



1 Sediment characteristics as an important factor for revealing carbon
2 storage in *Zostera marina* meadows: a comparison of four European
3 areas

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19 **Abstract.** The seagrass ecosystem is an important natural carbon sink but the efficiency varies greatly depending on
20 species composition and environmental conditions. What causes this variation is not fully known and could have
21 important implications for management and protection of the seagrass habitat to continue to act as a natural carbon
22 sink. Here, we assessed sedimentary organic carbon in *Zostera marina* meadows (and adjacent unvegetated sediment)
23 in four areas of Europe (Gullmar Fjord on the Swedish west coast, Askö in the Baltic Sea, Sozopol in Black Sea and Ria
24 Formosa in southern Portugal) down to ~35 cm depth. We also tested how sedimentary organic carbon in *Z. marina*
25 meadows relates to different sediment characteristics, a range of seagrass-associated variables and water depth. The
26 carbon storage varied both among and within areas, where the Gullmar Fjord had a 15 times higher carbon storage
27 compared to Askö and Sozopol. We found that high carbon content in *Z. marina* sediment is strongly related to a high
28 proportion of fine grain size, high porosity and low density of the sediment. We suggest that sediment characteristics
29 should be highlighted as an important factor when evaluating high priority areas in management of *Z. marina* generated
30 carbon sinks.

31

32 **Keywords:** Carbon storage variability, *Zostera marina*, grain size, sediment characteristics, natural carbon sinks.

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34 **1. Introduction**

35

36 Seagrass ecosystems are considered highly efficient natural carbon sinks (Mcleod et al., 2011) but there is a large
37 variation in the capacity to store carbon, depending on species composition and habitat characteristics (Lavery et al.,
38 2013; Rozaimi et al., 2013). While the carbon sequestration efficiency is quite well documented for many seagrass
39 species (e.g. Kennedy *et al.*, 2010; Fourqurean *et al.*, 2012) the effects of different factors influencing intraspecific
40 variation has only recently been investigated. To get a more accurate estimate of the global seagrass carbon sink
41 capacity cause-effect relationships need to be better understood, and as seagrass loss is accelerating (Waycott *et al.*,
42 2009) information on habitat characteristics affecting carbon storage are of importance for an efficient protection and
43 management strategy to increase carbon storage capacity (Duarte et al., 2011).

44

45 There are several environmental factors (e.g. water depth and hydrodynamic processes) and seagrass habitat variables
46 (e.g. canopy height and shoot density) that influence the carbon storage in seagrass sediments (Samper-Villarreal et al.,
47 2016). For example, seagrass meadows at shallower depths are known to have a high accumulation of sedimentary
48 carbon, which could be associated with higher primary production and larger standing biomass stock (Serrano et al.,
49 2014). Dense meadows have the ability to stabilize the sediment (and thereby preventing it from eroding) (Suykerbuyk
50 et al., 2015) and seagrass habitats with a high canopy can trap a high amount of suspended particles and thus
51 potentially increase the sedimentation of organic matter (Fonseca and Cahalan, 1992; Hendriks et al., 2008). Further,
52 as the belowground biomass largely contributes to the carbon storage due to its high production, fast turnover and
53 higher decay-resistant lignin content compared to the leaves (Duarte et al., 1998; Klap et al., 2000) a large root-rhizome
54 system could render a higher carbon storage (Kenworthy and Thayer, 1984). In the coastal environment, sediment
55 grain size is known to influence the aggregation of organic particles with finer grain sizes increasing the organic matter
56 content of the sediment (Mayer, 1994b). By reducing water velocity and facilitating sedimentation processes a seagrass
57 meadow could increase the amount of fine particles, which thus promote a high carbon storage. Grain size has, however,
58 proven a poor overall predictor of sedimentary carbon content in seagrass habitats, except for in smaller species such



59 as *Halodule uninervis*, *Zostera muelleri* and *Halophila* spp. (Serrano *et al.*, in review). Grain size is also strongly related
60 to sediment porosity and density, which influence the oxygen conditions in the sediment. Oxygen levels together with
61 the microbial community composition, biomass carbon and nutrient content are important factors for the degradation
62 rate of organic matter in the sediment (Benner *et al.*, 1984; Deming and Harass, 1993; Enriquez *et al.*, 1993) and
63 therefore influencing the carbon sequestration process.

