

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26

Ocean acidification decreases plankton respiration: evidence from a mesocosm experiment

K. Spilling^{1,2}, A. J. Paul³, N. Virkkala², T. Hastings⁴, S. Lischka³, A. Stühr³, R. Bermudez^{3,5}, J. Czerny³, T. Boxhammer³, K. G. Schulz⁶, A. Ludwig³, U. Riebesell³

[1] {Marine Research Centre, Finnish Environment Institute, P.O. Box 140, 00251 Helsinki, Finland}

[2] {Tvärminne Zoological Station, University of Helsinki, J. A. Palménin tie 260, 10900 Hanko, Finland}

[3] {GEOMAR Helmholtz Centre for Ocean Research Kiel, Düsternbrooker Weg 20, D-24105 Kiel, Germany}

[4] {Department of Biology, University of Portsmouth, University House, Winston Churchill Avenue, Portsmouth PO1 2UP, United Kingdom}

[5] {Facultad de Ingeniería Marítima, Ciencias Biológicas, Oceánicas y Recursos Naturales. ESPOL, Escuela Superior Politécnica del Litoral, Guayaquil, Ecuador}

[6] {Centre for Coastal Biogeochemistry, Southern Cross University, Military Road, East Lismore, NSW 2480, Australia}

Correspondence to: K. Spilling (kristian.spilling@environment.fi)

Running title: Ocean acidification decreases respiration

27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49
50
51
52
53
54

Abstract

Anthropogenic carbon dioxide (CO₂) emissions are reducing the pH in the world's oceans. The plankton community is a key component driving biogeochemical fluxes, and the effect of increased CO₂ on plankton is critical for understanding the ramifications of ocean acidification on global carbon fluxes. We determined the plankton community composition and measured primary production, respiration rates and carbon export (defined here as carbon sinking out of a shallow, coastal area) during an ocean acidification experiment. Mesocosms (~55 m³) were set up in the Baltic Sea with a gradient of CO₂ levels initially ranging from ambient (~240 μatm), used as control, to high CO₂ (up to ~1330 μatm). The phytoplankton community was dominated by dinoflagellates, diatoms, cyanobacteria and chlorophytes, and the zooplankton community by protozoans, heterotrophic dinoflagellates and cladocerans. The plankton community composition was relatively homogenous between treatments. Community respiration rates were lower at high CO₂ levels. The carbon-normalized respiration was approximately 40% lower in the high CO₂ environment compared with the controls during the latter phase of the experiment. We did not, however, detect any effect of increased CO₂ on primary production. This could be due to measurement uncertainty, as the measured total particular carbon (TPC) and combined results presented in this special issue suggest that the reduced respiration rate translated into higher net carbon fixation. The percent carbon derived from microscopy counts (both phyto- and zooplankton), of the measured total particular carbon (TPC) decreased from ~26% at *t0* to ~8% at *t31*, probably driven by a shift towards smaller plankton (<4 μm) not enumerated by microscopy. Our results suggest that reduced respiration lead to increased net carbon fixation at high CO₂. However, the increased primary production did not translate into increased carbon export, and did consequently not work as a negative feedback mechanism for increasing atmospheric CO₂ concentration.

55

56 **1 Introduction**

57 The ocean is a large sink of carbon dioxide (CO₂) and absorbs around 25 % of annual
58 anthropogenic CO₂ emissions (Le Quéré et al., 2009). CO₂ is a weak acid when dissolved in
59 water, and the increasing global atmospheric CO₂ concentration has reduced the average pH
60 in the ocean by approximately 0.1 since the start of the Industrial Revolution (Orr, 2011).
61 This pH reduction, with a concurrent increase in dissolved inorganic carbon, is called ocean
62 acidification. Following the same trajectory, the pH could decline further by as much as 0.7
63 by 2300 (Zeebe et al., 2008).

64 The topic of ocean acidification has received a lot of attention over the past decade. There is a
65 relatively good understanding of the rate of change and the effects on the ocean's carbon
66 chemistry (Zeebe and Ridgwell, 2011). There are also a range of studies documenting the
67 effects of decreasing pH on marine life, but the effect studied is often species or ecosystem
68 specific and based on short term perturbation experiments (Riebesell and Tortell, 2011).
69 There are still a lot of uncertainties as to what effect ocean acidification has on biological
70 processes.

71 The key driving force in marine biogeochemical element cycling is the planktonic community
72 that occupies the sunlit surface of the ocean. Primary producers use the energy from sunlight
73 to take up CO₂ and fix carbon into organic compounds. Respiration is the opposite process
74 where organic carbon is oxidized providing energy and releasing CO₂. This takes place at all
75 trophic levels, from bacteria through to zooplankton, fish and marine mammals. At steady
76 state, production and respiration are balanced. On a global scale, there is presently a surplus
77 of organic matter being produced in the upper ocean through photosynthesis. The extra
78 organic carbon is exported out of the surface layers to the deep ocean where it is sequestered
79 for the foreseeable future, a process referred to as the biological carbon pump. (Volk and
80 Hoffert, 1985; Siegenthaler and Sarmiento, 1993; Ducklow et al., 2001). In the case of
81 coastal seas, part of the carbon is buried at the sea floor (Dunne et al., 2007).

82 The greater the difference between primary production and respiration, the more carbon can
83 potentially be exported, and ocean acidification has the potential to affect this balance.
84 Generally, more CO₂ stimulates photosynthetic carbon fixation, as CO₂ becomes more
85 readily available for the key photosynthetic enzyme RubisCO (Falkowski and Raven, 2013),
86 however, increased primary production at high CO₂ concentration is not always recorded
87 (Sobrino et al., 2014) and the response is variable between different taxa (Mackey et al.,
88 2015). In cases where additional carbon is fixed, it may be excreted as dissolved organic

89 carbon (DOC), providing carbon for bacterial growth, and also increasing bacterial
90 respiration (Grossart et al., 2006; Piontek et al., 2010). Changes in pH might also directly
91 affect both primary production (Spilling, 2007) and respiration (Smith and Raven, 1979).
92 The Baltic Sea is an almost landlocked sea with low alkalinity (Beldowski et al., 2010), and
93 is thus particularly susceptible to variation in seawater pH. Because of the reduced water
94 exchange with the North Atlantic and the large catchment area (population ~80 million), it is
95 also subjected to a range of other environmental pressures, in particular increased nutrient
96 inputs from human activities, i.e. eutrophication. Eutrophication has led to increased primary
97 production and chlorophyll *a* (Chl *a*) biomass over the past decades in the Gulf of Finland
98 (Raateoja et al., 2005), benefitting chrysophytes, chlorophytes and cyanobacteria, (Suikkanen
99 et al., 2007). Dense blooms of diazotroph cyanobacteria are common in the summer, which
100 further aggravates the eutrophication problem as nitrogen fixation introduces substantial
101 amounts of new nitrogen into the system (Savchuk, 2005). The effect of ocean acidification
102 on this type of system is largely unexplored. In order to investigate the effect of increased
103 CO₂ (and lower pH) on primary production and total plankton respiration in the pelagic zone,
104 we measured carbon fixation, oxygen consumption and export/sedimentation rates during a
105 CO₂-manipulation study set up in the Gulf of Finland, Baltic Sea (further references within
106 this special issue).

107

108

109 **2 Materials and methods**

110 **2.1 Experimental set-up**

111 Six pelagic mesocosms (approximately 55 m³) were moored at Storfjärden, SW coast of
112 Finland (59° 51.5' N; 23° 15.5' E) on 12 June 2012. The water depth at the mooring site is
113 approximately 30 meters and the mesocosms extended from the surface down to 19 m depth.
114 A more detailed description of the mesocosm bags and the experimental area can be found in
115 Paul et al. (2015).

116 On *t*-5 (17 June 2012, 5 days before the first CO₂ enrichment), the mesocosms were bubbled
117 with compressed air to break down any existing pycnocline and ensure homogeneous water
118 mass distribution. Different CO₂ concentrations in the bags were achieved by adding filtered
119 (50 μm), CO₂-saturated seawater. This was done stepwise in four separate additions to reduce
120 the shock of rapid change in pH on the plankton community. The first addition took place
121 after sampling on *t*0, thus *t*1 was the first day with a CO₂ treatment. The CO₂ enriched water
122 was evenly distributed over the upper 17 m using a specially designed distribution device i.e.

