooplankton community response to  
882 increasing levels of CO<sub>2</sub> in the Baltic Sea: insights from a large-scale mesocosm experiment, *Biogeosciences*,  
883 Submitted, 2015.
- 884 Liss, P., Marandino, C. A., Dahl, E., Helmig, D., Hintsä, E. J., Hughes, C., Johnson, M., Moore, R. M., Plane, J. M. C.,  
885 Quack, B., Singh, H. B., Stefels, J., von Glasow, R. and Williams, J.: Short-lived trace gases in the surface ocean and  
886 the atmosphere, in *Ocean-Atmosphere Interactions of Gases and Particles*, edited by P. Liss and M. Johnson, pp. 55–  
887 112., 2014.
- 888 Lomans, B. P., Smolders, A., Intven, L. M., Pol, A., Op, D. and van der Drift, C.: Formation of dimethyl sulfide and  
889 methanethiol in anoxic freshwater sediments, *Appl. Environ. Microbiol.*, 63(12), 4741–4747, 1997.
- 890 Mackey, M. D., Mackey, D. J., Higgins, H. W. and Wright, S. W.: CHEMTAX a program for estimating class  
891 abundances from chemical markers: application to HPLC measurements of phytoplankton, *Mar. Ecol. Prog. Ser.*, 144,  
892 265–283 [online] Available from: <http://www.int-res.com/abstracts/meps/v144/p265-283> (Accessed 25 March 2014),  
893 1996.
- 894 Manley, S. L. and De La Cuesta, J. L.: Methyl iodide production from marine phytoplankton cultures, *Limnol.*  
895 *Oceanogr.*, 42(1), 142–147, doi:10.4319/lo.1997.42.1.0142, 1997.
- 896 Manley, S. L., Goodwin, K. and North, W. J.: Laboratory production of bromoform, methylene bromide, and methyl  
897 iodide by macroalgae in and distribution nearshore Southern California waters, *Limnol. Oceanogr.*, 37(8), 1652–1659,  
898 1992.
- 899 Martino, M., Liss, P. S. and Plane, J. M. C.: The photolysis of dihalomethanes in surface seawater., *Environ. Sci.*  
900 *Technol.*, 39(18), 7097–7101 [online] Available from: <http://www.ncbi.nlm.nih.gov/pubmed/16201634>, 2005.
- 901 Moore, R. M.: Methyl halide production and loss rates in sea water from field incubation experiments, *Mar. Chem.*,  
902 101(3-4), 213–219, doi:10.1016/j.marchem.2006.03.003, 2006.
- 903 Moore, R. M. and Tokarczyk, R.: Volatile biogenic halocarbons in the northwest Atlantic, *Global Biogeochem. Cycles*,  
904 7(1), 195–210, 1993.
- 905 Moore, R. M. and Zafiriou, O. C.: Photochemical production of methyl iodide in seawater, *J. Geophys. Res.*, 99(D8),  
906 16415–16420, doi:10.1029/94JD00786, 1994.
- 907 National Measurement Institute of Australia: Proficiency Study 12-23: DMS in seawater., 2013.