64

65 *Zostera marina* L. is the most widely spread seagrass species in the northern hemisphere, with a distribution in Europe
66 stretching from the southern Black Sea and the gulf of Cádiz (Southern Portugal) up to Iceland and the northern parts
67 of Norway (Green and Short, 2003). The plant biomass is generally larger at higher latitudes (Short *et al.*, 2007) because
68 of more optimal growth temperatures (Moore and Short, 2003). Large seagrass populations can be found along the
69 Swedish west coast and at the east coast of Denmark (Baden and Boström, 2001; Olesen and Sand-Jensen, 1994), where
70 they form extensive meadows with shoots over 1 m in length. Due to its wide distribution *Z. marina* populations have
71 adapted to a large range of environmental conditions, with potential differences in carbon storage capacity. The species
72 can tolerate salinity ranging from 5 to 35 (Boström *et al.*, 2003) and a depth distribution from the intertidal down to
73 30 m depending on water clarity (Phillips and Meñez, 1988). *Zostera marina* also grows in various substrates, from
74 courser stone-sand bottoms to finer silt and clay sediment. In this study, we aim to assess and compare carbon storage
75 in *Z. marina* meadows at four different areas in Europe as well as to examine relationships between sediment organic
76 carbon content and several explanatory predictors including seagrass structural complexity, carbon and nitrogen
77 content of the seagrass biomass and sediment characteristics (i.e. sediment porosity and density, and grain size) in
78 order to determine factors influencing the storage capacity of *Z. marina* meadows in these areas.

79



80 **2. Methods**

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82 **2.1 Study sites**

83

84 This study was conducted in four different areas in Europe (the Swedish Skagerrak and Baltic coasts, Black Sea in
85 Bulgaria and the southern coast of Portugal; Table 1 and Fig. 1) from June to October 2013 with one complimentary
86 field sampling performed in October 2014. Our areas roughly correspond to the edge zones of the *Z. marina* distribution
87 in Europe. In each area, two meadows and one unvegetated area (reference site) were sampled, except for Portugal
88 with one additional unvegetated area and the Baltic Sea where one meadow and one unvegetated area were added
89 (Table 1). The sampling on the Swedish west coast were carried out off the Sven Lovén Centre for Marine Sciences –
90 Kristineberg in the Gullmar Fjord (58°20'N, 11°33'E; Table 1). The area is comprised of small islands and shallow bays
91 making it highly productive and a suitable environment for seagrass growth with many sheltered soft bottoms covered
92 by extended *Z. marina* beds. In the Baltic Sea, samples were collected in the area around the Askö Laboratory in
93 Stockholm Archipelago (58°49'N, 17°39'E). The Baltic Sea is a brackish water system and the salinity is about 5-6
94 outside Askö, which is on the distribution limit for *Z. marina* (Boström et al., 2003). Low salinity is known to negatively
95 affect production and growth of the plant (Salo et al., 2014). In the Baltic Sea, *Z. marina* grows at approximately 2-5 m
96 depths (sometimes together with *Ruppia maritima*) and on more coarse sediment compared to the Skagerrak area
97 (Baden and Boström, 2001). In the Black Sea, sampling was carried out in two sites around the Laboratory of Marine
98 Ecology in Sozopol, Bulgaria (42°25'N, 27°41'E). The salinity is around 17 and commonly *Z. marina* grows in mixed
99 stands with *Z. noltii*. The Ropotamo (Rt) site is situated in the vicinity of the Ropotamo river mouth. Ria Formosa
100 (Algarve Marine Sciences Centre – Faro) is located in southern Portugal (36°59'N, 7°52'W) and is a coastal lagoon with
101 large intertidal areas and a tidal fluctuation of 2-3 m. This is the only area in the present study with pronounced tidal
102 variation, and the water depth for the Portugal sites was standardized to mean low water (MLW) by calculating the
103 difference between the measured water depth and the tide at the time of measurement. The tide values were obtained
104 from Ria Formosa tidal station (Faro-Olhão) with the mean water level as reference depth. Ria Formosa is a lagoon



105 with scarce *Z. marina* distribution (which at times grows together with *Cymodocea nodosa*) and apart from one other
106 area in Portugal (Óbidos Lagoon) the only one that still harbor *Z. marina*, which has decreased drastically during the
107 past 20 years (Cunha et al., 2013).