123 'spider' (Riebesell et al., 2013). Two controls and four treatment mesocosms were used.
124 Filtered water (with ambient CO₂ concentration) was added to the control mesocosms at the
125 time when CO₂ was manipulated in the treatment mesocosms. The CO₂ fugacity gradient on
126 *t4*, after the four additions, ranged from ambient (~240 μatm *f*CO₂) in the two control
127 mesocosms (M1 and M5), up to ~1650 μatm *f*CO₂, but we used the average *f*CO₂ throughout
128 the relevant part of this experiment (from *t1* – *t31*) to denote the different treatments: 346
129 (M1), 348 (M5), 494 (M7), 868 (M6), 1075 (M3) and 1333 (M8) μatm *f*CO₂. On *t15*,
130 additional CO₂-enriched seawater was added to the upper 7 m in the same manner as the
131 initial enrichment to counteract outgassing of CO₂. The mesocosm bags were regularly
132 cleaned to prevent wall growth. A more detailed description of the treatment and cleaning can
133 be found in Paul et al (2015).

134 Mesocosm sampling was carried out every day (or every second day for some variables) in
135 the morning. Two different depth-integrated water samples (0-10 m and 0-17 m) were taken
136 using integrating water samplers (IWS, HYDRO-BIOS, Kiel). The water was collected into
137 plastic carboys (10 L) and brought to the laboratory for sub-sampling and subsequent analysis
138 of plankton community composition, carbon concentration and for respiration and primary
139 production incubations. Sub-samples for primary production and respiration measurements
140 were treated and stored minimizing the contact with the air in order to prevent any gas
141 exchange.

142 Settling particles were quantitatively collected in the sediment traps at the bottom end of the
143 mesocosm units at 19 m water depth. Every 48 hours the accumulated material was vacuum
144 pumped through a silicon tube to the sea surface and transferred into 5 L glass bottles for
145 transportation to the laboratory. For a more detailed description of the sampling procedure
146 and sample processing of the sediment see Boxhammer et al. (2016).

147

148

149 **2.2 Phytoplankton community**

150 Phytoplankton cells were counted in 50 mL sub-samples, which were fixed with acidic
151 Lugol's iodine solution (1% final concentration) with an inverted microscope (ZEISS
152 Axiovert 100) after Utermöhl (1958). The cells >20 μm were counted either from half of the
153 chamber at 100-fold or on 3 to 4 stripes at 200-fold magnification. Filamentous cyanobacteria
154 were counted in 50 μm length units. Cells 12 - 20 μm were counted at half of the chamber at
155 200-fold magnification, and cells 4-12 μm were counted at 400-fold magnification on two

156 radial strips. The phytoplankton counts of the smaller size classes (<20µm) stopped on *t29*,
157 and these results have been used together with the *t31* results for larger (>20 µm)
158 phytoplankton as the end date of the experiment. Further details can be found in Bermúdez et
159 al. (2016)

160 Phytoplankton, heterotrophic dinoflagellates and protozoa were identified with the help of
161 Tomas (1997); Hoppenrath et al. (2009) and Kraberg et al. (2010). Biovolumes of counted
162 plankton cells were calculated according to Olenina et al. (2006) and converted to cellular
163 organic carbon quotas by the equations of Menden-Deuer and Lessard (2000).

164

165 **2.3 Microzooplankton community**

166 Ciliates were enumerated from 50 ml sub-samples every second day with a Zeiss Axiovert
167 100 inverted microscope (Utermöhl 1958) at 200 x magnification. At high cell numbers (>
168 400 cells), half the bottom plate area was counted. If less than 400 cells were found in the
169 first half of the bottom plate area, the entire chamber was counted. Rare species were counted
170 on the whole bottom plate. Ciliates were identified to the lowest possible taxonomic level
171 (genus/species) according to Setälä et al. (1992); Telesh et al. (2009) and to description plates
172 found at the planktonic ciliate project (<http://ciliate.zooplankton.cn/>). For more details see
173 Lischka et al. (2015) in this issue.

174

175 **2.4 Mesozooplankton community**

176 The term zooplankton includes here all metazoan species, i.e. organisms belonging strictly
177 speaking to the micro- or mesozooplankton, respectively. Zooplankton samples were
178 collected by net hauls from 17 m depth with an Apstein net of 17 cm diameter and 100 µm
179 mesh size. After closing of the mesocosm bags, zooplankton samples were taken prior to the
180 CO₂ addition on *t0* and at *t17* and *t31* (there were also other sampling days for zooplankton
181 but these are not included here). Samples were preserved in 70% ethanol. Zooplankton was
182 counted assuming 100% filtering efficiency of the net. The samples were divided with a
183 Folsom plankton splitter (1:2, 1:4, 1:8, 1:16, and 1:32) and aliquots were counted using a
184 WILD M3B stereomicroscope. Abundant species/taxa were enumerated from sub-samples (>
185 30 individuals in an aliquot) while less abundant and rare species/taxa were counted from the
186 whole sample. For more details on mesozooplankton collection, processing and species
187 determination, see Lischka et al. (2015). Carbon biomass (CB) in µmol C L⁻¹ was calculated
188 using the displacement volume (DV) and the equation of Wiebe (1988):

189 $(\text{Log DV} + 1.429)/0.82 = \text{log CB}$ (1)

190

191 **2.5 Total particulate carbon**

192 Samples for total particulate carbon (TPC) measurements were sub-sampled from 10 L
193 carboys and filtered onto GF/F filters (Whatman, nominal pore size of 0.7 μm , diameter = 25
194 mm) under reduced vacuum (< 200 mbar). Sampling for TPC occurred every 2nd day from $t-3$
195 until the end of the experiment. Filters were stored in glass petri dishes at -20°C directly after
196 filtration until preparation of samples for analyses. Petri dishes and filters were combusted at
197 450°C for 6 hours before use.

198 Samples were analyzed for total particulate carbon (organic + inorganic) as no acidifying step
199 was made to remove particulate inorganic carbon. Filters were dried at 60°C and packed into
200 tin capsules and stored in a desiccator until analysis on an elemental analyzer (EuroEA) as
201 described by (Sharp, 1974).

202 The particles collected from the sediment traps were allowed to settle down in the sampling
203 flasks at in-situ temperature before separation of supernatant and the dense particle
204 suspension at the bottom. TPC content of the supernatant was analysed from 10–50 mL sub-
205 samples as described above for water column measurements. The dense particle suspension
206 was concentrated by centrifugation, then freeze-dried and ground to a very fine powder of
207 homogeneous composition. From this material, small sub-samples of 1–2 mg were
208 transferred into tin capsules and TPC content was analysed analogue to the supernatant and
209 water column samples. Vertical carbon flux was calculated from the two measurements and is
210 given as the daily amount of TPC (mmol) collected in the sediment traps per square meter of
211 mesocosm surface area (3.142 m^2).

212

213 **2.6 Dissolved inorganic carbon**

214 Samples for dissolved inorganic carbon (DIC) were gently pressure-filtered (Saarstedt
215 Filtropur 0.2 μm) before measurements to remove all particulates. DIC concentrations were
216 determined by infrared absorption (LICOR LI-7000 on an AIRICA system, Marianda). Four
217 (2 mL) replicates were measured, and the final DIC concentration was calculated from the
218 mean of the three most consistent samples.

219

220 **2.7 Plankton community respiration**

221 Samples for respiration rate measurements were subsampled from the depth integrated
222 sample from the entire water column (0–17 m). Oxygen was measured using a fiber optical
223 dipping probe (PreSens, Fibox 3), which was calibrated against anoxic (0% O_2 , obtained by

224 adding sodium dithionite) and air saturated water (obtained by bubbling sampled water with
225 air for 5 minutes followed by 15 minutes of stirring with a magnetic stirrer). The final O₂
226 concentration was calculated using the Fibox 3 software including temperature compensation.
227 We filled three replicate 120 mL O₂ bottles (without headspace) for each mesocosm. After
228 the initial O₂ determination, the bottles were put in a dark, temperature controlled room, set to
229 the ambient water temperature at the surface. The O₂ concentration was determined again
230 after an incubation period of 48 hours, and the oxygen consumption (i.e. respiration rate) was
231 calculated from the difference between the O₂ concentration before and after the incubation
232 period. Respiration rates were measured every day *t-3* to *t31*, with the exception of days: *t2*
233 and *t14* because of technical problems.

234

235 **2.8 Primary production**

236 Primary production was measured using radio labeled NaH¹⁴CO₃ (Steeman-Nielsen, 1952)
237 from the 0-10 m depth integrated sample. The rationale for using the upper (0-10 m) part of
238 the mesocosm was the low light penetration depth, and 0-10 m was representative of the
239 euphotic zone. The water was gently filled into 12 small (8 mL) scintillation vials per
240 mesocosm and 10 µl of ¹⁴C bicarbonate solution (DHI Lab; 20 µCi mL⁻¹), was added. The
241 vials were filled completely and after adding the cap there was only a very small (2-3 mm) air
242 bubble remaining corresponding to ~0.1% of total volume.

