



1	Temperature exerted no influence on the organic carbon
2	isotope of surface soil along the isopleth of 400 mm mean
3	annual precipitation in China
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23 Abstract

Soil organic carbon is the largest pool of terrestrial ecosystem and its carbon isotope 24 composition is affected by many factors. However, the influence of environmental 25 factors, especially temperature, on soil organic carbon isotope ($\delta^{13}C_{SOM}$) is poorly 26 constrained. This impedes interpretations and application of variability of organic 27 carbon isotope in reconstructions of paleoclimate and paleoecology and global carbon 28 cycling. With a considerable temperature gradient along the 400 mm isohyet (isopleth 29 of mean annual precipitation - MAP) in China, this isohyet provides ideal 30 experimental sites for studying the influence of temperature on soil organic carbon 31 isotope. In this study, the effect of temperature on surface soil δ^{13} C was assessed by a 32 comprehensive investigation from 27 sites across a temperature gradient along the 33 isohvet. This work demonstrates that temperature did not play a role in soil δ^{13} C, this 34 suggests that organic carbon isotopes in sediments cannot be used for the 35 paleotemperature reconstruction, and that the effect of temperature on organic carbon 36 isotopes can be neglected in the reconstruction of paleoclimate and paleovegetation. 37 Multiple regression with MAT (mean annual temperature), MAP, altitude, latitude 38 and longitude as independent variables, and $\delta^{13}C_{\text{SOM}}$ as the dependent variable, shows 39 that the five environmental factors in total account for only 9% soil δ^{13} C variance. 40 However, One-way ANOVA analyses suggest that soil and vegetation types are 41 significant influential factors on soil δ^{13} C. Multiple regressions in which above five 42 environmental factors were taken as quantitative variables, vegetation type, Chinese 43 nomenclature soil type and WRB soil type were introduced as dummy variables 44 separately, show that 36.2%, 37.4%, 29.7% of the variability in soil δ^{13} C are 45 46 explained, respectively. Compared to the multiple regression in which only quantitative environmental variables were introduced, the multiple regressions in 47 48 which soil and vegetation were also introduced explain more variance, suggesting that soil type and vegetation type really exerted significant influences on $\delta^{13}C_{SOM}$. 49

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52 **1. Introduction**

53 Global climate change has recently received a great deal of attention, and effective predictions of future climate change depend on the relevant information from climate 54 in the geological past. Over recent decades, stable carbon isotopes in sediments, such 55 56 as loess, paleosol, lacustrine and marine sediments, have been widely used to reconstruct paleovegetation and paleoenvironments, and provided important insights 57 58 into patterns of past climate and environment changes. For examples, many 59 researchers have used organic carbon isotopes of loess to reconstruct paleovegetation 60 and paleoprecipitation. Vidic and Montañez (2004) conducted a reconstruction of paleovegetation at the central Chinese Loess Plateau during the Last Glaciation (LG) 61 and Holocene by using the organic carbon isotopes in loess from Jiaodao, Shanxi 62 Province. Hatt é and Guiot (2005) carried out a palaeoprecipitation reconstruction by 63 inverse modelling using the organic carbon isotopic signal of the Nußloch loess 64 sequence (Rhine Valley, Germany). Rao et al. (2013) reconstructed a high-resolution 65 summer precipitation variations in the western Chinese Loess Plateau during the Last 66 67 Glaciation using a well-dated organic carbon isotopic dataset. Yang et al. (2015) derived a minimum 300 km northwestward migration of the monsoon rain belt from 68 the Last Glacial Maximum to the Mid-Holocene using the organic carbon isotopes 69 from 21 loess sections across the Loess Plateau. However, to our knowledge, almost 70 71 no researchers have conducted paleotemperature reconstructions using organic carbon 72 isotope records of loess and paleosol, because it has been argued that temperature exerts slight, or even no influence on $\delta^{13}C_{SOM}$. While this statement may be likely, it 73





needs to be demonstrated because only few studies have addressed the influence of 74 temperature on organic carbon isotopes of modern surface soil; furthermore, these 75 studies do not appear to result in a conclusive statement. Lee et al. (2005) and Feng et 76 al. (2008) both reported no relationship between temperature and surface soil δ^{13} C in 77 78 central-east Asia. However, Lu et al. (2004) discovered a nonlinear relationship between annual mean temperature (MAT) and $\delta^{13}C_{SOM}$ from the Qinghai-Tibetan 79 80 Plateau; Sage et al. (1999) compiled the data from Bird and Pousai (1997) and also found a nonlinear trend for the variation in $\delta^{13}C_{SOM}$ along a temperature gradient in 81 Australian grasslands and savannas. 82