108 109 **2.2 Sediment sampling and biometrical measurements**

110
111 At each site, six sediment cores were taken with a push corer (h=50 cm, \varnothing =8 cm) at a distance of 10-30 m apart from
112 each other. Each core was sliced into a maximum of six segments (0-2.5 cm, 2.5-5 cm, 5-12.5 cm, 12.5-25 cm, 25-37.5
113 cm, 37.5-45 cm) with the majority of samples lacking the deepest segment. Within a few meters from each core at the
114 seagrass sites, shoot height (cm, n=20) was measured, percentage seagrass coverage (n=10) were estimated (in 0.5 x
115 0.5 m squares) and biomass samples (n=3) were collected (0.25 x 0.25 m). The biomass samples were used for
116 estimating above- and belowground seagrass biomass (as dry weight) and for counting number of shoots. Before
117 weighing the seagrass was cleaned and epiphytes removed, and the dry weight was measured after 24-48 h in 60°C
118 until constant weight. One out of the three biomass samples collected around each core were analyzed for carbon and
119 nitrogen content (n = 6 for each meadow). The sediment samples were cleaned from roots and rhizomes, larger shells
120 and benthic organisms prior of drying and dried in the same way as the biomass. The sediment was divided into two
121 subsamples, one for analysis of carbon and nitrogen content, and the other for grain size analysis. The carbon and
122 nitrogen contents in biomass and sediment were analyzed using an organic elemental analyzer (Flash 2000, Thermo
123 Fischer scientific). The sediment samples were pre-treated with 1 M HCl (direct addition) to remove inorganic carbon
124 prior to analysis for C_{org} content. Total nitrogen (N_T) was measured due to possible alteration of the nitrogen values
125 when treated with HCl (Harris et al., 2001). Sediment porosity was given as percentage (%) by calculating sediment
126 wet weight minus dry weight divided by the sample volume, whereas sediment density ($g\ DW\ mL^{-1}$) was derived from
127 dividing the dry weight of the sediment by the volume of the sample.

128 129 **2.3 Grain size analysis**



130

131 Three sediment cores in each habitat were used for particle size analysis and each depth section was separately
132 analyzed. Prior to analysis the total dry weight of sediment for each section was determined and 100 ml of 0.05 M
133 $\text{Na}_4\text{P}_2\text{O}_7$ was added to break down aggregates of clay particles. Each depth section was dry-sieved for 10 min using a
134 sieving tower (CISA electromagnetic sieve shaker, Spain) (including sieves of 0.074 mm, 0.125 mm, 0.25 mm, 0.5 mm,
135 1 mm and 2 mm) and the sediment of each sieve was weighed to determine the weight of the separate fractions. In
136 depth sections with high organic carbon content (>0.5%), the organic matter was removed prior to dry sieving, through
137 oxidation with 35% H_2O_2 , as the organic matter content leads to aggregation of particles (Gee and Bauder, 1986). After
138 the reaction with H_2O_2 had ceased the samples were centrifuged and washed in distilled water to remove the H_2O_2 .
139 Some of the samples from the Skagerrak and Ria Formosa regions had a high proportion of finer fractions (>15% was
140 assessed as %<0.074 mm) and had to be analysed with hydrometer for an accurate estimate of total grain size. The
141 samples were once more treated with 0.05 M $\text{Na}_4\text{P}_2\text{O}_7$ and placed in a 1L cylinder containing distilled water and kept
142 in suspension. At fixed time intervals (1, 2, 4, 10, 20, 50, 100, 200, 400 and 1000 min) the hydrometer was inserted and
143 the concentration of sediment (g L^{-1}) was noted. The mean grain size was presented in phi (ϕ) units.

144

145 2.4 Statistical analysis

146

147 All cores were standardized to a depth of 25 cm for the sediment characteristics (porosity, density, grain size and
148 organic carbon content) prior to statistical analysis. Some of the cores at Askö (both seagrass- and unvegetated sites)
149 lacked the 12.5-25 cm depth segment and in these cases logarithmic regressions were used to extrapolate the data
150 down to 25 cm depth. All data were checked for normal distribution using the Shapiro-Wilk normality test and
151 homogeneity of variances using Levene's test. When assumptions were not met the data was \log_{10} or $\log_{10}(x+1)$
152 transformed. To test differences in sedimentary carbon storage among areas (with site nested in area), among sites
153 (within each area separately) and among sediment depths (also within each area separately), one-way ANOVA was
154 used. For the analysis across areas and among sites the carbon content was analyzed as a mean (% C) or amount of



155 carbon per unit area (g C m^{-2}) for the top 25 cm of sediment. In those cases, where the ANOVA models were significant,
156 Tukey's HSD post hoc test was used to determine significant differences between specific areas, sites and sediment
157 depths, respectively. Partial Least Square (PLS) regression technique (by modeling of projections of latent structures;
158 Wold et al., 2001) and Principal Component Analysis (PCA) were used to test the influence of sediment characteristics,
159 water depth and seagrass-related variables on sediment carbon content (mean % C for the top 25 cm of sediment). The
160 advantage of using PLS model is that it can handle collinear explanatory data as well as a large number of predictors.

