

# Effect of ocean acidification and elevated $f\text{CO}_2$ on trace gas production by a Baltic Sea summer phytoplankton community

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## Abstract

The Baltic Sea is a unique environment as the largest body of brackish water in the world. Acidification of the surface oceans due to absorption of anthropogenic  $\text{CO}_2$  emissions is an additional stressor facing the pelagic community of the already challenging Baltic Sea. To investigate its impact on trace gas biogeochemistry, a large-scale mesocosm experiment was performed off Tvärminne Research Station, Finland in summer 2012. During the second half of the experiment, dimethylsulphide (DMS) concentrations in the highest  $f\text{CO}_2$  mesocosms

(1075 - 1333  $\mu\text{atm}$ ) were 34% lower than at ambient  $\text{CO}_2$  (350  $\mu\text{atm}$ ). However, the net production (as measured by concentration change) of seven halocarbons analysed was not significantly affected by even the highest  $\text{CO}_2$  levels after 5 weeks' exposure. Methyl iodide ( $\text{CH}_3\text{I}$ ) and diiodomethane ( $\text{CH}_2\text{I}_2$ ) showed 15% and 57% increases in mean mesocosm concentration ( $3.8 \pm 0.6 \text{ pmol L}^{-1}$  increasing to  $4.3 \pm 0.4 \text{ pmol L}^{-1}$  and  $87.4 \pm 14.9 \text{ pmol L}^{-1}$  increasing to  $134.4 \pm 24.1 \text{ pmol L}^{-1}$  respectively) during Phase II of the experiment, which were unrelated to  $\text{CO}_2$  and corresponded to 30% lower Chl-*a* concentrations compared to Phase I. No other iodocarbons increased or showed a peak, with mean chloriodomethane ( $\text{CH}_2\text{ClI}$ ) concentrations measured at  $5.3 (\pm 0.9) \text{ pmol L}^{-1}$  and iodoethane ( $\text{C}_2\text{H}_5\text{I}$ ) at  $0.5 (\pm 0.1) \text{ pmol L}^{-1}$ . Of the concentrations of bromoform ( $\text{CHBr}_3$ ; mean  $88.1 \pm 13.2 \text{ pmol L}^{-1}$ ), dibromomethane ( $\text{CH}_2\text{Br}_2$ ; mean  $5.3 \pm 0.8 \text{ pmol L}^{-1}$ ) and dibromochloromethane ( $\text{CHBr}_2\text{Cl}$ , mean  $3.0 \pm 0.5 \text{ pmol L}^{-1}$ ), only  $\text{CH}_2\text{Br}_2$  showed a decrease of 17% between Phases I and II, with  $\text{CHBr}_3$  and  $\text{CHBr}_2\text{Cl}$  showing similar mean concentrations in both Phases. Outside the mesocosms, an upwelling event was responsible for bringing colder, high  $\text{CO}_2$ , low pH water to the surface starting on day *t*16 of the experiment; this variable  $\text{CO}_2$  system with frequent upwelling events implies the community of the Baltic Sea is acclimated to regular significant declines in pH caused by up to 800  $\mu\text{atm } f\text{CO}_2$ . After this upwelling, DMS concentrations declined, but halocarbon concentrations remained similar or increased compared to measurements prior to the change in conditions. Based on our findings, with future acidification of Baltic Sea waters, biogenic halocarbon emissions are likely to remain at similar values to today, however emissions of biogenic sulphur could significantly decrease from this region.

## 1 Introduction

Anthropogenic activity has increased the fugacity of atmospheric carbon dioxide ( $f\text{CO}_2$ ) from 280  $\mu\text{atm}$  (pre-Industrial Revolution) to over 400  $\mu\text{atm}$  today (Hartmann *et al.*, 2013). The IPCC AR5 long-term projections for atmospheric  $p\text{CO}_2$  and associated changes to the climate have been established for a variety of scenarios of anthropogenic activity until the year 2300. As the largest global sink for atmospheric  $\text{CO}_2$ , the global ocean has absorbed an estimated 30% of excess  $\text{CO}_2$  produced (Canadell *et al.*, 2007). With atmospheric  $p\text{CO}_2$  projected to possibly exceed 2000  $\mu\text{atm}$  by the year 2300 (Collins *et al.*, 2013; Cubasch *et al.*, 2013), the ocean will take up increasing amounts of  $\text{CO}_2$ , with a potential lowering of surface ocean pH by over 0.8 units (Raven *et al.*, 2005). The overall effect of acidification on the biogeochemistry of surface ocean ecosystems is

62 unknown and currently unquantifiable, with a wide range of potential positive and negative impacts  
63 (Doney *et al.*, 2009; Hofmann *et al.*, 2010; Ross *et al.*, 2011).

64 A number of volatile organic compounds are produced by marine phytoplankton (Liss *et al.*, 2014),  
65 including the climatically important trace gas dimethylsulphide (DMS, C<sub>2</sub>H<sub>6</sub>S) and a number of  
66 halogen-containing organic compounds (halocarbons) including methyl iodide (CH<sub>3</sub>I) and  
67 bromoform (CHBr<sub>3</sub>). These trace gases are a source of sulphate particles and halide radicals when  
68 oxidised in the atmosphere, and have important roles as ozone catalysts in the troposphere and  
69 stratosphere (O'Dowd *et al.*, 2002; Solomon *et al.*, 1994) and as cloud condensation nuclei (CCNs;  
70 Charlson *et al.*, 1987).

71 DMS is found globally in surface waters originating from the algal-produced precursor  
72 dimethylsulphonioacetate (DMSP, C<sub>5</sub>H<sub>10</sub>O<sub>2</sub>S). Both DMS and DMSP provide the basis for  
73 major routes of sulphur and carbon flux through the marine microbial food web, and can provide up  
74 to 100% of the bacterial and phytoplanktonic sulphur demand (Simó *et al.*, 2009; Vila-Costa *et al.*,  
75 2006a). DMS is also a volatile compound which readily passes through the marine boundary layer  
76 to the troposphere, where oxidation results in a number of sulphur-containing particles important for  
77 atmospheric climate feedbacks (Charlson *et al.*, 1987; Quinn and Bates, 2011); for this reason, any  
78 change in the production of DMS may have significant implications for climate regulation. Several  
79 previous acidification experiments have shown differing responses of both compounds (e.g.  
80 Avgoustidi *et al.*, 2012; Hopkins *et al.*, 2010; Webb *et al.*, 2015), while others have shown delayed  
81 or more rapid responses as a direct effect of CO<sub>2</sub> (e.g. Archer *et al.*, 2013; Vogt *et al.*, 2008).  
82 Further, some laboratory incubations of coastal microbial communities showed increased DMS  
83 production with increased *f*CO<sub>2</sub> (Hopkins and Archer, 2014), but lower DMSP production. The  
84 combined picture arising from existing studies is that the response of communities to *f*CO<sub>2</sub>  
85 perturbation is not predictable and requires further study. Previous studies measuring DMS in the  
86 Baltic Sea measured concentrations up to 100 nmol L<sup>-1</sup> during the summer bloom, making the  
87 Baltic Sea a significant source of DMS (Orlikowska and Schulz-Bull, 2009).

88 In surface waters, halocarbons such as methyl iodide (CH<sub>3</sub>I), chloriodomethane (CH<sub>2</sub>ClI) and  
89 bromoform (CHBr<sub>3</sub>) are produced by biological and photochemical processes: many marine  
90 microbes (for example cyanobacteria; Hughes *et al.*, 2011, diatoms; Manley and De La Cuesta,  
91 1997 and haptophytes; Scarratt and Moore, 1998) and macroalgae (e.g. brown-algal *Fucus* species;  
92 Chance *et al.*, 2009 and red algae; Leedham *et al.*, 2013) utilise halides from seawater and emit a  
93 range of organic and inorganic halogenated compounds. This production can lead to significant flux  
94 to the marine boundary layer in the order of 10 Tg iodine-containing compounds ('iodocarbons';

95 O'Dowd *et al.*, 2002) and 1 Tg bromine-containing compounds ('bromocarbons'; Goodwin *et al.*,  
96 1997) into the atmosphere. The effect of acidification on halocarbon concentrations has received  
97 limited attention, but two acidification experiments measured lower concentrations of several  
98 iodocarbons while bromocarbons were unaffected by  $f\text{CO}_2$  up to 3000  $\mu\text{atm}$  (Hopkins *et al.*, 2010;  
99 Webb, 2015), whereas an additional mesocosm study did not elicit significant differences from any  
100 compound up to 1400  $\mu\text{atm } f\text{CO}_2$  (Hopkins *et al.*, 2013).

101 Measurements of the trace gases within the Baltic Sea are limited, with no prior study of DMSP  
102 concentrations in the region. The Baltic Sea is the largest body of brackish water in the world, and  
103 salinity ranges from 1 to 15. Furthermore, seasonal temperature variations of over 20 °C are  
104 common. A permanent halocline at 50-80 m separates  $\text{CO}_2$ -rich, bottom waters from fresher, lower  
105  $\text{CO}_2$  surface waters, and a summer thermocline at 20 m separates warmer surface waters from those  
106 below 4°C (Janssen *et al.*, 1999). Upwelling of bottom waters from below the summer thermocline  
107 is a common summer occurrence, replenishing the surface nutrients while simultaneously lowering  
108 surface temperature and pH (Brutemark *et al.*, 2011). Baltic organisms are required to adapt to  
109 significant variations in environmental conditions. The species assemblage in the Baltic Sea is  
110 different to those studied during previous mesocosm experiments in the Arctic, North Sea and  
111 Korea (Brussaard *et al.*, 2013; Engel *et al.*, 2008; Kim *et al.*, 2010), and are largely unstudied in  
112 terms of their community trace gas production during the summer bloom. Following the spring  
113 bloom (July-August), a low dissolved inorganic nitrogen (DIN) to dissolved inorganic phosphorous  
114 (DIP) ratio combines with high temperatures and light intensities to encourage the growth of  
115 heterocystous cyanobacteria, (Niemisto *et al.*, 1989; Raateoja *et al.*, 2011), in preference to nitrate-  
116 dependent groups.

117 Here we report the concentrations of DMS, DMSP and halocarbons from the 2012 summer post-  
118 bloom season mesocosm experiment aimed to assess the impact of elevated  $f\text{CO}_2$  on the microbial  
119 community and trace gas production in the Baltic Sea. Our objective was to assess how changes in  
120 the microbial community driven by changes in  $f\text{CO}_2$  impacted DMS and halocarbon concentrations.  
121 It is anticipated that any effect of  $\text{CO}_2$  on the growth of different groups within the phytoplankton  
122 assemblage will result in an associated change in trace gas concentrations measured in the  
123 mesocosms as  $f\text{CO}_2$  increases, which can potentially be used to predict future halocarbon and  
124 sulphur emissions from the Baltic Sea region.

