

Effect of ocean acidification and elevated $f\text{CO}_2$ on trace gas production by a Baltic Sea summer phytoplankton community

A.L. Webb^{1,2}, E. Leedham-Elvidge¹, C. Hughes³, F.E. Hopkins⁴, G. Malin¹, L.T. Bach⁵, K. Schulz⁶, K. Crawford⁷, C.P.D. Brussaard^{7,8}, A. Stühr⁵, U. Riebesell⁵, and P.S. Liss¹.

[1] {Centre for Ocean and Atmospheric Sciences, School of Environmental Science, University of East Anglia, Norwich, UK, NR4 7TJ}

[2] {Groningen Institute for Evolutionary Life Sciences, University of Groningen, 9700 CC Groningen, The Netherlands}

[3] {Environmental Department, University of York, York, UK, YO10 5DD}

[4] {Plymouth Marine Laboratory, Plymouth, UK, PL1 3DH}

[5] {GEOMAR Helmholtz Centre for Ocean Research Kiel, Düsternbrooker Weg 20, 24148 Kiel, Germany.}

[6] {Centre for Coastal Biogeochemistry, School of Environment, Science and Engineering, Southern Cross University, Lismore, NSW 2480, Australia.}

[7] {Department of Biological Oceanography, NIOZ – Royal Netherlands Institute for Sea Research, PO Box 59, 1790 AB Den Burg, Texel, The Netherlands}

[8] {Aquatic Microbiology, Institute for Biodiversity and Ecosystem Dynamics, University of Amsterdam, P.O. Box 94248, 1090 GE, Amsterdam, The Netherlands}

Correspondence to: Alison Webb (a.l.webb@rug.nl)

Abstract

The Baltic Sea is a unique environment as the largest body of brackish water in the world. Acidification of the surface oceans due to absorption of anthropogenic CO_2 emissions is an additional stressor facing the pelagic community of the already challenging Baltic Sea. To investigate its impact on trace gas biogeochemistry, a large-scale mesocosm experiment was performed off Tvärminne Research Station, Finland in summer 2012. During the second half of the experiment, dimethylsulphide (DMS) concentrations in the highest $f\text{CO}_2$ mesocosms

(1075 - 1333 μatm) were 34% lower than at ambient CO_2 (350 μatm). However, the net production (as measured by concentration change) of seven halocarbons analysed was not significantly affected by even the highest CO_2 levels after 5 weeks exposure. Methyl iodide (CH_3I) and diiodomethane (CH_2I_2) showed 15% and 57% increases in mean mesocosm concentration ($3.8 \pm 0.6 \text{ pmol L}^{-1}$ increasing to $4.3 \pm 0.4 \text{ pmol L}^{-1}$ and $87.4 \pm 14.9 \text{ pmol L}^{-1}$ increasing to $134.4 \pm 24.1 \text{ pmol L}^{-1}$ respectively) during Phase II of the experiment, which were unrelated to CO_2 and corresponded to 30% lower Chl-*a* concentrations compared to Phase I. No other iodocarbons increased or showed a peak, with mean chloriodomethane (CH_2ClI) concentrations measured at $5.3 (\pm 0.9) \text{ pmol L}^{-1}$ and iodoethane ($\text{C}_2\text{H}_5\text{I}$) at $0.5 (\pm 0.1) \text{ pmol L}^{-1}$. Of the concentrations of bromoform (CHBr_3 ; mean $88.1 \pm 13.2 \text{ pmol L}^{-1}$), dibromomethane (CH_2Br_2 ; mean $5.3 \pm 0.8 \text{ pmol L}^{-1}$) and dibromochloromethane (CHBr_2Cl , mean $3.0 \pm 0.5 \text{ pmol L}^{-1}$), only CH_2Br_2 showed a decrease of 17% between Phases I and II, with CHBr_3 and CHBr_2Cl showing similar mean concentrations in both Phases. Outside the mesocosms, an upwelling event was responsible for bringing colder, high CO_2 , low pH water to the surface starting on day *t*16 of the experiment; this variable CO_2 system with frequent upwelling events implies the community of the Baltic Sea is acclimated to regular significant declines in pH caused by up to 800 $\mu\text{atm } f\text{CO}_2$. After this upwelling, DMS concentrations declined, but halocarbon concentrations remained similar or increased compared to measurements prior to the change in conditions. Based on our findings, with future acidification of Baltic Sea waters, biogenic halocarbon emissions are likely to remain at similar values to today, however emissions of biogenic sulphur could significantly decrease from this region.

1 Introduction

Anthropogenic activity has increased the fugacity of atmospheric carbon dioxide ($f\text{CO}_2$) from 280 μatm (pre-Industrial Revolution) to over 400 μatm today (Hartmann *et al.*, 2013). The IPCC AR5 long-term projections for atmospheric $p\text{CO}_2$ and associated changes to the climate have been established for a variety of scenarios of anthropogenic activity until the year 2300. As the largest global sink for atmospheric CO_2 , the global ocean has absorbed an estimated 30% of excess CO_2 produced (Canadell *et al.*, 2007). With atmospheric $p\text{CO}_2$ projected to possibly exceed 2000 μatm by the year 2300 (Collins *et al.*, 2013; Cubasch *et al.*, 2013), the ocean will take up increasing amounts of CO_2 , with a potential lowering of surface ocean pH by over 0.8 units (Raven *et al.*, 2005). The overall effect of acidification on the biogeochemistry of surface ocean ecosystems is

62 unknown and currently unquantifiable, with a wide range of potential positive and negative impacts
63 (Doney *et al.*, 2009; Hofmann *et al.*, 2010; Ross *et al.*, 2011).

64 A number of volatile organic compounds are produced by marine phytoplankton (Liss *et al.*, 2014),
65 including the climatically important trace gas dimethylsulphide (DMS, C₂H₆S) and a number of
66 halogen-containing organic compounds (halocarbons) including methyl iodide (CH₃I) and
67 bromoform (CHBr₃). These trace gases are a source of sulphate particles and halide radicals when
68 oxidised in the atmosphere, and have important roles as ozone catalysts in the troposphere and
69 stratosphere (O'Dowd *et al.*, 2002; Solomon *et al.*, 1994) and as cloud condensation nuclei (CCNs;
70 Charlson *et al.*, 1987).

71 DMS is found globally in surface waters originating from the algal-produced precursor
72 dimethylsulphoniopropionate (DMSP, C₅H₁₀O₂S). Both DMS and DMSP provide the basis for
73 major routes of sulphur and carbon flux through the marine microbial food web, and can provide up
74 to 100% of the bacterial and phytoplanktonic sulphur demand (Simó *et al.*, 2009; Vila-Costa *et al.*,
75 2006a). DMS is also a volatile compound which readily passes through the marine boundary layer
76 to the troposphere, where oxidation results in a number of sulphur-containing particles important for
77 atmospheric climate feedbacks (Charlson *et al.*, 1987; Quinn and Bates, 2011); for this reason, any
78 change in the production of DMS may have significant implications for climate regulation. Several
79 previous acidification experiments have shown differing responses of both compounds (e.g.
80 Avgoustidi *et al.*, 2012; Hopkins *et al.*, 2010; Webb *et al.*, 2015), while others have shown delayed
81 or more rapid responses as a direct effect of CO₂ (e.g. Archer *et al.*, 2013; Vogt *et al.*, 2008).
82 Further, some laboratory incubations of coastal microbial communities showed increased DMS
83 production with increased *f*CO₂ (Hopkins and Archer, 2014), but lower DMSP production. The
84 combined picture arising from existing studies is that the response of communities to *f*CO₂
85 perturbation is not predictable and requires further study. Previous studies measuring DMS in the
86 Baltic Sea measured concentrations up to 100 nmol L⁻¹ during the summer bloom, making the
87 Baltic Sea a significant source of DMS (Orlikowska and Schulz-Bull, 2009).

88 In surface waters, halocarbons such as methyl iodide (CH₃I), chloriodomethane (CH₂ClI) and
89 bromoform (CHBr₃) are produced by biological and photochemical processes: many marine
90 microbes (for example cyanobacteria; Hughes *et al.*, 2011, diatoms; Manley and De La Cuesta,
91 1997 and haptophytes; Scarratt and Moore, 1998) and macroalgae (e.g. brown-algal *Fucus* species;
92 Chance *et al.*, 2009 and red algae; Leedham *et al.*, 2013) utilise halides from seawater and emit a
93 range of organic and inorganic halogenated compounds. This production can lead to significant
94 annual flux to the marine boundary layer in the order of 10 Tg iodine-containing compounds

95 ('iodocarbons'; O'Dowd *et al.*, 2002) and 1 Tg bromine-containing compounds ('bromocarbons';
96 Goodwin *et al.*, 1997) into the atmosphere. The effect of acidification on halocarbon concentrations
97 has received limited attention, but two acidification experiments measured lower concentrations of
98 several iodocarbons while bromocarbons were unaffected by $f\text{CO}_2$ up to 3000 μatm (Hopkins *et al.*,
99 2010; Webb, 2015), whereas an additional mesocosm study did not elicit significant differences
100 from any compound up to 1400 $\mu\text{atm } f\text{CO}_2$ (Hopkins *et al.*, 2013).

101 Measurements of the trace gases within the Baltic Sea are limited, with no prior study of DMSP
102 concentrations in the region. The Baltic Sea is the largest body of brackish water in the world, and
103 salinity ranges from 1 to 15. Furthermore, seasonal temperature variations of over 20 °C are
104 common. A permanent halocline at 50-80 m separates CO_2 -rich, bottom waters from fresher, lower
105 CO_2 surface waters, and a summer thermocline at 20 m separates warmer surface waters from those
106 below 4 °C (Janssen *et al.*, 1999). Upwelling of bottom waters from below the summer thermocline
107 is a common summer occurrence, replenishing the surface nutrients while simultaneously lowering
108 surface temperature and pH (Brutemark *et al.*, 2011). Baltic organisms are required to adapt to
109 significant variations in environmental conditions. The species assemblage in the Baltic Sea is
110 different to those studied during previous mesocosm experiments in the Arctic, North Sea and
111 Korea (Brussaard *et al.*, 2013; Engel *et al.*, 2008; Kim *et al.*, 2010), and are largely unstudied in
112 terms of their community trace gas production during the summer bloom. Following the spring
113 bloom (July-August), a low dissolved inorganic nitrogen (DIN) to dissolved inorganic phosphorous
114 (DIP) ratio combines with high temperatures and light intensities to encourage the growth of
115 heterocystous cyanobacteria, (Niemisto *et al.*, 1989; Raateoja *et al.*, 2011), in preference to nitrate-
116 dependent groups.

117 Here we report the concentrations of DMS, DMSP and halocarbons from the 2012 summer post-
118 bloom season mesocosm experiment aimed to assess the impact of elevated $f\text{CO}_2$ on the microbial
119 community and trace gas production in the Baltic Sea. Our objective was to assess how changes in
120 the microbial community driven by changes in $f\text{CO}_2$ impacted DMS and halocarbon concentrations.
121 It is anticipated that any effect of CO_2 on the growth of different groups within the phytoplankton
122 assemblage will result in an associated change in trace gas concentrations measured in the
123 mesocosms as $f\text{CO}_2$ increases, which can potentially be used to predict future halocarbon and
124 sulphur emissions from the Baltic Sea region.

125

126 **2 Methods**

127 **2.1 Mesocosm design and deployment**

128 Nine mesocosms were deployed on the 10th June 2012 (day $t-10$; days are numbered negative prior
129 to CO₂ addition and positive afterward) and moored near Tvärminne Zoological Station (59° 51.5'
130 N, 23° 15.5' E) in Tvärminne Storfjärden in the Baltic Sea. Each mesocosm comprised a
131 thermoplastic polyurethane (TPU) enclosure of 17 m depth, containing approximately 54,000 L of
132 seawater, supported by an 8m tall floating frame capped with a polyvinyl hood. For full technical
133 details of the mesocosms see Czerny *et al.* (2013) and Riebesell *et al.* (2013). The mesocosm bags
134 were filled by lowering through the stratified water column until fully submerged, with the opening
135 at both ends covered by 3 mm mesh to exclude organisms larger than 3 mm such as fish and large
136 zooplankton. The mesocosms were then left for 3 days ($t-10$ to $t-7$) with the mesh in position to
137 allow exchange with the external water masses and ensure the mesocosm contents were
138 representative of the phytoplankton community in the Storfjärden. On $t-7$ the bottom of the
139 mesocosm was sealed with a sediment trap and the upper opening was raised to approximately 1.5
140 m above the water surface. Stratification within the mesocosm bags was broken up on $t-5$ by the use
141 of compressed air for three and a half minutes to homogenise the water column and ensure an even
142 distribution of inorganic nutrients at all depths. Unlike in previous experiments, there was no
143 addition of inorganic nutrients to the mesocosms at any time during the experiment; mean inorganic
144 nitrate, inorganic phosphate and ammonium concentrations measured across all mesocosms at the
145 start of the experiment were 37.2 (\pm 18.8 s.d.) nmol L⁻¹, 323.9 (\pm 19.4 s.d.) nmol L⁻¹ and 413.8 (\pm
146 319.5 s.d.) nmol L⁻¹ respectively.

147 To obtain mesocosms with different $f\text{CO}_2$, the carbonate chemistry of the mesocosms was altered
148 by the addition of different volumes of 50 μm filtered, CO₂-enriched Baltic Sea water (sourced from
149 outside the mesocosms), to each mesocosm over a four-day period, with the first day of addition
150 being defined as day $t0$. Addition of the enriched CO₂ water was by the use of a bespoke dispersal
151 apparatus ('Spider') lowered through the bags to ensure even distribution throughout the water
152 column (further details are in Riebesell *et al.* 2013). Measurements of salinity in the mesocosms
153 throughout the experiment determined that three of the mesocosms were not fully sealed, and had
154 undergone unquantifiable water exchange with the surrounding waters. These three mesocosms
155 (M2, M4 and M9) were excluded from the analysis. Two mesocosms were designated as controls
156 (M1 and M5) and received only filtered seawater via the Spider; four mesocosms received addition
157 of CO₂-enriched waters, with the range of target $f\text{CO}_2$ levels between 600 and 1650 μatm (M7, 600

158 μatm ; M6, 950 μatm ; M3, 1300 μatm ; M8 1650 μatm). Mesocosms were randomly allocated a
159 target $f\text{CO}_2$; a noticeable decrease in $f\text{CO}_2$ was identified in the three highest $f\text{CO}_2$ mesocosms (M6,
160 M3 and M8) over the first half of the experiment, which required the addition of more CO_2 enriched
161 water on $t15$ to bring the $f\text{CO}_2$ back up to maximum concentrations (Fig. 1a; Paul *et al.*, 2015). A
162 summary of the $f\text{CO}_2$ in the mesocosms can be seen in Table 1. At the same time as this further CO_2
163 addition on $t15$, the walls of the mesocosms were cleaned using a bespoke wiper apparatus (See
164 Riebesell *et al.*, 2013 for more information), followed by weekly cleaning to remove aggregations
165 on the film which would block incoming light. Light measurements showed that over 95% of the
166 photosynthetically active radiation (PAR) was transmitted by the clean TPU and PVC materials
167 with 100% absorbance of UV light (Riebesell *et al.*, 2013). Samples for most parameters were
168 collected from the mesocosms at the same time every morning from $t-3$, and analysed daily or every
169 other day.