161



162 3. Results

163

164 3.1 Variation in sedimentary carbon storage

165

166 All areas, except Sozopol and Askö, were significantly different from each other in terms of % C_{org} ($P < 0.001$) and g C
167 cm^{-2} ($P < 0.05$; Fig. 2). The highest amount of sedimentary carbon was seen in the Gullmar Fjord (on average \pm SE, 2.79
168 ± 0.50 % C_{org} at 0-25 cm), followed by Ria Formosa (0.61 ± 0.09), Askö (0.18 ± 0.01) and Sozopol (0.17 ± 0.02). There
169 was a within-area variation in carbon storage of the seagrass sites in Gullmar Fjord, Sozopol and Askö, while the sites
170 at Ria Formosa did not differ from each other (Fig. 2; Table 2). There were clear within-area differences in both % C_{org}
171 and g C cm^{-2} between seagrass sites in the Gullmar Fjord as well as between sites in Sozopol (Fig. 2, Table 2). At Askö
172 the seagrass site Torö (T) differed from Storsand (S) and Långskär (L) in g C cm^{-2} but only to Storsand (S) in % C_{org} ,
173 while Storsand (S) and Långskär (L) were not significantly different from each other in any regards (Table 2). Both sites
174 in the Gullmar Fjord and at Ria Formosa showed significantly higher carbon storage (% C_{org} and g C cm^{-2}) compared to
175 their respective unvegetated areas ($P < 0.05$; Table 2). This was also seen in both sites at Sozopol in terms of % C_{org} but
176 Gradina (G) did not differ in g C cm^{-2} compared to Bay of Sozopol (SB, unvegetated area; Table 2). All sites at Askö
177 differed to one of the two unvegetated areas, Godahoppssudden (Gh), but not to Torö (Tr, unvegetated area), which was
178 significantly higher in gC cm^{-2} compared to Långskär (L) and Storsand (S).

179

180 The percentage sedimentary carbon decreased with depth in Ria Formosa ($P < 0.001$) and Askö ($P < 0.05$) while not in
181 the Gullmar Fjord and Sozopol (Fig. 3). For unvegetated areas the organic carbon decreased with sediment depth in Ria
182 Formosa and Askö ($P < 0.001$), while it increased in the Gullmar Fjord ($P < 0.05$) and was unaffected in Sozopol.

183

184 3.2 Influence of sediment characteristics and seagrass-associated variables on carbon storage

185



186 When the relationship between % C_{org} and explanatory variables (Tables 2, 3 and 4) was examined in a PLS (Partial
187 least square) regression model the sediment characteristics explained most of the model (with a variance of importance
188 value >1) where the proportion of sediment particles <0.074 mm (%) was the most important, followed by sediment
189 porosity (%), sediment density (g DW mL⁻¹) and mean grain size (ϕ) (Fig. 4). These variables characterizing the
190 sediment were all positively correlated to % C_{org} except sediment density that showed a negative relationship with %
191 C_{org} . The cumulative fraction explaining the % C_{org} variation (R_y^2 cum) of the predictor variables combined was 0.81
192 and the models cross-validated variance (Q^2 statistics) showed high predictability with Q^2 -value of 0.79, thus larger
193 than the significant level of 0.05. The results of the model with g C cm⁻² (not shown here) as response variable were
194 highly similar to the results of % C_{org} ($Q^2 = 0.77$, R_y^2 cum = 0.78) with the same predictor variables (i.e. sediment
195 characteristics) explaining most of the % C_{org} variation and correlated in the same way. All seagrass-associated
196 variables showed a positive relationship with % C_{org} except for belowground (Bg) biomass N (%), which was the least
197 influential variable in the model. In general, the seagrass-associated variables showed a lower contribution to the
198 overall model compared to the sediment characteristics. Water depth (m) was also negatively correlated to % C_{org} but
199 was, as with the seagrass-associated variables, of minor importance (Fig. 4).