Plant residues are the most important source of soil organic matter. $\delta^{13}C_{SOM}$ is 83 generally close to plant carbon isotope despite isotopic fractionation during 84 decomposition of organic matter (Nadelhoffer and Fry, 1988; Balesdent et al., 1993; 85 Ågren et al., 1996; Fernandez et al., 2003; Wynn, 2007). Thus, the influential factors 86 of plants δ^{13} C might also play a role in δ^{13} C_{SOM}, δ^{13} C in plants, especially C₃ plants, is 87 tightly associated with precipitation (Diefendorf et al., 2010; Kohn, 2010), so, 88 precipitation should have influence on soil δ^{13} C. In addition to effect of precipitation, 89 many factors, such as temperature, air pressure, atmospheric CO2 concentration, 90 altitude, latitude and longitude, may also exert influences on variance in plants δ^{13} C 91 (Körner et al., 1991; Hultine and Marshall, 2000; Zhu et al., 2010; Xu et al., 2015). 92 Although patterns of variation in plants $\delta^{13}C$ with temperature are unresolved so far 93 (e.g. Schleser et al., 1999; McCarroll and Loader, 2004; Treydte et al., 2007; Wang et 94 al., 2013), it has been widely accepted that, even if temperature has effect on plants 95





 δ^{13} C, this effect is slight. So, if the ¹³C enrichment during SOM decomposition is a 96 constant value, we expect a slight or no influence of temperature on soil δ^{13} C. 97 However, the fact is that this ¹³C-enrichment is affected by environmental and biotic 98 factors (Wang et al., 2015). Thus, it is difficult to expect whether or how temperature 99 affects soil δ^{13} C, and it needs specific investigations of focusing on this issue. 100 Although the relationship between temperature and $\delta^{13}C_{SOM}$ has been investigated in 101 102 these previous studies mentioned above, these studies were unable to effectively separate the influence of temperature from the effect of precipitation. Thus, new 103 104 investigations are necessary. The present study includes an intensive investigation of the variation in $\delta^{13}C_{SOM}$ with temperature across a temperature gradient along the 400 105 mm isohyet (isopleth of mean annual precipitation - MAP) in China. We sampled 106 107 surface soil along the specific isohyet to minimize the effect of precipitation changes on $\delta^{13}C_{SOM}$. 108

In addition, there are no meteorological stations near most of the sampling sites in 109 the previous studies mentioned above; thus, they had to interpolate meteorological 110 111 data, which could be unrealistic in regions with strong topographical variability. This interpolation could produce errors in the relationships between temperature and 112 $\delta^{13}C_{SOM}$ established in these studies. In the present investigation, we collected samples 113 only at those sites with meteorological stations; thus, the climatic data that we 114 obtained from these stations are probably more reliable compared with the 115 pseudo-data derived by interpolation. 116





118 2. Materials and methods

119 2.1. Study site

120	In this study, we set up a transect along the 400 mm isohyet from LangKaZi (site 1,
121	29°3.309'N, 90°23.469'E), on the Qinghai-Tibetan Plateau in southwest China, to
122	BeiJiCun (Site27, 53°17.458'N, 122°8.752'E), in Heilongjiang Province in northeast
123	China (Fig.1, Table 1). The straight-line distance between the above two sites is about
124	6000 km. Twenty-seven (27) sampling sites were set along the transect. Among these
125	sampling sites, 10 sites are located on the Qinghai-Tibetan Plateau, and the others are
126	in north China. BeiJiCun and KuDuEr have the lowest MAT of -5.5 $^{\mathrm{o}}\mathrm{C}$ and ShenMu
127	has the highest MAT of +8.9 $^{\rm o}\text{C}.$ The average MAP of these sites is 402 mm. In north
128	China, rainfall from June to September accounts for approximately 80% of the total
129	annual precipitation, and the dominant control over the amount of precipitation is the
130	strength of the East-Asian monsoon system. In the Qinghai-Tibetan Plateau, however,
131	precipitation is associated with both the Southwest monsoon and the Qinghai-Tibetan
132	Plateau monsoon, and approximately 80% - 90% rainfall occurs in the summer season
133	(from May to October).

