



1 **Cell-free extracellular enzymatic activity is linked to**  
2 **seasonal temperature changes in the Baltic Sea**

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10

11 **Abstract**

12 Extracellular enzymatic activities (EEA) are a crucial step on the degradation of organic  
13 matter. Dissolved (cell-free) extracellular enzymes in seawater can make up a significant  
14 contribution of the bulk EEA. However, the factors controlling the proportion of dissolved  
15 EEA in the marine environment remain unknown. Here we studied the seasonal changes in  
16 the proportion of dissolved relative to total EEA (of alkaline phosphatase [APase],  $\beta$ -  
17 glucosidase, [BGase], and leucine aminopeptidase, [LAPase]), in the Baltic Sea for 18  
18 months. The proportion of dissolved EEA ranged between 37-100%, 0-100%, 34-100% for  
19 APase, BGase and LAPase, respectively. A consistent seasonal pattern in the proportion of  
20 dissolved EEA was found among all the studied enzymes, with values up to 100% during  
21 winter and <40% during summer. A significant negative relation was found between the  
22 proportion of dissolved EEA and temperature, indicating that temperature might be a critical  
23 factor controlling the proportion of dissolved relative to total EEA in marine environments.  
24 Our results suggest a strong decoupling of hydrolysis rates from microbial dynamics in cold  
25 waters. This implies that under cold conditions, cell-free enzymes can contribute to substrate  
26 availability at large distances from the producing cell, increasing the dissociation between the  
27 hydrolysis of organic compounds and the actual microbes producing the enzymes. This also  
28 indicates that global warming could come to affect the hydrolysis of organic matter by  
29 reducing the hydrolytic activity of cell-free enzymes.



1

2 **1 Introduction**

3 Heterotrophic prokaryotes play a central role in the marine biogeochemical cycles by  
4 transforming dissolved organic matter (DOM) into living particulate organic matter (Azam  
5 and Cho, 1987). These organisms preferentially consume high molecular weight DOM, as  
6 explained by the DOM size-reactivity model (Amon and Benner, 1996; Benner and Amon,  
7 2015). However, since only molecules <600 Da can be directly transported across the  
8 prokaryotic cell membrane (Weiss et al., 1991), heterotrophic prokaryotes need to use  
9 extracellular enzymes (EE) for hydrolyzing high molecular weight DOM into low molecular  
10 weight compounds suitable for uptake. This is why the activity of extracellular enzymatic  
11 activity (EEA) has been recognized as the rate-limiting step in organic matter degradation  
12 (Arnosti, 2011).

13 EEA in aquatic environments can be cell-associated (i.e., EEs attached to the cell wall or in  
14 the periplasmic space), or dissolved (i.e., cell-free) in the surrounding waters (Hoppe et al.,  
15 2002)). Until recently, most EEA in the marine environment was believed to be associated to  
16 cells (Hoppe, 1983; Hoppe et al., 2002) leading to the perception that only cell-associated EEs  
17 were of ecological significance (Chrost and Rai, 1993; Rego et al., 1985; Someville and  
18 Billen, 1983). However, recent reports suggested a major contribution of dissolved EEA to  
19 the total oceanic EEA pool (Baltar et al., 2010; Baltar et al., 2013; Duhamel et al., 2010). This  
20 high proportion of cell-free EEA is important because it can decouple hydrolysis rates of  
21 organic material from microbial dynamics; this is, a high proportion of dissolved EEA could  
22 indicate a greater importance of the history of the water mass than of the actual processes  
23 occurring at the time of sampling (Baltar et al., 2013; Karner and Rassoulzadegan, 1995).

24 Different potential sources of cell-free EEA include direct EE release from cells in response  
25 to appropriate substrate (Alderkamp et al., 2007), to bacterial starvation (Albertson et al.,  
26 1990) to changes in cell permeability (Chrost, 1991), to viral lysis (Karner and  
27 Rassoulzadegan, 1995) and to protist grazing (Bochdansky et al., 1995). However, little is  
28 known about what happens once the enzymes are free in the marine environment, including  
29 information about their lifetimes. The few available studies on the EE lifetime, indicate a  
30 lifetime range between tens to hundreds of hours. Surface water EE lifetimes, when incubated  
31 at in situ temperature, ranged between >1 to 9 days (Bochdansky et al., 1995; Ziervogel and  
32 Arnosti, 2008; Ziervogel et al., 2010). However, EE lifetimes of surface waters were longer



1 (up to 40 d) when incubated in the dark at 4°C than at the in situ conditions of light and  
2 temperature (Li et al., 1998). This is consistent with the only available study comparing cold  
3 deep versus warm surface waters EE lifetimes, where EE lifetimes were about one order of  
4 magnitude longer in the deep waters (Baltar et al., 2013). These results suggest that  
5 temperature could be a critical factor preserving the activity of cell-free EEA and thereby  
6 controlling the proportion of dissolved EEA in the marine environment.