200

201 Mean grain size (ϕ) and sediment particles < 0.074 mm (%) both showed strong positive linear relationship with %
202 C_{org} in *Z. marina* meadows (mean phi (ϕ), $R^2 = 0.74$, $P < 0.001$; sediment particles < 0.074 mm (%), $R^2 = 0.91$, $P < 0.001$;
203 Fig. 5a and b). For unvegetated areas, mean grain size (ϕ) did not show any relationship with % C_{org} but was positively
204 related to sediment particles <0.074 mm (%) (linear regression, $R^2 = 0.42$, $P < 0.001$; Fig. 5c and d). The sediment
205 density (g DW mL⁻¹) had a negative effect on % C_{org} in the seagrass sites (linear regression, $R^2 = 0.84$, $P < 0.001$) and
206 sediment porosity (%) was positively related to % C_{org} (linear regression, $R^2 = 0.80$, $P < 0.001$; Fig. 6a and b). There was
207 no significant relationship between % C_{org} and sediment density (g DW mL⁻¹) in unvegetated areas while sediment
208 porosity (%) was significantly influencing % C_{org} but showed a low R^2 -value (linear regression, $R^2 = 0.08$, $P < 0.001$; Fig.
209 S2).

210



211 The sedimentary organic carbon content relationship to the different predictor variables was not uniform among sites.
212 In a PCA model, the Gullmar Fjord and Ria Formosa were grouped separately from other sites, while the Baltic- and
213 Black Seas sites overlapped each other (Fig. 7). For the fine grain size seagrass sites of the Gullmar Fjord (Table 4), the
214 sediment characteristics (i.e. sediment particles <0.074 mm (%), sediment porosity (%) and mean grain size (ϕ)) were
215 important for the carbon content while the sedimentary carbon in Ria Formosa was more related to seagrass cover (%)
216 and dry weight belowground biomass (m^2). The sedimentary organic carbon content in seagrass sites in Baltic- and
217 Black Seas were also more related to the seagrass-associated variables, such as dry weight aboveground biomass (m^2)
218 and shoot density (m^2), but also water depth (m) for one of the sites (Storsand, S).

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220



221 **4. Discussion**

222

223 In this assessment of four *Z. marina* areas in Europe, we found a large variation in organic carbon storage where the
224 carbon-rich sediment of the Gullmar Fjord on the Swedish west coast where 15 times higher compared to levels in the
225 Baltic- and Black Seas. The mean carbon content of the Gullmar Fjord was also higher than estimated global averages
226 (Fourqurean et al., 2012; Kennedy et al., 2010), demonstrating the high carbon capacity of the area. Along with recent
227 studies (Lavery et al., 2013; Samper-Villarreal et al., 2016), this study shows that the environmental conditions play an
228 essential role in determining the carbon sink capacity. Here we demonstrate that sediment characteristics influence
229 carbon storage in *Z. marina* meadows, where fine grain size, high sediment porosity and low sediment density
230 correspond with high sedimentary organic carbon content. Seagrass meadows situated in areas characterized by these
231 sediment properties are therefore suggested to have a high potential as natural carbon sinks.

232

233 Variation in sedimentary carbon content was also seen on a local scale where most of the sites differed from each other
234 (except for Ria Formosa and two of the sites at Askö) demonstrating that the local variability in carbon storage is as
235 high as the regional-based variation. Most of the sites, with the exception of the seagrass meadows with the lowest
236 carbon storage, showed higher carbon content than nearby unvegetated areas, which illustrates just as previous studies
237 have shown (e.g. Kennedy *et al.* 2010; Mcleod *et al.* 2011), that the seagrass ecosystem is a significant carbon sink. In
238 general, carbon content in seagrass meadows decreases logarithmically with sediment depth (Fourqurean et al., 2012)
239 due to degradation and remineralization of organic material with time (Burdige, 2007; Henrichs, 1992), a pattern seen
240 in two of the study areas, Ria Formosa and Askö. The Gullmar Fjord and Sozopol did not show decreasing carbon
241 content with sediment depth, which could be due to a slower degradation of organic matter in the sediment. Finer grain
242 sizes in combination with high organic matter and nutrient content, as seen in the Gullmar Fjord sites, could cause a
243 depletion of oxygen in the sediment because of increased oxygen consumption by detritus organisms and decreased
244 permeability (Pollard and Moriarty, 1991; Wilson et al., 2008), which slows down the degradation process of organic
245 matter. Sozopol, however, does not show such sediment characteristics but the stable carbon content with depth might



246 be because the Ropotamo site was situated close to a river mouth with an input of terrestrial organic matter with a
247 higher lignin content resulting in more decay-resistant carbon and a slower decomposition (Cowie and Hedges, 1984;
248 Ertel and Hedges, 1985).