- 134
- 135 Fig.1
- 136 Table 1
- 137 2.2 soil sampling

Soil samples were collected in the summer of 2013 between July 12th and August30th. In order to avoid disturbance of human activities, sample sites are 5-7





140	kilometers far from the towns where the meteorological stations are located. We set
141	three quadrats (0.5 m×0.5 m) within 200 m ² to collect surface mineral soil (0-5 cm)
142	using a ring knife. The O-horizon, including litters, moders and mors were removed
143	before collecting mineral soil. About 10 g air dried soils were sieved at 2 mm. Plant
144	fragments and the soil fraction coarser than 2 mm were removed. The rest of the soil
145	samples were immersed using excessive HCl (1 mol/L) for 24 h. In order to ensure
146	that all carbonate was cleared, we conducted artificial stirring 4 times during the
147	immersion. Then, the sample was washed to neutrality using distilled water. Finally it
148	was oven-dried at 50 $^\circ\!\mathrm{C}$ and ground. Carbon isotope ratios were determined on a
149	Delta ^{Plus} XP mass spectrometer (Thermo Scientific, Bremen, Germany) coupled with
150	an elemental analyzer (FlashEA 1112; CE Instruments,Wigan, UK) in continuous
151	flow mode. The elemental analyzer combustion temperature was 1020 $^{\rm o}{\rm C}.$

The carbon isotopic ratios are reported in delta notation relative to the V-PDB standard using the equation:

154
$$\delta^{13}C = (R_{\text{sample}}/R_{\text{standard}} - 1) \times 1000 \tag{1}$$

where δ^{13} C is the carbon isotope ratio of the sample (‰), and R_{sample} and R_{standard} are the 13 C/ 12 C ratios of the sample and the standard, respectively. For this measurement, we obtained a standard deviation of less than 0.15‰ among replicate measurements of the same soil sample.

159

160 **3. Results**

161 Except for one $\delta^{13}C_{SOM}$ value (-18.8‰), all other data range from -20.4‰ to -27.1‰





162	with a mean value of -23.3‰ (n =80, s.d. =1.45). Multiple regression with MAT,
163	MAP, altitude, latitude and longitude as independent variables, and $\delta^{13}C_{SOM}as$ the
164	dependent variable, shows that only 9% of the variability in soil δ^{13} C can be explained
165	as a linear combination of all five environmental factors ($p = 0.205$) (Table 2).
166	Considering the possibility of correlations among the five explanatory variables,
167	stepwise regression was used to eliminate the potential influence of collinearity
168	among them. Variables were incorporated into the model with P -value < 0.05 and
169	exclude with <i>P</i> -value > 0.1. Stepwise regression of soil δ^{13} C in the model consisting
170	only of latitude ($R^2 = 0.077$, $p = 0.012$). In order to constrain the relationship between
171	soil $\delta^{13}C$ and each environmental factor better, bivariate correlation analyses of soil
172	$\delta^{13}C$ against some environmental factors were conducted. The bivariate correlation
173	analyses show that $\delta^{13}C_{SOM}is$ not related to MAT (p = 0.114) or SMT (p = 0.697)
174	along the isohyet (Fig. 2a, b). In addition, in order to determine further the response of
175	$\delta^{13}C_{SOM}$ to temperature, we considered three subsets of our soil samples defined
176	according to the climate, topography or vegetation type: the Qinghai-Tibetan Plateau
177	(mainly alpine meadow, including 10 sites), steppe or grassland (11 sites) and
178	coniferous forest (6 sites) (Table 1). Bivariate correlation analyses within these
179	subsets also show no relationship between $\delta^{13}C_{\text{SOM}}$ and MAT for all categories. The
180	correlation analysis of $\delta^{13}C_{SOM}vs.$ altitude is shown in Fig.3, which displays no
181	relationship ($p = 0.132$). Although longitude is not found to exert influence on
182	$\delta^{13}C_{SOM}$ in the above stepwise regression, bivariate correlation analyses show that
183	latitude and longitude both are negatively related to $\delta^{13}C_{SOM}$ (p =0.012 and 0.034,





184 respectively) (Fig. 4a,b).