170 2.2 Trace gas extraction and analysis

171 2.2.1 DMS and halocarbons

172 A depth-integrated water sampler (IWS, HYDRO-BIOS, Kiel, Germany) was used to sample the
173 entire 17 m water column daily or alternative daily. As analysis of Chlorophyll-*a* (Chl-*a*) showed it
174 to be predominantly produced in the first 10 m of the water column, trace gas analysis was
175 conducted on only integrated samples collected from the surface 10 m, with all corresponding
176 community parameter analyses with the exception of pigment analysis performed also to this depth.
177 Water samples for trace gas analysis were taken from the first IWS from each mesocosm to
178 minimise the disturbance and bubble entrainment from taking multiple samples in the surface
179 waters. As in Hughes *et al.* (2009), samples were collected in 250 mL amber glass bottles in a
180 laminar flow with minimal disturbance to the water sample, using Tygon tubing from the outlet of
181 the IWS. Bottles were rinsed twice before being carefully filled from the bottom with minimal
182 stirring, and allowed to overflow the volume of the bottle approximately three times before sealing
183 with a glass stopper to prevent bubble formation and atmospheric contact. Samples were stored
184 below 10°C in the dark for 2 hours prior to analysis. Each day, a single sample was taken from each
185 mesocosm, with two additional samples taken from one randomly selected mesocosm to evaluate
186 the precision of the analysis ($<4\%$, no further data shown).

187 On return to the laboratory, 40 mL of water was injected into a purge and cryotrap system (Chuck *et al.*
188 *et al.*, 2005), filtered through a 25 mm Whatman glass fibre filter (GF/F; GE Healthcare Life Sciences,
189 Little Chalfont, England) and purged with oxygen-free nitrogen (OFN) at 80 mL min^{-1} for 10

190 minutes. Each gas sample passed through a glass wool trap to remove particles and aerosols, before
191 a dual nafion counterflow drier (180 mL min⁻¹ OFN) removed water vapour from the gas stream.
192 The gas sample was trapped in a stainless steel loop held at -150 °C in the headspace of a liquid
193 nitrogen-filled dewar. The sample was injected by immersion of the sample loop in boiling water
194 into an Agilent 6890 gas chromatograph equipped with a 60 m DB-VRX capillary column (0.32
195 mm ID, 1.8 µm film thickness, Agilent J&W Ltd) according to the programme outlined by Hopkins
196 *et al.* (2010). Analysis was performed by an Agilent 5973 quadrupole mass spectrometer operated
197 in electron ionisation, single ion mode. Liquid standards of CH₃I, diiodomethane (CH₂I₂), CH₂ClI,
198 iodoethane (C₂H₅I), iodopropane (C₃H₇I), CHBr₃, dibromoethane (CH₂Br₂), dibromochloromethane
199 (CHBr₂Cl), bromiodomethane (CH₂BrI) and DMS (Standards supplied by Sigma Aldrich Ltd, UK)
200 were gravimetrically prepared by dilution in HPLC-grade methanol (Table 2) and used for
201 calibration. The relative standard error was expressed as a percentage of the mean for the sample
202 analysis, calculated for each compound using triplicate analysis each day from a single mesocosm,
203 and was <7% for all compounds. GC-MS instrument drift was corrected by the use of a surrogate
204 analyte standard in every sample, comprising deuterated DMS (D₆-DMS), deuterated methyl iodide
205 (CD₃I) and ¹³C dibromoethane (¹³C₂H₄Br₂) via the method described in Hughes *et al.* (2006) and
206 Martino *et al.* (2005). Five-point calibrations were performed weekly for each compound with the
207 addition of the surrogate analyte, with a single standard analysed daily to check for instrument drift;
208 linear regression from calibrations typically produced r²>0.98. All samples measured within the
209 mesocosms were within the concentration ranges of the calibrations (Table 2).

210 **2.2.2 DMSP**

211 Samples for total DMSP (DMSP_T) were collected and stored for later analysis by the acidification
212 method of Curran *et al.* (1998). A 7 mL sub-sample was collected from the amber glass bottle into
213 an 8 mL glass sample vial (Labhut, Churcham, UK), into which 0.35 µL of 50% H₂SO₄ was added,
214 before storage at ambient temperature. Particulate DMSP (DMSP_P) samples were prepared by the
215 gravity filtration of 20 mL of sample through a 47 mm GF/F in a glass filter unit, before careful
216 removal and folding of the GF/F into a 7 mL sample vial filled with 7 mL of Milli-Q water and 0.35
217 µL of H₂SO₄ before storage at ambient temperature. Samples were stored for approximately 8
218 weeks prior to analysis. DMSP samples (total and particulate) were analysed on a PTFE purge and
219 cryotrap system using 2 mL of the sample purged with 1 mL of 10M NaOH for 5 minutes at 80 mL
220 min⁻¹. The sample gas stream passed through a glass wool trap and Nafion counterflow (Permapure)
221 drier before being trapped in a PTFE sample loop kept at -150 °C by suspension in the headspace of
222 a liquid nitrogen-filled dewar and controlled by feedback from a thermocouple. Immersion in

223 boiling water rapidly re-volatilised the sample for injection into a Shimadzu GC2010 gas
224 chromatograph with a Varian Chrompack CP-Sil-5CB column (30 m, 0.53 mm ID) and flame
225 photometric detector (FPD). The GC oven was operated isothermally at 60 °C which resulted in
226 DMS eluting at 2.1 minutes. Liquid DMSP standards were prepared and purged in the same manner
227 as the sample to provide weekly calibrations of the entire analytical system. Involvement in the
228 2013 AQA 12-23 international DMS analysis proficiency test (National Measurement Institute of
229 Australia, 2013) in February 2013 demonstrated excellent agreement between our method of DMSP
230 analysis and the mean from thirteen laboratories measuring DMS using different methods, with a
231 measurement error of 5%.

232 DMSP was not detected in any of the samples (total or particulate) collected and stored during the
233 experiment, and it was considered likely that this was due to an unresolved issue regarding
234 acidifying Baltic Sea samples for later DMSP analysis. This method had been used during a
235 previous mesocosm experiment (Bergen, Norway) and the results correlated well with those
236 measured immediately on a similar GC-FPD system (Webb *et al.* 2015). was considered unlikely
237 that rates of bacterial DMSP turnover through demethylation rather than through cleavage to
238 produce DMS (Curson *et al.*, 2011) were sufficiently high in the Baltic Sea to remove all detectable
239 DMSP, yet still produce measureable DMS concentrations. Also, rapid turnover of dissolved DMSP
240 in surface waters being the cause of low DMSP_T concentrations does not explain the lack of
241 intracellular particulate-phase DMSP. Although production of DMS is possible from alternate
242 sources, it is highly unlikely that there was a total absence of DMSP-producing phytoplankton
243 within the mesocosms or Baltic Sea surface waters around Tvärminne; DMSP has been measured in
244 surface waters of the Southern Baltic Sea at 22.2 nmol L⁻¹ in 2012, indicating that DMSP-producing
245 species are present within the Baltic Sea (Cathleen Zindler, GEOMAR, Pers. Comm.).

246 A previous study by del Valle *et al.* (2011) highlighted up to 94% loss of DMSP_T from acidified
247 samples of colonial *Phaeocystis globosa* culture, and field samples dominated by colonial
248 *Phaeocystis antarctica*. Despite filamentous, colonial cyanobacteria in the samples from Tvärminne
249 mesocosms potentially undergoing the same process, these species did not dominate the community
250 at only 6.6% of the total Chl-*a*, implying that the acidification method for DMSP fixation also failed
251 for unicellular phytoplankton species. The findings of this mesocosm study suggest that the
252 acidification method is unreliable in the Baltic Sea, and should be considered inadequate as the sole
253 method of DMSP fixation in future experiments in the region. The DMSP acidification method is
254 used worldwide as a simple and effective method of DMSP storage; the findings here, alongside
255 those of del Valle *et al.* (2011), question the applicability of this method in other marine

256 environments, and suggests significant testing prior to reliance on this method as a sole means of
257 DMSP storage.

258

259 **2.3 Measurement of carbonate chemistry and community dynamics**

260 Water samples were collected from the 10 m and 17 m IWS on a daily basis and analysed for
261 carbonate chemistry, fluorometric Chl-*a*, phytoplankton pigments (17 m IWS only) and cell
262 abundance to analyse the community structure and dynamics during the experiment. The carbonate
263 system was analysed through a suite of measurements (Paul *et al.*, 2015), including potentiometric
264 titration for total alkalinity (TA), infrared absorption for dissolved inorganic carbon (DIC) and
265 spectrophotometric determination for pH. For Chl-*a* analysis and pigment determination, 500 mL
266 sub-samples were filtered through a GF/F and stored frozen (-20 °C for two hours for Chl-*a* and -80
267 °C for up to 6 months for pigments), before homogenisation in 90 % acetone with glass beads. After
268 centrifuging (10 minutes at 800 x g at 4 °C) the Chl-*a* concentrations were determined using a
269 Turner AU-10 fluorometer by the methods of Welschmeyer (1994), and the phytoplankton pigment
270 concentrations by reverse phase high performance liquid chromatography (WATERS HPLC with a
271 Varian Microsorb-MV 100-3 C8 column) as described by Barlow *et al.* (1997). Phytoplankton
272 community composition was determined by the use of the CHEMTAX algorithm to convert the
273 concentrations of marker pigments to Chl-*a* equivalents (Mackey *et al.*, 1996; Schulz *et al.*, 2013).
274 Microbes were enumerated using a Becton Dickinson FACSCalibur flow cytometer (FCM)
275 equipped with a 488 nm argon laser (Crawford *et al.*, 2016) and counts of phytoplankton cells >20
276 µm were made on concentrated (50 mL) sample water, fixed with acidic Lugol's iodine solution
277 with an inverted microscope. Filamentous cyanobacteria were counted in 50 µm length units.

278 **2.4 Statistical Analysis**

279 All statistical analysis was performed using Minitab V16. In analysis of the measurements between
280 mesocosms, one-way ANOVA was used with Tukey's post-hoc analysis test to determine the effect
281 of different $f\text{CO}_2$ on concentrations measured in the mesocosms and the Baltic Sea (H_0 assumes no
282 significant difference in the mean concentrations of trace gases measured through the duration of
283 the experiment). Spearman's Rank Correlation Coefficients were calculated to compare the
284 relationships between trace gas concentrations, $f\text{CO}_2$, and a number of biological parameters, and
285 the resulting ρ -values for each correlation are given in Supplementary table S1 for the mesocosms
286 and S2 for the Baltic Sea data.

287

288 **3 Results and Discussion**

289 **3.1 Biogeochemical changes within the mesocosms**

290 The mesocosm experiment was split into three phases based on the temporal variation in Chl-*a* (Fig.
291 2; Paul *et al.*, 2015) evaluated after the experiment was completed:

- 292 • Phase 0 (days $t-5$ to $t0$) – pre-CO₂ addition
- 293 • Phase I (days $t1$ to $t16$) – ‘productive phase’
- 294 • Phase II (days $t17$ to $t30$) – temperature induced autotrophic decline.

295 **3.1.1 Physical Parameters**

296 $f\text{CO}_2$ decreased over Phase I in the three highest $f\text{CO}_2$ mesocosms, mainly through air-sea gas
297 exchange and carbon fixation by phytoplankton (Fig. 1a). All mesocosms still showed distinct
298 differences in $f\text{CO}_2$ levels throughout the experiment (Table 1), and there was no overlap of
299 mesocosm $f\text{CO}_2$ values on any given day, save for the two controls (M1 and M5). The control
300 mesocosm $f\text{CO}_2$ increased through Phase I of the experiment, likely as a result of undersaturation of
301 the water column encouraging dissolution of atmospheric CO₂ (Paul *et al.*, 2015). Salinity in the
302 mesocosms remained constant throughout the experiment at 5.70 ± 0.004 , and showed no variation
303 with depth (data not shown but available in Paul *et al.* 2015). It remained similar to salinity in the
304 Baltic Sea surrounding the mesocosms, which was 5.74 ± 0.14 . Water temperature varied from a
305 low of 8.6 ± 0.4 °C during Phase 0 to a high of 15.9 ± 2.2 °C measured on day $t16$, before
306 decreasing once again (Fig. 1b).

307 Summertime upwelling events are common and well described (Gidhagen, 1987; Lehmann and
308 Myrberg, 2008), and induce a significant temperature decrease in surface waters; such an event
309 appears to have commenced around $t16$, as indicated by significantly decreasing temperatures
310 inside and out of the mesocosms (Fig. 1b) and increased salinity in the Baltic Sea from 5.5 to 6.1
311 over the following 15 days to the end of the experiment. Due to the enclosed nature of the
312 mesocosms, the upwelling affected only the temperature and not pH, $f\text{CO}_2$ or the microbial
313 community. However, the temperature decrease after $t16$ was likely to have had a significant effect
314 on phytoplankton growth (and biogenic gas production), explaining the lower Chl-*a* in Phase II.

3.1.2 Community Dynamics

Mixing of the mesocosms and redistribution of the nutrients throughout the water column after closure (prior to $t-3$) did not trigger a notable increase in total Chl-*a* in Phase 0 as was identified in previous mesocosm experiments. During Phase I, light availability, combined with increasing water temperatures favoured the growth of phytoplankton in all mesocosms (Paul *et al.* 2015), and was unlikely to be a direct result of the CO₂ enrichment, as no difference was identified between enriched mesocosms and controls. Mean Chl-*a* during Phase I was 1.98 (\pm 0.29) $\mu\text{g L}^{-1}$ from all mesocosms, decreasing to 1.44 (\pm 0.46) $\mu\text{g L}^{-1}$ in Phase II: this decrease was attributed to a temperature induced decrease in phytoplankton growth rates and higher grazing rates as a result of higher zooplankton reproduction rates during Phase I (Lischka *et al.*, 2015; Paul *et al.*, 2015). Mesocosm Chl-*a* decreased until the end of the experiment on $t31$.

The largest contributors to Chl-*a* in the mesocosms during the summer of 2012 were the chlorophytes and cryptophytes, with up to 40% and 21% contributions to the Chl-*a* respectively (Table 3; Paul *et al.*, 2015). Significant long-term differences in abundance between mesocosms developed as a result of elevated $f\text{CO}_2$ in only two groups: picoeukaryotes I showed higher abundance at high $f\text{CO}_2$ ($F=8.2$, $p<0.01$; Crawford *et al.*, 2016 and Supplementary Fig. S2), as seen in previous mesocosm experiments (Brussaard *et al.*, 2013; Newbold *et al.*, 2012) and picoeukaryotes III the opposite trend ($F=19.6$, $p<0.01$; Crawford *et al.*, 2016). Temporal variation in phytoplankton abundance was similar between all mesocosms (Supplementary Fig. S1 and S2).

Diazotrophic, filamentous cyanobacterial blooms in the Baltic Sea are an annual event in summer (Finni *et al.*, 2001), and single-celled cyanobacteria have been found to comprise as much as 80% of the cyanobacterial biomass and 50% of the total primary production during the summer in the Baltic Sea (Stal *et al.*, 2003). However, CHEMTAX analysis identified cyanobacteria as contributing less than 10% of the total Chl-*a* in the mesocosms (Crawford *et al.*, 2016; Paul *et al.*, 2015). These observations were backed up by satellite observations showing reduced cyanobacterial abundance throughout the Baltic Sea in 2012 compared to previous and later years (Oberg, 2013). It was proposed that light availability and surface water temperatures during the summer of 2012 were sub-optimal for triggering a filamentous cyanobacteria bloom (Wasmund, 1997).