249
250 A high carbon content in *Zostera marina* sediment seems to be related to the sediment characteristics of the area. A
251 high proportion of finer grain size particles leads to preservation and accumulation of organic matter (Keil et al., 1994;
252 Mayer, 1994a, 1994b), which also explains the strong correlation between finer grain size content and sedimentary
253 carbon seen in our study. Sediment grain size has recently been described as a strong predictor for carbon storage in
254 another blue carbon habitat, i.e. saltmarshes (Kelleway et al., 2016). For seagrass meadows, sediment grain size has
255 shown to have little influence on sedimentary carbon content (Samper-Villarreal *et al.* 2016), except for sediment with
256 a high proportion of finer particle sizes (Serrano *et al.* in review). Also in the present study the proportion of fine
257 sediment particles was strongly coupled to a high carbon content. The grain size is directly linked to the sediment
258 porosity and density where the organic carbon has a negative effect on sediment density (Avnimelech et al., 2001;
259 Gullström et al. submitted). This was also seen in our study as higher sedimentary carbon values were found in areas
260 with lower sediment density (and hence higher porosity). For these reasons, we suggest that sediment characteristics
261 of the area where *Z. marina* meadows are situated is relevant for revealing the carbon storage potential. For example,
262 the *Z. marina* meadows at Askö are usually found on coarser sand and in more exposed areas whereas the most
263 sheltered bays with finer grain sizes are dominated by brackish water plants, such as *Potamogeton pectinatus* and
264 *Zannichellia palustris* (Idestam-Almquist, 2000), which might explain the low carbon storage potential of the area. A
265 high organic content in the sediment could, however, cause a depletion of oxygen (Holmer, 1999) and at too low oxygen
266 levels seagrass can no longer maintain the aerobic conditions of the rhizosphere, eventually leading to seagrass
267 mortality (Terrados et al., 1999) with consequences for the carbon storage capacity. The seagrasses could adapt to
268 lower oxygen concentrations by reducing the shoot density (Folmer et al., 2012) and thereby lower the oxygen demand
269 of the root-rhizome system, which may explain why the areas with high proportion of fine grain size particles in this
270 study had the lowest shoot density. High canopy height, high shoot density and shallow depths are generally considered



271 to increase sedimentation rates and thus promote accumulation of finer grain size particles (Bos et al., 2007; Fonseca
272 and Cahalan, 1992; Peralta et al., 2008). This implies that aboveground seagrass structure and water depth should
273 influence the sediment carbon storage, however, in our study these variables were of minor influence. The influence of
274 seagrass meadow structure on sediment composition is complex and hard to predict, and may be highly influenced by
275 environmental conditions (van Katwijk et al., 2010).

276

277 The carbon storage in *Z. marina* meadows in our study were clearly related to sediments with high proportion of fine
278 grain size particles, high porosity and low density. In areas without these sediment properties other variables, i.e.
279 above- and below-ground seagrass biomass, seagrass cover and shoot density, have an influence. For example, the
280 influence of belowground biomass and seagrass cover on sedimentary carbon content in Ria Formosa could be due to
281 the stabilizing properties of dense meadows (Suykerbuyk et al., 2015), the binding of sediment by the root-rhizome
282 system (Christianen et al., 2013) and the high lignin content of the belowground biomass (Klap et al., 2000).

283

284 The continuous loss of seagrass areas (Waycott et al., 2009) leads to a decline in natural carbon sinks (Dahl et al., 2016;
285 Marbà et al., 2015), and to ensure efficient management factors for high carbon storage capacity should be evaluated.
286 Several environmental and seagrass-related factors have shown to be of importance, i.e. water depth (Serrano et al.,
287 2014), meadow size (Ricart et al., 2015), hydrodynamics and seagrass canopy complexity (Samper-Villarreal et al.,
288 2016). In our study, the main factors related to high carbon storage were the sediment density and porosity, and
289 amount of fine grain size particles in the sediment, whereas the seagrass-associated variables had a minor influence.
290 Therefore, we highlight that the sediment characteristics is an important factor for a high carbon storage potential in
291 these types of *Z. marina* meadows, and should be taking into consideration when evaluating high priority areas for
292 protection of efficient carbon storage *Z. marina* areas.

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295



296 **Data availability**

297

298 All data is presented in the manuscript and supplementary figures. Figure S1, shows the variance of importance (VIP)
299 for the response variables in the PLS model (see Fig. 4). Figure S2, semi-log plots showing the relationship between
300 sediment porosity and density, and sedimentary organic carbon for unvegetated areas.