- 908 Nausch, M., Nausch, G., Lass, H. U., Mohrholz, V., Nagel, K., Siegel, H. and Wasmund, N.: Phosphorus input by  
909 upwelling in the eastern Gotland Basin (Baltic Sea) in summer and its effects on filamentous cyanobacteria, *Estuar.  
910 Coast. Shelf Sci.*, 83(4), 434–442, doi:10.1016/j.ecss.2009.04.031, 2009.
- 911 Newbold, L. K., Oliver, A. E., Booth, T., Tiwari, B., DeSantis, T., Maguire, M., Andersen, G., van der Gast, C. J. and  
912 Whiteley, A. S.: The response of marine picoplankton to ocean acidification, *Environ. Microbiol.*, 14(9), 2293–2307,  
913 doi:10.1111/j.1462-2920.2012.02762.x, 2012.
- 914 Niemisto, L., Rinne, I. and Melvasalo, T.: Blue-green algae and their nitrogen fixation in the Baltic Sea in 1980, 1982  
915 and 1984, *Meri*, 17, 1–59, 1989.
- 916 O'Dowd, C. D., Jimenez, J. L., Bahreini, R., Flagan, R. C., Seinfeld, J. H., Hameri, K., Pirjola, L., Kulmala, M.,  
917 Jennings, S. G. and Hoffmann, T.: Marine aerosol formation from biogenic iodine emissions, *Nature*, 417(6889), 632–  
918 636, doi:10.1038/nature00773.1.2.3.4.5.6.7.8.9.10., 2002.
- 919 Oberg, J.: Cyanobacterial blooms in the Baltic Sea in 2013, *HELCOM Balt. Sea Environ. Fact Sheet*, 2013.
- 920 Orlikowska, A. and Schulz-Bull, D. E.: Seasonal variations of volatile organic compounds in the coastal Baltic Sea,  
921 *Environ. Chem.*, 6, 495–507, doi:10.1071/EN09107, 2009.
- 922 Park, K.-T., Lee, K., Shin, K., Yang, E. J., Hyun, B., Kim, J.-M., Noh, J. H., Kim, M., Kong, B., Choi, D. H., Choi, S.-  
923 J., Jang, P.-G. and Jeong, H. J.: Direct linkage between dimethyl sulfide production and microzooplankton grazing,  
924 resulting from prey composition change under high partial pressure of carbon dioxide conditions., *Environ. Sci.  
925 Technol.*, 48(9), 4750–4756, doi:10.1021/es403351h, 2014.
- 926 Passow, U. and Riebesell, U.: Mesocosm perturbation experiments and the sensitivity of marine biological systems to  
927 global change, *Solas News*, (1), 12–13, doi:10.1029/2003JC002120, 2005.
- 928 Paul, A. J., Bach, L. T., Schulz, K.-G., Boxhammer, T., Czerny, J., Achterberg, E., Hellemann, D., Trense, Y., Nausch,  
929 M., Sswat, M. and Riebesell, U.: Effect of elevated CO<sub>2</sub> on organic matter pools and fluxes in a summer Baltic Sea  
930 plankton community., *Biogeosciences*, 12, 6181 – 6203, 2015.
- 931 Pedersen, M., Collen, J., Abrahamsson, K. and Ekdahl, A.: Production of halocarbons from seaweeds: an oxidative  
932 stress reaction?, *Sci. Mar.*, 60(Supplement 1), 257–263, 1996.
- 933 Piontek, J., Borchard, C., Sperling, M., Schulz, K. G., Riebesell, U. and Engel, A.: Response of bacterioplankton  
934 activity in an Arctic fjord system to elevated pCO<sub>2</sub>: results from a mesocosm perturbation study, *Biogeosciences*, 10,  
935 297–314, doi:10.5194/bg-10-297-2013, 2013.
- 936 Quack, B., Peeken, I., Petrick, G. and Nachtigall, K.: Oceanic distribution and sources of bromoform and  
937 dibromomethane in the Mauritanian upwelling, *J. Geophys. Res.*, 112, C10006, doi:10.1029/2006JC003803, 2007.
- 938 Quinn, P. K. and Bates, T. S.: The case against climate regulation via oceanic phytoplankton sulphur emissions.,  
939 *Nature*, 480(7375), 51–56, doi:10.1038/nature10580, 2011.
- 940 Raateoja, M., Kuosa, H. and Hällfors, S.: Fate of excess phosphorus in the Baltic Sea: A real driving force for  
941 cyanobacterial blooms?, *J. Sea Res.*, 65(2), 315–321, doi:10.1016/j.seares.2011.01.004, 2011.
- 942 Raven, J. R., Caldeira, K., Elderfield, H., Hoegh-Guldberg, O., Liss, P. S., Riebesell, U., Shepherd, J., Turley, C. and  
943 Watson, A.: Ocean acidification due to increasing atmospheric carbon dioxide, *R. Soc. Policy Doc.* 12/05, (June)  
944 [online] Available from: [http://eprints.uni-  
945 kiel.de/7878/1/965\\_Raven\\_2005\\_OceanAcidificationDueToIncreasing\\_Monogr\\_pubid13120.pdf](http://eprints.uni-kiel.de/7878/1/965_Raven_2005_OceanAcidificationDueToIncreasing_Monogr_pubid13120.pdf) (Accessed 26 July  
946 2013), 2005.
- 947 Riebesell, U., Czerny, J., von Bröckel, K., Boxhammer, T., Büdenbender, J., Deckelnick, M., Fischer, M., Hoffmann,  
948 D., Krug, S. A., Lentz, U., Ludwig, A., Mücke, R. and Schulz, K. G.: Technical Note: A mobile sea-going mesocosm  
949 system – new opportunities for ocean change research, *Biogeosciences*, 10(3), 1835–1847, doi:10.5194/bg-10-1835-  
950 2013, 2013.
- 951 Ross, P. M., Parker, L., O'Connor, W. A. and Bailey, E. A.: The impact of ocean acidification on reproduction, early  
952 development and settlement of marine organisms, *Water*, 3(4), 1005–1030, doi:10.3390/w3041005, 2011.
- 953 Roy, R., Pratihary, A., Narvenkar, G., Mochemadkar, S., Gauns, M. and Naqvi, S. W. A.: The relationship between  
954 volatile halocarbons and phytoplankton pigments during a *Trichodesmium* bloom in the coastal eastern Arabian Sea,  
955 *Estuar. Coast. Shelf Sci.*, 95(1), 110–118, doi:10.1016/j.ecss.2011.08.025, 2011.
- 956 Scarratt, M. G. and Moore, R. M.: Production of methyl bromide and methyl chloride in laboratory cultures of marine