125

## 126 2 Methods

### 127 2.1 Mesocosm design and deployment

128 Nine mesocosms were deployed on the 10th June 2012 (day  $t-10$ ; days are numbered negative prior  
129 to CO<sub>2</sub> addition and positive afterward) and moored near Tvärminne Zoological Station (59° 51.5'  
130 N, 23° 15.5' E) in Tvärminne Storfjärden in the Baltic Sea. Each mesocosm comprised a  
131 thermoplastic polyurethane (TPU) enclosure of 17 m depth, containing approximately 54,000 L of  
132 seawater, supported by an 8m tall floating frame capped with a polyvinyl hood. For full technical  
133 details of the mesocosms see Czerny *et al.* (2013) and Riebesell *et al.* (2013). The mesocosm bags  
134 were filled by lowering through the stratified water column until fully submerged, with the opening  
135 at both ends covered by 3 mm mesh to exclude organisms larger than 3 mm such as fish and large  
136 zooplankton. The mesocosms were then left for 3 days ( $t-10$  to  $t-7$ ) with the mesh in position to  
137 allow exchange with the external water masses and ensure the mesocosm contents were  
138 representative of the phytoplankton community in the Storfjärden. On  $t-7$  the bottom of the  
139 mesocosm was sealed with a sediment trap and the upper opening was raised to approximately 1.5  
140 m above the water surface. Stratification within the mesocosm bags was broken up on  $t-5$  by the use  
141 of compressed air for three and a half minutes to homogenise the water column and ensure an even  
142 distribution of inorganic nutrients at all depths. Unlike in previous experiments, there was no  
143 addition of inorganic nutrients to the mesocosms at any time during the experiment; mean inorganic  
144 nitrate, inorganic phosphate and ammonium concentrations measured across all mesocosms at the  
145 start of the experiment were 37.2 ( $\pm$  18.8 s.d.) nmol L<sup>-1</sup>, 323.9 ( $\pm$  19.4 s.d.) nmol L<sup>-1</sup> and 413.8 ( $\pm$   
146 319.5 s.d.) nmol L<sup>-1</sup> respectively.

147 To obtain mesocosms with different  $f\text{CO}_2$ , the carbonate chemistry of the mesocosms was altered  
148 by the addition of different volumes of 50  $\mu\text{m}$  filtered, CO<sub>2</sub>-enriched Baltic Sea water (sourced from  
149 outside the mesocosms), to each mesocosm over a four-day period, with the first day of addition  
150 being defined as day  $t0$ . Addition of the enriched CO<sub>2</sub> water was by the use of a bespoke dispersal  
151 apparatus ('Spider') lowered through the bags to ensure even distribution throughout the water  
152 column (further details are in Riebesell *et al.* 2013). Measurements of salinity in the mesocosms  
153 throughout the experiment determined that three of the mesocosms were not fully sealed, and had  
154 undergone unquantifiable water exchange with the surrounding waters. These three mesocosms  
155 (M2, M4 and M9) were excluded from the analysis. Two mesocosms were designated as controls  
156 (M1 and M5) and received only filtered seawater via the Spider; four mesocosms received addition  
157 of CO<sub>2</sub>-enriched waters, with the range of target  $f\text{CO}_2$  levels between 600 and 1650  $\mu\text{atm}$  (M7, 600

158  $\mu\text{atm}$ ; M6, 950  $\mu\text{atm}$ ; M3, 1300  $\mu\text{atm}$ ; M8 1650  $\mu\text{atm}$ ). Mesocosms were randomly allocated a  
159 target  $f\text{CO}_2$ ; a noticeable decrease in  $f\text{CO}_2$  was identified in the three highest  $f\text{CO}_2$  mesocosms (M6,  
160 M3 and M8) over the first half of the experiment, which required the addition of more  $\text{CO}_2$  enriched  
161 water on  $t15$  to bring the  $f\text{CO}_2$  back up to maximum concentrations (Fig. 1a; Paul *et al.*, 2015). A  
162 summary of the  $f\text{CO}_2$  in the mesocosms can be seen in Table 1. At the same time as this further  $\text{CO}_2$   
163 addition on  $t15$ , the walls of the mesocosms were cleaned using a bespoke wiper apparatus (See  
164 Riebesell *et al.*, 2013 for more information), followed by weekly cleaning to remove aggregations  
165 on the film which would block incoming light. Light measurements showed that over 95% of the  
166 photosynthetically active radiation (PAR) was transmitted by the clean TPU and PVC materials  
167 with 100% absorbance of UV light (Riebesell *et al.*, 2013). Samples for most parameters were  
168 collected from the mesocosms at the same time every morning from  $t-3$ , and analysed daily or every  
169 other day.

## 170 2.2 Trace gas extraction and analysis

### 171 2.2.1 DMS and halocarbons

172 A depth-integrated water sampler (IWS, HYDRO-BIOS, Kiel, Germany) was used to sample the  
173 entire 17 m water column daily or alternative daily. As analysis of Chlorophyll-*a* (Chl-*a*) showed it  
174 to be predominantly produced in the first 10 m of the water column, trace gas analysis was  
175 conducted on only integrated samples collected from the surface 10 m, with all corresponding  
176 community parameter analyses with the exception of pigment analysis performed also to this depth.  
177 Water samples for trace gas analysis were taken from the first IWS from each mesocosm to  
178 minimise the disturbance and bubble entrainment from taking multiple samples in the surface  
179 waters. As in Hughes *et al.* (2009), samples were collected in 250 mL amber glass bottles in a  
180 laminar flow with minimal disturbance to the water sample, using Tygon tubing from the outlet of  
181 the IWS. Bottles were rinsed twice before being carefully filled from the bottom with minimal  
182 stirring, and allowed to overflow the volume of the bottle approximately three times before sealing  
183 with a glass stopper to prevent bubble formation and atmospheric contact. Samples were stored  
184 below 10°C in the dark for 2 hours prior to analysis. Each day, a single sample was taken from each  
185 mesocosm, with two additional samples taken from one randomly selected mesocosm to evaluate  
186 the precision of the analysis (<4%, no further data shown).

187 On return to the laboratory, 40 mL of water was injected into a purge and cryotrap system (Chuck *et al.*  
188 *et al.*, 2005), filtered through a 25 mm Whatman glass fibre filter (GF/F; GE Healthcare Life Sciences,  
189 Little Chalfont, England) and purged with oxygen-free nitrogen (OFN) at 80 mL min<sup>-1</sup> for 10

minutes. Each gas sample passed through a glass wool trap to remove particles and aerosols, before a dual nafion counterflow drier (180 mL min<sup>-1</sup> OFN) removed water vapour from the gas stream. The gas sample was trapped in a stainless steel loop held at -150 °C in the headspace of a liquid nitrogen-filled dewar. The sample was injected by immersion of the sample loop in boiling water into an Agilent 6890 gas chromatograph equipped with a 60 m DB-VRX capillary column (0.32 mm ID, 1.8 µm film thickness, Agilent J&W Ltd) according to the programme outlined by Hopkins *et al.* (2010). Analysis was performed by an Agilent 5973 quadrupole mass spectrometer operated in electron ionisation, single ion mode. Liquid standards of CH<sub>3</sub>I, diiodomethane (CH<sub>2</sub>I<sub>2</sub>), CH<sub>2</sub>ClI, iodoethane (C<sub>2</sub>H<sub>5</sub>I), iodopropane (C<sub>3</sub>H<sub>7</sub>I), CHBr<sub>3</sub>, dibromoethane (CH<sub>2</sub>Br<sub>2</sub>), dibromochloromethane (CHBr<sub>2</sub>Cl), bromiodomethane (CH<sub>2</sub>BrI) and DMS (Standards supplied by Sigma Aldrich Ltd, UK) were gravimetrically prepared by dilution in HPLC-grade methanol (Table 2) and used for calibration. The relative standard error was expressed as a percentage of the mean for the sample analysis, calculated for each compound using triplicate analysis each day from a single mesocosm, and was <7% for all compounds. GC-MS instrument drift was corrected by the use of a surrogate analyte standard in every sample, comprising deuterated DMS (D<sub>6</sub>-DMS), deuterated methyl iodide (CD<sub>3</sub>I) and <sup>13</sup>C dibromoethane (<sup>13</sup>C<sub>2</sub>H<sub>4</sub>Br<sub>2</sub>) via the method described in Hughes *et al.* (2006) and Martino *et al.* (2005). Five-point calibrations were performed weekly for each compound with the addition of the surrogate analyte, with a single standard analysed daily to check for instrument drift; linear regression from calibrations typically produced r<sup>2</sup>>0.98. All samples measured within the mesocosms were within the concentration ranges of the calibrations (Table 2).

## 2.2.2 DMSP

Samples for total DMSP (DMSP<sub>T</sub>) were collected and stored for later analysis by the acidification method of Curran *et al.* (1998). A 7 mL sub-sample was collected from the amber glass bottle into an 8 mL glass sample vial (Labhut, Churcham, UK), into which 0.35 µL of 50% H<sub>2</sub>SO<sub>4</sub> was added, before storage at ambient temperature. Particulate DMSP (DMSP<sub>P</sub>) samples were prepared by the gravity filtration of 20 mL of sample through a 47 mm GF/F in a glass filter unit, before careful removal and folding of the GF/F into a 7 mL sample vial filled with 7 mL of Milli-Q water and 0.35 µL of H<sub>2</sub>SO<sub>4</sub> before storage at ambient temperature. Samples were stored for approximately 8 weeks prior to analysis. DMSP samples (total and particulate) were analysed on a PTFE purge and cryotrap system using 2 mL of the sample purged with 1 mL of 10M NaOH for 5 minutes at 80 mL min<sup>-1</sup>. The sample gas stream passed through a glass wool trap and Nafion counterflow (Permapure) drier before being trapped in a PTFE sample loop kept at -150 °C by suspension in the headspace of a liquid nitrogen-filled dewar and controlled by feedback from a thermocouple. Immersion in

223 boiling water rapidly re-volatilised the sample for injection into a Shimadzu GC2010 gas  
224 chromatograph with a Varian Chrompack CP-Sil-5CB column (30 m, 0.53 mm ID) and flame  
225 photometric detector (FPD). The GC oven was operated isothermally at 60 °C which resulted in  
226 DMS eluting at 2.1 minutes. Liquid DMSP standards were prepared and purged in the same manner  
227 as the sample to provide weekly calibrations of the entire analytical system. Involvement in the  
228 2013 AQA 12-23 international DMS analysis proficiency test (National Measurement Institute of  
229 Australia, 2013) in February 2013 demonstrated excellent agreement between our method of DMSP  
230 analysis and the mean from thirteen laboratories measuring DMS using different methods, with a  
231 measurement error of 5%.