243 Duplicate samples for each mesocosm were incubated just below the surface and at 2, 4, 6, 8
244 and 10 m depths for 24 h on small incubation platforms moored next to the mesocosms (Fig.
245 S1). In addition, a dark incubation (vials covered with aluminium foil) was incubated at the
246 same location at 11 m depth.

247 After incubation, 3 mL of the sample was removed from each vial and acidified with 100 µl 1
248 mol L⁻¹ HCl, and left without a lid for 24 h to ensure removal of remaining inorganic ¹⁴C.
249 Four mL of scintillation cocktail (Instagel Plus, Perkin Elmer) was added, and the
250 radioactivity was determined using a scintillation counter (Wallac 1414, Perkin Elmer).
251 Primary production was calculated knowing the ¹⁴C incorporation (with dark values
252 subtracted) and the fraction of the ¹⁴C addition to the total inorganic carbon pool according to
253 Gargas (1975). The primary production incubations were set up at the same time as the
254 respiration incubations, but here we missed measurements for two periods: *t1- t3* and *t6- t8*,
255 due to loss of the incubation platform.

256

257 **2.9 Data treatment**

258 The average of the three respiration bottles was used to calculate the respiration rate. There
259 were two days without measurements: t_2 and t_{14} and for these days we estimated the
260 respiration rate by using the average of the day before and after this day. TPC was measured
261 only every second day, therefore for the days without TPC measurements we normalized
262 respiration to average TPC from the day before and the day after the respiration
263 measurement.

264 The cumulative respiration was calculated by adding the total oxygen consumption for each
265 day. When evaluating the data, there were two clear periods emerging from the experiment:
266 the initial period $t_0 - t_{16}$ (Phase I) and period from $t_{17} - t_{31}$ (Phase II) when the effect of the
267 CO_2 addition was more evident. This division was also seen in e.g. Chl a and temperature
268 (Paul et al. 2015). Using the respiration data from Phase II we calculated the average
269 respiration for each treatment by linear regression. From the linear regression, the standard
270 error (SE) from the residuals and the coefficient of determination (R^2) were calculated, in
271 addition to a statistical test comparing the linear regression with a flat line, using Sigma Plot
272 software.

273 The areal primary production was calculated based on a simple linear model of the
274 production measurements from the different depths (Fig S2). The cumulative primary
275 production was carried out similar to respiration, but as the two missing periods were >1 day,
276 we did not estimate missing values, and the final cumulative production is therefore a slight
277 underestimate (missing 6 days of production). We normalized the production data to the TPC
278 in the euphotic zone, defined by the areal production model (Fig S2).

279 From the two different phases of the experiment (Phases I and II; $t_0 - t_{16}$ and $t_{17} - t_{31}$
280 respectively) we calculated the average for the different parameters and SE, with 9 and 7
281 sampling points during Phase I and II respectively.

282

283 **3 Results**

284 **3.1 Phytoplankton community composition**

285 The phytoplankton community in the mesocosms was dominated by dinoflagellates,
286 cyanobacteria, diatoms, chrysophytes and chlorophytes at the start of the experiment (Fig 1).
287 The two latter groups consisted almost exclusively of small cells ($<20 \mu\text{m}$). There was an
288 initial increase in phytoplankton biomass from an average of $3 \mu\text{mol C L}^{-1}$ to a maximum of
289 $\sim 4.1 \mu\text{mol C L}^{-1}$ in the two controls (M1 and M5), but at the end of Phase I (t_0-t_{16}) the
290 biomass had declined and at t_{17} it ranged between 3.2 to $3.5 \mu\text{mol C L}^{-1}$. During Phase I,
291 large ($>20\mu\text{m}$) diatoms decreased in abundance and euglenophytes increased from a

292 negligible group initially (0.5% of the biomass) to constituting 15-25% of the autotrophic
293 biomass at *t17*. It was, however, the small (<20 μm) phytoplankton cells (small diatoms,
294 chrysophytes and chlorophytes) that made up the majority (70-80%) of the counted autotroph
295 biomass during Phase I.

296 During Phase II (*t17–t31*), there was a decline in phytoplankton biomass to 0.5-1 $\mu\text{mol C L}^{-1}$
297 and at *t31* dinoflagellates had become the dominating group in all treatments except at the
298 highest CO_2 level. Cyanobacteria and chlorophytes were also abundant and the dominating
299 groups in the highest CO_2 . There was no consistent difference between phytoplankton
300 communities in the different CO_2 treatments, but dinoflagellate abundance was lower in the
301 highest CO_2 treatment (M8), and consequently the total phytoplankton biomass was lower in
302 this treatment at *t31*. The relative increase of large dinoflagellates decreased the contribution
303 of the smaller autotroph size class (4-20 μm) to 40-60% of the counted phytoplankton
304 biomass at *t31*.

305

306 **3.2 Zooplankton community composition**

307 Protozoans, ciliates and heterotrophic dinoflagellates dominated the microzooplankton and
308 constituted a major part (2.8 $\mu\text{mol C L}^{-1}$) of the whole zooplankton community at the start of
309 the experiment (Fig. 2). Protozoans, dominated by the choanoflagellate *Calliacantha natans*,
310 decreased from the initial high concentrations during Phase I, in particular in the M1 control
311 bag. The photosynthesizing, *Myrionecta rubra* (syn. *Mesodinium rubrum*) made up
312 approximately half of the ciliate biomass at *t0*, but both this species and the total biomass of
313 ciliates decreased during Phase I. The biomass of heterotrophic dinoflagellates was relatively
314 stable throughout Phase I, but started to decrease during Phase II.

315 The mesozooplankton community was initially dominated by copepods, cladocerans and
316 rotifers (Fig. 2). The average initial biomass was 0.05 $\mu\text{mol C L}^{-1}$ and increased to 0.13 $\mu\text{mol C L}^{-1}$
317 C L^{-1} at *t17*. During Phase I, copepods became the dominating group with >50% of the
318 mesozooplankton biomass. In Phase II of the experiment, mesozooplankton biomass
319 increased and was on average 0.27 $\mu\text{mol C L}^{-1}$ at *t31*. This was caused by an increase in
320 cladocerans, mainly *Bosmina* sp., whereas copepod biomass was more constant over the
321 course of the experiment. The population peak of *Bosmina* sp. had slightly different timing in
322 the different mesocosms but was higher in the mesocosms with added CO_2 , except for the
323 highest CO_2 addition (M8).

324

325 **3.3 Total particulate carbon and export of carbon**

326 Average TPC was $22.5 \mu\text{mol C L}^{-1}$ at the beginning of the experiment and after an initial
327 increase to $32 \mu\text{mol C L}^{-1}$ it decreased to $19.2 \mu\text{mol C L}^{-1}$ at *t17* (Fig 3). In the beginning of
328 Phase II it was relatively stable and with no clear effect of CO_2 treatment, but at the end of
329 the study period (*t31*) there was more TPC in the higher CO_2 treatments, and the increase in
330 TPC during Phase II was highest in the CO_2 additions (Table 1). At *t31* the average TPC was
331 $19.9 \mu\text{mol C L}^{-1}$, ranging from 18.9 ± 0.6 (SE) $\mu\text{mol C L}^{-1}$ in the controls to $22.1 \mu\text{mol C L}^{-1}$
332 in the highest CO_2 treatment.

333 The carbon accounted for by biologically active organisms counted in the microscope
334 (phytoplankton and zooplankton) was initially 26% of the TPC. At *t17* and *t31* this
335 percentage decreased to ~20% and ~8% respectively.

336 The export of carbon, defined here as carbon settling out of the mesocosms, decreased during
337 the experiment and there was no effect of CO_2 concentration. The average export of TPC was
338 in the range of $6.1 - 7.4 \text{ mmol C m}^{-2} \text{ d}^{-1}$ during Phase I (Table 1). This decreased to $2.5 - 3.3$
339 $\text{mmol C m}^{-2} \text{ d}^{-1}$ during Phase II.

340

341 **3.4 Primary production and respiration**

342 There was no clear effect of CO_2 addition on primary production (Fig. 4). There were
343 relatively large daily variations in depth-integrated primary production depending on the light
344 environment, and days with clear skies and more light increased carbon fixation. One of the
345 control bags (M1) had clearly lower primary production from the very start of the
346 experiment, and this was evident even before the initiation of the CO_2 addition (Fig 4). The
347 average production during the whole experiment was 3.67 ± 0.42 (SE) $\text{mmol C m}^{-2} \text{ d}^{-1}$ in M1,
348 and for all other bags 10.5 ± 0.67 (SE) $\text{mmol C m}^{-2} \text{ d}^{-1}$. Production on clear, sunny days was
349 (except for M1) approximately $25 \text{ mmol C m}^{-2} \text{ d}^{-1}$. The general pattern in areal primary
350 production was similar to TPC-normalized production (Table 1). Cumulative production
351 values in mol C m^{-2} are presented in the supplementary material (Fig S3).