- 957 phytoplankton II, *Mar. Chem.*, 59(3-4), 311–320, doi:10.1016/S0304-4203(97)00092-3, 1998.
- 958 Schall, C., Heumann, K. G. and Kirst, G. O.: Biogenic volatile organoiodine and organobromine hydrocarbons in the  
959 Atlantic Ocean from 42°N to 72°S, *Fresenius. J. Anal. Chem.*, 359(3), 298–305, doi:10.1007/s002160050577, 1997.
- 960 Schneider, B., Nausch, G., Kubsch, H. and Petersohn, I.: Accumulation of total CO<sub>2</sub> during stagnation in the Baltic Sea  
961 deep water and its relationship to nutrient and oxygen concentrations, *Mar. Chem.*, 77, 277–291, 2002.
- 962 Schulz, K. G., Bellerby, R. G. J., Brussaard, C. P. D., Büdenbender, J., Czerny, J., Engel, A., Fischer, M., Koch-  
963 Klavsen, S., Krug, S. A., Lischka, S., Ludwig, A., Meyerhöfer, M., Nondal, G., Silyakova, A., Stühr, A. and Riebesell,  
964 U.: Temporal biomass dynamics of an Arctic plankton bloom in response to increasing levels of atmospheric carbon  
965 dioxide, *Biogeosciences*, 10(1), 161–180, doi:10.5194/bg-10-161-2013, 2013.
- 966 Simó, R., Archer, S. D., Pedros-Alio, C., Gilpin, L. and Stelfox-Widdicombe, C. E.: Coupled dynamics of  
967 dimethylsulfoniopropionate and dimethylsulfide cycling and the microbial food web in surface waters of the North  
968 Atlantic, *Limnol. Oceanogr.*, 47(1), 53–61 [online] Available from: <http://cedadocs.badc.rl.ac.uk/67/> (Accessed 25  
969 March 2014), 2002.
- 970 Simó, R., Vila-Costa, M., Alonso-Sáez, L., Cardelús, C., Guadayol, Ó., Vázquez-Dominguez, E. and Gasol, J. M.:  
971 Annual DMSP contribution to S and C fluxes through phytoplankton and bacterioplankton in a NW Mediterranean  
972 coastal site, *Aquat. Microb. Ecol.*, 57(October), 43–55, doi:10.3354/ame01325, 2009.
- 973 Six, K. D., Kloster, S., Ilyina, T., Archer, S. D., Zhang, K. and Maier-Reimer, E.: Global warming amplified by reduced  
974 sulphur fluxes as a result of ocean acidification, *Nat. Clim. Chang.*, 3(8), 1–4, doi:10.1038/nclimate1981, 2013.
- 975 Smith, D. C., Simon, M., Alldredge, A. L. and Azam, F.: Intense hydrolytic enzyme activity on marine aggregates and  
976 implications for rapid particle dissolution, *Nature*, 359, 139 – 142 [online] Available from:  
977 <http://www.gso.uri.edu/dcsmith/page3/page19/assets/Smithetal92.pdf> (Accessed 11 September 2014), 1992.
- 978 Solomon, S., Garcia, R. R. and Ravishankara, A. R.: On the role of iodine in ozone depletion, *J. Geophys. Res.*,  
979 99(D10), 20491–20499, doi:10.1029/94JD02028, 1994.
- 980 Stal, L. J., Albertano, P., Bergman, B., von Bröckel, K., Gallon, J. R., Hayes, P. K., Sivonen, K. and Walsby, A. E.:  
981 BASIC: Baltic Sea cyanobacteria. An investigation of the structure and dynamics of water blooms of cyanobacteria in  
982 the Baltic Sea—responses to a changing environment, *Cont. Shelf Res.*, 23(17-19), 1695–1714,  
983 doi:10.1016/j.csr.2003.06.001, 2003.
- 984 Stets, E. G., Hines, M. E. and Kiene, R. P.: Thiol methylation potential in anoxic, low-pH wetland sediments and its  
985 relationship with dimethylsulfide production and organic carbon cycling., *FEMS Microbiol. Ecol.*, 47(1), 1–11,  
986 doi:10.1016/S0168-6496(03)00219-8, 2004.
- 987 Sunda, W., Kieber, D. J., Kiene, R. P. and Huntsman, S.: An antioxidant function for DMSP and DMS in marine algae,  
988 *Nature*, 418(6895), 317–320, doi:10.1038/nature00851, 2002.
- 989 Theiler, R., Cook, J. C., Hager, L. P. and Siuda, J. F.: Halohydrocarbon synthesis by bromoperoxidase, *Science.*,  
990 202(December), 1094 – 1096, 1978.
- 991 Turner, S. M., Malin, G., Liss, P. S., Harbour, D. S. and Holligan, P. M.: The seasonal variation of dimethyl sulfide and  
992 dimethylsulfoniopropionate concentrations in nearshore waters, *Limnol. Oceanogr.*, 33(3), 364–375, 1988.
- 993 Urhahn, T. and Ballschmiter, K.: Chemistry of the biosynthesis of halogenated methanes: C1-organohalogenes as pre-  
994 industrial chemical stressors in the environment?, *Chemosphere*, 37(6), 1017–1032, doi:10.1016/S0045-  
995 6535(98)00100-3, 1998.
- 996 del Valle, D. A., Slezak, D., Smith, C. M., Rellinger, A. N., Kieber, D. J. and Kiene, R. P.: Effect of acidification on  
997 preservation of DMSP in seawater and phytoplankton cultures: Evidence for rapid loss and cleavage of DMSP in  
998 samples containing *Phaeocystis* sp., *Mar. Chem.*, 124, 57–67, doi:10.1016/j.marchem.2010.12.002, 2011.
- 999 Vila-Costa, M., Simó, R., Harada, H., Gasol, J. M., Slezak, D. and Kiene, R. P.: Dimethylsulfoniopropionate uptake by  
1000 marine phytoplankton, *Science*, 314(5799), 652–4, doi:10.1126/science.1131043, 2006a.
- 1001 Vila-Costa, M., del Valle, D. A., González, J. M., Slezak, D., Kiene, R. P., Sánchez, O. and Simó, R.: Phylogenetic  
1002 identification and metabolism of marine dimethylsulfide-consuming bacteria, *Environ. Microbiol.*, 8(12), 2189–2200,  
1003 doi:10.1111/j.1462-2920.2006.01102.x, 2006b.
- 1004 Visscher, P. T., Baumgartner, L. K., Buckley, D. H., Rogers, D. R., Hogan, M. E., Raleigh, C. D., Turk, K. A. and Des  
1005 Marais, D. J.: Dimethyl sulphide and methanethiol formation in microbial mats: potential pathways for biogenic