232 DMSP was not detected in any of the samples (total or particulate) collected and stored during the  
233 experiment, and it was considered likely that this was due to an unresolved issue regarding  
234 acidifying the samples for later DMSP analysis. It was considered unlikely that rates of bacterial  
235 DMSP turnover through demethylation rather than through cleavage to produce DMS (Curson *et al.*, 2011)  
236 were sufficiently high in the Baltic Sea to remove all detectable DMSP, yet still produce  
237 measureable DMS concentrations. Also, rapid turnover of dissolved DMSP in surface waters being  
238 the cause of low DMSP<sub>T</sub> concentrations does not explain the lack of intracellular particulate-phase  
239 DMSP. Although production of DMS is possible from alternate sources, it is highly unlikely that  
240 there was a total absence of DMSP-producing phytoplankton within the mesocosms or Baltic Sea  
241 surface waters around Tvärminne; DMSP has been measured in surface waters of the Southern  
242 Baltic Sea at 22.2 nmol L<sup>-1</sup> in 2012, indicating that DMSP-producing species are present within the  
243 Baltic Sea (Cathleen Zindler, GEOMAR, Pers. Comm.).

244 A previous study by del Valle *et al.* (2011) highlighted up to 94% loss of DMSP<sub>T</sub> from acidified  
245 samples of colonial *Phaeocystis globosa* culture, and field samples dominated by colonial  
246 *Phaeocystis antarctica*. Despite filamentous, colonial cyanobacteria in the samples from Tvärminne  
247 mesocosms potentially undergoing the same process, these species did not dominate the community  
248 at only 6.6% of the total Chl-*a*, implying that the acidification method for DMSP fixation also failed  
249 for unicellular phytoplankton species. This suggests that the acidification method is unreliable in the  
250 Baltic Sea, and should be considered inadequate as the sole method of DMSP fixation in future  
251 experiments in the region. The question of its applicability in other marine waters also needs further  
252 investigation.

253



## 2.3 Measurement of carbonate chemistry and community dynamics

Water samples were collected from the 10 m and 17 m IWS on a daily basis and analysed for carbonate chemistry, fluorometric Chl-*a*, phytoplankton pigments (17 m IWS only) and cell abundance to analyse the community structure and dynamics during the experiment. The carbonate system was analysed through a suite of measurements (Paul *et al.*, 2015), including potentiometric titration for total alkalinity (TA), infrared absorption for dissolved inorganic carbon (DIC) and spectrophotometric determination for pH. For Chl-*a* analysis and pigment determination, 500 mL sub-samples were filtered through a GF/F and stored frozen (-20 °C for two hours for Chl-*a* and -80 °C for up to 6 months for pigments), before homogenisation in 90 % acetone with glass beads. After centrifuging (10 minutes at 800 x g at 4 °C) the Chl-*a* concentrations were determined using a Turner AU-10 fluorometer by the methods of Welschmeyer (1994), and the phytoplankton pigment concentrations by reverse phase high performance liquid chromatography (WATERS HPLC with a Varian Microsorb-MV 100-3 C8 column) as described by Barlow *et al.* (1997). Phytoplankton community composition was determined by the use of the CHEMTAX algorithm to convert the concentrations of marker pigments to Chl-*a* equivalents (Mackey *et al.*, 1996; Schulz *et al.*, 2013). Microbes were enumerated using a Becton Dickinson FACSCalibur flow cytometer (FCM) equipped with a 488 nm argon laser (Crawford *et al.*, 2016) and counts of phytoplankton cells >20 µm were made on concentrated (50 mL) sample water, fixed with acidic Lugol's iodine solution with an inverted microscope. Filamentous cyanobacteria were counted in 50 µm length units.

## 2.4 Statistical Analysis

All statistical analysis was performed using Minitab V16. In analysis of the measurements between mesocosms, one-way ANOVA was used with Tukey's post-hoc analysis test to determine the effect of different  $f\text{CO}_2$  on concentrations measured in the mesocosms and the Baltic Sea ( $H_0$  assumes no significant difference in the mean concentrations of trace gases measured through the duration of the experiment). Spearman's Rank Correlation Coefficients were calculated to compare the relationships between trace gas concentrations,  $f\text{CO}_2$ , and a number of biological parameters, and the resulting  $p$ -values for each correlation are given in Supplementary table S1 for the mesocosms and S2 for the Baltic Sea data.

## 283 3 Results and Discussion

### 284 3.1 Biogeochemical changes within the mesocosms

285 The mesocosm experiment was split into three phases based on the temporal variation in Chl-*a* (Fig.  
286 2; Paul *et al.*, 2015) evaluated after the experiment was completed:

- 287 • Phase 0 (days  $t-5$  to  $t0$ ) – pre-CO<sub>2</sub> addition
- 288 • Phase I (days  $t1$  to  $t16$ ) – ‘productive phase’
- 289 • Phase II (days  $t17$  to  $t30$ ) – temperature induced autotrophic decline.

#### 290 3.1.1 Physical Parameters

291  $f\text{CO}_2$  decreased over Phase I in the three highest  $f\text{CO}_2$  mesocosms, mainly through air-sea gas  
292 exchange and carbon fixation by phytoplankton (Fig. 1a). All mesocosms still showed distinct  
293 differences in  $f\text{CO}_2$  levels throughout the experiment (Table 1), and there was no overlap of  
294 mesocosm  $f\text{CO}_2$  values on any given day, save for the two controls (M1 and M5). The control  
295 mesocosm  $f\text{CO}_2$  increased through Phase I of the experiment, likely as a result of undersaturation of  
296 the water column encouraging dissolution of atmospheric CO<sub>2</sub> (Paul *et al.*, 2015). Salinity in the  
297 mesocosms remained constant throughout the experiment at  $5.70 \pm 0.004$ , and showed no variation  
298 with depth (data not shown but available in Paul *et al.* 2015). It remained similar to salinity in the  
299 Baltic Sea surrounding the mesocosms, which was  $5.74 \pm 0.14$ . Water temperature varied from a  
300 low of  $8.6 \pm 0.4$  °C during Phase 0 to a high of  $15.9 \pm 2.2$  °C measured on day  $t16$ , before  
301 decreasing once again (Fig. 1b).

302 Summertime upwelling events are common and well described (Gidhagen, 1987; Lehmann and  
303 Myrberg, 2008), and induce a significant temperature decrease in surface waters; such an event  
304 appears to have commenced around  $t16$ , as indicated by significantly decreasing temperatures  
305 inside and out of the mesocosms (Fig. 1b) and increased salinity in the Baltic Sea from 5.5 to 6.1  
306 over the following 15 days to the end of the experiment. Due to the enclosed nature of the  
307 mesocosms, the upwelling affected only the temperature and not pH,  $f\text{CO}_2$  or the microbial  
308 community. However, the temperature decrease after  $t16$  was likely to have had a significant effect  
309 on phytoplankton growth (and biogenic gas production), explaining the lower Chl-*a* in Phase II.

#### 310 3.1.2 Community Dynamics

311 Mixing of the mesocosms and redistribution of the nutrients throughout the water column after  
312 closure (prior to  $t-3$ ) did not trigger a notable increase in total Chl-*a* in Phase 0 as was identified in

313 previous mesocosm experiments. During Phase I, light availability, combined with increasing water  
 314 temperatures favoured the growth of phytoplankton in all mesocosms (Paul *et al.* 2015), and was  
 315 unlikely to be a direct result of the CO<sub>2</sub> enrichment, as no difference was identified between  
 316 enriched mesocosms and controls. Mean Chl-*a* during Phase I was 1.98 ( $\pm$  0.29)  $\mu\text{g L}^{-1}$  from all  
 317 mesocosms, decreasing to 1.44 ( $\pm$  0.46)  $\mu\text{g L}^{-1}$  in Phase II: this decrease was attributed to a  
 318 temperature induced decrease in phytoplankton growth rates and higher grazing rates as a result of  
 319 higher zooplankton reproduction rates during Phase I (Lischka *et al.*, 2015; Paul *et al.*, 2015).  
 320 Mesocosm Chl-*a* decreased until the end of the experiment on *t*31.

321 The largest contributors to Chl-*a* in the mesocosms during the summer of 2012 were the  
 322 chlorophytes and cryptophytes, with up to 40% and 21% contributions to the Chl-*a* respectively  
 323 (Table 3; Paul *et al.*, 2015). Significant long-term differences in abundance between mesocosms  
 324 developed as a result of elevated *f*CO<sub>2</sub> in only two groups: picoeukaryotes I showed higher  
 325 abundance at high *f*CO<sub>2</sub> ( $F=8.2$ ,  $p<0.01$ ; Crawford *et al.*, 2016 and Supplementary Fig. S2), as seen  
 326 in previous mesocosm experiments (Brussaard *et al.*, 2013; Newbold *et al.*, 2012) and  
 327 picoeukaryotes III the opposite trend ( $F=19.6$ ,  $p<0.01$ ; Crawford *et al.*, 2016). Temporal variation in  
 328 phytoplankton abundance was similar between all mesocosms (Supplementary Fig. S1 and S2).

329 Diazotrophic, filamentous cyanobacterial blooms in the Baltic Sea are an annual event in summer  
 330 (Finni *et al.*, 2001), and single-celled cyanobacteria have been found to comprise as much as 80%  
 331 of the cyanobacterial biomass and 50% of the total primary production during the summer in the  
 332 Baltic Sea (Stal *et al.*, 2003). However, CHEMTAX analysis identified cyanobacteria as  
 333 contributing less than 10% of the total Chl-*a* in the mesocosms (Crawford *et al.*, 2016; Paul *et al.*,  
 334 2015). These observations were backed up by satellite observations showing reduced cyanobacterial  
 335 abundance throughout the Baltic Sea in 2012 compared to previous and later years (Oberg, 2013). It  
 336 was proposed that light availability and surface water temperatures during the summer of 2012 were  
 337 sub-optimal for triggering a filamentous cyanobacteria bloom (Wasmund, 1997).

## 338 **3.2 DMS and DMSP**

### 339 **3.2.1 Mesocosm DMS**

340 A significant 34% reduction in DMS concentrations was detected in the high *f*CO<sub>2</sub> treatments  
 341 during Phase II compared to the ambient *f*CO<sub>2</sub> mesocosms ( $F=31.7$ ,  $p<0.01$ ). Mean DMS  
 342 concentrations of 5.0 ( $\pm$  0.8; range 3.5 – 6.8) nmol L<sup>-1</sup> in the ambient treatments compared to 3.3 ( $\pm$   
 343 0.3; range 2.9 – 3.9) nmol L<sup>-1</sup> in the 1333 and 1075  $\mu\text{atm}$  mesocosms (Fig. 2a). The primary

344 differences identified were apparent from the start of Phase II on *t*17, after which maximum  
345 concentrations were observed in the ambient mesocosms on *t*21. The relationship between DMS  
346 and increasing *f*CO<sub>2</sub> during Phase II was found to be linear (Fig. 2b), a finding also identified in  
347 previous mesocosm experiments (Archer *et al.*, 2013; Webb *et al.*, 2015). Furthermore, increases in  
348 DMS concentrations under high *f*CO<sub>2</sub> were delayed by three days relative to the ambient and  
349 medium *f*CO<sub>2</sub> treatments, a situation which has been observed in a previous mesocosm experiment.  
350 This was attributed to small-scale shifts in community composition and succession which could not  
351 be identified with only a once-daily measurement regime (Vogt *et al.*, 2008). DMS measured in all  
352 mesocosms fell within the range 2.7 to 6.8 nmol L<sup>-1</sup> across the course of the experiment. During  
353 Phase I, no difference was identified in DMS concentrations between *f*CO<sub>2</sub> treatments with the  
354 mean of all mesocosms 3.1 (± 0.2) nmol L<sup>-1</sup>. Concentrations in all mesocosms gradually declined  
355 from *t*21 until the end of DMS measurements on *t*31. DMS concentrations measured in the  
356 mesocosms and Baltic Sea were comparable to those measured in temperate coastal conditions in  
357 the North Sea (Turner *et al.*, 1988), the Mauritanian upwelling (Franklin *et al.*, 2009; Zindler *et al.*,  
358 2012) and South Pacific (Lee *et al.*, 2010).