352 The respiration rate was higher in the ambient than the high CO_2 treatments (Fig 5). In one of
353 the two controls (M1), the respiration rate was clearly higher compared to all other treatments
354 from the beginning of the experiment. The respiration rate in the other control (M5) increased
355 approximately two weeks later than the CO_2 treatments. After *t17*, the mesocosm with
356 highest CO_2 concentration (average of $1333 \mu\text{atm } f\text{CO}_2$) started to have lower cumulative
357 respiration compared to those with intermediate CO_2 levels ($494\text{-}1075 \mu\text{atm } f\text{CO}_2$). After
358 another week (~*t27*), differences between the intermediate CO_2 treatments became apparent.
359 At the end of Phase II (*t20-t31*), there was a 40% difference in respiration rate between the

360 lowest and highest $f\text{CO}_2$ treatments (slope -0.0002; $p = 0.02$; $R^2 = 0.77$; Fig. 6). The
361 volumetric respiration during Phase II was 7.6 and 7.1 $\mu\text{mol O}_2 \text{ L}^{-1} \text{ d}^{-1}$ for the two controls,
362 and 4.7 - 5.7 $\mu\text{mol O}_2 \text{ L}^{-1} \text{ d}^{-1}$ for the CO_2 treatment mesocosms. Outside the mesocosms, at
363 ambient CO_2 concentration (average of 343 $\mu\text{atm } f\text{CO}_2$ but with larger variability than inside
364 the mesocosms), the carbon normalized respiration rate was lower than inside the mesocosms
365 and the cumulative, carbon-normalized respiration was approximately half of that measured
366 in the control bags at the end of the experiment (Fig 5). The general pattern of lower
367 respiration rates at high CO_2 concentration was the same without normalization to TPC
368 (Table 1, Fig S4).

369

370 **4 Discussion**

371 **4.1 Plankton community**

372 The particulate and dissolved standing stocks during this experiment are presented in Paul et
373 al. (2015). In the initial Phase I of the experiment the Chl *a* concentration was relatively high
374 ($\sim 2 \mu\text{g Chl } a \text{ L}^{-1}$), but started to decrease during Phase II, and reached $\sim 1 \mu\text{g Chl } a \text{ L}^{-1}$ at *t31*
375 in all of the treatments. During this transition there was a shift in the plankton community
376 with decreasing phytoplankton and microzooplankton, and increasing abundance of
377 mesozooplankton, primarily cladocerans (Figs 1 and 2).

378 The phytoplankton community composition was dominated by common species in the area
379 (Hällfors, 2004). In the latter part (Phase II), the relative dominance by dinoflagellates was
380 mainly due to reduction in biomass of the other groups, with the exception of the highest CO_2
381 concentration where also the dinoflagellates decreased in abundance. Dinoflagellates are
382 generally favored in low turbulence (Margalef, 1978; Smayda and Reynolds, 2001), and were
383 probably benefitting from the relative stable conditions within the mesocosms. Blooms of
384 filamentous cyanobacteria do occur in the area, but did not develop within the mesocosms.
385 The relatively low temperature (mostly $<15^\circ\text{C}$; Paul et al., 2015) could be a reason for that
386 (Kanoshina et al., 2003).

387 Protozoans, ciliates and heterotrophic dinoflagellates dominated the microzooplankton, and
388 *Myrionecta rubra* initially made up a large proportion of the ciliates. *M. rubra* can be
389 regarded as mixotrophic and would also have contributed to the carbon fixation (Johnson et al.,
390 2006). Copepods and cladocerans initially dominated the mesozooplankton, and during Phase
391 II, cladocerans became the dominant mesozooplankton group. Cladocerans are typically
392 predominant in freshwater but in the brackish Baltic Sea they can be common, in particular
393 when stability in the water column is high (Viitasalo et al., 1995).

394 The combined phyto- and zooplankton carbon derived from microscope counts decreased
395 during the experiment. TPC did not decrease to the same extent, and the percentage
396 microscope-derived carbon of TPC decreased from 26% at t_0 to only ~8% of the measured
397 TPC at t_{31} . These numbers are not directly comparable, as detritus, i.e. non-living carbon
398 particles, are included in TPC. However, any large aggregates sink rapidly and are not
399 expected to have contributed much to the TPC. The reduction of microscopy-derived carbon
400 to TPC indicate rather increasing importance of smaller size classes ($<4 \mu\text{m}$), not enumerated
401 by the microscope counts. This conclusion is also supported by flow cytometer data from this
402 experiment (Crawford et al., 2016), increasing uptake of PO_4 by the $<3 \mu\text{m}$ fraction (Nausch
403 et al., 2016) and the increasing proportion of the smallest ($<2 \mu\text{m}$) size class of Chl *a* (Paul et
404 al., 2015).

405

406 **4.2. Primary production and respiration**

407 Primary production and respiration rates were comparable to values obtained under similar
408 conditions in the area (Kivi et al., 1993). There are relatively few records of respiration, but
409 the measured respiration rates in the control bags were similar to the average respiration rate
410 obtained for a range of coastal waters of $7.4 \pm 0.54 \text{ mmol O}_2 \text{ m}^{-3} \text{ d}^{-1}$ ($n=323$) (Robinson and
411 Williams, 2005). The incubation period we used for primary production measurements (24 h)
412 provides production rates close to net production (Marra, 2009).

413 The higher respiration and lower production in the M1 control bag was probably connected,
414 i.e. higher respiration lead to lower net carbon fixation, however, the reason for the M1 bag
415 being very different from the very start is not clear. Most of the other variables were similar
416 in the M1 bag compared to the rest (Paul et al., 2015), but there was some indication of
417 difference in community. In particular, protozoans were lower in the M1 bag compared with
418 the rest of the mesocosms throughout the experiment. However, judging from the
419 development in carbon pools (Paul et al., 2015) and fluxes in the system (Spilling et al.,
420 2016), the NPP measurements for the M1 bag must be an underestimate. Bacterial production
421 during Phase II was highest in the ambient CO_2 , in particular in M1 (Hornick et al., 2016),
422 and could partly be the reason for the elevated respiration rate in this mesocosm bag.

423 Having the respiration incubation at a fixed temperature might have caused a slight bias as
424 there was varying thermal stratification throughout the experiment and the temperature was
425 not even throughout the mesocosm bags. A better approach would have been to have
426 respiration incubations in temperatures above and below the thermocline, but logistical
427 constrains prevented us from doing this.

428 Another factor that could have influenced our incubations is UV light, which is a known
429 inhibitor of primary production (Vincent and Roy, 1993), and elevated CO₂ concentration
430 may increase the sensitivity to UV light (Sobrino et al., 2009). Additionally, UV light reduces
431 the release of DOC by phytoplankton, in particular at high CO₂ concentration (Sobrino et al.,
432 2014), but also cause photochemical mineralization of dissolved organic matter (DOM)
433 (Vahatalo and Jarvinen, 2007). Both DOC release and DOM break down may have
434 implications for bacterial production and nutrient cycling. The mesocosm bags were made in
435 a material absorbing UV light (thermoplastic polyurethane) whereas our primary production
436 incubations were done in glass vials (transmitting some UV light) moored outside the
437 mesocosm bags. The difference in UV transmittance could have produced a bias in the
438 primary production measurements. However, the DOM concentration in the Baltic Sea is
439 very high compared with most other oceans and coastal seas (Hoikkala et al., 2015). Most of
440 this is terrestrial derived, refractory DOM, which effectively absorbs in the UV region, and
441 typically the depth at which 1% of UVB remains is <50 cm (Piazena and Häder, 1994). UVA
442 penetrates a little deeper and may have affected slightly the incubation platform moored at 2
443 m depth, but we do not believe that UV light caused major inhibition of our primary
444 production measurements or affected phytoplankton DOC production.