185 In addition to effects of quantifiable environmental factors, qualitative factors, such as soil type and vegetation type, may have influence on $\delta^{13}C_{SOM}$. Varied concepts 186 have been introduced in soil taxonomy, leaving varied soil nomenclatures in use. In 187 this study we adopted Chinese soil nomenclature and the World Reference Base 188 (WRB) to describe the observed soil. The soil was divided into 8 types and 6 types 189 190 based on the Chinese Soil Taxonomy and WRB, respectively (Table 1). One-way 191 ANOVA analyses suggest that soil type and vegetation type both played a significant role in $\delta^{13}C_{SOM}$ (p = 0.002 for soil types based on the Chinese Soil Taxonomy, p = 192 0.003 for soil type based on WRB and p = 0.001 for vegetation types) (Fig. 5). 193

In order to constrain further the effects of soil type and vegetation type on $\delta^{13}C_{SOM}$, 194 multiple regressions with soil type and vegetation type as dummy variables were 195 196 conducted. Considering the tight relationship between soil type and vegetation type, especially in Chinese soil taxonomy, the soil variable and the vegetation variable were 197 separately introduced into the statistical analyses. Multiple regression, in which the 198 199 above five explanatory environmental factors were taken as quantitative variables and the 8 soil types of the Chinese nomenclature as values of a dummy variable, shows 200 that environmental factors and soil types in total account for 37.4% soil δ^{13} C variance 201 (p < 0.001) (Table 2). 29.7% (p = 0.003) of the variability is explained using the 6 soil 202 203 types based on WRB rather than the Chinese nomenclature (Table 2). Similarly, multiple regression with vegetation types as dummy variables shows that the five 204 environmental factors and vegetation types in total can explain 36.2% of the 205





206	variability in soil $\delta^{13}C$ (p = 0.001) (Table 2). Compared to the multiple regressions in
207	which only quantitative environmental variables were introduced, the multiple
208	regressions in which soil and vegetation were also introduced explain more variance,
209	suggesting that soil type and vegetation type really played a significant role in
210	$\delta^{13}C_{SOM}$ variability.
211	Table 2
212	Fig.2a, b

- 213 Fig.3
- 214 Fig.4a, b
- 215 Fig.5
- 216

217 **4. Discussion**

Soil δ^{13} C depends on plants δ^{13} C and carbon isotopic fractionation during organic 218 matter decomposition. δ^{13} C values of C₃ plants vary between -22‰ and -34‰ with a 219 mean of -27%, and C₄ plants range from -9% to -19% with a mean of -13%220 (Dienes, 1980). Carbon isotope fractionation occurs in the process of plant litter 221 decomposition into soil organic matter in most environments, especially in non-arid 222 environments, causing ¹³C-enrichment in soil organic matter compared with the plant 223 sources (Nadelhoffer, 1988; Balesdent et al., 1993; Ågren et al., 1996; Fernandez et 224 al., 2003; Wynn et al., 2005; Wynn, 2007). An intensive investigation of isotope 225 fractionation during organic matter decomposition, which was conducted in Mount 226 Gongga, an area in the Qinghai-Tibetan Plateau dominated by C3 vegetation with 227