359 The majority of DMS production is presumed to be from DMSP. However, an alternative  
360 production route for DMS is available through the methylation of methanethiol (Drotar *et al.*, 1987;  
361 Kiene and Hines, 1995; Stets *et al.*, 2004) predominantly identified in anaerobic environments such  
362 as freshwater lake sediments (Lomans *et al.*, 1997), saltmarsh sediments (Kiene and Visscher,  
363 1987) and microbial mats (Visscher *et al.*, 2003; Zinder *et al.*, 1977). Recent studies have also  
364 identified this pathway of DMS production from *Pseudomonas deceptionensis* in an aerobic  
365 environment (Carrión *et al.*, 2015), where *P. deceptionensis* was unable to synthesise or catabolise  
366 DMSP, but was able to enzymatically mediate DMS production from methanethiol (MeSH). The  
367 same enzyme has also been identified in a wide range of other bacterial taxa, including the  
368 cyanobacterial *Pseudanabaena*, which was identified in the Baltic Sea during this and previous  
369 investigations (Stuhr, pers. comm.; Kangro *et al.*, 2007; Nausch *et al.*, 2009). Correlations between  
370 DMS and the cyanobacterial equivalent Chl-*a* ( $\rho=0.42$ ,  $p<0.01$ ; Supplementary Figure S1g) and  
371 DMS and single-celled cyanobacteria ( $\rho=0.58$ ,  $p<0.01$ ; Supplementary Figure S2a) suggest that the  
372 methylation pathway may be a potential source of DMS within the Baltic Sea community. In  
373 addition to the methylation pathway, DMS production has been identified from S-methylmethionine  
374 (Bentley and Chasteen, 2004), as well as from the reduction of dimethylsulphoxide (DMSO) in both  
375 surface and deep waters by bacterial metabolism (Hatton *et al.*, 2004). As these compounds were

not measured in the mesocosms, it is impossible to determine if they were significant sources of DMS.

### 3.2.2 DMS and Community Interactions

Throughout Phase I, DMS showed no correlation with any measured variables of biological activity or cell abundance, and was unaffected by elevated  $f\text{CO}_2$ , indicating measured DMS concentrations were not directly related to the perturbation of the system and associated cellular stress (Sunda *et al.*, 2002). Of the studied phytoplankton groupings, neither the cryptophytes or chlorophytes as the largest contributors of Chl-*a* were identified as significant producers of DMSP. During Phase II, DMS was negatively correlated with Chl-*a* in the ambient and medium  $f\text{CO}_2$  mesocosms ( $\rho=-0.60$ ,  $p<0.01$ ). During Phase II, a significant correlation was seen between DMS and single-celled cyanobacteria identified predominantly as *Synechococcus* ( $\rho=0.53$ ,  $p<0.01$ ; Crawford *et al.* 2016 and supplementary table S1) and picoeukaryotes III ( $\rho=0.75$ ,  $p<0.01$ ). The peak in DMS concentrations is unlikely to be a delayed response to the increased Chl-*a* on *t*16 due to the time lag of 7 days. These higher concentrations were also likely connected to a peak in dissolved organic carbon (DOC) on *t*15, as well as increasing bacterial abundance during Phase II (Hornick *et al.*, 2016), as well as a response to the mesocosm wall cleaning which took place on *t*16. The variation in inorganic nutrient concentrations between mesocosms at the start of the experiment did not have an effect on DMS concentrations during Phase I, and by the start of Phase II the variation between mesocosms had decreased.

In previous mesocosm experiments (Archer *et al.*, 2013; Hopkins *et al.*, 2010; Webb *et al.*, 2015), DMS has shown poor correlations with many of the indicators of primary production and phytoplankton abundance, as well as showing the same trend of decreased concentrations in high  $f\text{CO}_2$  mesocosms compared to ambient. DMS production is often uncoupled from measurements of primary production in open waters (Lana *et al.*, 2012), and also often from production of its precursor DMSP (Archer *et al.*, 2009). DMS and DMSP are important sources of sulphur and carbon in the microbial food web for both bacteria and algae (Simó *et al.*, 2002, 2009), and since microbial turnover of DMSP and DMS play a significant role in net DMS production, it is unsurprising that DMS concentrations have shown poor correlation with DMSP-producing phytoplankton groups in past experiments and open waters.

DMS concentrations have been reported lower under conditions of elevated  $f\text{CO}_2$  compared to ambient controls, in both mesocosm experiments (Table 4) and phytoplankton monocultures

(Arnold *et al.*, 2013; Avgoustidi *et al.*, 2012). However, the varying response of the community within each experiment limit our ability to generalise the response of algal production of DMS and DMSP in all situations due to the characteristic community dynamics of each experiment in specific geographical areas and temporal periods. Previous experiments in the temperate Raunefjord of Bergen, Norway, showed lower abundance of DMSP-producing algal species, and subsequently DMSP-dependent DMS concentrations (Avgoustidi *et al.*, 2012; Hopkins *et al.*, 2010; Vogt *et al.*, 2008; Webb *et al.*, 2015). In contrast mesocosm experiments in the Arctic and Korea have shown increased abundance of DMSP producers (Archer *et al.*, 2013; Kim *et al.*, 2010) but lower DMS concentrations, while incubation experiments by Hopkins and Archer (2014) showed lower DMSP production but higher DMS concentrations at high  $f\text{CO}_2$ . However, in all previous experiments with DMSP as the primary precursor of DMS, elevated  $f\text{CO}_2$  had a less marked effect on measured DMSP concentrations than on measured DMS concentrations. Hopkins *et al.* (2010) suggested that ‘the perturbation of the system has a greater effect on the processes that control the conversion of DMSP to DMS rather than the initial production of DMSP itself’.

Previous mesocosm experiments have suggested significant links between increased bacterial production through greater availability of organic substrates at high  $f\text{CO}_2$  (Engel *et al.*, 2013; Piontek *et al.*, 2013). Further, Endres *et al.* (2014) identified significant enhanced enzymatic hydrolysis of organic matter with increasing  $f\text{CO}_2$ , with higher bacterial abundance. Higher bacterial abundance will likely result in greater bacterial demand for sulphur, and therefore greater consumption of DMS and conversion to DMSO. This was suggested as a significant sink for DMS in a previous experiment (Webb *et al.*, 2015), but during the present experiment, both bacterial abundance and bacterial production were lower at high  $f\text{CO}_2$  (Hornick *et al.*, 2016). However, as it has been proposed that only specialist bacterial groups are DMS consumers (Vila-Costa *et al.*, 2006b), and there is no determination of the DMS consumption characteristics of the bacterial community in the Baltic Sea, it is not known if this loss pathway is stimulated at high  $f\text{CO}_2$ . As microbial DMS yields can vary between 5-40% depending on the sulphur and carbon demand (Kiene and Linn, 2000), a change in the bacterial sulphur requirements could change DMS turnover despite lower abundance.

### 3.3 Iodocarbons in the mesocosms and relationships with community composition

Elevated  $f\text{CO}_2$  did not affect the concentration of iodocarbons in the mesocosms significantly at any time during the experiment, which is in agreement with the findings of Hopkins *et al.* (2013) in the Arctic, but in contrast to Hopkins *et al.* (2010) and Webb (2015), where iodocarbons were measured

440 significantly lower under elevated  $f\text{CO}_2$  (Table 4). Concentrations of all iodocarbons measured in  
441 the mesocosms and the Baltic Sea fall within the range of those measured previously in the region  
442 (Table 5). Mesocosm concentrations of  $\text{CH}_3\text{I}$  (Fig. 3a) and  $\text{C}_2\text{H}_5\text{I}$  (Fig. 3b) showed concentration  
443 ranges of 2.91 to 6.25 and 0.23 to 0.76  $\text{pmol L}^{-1}$  respectively.  $\text{CH}_3\text{I}$  showed a slight increase in all  
444 mesocosms during Phase I, peaking on  $t16$  which corresponded with higher Chl-*a* concentrations,  
445 and correlated throughout the entire experiment with picoeukaryote groups II ( $\rho=0.59$ ,  $p<0.01$ ) and  
446 III ( $\rho=0.23$ ,  $p<0.01$ ; Crawford *et al.* 2016) and nanoeukaryotes I ( $\rho=0.37$ ,  $p<0.01$ ). Significant  
447 differences identified between mesocosms for  $\text{CH}_3\text{I}$  were unrelated to elevated  $f\text{CO}_2$  ( $F=3.1$ ,  
448  $p<0.05$ ), but concentrations were on average 15% higher in Phase II than Phase I.  $\text{C}_2\text{H}_5\text{I}$  decreased  
449 slightly during Phases I and II, although concentrations of this halocarbon were close to its  
450 detection limit (0.2  $\text{pmol L}^{-1}$ ), remaining below 1  $\text{pmol L}^{-1}$  at all times. As this compound showed  
451 no significant effect of elevated  $f\text{CO}_2$ , and was identified by Orlikowska and Schulz-Bull (2009) as  
452 having extremely low concentrations in the Baltic Sea (Table 5), it will not be discussed further.

453 No correlation was found between  $\text{CH}_3\text{I}$  and Chl-*a* at any phase, and the only correlation of any  
454 phytoplankton grouping was with nanoeukaryotes II ( $\rho=0.88$ ,  $p<0.01$ ; Crawford *et al.*, 2016). These  
455  $\text{CH}_3\text{I}$  concentrations compare well to the 7.5  $\text{pmol L}^{-1}$  measured by Karlsson *et al.* (2008) during a  
456 cyanobacterial bloom in the Baltic Sea (Table 5), and the summer maximum of 16  $\text{pmol L}^{-1}$   
457 identified by Orlikowska and Schulz-Bull (2009).