445
446

447 **4.3. Effect of CO₂ on the balance between respiration and carbon fixation**

448 Increased CO₂ concentration has increased carbon fixation in some studies (Egge et al., 2009;
449 Engel et al., 2013). This was not observed in this study, but the higher Chl *a*, TPC and DOC
450 in the high CO₂ treatments at the end of the experiment (Paul et al., 2015) could have been
451 caused by the lower respiration rate in the highest CO₂ enriched mesocosms, rather than
452 increased primary production. Bacterial production was higher in the low CO₂ after *t*20
453 during this experiment (Hornick et al., 2016), which fits with the higher respiration rate at
454 ambient CO₂ concentration. The biomass of the smallest plankton size fraction (<4 μm, not
455 counted by microscope) increased in relative importance with CO₂ addition in the latter part
456 of the experiment, in particular two groups of pico-eukaryotes (Crawford et al., 2016), and
457 seems to have benefitted most by elevated CO₂ concentration, similar to findings in the
458 Arctic (Brussaard et al., 2013). Temporal changes in bacterial abundances followed largely
459 that of phytoplankton biomass, and there were significant increases in viral lysis rates in the
460 high CO₂ treatment (Crawford et al., 2016). This was most likely a consequence of higher

461 abundances of pico-eukaryotes and pointing towards a more productive but regenerative
462 system (Crawford et al., 2016).

463 This study is, to our knowledge, the first one describing reduced respiration rates with ocean
464 acidification on a plankton community scale. There are relatively few measurements of
465 community respiration in ocean acidification experiments, and existing studies have revealed
466 no specific responses in respiration (Egge et al., 2009; Tanaka et al., 2013; Mercado et al.,
467 2014). Some of these studies have been relatively short (<2 weeks) compared to the current
468 study. Our results revealed a CO₂ effect only two weeks into the experiment, suggesting that
469 potential effects may have been present but remained below the detection limits in previous
470 studies.

471 The effect of increasing CO₂ concentration on respiration has mostly been documented for
472 single species. For example, the copepod *Centropages tenuiremis* (Li and Gao, 2012) and the
473 diatom *Phaeodactylum tricornutum* (Wu et al., 2010) exhibited increased respiration rates in
474 a high CO₂ environment ($\geq 1000 \mu\text{atm } f\text{CO}_2$), contrary to our findings. However, these types
475 of studies have revealed different responses even when comparing different populations of
476 the same species (Thor and Oliva, 2015), and any interpolation from single-species,
477 laboratory-studies should be carried out with great caution. The larger-scale mesocosm
478 approach taken here has the advantage that the whole plankton community and possible
479 interacting effects between different components of the food web are included.

480 In higher plants, it is known that elevated CO₂ decreases mitochondrial respiration in the
481 foliage (Puhe and Ulrich, 2012). In their review, Drake et al. (1999) outlined two CO₂ effects
482 on respiration: an immediate, reversible effect and a longer term, irreversible effect, both
483 decreasing respiration in a high CO₂ environment. In our study it was only a longer term
484 effect that was observed. It is not known what cause this reduced respiration in plant foliage,
485 but Amthor (1991) pointed out seven potential mechanisms for how changes in the CO₂
486 concentration could reduce plant respiration, for example by affecting respiratory enzymes. A
487 doubling of present day CO₂ concentration could decrease foliage respiration rate by 15 to
488 30% (Drake et al., 1999; Puhe and Ulrich, 2012), but other parts e.g. root system are
489 projected to increase respiration so the net effect of elevated CO₂ on plant respiration is
490 uncertain (Puhe and Ulrich, 2012). Phytoplankton lacks any specialized structures like root
491 system and may consequently function more like plant foliage.

492 The intracellular pH can be highly variable between different cellular compartments and
493 organelles, but in the cytosol the pH is normally close to neutral (pH ~7.0), and is to a large
494 extent independent of the external pH (Roos and Boron, 1981). In plants, animals and also

495 bacteria, there is a complex set of pH regulatory mechanisms that is fundamentally controlled
496 by physiological processes such as membrane transport of H^+ or OH^- and intracellular
497 metabolism (Smith and Raven, 1979; Kurkdjian and Guern, 1989). Internal pH regulation can
498 be a considerable part of baseline respiration (Pörtner et al., 2000). With ocean acidification,
499 the external pH becomes closer to the intracellular pH, and this might reduce the metabolic
500 cost (respiration) related to internal pH regulation. However, this is not straight forward and
501 more studies of the effect of changed external pH on membrane transport are needed (Taylor
502 et al., 2012). There might additionally be considerable difference between marine organisms
503 depending on e.g. size, metabolic activity and growth rates, which directly affect pH in the
504 diffusive boundary layer surrounding the organism (Flynn et al., 2012).

505 Changes in carbonate chemistry speciation might also affect the availability of the sole
506 substrate, i.e. CO_2 , at the site of photosynthetic carbon fixation. At present, marine waters
507 typically have a pH of 8 or above, and most of the carbon is in the form of bicarbonate
508 (HCO_3^-). Many phytoplankton groups have developed carbon concentrating mechanisms
509 (CCMs), such as the active uptake of bicarbonate, as a way to increase substrate availability
510 at the site of carbon fixation (Singh et al., 2014). Increased CO_2 availability may reduce
511 metabolic activity related to CCMs, which would affect the respiration rate of primary
512 producers.

513 Judging from the importance of the smallest size class in this study, bacterial and
514 picophytoplankton community (Crawford et al., 2016) and bacterial production (Hornick et
515 al., 2016), the decreased respiration at higher CO_2 concentration was probably mostly due to
516 reduced picoplankton respiration. The underlying mechanisms behind the reduced respiration
517 are unclear and this is an underexplored research avenue that deserves further study.

518

519

520 **4.4. Interacting effects and community composition**

521 Our measurements outside the mesocosm bags demonstrate that plankton physiology and
522 community composition can have a big impact on both primary production and respiration.
523 The plankton community was relatively uniform across all mesocosm bags. Unfortunately,
524 we do not have any community data from outside the mesocosm bags, but the amplitude of
525 Chl *a* dynamics was different, with an upwelling event leading to a doubling of the Chl *a*
526 concentration ($\sim 5 \mu g \text{ Chl } a \text{ L}^{-1}$) around *t17* (Paul et al., 2015). This suggests a different
527 availability of inorganic nutrients and different plankton community as other environmental
528 variables such as light and temperature were similar both inside and outside the mesocosm

529 bags, except that UV light was absent inside the mesocosm bags. The carbon-normalized
530 respiration rate outside the mesocosm bags (with ambient $f\text{CO}_2$) was approximately half of
531 the respiration rates in the controls with the same average $f\text{CO}_2$, and also absolute respiration
532 was clearly lower during Phase II, when nitrate was depleted inside the bags and plankton
533 biomass was decreasing. However, the $f\text{CO}_2$ was more variable outside the mesocosm bags
534 compared with the control bags (although their averages were similar), and the $f\text{CO}_2$
535 increased throughout Phase II outside the bags to approximately 700 μatm by *t31* (Paul et al.
536 2015). This could have influenced the carbon normalized respiration, which started to deviate
537 outside the bags during Phase II, but it could also have been interacting effects of different
538 environmental changes (different nutrient dynamics) leading to this lower respiration rate. An
539 often overlooked aspect is the importance of the plankton community composition, which can
540 be more important than changes in external factors (Verity and Smetacek, 1996; Eggers et al.,
541 2014).

542 Bacterial production (Grossart et al., 2006) and bacterial degradation of polysaccharides
543 (Piontek et al., 2010) have been demonstrated to increase under elevated CO_2 concentration,
544 contrary to the findings during this experiment (Hornick et al., 2016). All of these responses
545 are to a large extent dependent on the plankton community composition. For example, the
546 increased bacterial production observed in a mesocosm study in a Norwegian fjord was
547 probably a response to increased carbon availability produced by phytoplankton (Grossart et
548 al., 2006). DOC production by phytoplankton is determined by the physiological state and the
549 composition of the community (Thornton, 2014); in particular diatoms have been intensively
550 studied in this respect and are known to be important DOC producers (Hoagland et al., 1993).
551 Shifts in the phytoplankton community may alter the DOC production (Spilling et al., 2014),
552 and any shifts in the plankton community composition, caused by ocean acidification, may
553 have greater effects on ecosystem functioning than any direct effect of increasing $f\text{CO}_2$ /
554 decreasing pH (Eggers et al., 2014).

555 It is evident that there were other variables that influence the physiology of the plankton
556 community as a whole outside the mesocosms. Changes in community composition and
557 nutrient availability seem the most plausible reasons. A better understanding of how different
558 physical, chemical and biological factors interact with each other is needed in order to
559 improve our understanding of how marine ecosystems change under the influence of a range
560 of environmental pressures.

561

562 **4.5. Potential implications for carbon cycling**

563 A lot of attention during past decades has been directed to understanding the biological
564 carbon pump, as it is a key mechanism for sequestering atmospheric CO₂. The potential
565 export is ultimately determined by gross primary production minus total community
566 respiration. Even small changes in the production or loss term of this equation have the
567 potential to greatly affect biogeochemical cycling of carbon.