3.2 DMS and DMSP

3.2.1 Mesocosm DMS

A significant 34% reduction in DMS concentrations was detected in the high $f\text{CO}_2$ treatments during Phase II compared to the ambient $f\text{CO}_2$ mesocosms ($F=31.7$, $p<0.01$). Mean DMS concentrations of $5.0 (\pm 0.8)$; range $3.5 - 6.8$ nmol L^{-1} in the ambient treatments compared to $3.3 (\pm 0.3)$; range $2.9 - 3.9$ nmol L^{-1} in the 1333 and 1075 μatm mesocosms (Fig. 2a). The primary differences identified were apparent from the start of Phase II on $t17$, after which maximum concentrations were observed in the ambient mesocosms on $t21$. The relationship between DMS and increasing $f\text{CO}_2$ during Phase II was found to be linear (Fig. 2b), a finding also identified in previous mesocosm experiments (Archer *et al.*, 2013; Webb *et al.*, 2015). Furthermore, increases in DMS concentrations under high $f\text{CO}_2$ were delayed by three days relative to the ambient and medium $f\text{CO}_2$ treatments, a situation which has been observed in a previous mesocosm experiment. This was attributed to small-scale shifts in community composition and succession which could not be identified with only a once-daily measurement regime (Vogt *et al.*, 2008). DMS measured in all mesocosms fell within the range 2.7 to 6.8 nmol L^{-1} across the course of the experiment. During Phase I, no difference was identified in DMS concentrations between $f\text{CO}_2$ treatments with the mean of all mesocosms $3.1 (\pm 0.2)$ nmol L^{-1} . Concentrations in all mesocosms gradually declined from $t21$ until the end of DMS measurements on $t31$. DMS concentrations measured in the mesocosms and Baltic Sea were comparable to those measured in temperate coastal conditions in the North Sea (Turner *et al.*, 1988), the Mauritanian upwelling (Franklin *et al.*, 2009; Zindler *et al.*, 2012) and South Pacific (Lee *et al.*, 2010).

The majority of DMS production is presumed to be from DMSP. However, an alternative production route for DMS is available through the methylation of methanethiol (Drotar *et al.*, 1987; Kiene and Hines, 1995; Stets *et al.*, 2004) predominantly identified in anaerobic environments such as freshwater lake sediments (Lomans *et al.*, 1997), saltmarsh sediments (Kiene and Visscher, 1987) and microbial mats (Visscher *et al.*, 2003; Zinder *et al.*, 1977). Recent studies have also identified this pathway of DMS production from *Pseudomonas deceptionensis* in an aerobic environment (Carrión *et al.*, 2015), where *P. deceptionensis* was unable to synthesise or catabolise DMSP, but was able to enzymatically mediate DMS production from methanethiol (MeSH). The same enzyme has also been identified in a wide range of other bacterial taxa, including the cyanobacterial *Pseudanabaena*, which was identified in the Baltic Sea during this and previous investigations (Stuhr, pers. comm.; Kangro *et al.*, 2007; Nausch *et al.*, 2009). Correlations between

375 DMS and the cyanobacterial equivalent Chl-*a* ($\rho=0.42$, $p<0.01$; Supplementary Figure S1g) and
376 DMS and single-celled cyanobacteria ($\rho=0.58$, $p<0.01$; Supplementary Figure S2a) suggest that the
377 methylation pathway may be a potential source of DMS within the Baltic Sea community. In
378 addition to the methylation pathway, DMS production has been identified from S-methylmethionine
379 (Bentley and Chasteen, 2004), as well as from the reduction of dimethylsulphoxide (DMSO) in both
380 surface and deep waters by bacterial metabolism (Hatton *et al.*, 2004). As these compounds were
381 not measured in the mesocosms, it is impossible to determine if they were significant sources of
382 DMS.

383

384 3.2.2 DMS and Community Interactions

385 Throughout Phase I, DMS showed no correlation with any measured variables of biological activity
386 or cell abundance, and was unaffected by elevated $f\text{CO}_2$, indicating measured DMS concentrations
387 were not directly related to the perturbation of the system and associated cellular stress (Sunda *et al.*,
388 2002). Of the studied phytoplankton groupings, neither the cryptophytes or chlorophytes as the
389 largest contributors of Chl-*a* were identified as significant producers of DMSP. During Phase II,
390 DMS was negatively correlated with Chl-*a* in the ambient and medium $f\text{CO}_2$ mesocosms ($\rho=-0.60$,
391 $p<0.01$). During Phase II, a significant correlation was seen between DMS and single-celled
392 cyanobacteria identified predominantly as *Synechococcus* ($\rho=0.53$, $p<0.01$; Crawford *et al.* 2016
393 and supplementary table S1) and picoeukaryotes III ($\rho=0.75$, $p<0.01$). The peak in DMS
394 concentrations on *t*21 is unlikely to be a delayed response to the increased Chl-*a* on *t*16 due to the
395 time lag of 7 days. These higher DMS concentrations were likely connected to a peak in dissolved
396 organic carbon (DOC) on *t*15, as well as increasing bacterial abundance during Phase II (Hornick *et al.*,
397 2016). It is also likely that DMS concentrations increased as a response to the mesocosm wall
398 cleaning which took place on *t*16. The variation in inorganic nutrient concentrations between
399 mesocosms at the start of the experiment did not have an effect on DMS concentrations during
400 Phase I, and by the start of Phase II the variation between mesocosms had decreased.

401 In previous mesocosm experiments (Archer *et al.*, 2013; Hopkins *et al.*, 2010; Webb *et al.*, 2015),
402 DMS has shown poor correlations with many of the indicators of primary production and
403 phytoplankton abundance, as well as showing the same trend of decreased concentrations in high
404 $f\text{CO}_2$ mesocosms compared to ambient. DMS production is often uncoupled from measurements of
405 primary production in open waters (Lana *et al.*, 2012), and also often from production of its
406 precursor DMSP (Archer *et al.*, 2009). DMS and DMSP are important sources of sulphur and

407 carbon in the microbial food web for both bacteria and algae (Simó *et al.*, 2002, 2009), and since
408 microbial turnover of DMSP and DMS play a significant role in net DMS production, it is
409 unsurprising that DMS concentrations have shown poor correlation with DMSP-producing
410 phytoplankton groups in past experiments and open waters.

411 DMS concentrations have been reported lower under conditions of elevated $f\text{CO}_2$ compared to
412 ambient controls, in both mesocosm experiments (Table 4) and phytoplankton monocultures
413 (Arnold *et al.*, 2013; Avgoustidi *et al.*, 2012). However, the varying response of the community
414 within each experiment limit our ability to generalise the response of algal production of DMS and
415 DMSP in all situations due to the characteristic community dynamics of each experiment in specific
416 geographical areas and temporal periods. Previous experiments in the temperate Raunefjord of
417 Bergen, Norway, showed lower abundance of DMSP-producing algal species, and subsequently
418 DMSP-dependent DMS concentrations (Avgoustidi *et al.*, 2012; Hopkins *et al.*, 2010; Vogt *et al.*,
419 2008; Webb *et al.*, 2015). In contrast mesocosm experiments in the Arctic and Korea have shown
420 increased abundance of DMSP producers (Archer *et al.*, 2013; Kim *et al.*, 2010) but lower DMS
421 concentrations, while incubation experiments by Hopkins and Archer (2014) showed lower DMSP
422 production but higher DMS concentrations at high $f\text{CO}_2$. However, in all previous experiments with
423 DMSP as the primary precursor of DMS, elevated $f\text{CO}_2$ had a less marked effect on measured
424 DMSP concentrations than on measured DMS concentrations. Hopkins *et al.* (2010) suggested that
425 ‘the perturbation of the system has a greater effect on the processes that control the conversion of
426 DMSP to DMS rather than the initial production of DMSP itself’.

427 Previous mesocosm experiments have suggested significant links between increased bacterial
428 production through greater availability of organic substrates at high $f\text{CO}_2$ (Engel *et al.*, 2013;
429 Piontek *et al.*, 2013). Further, Endres *et al.* (2014) identified significant enhanced enzymatic
430 hydrolysis of organic matter with increasing $f\text{CO}_2$, with higher bacterial abundance. Higher
431 bacterial abundance will likely result in greater bacterial demand for sulphur, and therefore greater
432 consumption of DMS and conversion to DMSO. This was suggested as a significant sink for DMS
433 in a previous experiment (Webb *et al.*, 2015), but during the present experiment, both bacterial
434 abundance and bacterial production were lower at high $f\text{CO}_2$ (Hornick *et al.*, 2016). However, as it
435 has been proposed that only specialist bacterial groups are DMS consumers (Vila-Costa *et al.*,
436 2006b), and there is no determination of the DMS consumption characteristics of the bacterial
437 community in the Baltic Sea, it is not known if this loss pathway is stimulated at high $f\text{CO}_2$. As
438 microbial DMS yields can vary between 5-40% depending on the sulphur and carbon demand

439 (Kiene and Linn, 2000), a change in the bacterial sulphur requirements could change DMS turnover
440 despite lower abundance.

441 **3.3 Iodocarbons in the mesocosms and relationships with community composition**

442 Elevated $f\text{CO}_2$ did not affect the concentration of iodocarbons in the mesocosms significantly at any
443 time during the experiment, which is in agreement with the findings of Hopkins *et al.* (2013) in the
444 Arctic, but in contrast to Hopkins *et al.* (2010) and Webb (2015), where iodocarbons were measured
445 significantly lower under elevated $f\text{CO}_2$ (Table 4). Concentrations of all iodocarbons measured in
446 the mesocosms and the Baltic Sea fall within the range of those measured previously in the region
447 (Table 5). Mesocosm concentrations of CH_3I (Fig. 3a) and $\text{C}_2\text{H}_5\text{I}$ (Fig. 3b) showed concentration
448 ranges of 2.91 to 6.25 and 0.23 to 0.76 pmol L^{-1} respectively. CH_3I showed a slight increase in all
449 mesocosms during Phase I, peaking on $t16$ which corresponded with higher Chl-*a* concentrations,
450 and correlated throughout the entire experiment with picoeukaryote groups II ($\rho=0.59$, $p<0.01$) and
451 III ($\rho=0.23$, $p<0.01$; Crawford *et al.* 2016) and nanoeukaryotes I ($\rho=0.37$, $p<0.01$). Significant
452 differences identified between mesocosms for CH_3I were unrelated to elevated $f\text{CO}_2$ ($F=3.1$,
453 $p<0.05$), but concentrations were on average 15% higher in Phase II than Phase I. $\text{C}_2\text{H}_5\text{I}$ decreased
454 slightly during Phases I and II, although concentrations of this halocarbon were close to its
455 detection limit (0.2 pmol L^{-1}), remaining below 1 pmol L^{-1} at all times. As this compound showed
456 no significant effect of elevated $f\text{CO}_2$, and was identified by Orlikowska and Schulz-Bull (2009) as
457 having extremely low concentrations in the Baltic Sea (Table 5), it will not be discussed further.

458 No correlation was found between CH_3I and Chl-*a* at any phase, and the only correlation of any
459 phytoplankton grouping was with nanoeukaryotes II ($\rho=0.88$, $p<0.01$; Crawford *et al.*, 2016). These
460 CH_3I concentrations compare well to the 7.5 pmol L^{-1} measured by Karlsson *et al.* (2008) during a
461 cyanobacterial bloom in the Baltic Sea (Table 5), and the summer maximum of 16 pmol L^{-1}
462 identified by Orlikowska and Schulz-Bull (2009).

463 Karlsson *et al.* (2008) showed Baltic Sea halocarbon production occurring predominately during
464 daylight hours, with concentrations at night decreasing by 70% compared to late afternoon. Light
465 dependent production of CH_3I has been shown to take place through abiotic processes, including
466 radical recombination of CH_3 and I (Moore and Zafiriou, 1994). However, since samples were
467 integrated over the surface 10m of the water column, it was impossible to determine if
468 photochemistry was affecting iodocarbon concentrations near the surface where some UV light was
469 able to pass between the top of the mesocosm film material and the cover. For the same reason,
470 photodegradation of halocarbons (Zika *et al.*, 1984) within the mesocosms was also likely to have

471 been significantly restricted. Thus, as photochemical production was expected to be minimal,
472 biogenic production was likely to have been the dominant source of these compounds. Karlsson *et*
473 *al.* (2008) identified *Pseudanabaena* as a key producer of CH₃I in the Baltic Sea. However, the
474 abundance of *Pseudanabaena* was highest during Phase I of the experiment (A. Stühr, Pers.
475 Comm.) when CH₃I concentrations were lower, and as discussed previously, the abundance of these
476 species constituted only a very small proportion of the community. Previous investigations in the
477 laboratory have identified diatoms as significant producers of CH₃I (Hughes *et al.*, 2013; Manley
478 and De La Cuesta, 1997), and the low, steady-state abundance of the diatom populations in the
479 mesocosms could have produced the same relatively steady-state trends in the iodocarbon
480 concentrations.

481 Measured in the range 57.2 – 202.2 pmol L⁻¹ in the mesocosms, CH₂I₂ (Fig. 3c) showed the clearest
482 increase in concentration during Phase II, when it peaked on *t*21 in all mesocosms, with a maximum
483 of 202.2 pmol L⁻¹ in M5 (348 µatm). During Phase II, concentrations of CH₂I₂ were 57% higher
484 than Phase I, and were therefore negatively correlated with Chl-*a*. The peak on *t*21 corresponds
485 with the peak identified in DMS on *t*21, and concentrations through all three phases correlate with
486 picoeukaryotes II ($\rho=0.62$, $p<0.01$) and III ($\rho=0.47$, $p<0.01$) and nanoeukaryotes I ($\rho=0.88$, $p<0.01$;
487 Crawford *et al.*, 2015). CH₂ClI (Fig. 3d) showed no peaks during either Phase I or Phase II,
488 remaining within the range 3.81 to 8.03 pmol L⁻¹, and again correlated with picoeukaryotes groups
489 II ($\rho=0.34$, $p<0.01$) and III ($\rho=0.38$, $p<0.01$). These results may suggest that these groups possessed
490 halo-peroxidase enzymes able to oxidise I⁻, most likely as an anti-oxidant mechanism within the cell
491 to remove H₂O₂ (Butler and Carter-Franklin, 2004; Pedersen *et al.*, 1996; Theiler *et al.*, 1978).
492 However, given the lack of response of these compounds to elevated *f*CO₂ ($F=1.7$, $p<0.01$), it is
493 unlikely that production was increased in relation to elevated *f*CO₂. Production of all iodocarbons
494 increased during Phase II when total Chl-*a* decreased, particularly after the walls of the mesocosms
495 were cleaned for the first time, releasing significant volumes of organic aggregates into the water
496 column. Aggregates have been suggested as a source of CH₃I and C₂H₅I (Hughes *et al.*, 2008),
497 likely through the alkylation of inorganic iodide (Urhahn and Ballschmiter, 1998) or through the
498 breakdown of organic matter by microbial activity to supply the precursors required for iodocarbon
499 production (Smith *et al.*, 1992). Hughes *et al.* (2008) did not identify this route as a pathway for
500 CH₂I₂ or CH₂ClI production, but Carpenter *et al.* (2005) suggested a production pathway for these
501 compounds through the reaction of HOI with aggregated organic materials.