7 Despite the importance and implications of the cell-free EEA in marine environments, little is  
8 known about what are the factors that control changes in the proportion of dissolved EEA. To  
9 resolve this question, a long temporal sampling strategy is needed that accounts for the long  
10 lifetime of EEs. Here we studied the temporal changes in the proportion of dissolved relative  
11 to total EEA (of glycolytic enzymes [ $\beta$ -glucosidase, BGase], a proteolytic enzyme [leucine  
12 aminopeptidase, LAPase]) and alkaline phosphatase [APase]), in a continuous biweekly  
13 sampling, with water from the Baltic Sea, for 18 months. We aimed to reveal the seasonal  
14 variability of dissolved EEA and to decipher what factors could control the proportion of  
15 dissolved relative to total EEA. Based on previous research suggesting longer EE lifetimes in  
16 cold environments, we hypothesized to find a strong link between temperature and the  
17 proportion of dissolved relative to total EEA, with lower proportions of dissolved EEA during  
18 warm periods (e.g. summer) than during cold periods (e.g. winter).

19

## 20 **2 Materials and methods**

### 21 **2.1 Study site and sampling**

22 Seawater from the Baltic Sea proper was collected twice weekly for almost 18 months, from  
23 March 22, 2012 to the August 15, 2013. Samples were taken at 2 m depth at the Linnaeus  
24 Microbial Observatory (LMO) (N 56°55.851, E 17°03.640), 10 km off the east coast of  
25 Öland, Sweden, using a Ruttner sampler. Temperature was measured on site, and the water  
26 was transported to the laboratory in acid-washed Milli-Q-rinsed polycarbonate bottles within  
27 1 h.

### 28 **2.2 Chlorophyll a concentration**

29 Chlorophyll a (Chl *a*) concentration was analyzed following extraction using ethanol  
30 (Jespersen and Christoffersen, 1987).



### 1     **2.3 Measurement of total and dissolved extracellular enzymatic activity** 2           **(EEA)**

3     The hydrolysis of the fluorogenic substrate analogues 4-methylcoumarinyl-7-amide (MCA)-  
4     L-leucine-7-amido-4-methylcoumarin, 4-methylumbelliferyl (MUF)- $\beta$ -D-glucoside and  
5     MUF-phosphate was analyzed to estimate potential activity rates of leucine aminopeptidase  
6     (LAPase),  $\beta$ -glucosidase (BGase), and alkaline phosphatase (APase), respectively (Hoppe,  
7     1983). The procedure was followed as previously described (Baltar et al., 2010; Baltar et al.,  
8     2013; Baltar et al., 2009). Briefly, EEA was determined after substrate addition and  
9     incubation using a spectrofluorometer with a microwell plate reader (FLUOstar – BMG  
10    Labtech) at excitation and emission wavelengths of 365 and 445 nm, respectively. Samples  
11    (300  $\mu$ l) were incubated in the dark at *in situ* temperature for 1.5-3 h. The increase in  
12    fluorescence over time was transformed into hydrolysis activity using a standard curve  
13    established with different concentrations of the fluorochromes MUF and MCA added to 0.2  
14     $\mu$ m filtered sample water. A final substrate concentration of 31.2  $\mu$ mol  $l^{-1}$  was used to  
15    measure BGase activities, 100  $\mu$ mol  $l^{-1}$  for APase and 500  $\mu$ mol  $l^{-1}$  for LAPase. These  
16    concentrations were previously determined as saturating substrate concentrations.

17    The total and the dissolved fraction of the EEA was distinguished as previously described  
18    (Baltar et al., 2010; Baltar et al., 2013). Briefly, raw seawater was used for total EEA;  
19    whereas for dissolved EEA, samples were gently filtered through a low protein-binding 0.2  
20     $\mu$ m Acrodisc Syringe filter (Pall) for dissolved EEA following the protocol of (Kim et al.,  
21    2007). The use of low protein-binding filters is important in this context since the adsorption  
22    of extracellular enzymes depends on the type of the filter material used for size fractionation  
23    (Obayashi and Suzuki, 2008). In the present study, dissolved (cell-free) EEA is defined as the  
24    EEA recovered in the filtrate. Total and dissolved EEA were determined on six replicate  
25    samples each.