228	herbs, shrubs and trees, showed that the mean 13 C-enrichment in surface soil (0-5 cm
229	depth) relative to the vegetation was 2.87‰ (Chen et al., 2009). Another
230	investigation of 13 soil profiles from the Tibetan Plateau and north China showed the
231	$\delta^{13}C$ difference between surface soil (0-5 cm depth) and the original biomass varied
232	from 0.6 to 3.5‰ with a mean of 1.8‰ (Wang et al., 2008). Thus, the $\delta^{13}C_{SOM}$ data set
233	of this study ($\delta^{13}C_{SOM}$ ranges from -20.4‰ to -27.1‰) indicates that the modern
234	terrestrial ecosystem along the isohyet is greatly dominated by C ₃ plants. This result is
235	consistent with the observations of vegetation along the isohyet done in our previous
236	study (Wang et al., 2013) and in this present study. Yin and Li (1997), Lu et al. (2004)
237	and Wang et al. (2004) have reported that a small number of C_4 species occurred in
238	the Qinghai-Tibetan Plateau; however, in this present study we found no C4 plants in
239	the Qinghai-Tibetan Plateau. We are also very surprised at such high soil $\delta^{13}C$ values
240	at RiKaZe (site 2) (Fig.3 and Table 1) because only four C_3 plants grow there, no C_4
241	species. The abnormal observation suggests that a very high carbon isotope
242	fractionation with SOM degradation have taken place in the local ecosystem.
243	Although the mechanism accounted for the unusually high isotopic fractionation
244	remains unclear, it is not surprising. For example, Wynn (2007) has reported that the
245	fractionation leaved soil organic carbon 13 C-enriched by up to ~6% with respect to
246	the original biomass. Rao et al. (2008) has suggested that mid-latitude area
247	(31 N-40 N) in east China provides relatively favorable condition for C_4 plant growth
248	But we observed that a small number of C_4 species occur only in the temperate
249	meadow steppe and the temperate typical steppe in north China, while no C ₄ species





are distributed in the coniferous forests in north China. In short, the contribution of C_4

biomass to the local vegetation along the isohyet is very low, and can be neglected.

The MAT, MAP, altitude, latitude and longitude, combined, are responsible for 252 only 9% variability in soil δ^{13} C in the multiple regression model, suggesting that the 253 contribution of the five environmental factors to the soil δ^{13} C variance is very small. 254 Our previous study conducted along the isohyet resulted in a strong positive 255 relationship between C₃ plant δ^{13} C and MAT with a coefficient of 0.104‰/°C (Wang 256 et al., 2013). The difference between maximum and minimum temperature along the 257 isohyet is 15°C, so the greatest possible effect of temperature on plant δ^{13} C along the 258 temperature gradient is 1.56‰, which is not very great. Since the main source of soil 259 organic matter along the isohyet is C₃ plants, the induced variance in soil δ^{13} C by 260 plant δ^{13} C also cannot be very great. On the other hand, although the ¹³C-enrichment 261 with SOM degradation follows a Rayleigh distillation process (Wynn, 2007), our 262 recent study shows that temperature does not influence carbon isotopic fractionation 263 during decomposition (Wang et al., 2015), which is also a reason for the lack of a 264 relationship between soil δ^{13} C and temperature. Feng et al. (2008) and Lee et al. (2005) 265 respectively, reported no relationships between soil δ^{13} C and MAT and SMT, which 266 is consistent with our result. Their field campaigns were conducted in central Asia, 267 which is also dominated by C₃ plants, similar to the area along the 400 mm isohyet. 268 269 This is the reason why the same pattern exists in central Asia and the area along the 270 400 mm isohyet.

The observations in Bird and Pousai (1997) and Sage et al. (1999) appear to be





272	inconsistent with our findings; they found a nonlinear relationship between soil $\delta^{13}C$
273	and MAT in Australian grasslands. However, if considering only soil with pure $\ensuremath{C_3}$
274	plants (MAT is below 16°C), soil $\delta^{13}C$ and temperature are not related in Australian
275	grasslands, which is in agreement with our result. Below15 $^\circ\! \mathbb{C}$, C_4 contribution to
276	productivity in Australian grasslands is negligible, whereas above 23 $^\circ\! C$, C_3
277	contribution is negligible; Between 14 $^\circ C$ and 23 $^\circ C,$ soil $\delta^{13}C$ is positively correlated
278	with MAT, indicating C_4 representation increasing with MAT (Sage et al., 1999). Lu
279	et al. (2004) also reported a nonlinear relationship between soil $\delta^{13}C$ and MAT.
280	Similarly, if the soil data with C_4 plants are excluded from the nonlinear correlation,
281	soil δ^{13} C is also not related to MAT in Lu et al. (2004) (see Fig.5b in Lu et al., 2004).
282	Thus, this present study and the previous observations are consistent in showing that
283	in a terrestrial ecosystem in which the vegetation is dominated by C_3 plants,
284	temperature does not influence soil δ^{13} C variance.