1006 signatures, *Environ. Microbiol.*, 5(4), 296–308 [online] Available from:  
1007 <http://www.ncbi.nlm.nih.gov/pubmed/12662177>, 2003.

1008 Vogt, M., Steinke, M., Turner, S. M., Paulino, A., Meyerhöfer, M., Riebesell, U., LeQuéré, C. and Liss, P. S.:  
1009 Dynamics of dimethylsulphoniopropionate and dimethylsulphide under different CO<sub>2</sub> concentrations during a  
1010 mesocosm experiment, *Biogeosciences*, 5(2), 407–419, doi:10.5194/bg-5-407-2008, 2008.

1011 Wasmund, N.: Occurrence of cyanobacterial blooms in the Baltic Sea in relation to environmental conditions, *Iny. Rev.*  
1012 *ges. Hydrobiol.*, 82(2), 169–184 [online] Available from:  
1013 <http://onlinelibrary.wiley.com/doi/10.1002/iroh.19970820205/full> (Accessed 25 March 2014), 1997.

1014 Webb, A.: The effects of elevated CO<sub>2</sub> and ocean acidification on the production of marine biogenic trace gases, PhD  
1015 Thesis, Univ. East Angl., (March), 2015.

1016 Webb, A. L., Malin, G., Hopkins, F. E., Ho, K.-L., Riebesell, U., Schulz, K., Larsen, A. and Liss, P.: Ocean  
1017 acidification has different effects on the production of dimethylsulphide and dimethylsulphoniopropionate measured in  
1018 cultures of *Emiliania huxleyi* RCC1229 and mesocosm study: a comparison of laboratory monocultures and community  
1019 interactions, *Environ. Chem.*, EN14268, doi:<http://dx.doi.org/10.1071/EN14268>, 2015.

1020 Welschmeyer, N. A.: Fluorometric analysis of chlorophyll a in the presence of chlorophyll b and pheopigments,  
1021 *Limnol. Oceanogr.*, 39(8), 1985–1992, 1994.

1022 Zika, R. G., Gidel, L. T. and Davis, D. D.: A comparison of photolysis and substitution decomposition rates of methyl  
1023 iodide in the ocean, *Geophys. Res. Lett.*, 11(4), 353–356, 1984.

1024 Zinder, S. H., Doemel, W. N. and Brock, T. D.: Production of volatile sulfur compounds during the decomposition of  
1025 algal mats, *Appl. Environ. Microbiol.*, 34(6), 859–861 [online] Available from:  
1026 <http://aem.asm.org/content/34/6/859.short> (Accessed 25 March 2014), 1977.