3.4 Bromocarbons in the mesocosms and the relationships with community composition

No effect of elevated $f\text{CO}_2$ was identified for any of the three bromocarbons, which compared with the findings from previous mesocosms where bromocarbons were studied (Hopkins *et al.*, 2010, 2013; Webb, 2015; Table 4). Measured concentrations were comparable to those of Orlikowska and Schulz-Bull (2009) and Karlsson *et al.* (2008) measured in the Southern part of the Baltic Sea (Table 3). The concentrations of CHBr_3 , CH_2Br_2 and CHBr_2Cl showed no major peaks of production in the mesocosms. CHBr_3 (Fig. 4a) decreased rapidly in all mesocosms over Phase 0 from a maximum measured concentration of $147.5 \text{ pmol L}^{-1}$ in M1 (mean of $138.3 \text{ pmol L}^{-1}$ in all mesocosms) to a mean of $85.7 (\pm 8.2 \text{ s.d.}) \text{ pmol L}^{-1}$ in all mesocosms for the period t_0 to t_{31} (Phases I and II). The steady-state CHBr_3 concentrations indicated a production source, however there was no clear correlation with any measured algal groups. CH_2Br_2 concentrations (Fig. 4b) decreased steadily in all mesocosms from t_{-3} through to t_{31} , over the range 4.0 to 7.7 pmol L^{-1} , and CHBr_2Cl followed a similar trend in the range 1.7 to 4.7 pmol L^{-1} (Fig. 4c). Of the three bromocarbons, only CH_2Br_2 showed correlation with total Chl-*a* ($\rho=0.52$, $p<0.01$), and with cryptophyte ($\rho=0.86$, $p<0.01$) and dinoflagellate ($\rho=0.65$, $p<0.01$) derived Chl-*a*. Concentrations of CH_2BrI were below detection limit for the entire experiment.

CH_2Br_2 showed positive correlation with Chl-*a* ($\rho=0.52$, $p<0.01$), nanoeukaryotes II ($\rho=0.34$, $p<0.01$) and cryptophytes ($\rho=0.86$, $p<0.01$; see supplementary material), whereas CHBr_3 and CHBr_2Cl showed very weak or no correlation with any indicators of algal biomass. Schall *et al.* (1997) have proposed that CHBr_2Cl is produced in seawater by the nucleophilic substitution of bromide by chloride in CHBr_3 , which given the steady-state concentrations of CHBr_3 would explain the similar distribution of CHBr_2Cl concentrations. Production of all three bromocarbons was identified from large-size cyanobacteria such as *Aphanizomenon flos-aquae* by Karlsson *et al.* (2008), and in addition, significant correlations were found in the Arabian Sea between the abundance of the cyanobacterium *Trichodesmium* and several bromocarbons (Roy *et al.*, 2011), and the low abundance of such bacteria in the mesocosms would explain the low variation in bromocarbon concentrations through the experiment.

Halocarbon loss processes such as nucleophilic substitution (Moore, 2006), hydrolysis (Elliott and Rowland, 1995), sea-air exchange and microbial degradation are suggested as of greater importance than production of these compounds by specific algal groups, particularly given the relatively low growth rates and low net increase in total Chl-*a*. Hughes *et al.* (2013) identified bacterial inhibition of CHBr_3 production in laboratory cultures of *Thalassiosira* diatoms, but that it was not subject to

bacterial breakdown; which could explain the relative steady state of CHBr_3 concentrations in the mesocosms. In contrast, significant bacterial degradation of CH_2Br_2 in the same experiments could explain the steady decrease in CH_2Br_2 concentrations seen in the mesocosms. Bacterial oxidation was also identified by Goodwin *et al.* (1998) as a significant sink for CH_2Br_2 . As discussed for the iodocarbons, photolysis was unlikely due to the UV absorption of the mesocosm film, and limited UV exposure of the surface waters within the mesocosm due to the mesocosm cover. The ratio of CH_2Br_2 to CHBr_3 was also unaffected by increased $f\text{CO}_2$, staying within the range 0.04 to 0.08. This range in ratios is consistent with that calculated by Hughes *et al.* (2009) in the surface waters of an Antarctic depth profile, and attributed to higher sea-air flux of CHBr_3 than CH_2Br_2 due to a greater concentrations gradient, despite the similar transfer velocities of the two compounds (Quack *et al.*, 2007). Using cluster analysis in a time-series in the Baltic Sea, Orlikowska and Schulz-Bull (2009) identified both these compounds as originating from different sources and different pathways of production.

Macroalgal production would not have influenced the mesocosm concentrations after the bags were sealed due to the isolation from the coastal environment. However macroalgal production into the water column prior to mesocosm installation (Klick, 1992; Leedham *et al.*, 2013; Moore and Tokarczyk, 1993) could account for the high initial concentrations with concentrations decreasing through the duration of the experiment via turnover and transfer to the atmosphere.

3.5 Natural variations in Baltic Sea $f\text{CO}_2$ and the effect on biogenic trace gases

3.5.1 Physical variation and community dynamics

Baltic Sea deep waters have high $f\text{CO}_2$ and subsequently lower pH (Schneider *et al.*, 2002), and the influx to the surface waters surrounding the mesocosms resulted in $f\text{CO}_2$ increasing to 725 μatm on t_{31} , close to the average $f\text{CO}_2$ of the third highest mesocosm (M6: 868 μatm). The input of upwelled water into the region mid-way through the experiment significantly altered the biogeochemical properties of the waters surrounding the mesocosms, and as a result it is inappropriate to directly compare the community structure and trace gas production of the Baltic Sea and the mesocosms. These conditions imply that pelagic communities in the Baltic Sea are regularly exposed to rapid changes in $f\text{CO}_2$ and the associated pH, as well as having communities associated with the elevated $f\text{CO}_2$ conditions. The changes in biological parameters and trace gas

565 concentrations are therefore discussed here separately from the concentrations measured in the
566 mesocosms.

567 Given the separation of the waters within the mesocosms, and the movement of water masses within
568 the Baltic Sea, it is expected that phytoplankton population structure could be significantly different
569 inside the mesocosms compared to the external waters. Chl-*a* followed the pattern of the
570 mesocosms until *t*₄, after which concentrations were significantly higher than any mesocosm,
571 peaking at 6.48 µg L⁻¹ on *t*₁₆, corresponding to the maximum Chl-*a* peak in the mesocosms and the
572 maximum peak of temperature. As upwelled water intruded into the surface waters, the surface Chl-
573 *a* was diluted with low Chl-*a* deep water: Chl-*a* in the surface 10m decreased from around *t*₁₆ at the
574 start of the upwelling until *t*₃₁ when concentrations were once again equivalent to those found in
575 the mesocosms at 1.30 µg L⁻¹. In addition, there was potential introduction of different algal groups
576 to the surface, but chlorophytes and cryptophytes were the major contributors to the Chl-*a* in the
577 Baltic Sea, as in the mesocosms. Cyanobacteria contributed less than 2% of the total Chl-*a* in the
578 Baltic Sea (Crawford *et al.*, 2016; Paul *et al.*, 2015).

579 Temporal community dynamics in the Baltic Sea were very different to that in the mesocosms
580 across the experiment, with euglenophytes, chlorophytes, diatoms and prasinophytes all showing
581 distinct peaks at the start of Phase II, with these same peaks identified in the nanoeukaryotes I and
582 II, and picoeukaryotes II (Crawford *et al.*, 2016; Paul *et al.*, 2015; Supplementary Figs. S1 and S2).
583 The decrease in abundance of many groups during Phase II was attributed to the decrease in
584 temperature and dilution with low-abundance deep waters.

585 3.5.2 DMS in the Baltic Sea

586 The Baltic Sea samples gave a mean DMS concentration of 4.6 ± 2.6 nmol L⁻¹ but peaked at 11.2
587 nmol L⁻¹ on *t*₁₆, and were within the range of previous measurements for the region (Table 5).
588 Strong correlations were seen between DMS and Chl-*a* ($\rho=0.84$, $p<0.01$), with the ratio of DMS:
589 Chl-*a* at 1.6 (± 0.3) nmol µg⁻¹. Other strong correlations were seen with euglenophytes ($\rho=0.89$,
590 $p<0.01$), dinoflagellates ($\rho=0.61$, $p<0.05$) and nanoeukaryotes II ($\rho=0.88$, $p<0.01$), but no
591 correlation was found between DMS and cyanobacterial abundance, or with picoeukaryotes III
592 which was identified in the mesocosms, suggesting that DMS had a different origin in the Baltic
593 Sea community than in the mesocosms. In addition, the community demands of sulphur are likely to
594 be very different in the Baltic Sea compared to the mesocosms, due to differences in community
595 composition and sulphur availability, and therefore direct comparisons with mesocosm
596 concentrations are inappropriate.

597 As CO₂ levels increased after *t*16 the DMS concentration measured in the Baltic Sea decreased,
598 from the peak on *t*16 to the lowest recorded sample of the entire experiment at 1.85 nmol L⁻¹ on *t*31.
599 As with Chl-*a*, DMS concentrations in the surface of the Baltic Sea may have been diluted with
600 low-DMS deep water

601 3.5.3 Halocarbon concentrations in the Baltic Sea

602 Outside the mesocosms in the Baltic Sea, CH₃I was measured at a maximum concentration of 8.65
603 pmol L⁻¹, during Phase II, and showed limited effect of the upwelling event. Both CH₂I₂ and
604 CH₂ClI showed higher concentrations in the Baltic Sea samples than the mesocosms (CH₂I₂: 373.9
605 pmol L⁻¹ and CH₂ClI: 18.1 pmol L⁻¹), and were correlated with the euglenophytes (CH₂I₂; ρ =0.63,
606 p <0.05 and CH₂ClI; ρ =0.68, p <0.01) and nanoeukaryotes II (CH₂I₂; ρ =0.53, p <0.01 and CH₂ClI;
607 ρ =0.58, p <0.01), but no correlation with Chl-*a*. Both polyhalogenated compounds showed
608 correlation with picoeukaryote groups II and III, indicating that production was probably not limited
609 to a single source. These concentrations of CH₂I₂ and CH₂ClI compared well to those measured
610 over a macroalgal bed in the higher saline waters of the Kattegat by Klick and Abrahamsson (1992),
611 suggesting that macroalgae were a significant iodocarbon source in the Baltic Sea. Macroalgal
612 production in the Baltic Sea is likely the predominant iodocarbon source, compared to the
613 mesocosms where macroalgae are excluded.

614 As with the iodocarbons, the Baltic Sea showed significantly higher concentrations of CHBr₃
615 (F =28.1, p <0.01), CH₂Br₂ (F =208.8, p <0.01) and CHBr₂Cl (F =23.5, p <0.01) than the mesocosms,
616 with maximum concentrations 191.6 pmol L⁻¹, 10.0 pmol L⁻¹ and 5.0 pmol L⁻¹ respectively. In the
617 Baltic Sea, only CHBr₃ was correlated with Chl-*a* (ρ =0.65, p <0.05), cyanobacteria (ρ =0.61, p <0.01;
618 Paul *et al.*, 2015) and nanoeukaryotes II (ρ =0.56, p <0.01; Crawford *et al.*, 2016), with the other two
619 bromocarbons showing little to no correlations with any parameter of community activity.
620 Production of bromocarbons from macroalgal sources (Laternus *et al.*, 2000; Leedham *et al.*, 2013;
621 Manley *et al.*, 1992) was likely a significant contributor to the concentrations detected in the Baltic
622 Sea; over the macroalgal beds in the Kattegat, Klick (1992) measured concentrations an order of
623 magnitude higher than seen in this experiment for CH₂Br₂ and CHBr₂Cl. There was only a slight
624 increase in bromocarbon concentrations as a result of the upwelling, indicating that the upwelled
625 water had similar concentrations to the surface waters. These data from the Baltic Sea are presented
626 as an important time-series of halocarbon measurements during the summer of 2012, which are
627 expected to add to existing Baltic Sea trace gas datasets.

628

629 **4 The Baltic Sea as a natural analogue to future ocean acidification?**

630 Mesocosm experiments are a highly valuable tool in assessing the potential impacts of elevated CO₂
631 on complex marine communities, however they are limited in that the rapid change in *f*CO₂
632 experienced by the community may not be representative of changes in the future ocean (Passow
633 and Riebesell, 2005). This inherent problem with mesocosm experiments can be overcome through
634 using naturally low pH/ high CO₂ areas such as upwelling regions or vent sites (Hall-Spencer *et al.*,
635 2008), which can give an insight into populations already living and acclimated to high CO₂
636 regimes by exposure over timescales measured in years. This mesocosm experiment was performed
637 at such a location with a relatively high *f*CO₂ excursion, however still low compared to some sites
638 (800 µatm compared to >2000 µatm; Hall-Spencer *et al.*, 2008), and it was clear through the
639 minimal variation in Chl-*a* between all mesocosms that the community was relatively unaffected by
640 elevated *f*CO₂, although variation could be identified in some phytoplankton groups and some shifts
641 in community composition. The upwelling event occurring mid-way through our experiment
642 allowed comparison of the mesocosm findings with a natural analogue of the system, as well as
643 showing the extent to which the system perturbation can occur (up to 800 µatm). This event was a
644 fortuitous occurrence during this mesocosm experiment, but as the scale and timing of these
645 upwelling events is difficult to determine, and therefore these upwelling events are extremely
646 challenging to study as natural high CO₂ analogues.

647 In this paper, we described the temporal changes in concentrations of DMS and halocarbons in
648 natural Baltic phytoplankton communities exposed to elevated *f*CO₂ treatments. In contrast to the
649 halocarbons, concentrations of DMS were significantly lower in the highest *f*CO₂ treatments
650 compared to the control. Despite very different physicochemical and biological characteristics of
651 the Baltic Sea (e.g. salinity, community composition and nutrient concentrations), this is a very
652 similar outcome to that seen in several other high *f*CO₂ experiments. The Baltic Sea trace gas
653 samples give a good record of trace gas cycling during the injection of high *f*CO₂ deep water into
654 the surface community during upwelling events. For the concentrations of halocarbons, the
655 measured concentrations did not change during the upwelling event in the Baltic Sea, which may
656 indicate that emissions of organic iodine and bromine are unlikely to change with future
657 acidification of the Baltic Sea without significant alteration to the meteorological conditions.
658 Further studies of these compounds are important to determine rates of production and consumption
659 to include in prognostic and predictive models. However, net production of organic sulphur within
660 the Baltic Sea region is likely to decrease with an acidified future ocean scenario, despite the
661 possible acclimation of the microbial community to elevated *f*CO₂. This will potentially impact the

662 flux of DMS to the atmosphere over Northern Europe, and could have significant impacts on the
663 local climate through the reduction of atmospheric sulphur aerosols. Data from a previous
664 mesocosm experiment has been used to estimate future global changes in DMS production, and
665 predicted that global warming would be amplified (Six *et al.*, 2013); utilising the data from this
666 experiment combined with those of other mesocosm, field and laboratory experiments and
667 associated modelling provide the basis for a better understanding of the future changes in global
668 DMS production and their climatic impacts.

669
670

671 **Acknowledgements**

672 The Tvärminne 2012 mesocosm experiment was part of the SOPRAN II (Surface Ocean Processes
673 in the Anthropocene) Programme (FKZ 03F0611) and BIOACID II (Biological Impacts of Ocean
674 Acidification) project (FKZ 03F06550), funded by the German Ministry for Education and
675 Research (BMBF) and led by the GEOMAR Helmholtz Centre for Ocean Research Kiel, Germany.
676 The authors thank all participants in the SOPRAN Tvärminne experiment for their assistance,
677 including A. Ludwig for logistical support, the diving team, and the staff of Tvärminne Zoological
678 Research Station for hosting the experiment. We also acknowledge the captain and crew of RV
679 *ALKOR* (**AL394 and AL397**) for their work transporting, deploying and recovering the mesocosms.

680 This work was funded by a UK Natural Environment Research Council Directed Research
681 Studentship (NE/H025588/1) through the UK Ocean Acidification Research Programme, with
682 CASE funding from Plymouth Marine Laboratory. Additional funding was supplied by the EU
683 Seventh Framework Program (FP7/2007-2013) MESOAQUA (EC Contract No. 228224).