This study shows that the contribution of precipitation to the variability in soil δ^{13} C 285 is neglected. The reason for this is that the soil was sampled along the 400 mm 286 isohyet, and the MAP difference among sites is very small. It should be pointed out 287 here that the no MAP influence on the soil $\delta^{13}C$ does not mean no moisture control of 288 the soil δ^{13} C. Because the temperature varies greatly across the temperature gradient 289 290 although the MAP is almost the same for each sampling site ; this would cause a big difference in relative humidity among sites. We expect that relative humidity would 291 explain a great variability in soil δ^{13} C. But we did not take relative humidity as an 292 explanatory variable in the statistical analyses, because we lack the complete data of 293





relative humidity, and we do not want to use the pseudo-data derived by

interpolation.

Although stepwise regression and correlation analysis both show a significant 296 influence of latitude on soil δ^{13} C, the five environmental variables including latitude 297 were responsible for only 9% variability in soil δ^{13} C in a multiple regression model 298 (Table 2), suggesting that the contribution of latitude to soil δ^{13} C was also slight. This 299 study shows a negative correlation between latitude and $\delta^{13}C_{SOM}$ (p=0.012). Bird and 300 301 Pausai (1997) and Tieszen et al. (1979) reported a similar pattern. Latitude is a 302 comprehensive environmental factor, and change in latitude can bring about changes in other environmental factors, such as temperature, irradiation, cloud amount, and 303 moisture, but temperature or irradiation should be most strongly related to latitude, 304 305 and obviously change with latitude. The observed significant relationship between latitude and soil δ^{13} C (Fig.4a) suggests that environmental factors other than 306 temperature might contribute more or less to the variance in soil δ^{13} C. 307

Vegetation type control of the soil δ^{13} C mainly reflected the effects of life-form on plant δ^{13} C and substrate quality on isotope fractionation during organic matter decomposition. Communities in which life-form of the dominant plants is similar are generally treated as the same vegetation type. Plant δ^{13} C is tightly related to life-form (Diefendorf et al., 2010; Ehleringer and Cooper, 1988) and this causes δ^{13} C differences among varying vegetation types, consequently resulting in the observed effect of vegetation type on the soil δ^{13} C.

315 Substrate quality partly quantifies how easily organic carbon is used by soil





microbes (Poage and Feng, 2004). It can be related to plant type and is often defined 316 317 using a C/N ratio, lignin content, cellulose content, and/or lignin content/N ratio (Melillo et al., 1989; Gartern et al., 2000). Our study in Mount Gongga, China, 318 showed that litter quality play a significant role in isotope fractionation during organic 319 320 matter decomposition, and the carbon isotope fractionation factor, α , increases with litter quality (Wang et al., 2015). Thus, the isotope fractionation factor should be 321 322 different among varying sites because litter quality is dependent on vegetation and this makes soil change its δ^{13} C with vegetation type. 323

Control of soil type on soil δ^{13} C could be associated with the effect of soil type on 324 isotope fractionation during organic matter decomposition, and involve at least two 325 mechanisms. (1) Properties and compositions of microbial decomposer communities 326 327 are dependent on soil type (Gelsomino et al., 1999). Different microbes could have 328 different metabolic pathways even when they decompose the same organic compound (Macko and Estep, 1984), and the extent of isotope fractionation during 329 decomposition may be tightly related to the metabolic pathways of microbes (Macko 330 331 and Estep, 1984). For example, Morasch et al. (2001) observed a greater hydrogen isotope fractionation for toluene degradation in growth experiments with the aerobic 332 bacterium P. putida mt-2 and less fractionation in toluene degradation by anaerobic 333 bacteria. (2) Physical and chemical properties, such as pH, particle size fraction, 334 335 water-holding capacity, display striking differences among soil types and this causes organic compounds to be decayed at different rate in different soil environments. The 336 magnitude of isotope fractionation during decomposition is linked to degree of 337





- 338 organic matter decomposition (Feng, 2002), thus, soil type plays a significant role in
- 339 fractionation.
- 340

341 5. Conclusions

The present study measured organic carbon isotopes in surface soil along a 400 mm isohyet of mean annual precipitation in China, and observed that soil type and vegetation type both had significant influence on soil organic carbon isotopes. However, temperature is found to have no observable impact on $\delta^{13}C_{SOM}$, suggesting that $\delta^{13}C$ signals in sediments cannot be used for the reconstruction of temperature, and that the effect of temperature on $\delta^{13}C_{SOM}$ could be neglected in the reconstruction of paleoclimate and paleovegetation using carbon isotopes of soil organic matter.