1027 Zindler, C., Peeken, I., Marandino, C. A. and Bange, H. W.: Environmental control on the variability of DMS and  
1028 DMSP in the Mauritanian upwelling region, *Biogeosciences*, 9, 1041–1051, doi:10.5194/bg-9-1041-2012, 2012.

1029

1030

1031 Table 1. Summary of  $f\text{CO}_2$  and  $\text{pH}_T$  (total scale) during phases 0, 1 and 2 of the mesocosm  
 1032 experiment.

Mesocosm <sup>a</sup>	Target $f\text{CO}_2$ ( $\mu\text{atm}$ )	Whole Experiment ( $t-3$ to $t31$ )		Phase 0 ( $t-3$ to $t0$ )		Phase I ( $t1$ – $t16$ )		Phase II ( $t16$ – $t31$ )	
		Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_T$	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_T$	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_T$	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_T$
<b>M1</b>	Control	331	7.91	231	8.00	328	7.95	399	7.86
<b>M5</b>	Control	334	7.91	244	7.98	329	7.94	399	7.52
<b>M7</b>	390	458	7.80	239	7.99	494	7.81	532	7.76
<b>M6</b>	840	773	7.63	236	7.99	932	7.59	855	7.59
<b>M3</b>	1120	950	7.56	243	7.98	1176	7.51	1027	7.52
<b>M8</b>	1400	1166	7.49	232	8.00	1481	7.43	1243	7.45
<b>Baltic Sea</b>	380	350	7.91	298	7.91	277	7.98	436	7.86

1033 <sup>a</sup> listed in order of increasing  $f\text{CO}_2$

1034

1035 Table 2. Calibration ranges and calculated percentage mean relative standard error for the trace  
1036 gases measured in the mesocosms.

<b>Compound</b>	<b>Calibration range (<math>\mu\text{mol L}^{-1}</math>)</b>	<b>% Mean relative standard error</b>
<b>DMS</b>	600 – 29300*	6.33
<b>DMSP</b>	2030 – 405900*	
<b>CH<sub>3</sub>I</b>	0.11 – 11.2	4.62
<b>CH<sub>2</sub>I<sub>2</sub></b>	5.61 – 561.0	4.98
<b>C<sub>2</sub>H<sub>5</sub>I</b>	0.10 – 4.91	5.61
<b>CH<sub>2</sub>ClI</b>	1.98 – 99.0	3.64
<b>CHBr<sub>3</sub></b>	8.61 – 816.0	4.03
<b>CH<sub>2</sub>Br<sub>2</sub></b>	0.21 – 20.9	5.30
<b>CHBr<sub>2</sub>Cl</b>	0.07 – 7.00	7.20

1037 \* throughout the rest of this paper, these measurements are given in  $\text{nmol L}^{-1}$ .

1038

Table 3. Abundance and contributions of different phytoplankton groups to the total phytoplankton community assemblage, showing the range of measurements from total Chl-*a* (Paul *et al.*, 2015), CHEMTAX analysis of derived Chl-*a* (Paul *et al.*, 2015) and phytoplankton abundance (Crawford *et al.*, 2016). Data are split into the range of all the mesocosm measurements and those from the Baltic Sea.

	Mesocosm			Baltic Sea		
	Range	Range	% Contribution to Chl- <i>a</i>	Range	Range	% Contribution to Chl- <i>a</i>
	Integrated 10 m	Integrated 17 m		Integrated 10 m	Integrated 17 m	
<b>Chl-<i>a</i></b>	0.9 – 2.9	0.9 – 2.6	100	1.3 – 6.5	1.12 – 5.5	100
<b>Phytoplankton Taxonomy (Equivalent Chlorophyll <math>\mu\text{g L}^{-1}</math>)</b>						
<b>Cyanobacteria</b>		0.01 – 0.4	8		0.0 – 0.1	1
<b>Prasinophytes</b>		0.04 – 0.3	7		0.01 – 0.3	4
<b>Euglenophytes</b>		0.0 – 1.6	15		0.0 – 2.6	21
<b>Dinoflagellates</b>		0.0 – 0.3	3		0.04 – 0.6	9
<b>Diatoms</b>		0.1 – 0.3	7		0.04 – 0.9	9
<b>Chlorophytes</b>		0.3 – 2.0	40		0.28 – 3.1	41
<b>Cryptophytes</b>		0.1 – 1.4	21		0.1 – 1.0	15
<b>Small Phytoplankton (&lt;10 <math>\mu\text{m}</math>) abundance <math>\neq</math> (cells <math>\text{mL}^{-1}</math>)</b>						
<b>Cyanobacteria</b>	55000 – 380000	65000 – 470000		30000 – 180000	30000 – 250000	
<b>Picoeukaryotes I</b>	15000 – 100000	17000 – 111000		5000 – 70000	6100 – 78000	
<b>Picoeukaryotes II</b>	700 – 4000	600 – 4000		400 – 3000	460 – 3700	
<b>Picoeukaryotes III</b>	1000 – 9000	1100 – 8500		1000 – 6000	950 – 7500	
<b>Nanoeukaryotes I</b>	400 – 1400	270 – 1500		200 – 4000	210 – 4100	
<b>Nanoeukaryotes II</b>	0 – 400	4 – 400		100 – 1100	60 – 1300	