684

685 Archer, S. D., Cummings, D., Llewellyn, C. and Fishwick, J.: Phytoplankton taxa, irradiance and nutrient availability
686 determine the seasonal cycle of DMSP in temperate shelf seas, *Mar. Ecol. Prog. Ser.*, 394, 111–124,
687 doi:10.3354/meps08284, 2009.

688 Archer, S. D., Kimmance, S. A., Stephens, J. A., Hopkins, F. E., Bellerby, R. G. J., Schulz, K. G., Piontek, J. and Engel,
689 A.: Contrasting responses of DMS and DMSP to ocean acidification in Arctic waters, *Biogeosciences*, 10(3), 1893–
690 1908, doi:10.5194/bg-10-1893-2013, 2013.

691 Arnold, H. E., Kerrison, P. and Steinke, M.: Interacting effects of ocean acidification and warming on growth and
692 DMS-production in the haptophyte coccolithophore *Emiliania huxleyi*., *Glob. Chang. Biol.*, 19(4), 1007–16,
693 doi:10.1111/gcb.12105, 2013.

694 Avgoustidi, V., Nightingale, P. D., Joint, I., Steinke, M., Turner, S. M., Hopkins, F. E. and Liss, P. S.: Decreased
695 marine dimethyl sulfide production under elevated CO₂ levels in mesocosm and in vitro studies, *Environ. Chem.*, 9(4),
696 399–404, doi:10.1071/EN11125, 2012.

697 Barlow, R. G., Cummings, D. G. and Gibb, S. W.: Improved resolution of mono- and divinyl chlorophylls a and b and
698 zeaxanthin and lutein in phytoplankton extracts using reverse phase C-8 HPLC, *Mar. Ecol. Prog. Ser.*, 161, 303–307,
699 1997.

700 Bentley, R. and Chasteen, T. G.: Environmental VOSCs—formation and degradation of dimethyl sulfide, methanethiol
701 and related materials, *Chemosphere*, 55(3), 291–317, doi:10.1016/j.chemosphere.2003.12.017, 2004.

702 Brussaard, C. P. D., Noordeloos, A. A. M., Witte, H., Collenteur, M. C. J., Schulz, K., Ludwig, A. and Riebesell, U.:
703 Arctic microbial community dynamics influenced by elevated CO₂ levels, *Biogeosciences*, 10(2), 719–731,
704 doi:10.5194/bg-10-719-2013, 2013.

705 Brutemark, A., Engström-Öst, J. and Vehmaa, A.: Long-term monitoring data reveal pH dynamics, trends and
706 variability in the western Gulf of Finland, *Oceanol. Hydrobiol. Stud.*, 40(3), 91–94, doi:10.2478/s13545-011-0034-3,
707 2011.

708 Butler, A. and Carter-Franklin, J. N.: The role of vanadium bromoperoxidase in the biosynthesis of halogenated marine
709 natural products, *Nat. Prod. Rep.*, 21(1), 180–188, doi:10.1039/b302337k, 2004.

710 Canadell, J. G., Le Quéré, C., Raupach, M. R., Field, C. B., Buitenhuis, E. T., Ciais, P., Conway, T. J., Gillett, N. P.,
711 Houghton, R. A. and Marland, G.: Contributions to accelerating atmospheric CO₂ growth from economic activity,
712 carbon intensity, and efficiency of natural sinks., *Proc. Natl. Acad. Sci. U. S. A.*, 104(47), 18866–18870,
713 doi:10.1073/pnas.0702737104, 2007.

714 Carpenter, L. J., Hopkins, J. R., Jones, C. E., Lewis, A. C., Parthipan, R., Wevill, D. J., Poissant, L., Pilote, M. and
715 Constant, P.: Abiotic source of reactive organic halogens in the sub-arctic atmosphere?, *Environ. Sci. Technol.*, 39(22),
716 8812–8816 [online] Available from: <http://www.ncbi.nlm.nih.gov/pubmed/16323781>, 2005.

717 Carrión, O., Curson, A. R. J., Kumaresan, D., Fu, Y., Lang, A. S., Mercadé, E. and Todd, J. D.: A novel pathway
718 producing dimethylsulphide in bacteria is widespread in soil environments, *Nat. Commun.*, 6, 6579,
719 doi:10.1038/ncomms7579, 2015.

720 Chance, R., Baker, A. R., Küpper, F. C., Hughes, C., Kloareg, B. and Malin, G.: Release and transformations of
721 inorganic iodine by marine macroalgae, *Estuar. Coast. Shelf Sci.*, 82, 406–414, doi:10.1016/j.ecss.2009.02.004, 2009.

722 Charlson, R. J., Lovelock, J. E., Andreae, M. O. and Warren, S. G.: Oceanic phytoplankton, atmospheric sulphur, cloud
723 albedo and climate, *Nature*, 326(6114), 655–661 [online] Available from:
724 http://www.atmos.washington.edu/~sgw/PAPERS/1987_CLAW.pdf (Accessed 15 July 2011), 1987.

725 Chuck, A. L., Turner, S. M. and Liss, P. S.: Oceanic distributions and air-sea fluxes of biogenic halocarbons in the open
726 ocean, *J. Geophys. Res.*, 110(C10022), doi:10.1029/2004JC002741, 2005.

727 Collins, M., Knutti, R., Arblaster, J., Dufresne, J.-L., Fichet, T., Frielingstein, P., Gao, X., Gutowski, W. J., Johns, T.,
728 Krinner, G., Shongwe, M., Tebaldi, C., Weaver, A. J. and Wehner, M.: Long-term climate change: projections,
729 commitments and irreversibility, in *Climate Change 2013: The Physical Science Basis. Contribution of Working Group*
730 *1 to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, edited by T. Stocker, D. Qin, G.-
731 K. Plattner, M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, and P. M. Midgley, Cambridge University
732 Press, Cambridge, UK. [online] Available from: <http://dia.academielouvain.be/handle/boreal:140396> (Accessed 12
733 June 2014), 2013.

734 Crawford, K., Brussaard, C. P. D. and Riebesell, U.: Shifts in the microbial community in the Baltic Sea with increasing
735 CO₂, *Biogeosciences*, In Press, 2016.

- Cubasch, U., Wuebbles, D., Chen, D., Facchini, M. C., Frame, D., Mahowald, N. and Winther, J.-G.: Introduction, in *Climate Change 2013: The Physical Science Basis. Contribution of Working Group 1 to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, edited by T. Stocker, D. Qin, G.-K. Plattner, M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, and P. M. Midgley, Cambridge University Press, Cambridge, UK., 2013.
- Curran, M. A. J., Jones, G. B. and Burton, H.: Spatial distribution of dimethylsulfide and dimethylsulfoniopropionate in the Australasian sector of the Southern Ocean, *J. Geophys. Res.*, 103(D13), 16677 – 16689, 1998.
- Curson, A. R. J., Todd, J. D., Sullivan, M. J. and Johnston, A. W. B.: Catabolism of dimethylsulphoniopropionate: microorganisms, enzymes and genes, *Nat. Rev. Microbiol.*, 9(12), 849–859, doi:10.1038/nrmicro2653, 2011.
- Czerny, J., Schulz, K. G., Krug, S. A., Ludwig, A. and Riebesell, U.: Technical Note: The determination of enclosed water volume in large flexible-wall mesocosms “KOSMOS,” *Biogeosciences*, 10, 1937–1941, doi:10.5194/bg-10-1937-2013, 2013.
- Doney, S. C., Fabry, V. J., Feely, R. A. and Kleypas, J. A.: Ocean acidification: the other CO₂ problem., *Ann. Rev. Mar. Sci.*, 1, 169–192, doi:10.1146/annurev.marine.010908.163834, 2009.
- Drotar, A., Burton, G. A., Tavernier, J. E. and Fall, R.: Widespread occurrence of bacterial thiol methyltransferases and the biogenic emission of methylated sulfur gases, *Appl. Environ. Microbiol.*, 53(7), 1626–1631 [online] Available from: <http://aem.asm.org/content/53/7/1626.short> (Accessed 25 March 2014), 1987.
- Elliot, S. and Rowland, F. S.: Methyl halide hydrolysis rates in natural waters, *J. Atmos. Chem.*, 20, 229–236, 1995.
- Endres, S., Galgani, L., Riebesell, U., Schulz, K.-G. and Engel, A.: Stimulated bacterial growth under elevated pCO₂: results from an off-shore mesocosm study., *PLoS One*, 9(6), e99228, doi:10.1371/journal.pone.0099228, 2014.
- Engel, A., Schulz, K. G., Riebesell, U., Bellerby, R. G. J., Delille, B. and Schartau, M.: Effects of CO₂ on particle size distribution and phytoplankton abundance during a mesocosm bloom experiment (PeECE II), *Biogeosciences*, 5(2), 509–521, doi:10.5194/bg-5-509-2008, 2008.
- Engel, A., Borchard, C., Piontek, J., Schulz, K. G., Riebesell, U. and Bellerby, R. G. J.: CO₂ increases ¹⁴C primary production in an Arctic plankton community, *Biogeosciences*, 10(3), 1291–1308, doi:10.5194/bg-10-1291-2013, 2013.
- Finni, T., Kononen, K., Olsonen, R. and Wallström, K.: The History of Cyanobacterial Blooms in the Baltic Sea, *AMBIO A J. Hum. Environ.*, 30(4), 172–178, doi:10.1579/0044-7447-30.4.172, 2001.
- Franklin, D. J., Poulton, A. J., Steinke, M., Young, J., Peeken, I. and Malin, G.: Dimethylsulphide, DMSP-lyase activity and microplankton community structure inside and outside of the Mauritanian upwelling, *Prog. Oceanogr.*, 83(1-4), 134–142, doi:10.1016/j.pocean.2009.07.011, 2009.
- Gidhagen, L.: Coastal upwelling in the Baltic Sea—Satellite and in situ measurements of sea-surface temperatures indicating coastal upwelling, *Estuar. Coast. Shelf Sci.*, 24, 449–462 [online] Available from: <http://www.sciencedirect.com/science/article/pii/0272771487901272> (Accessed 16 August 2014), 1987.
- Goodwin, K., Schaefer, J. K. and Oremland, R. S.: Bacterial oxidation of dibromomethane and methyl bromide in natural waters and enrichment cultures, *Appl. Environ. Microbiol.*, 64(12), 4629 –4636 [online] Available from: <http://aem.asm.org/content/64/12/4629.short> (Accessed 30 July 2014), 1998.
- Goodwin, K. D., North, W. J. and Lidstrom, M. E.: Production of bromoform and dibromomethane by giant kelp: factors affecting release and comparison to anthropogenic bromine sources, *Limnol. Oceanogr.*, 42(8), 1725–1734, doi:10.4319/lo.1997.42.8.1725, 1997.
- Hall-Spencer, J. M., Rodolfo-Metalpa, R., Martin, S., Ransome, E., Fine, M., Turner, S. M., Rowley, S. J., Tedesco, D. and Buia, M.-C.: Volcanic carbon dioxide vents show ecosystem effects of ocean acidification, *Nature*, 454(7200), 96–99, doi:10.1038/nature07051, 2008.
- Hartmann, D. L., Klein Tank, A. M. G., Rusticucci, M., Alexander, L. V., Bronnimann, S., Charabi, Y., Dentener, F. J., Dlugokencky, E. J., Easterling, D. R., Kaplan, A., Soden, B. J., Thorne, P. W., Wild, M. and Zhai, P. M.: Observations: Atmosphere and Surface, in *Climate Change 2013: The Physical Science Basis. Contribution of Working Group 1 to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, edited by T. F. Stocker, D. Qin, G.-K. Plattner, M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, and P. M. Midgley, Cambridge University Press, Cambridge, Cambridge, UK., 2013.
- Hatton, A. D., Darroch, L. and Malin, G.: The role of dimethylsulphoxide in the marine biogeochemical cycle of dimethylsulphide, *Oceanogr. Mar. Biol. Annu. Rev.*, 42, 29–56, 2004.

785 Hofmann, G. E., Barry, J. P., Edmunds, P. J., Gates, R. D., Hutchins, D. A., Klinger, T. and Sewell, M. A.: The effect of
786 ocean acidification on calcifying organisms in marine ecosystems: an organism-to-ecosystem perspective, *Annu. Rev.*
787 *Ecol. Evol. Syst.*, 41(1), 127–147, doi:10.1146/annurev.ecolsys.110308.120227, 2010.

788 Hopkins, F. E. and Archer, S. D.: Consistent increase in dimethyl sulphide (DMS) in response to high CO₂ in five
789 shipboard bioassays from contrasting NW European waters, *Biogeosciences*, 11(2), 4925 – 4940, doi:10.5194/bg-11-
790 2267-2014, 2014.

791 Hopkins, F. E., Turner, S. M., Nightingale, P. D., Steinke, M., Bakker, D. and Liss, P. S.: Ocean acidification and
792 marine trace gas emissions., *Proc. Natl. Acad. Sci. U. S. A.*, 107(2), 760–765, doi:10.1073/pnas.0907163107, 2010.

793 Hopkins, F. E., Kimmance, S. A., Stephens, J. A., Bellerby, R. G. J., Brussaard, C. P. D., Czerny, J., Schulz, K. G. and
794 Archer, S. D.: Response of halocarbons to ocean acidification in the Arctic, *Biogeosciences*, 10(4), 2331–2345,
795 doi:10.5194/bg-10-2331-2013, 2013.

796 Hornick, T., Bach, L. T., Crawford, K. J., Spilling, K., Achterberg, E. P., Brussaard, C. P. D., Riebesell, U. and
797 Grossart, H.-P.: Ocean acidification indirectly alters trophic interaction of heterotrophic bacteria at low nutrient
798 conditions, *Biogeosciences Discuss.*, (March), 1–37, doi:10.5194/bg-2016-61, 2016.

799 Hughes, C., Malin, G., Nightingale, P. D. and Liss, P. S.: The effect of light stress on the release of volatile iodocarbons
800 by three species of marine microalgae, *Limnol. Oceanogr.*, 51(6), 2849–2854 [online] Available from:
801 <http://cat.inist.fr/?aModele=afficheN&cpsidt=18312251> (Accessed 26 July 2013), 2006.

802 Hughes, C., Malin, G., Turley, C. M., Keely, B. J., Nightingale, P. D. and Liss, P. S.: The production of volatile
803 iodocarbons by biogenic marine aggregates, *Limnol. Oceanogr.*, 53(2), 867–872, 2008.

804 Hughes, C., Chuck, A. L., Rossetti, H., Mann, P. J., Turner, S. M., Clarke, A., Chance, R. and Liss, P. S.: Seasonal
805 cycle of seawater bromoform and dibromomethane concentrations in a coastal bay on the western Antarctic Peninsula,
806 *Global Biogeochem. Cycles*, 23, doi:10.1029/2008GB003268, 2009.

807 Hughes, C., Franklin, D. J. and Malin, G.: Iodomethane production by two important marine cyanobacteria:
808 *Prochlorococcus marinus* (CCMP 2389) and *Synechococcus* sp. (CCMP 2370), *Mar. Chem.*, 125(1-4), 19–25,
809 doi:10.1016/j.marchem.2011.01.007, 2011.

810 Hughes, C., Johnson, M., Utting, R., Turner, S., Malin, G., Clarke, a. and Liss, P. S.: Microbial control of bromocarbon
811 concentrations in coastal waters of the western Antarctic Peninsula, *Mar. Chem.*, 151, 35–46,
812 doi:10.1016/j.marchem.2013.01.007, 2013.