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Fig.1. Sketch of sampled region. Sample sites are indicated with numbers. 1, LangKaZi; 2,
RiKaZe; 3, NaQu; 4, NieRong; 5, ZhiDuo; 6, QuMaLai; 7, TongDe; 8, TongRen; 9, HuangYuan;
10, HaiYan; 11, YuZhong; 12, XiJi; 13, JingBian; 14, HengShan; 15, ShenMu; 16, HeQu; 17,
ZhunGeErQi; 18, FengZhen; 19, DuoLun; 20, LinXi; 21, ZhaLuTeQi; 22, WuLanHaoTe; 23,
AErShan; 24, YaKeShi; 25, KuDuEr; 26, GenHe; 27, BeiJiCun. Detailed information of sites is
shown in Table 1.







541 Fig.2 shows the variance in surface soil δ^{13} C with MAT (a) and SMT (b) along the 400 mm isoline 542 in China. Circle represents alpine and subalpine; diamond indicates temperate steppe and 543 grassland;, triangle is coniferous forest.







Fig.3 shows the variance in surface soil $\delta^{13}C$ with altitude.







Fig.4 shows the relationships between the soil $\delta^{13}C$ and latitude (a) and longitude (b).







Fig.5 shows the effects of soil types and vegetation types on the soil δ^{13} C. (a.) soil types based on 596 597 Chinese soil taxonomy. 1. Matti-Gelic Cambosols; 2. Hapli- Cryic Aridosolsl; 3. Calci-Orthic Aridosols; 4. Mottlic Calci-Orthic Aridosols; 5. Typic Calci-Ustic Isohumosols; 6. Pachi-Ustic 598 599 Isohumosols; 7. Umbri-Gelic Cambosols; 8. Hapli-Ustic Argosols. (b.) soil types based on WRB. 1. Cambisols; 2. Leptosols; 3. Calcisols; 4. Chernozems; 5. Umbrisols; 6. Luvisols. (c.) vegetation 600 601 type. 1. Alpine grassland; 2. Alpine meadow; 3. Subalpine grassland; 4. Temperate coniferous and 602 broad-leaved mixed forests; 5. Temperate meadow steppe; 6. Semi-desert grasslands; 7. Temperate 603 typical steppe; 8. Frigid temperate coniferous forest. The bar in Fig.5 indicates ±1SD. 604