301

302 **Author contribution**

303

304 The design of this study was carried out by Martin Dahl, Martin Gullström, Mats Björk and Diana Deyanova. The
305 collection of data was done by Martin Dahl, Diana Deyanova, Liberatus D. Lyimo, Martin Gullström, Maria E. Asplund
306 and Ventzislav Karamfilov, and was analyzed by Martin Dahl, Martin Gullström, Mats Björk, Silvia Gütschow and Maria
307 E. Asplund. Martin Dahl prepared the manuscript with contribution from Martin Gullström, Mats Björk, Maria E.
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309

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311

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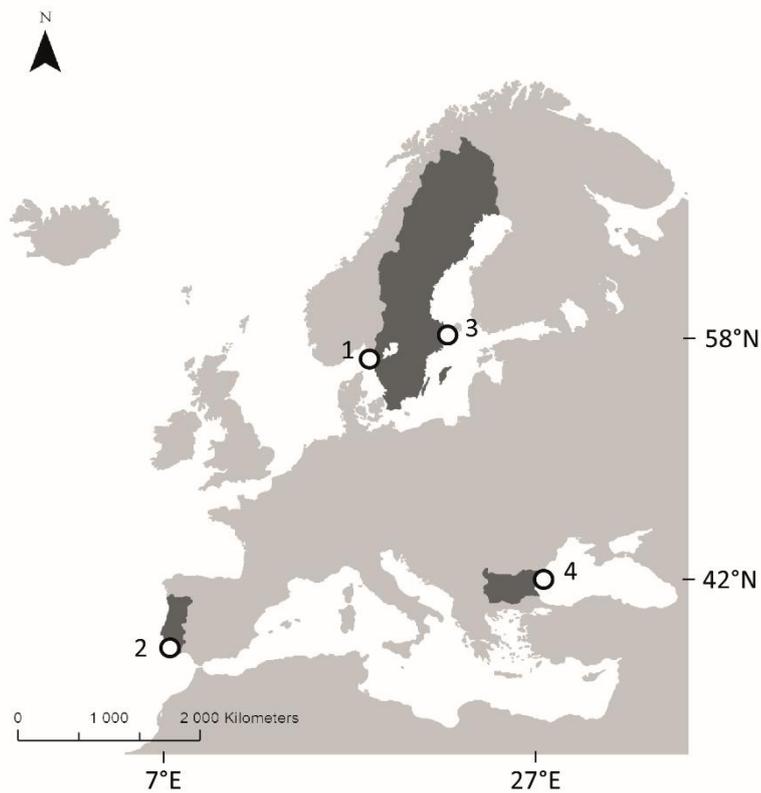
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449 Figure 1. The four study regions, Gullmar Fjord (Skagerrak, Sweden) (1), Ria Formosa (gulf of Cádiz, Portugal) (2), Askö
450 (Baltic Sea, Sweden) (3) and Sozopol (Black Sea, Bulgaria) (4).