Table 4. Concentration ranges of trace gases measured in the mesocosms compared to other open water ocean acidification experiments, showing the range of concentrations for each gas and the percentage change between the control and the highest  $f\text{CO}_2$  treatment.

	Range $f\text{CO}_2$		DMS	$\text{CH}_3\text{I}$	$\text{CH}_2\text{I}_2$	$\text{CH}_2\text{ClI}$	$\text{CHBr}_3$	$\text{CH}_2\text{Br}_2$	$\text{CH}_2\text{Br}_2\text{Cl}$
	( $\mu\text{atm}$ )		( $\text{nmol L}^{-1}$ )						
<b>SOPRAN Tvärminne Mesocosm (this study)</b>	346 – 1333	Range	2.7-6.8	2.9-6.4	57-202	3.8-8.0	69-148	4.0-7.7	1.7-3.1
		% change	-34	-0.3	1.3	-11	-9	-3	-4
<b>SOPRAN Bergen 2011 (Webb <i>et al.</i>, 2015)</b>	280 – 3000	Range	0.1-4.9	4.9-32	5.8-321	9.0-123	64-306	6.3-30.8	3.9-14
		% change	-60	-37	-48	-27	-2	-4	-6
<b>NERC Microbial Metagenomics Experiment, Bergen 2006 (Hopkins <i>et al.</i>, 2010)</b>	300 - 750	Range	ND-50	2.0-25	ND-750	ND-700	5.0-80	ND-5.5	0.2-1.2
		% change	-57	-41	-33	-28	13	8	22
<b>EPOCA Svalbard 2010 (Archer <i>et al.</i>, 2013; Hopkins <i>et al.</i>, 2013)</b>	180 - 1420	Range	ND-14	0.04-10	0.01-2.5	0.3-1.6	35-151	6.3-33.3	1.6-4.7
		% change	-60	NS		NS	NS	NS	NS
<b>UKOA European Shelf 2011 (Hopkins and Archer, 2014)</b>	340 - 1000	Range	0.5-12						
		% change	225						
<b>Korean Mesocosm Experiment 2012 (Park <i>et al.</i>, 2014)</b>	160 - 830	Range	1.0-100						
		% change	-82						

Table 5. Concentration ranges of trace gases measured in the Baltic Sea compared to concentrations measured in the literature. ND – Not Detected.

Study	DMS concentration range (nmol L <sup>-1</sup> )	Halocarbon concentration range (pmol L <sup>-1</sup> )						
		CH <sub>3</sub> I	CH <sub>2</sub> I <sub>2</sub>	C <sub>2</sub> H <sub>5</sub> I	CH <sub>3</sub> Cl	CHBr <sub>3</sub>	CH <sub>2</sub> Br <sub>2</sub>	CH <sub>2</sub> Br <sub>2</sub> Cl
<b>SOPRAN Tvärminne Baltic Sea (This Study)</b>	1.9-11	4.3-8.6	66.9-374	0.6 – 1.0	7.0-18	93-192	7.1-10	3.3-5.0
<b>Orlikowska and Schulz-Bull (2009)</b>	0.3-120	1-16	0-85	0.4 – 1.2	5-50	5.0-40	2.0-10	0.8-2.5
<b>Karlsson <i>et al.</i> (2008)</b>		3.0-7.5				35-60	4.0-7.0	2.0-6.5
<b>Klick and Abrahamsson (1992)</b>			15-709		11-74	14-585		
<b>Klick (1992)</b>			ND-243		ND-57	40-790	ND-86	ND-29
<b>Leck and Rodhe (1991)</b>	0.4-2.8							
<b>Leck <i>et al.</i> (1990)</b>	ND-3.2							