458 Karlsson *et al.* (2008) showed Baltic Sea halocarbon production occurring predominately during  
459 daylight hours, with concentrations at night decreasing by 70% compared to late afternoon. Light  
460 dependent production of  $\text{CH}_3\text{I}$  has been shown to take place through abiotic processes, including  
461 radical recombination of  $\text{CH}_3$  and I (Moore and Zafiriou, 1994). However, since samples were  
462 integrated over the surface 10m of the water column, it was impossible to determine if  
463 photochemistry was affecting iodocarbon concentrations near the surface where some UV light was  
464 able to pass between the top of the mesocosm film material and the cover. For the same reason,  
465 photodegradation of halocarbons (Zika *et al.*, 1984) within the mesocosms was also likely to have  
466 been significantly restricted. Thus, as photochemical production was expected to be minimal,  
467 biogenic production was likely to have been the dominant source of these compounds. Karlsson *et al.*  
468 (2008) identified *Pseudanabaena* as a key producer of  $\text{CH}_3\text{I}$  in the Baltic Sea. However, the  
469 abundance of *Pseudanabaena* was highest during Phase I of the experiment (A. Stühr, Pers.  
470 Comm.) when  $\text{CH}_3\text{I}$  concentrations were lower, and as discussed previously, the abundance of these  
471 species constituted only a very small proportion of the community. Previous investigations in the  
472 laboratory have identified diatoms as significant producers of  $\text{CH}_3\text{I}$  (Hughes *et al.*, 2013; Manley

473 and De La Cuesta, 1997), and the low, steady-state abundance of the diatom populations in the  
474 mesocosms could have produced the same relatively steady-state trends in the iodocarbon  
475 concentrations.

476 Measured in the range 57.2 – 202.2 pmol L<sup>-1</sup> in the mesocosms, CH<sub>2</sub>I<sub>2</sub> (Fig. 3c) showed the clearest  
477 increase in concentration during Phase II, when it peaked on *t*21 in all mesocosms, with a maximum  
478 of 202.2 pmol L<sup>-1</sup> in M5 (348 µatm). During Phase II, concentrations of CH<sub>2</sub>I<sub>2</sub> were 57% higher  
479 than Phase I, and were therefore negatively correlated with Chl-*a*. The peak on *t*21 corresponds  
480 with the peak identified in DMS on *t*21, and concentrations through all three phases correlate with  
481 picoeukaryotes II ( $\rho=0.62$ ,  $p<0.01$ ) and III ( $\rho=0.47$ ,  $p<0.01$ ) and nanoeukaryotes I ( $\rho=0.88$ ,  $p<0.01$ ;  
482 Crawford *et al.*, 2015). CH<sub>2</sub>ClI (Fig. 3d) showed no peaks during either Phase I or Phase II,  
483 remaining within the range 3.81 to 8.03 pmol L<sup>-1</sup>, and again correlated with picoeukaryotes groups  
484 II ( $\rho=0.34$ ,  $p<0.01$ ) and III ( $\rho=0.38$ ,  $p<0.01$ ). These results may suggest that these groups possessed  
485 halo-peroxidase enzymes able to oxidise I<sup>-</sup>, most likely as an anti-oxidant mechanism within the cell  
486 to remove H<sub>2</sub>O<sub>2</sub> (Butler and Carter-Franklin, 2004; Pedersen *et al.*, 1996; Theiler *et al.*, 1978).  
487 However, given the lack of response of these compounds to elevated *f*CO<sub>2</sub> ( $F=1.7$ ,  $p<0.01$ ), it is  
488 unlikely that production was increased in relation to elevated *f*CO<sub>2</sub>. Production of all iodocarbons  
489 increased during Phase II when total Chl-*a* decreased, particularly after the walls of the mesocosms  
490 were cleaned for the first time, releasing significant volumes of organic aggregates into the water  
491 column. Aggregates have been suggested as a source of CH<sub>3</sub>I and C<sub>2</sub>H<sub>5</sub>I (Hughes *et al.*, 2008),  
492 likely through the alkylation of inorganic iodide (Urhahn and Ballschmiter, 1998) or through the  
493 breakdown of organic matter by microbial activity to supply the precursors required for iodocarbon  
494 production (Smith *et al.*, 1992). Hughes *et al.* (2008) did not identify this route as a pathway for  
495 CH<sub>2</sub>I<sub>2</sub> or CH<sub>2</sub>ClI production, but Carpenter *et al.* (2005) suggested a production pathway for these  
496 compounds through the reaction of HOI with aggregated organic materials.

### 497 **3.4 Bromocarbons in the mesocosms and the relationships with community** 498 **composition**

499 No effect of elevated *f*CO<sub>2</sub> was identified for any of the three bromocarbons, which compared with  
500 the findings from previous mesocosms where bromocarbons were studied (Hopkins *et al.*, 2010,  
501 2013; Webb, 2015; Table 4). Measured concentrations were comparable to those of Orlikowska and  
502 Schulz-Bull (2009) and Karlsson *et al.* (2008) measured in the Southern part of the Baltic Sea  
503 (Table 3). The concentrations of CHBr<sub>3</sub>, CH<sub>2</sub>Br<sub>2</sub> and CHBr<sub>2</sub>Cl showed no major peaks of  
504 production in the mesocosms. CHBr<sub>3</sub> (Fig. 4a) decreased rapidly in all mesocosms over Phase 0



505 from a maximum measured concentration of 147.5 pmol L<sup>-1</sup> in M1 (mean of 138.3 pmol L<sup>-1</sup> in all  
506 mesocosms) to a mean of 85.7 (±8.2 s.d.) pmol L<sup>-1</sup> in all mesocosms for the period *t*0 to *t*31 (Phases  
507 I and II). The steady-state CHBr<sub>3</sub> concentrations indicated a production source, however there was  
508 no clear correlation with any measured algal groups. CH<sub>2</sub>Br<sub>2</sub> concentrations (Fig. 4b) decreased  
509 steadily in all mesocosms from *t*-3 through to *t*31, over the range 4.0 to 7.7 pmol L<sup>-1</sup>, and CHBr<sub>2</sub>Cl  
510 followed a similar trend in the range 1.7 to 4.7 pmol L<sup>-1</sup> (Fig. 4c). Of the three bromocarbons, only  
511 CH<sub>2</sub>Br<sub>2</sub> showed correlation with total Chl-*a* ( $\rho=0.52$ ,  $p<0.01$ ), and with cryptophyte ( $\rho=0.86$ ,  
512  $p<0.01$ ) and dinoflagellate ( $\rho=0.65$ ,  $p<0.01$ ) derived Chl-*a*. Concentrations of CH<sub>2</sub>BrI were below  
513 detection limit for the entire experiment.

514 CH<sub>2</sub>Br<sub>2</sub> showed positive correlation with Chl-*a* ( $\rho=0.52$ ,  $p<0.01$ ), nanoeukaryotes II ( $\rho=0.34$ ,  
515  $p<0.01$ ) and cryptophytes ( $\rho=0.86$ ,  $p<0.01$ ; see supplementary material), whereas CHBr<sub>3</sub> and  
516 CHBr<sub>2</sub>Cl showed very weak or no correlation with any indicators of algal biomass. Schall *et al.*  
517 (1997) have proposed that CHBr<sub>2</sub>Cl is produced in seawater by the nucleophilic substitution of  
518 bromide by chloride in CHBr<sub>3</sub>, which given the steady-state concentrations of CHBr<sub>3</sub> would explain  
519 the similar distribution of CHBr<sub>2</sub>Cl concentrations. Production of all three bromocarbons was  
520 identified from large-size cyanobacteria such as *Aphanizomenon flos-aquae* by Karlsson *et al.*  
521 (2008), and in addition, significant correlations were found in the Arabian Sea between the  
522 abundance of the cyanobacterium *Trichodesmium* and several bromocarbons (Roy *et al.*, 2011), and  
523 the low abundance of such bacteria in the mesocosms would explain the low variation in  
524 bromocarbon concentrations through the experiment.

525 Halocarbon loss processes such as nucleophilic substitution (Moore, 2006), hydrolysis (Elliott and  
526 Rowland, 1995), sea-air exchange and microbial degradation are suggested as of greater importance  
527 than production of these compounds by specific algal groups, particularly given the relatively low  
528 growth rates and low net increase in total Chl-*a*. Hughes *et al.* (2013) identified bacterial inhibition  
529 of CHBr<sub>3</sub> production in laboratory cultures of *Thalassiosira* diatoms, but that it was not subject to  
530 bacterial breakdown; which could explain the relative steady state of CHBr<sub>3</sub> concentrations in the  
531 mesocosms. In contrast, significant bacterial degradation of CH<sub>2</sub>Br<sub>2</sub> in the same experiments could  
532 explain the steady decrease in CH<sub>2</sub>Br<sub>2</sub> concentrations seen in the mesocosms. Bacterial oxidation  
533 was also identified by Goodwin *et al.* (1998) as a significant sink for CH<sub>2</sub>Br<sub>2</sub>. As discussed for the  
534 iodocarbons, photolysis was unlikely due to the UV absorption of the mesocosm film, and limited  
535 UV exposure of the surface waters within the mesocosm due to the mesocosm cover. The ratio of  
536 CH<sub>2</sub>Br<sub>2</sub> to CHBr<sub>3</sub> was also unaffected by increased *f*CO<sub>2</sub>, staying within the range 0.04 to 0.08. This  
537 range in ratios is consistent with that calculated by Hughes *et al.* (2009) in the surface waters of an

Antarctic depth profile, and attributed to higher sea-air flux of  $\text{CHBr}_3$  than  $\text{CH}_2\text{Br}_2$  due to a greater concentrations gradient, despite the similar transfer velocities of the two compounds (Quack *et al.*, 2007). Using cluster analysis in a time-series in the Baltic Sea, Orlikowska and Schulz-Bull (2009) identified both these compounds as originating from different sources and different pathways of production.

Macroalgal production would not have influenced the mesocosm concentrations after the bags were sealed due to the isolation from the coastal environment, however macroalgal production into the water column prior to mesocosm installation (Klick, 1992; Leedham *et al.*, 2013; Moore and Tokarczyk, 1993) could account for the high initial concentrations with concentrations decreasing through the duration of the experiment via turnover and transfer to the atmosphere.

### 3.5 Natural variations in Baltic Sea $f\text{CO}_2$ and the effect on biogenic trace gases

#### 3.5.1 Physical variation and community dynamics

Baltic Sea deep waters have high  $f\text{CO}_2$  and subsequently lower pH (Schneider *et al.*, 2002), and the influx to the surface waters surrounding the mesocosms resulted in  $f\text{CO}_2$  increasing to 725  $\mu\text{atm}$  on  $t_{31}$ , close to the average  $f\text{CO}_2$  of the third highest mesocosm (M6: 868  $\mu\text{atm}$ ). These conditions imply that pelagic communities in the Baltic Sea are regularly exposed to rapid changes in  $f\text{CO}_2$  and the associated pH, as well as having communities associated with the elevated  $f\text{CO}_2$  conditions.

Chl-*a* followed the pattern of the mesocosms until  $t_4$ , after which concentrations were significantly higher than any mesocosm, peaking at 6.48  $\mu\text{g L}^{-1}$  on  $t_{16}$ , corresponding to the maximum Chl-*a* peak in the mesocosms and the maximum peak of temperature. As upwelled water intruded into the surface waters, the surface Chl-*a* was diluted with low Chl-*a* deep water: Chl-*a* in the surface 10m decreased from around  $t_{16}$  at the start of the upwelling until  $t_{31}$  when concentrations were once again equivalent to those found in the mesocosms at 1.30  $\mu\text{g L}^{-1}$ . In addition, there was potential introduction of different algal groups to the surface, but chlorophytes and cryptophytes were the major contributors to the Chl-*a* in the Baltic Sea, as in the mesocosms. Cyanobacteria contributed less than 2% of the total Chl-*a* in the Baltic Sea (Crawfurd *et al.*, 2016; Paul *et al.*, 2015).