568 The exported carbon decreased during the experiment. Part of this decrease was probably due
569 to sinking of existing organic material at the start of the experiment and can be seen as the
570 reduction in TPC. However, this also coincided with the shift towards increased dominance
571 of picoplankton. Size is a key parameter determining sinking speed, and picoplankton is very
572 inefficient in transporting carbon out of the euphotic layer (Michaels and Silver, 1988). The
573 shift towards smaller size classes was likely also contributing to the reduction in exported
574 carbon.

575 The 40% reduction in respiration with increasing *f*CO₂ found in our study could have great
576 implications for net export of carbon in the future ocean. There is, however, uncertainty in the
577 results, in particular that the measured net carbon fixation under increased CO₂ was not
578 higher than in the controls. In the case of reduced respiration, an increase in net primary
579 production can be expected, as loss rates are reduced. That the measured carbon fixation was
580 not evidently different between treatments could be due to similar reduction in GPP, as
581 indicated by carbon flux estimates (Spilling et al., 2016). Alternatively, the measurement
582 uncertainty in our small scale incubations (8 mL), involving several pipetting steps, was
583 likely higher than the respiration measurements, which could have prevented us from picking
584 up any CO₂ effect on primary production. Another complicating factor is what the ¹⁴C
585 method is actually measuring (Sakshaug et al., 1997; Falkowski and Raven, 2013). The
586 consensus seems to be somewhere between gross and net production, but leaning towards net
587 production with long incubation times (Marra, 2009).

588 There was evidence of a positive CO₂ effect on the amount of Chl *a*, TPC and DOC pools
589 (Paul et al., 2015), suggesting that the reduced respiration does translate into higher net
590 carbon fixation. This effect was seen from the latter part of Phase II and the trend continued
591 after *t*₃₁ (these variables were sampled until *t*₄₃). This increased net carbon fixation did not,
592 however, affect carbon export as there was no detectable difference in the sinking flux
593 measurements (Table 1 and Paul et al. 2015). The results suggest that the increased carbon
594 fixation ended up in the smallest size fraction of TPC not being exported and/or into the
595 dissolved organic carbon pool. Further support for this conclusion is presented in Paul et al.
596 (2015), Crawford et al. (2016) and Lischka et al. (2015).

597 In conclusion, this study suggests that elevated CO₂ reduced respiration which in turn
598 increased net carbon fixation. However, the increased primary production did not translate
599 into increased carbon export, and did consequently not work as a negative feedback
600 mechanism for increasing atmospheric CO₂ concentration.

601

602 **Acknowledgements**

603 We would like to thank all of the staff at Tvärminne Zoological station for great help during
604 this experiment, and Michael Sswat for carrying out the TPC filtrations. We also gratefully
605 acknowledge the captain and crew of R/V ALKOR (AL394 and AL397) for their work
606 transporting, deploying and recovering the mesocosms. The collaborative mesocosm
607 campaign was funded by BMBF projects BIOACID II (FKZ 03F06550) and SOPRAN Phase
608 II (FKZ 03F0611). Additional financial support for this study came from Academy of Finland
609 (KS - Decisions no: 259164 and 263862) and Walter and Andrée de Nottbeck Foundation
610 (KS, NV).

611

612

613 **References**

- 614 Amthor, J.: Respiration in a future, higher-CO₂ world, *Plant, Cell & Environment*, 14, 13-20,
615 1991.
- 616 Beldowski, J., Löffler, A., Schneider, B., and Joensuu, L.: Distribution and biogeochemical
617 control of total CO₂ and total alkalinity in the Baltic Sea, *J Mar Sys*, 81, 252-259, 2010.
- 618 Bermúdez, R., Winder, M., Stühr, A., Almén, A.-K., Engström-Öst, J., and Riebesell, U.:
619 Effect of ocean acidification on the structure and fatty acid composition of a natural
620 plankton community in the Baltic Sea, *Biogeosciences Discuss*, 10.5194/bg-2015-669,
621 2016.
- 622 Boxhammer, T., Bach, L. T., Czerny, J., and Riebesell, U.: Technical Note: Sampling and
623 processing of mesocosm sediment trap material for quantitative biogeochemical analyses,
624 *Biogeosciences* 13, 2849-2858, 2016.
- 625 Brussaard, C., Noordeloos, A., Witte, H., Collenteur, M., Schulz, K. G., Ludwig, A., and
626 Riebesell, U.: Arctic microbial community dynamics influenced by elevated CO₂ levels,
627 *Biogeosciences*, 10, 719-731, 2013.
- 628 Crawford, K. J., Riebesell, U., and Brussaard, C. P. D.: Shifts in the microbial community in
629 the Baltic Sea with increasing CO₂ *Biogeosciences Discuss*, 10.5194/bg2015-606, 2016.

630 Drake, B., Azcon-Bieto, J., Berry, J., Bunce, J., Dijkstra, P., Farrar, J., Gifford, R., Gonzalez-
631 Meler, M., Koch, G., and Lambers, H.: Does elevated atmospheric CO₂ concentration
632 inhibit mitochondrial respiration in green plants?, *Plant Cell Env*, 22, 649-657, 1999.

633 Ducklow, H. W., Steinberg, D. K., and Buesseler, K. O.: Upper ocean carbon export and the
634 biological pump, *Oceanogr*, 14, 50-58, 2001.

635 Dunne, J. P., Sarmiento, J. L., and Gnanadesikan, A.: A synthesis of global particle export
636 from the surface ocean and cycling through the ocean interior and on the seafloor, *Global*
637 *Biogeochem Cy*, 21, GB4006, 10.1029/2006GB002907, 2007.

638 Egge, J., Thingstad, J., Larsen, A., Engel, A., Wohlers, J., Bellerby, R., and Riebesell, U.:
639 Primary production during nutrient-induced blooms at elevated CO₂ concentrations,
640 *Biogeosciences*, 6, 877-885, 2009.

641 Eggers, S. L., Lewandowska, A. M., Barcelos e Ramos, J., Blanco-Ameijeiras, S., Gallo, F.,
642 and Matthiessen, B.: Community composition has greater impact on the functioning of
643 marine phytoplankton communities than ocean acidification, *Global Change Biol*, 20, 713-
644 723, 2014.

645 Engel, A., Borchard, C., Piontek, J., Schulz, K. G., Riebesell, U., and Bellerby, R.: CO₂
646 increases ¹⁴C-primary production in an Arctic plankton community, *Biogeosciences*, 10,
647 1291-1308, 2013.

648 Falkowski, P. G., and Raven, J. A.: *Aquatic photosynthesis; Second edition*, Princeton
649 University Press, Princeton, 501 pp., 2013.

650 Flynn, K. J., Blackford, J. C., Baird, M. E., Raven, J. A., Clark, D. R., Beardall, J., Brownlee,
651 C., Fabian, H., and Wheeler, G. L.: Changes in pH at the exterior surface of plankton with
652 ocean acidification, *Nature Clim Change*, 2, 510-513, 2012.

653 Gargas, E.: *A manual for phytoplankton primary production studies in the Baltic*, The Baltic
654 *Marine Biologist*, Hørsholm, Denmark, 88 pp., 1975.

655 Grossart, H.-P., Allgaier, M., Passow, U., and Riebesell, U.: Testing the effect of CO₂
656 concentration on the dynamics of marine heterotrophic bacterioplankton, *Limnol*
657 *Oceanogr*, 51, 1-11, 2006.

658 Hoagland, K. D., Rosowski, J. R., Gretz, M. R., and Roemer, S. C.: Diatom extracellular
659 polymeric substances: function, fine structure, chemistry, and physiology, *J Phycol*, 29,
660 537-566, 1993.

661 Hoikkala, L., Kortelainen, P., Soenne, H., and Kuosa, H.: Dissolved organic matter in the
662 Baltic Sea, *J Mar Sys*, 142, 47-61, 2015.

663 Hornick, T., Bach, L. T., Crawford, K. J., Spilling, K., Achterberg, E. P., Brussaard, C.,
664 Riebesell, U., and Grossart, H.-P.: Ocean acidification indirectly alters trophic interaction
665 of heterotrophic bacteria at low nutrient conditions, *Biogeosciences Discuss*,
666 doi:10.5194/bg-2016-61, 2016.

667 Johnson, M. D., Tengs, T., Oldach, D., and Stoecker, D. K.: Sequestration, performance, and
668 functional control of cryptophyte plastids in the ciliate *Myrionecta rubra* (ciliophora), *J*
669 *Phycol*, 42, 1235-1246, 2006.