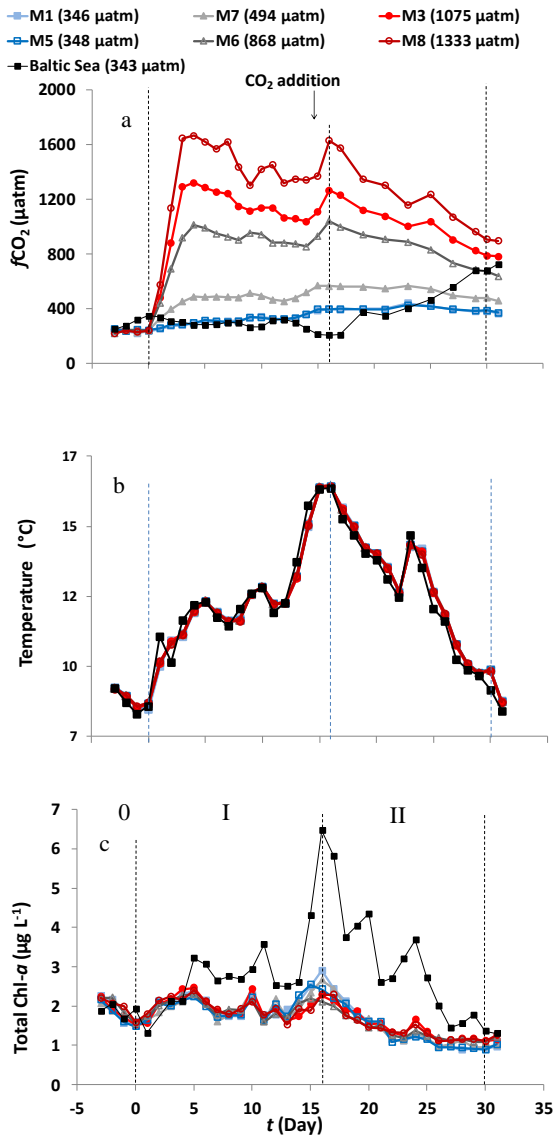


Figure 1. Daily measurements of (a)  $f\text{CO}_2$ , (b) mean temperature and (c) total Chlorophyll- $a$  in the mesocosms and surrounding Baltic Sea waters. Dashed lines represent the three Phases of the experiment, based on the Chl- $a$  data.



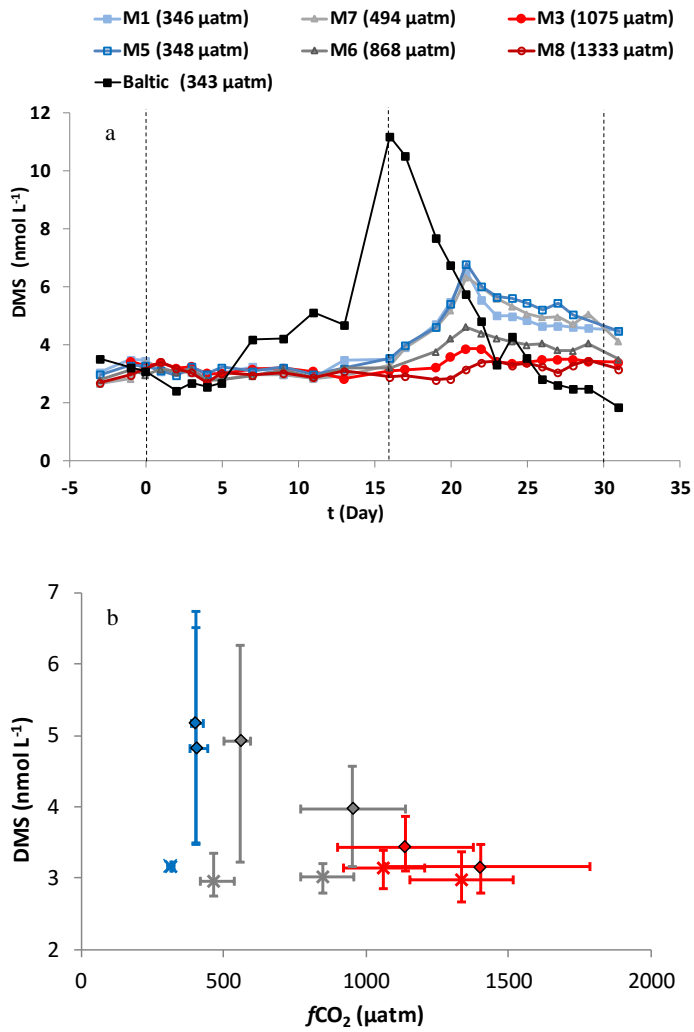


Figure 32. (a) Mean DMS concentrations measured daily in the mesocosms and Baltic Sea from an integrated water sample of the surface 10 m. Dashed lines show the Phases of the experiment as given in Fig. 12,  $f\text{CO}_2$  shown in the legend are mean  $f\text{CO}_2$  across the duration of the experiment. (b) Mean DMS concentrations from each mesocosm during Phase I (crosses) and Phase II (diamonds), for ambient (blue), medium (grey) and high  $f\text{CO}_2$  (red), with error bars showing the range of both the DMS and  $f\text{CO}_2$ .