### 26    **2.4 Statistical analyses**

27    The relations between variables were examined by means of correlation analysis computing  
28    Pearson pairwise statistics. Normality was checked with a Shapiro-Wilks test before Pearson  
29    correlations were calculated.

30



### 1 **3 Results and discussion**

2 A clear seasonal pattern in temperature was observed, with lower and relatively stable  
3 temperatures during winter (3-4°C), and strong increases during spring-summer (up to 20°C),  
4 followed by a quick temperature drop in autumn (Fig. 1A). Chlorophyll-a concentration  
5 varied between 0.4 to 4.8  $\mu\text{g l}^{-1}$ , with maximum peaks during the two types of blooms that  
6 typically occur in the Baltic (Lindh et al., 2015), the diatom and dinoflagellate spring bloom  
7 and the cyanobacterial summer bloom (Legrand et al., 2015) (Fig. 1B).

8 The total (bulk) EEA of APase, LAPase and BGase followed a similar temporal pattern (Fig.  
9 2A). APase was the EEA with the highest rates (ranging from 1.5-32  $\text{nmol l}^{-1} \text{h}^{-1}$ ), followed  
10 by LAPase (from 0.6-9.3  $\text{nmol l}^{-1} \text{h}^{-1}$ ) and BGase (from 0.1-21  $\text{nmol l}^{-1} \text{h}^{-1}$ ), indicating a  
11 potential significant P limitation in the Baltic Sea (Granéli et al., 1990; Hagström et al., 2001).  
12 The strongest peaks in BGase and LAPase co-occurred (June and August 2012, May and  
13 August 2013), together with some of the APase and the Chl-a peaks, suggesting a link  
14 between these different enzymes and the phytoplankton bloom. However, the BGase:LAPase  
15 ratio, indicative of the relative degradation of polysaccharides relative to proteinaceous  
16 material, peaked when the highest BGase and LAPase rates were observed, coinciding with  
17 Chl-a peaks. This is agreement with the results obtained in a recent seasonal study in the  
18 Adriatic Sea, where BGase prevailed over LAPase associated to phytoplankton blooms  
19 (Celussi and Del Negro, 2012). There was a tendency for higher EEA rates (and  
20 BGase:LAPase ratio) in 2013 as compared to 2012, likely explained by the reported  
21 interannual variability of phytoplankton communities linked to environmental conditions in  
22 the Baltic Sea (Kahru and Elmgren, 2014; Legrand et al., 2015).

23 The proportion of dissolved relative to total EEA ranged between 0-100%, where LAPase and  
24 APase showed a similar range (37-100% and 37-100%, respectively) and BGase showed the  
25 broadest range (0-100%) (Fig. 3). LAPase was the EEA with the lowest seasonal amplitude  
26 variability in the proportion of dissolved EEA, whereas BGase showed the largest seasonal  
27 variability. These values are within the same ranges reported in the surface coastal North Sea  
28 waters (Someville and Billen, 1983), Tokyo Bay (Hashimoto et al., 1985), Mediterranean Sea  
29 (Karner and Rassoulzadegan, 1995), Elbe estuary (Karrasch et al., 2003), Gulf of Mexico  
30 (Ziervogel and Arnosti, 2008; Ziervogel et al., 2010), North Pacific Subtropical Gyre  
31 (Duhamel et al., 2010) and the epipelagic to bathypelagic waters of the Atlantic (Baltar et al.,  
32 2010; Baltar et al., 2013).



1 The proportion of dissolved EEA showed a clear seasonal pattern in our study (despite the  
2 great variability in the bulk EEA rates), with higher values in winter and a pronounced  
3 decrease in summer. This seasonal pattern in the % of dissolved EEA was conserved among  
4 all the enzymes studied (APase, BGase and LAPase), suggesting that the main factors  
5 regulating the proportion of dissolved EEA affect all enzymes equally, irrespectively of their  
6 metabolic function. This advocates for some environmental factor rather than a biological  
7 factor controlling the proportion of dissolved EEA. The most logical factors that could be  
8 responsible for this seasonal pattern would be light and/or temperature. Inactivation of  
9 extracellular enzymes by photochemical reactions have been found in biofilm microbiota  
10 (Espeland and Wetzel, 2001). However, in laboratory experiments with Arctic surface waters,  
11 the activity of cell-free EEs (APase and LAPase) were not affected by light (under full  
12 spectrum natural sunlight), suggesting that photochemical reactions are not relevant pathway  
13 for the decay of cell-free EE in seawater (Steen and Arnosti, 2011).