Temporal community dynamics in the Baltic Sea were very different to that in the mesocosms across the experiment, with euglenophytes, chlorophytes, diatoms and prasinophytes all showing distinct peaks at the start of Phase II, with these same peaks identified in the nanoeukaryotes I and II, and picoeukaryotes II (Crawfurd *et al.*, 2016; Paul *et al.*, 2015; Supplementary Figs. S1 and S2).

569 The decrease in abundance of many groups during Phase II was attributed to the decrease in  
570 temperature and dilution with low-abundance deep waters.

### 571 **3.5.2 DMS in the Baltic Sea**

572 The input of upwelled water into the region mid-way through the experiment significantly altered  
573 the biogeochemical properties of the waters surrounding the mesocosms, and as a result it is  
574 inappropriate to directly compare the community structure and trace gas production of the Baltic  
575 Sea and the mesocosms. The Baltic Sea samples gave a mean DMS concentration of  $4.6 \pm 2.6$  nmol  
576  $L^{-1}$  but peaked at  $11.2$  nmol  $L^{-1}$  on *t*16, and were within the range of previous measurements for the  
577 region (Table 5). Strong correlations were seen between DMS and Chl-*a* ( $\rho=0.84$ ,  $p<0.01$ ), with the  
578 ratio of DMS: Chl-*a* at  $1.6 (\pm 0.3)$  nmol  $\mu g^{-1}$ . Other strong correlations were seen with  
579 euglenophytes ( $\rho=0.89$ ,  $p<0.01$ ), dinoflagellates ( $\rho=0.61$ ,  $p<0.05$ ) and nanoeukaryotes II ( $\rho=0.88$ ,  
580  $p<0.01$ ), but no correlation was found between DMS and cyanobacterial abundance, or with  
581 picoeukaryotes III which was identified in the mesocosms, suggesting that DMS had a different  
582 origin in the Baltic Sea community than in the mesocosms.

583 As CO<sub>2</sub> levels increased after *t*17 the DMS concentration measured in the Baltic Sea decreased,  
584 from the peak on *t*16 to the lowest recorded sample of the entire experiment at  $1.85$  nmol  $L^{-1}$ . As  
585 with Chl-*a*, DMS concentrations in the surface of the Baltic Sea may have been diluted with low-  
586 DMS deep water

### 587 **3.5.3 Halocarbon concentrations in the Baltic Sea**

588 Outside the mesocosms in the Baltic Sea, CH<sub>3</sub>I was measured at a maximum concentration of  $8.65$   
589 pmol  $L^{-1}$ , during Phase II, and showed limited effect of the upwelling event. Both CH<sub>2</sub>I<sub>2</sub> and  
590 CH<sub>2</sub>ClI showed higher concentrations in the Baltic Sea samples than the mesocosms (CH<sub>2</sub>I<sub>2</sub>:  $373.9$   
591 pmol  $L^{-1}$  and CH<sub>2</sub>ClI:  $18.1$  pmol  $L^{-1}$ ), and were correlated with the euglenophytes (CH<sub>2</sub>I<sub>2</sub>;  $\rho=0.63$ ,  
592  $p<0.05$  and CH<sub>2</sub>ClI;  $\rho=0.68$ ,  $p<0.01$ ) and nanoeukaryotes II (CH<sub>2</sub>I<sub>2</sub>;  $\rho=0.53$ ,  $p<0.01$  and CH<sub>2</sub>ClI;  
593  $\rho=0.58$ ,  $p<0.01$ ), but no correlation with Chl-*a*. Both polyhalogenated compounds showed  
594 correlation with picoeukaryote groups II and III, indicating that production was probably not limited  
595 to a single source. These concentrations of CH<sub>2</sub>I<sub>2</sub> and CH<sub>2</sub>ClI compared well to those measured  
596 over a macroalgal bed in the higher saline waters of the Kattegat by Klick and Abrahamsson (1992),  
597 suggesting that macroalgae were a significant iodocarbon source in the Baltic Sea.

598 As with the iodocarbons, the Baltic Sea showed significantly higher concentrations of CHBr<sub>3</sub>  
599 ( $F=28.1$ ,  $p<0.01$ ), CH<sub>2</sub>Br<sub>2</sub> ( $F=208.8$ ,  $p<0.01$ ) and CHBr<sub>2</sub>Cl ( $F=23.5$ ,  $p<0.01$ ) than the mesocosms,

with maximum concentrations 191.6 pmol L<sup>-1</sup>, 10.0 pmol L<sup>-1</sup> and 5.0 pmol L<sup>-1</sup> respectively. In the Baltic Sea, only CHBr<sub>3</sub> was correlated with Chl-*a* ( $\rho=0.65$ ,  $p<0.05$ ), cyanobacteria ( $\rho=0.61$ ,  $p<0.01$ ; Paul *et al.*, 2015) and nanoeukaryotes II ( $\rho=0.56$ ,  $p<0.01$ ; Crawford *et al.*, 2016), with the other two bromocarbons showing little to no correlations with any parameter of community activity. Production of bromocarbons from macroalgal sources (Laternus *et al.*, 2000; Leedham *et al.*, 2013; Manley *et al.*, 1992) was likely a significant contributor to the concentrations detected in the Baltic Sea; over the macroalgal beds in the Kattegat, Klick (1992) measured concentrations an order of magnitude higher than seen in this experiment for CH<sub>2</sub>Br<sub>2</sub> and CHBr<sub>2</sub>Cl. There was only a slight increase in bromocarbon concentrations as a result of the upwelling, indicating that the upwelled water had similar concentrations to the surface waters.

610

#### 611 **4 The Baltic Sea as a natural analogue to future ocean acidification?**

Mesocosm experiments are a highly valuable tool in assessing the potential impacts of elevated CO<sub>2</sub> on complex marine communities, however they are limited in that the rapid change in *f*CO<sub>2</sub> experienced by the community may not be representative of changes in the future ocean (Passow and Riebesell, 2005). This inherent problem with mesocosm experiments can be overcome through using naturally low pH/ high CO<sub>2</sub> areas such as upwelling regions or vent sites (Hall-Spencer *et al.*, 2008), which can give an insight into populations already living and acclimated to high CO<sub>2</sub> regimes by exposure over timescales measured in years. This mesocosm experiment was performed at such a location with a relatively high *f*CO<sub>2</sub> excursion, however still low compared to some sites (800  $\mu$ atm compared to >2000  $\mu$ atm; Hall-Spencer *et al.*, 2008), and it was clear through the minimal variation in Chl-*a* between all mesocosms that the community was relatively unaffected by elevated *f*CO<sub>2</sub>, although variation could be identified in some phytoplankton groups and some shifts in community composition. The upwelling event occurring mid-way through our experiment allowed comparison of the mesocosm findings with a natural analogue of the system, as well as showing the extent to which the system perturbation can occur (up to 800  $\mu$ atm), but the scale and timing of these upwelling events is difficult to determine, and therefore it will be hard to utilise these events as natural high CO<sub>2</sub> analogues.

In this paper, we described the temporal changes in concentrations of DMS and halocarbons in natural Baltic phytoplankton communities exposed to elevated *f*CO<sub>2</sub> treatments. In contrast to the halocarbons, concentrations of DMS were significantly lower in the highest *f*CO<sub>2</sub> treatments compared to the control. Despite very different physicochemical and biological characteristics of

632 the Baltic Sea (e.g. salinity, community composition and nutrient concentrations), this is a very  
633 similar outcome to that seen in several other high  $f\text{CO}_2$  experiments. The Baltic Sea trace gas  
634 samples give a good record of trace gas cycling during the injection of high  $f\text{CO}_2$  deep water into  
635 the surface community during upwelling events. For the concentrations of halocarbons, the  
636 measured concentrations did not change during the upwelling event in the Baltic Sea, which may  
637 indicate that emissions of organic iodine and bromine are unlikely to change with future  
638 acidification of the Baltic Sea without significant alteration to the meteorological conditions.  
639 Further studies of these compounds are important to determine rates of production and consumption  
640 to include in prognostic and predictive models. However, net production of organic sulphur within  
641 the Baltic Sea region is likely to decrease with an acidified future ocean scenario, despite the  
642 possible acclimation of the microbial community to elevated  $f\text{CO}_2$ . This will potentially impact the  
643 flux of DMS to the atmosphere over Northern Europe, and could have significant impacts on the  
644 local climate through the reduction of atmospheric sulphur aerosols. Data from a previous  
645 mesocosm experiment has been used to estimate future global changes in DMS production, and  
646 predicted that global warming would be amplified (Six *et al.*, 2013); utilising the data from this  
647 experiment combined with those of other mesocosm, field and laboratory experiments and  
648 associated modelling provide the basis for a better understanding of the future changes in global  
649 DMS production and their climatic impacts.

650

651

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986

987 Table 1. Summary of  $f\text{CO}_2$  and  $\text{pH}_\text{T}$  (total scale) during phases 0, 1 and 2 of the mesocosm  
 988 experiment.

		Whole Experiment		Phase 0 ( $t\text{-}3$ to $t0$ )		Phase I ( $t1$ – $t16$ )		Phase II ( $t16$ – $t31$ )	
		( $t\text{-}3$ to $t31$ )							
Mesocosm <sup>a</sup>	Target $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_\text{T}$	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_\text{T}$	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_\text{T}$	Mean $f\text{CO}_2$ ( $\mu\text{atm}$ )	Mean $\text{pH}_\text{T}$
M1	Control	331	7.91	231	8.00	328	7.95	399	7.86
M5	Control	334	7.91	244	7.98	329	7.94	399	7.52
M7	390	458	7.80	239	7.99	494	7.81	532	7.76
M6	840	773	7.63	236	7.99	932	7.59	855	7.59
M3	1120	950	7.56	243	7.98	1176	7.51	1027	7.52
M8	1400	1166	7.49	232	8.00	1481	7.43	1243	7.45
Baltic Sea	380	350	7.91	298	7.91	277	7.98	436	7.86

989 <sup>a</sup> listed in order of increasing  $f\text{CO}_2$

990

991 Table 2. Calibration ranges and calculated percentage mean relative standard error for the trace  
 992 gases measured in the mesocosms.

Compound	Calibration range (pmol L <sup>-1</sup> )	% Mean relative standard error
DMS	600 – 29300*	6.33
DMSP	2030 – 405900*	
CH <sub>3</sub> I	0.11 – 11.2	4.62
CH <sub>2</sub> I <sub>2</sub>	5.61 – 561.0	4.98
C <sub>2</sub> H <sub>5</sub> I	0.10 – 4.91	5.61
CH <sub>2</sub> ClI	1.98 – 99.0	3.64
CHBr <sub>3</sub>	8.61 – 816.0	4.03
CH <sub>2</sub> Br <sub>2</sub>	0.21 – 20.9	5.30
CHBr <sub>2</sub> Cl	0.07 – 7.00	7.20

993 \* throughout the rest of this paper, these measurements are given in nmol L<sup>-1</sup>.