670 Kanoshina, I., Lips, U., and Leppänen, J.-M.: The influence of weather conditions
671 (temperature and wind) on cyanobacterial bloom development in the Gulf of Finland
672 (Baltic Sea), *Harmful Algae*, 2, 29-41, 2003.

673 Kivi, K., Kaitala, S., Kuosa, H., Kuparinen, J., Leskinen, E., Lignell, R., Marcussen, B., and
674 Tamminen, T.: Nutrient limitation and grazing control of the Baltic plankton community
675 during annual succession, *Limnol Oceanogr*, 38, 893-905, 1993.

676 Kraberg, A., Baumann, M., and Dürselen, C.-D.: Coastal phytoplankton: photo guide for
677 Northern European seas, Pfeil, München, 2010.

678 Kurkdjian, A., and Guern, J.: Intracellular pH: measurement and importance in cell activity,
679 *Ann Rev Plant Biol*, 40, 271-303, 1989.

680 Le Quéré, C., Raupach, M. R., Canadell, J. G., Marland, G., Bopp, L., Ciais, P., Conway, T.
681 J., Doney, S. C., Feely, R. A., and Foster, P.: Trends in the sources and sinks of carbon
682 dioxide, *Nature Geosci*, 2, 831-836, 2009.

683 Li, W., and Gao, K.: A marine secondary producer respire and feeds more in a high CO₂
684 ocean, *Mar Pollut Bull*, 64, 699-703, 2012.

685 Lischka, S., Bach, L. T., Schulz, K.-G., and Riebesell, U.: Micro- and mesozooplankton
686 community response to increasing levels of *f*CO₂ in the Baltic Sea: insights from a large-
687 scale mesocosm experiment, *Biogeosciences Discuss*, 10.5194/bgd-12-20025-2015, 2015.

688 Mackey, K. R., Morris, J. J., Morel, F. M., and Kranz, S. A.: Response of photosynthesis to
689 ocean acidification, *Oceanogr*, 28, 74-91, 2015.

690 Margalef, R.: Life-forms of phytoplankton as survival alternatives in an unstable
691 environment, *Oceanol. Acta*, 1, 493-509, 1978.

692 Marra, J.: Net and gross productivity: weighing in with ¹⁴C, *Aquat Microb Ecol*, 56, 123-131,
693 2009.

694 Menden-Deuer, S., and Lessard, E. J.: Carbon to volume relationships for dinoflagellates,
695 diatoms, and other protist plankton, *Limnol Oceanogr*, 45, 569-579, 2000.

696 Mercado, J., Sobrino, C., Neale, P., Segovia, M., Reul, A., Amorim, A., Carrillo, P., Claquin,
697 P., Cabrerizo, M., and León, P.: Effect of CO₂, nutrients and light on coastal plankton. II.
698 Metabolic rates, *Aquat Biol*, 22, 43-57, 2014.

699 Michaels, A. F., and Silver, M. W.: Primary production, sinking fluxes and the microbial
700 food web, *Deep-Sea Res*, 35, 473-490, 1988.

701 Nausch, M., Bach, L. T., Czerny, J., Goldstein, J., Grossart, H.-P., Hellemann, D., Hornick,
702 T., Achterberg, E. P., Schulz, K.-G., and Riebesell, U.: Effects of CO₂ perturbation on
703 phosphorus pool sizes and uptake in a mesocosm experiment during a low productive
704 summer season in the northern Baltic Sea, *Biogeosciences* 13, 3035-3050, 2016.

705 Olenina, I., Hajdu, S., Edler, L., Andersson, A., Wasmund, N., Busch, S., Göbel, J., Gromisz,
706 S., Huseby, S., Huttunen, M., Jaanus, A., Kokkonen, P., Ledaine, I., and Niemkiewicz, E.:
707 Biovolumes and size-classes of phytoplankton in the Baltic Sea, *Balt. Sea Environ. Proc.*,
708 HELCOM, 144 pp., 2006.

709 Orr, J. C.: Recent and future changes in ocean carbonate chemistry, in: *Ocean acidification*,
710 edited by Gattuso J-P. and Hansson L., Oxford University Press, Oxford, 41-66, 2011.

711 Paul, A. J., Bach, L. T., Schulz, K.-G., Boxhammer, T., Czerny, J., Achterberg, E. P.,
712 Hellemann, D., Trense, Y., Nausch, M., Sswat, M., and Riebesell, U.: Effect of elevated
713 CO₂ on organic matter pools and fluxes in a summer Baltic Sea plankton community,
714 *Biogeosciences*, 12, 6181-6203, 2015.

715 Piazena, H., and Häder, D. P.: Penetration of solar UV irradiation in coastal lagoons of the
716 southern Baltic Sea and its effect on phytoplankton communities, *Photochem Photobiol*,
717 60, 463-469, 1994.

718 Piontek, J., Lunau, M., Handel, N., Borchard, C., Wurst, M., and Engel, A.: Acidification
719 increases microbial polysaccharide degradation in the ocean, *Biogeosciences*, 7, 1615–
720 1624, 2010.

721 Puhe, J., and Ulrich, B.: *Global climate change and human impacts on forest ecosystems:*
722 *postglacial development, present situation and future trends in Central Europe*, *Ecological*
723 *studies – analysis and synthesis*, Springer, Berlin, 476 pp., 2012.

724 Pörtner, H., Bock, C., and Reipschlager, A.: Modulation of the cost of pHi regulation during
725 metabolic depression: a (31) P-NMR study in invertebrate (*Sipunculus nudus*) isolated
726 muscle, *J Exp Biol*, 203, 2417-2428, 2000.

727 Raateoja, M., Seppala, J., Kuosa, H., and Myrberg, K.: Recent changes in trophic state of the
728 Baltic Sea along SW coast of Finland, *Ambio*, 34, 188-191, 2005.

729 Riebesell, U., and Tortell, P. D.: Effects of ocean acidification on pelagic organisms and
730 ecosystems, in: *Ocean acidification*, edited by Gattuso J-P. and Hansson L., Oxford
731 University Press, Oxford, 99-121, 2011.

732 Riebesell, U., Czerny, J., Bröckel, K. v., Boxhammer, T., Büdenbender, J., Deckelnick, M.,
733 Fischer, M., Hoffmann, D., Krug, S., and Lentz, U.: Technical Note: A mobile sea-going
734 mesocosm system—new opportunities for ocean change research, *Biogeosciences*, 10,
735 1835-1847, 2013.

736 Robinson, C., and Williams, P. I. B.: Respiration and its measurement in surface marine
737 waters, in: *Respiration in aquatic ecosystems*, edited by: Giorgio, P. A. and Williams, P. I.
738 B., Oxford University Press, Oxford, 147-181, 2005.

739 Roos, A., and Boron, W. F.: Intracellular pH, *Am Physiological Soc*, 61, 296-434, 1981.

740 Sakshaug, E., Bricaud, A., Dandonneau, Y., Falkowski, P. G., Kiefer, D. A., Legendre, L.,
741 Morel, A., Parslow, J., and Takahashi, M.: Parameters of photosynthesis: definitions,
742 theory and interpretation of results, *J Plankton Res*, 19, 1637-1670, 1997.

743 Savchuk, O. P.: Resolving the Baltic Sea into seven subbasins: N and P budgets for 1991-
744 1999, *J Mar Sys*, 56, 1-15, 2005.

745 Siegenthaler, U., and Sarmiento, J.: Atmospheric carbon dioxide and the ocean, *Nature*, 365,
746 119-125, 1993.

747 Singh, S. K., Sundaram, S., and Kishor, K.: *Photosynthetic microorganisms: Mechanism for*
748 *carbon concentration*, Springer, Berlin, 131 pp., 2014.

749 Smayda, T. J., and Reynolds, C. S.: Community assembly in marine phytoplankton:
750 application of recent models to harmful dinoflagellate blooms, *J Plankton Res*, 23, 447-
751 461, 2001.

752 Smith, F., and Raven, J. A.: Intracellular pH and its regulation, *Ann Rev Plant Physiol*, 30,
753 289-311, 1979.

754 Sobrino, C., Neale, P., Phillips-Kress, J., Moeller, R., and Porter, J.: Elevated CO₂ increases
755 sensitivity to ultraviolet radiation in lacustrine phytoplankton assemblages, *Limnol*
756 *Oceanogr*, 54, 2448, 2009.

757 Sobrino, C., Segovia, M., Neale, P., Mercado, J., García-Gómez, C., Kulk, G., Lorenzo, M.,
758 Camarena, T., van de Poll, W., Spilling, K., Ruan, Z.: Effect of CO₂, nutrients and light on
759 coastal plankton. IV. Physiological responses, *Aquat Biol*, 22, 77-93, 2014.