14 Indeed, a significant negative correlation was found between the proportion of dissolved EEA  
15 and temperature for APase (Pearson's  $r = -0.60$ ,  $N = 22$ ,  $p=0.0035$ ), BGase (Pearson's  $r = -$   
16  $0.71$ ,  $N = 22$ ,  $p=0.0002$ ) and LAPase (Pearson's  $r = -0.67$ ,  $N = 22$ ,  $p=0.0006$ ). These results  
17 are consistent with the negative effect of temperature on EE lifetimes found in an incubation  
18 experiment with Red Sea surface water (Li et al., 1998), the extended life time of cell-free EE  
19 in Arctic waters (Steen and Arnosti, 2011), and with the order of magnitude higher EE  
20 lifetimes found in the deep as compared to the surface waters of the Atlantic (Baltar et al.,  
21 2013). This suggests that low temperature preserves better (than warm temperature) the  
22 constitutive activity of the cell-free enzymes, allowing them to remain active for longer  
23 periods. This might be linked to a reduction in the metabolism of heterotrophic microbes that  
24 would reduce the consumption/degradation rates of dissolved EEs. The higher proportion of  
25 dissolved EEA during winter suggests that the decoupling of *in situ* hydrolysis rates from  
26 actual microbial dynamics is stronger during winter (or in cold waters). Thus, these results  
27 underpin the importance of considering this stronger cold-related decoupling when relating  
28 EEA to other microbial processes. These results also suggest that, under the projected global  
29 warming scenario, it is likely that the hydrolysis of organic matter due to cell-free EE might  
30 be reduced due to a shorter lifetime of the EEs.

31



#### 1 **4 Conclusions**

2 Overall, the results of this study suggest that a relevant fraction of the total EEA measured in  
3 a particular environment can be due to free EEs, which might be a consequence of the  
4 substrate history of the water masses. Thus, advection of dissolved EEA might be a critical  
5 source of EEA at any given environment (Baltar et al., 2010; Baltar et al., 2013; Steen and  
6 Arnosti, 2011). Other factors might also allow for extended EE lifetimes during advection,  
7 like association of EEs to particles (Gianfreda and Scarfi, 1991; Naidja et al., 2000; Ziervogel  
8 et al., 2007), to exopolymeric matrix (Decho, 1990), and to detrital particle complexes  
9 (Nagata and Kirchman, 1992). Low temperature seems to be a critical factor favoring longer  
10 EE lifetimes and thereby higher proportions of dissolved relative to total EEA. This implies  
11 that under cold conditions, cell-free enzymes can contribute to substrate availability at large  
12 distances from the producing cell, potentiating the disconnection between the hydrolysis of  
13 organic compounds and the actual microbes producing the enzymes. Moreover, under warmer  
14 conditions, like those predicted to occur due to global warming, the hydrolysis of organic  
15 matter (i.e., rate limiting step in the degradation of organic matter) can be reduced due to a  
16 lower contribution of the cell-free EE hydrolysis.

17

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27



## 1 **References**

- 2 Albertson NH, Nystrom T & Kjelleberg S (1990) Macromolecular-synthesis during recovery  
3 of the marine *Vibrio* sp. S14 from starvation. *Journal of General Microbiology* **136**: 2201-  
4 2207.
- 5 Alderkamp AC, Rijssel Mv & Bolhuis H (2007) Characterization of marine bacteria and the  
6 activity of their enzyme systems involved in degradation of the algal storage glucan  
7 laminarin. *FEMS Microbiology Ecology* **59**: 108-117.
- 8 Amon RMW & Benner R (1996) Bacterial utilization of different size classes of dissolved  
9 organic matter. *Limnology and Oceanography* **41**: 41-51.
- 10 Arnosti C (2011) Microbial extracellular enzymes and the marine carbon cycle. *Annual*  
11 *Review of Marine Science* **3**: 401-425.
- 12 Azam F & Cho BC (1987) Bacterial utilization of organic matter in the sea. *Ecology of*  
13 *microbial communities*, pp. 261-281. Cambridge University Press, Cambridge.
- 14 Baltar F, Aristegui J, Gasol JM, Yokokawa T & Herndl GJ (2013) Bacterial Versus Archaeal  
15 Origin of Extracellular Enzymatic Activity in the Northeast Atlantic Deep Waters. *Microbial*  
16 *ecology* **65**: 277-288.
- 17 Baltar F, E. Sintes, H. Van Aken, J. M. Gasol, J. Aristegui & Herndl GJ (2009) Prokaryotic  
18 extracellular enzymatic activity in relation to biomass production and respiration in the meso-  
19 and bathypelagic waters of the (sub)tropical Atlantic. *Environmental Microbiology* **11**: 1998-  
20 2014.
- 21 Baltar F, Aristegui J, Gasol JM, Sintes E, Aken HMv & Herndl GJ (2010) High dissolved  
22 extracellular enzymatic activity in the deep central Atlantic Ocean. *Aquatic Microbial*  
23 *Ecology* **58**: 287-302.
- 24 Benner R & Amon RM (2015) The size-reactivity continuum of major bioelements in the  
25 ocean. *Annual review of marine science* **7**: 185-205.
- 26 Bochdansky AB, Puskaric S & Herndl GJ (1995) Influence of zooplankton grazing on free  
27 dissolved enzymes in the sea. *Marine Ecology Progress Series* **121**: 53-63.