994

Table 3. Abundance and contributions of different phytoplankton groups to the total phytoplankton community assemblage, showing the range of measurements from total Chl-*a* (Paul *et al.*, 2015), CHEMTAX analysis of derived Chl-*a* (Paul *et al.*, 2015) and phytoplankton abundance (Crawfurd *et al.*, 2016). Data are split into the range of all the mesocosm measurements and those from the Baltic Sea.

Mesocosm				Baltic Sea		
	Range	Range	% Contribution to Chl-	Range	Range	% Contribution to Chl- <i>a</i>
	Integrated 10 m	Integrated 17 m	<i>a</i>	Integrated 10 m	Integrated 17 m	
<b>Chl-<i>a</i></b>	0.9 – 2.9	0.9 – 2.6	100	1.3 – 6.5	1.12 – 5.5	100
<b>Phytoplankton Taxonomy (Equivalent Chlorophyll µg L<sup>-1</sup>)</b>						
<b>Cyanobacteria</b>		0.01 – 0.4	8		0.0 – 0.1	1
<b>Prasinophytes</b>		0.04 – 0.3	7		0.01 – 0.3	4
<b>Euglenophytes</b>		0.0 – 1.6	15		0.0 – 2.6	21
<b>Dinoflagellates</b>		0.0 – 0.3	3		0.04 – 0.6	9
<b>Diatoms</b>		0.1 – 0.3	7		0.04 – 0.9	9
<b>Chlorophytes</b>		0.3 – 2.0	40		0.28 – 3.1	41
<b>Cryptophytes</b>		0.1 – 1.4	21		0.1 – 1.0	15
<b>Small Phytoplankton (&lt;10 µm) abundance (cells mL<sup>-1</sup>)</b>						
<b>Cyanobacteria</b>	55000 – 380000	65000 – 470000		30000 – 180000	30000 – 250000	
<b>Picoeukaryotes I</b>	15000 – 100000	17000 – 111000		5000 – 70000	6100 – 78000	
<b>Picoeukaryotes II</b>	700 – 4000	600 – 4000		400 – 3000	460 – 3700	
<b>Picoeukaryotes III</b>	1000 - 9000	1100 – 8500		1000 – 6000	950 – 7500	
<b>Nanoeukaryotes I</b>	400 – 1400	270 – 1500		200 – 4000	210 – 4100	
<b>Nanoeukaryotes II</b>	0 – 400	4 – 400		100 – 1100	60 – 1300	



Table 4. Concentration ranges of trace gases measured in the mesocosms compared to other open water ocean acidification experiments, showing the range of concentrations for each gas and the percentage change between the control and the highest  $f\text{CO}_2$  treatment.

	Range $f\text{CO}_2$		DMS	$\text{CH}_3\text{I}$	$\text{CH}_2\text{I}_2$	$\text{CH}_2\text{ClI}$	$\text{CHBr}_3$	$\text{CH}_2\text{Br}_2$	$\text{CH}_2\text{Br}_2\text{Cl}$
	( $\mu\text{atm}$ )		( $\text{nmol L}^{-1}$ )						
<b>SOPRAN Tvärminne Mesocosm (this study)</b>	346 – 1333	Range	2.7-6.8	2.9-6.4	57-202	3.8-8.0	69-148	4.0-7.7	1.7-3.1
		% change	-34	-0.3	1.3	-11	-9	-3	-4
<b>SOPRAN Bergen 2011 (Webb <i>et al.</i>, 2015)</b>	280 – 3000	Range	0.1-4.9	4.9-32	5.8-321	9.0-123	64-306	6.3-30.8	3.9-14
		% change	-60	-37	-48	-27	-2	-4	-6
<b>NERC Microbial Metagenomics Experiment, Bergen 2006 (Hopkins <i>et al.</i>, 2010)</b>	300 - 750	Range	ND-50	2.0-25	ND-750	ND-700	5.0-80	ND-5.5	0.2-1.2
		% change	-57	-41	-33	-28	13	8	22
<b>EPOCA Svalbard 2010 (Archer <i>et al.</i>, 2013; Hopkins <i>et al.</i>, 2013)</b>	180 - 1420	Range	ND-14	0.04-10	0.01-2.5	0.3-1.6	35-151	6.3-33.3	1.6-4.7
		% change	-60	NS		NS	NS	NS	NS
<b>UKOA European Shelf 2011 (Hopkins and Archer, 2014)</b>	340 - 1000	Range	0.5-12						
		% change	225						
<b>Korean Mesocosm Experiment 2012 (Park <i>et al.</i>, 2014)</b>	160 - 830	Range	1.0-100						
		% change	-82						

Table 5. Concentration ranges of trace gases measured in the Baltic Sea compared to concentrations measured in the literature. ND – Not Detected.

Study	DMS concentration range (nmol L <sup>-1</sup> )	Halocarbon concentration range (pmol L <sup>-1</sup> )						
		CH <sub>3</sub> I	CH <sub>2</sub> I <sub>2</sub>	C <sub>2</sub> H <sub>5</sub> I	CH <sub>3</sub> Cl	CHBr <sub>3</sub>	CH <sub>2</sub> Br <sub>2</sub>	CH <sub>2</sub> Br <sub>2</sub> Cl
<b>SOPRAN Tvärminne Baltic Sea (This Study)</b>	1.9-11	4.3-8.6	66.9-374	0.6 – 1.0	7.0-18	93-192	7.1-10	3.3-5.0
<b>Orlikowska and Schulz-Bull (2009)</b>	0.3-120	1-16	0-85	0.4 – 1.2	5-50	5.0-40	2.0-10	0.8-2.5
<b>Karlsson <i>et al.</i> (2008)</b>		3.0-7.5				35-60	4.0-7.0	2.0-6.5
<b>Klick and Abrahamsson (1992)</b>			15-709		11-74	14-585		
<b>Klick (1992)</b>			ND-243		ND-57	40-790	ND-86	ND-29
<b>Leck and Rodhe (1991)</b>	0.4-2.8							
<b>Leck <i>et al.</i> (1990)</b>	ND-3.2							

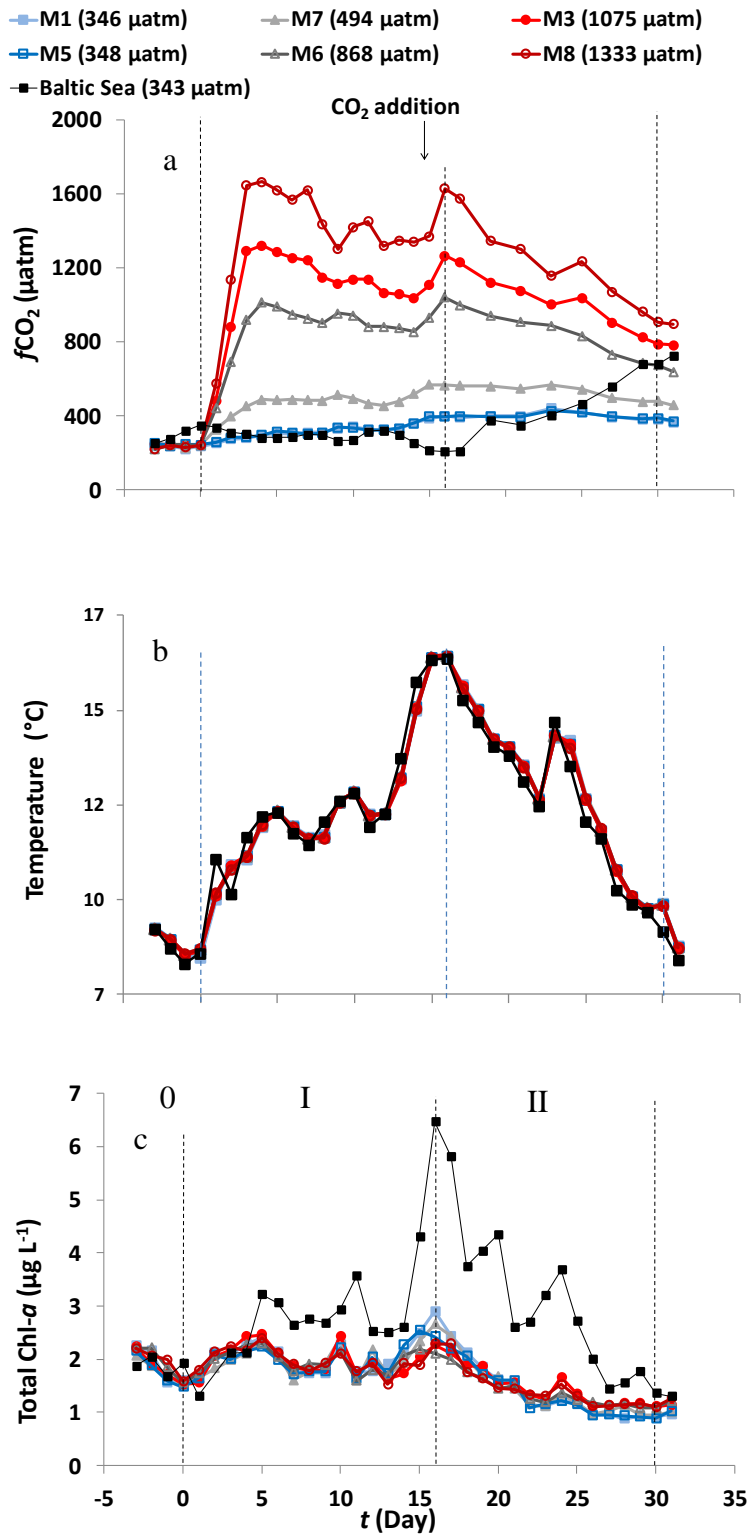


Figure 1. Daily measurements of (a)  $f\text{CO}_2$ , (b) mean temperature and (c) total Chlorophyll- $a$  in the mesocosms and surrounding Baltic Sea waters. Dashed lines represent the three Phases of the experiment, based on the Chl- $a$  data.