813 Janssen, F., Schrum, C. and Backhaus, J.: A climatological data set of temperature and salinity for the Baltic Sea and
814 the North Sea, *Dtsch. Hydrogr. Zeitschrift, Supplement* [online] Available from:
815 <http://link.springer.com/article/10.1007/BF02933676> (Accessed 16 August 2014), 1999.

816 Kangro, K., Olli, K., Tamminen, T. and Lignell, R.: Species-specific responses of a cyanobacteria-dominated
817 phytoplankton community to artificial nutrient limitation in the Baltic Sea, *Mar. Ecol. Prog. Ser.*, 336, 15–27

818 Karlsson, A., Auer, N., Schulz-Bull, D. and Abrahamsson, K.: Cyanobacterial blooms in the Baltic — A source of
819 halocarbons, *Mar. Chem.*, 110, 129–139, doi:10.1016/j.marchem.2008.04.010, 2008.

820 Kiene, R. P. and Hines, M. E.: Microbial formation of dimethyl sulfide in anoxic sphagnum peat, *Appl. Environ.*
821 *Microbiol.*, 61(7), 2720–2726 [online] Available from: <http://aem.asm.org/content/61/7/2720.short> (Accessed 6
822 February 2014), 1995.

823 Kiene, R. P. and Linn, L. J.: Distribution and turnover of dissolved DMSP and its relationship with bacterial production
824 and dimethylsulfide in the Gulf of Mexico, *Limnol. Oceanogr.*, 45(4), 849–861, 2000.

825 Kiene, R. P. and Visscher, P. T.: Production and fate of methylated sulfur compounds from methionine and
826 dimethylsulfoniopropionate in anoxic salt marsh sediments., *Appl. Environ. Microbiol.*, 53(10), 2426–2434, 1987.

827 Kim, J.-M., Lee, K., Yang, E. J., Shin, K., Noh, J. H., Park, K.-T., Hyun, B., Jeong, H.-J., Kim, J.-H., Kim, K. Y., Kim,
828 M., Kim, H.-C., Jang, P.-G. and Jang, M.-C.: Enhanced production of oceanic dimethylsulfide resulting from CO₂-
829 induced grazing activity in a high CO₂ world., *Environ. Sci. Technol.*, 44(21), 8140–8143, doi:10.1021/es102028k,
830 2010.

831 Klick, S.: Seasonal variations of biogenic and anthropogenic halocarbons in seawater from a coastal site, *Limnol.*
832 *Oceanogr.*, 37(7), 1579–1585 [online] Available from: <http://cat.inist.fr/?aModele=afficheN&cpsidt=4788349>
833 (Accessed 4 August 2014), 1992.

- 834 Klick, S. and Abrahamsson, K.: Biogenic volatile iodated hydrocarbons in the ocean, *J. Geophys. Res.*, 97(C8), 12683–
835 12687 [online] Available from: <http://www.agu.org/journals/ABS/1992/92JC00948.shtml> (Accessed 26 July 2013),
836 1992.
- 837 Lana, A., Simó, R., Vallina, S. M. and Dachs, J.: Re-examination of global emerging patterns of ocean DMS
838 concentration, *Biogeochemistry*, 110, 173–182, doi:10.1007/s10533-011-9677-9, 2012.
- 839 Laturnus, F., Giese, B., Wiencke, C. and Adams, F. C.: Low-molecular-weight organoiodine and organobromine
840 compounds released by polar macroalgae-the influence of abiotic factors, *Fresenius. J. Anal. Chem.*, 368, 297–302
841 [online] Available from: <http://www.ncbi.nlm.nih.gov/pubmed/11220596>, 2000.
- 842 Leck, C. and Rodhe, H.: Emissions of marine biogenic sulfur to the atmosphere of northern Europe, *J. Atmos. Chem.*,
843 12, 63–86 [online] Available from: <http://www.springerlink.com/index/H5472PG48150N025.pdf> (Accessed 26 July
844 2013), 1991.
- 845 Leck, C., Larsson, U., Bågander, L. E., Johansson, S. and Hajdu, S.: Dimethyl sulfide in the Baltic Sea: annual
846 variability in relation to biological activity, *J. Geophys. Res.*, 95(C3), 3353–3363, doi:10.1029/JC095iC03p03353,
847 1990.
- 848 Lee, G., Park, J., Jang, Y., Lee, M., Kim, K. R., Oh, J. R., Kim, D., Yi, H. II and Kim, T. Y.: Vertical variability of
849 seawater DMS in the South Pacific Ocean and its implication for atmospheric and surface seawater DMS,
850 *Chemosphere*, 78(8), 1063–1070, doi:10.1016/j.chemosphere.2009.10.054, 2010.
- 851 Leedham, E. C., Hughes, C., Keng, F. S. L., Phang, S.-M., Malin, G. and Sturges, W. T.: Emission of atmospherically
852 significant halocarbons by naturally occurring and farmed tropical macroalgae, *Biogeosciences*, 10(6), 3615–3633,
853 doi:10.5194/bg-10-3615-2013, 2013.
- 854 Lehmann, A. and Myrberg, K.: Upwelling in the Baltic Sea — A review, *J. Mar. Syst.*, 74, S3–S12,
855 doi:10.1016/j.jmarsys.2008.02.010, 2008.
- 856 Lischka, S., Riebesell, U., Stühr, A. and Bermudez, J. R.: Micro- and mesozooplankton community response to
857 increasing levels of CO₂ in the Baltic Sea: insights from a large-scale mesocosm experiment, *Biogeosciences*,
858 Submitted, 2015.
- 859 Liss, P., Marandino, C. A., Dahl, E., Helmig, D., Hintsä, E. J., Hughes, C., Johnson, M., Moore, R. M., Plane, J. M. C.,
860 Quack, B., Singh, H. B., Stefels, J., von Glasow, R. and Williams, J.: Short-lived trace gases in the surface ocean and
861 the atmosphere, in *Ocean-Atmosphere Interactions of Gases and Particles*, edited by P. Liss and M. Johnson, pp. 55–
862 112., 2014.
- 863 Lomans, B. P., Smolders, A., Intven, L. M., Pol, A., Op, D. and van der Drift, C.: Formation of dimethyl sulfide and
864 methanethiol in anoxic freshwater sediments, *Appl. Environ. Microbiol.*, 63(12), 4741–4747, 1997.
- 865 Mackey, M. D., Mackey, D. J., Higgins, H. W. and Wright, S. W.: CHEMTAX a program for estimating class
866 abundances from chemical markers: application to HPLC measurements of phytoplankton, *Mar. Ecol. Prog. Ser.*, 144,
867 265–283 [online] Available from: <http://www.int-res.com/abstracts/meps/v144/p265-283> (Accessed 25 March 2014),
868 1996.
- 869 Manley, S. L. and De La Cuesta, J. L.: Methyl iodide production from marine phytoplankton cultures, *Limnol.*
870 *Oceanogr.*, 42(1), 142–147, doi:10.4319/lo.1997.42.1.0142, 1997.
- 871 Manley, S. L., Goodwin, K. and North, W. J.: Laboratory production of bromoform, methylene bromide, and methyl
872 iodide by macroalgae in and distribution nearshore Southern California waters, *Limnol. Oceanogr.*, 37(8), 1652–1659,
873 1992.
- 874 Martino, M., Liss, P. S. and Plane, J. M. C.: The photolysis of dihalomethanes in surface seawater., *Environ. Sci.*
875 *Technol.*, 39(18), 7097–7101 [online] Available from: <http://www.ncbi.nlm.nih.gov/pubmed/16201634>, 2005.
- 876 Moore, R. M.: Methyl halide production and loss rates in sea water from field incubation experiments, *Mar. Chem.*,
877 101(3-4), 213–219, doi:10.1016/j.marchem.2006.03.003, 2006.
- 878 Moore, R. M. and Tokarczyk, R.: Volatile biogenic halocarbons in the northwest Atlantic, *Global Biogeochem. Cycles*,
879 7(1), 195–210, 1993.
- 880 Moore, R. M. and Zafiriou, O. C.: Photochemical production of methyl iodide in seawater, *J. Geophys. Res.*, 99(D8),
881 16415–16420, doi:10.1029/94JD00786, 1994.
- 882 National Measurement Institute of Australia: Proficiency Study 12-23: DMS in seawater., 2013.

883 Nausch, M., Nausch, G., Lass, H. U., Mohrholz, V., Nagel, K., Siegel, H. and Wasmund, N.: Phosphorus input by
884 upwelling in the eastern Gotland Basin (Baltic Sea) in summer and its effects on filamentous cyanobacteria, *Estuar.
885 Coast. Shelf Sci.*, 83(4), 434–442, doi:10.1016/j.ecss.2009.04.031, 2009.

886 Newbold, L. K., Oliver, A. E., Booth, T., Tiwari, B., DeSantis, T., Maguire, M., Andersen, G., van der Gast, C. J. and
887 Whiteley, A. S.: The response of marine picoplankton to ocean acidification, *Environ. Microbiol.*, 14(9), 2293–2307,
888 doi:10.1111/j.1462-2920.2012.02762.x, 2012.

889 Niemisto, L., Rinne, I. and Melvasalo, T.: Blue-green algae and their nitrogen fixation in the Baltic Sea in 1980, 1982
890 and 1984, *Meri*, 17, 1–59, 1989.

891 O'Dowd, C. D., Jimenez, J. L., Bahreini, R., Flagan, R. C., Seinfeld, J. H., Hameri, K., Pirjola, L., Kulmala, M.,
892 Jennings, S. G. and Hoffmann, T.: Marine aerosol formation from biogenic iodine emissions, *Nature*, 417(6889), 632–
893 636, doi:10.1038/nature00773.1.2.3.4.5.6.7.8.9.10., 2002.

894 Oberg, J.: Cyanobacterial blooms in the Baltic Sea in 2013, HELCOM Balt. Sea Environ. Fact Sheet, 2013.

895 Orlikowska, A. and Schulz-Bull, D. E.: Seasonal variations of volatile organic compounds in the coastal Baltic Sea,
896 *Environ. Chem.*, 6, 495–507, doi:10.1071/EN09107, 2009.

897 Park, K.-T., Lee, K., Shin, K., Yang, E. J., Hyun, B., Kim, J.-M., Noh, J. H., Kim, M., Kong, B., Choi, D. H., Choi, S.-
898 J., Jang, P.-G. and Jeong, H. J.: Direct linkage between dimethyl sulfide production and microzooplankton grazing,
899 resulting from prey composition change under high partial pressure of carbon dioxide conditions., *Environ. Sci.
900 Technol.*, 48(9), 4750–4756, doi:10.1021/es403351h, 2014.

901 Passow, U. and Riebesell, U.: Mesocosm perturbation experiments and the sensitivity of marine biological systems to
902 global change, *Solas News*, (1), 12–13, doi:10.1029/2003JC002120, 2005.

903 Paul, A. J., Bach, L. T., Schulz, K.-G., Boxhammer, T., Czerny, J., Achterberg, E., Hellemann, D., Trense, Y., Nausch,
904 M., Sswat, M. and Riebesell, U.: Effect of elevated CO₂ on organic matter pools and fluxes in a summer Baltic Sea
905 plankton community., *Biogeosciences*, 12, 6181 – 6203, 2015.

906 Pedersen, M., Collen, J., Abrahamsson, K. and Ekdahl, A.: Production of halocarbons from seaweeds: an oxidative
907 stress reaction?, *Sci. Mar.*, 60(Supplement 1), 257–263, 1996.

908 Piontek, J., Borchard, C., Sperling, M., Schulz, K. G., Riebesell, U. and Engel, A.: Response of bacterioplankton
909 activity in an Arctic fjord system to elevated pCO₂: results from a mesocosm perturbation study, *Biogeosciences*, 10,
910 297–314, doi:10.5194/bg-10-297-2013, 2013.

911 Quack, B., Peeken, I., Petrick, G. and Nachtigall, K.: Oceanic distribution and sources of bromoform and
912 dibromomethane in the Mauritanian upwelling, *J. Geophys. Res.*, 112, C10006, doi:10.1029/2006JC003803, 2007.

913 Quinn, P. K. and Bates, T. S.: The case against climate regulation via oceanic phytoplankton sulphur emissions.,
914 *Nature*, 480(7375), 51–56, doi:10.1038/nature10580, 2011.

915 Raateoja, M., Kuosa, H. and Hällfors, S.: Fate of excess phosphorus in the Baltic Sea: A real driving force for
916 cyanobacterial blooms?, *J. Sea Res.*, 65(2), 315–321, doi:10.1016/j.seares.2011.01.004, 2011.

917 Raven, J. R., Caldeira, K., Elderfield, H., Hoegh-Guldberg, O., Liss, P. S., Riebesell, U., Shepherd, J., Turley, C. and
918 Watson, A.: Ocean acidification due to increasing atmospheric carbon dioxide, *R. Soc. Policy Doc.* 12/05, (June)
919 [online] Available from: [http://eprints.uni-](http://eprints.uni-kiel.de/7878/1/965_Raven_2005_OceanAcidificationDueToIncreasing_Monogr_pubid13120.pdf)
920 [kiel.de/7878/1/965_Raven_2005_OceanAcidificationDueToIncreasing_Monogr_pubid13120.pdf](http://eprints.uni-kiel.de/7878/1/965_Raven_2005_OceanAcidificationDueToIncreasing_Monogr_pubid13120.pdf) (Accessed 26 July
921 2013), 2005.

922 Riebesell, U., Czerny, J., von Bröckel, K., Boxhammer, T., Büdenbender, J., Deckelnick, M., Fischer, M., Hoffmann,
923 D., Krug, S. A., Lentz, U., Ludwig, A., Mücke, R. and Schulz, K. G.: Technical Note: A mobile sea-going mesocosm
924 system – new opportunities for ocean change research, *Biogeosciences*, 10(3), 1835–1847, doi:10.5194/bg-10-1835-
925 2013, 2013.

926 Ross, P. M., Parker, L., O'Connor, W. A. and Bailey, E. A.: The impact of ocean acidification on reproduction, early
927 development and settlement of marine organisms, *Water*, 3(4), 1005–1030, doi:10.3390/w3041005, 2011.

928 Roy, R., Pratihary, A., Narvenkar, G., Mochemadkar, S., Gauns, M. and Naqvi, S. W. A.: The relationship between
929 volatile halocarbons and phytoplankton pigments during a *Trichodesmium* bloom in the coastal eastern Arabian Sea,
930 *Estuar. Coast. Shelf Sci.*, 95(1), 110–118, doi:10.1016/j.ecss.2011.08.025, 2011.