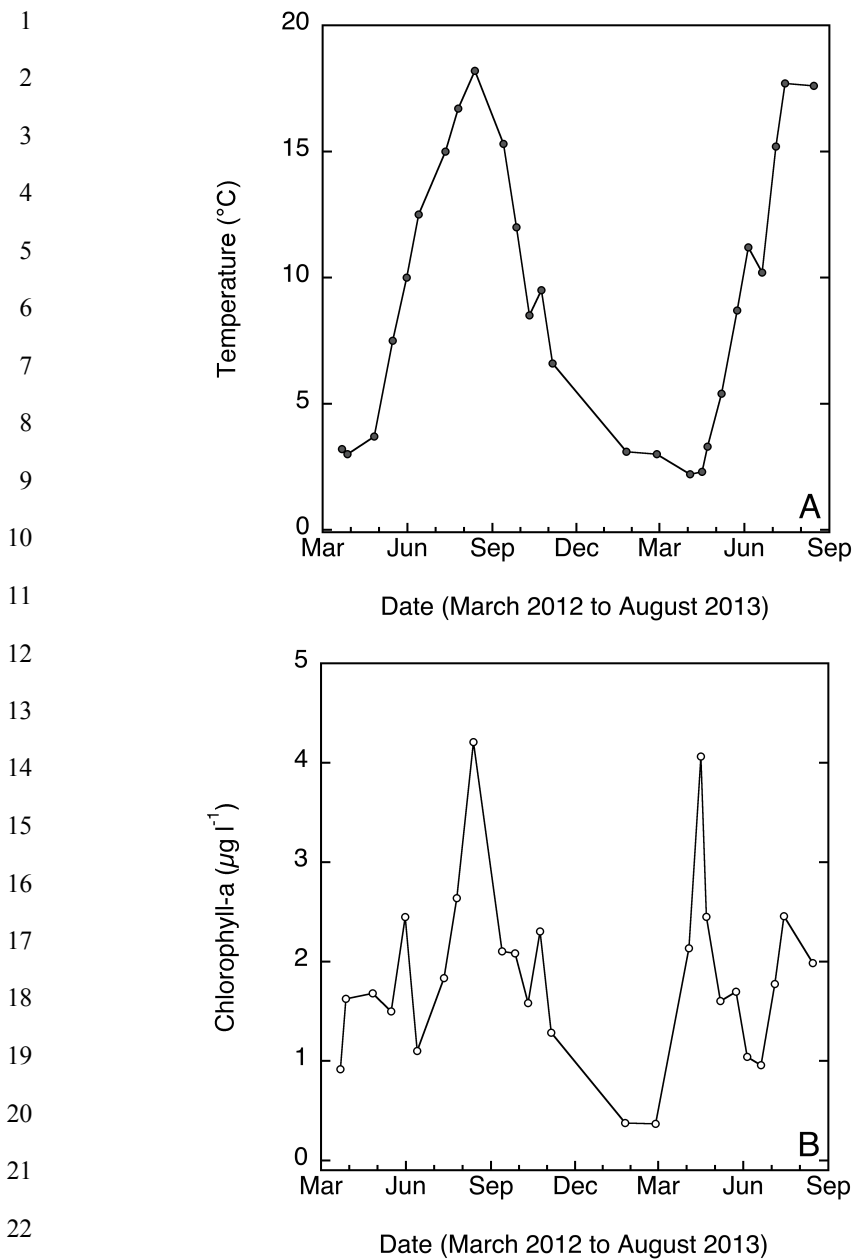
- 1 Celussi M & Del Negro P (2012) Microbial degradation at a shallow coastal site: Long-term  
2 spectra and rates of exoenzymatic activities in the NE Adriatic Sea. *Estuarine, Coastal and*  
3 *Shelf Science* **115**: 75-86.
- 4 Chrost RJ (1991) Environmental control of the synthesis and activity of aquatic microbial  
5 ectoenzymes. *Microbial enzymes in aquatic environments*, pp. 29-59. Springer Verlag, New  
6 York.
- 7 Chrost RJ & Rai H (1993) Ectoenzyme activity and bacterial secondary production in  
8 nutrient-improverished and nutrient-enriched mesocosms. *Microbial Ecology* **25**: 131-150.
- 9 Decho AW (1990) Microbial exopolymers secretions in ocean environments: their role(s) in  
10 food webs and marine processes. *Oceanogr Mar Biol Ann Rev* **28**: 73-153.
- 11 Degerholm J, Gundersen K, Bergman B & Söderbäck E (2006) Phosphorus-limited growth  
12 dynamics in two Baltic Sea cyanobacteria, *Nodularia* sp. and *Aphanizomenon* sp. *FEMS*  
13 *microbiology ecology* **58**: 323-332.
- 14 Duhamel S, Dyhrman ST & Karl DM (2010) Alkaline phosphatase activity and regulation in  
15 the North Pacific Subtropical Gyre. *Limnology and Oceanography* **55**: 1414-1425.
- 16 Espeland E & Wetzel R (2001) Complexation, stabilization, and UV photolysis of  
17 extracellular and surface-bound glucosidase and alkaline phosphatase: implications for  
18 biofilm microbiota. *Microbial ecology* **42**: 572-585.
- 19 Gianfreda L & Scarfi MR (1991) Enzyme stabilization: state of the art. *Molecular and*  
20 *Cellular Biochemistry* **100**.
- 21 Granéli E, Wallström K, Larsson U, W. Granéli & Elmgren R (1990) Nutrient limitation of  
22 primary production in Baltic Sea area. *Ambio* **19**: 142-151.
- 23 Hagström Å, Azam F, Kuparinen J & Zweifel U-L (2001) Pelagic plankton growth and  
24 resource limitations in the Baltic Sea. *A systems analysis of the Baltic Sea*, pp. 177-210.  
25 Springer-Verlag.
- 26 Hashimoto S, Fujiwara K, Fuwa K & Saino T (1985) Distribution and characteristics of  
27 carboxypeptidase activity in pond, river, and seawater in the vicinity of Tokyo. *Limnology*  
28 *and Oceanography* **30**: 631-645.