Table	e 1 Informa	tion of t	he samp	oling sites							
								Mean			
No.	Site name	MAT/°C	J./LWS	MAP/mm	Alt./m	Lat./N°	Lon/E°	$\delta^{13}C$ (‰)	Vegetation type	Dominate species	Soil types
1	LangKaZi	-5.3	10.8	376	4492	29.06	90.39	-23.0	Alpine grassland	Stipa 、 Festuca and Carex	Matti-Gelic Cambosols (Cambisols)
7	RiKaZe	6.3	14	420	3865	29.33	88.98	-20.4	Alpine grassland	Stipa 、 Festuca and Carex	Matti-Gelic Cambosols (Cambisols)
3	NaQu	-2.2	8.2	406	4519	31.41	91.96	-23.0	Alpine meadow	Kobresia	Matti-Gelic Cambosols (Cambisols)
4	NieRong	-1.2	6.3	400	4731	32.09	92.27	-22.9	Alpine meadow	Kobresia	Matti-Gelic Cambosols (Cambisols)
5	ZhiDuo	-0.5	٢	394	4360	33.77	95.66	-24.0	Alpine meadow	Kobresia	Matti-Gelic Cambosols (Cambisols)
9	QuMaLai	-3.3	7.4	391.7	4727	34.16	95.9	-24.0	Alpine meadow	Kobresia	Matti-Gelic Cambosols (Cambisols)
٢	TongDe	1.2	15.8	371	3258	35.27	100.64	-21.6	Subalpine grassland	Stipa and Hippolytia	Hapli- Cryic Aridosolsl (Leptosols)
8	TongRen	5.2	19.5	425.7	2467	35.55	102.03	-24.0	Subalpine grassland	Stipa and Hippolytia	Hapli- Cryic Aridosols1 (Leptosols)
6	HuangYuan	3	13.9	408.9	2725	37.02	100.8	-22.9	Subalpine grassland	Stipa and Hippolytia	Hapli- Cryic Aridosolsl (Leptosols)
10	HaiYan	1.5	9.5	400	3233	36.69	101.3	-23.1	Subalpine grassland	Stipa and Hippolytia	Hapli- Cryic Aridosolsl (Leptosols)
									Temperate coniferous and		
									broad-leaved mixed forests		
11	YuZhong	6.6	20.1	403	2361	35.78	104.05	-23.6		Pinus tabulaeformis	Hapli-Ustic Argosols (Luvisols)
12	XiJi	5.3	16.8	400	1982	36.02	105.88	-24.4	Temperate meadow steppe	Stipa and Hippolytia	Calci-Orthic Aridosols(Calcisols)
13	JingBian	7.8	20	395.4	1394	37.74	108.91	-24.2	Semi-desert grasslands	Stipa、 Hippolytia and Ajania	Calci-Orthic Aridosols(Calcisols)
14	HengShan	8.5	22.2	397	1131	38.04	109.24	-23.8	Semi-desert grasslands	Stipa、 Hippolytia and Ajania	Calci-Orthic Aridosols(Calcisols)
15	ShenMu	8.9	21.5	393	1131	38.84	110.44	-24.8	Semi-desert grasslands	Stipa、 Hippolytia and Ajania	Calci-Orthic Aridosols(Calcisols)
16	HeQu	8.8	22.3	426	912	39.33	111.19	-22.5	Temperate meadow steppe	Bothriochloa and Pennisetum	Mottlic Calci-Orthic Aridosols(Calcisols)
17	ZhunGeErQi	7.5	20	400	1236	39.87	111.18	-23.2	Temperate meadow steppe	Stipa and Aneuralepidium	Mottlic Calci-Orthic Aridosols(Calcisols)
18	FengZhen	4.7	19.4	413	1236	40.45	113.19	-22.0	Temperate typical steppe	Stipa and Aneuralepidium	Typic Calci-Ustic Isohumosols (Chernozems)
19	DuoLun	2.4	17.7	407	1245	42.18	116.47	-22.7	Temperate typical steppe	Stipa and Aneuralepidium	Typic Calci-Ustic Isohumosols (Chernozems)
20	LinXi	2.2	19.9	370	928	43.44	110.08	-23.4	Temperate typical steppe	Stipa and Aneuralepidium	Typic Calci-Ustic Isohumosols (Chernozems)
21	ZhaLuTeQi	2.8	22.4	387	332	44.61	120.97	-22.3	Temperate meadow steppe	Stipa、Aneuralepidium and	Pachi-Ustic Isohumosols (Chemozems)

Filifolium



longitude, respectively. Longitude, latitude and altitude of each site were from the portable GPS; MAT and MAP represent the average values of more than 30 years, Note: MAT, SMT, MAP, Alt, Lat and Lon are the abbreviations of mean annual temperature, summer mean temperature, mean annual precipitation, altitude, latitude,

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SMT presents the average value of June, July and August for more than 30 years. All climatic data were from the local meteorological stations and the China Meteorological Data Sharing Service System (http://cdc.cma.gov.cn/shishi/climate.jsp); The soil types are based on Chinese soil taxonomy and WRB (in the 608 609

610 Meteorological Data Sharing Service System (http://cdc.cma.gov.cn/shishi/climate.jsp); Th
611 brackets).
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624		Table 2 shows	s the results from mu	ltiple regression	ns.
	Model	\mathbb{R}^2	Adjusted R ²	ц	p-value
	1	0.091	0.030	1.484	0.205
	2	0.374	0.273	3.690	< 0.001
	ŝ	0.297	0.195	2.911	0.004
	4	0.362	0.247	3.164	0.001
625	Note: Model-1 is the mu	ultiple regression	of soil \delta ¹³ C against	MAT, MAP, al	ltitude, latitude and longitude; For Model-2, Model-3 and Model-4, i
626	addition to taking the	se five environm	iental factors as inde	pendent variab	ples, the soil types based on Chinese nomenclature and WRB, and the
627	vegetation types as du	ummy variables w	vere separately introc	luced in the mu	ultiple regressions.
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