931 Scarratt, M. G. and Moore, R. M.: Production of methyl bromide and methyl chloride in laboratory cultures of marine

- 932 phytoplankton II, *Mar. Chem.*, 59(3-4), 311–320, doi:10.1016/S0304-4203(97)00092-3, 1998.
- 933 Schall, C., Heumann, K. G. and Kirst, G. O.: Biogenic volatile organoiodine and organobromine hydrocarbons in the
934 Atlantic Ocean from 42°N to 72°S, *Fresenius. J. Anal. Chem.*, 359(3), 298–305, doi:10.1007/s002160050577, 1997.
- 935 Schneider, B., Nausch, G., Kubsch, H. and Petersohn, I.: Accumulation of total CO₂ during stagnation in the Baltic Sea
936 deep water and its relationship to nutrient and oxygen concentrations, *Mar. Chem.*, 77, 277–291, 2002.
- 937 Schulz, K. G., Bellerby, R. G. J., Brussaard, C. P. D., Büdenbender, J., Czerny, J., Engel, A., Fischer, M., Koch-
938 Klavsen, S., Krug, S. A., Lischka, S., Ludwig, A., Meyerhöfer, M., Nondal, G., Silyakova, A., Stühr, A. and Riebesell,
939 U.: Temporal biomass dynamics of an Arctic plankton bloom in response to increasing levels of atmospheric carbon
940 dioxide, *Biogeosciences*, 10(1), 161–180, doi:10.5194/bg-10-161-2013, 2013.
- 941 Simó, R., Archer, S. D., Pedros-Alio, C., Gilpin, L. and Stelfox-Widdicombe, C. E.: Coupled dynamics of
942 dimethylsulfoniopropionate and dimethylsulfide cycling and the microbial food web in surface waters of the North
943 Atlantic, *Limnol. Oceanogr.*, 47(1), 53–61 [online] Available from: <http://cedadocs.badc.rl.ac.uk/67/> (Accessed 25
944 March 2014), 2002.
- 945 Simó, R., Vila-Costa, M., Alonso-Sáez, L., Cardelús, C., Guadayol, Ó., Vázquez-Dominguez, E. and Gasol, J. M.:
946 Annual DMSP contribution to S and C fluxes through phytoplankton and bacterioplankton in a NW Mediterranean
947 coastal site, *Aquat. Microb. Ecol.*, 57(October), 43–55, doi:10.3354/ame01325, 2009.
- 948 Six, K. D., Kloster, S., Ilyina, T., Archer, S. D., Zhang, K. and Maier-Reimer, E.: Global warming amplified by reduced
949 sulphur fluxes as a result of ocean acidification, *Nat. Clim. Chang.*, 3(8), 1–4, doi:10.1038/nclimate1981, 2013.
- 950 Smith, D. C., Simon, M., Alldredge, A. L. and Azam, F.: Intense hydrolytic enzyme activity on marine aggregates and
951 implications for rapid particle dissolution, *Nature*, 359, 139 – 142 [online] Available from:
952 <http://www.gso.uri.edu/dcsmith/page3/page19/assets/Smithetal92.pdf> (Accessed 11 September 2014), 1992.
- 953 Solomon, S., Garcia, R. R. and Ravishankara, A. R.: On the role of iodine in ozone depletion, *J. Geophys. Res.*,
954 99(D10), 20491–20499, doi:10.1029/94JD02028, 1994.
- 955 Stal, L. J., Albertano, P., Bergman, B., von Bröckel, K., Gallon, J. R., Hayes, P. K., Sivonen, K. and Walsby, A. E.:
956 BASIC: Baltic Sea cyanobacteria. An investigation of the structure and dynamics of water blooms of cyanobacteria in
957 the Baltic Sea—responses to a changing environment, *Cont. Shelf Res.*, 23(17-19), 1695–1714,
958 doi:10.1016/j.csr.2003.06.001, 2003.
- 959 Stets, E. G., Hines, M. E. and Kiene, R. P.: Thiol methylation potential in anoxic, low-pH wetland sediments and its
960 relationship with dimethylsulfide production and organic carbon cycling, *FEMS Microbiol. Ecol.*, 47(1), 1–11,
961 doi:10.1016/S0168-6496(03)00219-8, 2004.
- 962 Sunda, W., Kieber, D. J., Kiene, R. P. and Huntsman, S.: An antioxidant function for DMSP and DMS in marine algae,
963 *Nature*, 418(6895), 317–320, doi:10.1038/nature00851, 2002.
- 964 Theiler, R., Cook, J. C., Hager, L. P. and Siuda, J. F.: Halohydrocarbon synthesis by bromoperoxidase, *Science.*,
965 202(December), 1094 – 1096, 1978.
- 966 Turner, S. M., Malin, G., Liss, P. S., Harbour, D. S. and Holligan, P. M.: The seasonal variation of dimethyl sulfide and
967 dimethylsulfoniopropionate concentrations in nearshore waters, *Limnol. Oceanogr.*, 33(3), 364–375, 1988.
- 968 Urhahn, T. and Ballschmiter, K.: Chemistry of the biosynthesis of halogenated methanes: C1-organohalogenes as pre-
969 industrial chemical stressors in the environment?, *Chemosphere*, 37(6), 1017–1032, doi:10.1016/S0045-
970 6535(98)00100-3, 1998.
- 971 del Valle, D. A., Slezak, D., Smith, C. M., Rellinger, A. N., Kieber, D. J. and Kiene, R. P.: Effect of acidification on
972 preservation of DMSP in seawater and phytoplankton cultures: Evidence for rapid loss and cleavage of DMSP in
973 samples containing *Phaeocystis* sp., *Mar. Chem.*, 124, 57–67, doi:10.1016/j.marchem.2010.12.002, 2011.
- 974 Vila-Costa, M., Simó, R., Harada, H., Gasol, J. M., Slezak, D. and Kiene, R. P.: Dimethylsulfoniopropionate uptake by
975 marine phytoplankton, *Science*, 314(5799), 652–4, doi:10.1126/science.1131043, 2006a.
- 976 Vila-Costa, M., del Valle, D. A., González, J. M., Slezak, D., Kiene, R. P., Sánchez, O. and Simó, R.: Phylogenetic
977 identification and metabolism of marine dimethylsulfide-consuming bacteria, *Environ. Microbiol.*, 8(12), 2189–2200,
978 doi:10.1111/j.1462-2920.2006.01102.x, 2006b.
- 979 Visscher, P. T., Baumgartner, L. K., Buckley, D. H., Rogers, D. R., Hogan, M. E., Raleigh, C. D., Turk, K. A. and Des
980 Marais, D. J.: Dimethyl sulphide and methanethiol formation in microbial mats: potential pathways for biogenic

981 signatures, Environ. Microbiol., 5(4), 296–308 [online] Available from:
982 <http://www.ncbi.nlm.nih.gov/pubmed/12662177>, 2003.

983 Vogt, M., Steinke, M., Turner, S. M., Paulino, A., Meyerhöfer, M., Riebesell, U., LeQuéré, C. and Liss, P. S.:
984 Dynamics of dimethylsulphoniopropionate and dimethylsulphide under different CO₂ concentrations during a
985 mesocosm experiment, Biogeosciences, 5(2), 407–419, doi:10.5194/bg-5-407-2008, 2008.

986 Wasmund, N.: Occurrence of cyanobacterial blooms in the Baltic Sea in relation to environmental conditions, Iny. Rev.
987 ges. Hydrobiol., 82(2), 169–184 [online] Available from:
988 <http://onlinelibrary.wiley.com/doi/10.1002/iroh.19970820205/full> (Accessed 25 March 2014), 1997.

989 Webb, A.: The effects of elevated CO₂ and ocean acidification on the production of marine biogenic trace gases, PhD
990 Thesis, Univ. East Angl., (March), 2015.

991 Webb, A. L., Malin, G., Hopkins, F. E., Ho, K.-L., Riebesell, U., Schulz, K., Larsen, A. and Liss, P.: Ocean
992 acidification has different effects on the production of dimethylsulphide and dimethylsulphoniopropionate measured in
993 cultures of *Emiliania huxleyi* RCC1229 and mesocosm study: a comparison of laboratory monocultures and community
994 interactions, Environ. Chem., EN14268, doi:<http://dx.doi.org/10.1071/EN14268>, 2015.

995 Welschmeyer, N. A.: Fluorometric analysis of chlorophyll a in the presence of chlorophyll b and pheopigments,
996 Limnol. Oceanogr., 39(8), 1985–1992, 1994.

997 Zika, R. G., Gidel, L. T. and Davis, D. D.: A comparison of photolysis and substitution decomposition rates of methyl
998 iodide in the ocean, Geophys. Res. Lett., 11(4), 353–356, 1984.

999 Zinder, S. H., Doemel, W. N. and Brock, T. D.: Production of volatile sulfur compounds during the decomposition of
1000 algal mats, Appl. Environ. Microbiol., 34(6), 859–861 [online] Available from:
1001 <http://aem.asm.org/content/34/6/859.short> (Accessed 25 March 2014), 1977.

1002 Zindler, C., Peeken, I., Marandino, C. A. and Bange, H. W.: Environmental control on the variability of DMS and
1003 DMSP in the Mauritanian upwelling region, Biogeosciences, 9, 1041–1051, doi:10.5194/bg-9-1041-2012, 2012.

1006 Table 1. Summary of $f\text{CO}_2$ and pH_T (total scale) during phases 0, 1 and 2 of the mesocosm
 1007 experiment.

		Whole Experiment		Phase 0 ($t\text{-}3$ to $t0$)		Phase I ($t1$ – $t16$)		Phase II ($t16$ – $t31$)	
		Target	Mean	Mean	Mean	Mean	Mean	Mean	Mean
Mesocosm ^a	$f\text{CO}_2$	$f\text{CO}_2$	pH_T	$f\text{CO}_2$	pH_T	$f\text{CO}_2$	pH_T	$f\text{CO}_2$	pH_T
	(μatm)	(μatm)		(μatm)		(μatm)		(μatm)	
M1	Control	331	7.91	231	8.00	328	7.95	399	7.86
M5	Control	334	7.91	244	7.98	329	7.94	399	7.52
M7	390	458	7.80	239	7.99	494	7.81	532	7.76
M6	840	773	7.63	236	7.99	932	7.59	855	7.59
M3	1120	950	7.56	243	7.98	1176	7.51	1027	7.52
M8	1400	1166	7.49	232	8.00	1481	7.43	1243	7.45
Baltic Sea	380	350	7.91	298	7.91	277	7.98	436	7.86

1008 ^a listed in order of increasing $f\text{CO}_2$

1009

1010 Table 2. Calibration ranges and calculated percentage mean relative standard error for the trace
 1011 gases measured in the mesocosms.

Compound	Calibration range (pmol L ⁻¹)	% Mean relative standard error
DMS	600 – 29300*	6.33
DMSP	2030 – 405900*	
CH ₃ I	0.11 – 11.2	4.62
CH ₂ I ₂	5.61 – 561.0	4.98
C ₂ H ₅ I	0.10 – 4.91	5.61
CH ₂ ClI	1.98 – 99.0	3.64
CHBr ₃	8.61 – 816.0	4.03
CH ₂ Br ₂	0.21 – 20.9	5.30
CHBr ₂ Cl	0.07 – 7.00	7.20

1012 * throughout the rest of this paper, these measurements are given in nmol L⁻¹.

1013

Table 3. Abundance and contributions of different phytoplankton groups to the total phytoplankton community assemblage, showing the range of measurements from total Chl-*a* (Paul *et al.*, 2015), CHEMTAX analysis of derived Chl-*a* (Paul *et al.*, 2015) and phytoplankton abundance (Crawfurd *et al.*, 2016). Data are split into the range of all the mesocosm measurements and those from the Baltic Sea.

Mesocosm				Baltic Sea		
	Range	Range	% Contribution to Chl-	Range	Range	% Contribution to Chl- <i>a</i>
	Integrated 10 m	Integrated 17 m	<i>a</i>	Integrated 10 m	Integrated 17 m	
Chl-<i>a</i>	0.9 – 2.9	0.9 – 2.6	100	1.3 – 6.5	1.12 – 5.5	100
Phytoplankton Taxonomy (Equivalent Chlorophyll µg L⁻¹)						
Cyanobacteria		0.01 – 0.4	8		0.0 – 0.1	1
Prasinophytes		0.04 – 0.3	7		0.01 – 0.3	4
Euglenophytes		0.0 – 1.6	15		0.0 – 2.6	21
Dinoflagellates		0.0 – 0.3	3		0.04 – 0.6	9
Diatoms		0.1 – 0.3	7		0.04 – 0.9	9
Chlorophytes		0.3 – 2.0	40		0.28 – 3.1	41
Cryptophytes		0.1 – 1.4	21		0.1 – 1.0	15
Small Phytoplankton (<10 µm) abundance (cells mL⁻¹)						
Cyanobacteria	55000 – 380000	65000 – 470000		30000 – 180000	30000 – 250000	
Picoeukaryotes I	15000 – 100000	17000 – 111000		5000 – 70000	6100 – 78000	
Picoeukaryotes II	700 – 4000	600 – 4000		400 – 3000	460 – 3700	
Picoeukaryotes III	1000 - 9000	1100 – 8500		1000 – 6000	950 – 7500	
Nanoeukaryotes I	400 – 1400	270 – 1500		200 – 4000	210 – 4100	
Nanoeukaryotes II	0 – 400	4 – 400		100 – 1100	60 – 1300	

Table 4. Concentration ranges of trace gases measured in the mesocosms compared to other open water ocean acidification experiments, showing the range of concentrations for each gas and the percentage change between the control and the highest $f\text{CO}_2$ treatment.

	Range $f\text{CO}_2$		DMS	CH_3I	CH_2I_2	CH_2ClI	CHBr_3	CH_2Br_2	$\text{CH}_2\text{Br}_2\text{Cl}$
	(μatm)		(nmol L^{-1})						
SOPRAN Tvärminne Mesocosm (this study)	346 – 1333	Range	2.7-6.8	2.9-6.4	57-202	3.8-8.0	69-148	4.0-7.7	1.7-3.1
		% change	-34	-0.3	1.3	-11	-9	-3	-4
SOPRAN Bergen 2011 (Webb <i>et al.</i>, 2015)	280 – 3000	Range	0.1-4.9	4.9-32	5.8-321	9.0-123	64-306	6.3-30.8	3.9-14
		% change	-60	-37	-48	-27	-2	-4	-6
NERC Microbial Metagenomics Experiment, Bergen 2006 (Hopkins <i>et al.</i>, 2010)	300 - 750	Range	ND-50	2.0-25	ND-750	ND-700	5.0-80	ND-5.5	0.2-1.2
		% change	-57	-41	-33	-28	13	8	22
EPOCA Svalbard 2010 (Archer <i>et al.</i>, 2013; Hopkins <i>et al.</i>, 2013)	180 - 1420	Range	ND-14	0.04-10	0.01-2.5	0.3-1.6	35-151	6.3-33.3	1.6-4.7
		% change	-60	NS		NS	NS	NS	NS
UKOA European Shelf 2011 (Hopkins and Archer, 2014)	340 - 1000	Range	0.5-12						
		% change	225						
Korean Mesocosm Experiment 2012 (Park <i>et al.</i>, 2014)	160 - 830	Range	1.0-100						
		% change	-82						

Table 5. Concentration ranges of trace gases measured in the Baltic Sea compared to concentrations measured in the literature. ND – Not Detected.