- 1 Hoppe H-G (1983) Significance of exoenzymatic activities in the ecology of brackish water:  
2 measurements by means of methylumbelliferyl-substrates. *Marine Ecology Progress Series*  
3 **11**: 299-308.
- 4 Hoppe H-G, Arnosti C & Herndl GJ (2002) Ecological significance of bacterial enzymes in  
5 the marine environment. *Enzymes in the environment: activity, ecology, and*  
6 *applications*, (Burns RG & Dick RP, eds.), pp. 73-108. Marcel Dekker, Inc., New York.
- 7 Jespersen A-M & Christoffersen K (1987) Measurements of chlorophyll—a from  
8 phytoplankton using ethanol as extraction solvent. *Archiv für Hydrobiologie* **109**: 445-454.
- 9 Kahru M & Elmgren R (2014) Multidecadal time series of satellite-detected accumulations of  
10 cyanobacteria in the Baltic Sea. *Biogeosciences* **11**: 3619-3633.
- 11 Karner M & Rassoulzadegan F (1995) Extracellular enzyme activity: indications for high  
12 short-term variability in a coastal marine ecosystem. *Microbial Ecology* **30**: 143-156.
- 13 Karrasch B, Ullrich S, Mehrens M & Zimmermann-Timm H (2003) Free and particle-  
14 associated extracellular enzyme activity and bacterial production in the lower Elbe estuary,  
15 Germany. *Acta Hydroch Hydrob* **31**: 297-306.
- 16 Kim C, Nishimura Y & Nagata T (2007) High potential activity of alkaline phosphatase in the  
17 benthic nepheloid layer of a large mesotrophic lake: implications for phosphorus regeneration  
18 in oxygenated hypolimnion. *Aquatic Microbial Ecology* **49**: 303-311.
- 19 Legrand C, Fridolfsson E, Bertos-Fortis M, Lindehoff E, Larsson P, Pinhassi J & Andersson  
20 A (2015) Interannual variability of phyto-bacterioplankton biomass and production in coastal  
21 and offshore waters of the Baltic Sea. *Ambio* **44**: 427-438.
- 22 Li H, Veldhuis MJW & Post AF (1998) Alkaline phosphatase activities among planktonic  
23 communities in the northern Red Sea. *Marine Ecology Progress Series* **173**: 107-115.
- 24 Lindh MV, Sjöstedt J, Andersson AF, Baltar F, Hugerth LW, Lundin D, Muthusamy S,  
25 Legrand C & Pinhassi J (2015) Disentangling seasonal bacterioplankton population dynamics  
26 by high-frequency sampling. *Environmental microbiology* **17**: 2459-2476.
- 27 Nagata T & Kirchman DL (1990) Filtration-induced release of dissolved free amino acids:  
28 application to cultures of marine protozoa. *Mar Ecol Prog Ser* **68**: 1-5.
- 29 Nagata T & Kirchman DL (1992) Release of macromolecular organic complexes by  
30 heterotrophic marine flagellates. *Marine Ecology Progress Series* **83**: 233-240.