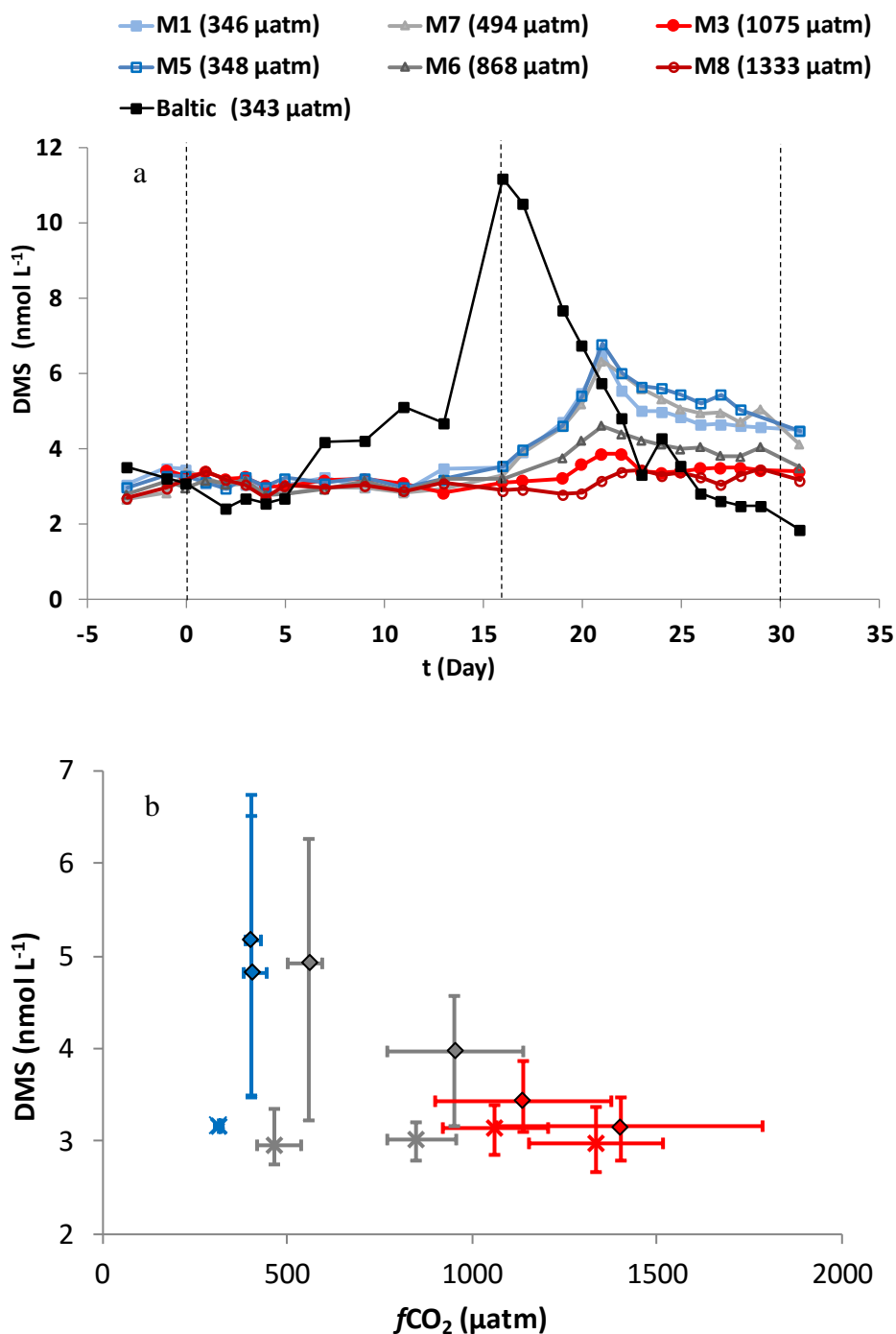


Figure 2. (a) Mean DMS concentrations measured daily in the mesocosms and Baltic Sea from an integrated water sample of the surface 10m. Dashed lines show the Phases of the experiment as given in Fig. 1,  $f\text{CO}_2$  shown in the legend are mean  $f\text{CO}_2$  across the duration of the experiment. (b) Mean DMS concentrations from each mesocosm during Phase I (crosses) and Phase II (diamonds), for ambient (blue), medium (grey) and high  $f\text{CO}_2$  (red), with error bars showing the range of both the DMS and  $f\text{CO}_2$ .

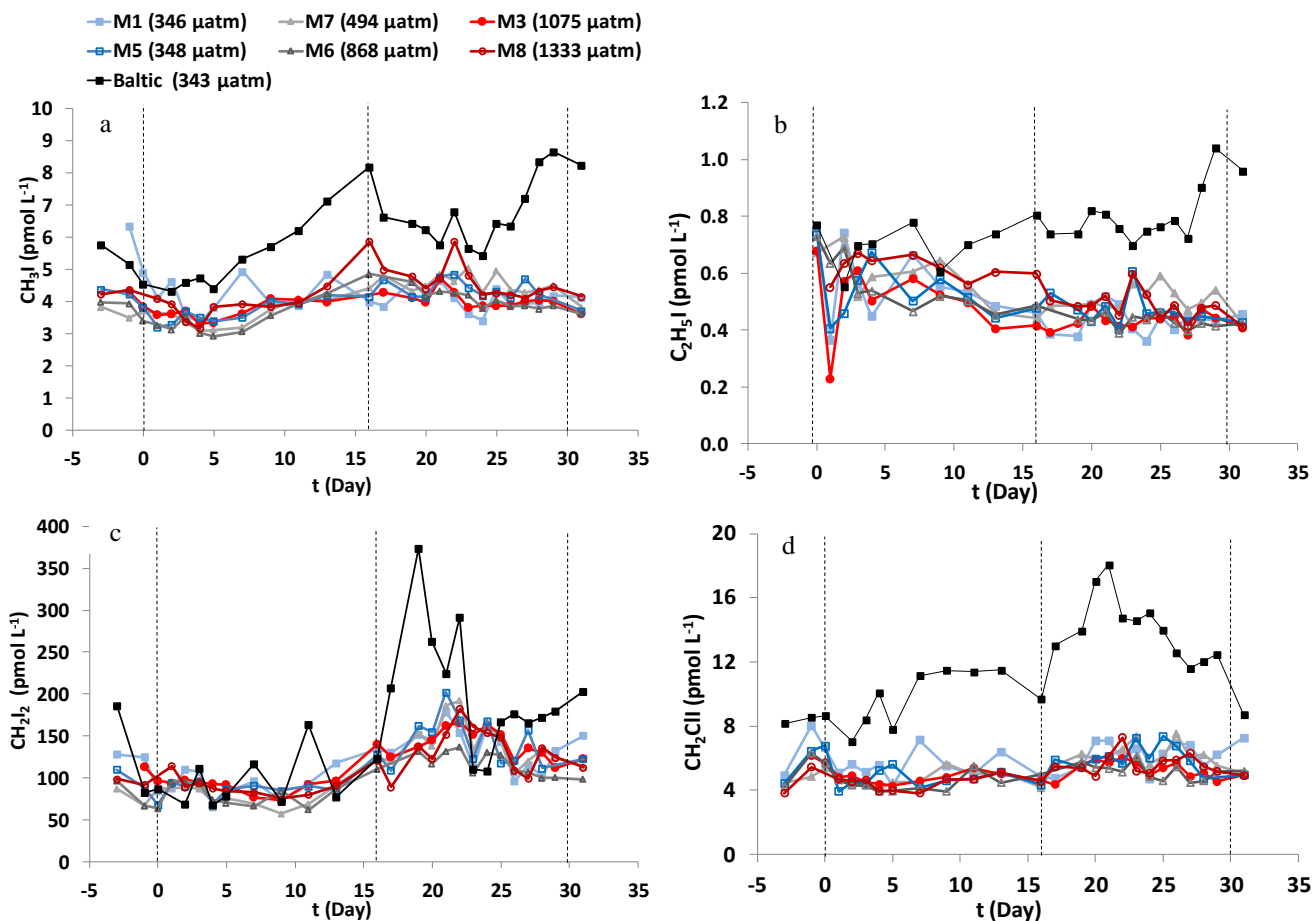


Figure 3. Mean concentrations ( $\text{pmol L}^{-1}$ ) of (a)  $\text{CH}_3\text{I}$ , (b)  $\text{C}_2\text{H}_5\text{I}$ , (c)  $\text{CH}_2\text{I}_2$  and (d)  $\text{CH}_2\text{ClI}$  taken from a water sample integrated from the surface 10m. Dashed lines indicate the Phases of the experiment, as given in Fig. 2.  $f\text{CO}_2$  shown in the legend are mean  $f\text{CO}_2$  across the duration of the experiment.

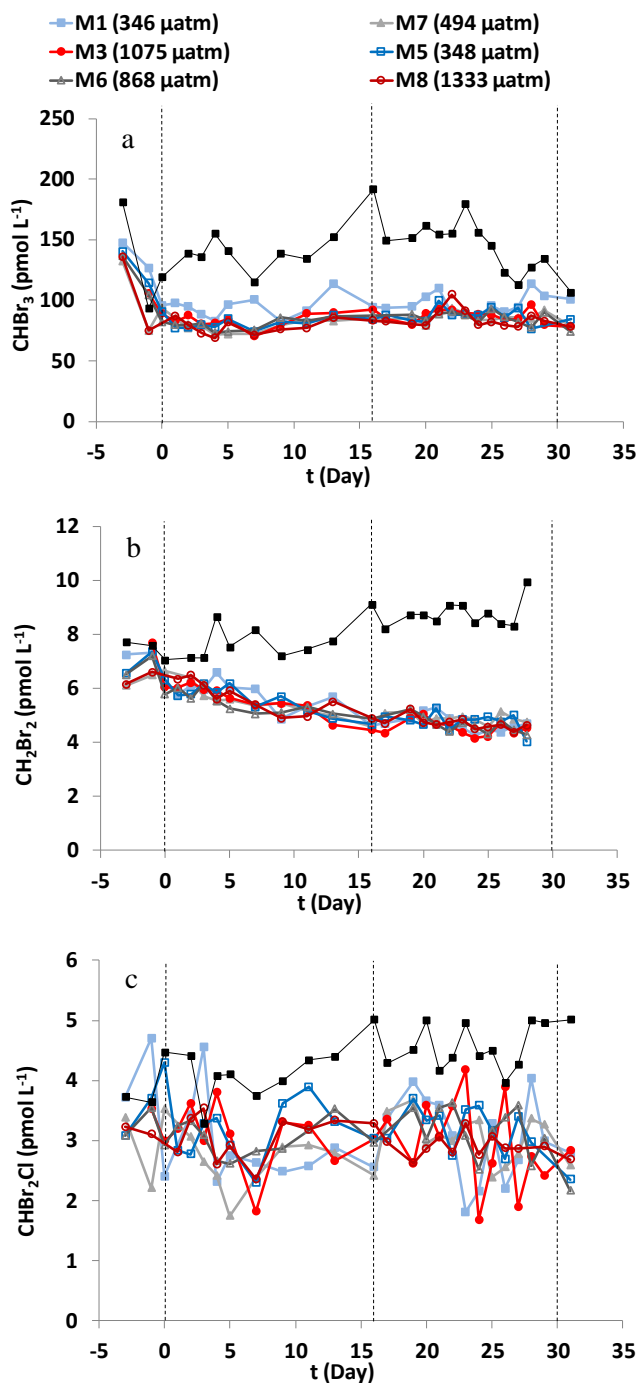


Figure 4. Mean concentrations ( $\text{pmol L}^{-1}$ ) of (a)  $\text{CHBr}_3$ , (b)  $\text{CH}_2\text{Br}_2$  and (c)  $\text{CHBr}_2\text{Cl}$  taken from a water sample integrated from the surface 10m. Dashed lines indicate the phases of the experiment as defined in Fig. 2,  $f\text{CO}_2$  shown in the legend are mean  $f\text{CO}_2$  across the duration of the experiment.