760 Spilling, K.: Dense sub-ice bloom of dinoflagellates in the Baltic Sea, potentially limited by
761 high pH, *J Plankton Res*, 29, 895-901, 2007.

762 Spilling, K., Kremp, A., Klais, R., Olli, K., and Tamminen, T.: Spring bloom community
763 change modifies carbon pathways and C: N: P: Chl a stoichiometry of coastal material
764 fluxes, *Biogeosciences*, 11, 7275-7289, 2014.

765 Spilling, K., Schulz, K. G., Paul, A. J., Boxhammer, T., Achterberg, E. P., Hornick, T.,
766 Lischka, S., Stuhr, A., Bermúdez, R., Czerny, J., Crawford, K., Brussaard, C. P. D.,
767 Grossart, H.-P., and Riebesell, U.: Effects of ocean acidification on pelagic carbon fluxes
768 in a mesocosm experiment, *Biogeosciences Discuss.*, doi:10.5194/bg-2016-56, 2016.

769 Steeman-Nielsen, E.: The use of radioactive carbon for measuring organic production in the
770 sea, *J Cons Int Explor Mer*, 18, 117-140, 1952.

771 Suikkanen, S., Laamanen, M., and Huttunen, M.: Long-term changes in summer
772 phytoplankton communities of the open northern Baltic Sea, *Estuar Coast Shelf Sci*, 71,
773 580-592, 2007.

774 Tanaka, T., Alliouane, S., Bellerby, R., Czerny, J., De Kluijver, A., Riebesell, U., Schulz, K.
775 G., Silyakova, A., and Gattuso, J.-P.: Effect of increased pCO₂ on the planktonic
776 metabolic balance during a mesocosm experiment in an Arctic fjord, *Biogeosciences*, 10,
777 315-325, 2013.

778 Taylor, A. R., Brownlee, C., and Wheeler, G. L.: Proton channels in algae: reasons to be
779 excited, *Trends Plant Sci*, 17, 675-684, 2012.

780 Telesh, I., Postel, L., Heerkloss, R., Mironova, E., and Skarlato, S.: Zooplankton of the open
781 Baltic Sea: Extended Atlas. Marine Science Reports, Leibniz institute for Baltic Sea
782 research, Warnemünde, 290, 2009.

783 Thor, P., and Oliva, E. O.: Ocean acidification elicits different energetic responses in an
784 Arctic and a boreal population of the copepod *Pseudocalanus acuspes*, *Mar Biol*, 162,
785 799-807, 2015.

786 Thornton, D. C.: Dissolved organic matter (DOM) release by phytoplankton in the
787 contemporary and future ocean, *Eur J Phycol*, 49, 20-46, 2014.

788 Tomas, C. R.: Identifying marine phytoplankton, Academic press, New York, 858 pp., 1997.

789 Utermöhl, H.: Zur Vervollkommnung der quantitativen Phytoplanktonmethodik,
790 Mitteilungen der Internationale Vereinigung für Theoretische und Angewandte
791 Limnologie, 9, 1-38, 1958.

792 Vähätalo, A. V., and Jarvinen, M.: Photochemically produced bioavailable nitrogen from
793 biologically recalcitrant dissolved organic matter stimulates production of a nitrogen-
794 limited microbial food web in the Baltic Sea, *Limnol Oceanogr*, 52, 132-143, 2007.

795 Verity, P., and Smetacek, V.: Organism life cycles, predation, and the structure of marine
796 pelagic ecosystems, *Mar Ecol Prog Ser*, 130, 277-293, 1996.

797 Viitasalo, M., Vuorinen, I., and Saesmaa, S.: Mesozooplankton dynamics in the northern
798 Baltic Sea: implications of variations in hydrography and climate, *J Plankton Res*, 17,
799 1857-1878, 1995.

800 Vincent, W. F., and Roy, S.: Solar ultraviolet-B radiation and aquatic primary production:
801 damage, protection, and recovery, *Environmen Rev*, 1, 1-12, 1993.

802 Volk, T., and Hoffert, M. I.: Ocean carbon pumps: analysis of relative strengths and
803 efficiencies in ocean-driven atmospheric CO₂ changes, in: *The carbon cycle and*
804 *atmospheric CO₂: natural variations Archean to Present*, edited by: Sundquist, E. T., and
805 Broecker, W. S., American Geophysical Union, Washington, 99-110, 1985.

806 Wu, Y., Gao, K., and Riebesell, U.: CO₂-induced seawater acidification affects physiological
807 performance of the marine diatom *Phaeodactylum tricornutum*, *Biogeosciences*, 7, 2915-
808 2923, 2010.

809 Zeebe, R. E., Zachos, J. C., Caldeira, K., and Tyrrell, T.: Carbon emissions and acidification,
810 *Science*, 321, 51-52, 2008.

811 Zeebe, R. E., and Ridgwell, A.: Past changes of ocean carbonate chemistry, in: *Ocean*
812 *acidification*, edited by: Gattuso, J.-P., and Hansson, L., Oxford University Press, Oxford,
813 1-28, 2011.

814

815

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19

Table 1. Average net primary production (NPP), total respiration (TR) and exported total particulate carbon (Exp_{TPC}) in mmol C m⁻² d⁻¹ ± SE during Phase I and Phase II of the experiment. The pool of total particulate carbon (TPC) is the average during the two periods in mmol C m⁻² ± SE. The standard error was calculated throughout the period: Phase I, n = 9 and Phase II, n = 7. NPP and TR was corrected for the missing measuring days during Phase I. TR was measured as O₂ consumption and for comparison with carbon fixation we used a respiratory quotient (RQ) of 1.

Phase I (t0-t16)

CO₂ treatment (μatm fCO₂)	346	348	494	868	1075	1333
NPP	4.8±0.8	11.4±2.1	14.9±3.6	12.3±2.3	11.3±2.4	14.5±2.7
TR	107±9	82±7	81±6	80±8	75±8	74±8
Exp _{TPC}	6.6±0.10	5.6±0.04	5.4±0.07	6.0±0.07	5.6±0.06	6.0±0.05
TPC	410±25	385±25	402±31	415±33	408±27	424±38

Phase II (t17-t31)

CO₂ treatment (μatm fCO₂)	346	348	494	868	1075	1333
NPP	3.8±0.6	11.2±1.9	10.8±2.0	14.3±2.8	10.4±2.1	12.0±2.5
TR	140±7	127±5	103±3	103±4	101±5	86±4
Exp _{TPC}	3.3±0.08	2.6±0.06	2.5±0.08	2.6±0.06	2.8±0.07	2.9±0.06
TPC	301±11	313±11	305±16	316±7	317±5	326±10

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34

Figure legends

Fig 1. The main phytoplankton groups at the start of the experiment, t_0 , and t_{17} (upper panel) and t_{31} (lower panel). The initial (t_0) was the average of all mesocosm bags. A more detailed description of the temporal development in the phytoplankton community can be found in Bermúdez et al. (2016).

Fig 2. The main micro- and mesozooplankton groups at the start of the experiment, t_0 , and t_{17} (upper panel) and t_{31} (lower panel). The initial (t_0) was the average of all mesocosm bags. A more detailed description of the temporal development in the phytoplankton community can be found in Lischka et al. (2015).

Fig 3. The development of total particulate carbon (TPC) during the experiment.

Fig. 4. The cumulative primary production in the different $f\text{CO}_2$ treatments normalized to total particulate carbon (TPC) in the euphotic zone. The $f\text{CO}_2$ (μatm) were the average measured over the duration of the experiment. The two lowest $f\text{CO}_2$ treatments (346 and 348 μatm) were controls without any CO_2 addition. The two phases of the experiment is indicated by the horizontal bars on top.

Fig. 5. The cumulative respiration in the different $f\text{CO}_2$ treatments normalized to total particulate carbon (TPC). The $f\text{CO}_2$ (μatm) were the average measured over the duration of the experiment. The two lowest $f\text{CO}_2$ treatments (346 and 348 μatm) were controls without any CO_2 addition. The two phases of the experiment is indicated by the horizontal bars on top.

Fig 6. The respiration rate, normalized to total particulate carbon (TPC), in the different $f\text{CO}_2$ treatments during the latter half of the experiment ($t_{20} - t_{31}$). Respiration was estimated by linear regression from the data presented in Fig. 4 from the time when an effect of increased CO_2 concentration was first observed. The error bars represent standard error (SE) of the residuals from the linear regression. The solid line represents the linear regression (slope - 0.0002; $p = 0.02$; $R^2 = 0.77$) and dotted lines the 95% confidence intervals.