Study	DMS concentration range (nmol L ⁻¹)	Halocarbon concentration range (pmol L ⁻¹)						
		CH ₃ I	CH ₂ I ₂	C ₂ H ₅ I	CH ₃ Cl	CHBr ₃	CH ₂ Br ₂	CH ₂ Br ₂ Cl
SOPRAN Tvärminne Baltic Sea (This Study)	1.9-11	4.3-8.6	66.9-374	0.6 – 1.0	7.0-18	93-192	7.1-10	3.3-5.0
Orlikowska and Schulz-Bull (2009)	0.3-120	1-16	0-85	0.4 – 1.2	5-50	5.0-40	2.0-10	0.8-2.5
Karlsson <i>et al.</i> (2008)		3.0-7.5				35-60	4.0-7.0	2.0-6.5
Klick and Abrahamsson (1992)			15-709		11-74	14-585		
Klick (1992)			ND-243		ND-57	40-790	ND-86	ND-29
Leck and Rodhe (1991)	0.4-2.8							
Leck <i>et al.</i> (1990)	ND-3.2							

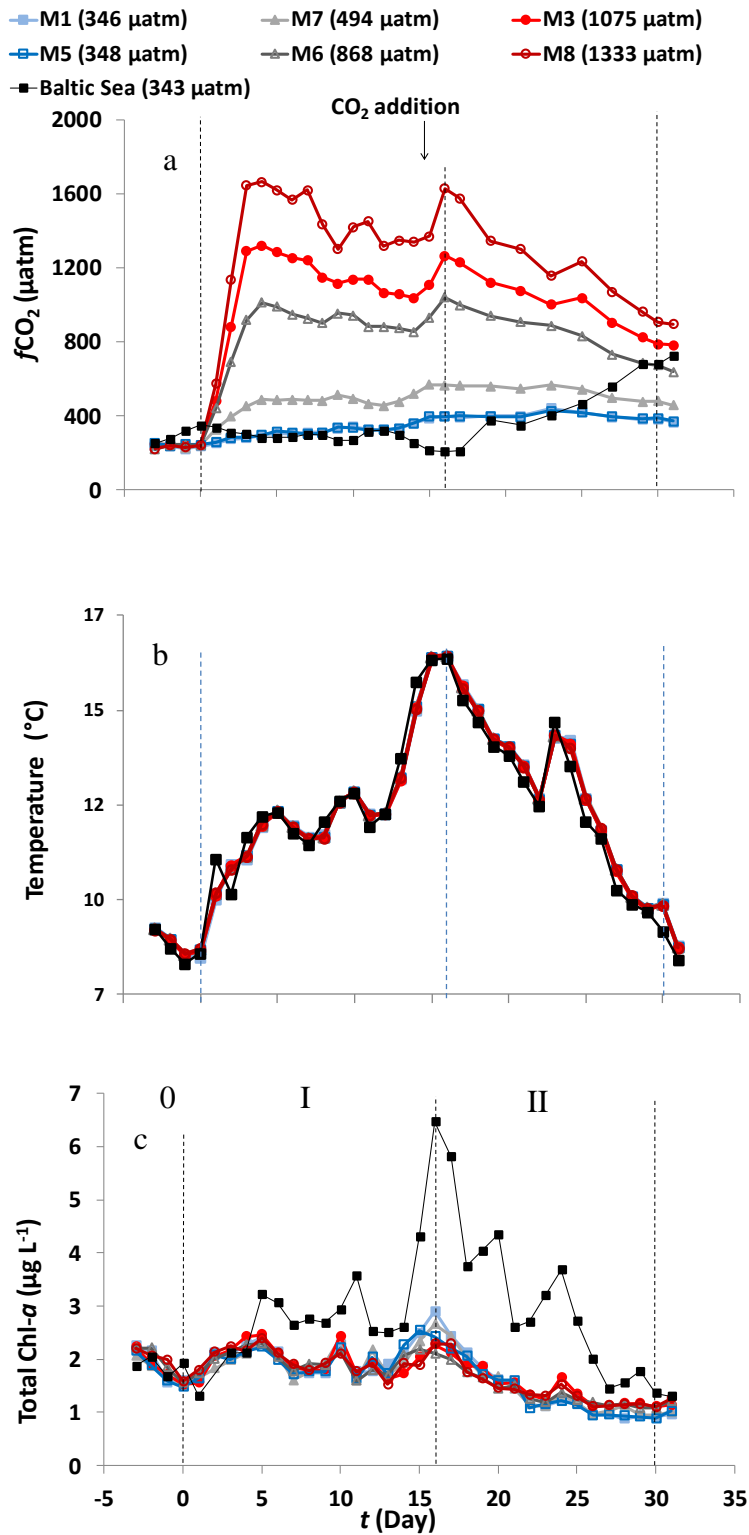


Figure 1. Daily measurements of (a) $f\text{CO}_2$, (b) mean temperature and (c) total Chlorophyll- a in the mesocosms and surrounding Baltic Sea waters. Dashed lines represent the three Phases of the experiment, based on the Chl- a data.

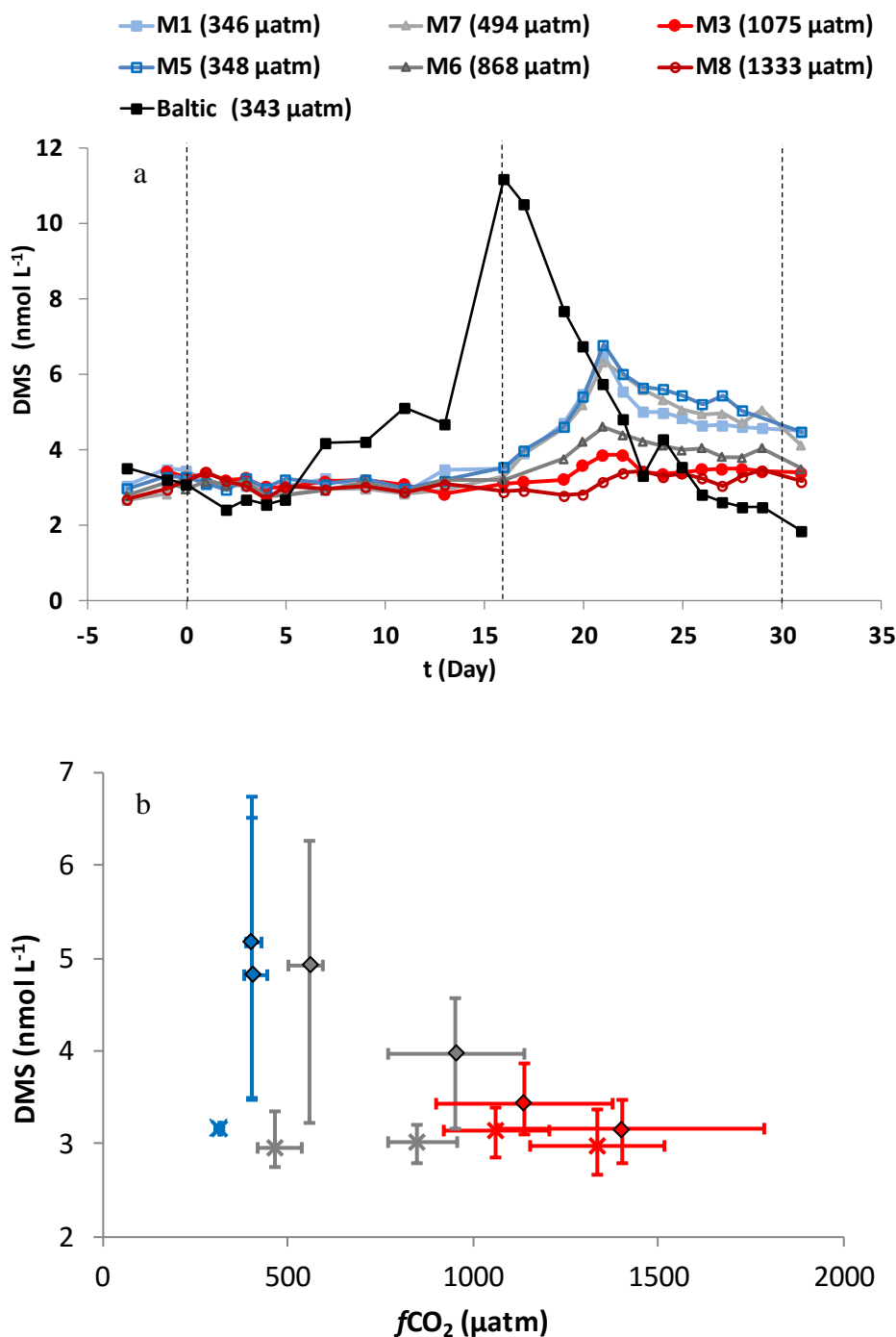


Figure 2. (a) Mean DMS concentrations measured daily in the mesocosms and Baltic Sea from an integrated water sample of the surface 10m. Dashed lines show the Phases of the experiment as given in Fig. 1, $f\text{CO}_2$ shown in the legend are mean $f\text{CO}_2$ across the duration of the experiment. (b) Mean DMS concentrations from each mesocosm during Phase I (crosses) and Phase II (diamonds), for ambient (blue), medium (grey) and high $f\text{CO}_2$ (red), with error bars showing the range of both the DMS and $f\text{CO}_2$.

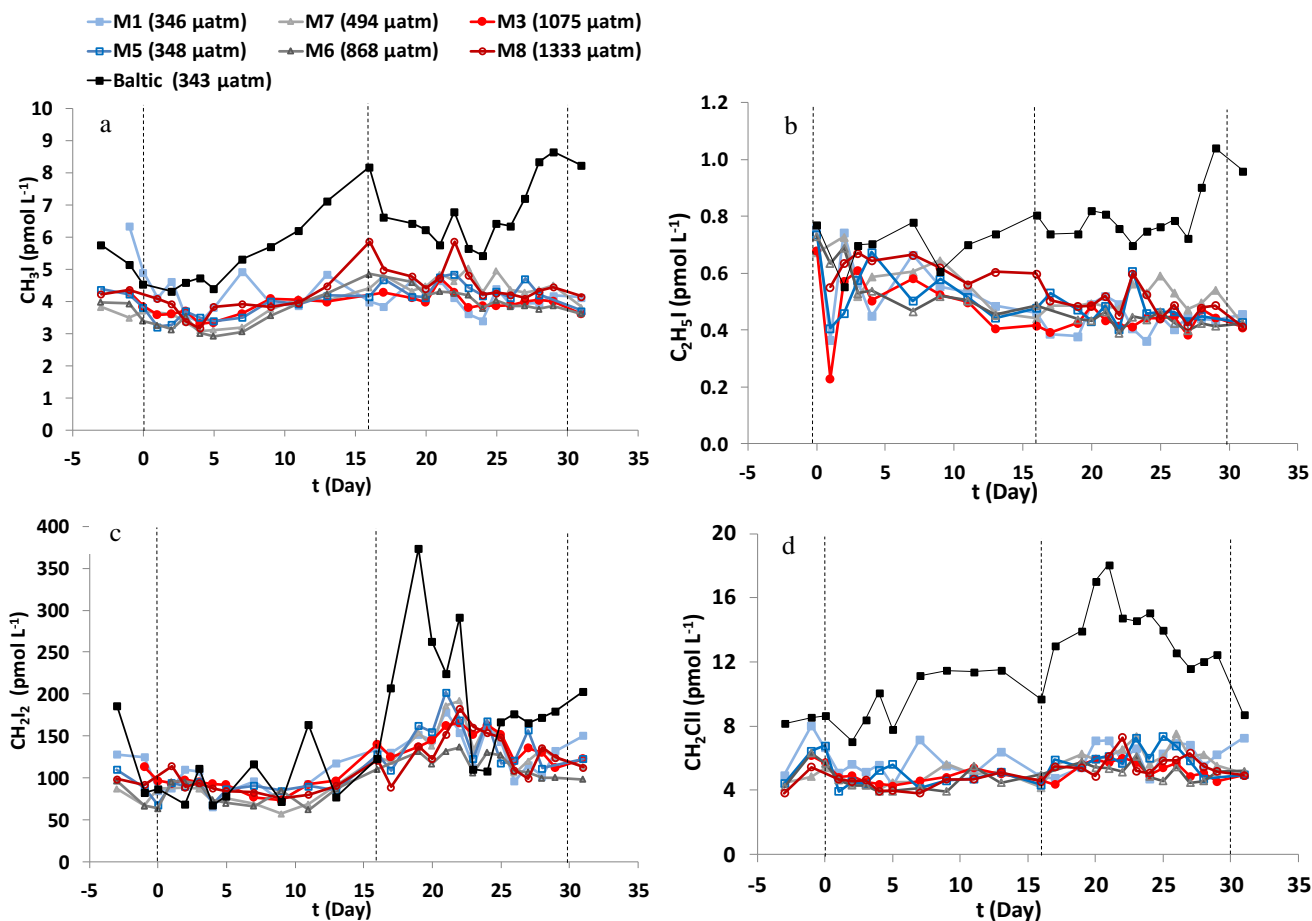


Figure 3. Mean concentrations (pmol L⁻¹) of (a) CH₃I, (b) C₂H₅I, (c) CH₂I₂ and (d) CH₂ClI taken from a water sample integrated from the surface 10m. Dashed lines indicate the Phases of the experiment, as given in Fig. 2. $f\text{CO}_2$ shown in the legend are mean $f\text{CO}_2$ across the duration of the experiment.

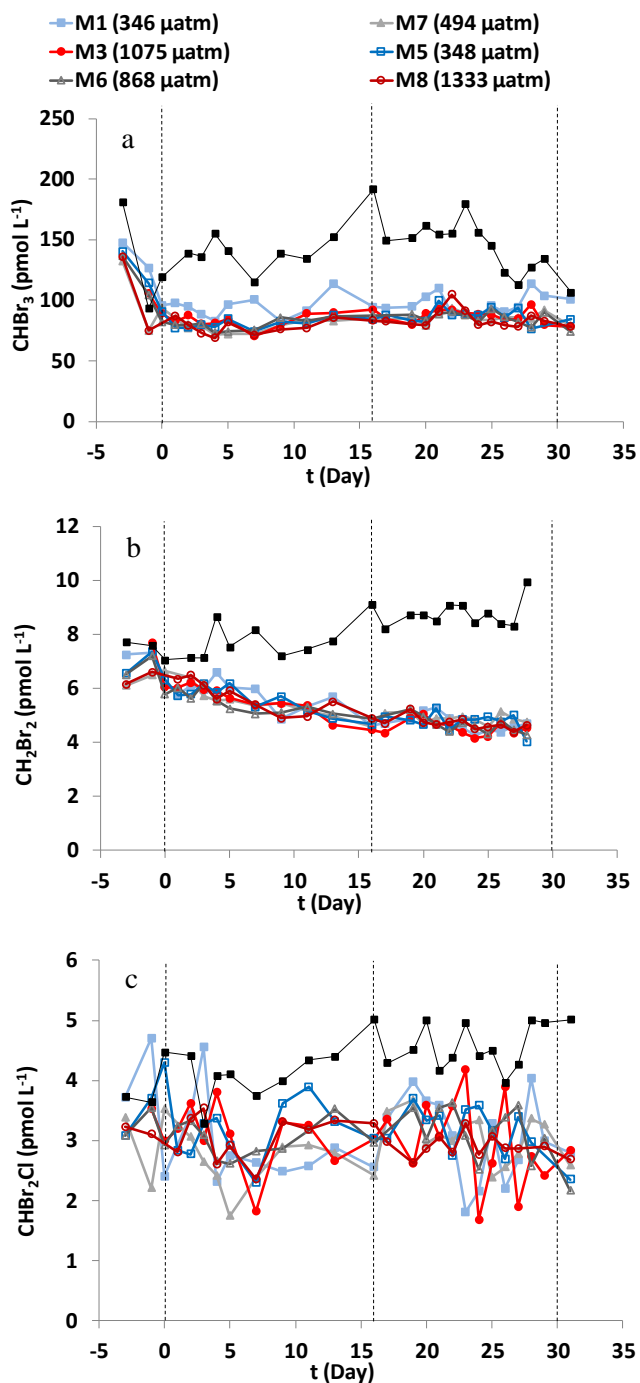


Figure 4. Mean concentrations (pmol L^{-1}) of (a) CHBr_3 , (b) CH_2Br_2 and (c) CHBr_2Cl taken from a water sample integrated from the surface 10m. Dashed lines indicate the phases of the experiment as defined in Fig. 2, $f\text{CO}_2$ shown in the legend are mean $f\text{CO}_2$ across the duration of the experiment.