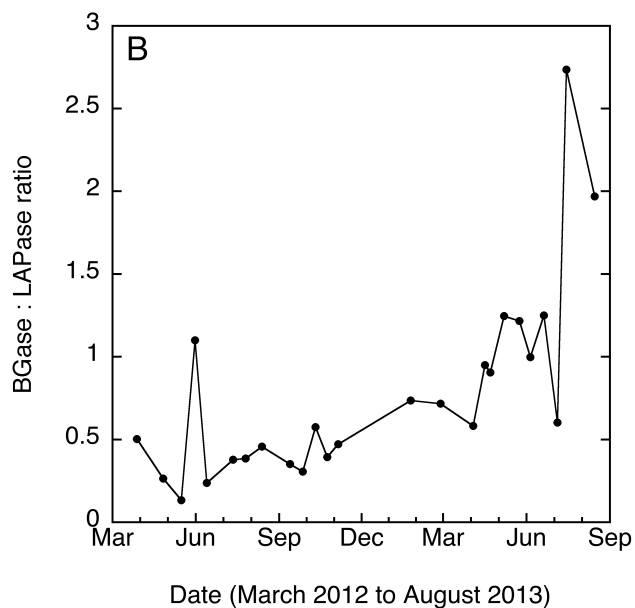
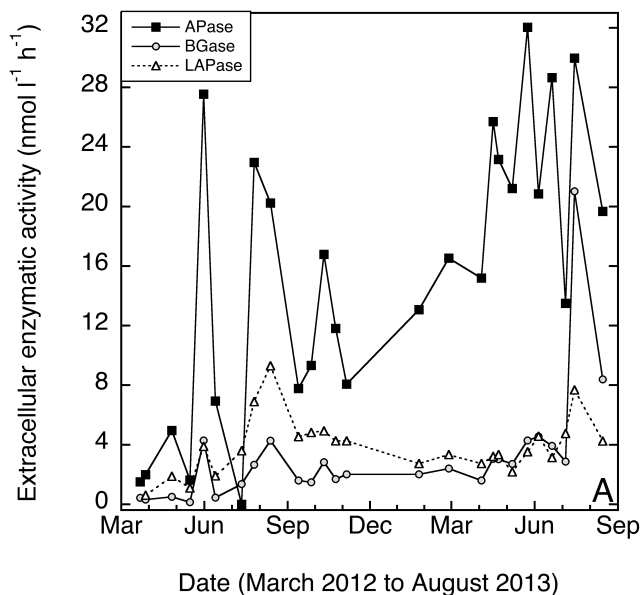
- 1 Naidja A, Huang PM & Bollag JM (2000) Enzyme-clay interactions and their impact on  
2 transformations of natural and anthropogenic organic compounds in soil. *Journal of*  
3 *Environmental Quality* **29**: 677-691.
- 4 Obayashi Y & Suzuki S (2008) Adsorption of extracellular proteases in seawater onto filters  
5 during size fractionation. *Journal of Oceanography* **64**: 367-372.
- 6 Rego JV, Billen G, Fontigny A & Someville M (1985) Free and attached proteolytic activity  
7 in water environments. *Marine Ecology Progress Series* **21**: 245-249.
- 8 Someville M & Billen G (1983) A method for determining exoproteolytic activity in natural  
9 waters. *Limnology and Oceanography* **28**: 190-193.
- 10 Steen AD & Arnosti C (2011) Long lifetimes of beta-glucosidase, leucine aminopeptidase,  
11 and phosphatase in Arctic seawater. *Mar Chem* **123**: 127-132.
- 12 Weiss M, Abele U, Weckesser J, Welte W, Schiltz E & Schulz GE (1991) Molecular  
13 architecture and electrostatic properties of bacterial porin. *Science* **254**: 1627-1630.
- 14 Ziervogel K & Arnosti C (2008) Polysaccharide hydrolysis in aggregates and free enzyme  
15 activity in aggregate-free seawater from the north-eastern Gulf of Mexico. *Environmental*  
16 *Microbiology* **10**: 289-299.
- 17 Ziervogel K, Karlsson E & Arnosti C (2007) Surface associations of enzymes and of organic  
18 matter: consequences for hydrolytic activity and organic matter remineralization in marine  
19 systems. *Marine Chemistry* **104**: 241-252.
- 20 Ziervogel K, Steen AD & Arnosti C (2010) Changes in the spectrum and rates of extracellular  
21 enzyme activities in seawater following aggregate formation. *Biogeosciences* **7**: 1007-1015.
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24 Figure 1. Temporal dynamics in (A) temperature and (B) chlorophyll-a concentrations, at the  
25 Linnaeus Microbial Observatory (LMO, Baltic Sea) from March 2012 to August 2013