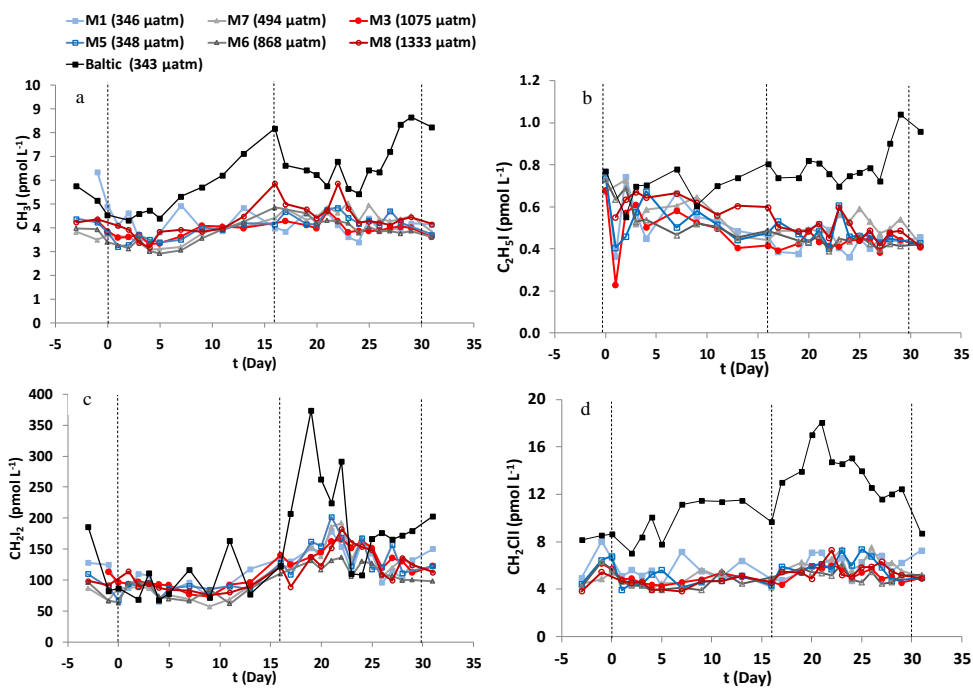


Figure 3. Mean  $\text{C}_i$  concentrations (pmol L<sup>-1</sup>) of (a) CH<sub>3</sub>I, (b) C<sub>2</sub>H<sub>5</sub>I, (c) CH<sub>2</sub>I<sub>2</sub> and (d) CH<sub>2</sub>ClI taken from a water sample integrated from the surface 10m. Dashed lines indicate the Phases of the experiment, as given in Fig. 2.  $f\text{CO}_2$  shown in the legend are mean  $f\text{CO}_2$  across the duration of the experiment.

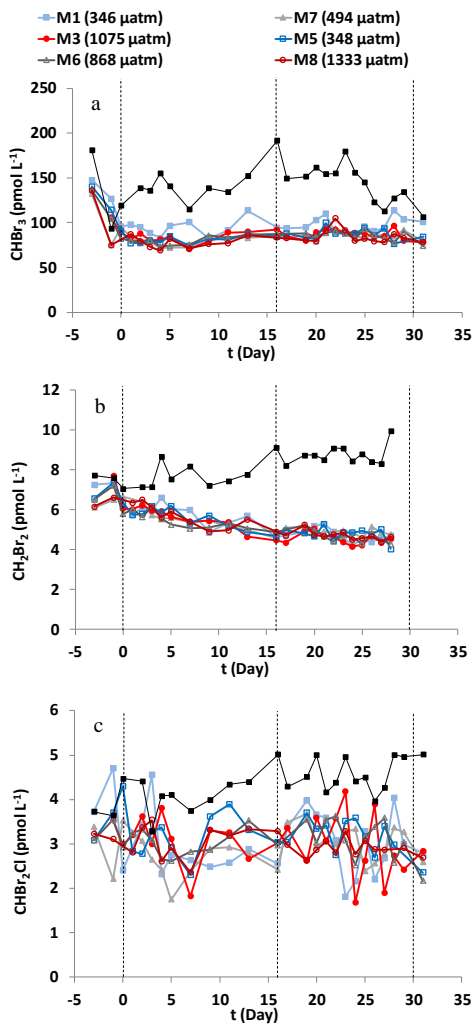


Figure 54. Mean concentrations (pmol L<sup>-1</sup>) of (a) CHBr<sub>3</sub>, (b) CH<sub>2</sub>Br<sub>2</sub> and (c) CHBr<sub>2</sub>Cl taken from a water sample integrated from the surface 10m. Dashed lines indicate the phases of the experiment as defined in Fig. 2,  $f\text{CO}_2$  shown in the legend are mean  $f\text{CO}_2$  across the duration of the experiment.