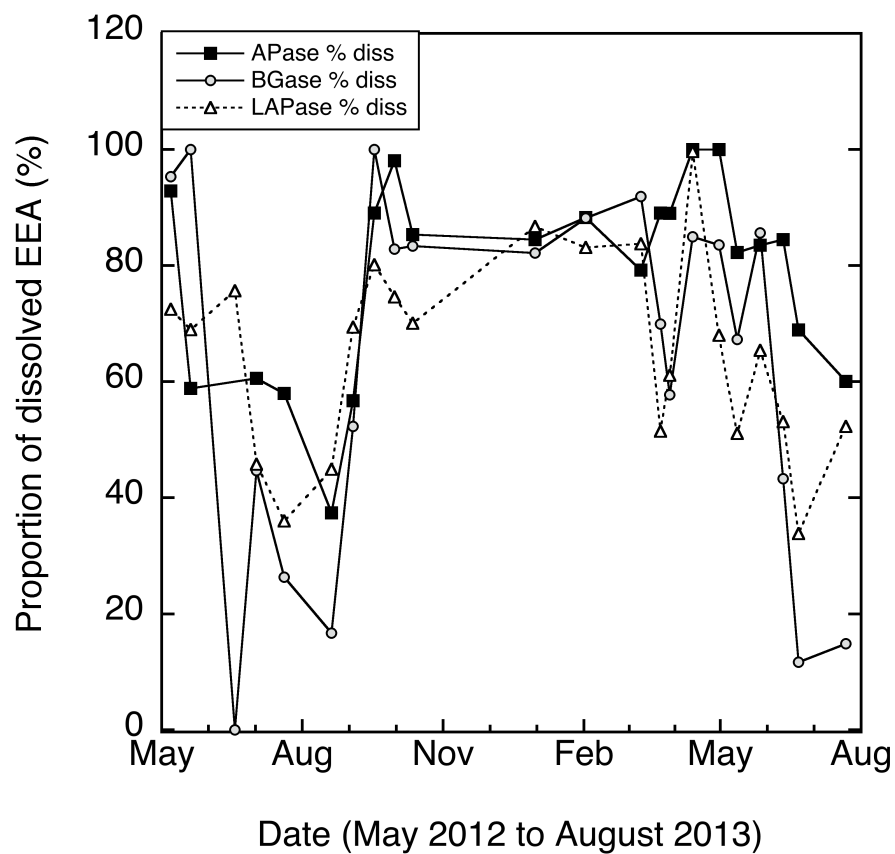
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23 Figure 2. Temporal dynamics in (A) extracellular enzymatic activities of alkaline phosphatase  
24 (APase),  $\beta$ -glucosidase, (BGase), and leucine aminopeptidase, (LAPase) and), and (B) the  
25 BGase:LAPase ratio, at the Linnaeus Microbial Observatory (LMO, Baltic Sea) from March  
26 2012 to August 2013.,



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3

4 Figure 3. Temporal dynamics in the proportion of dissolved relative to total extracellular  
 5 enzymatic activities (EEA) of alkaline phosphatase (APase),  $\beta$ -glucosidase, (BGase), and  
 6 leucine aminopeptidase, (LAPase), at the Linnaeus Microbial Observatory (LMO, Baltic Sea)  
 7 from May 2012 to August 2013.

8