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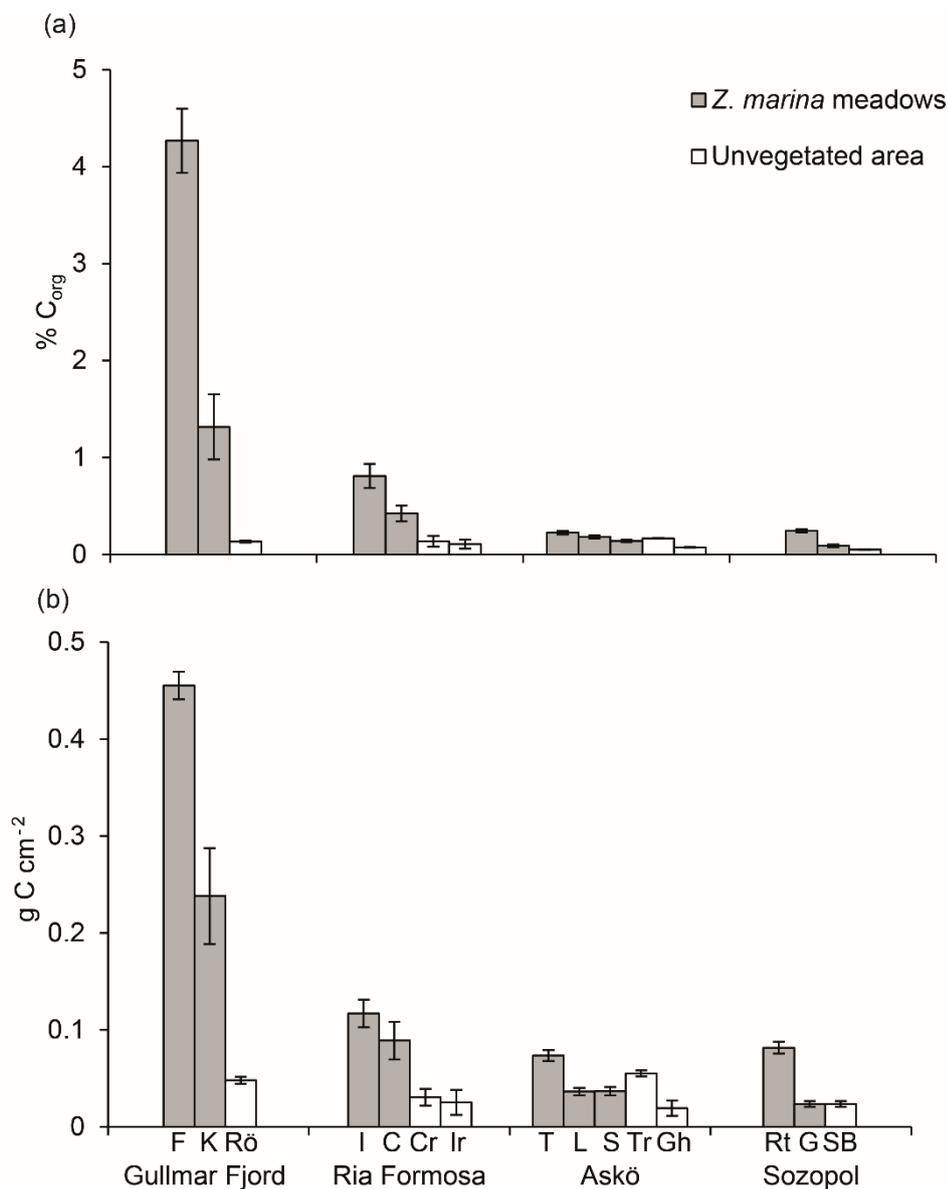
452 Table 1. Description of study sites in the four areas of Europe.

Area	Site	Vegetation	Coordinates	Mean depth (m)
Gullmar Fjord (Skagerrak, Sweden)				
	Finnsbo (F)	<i>Z. marina</i>	58°17'55N, 11°29'34E	2.8
	Kristineberg (K)	<i>Z. marina</i>	58°14'53N, 11°26'51E	3.0
	Rödberget (Rö) (r)	Unvegetated	58°15'06N, 11°27'54E	2.5
Ria Formosa (gulf of Cádiz, Portugal) ¹				
	Culatra channel (C)	<i>Z. marina</i> / <i>C. nodosa</i>	37°00'14N 7°49'36W	1.9
	Ilha da Culatra (I)	<i>Z. marina</i>	36°59'50N, 7°49'41W	1.0
	Culatra channel (Cr) (r)	Unvegetated	37°00'15N, 7°49'33W	2.6
	Ilha da Culatra (Ir) (r)	Unvegetated	36°59'51N, 7°49'40W	1.8
Askö (Baltic Sea, Sweden)				
	Torö (T)	<i>Z. marina</i> / <i>R. maitima</i>	58°48'14N, 17°47'32E	3.2
	Långskär (L)	<i>Z. marina</i> / <i>R. maitima</i>	58°48'00N, 17°40'48E	2.2
	Storsand (S)	<i>Z. marina</i>	58°48'26N, 17°41'40E	3.8
	Torö (Tr) (r)	Unvegetated	58°48'21N, 17°47'31E	6
	Godahoppsudden (Gh) (r)	Unvegetated	58°48'09N, 17°42'24E	2.9
Sozopol (Black Sea, Bulgaria)				
	Ropotamo (Rt)	<i>Z. marina</i> / <i>Z. noltii</i>	42°19'49N, 27°45'20E	2.7
	Gradina (G)	<i>Z. marina</i> / <i>Z. noltii</i>	42°25'39N, 27°39'05E	4.2
	Bay of Sozopol (r)	Unvegetated	42°24'42N, 27°39'48E	5.7

453 r = reference site (unvegetated area)

454 ¹Depth values standardized to mean low water (MLW).

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Figure 2. Mean (\pm SE) $\% C_{org}$ (a) and $g C cm^{-2}$ (b) in sediment (for 0-25 cm sediment depth). The percent organic carbon ($\% C_{org}$) is presented as a mean of the content of the top 25 cm sediment, while carbon per unit area ($g C cm^{-2}$) is the total (accumulated) amount of carbon in the top 25 cm of sediment. For full names of the sites see Table 1.

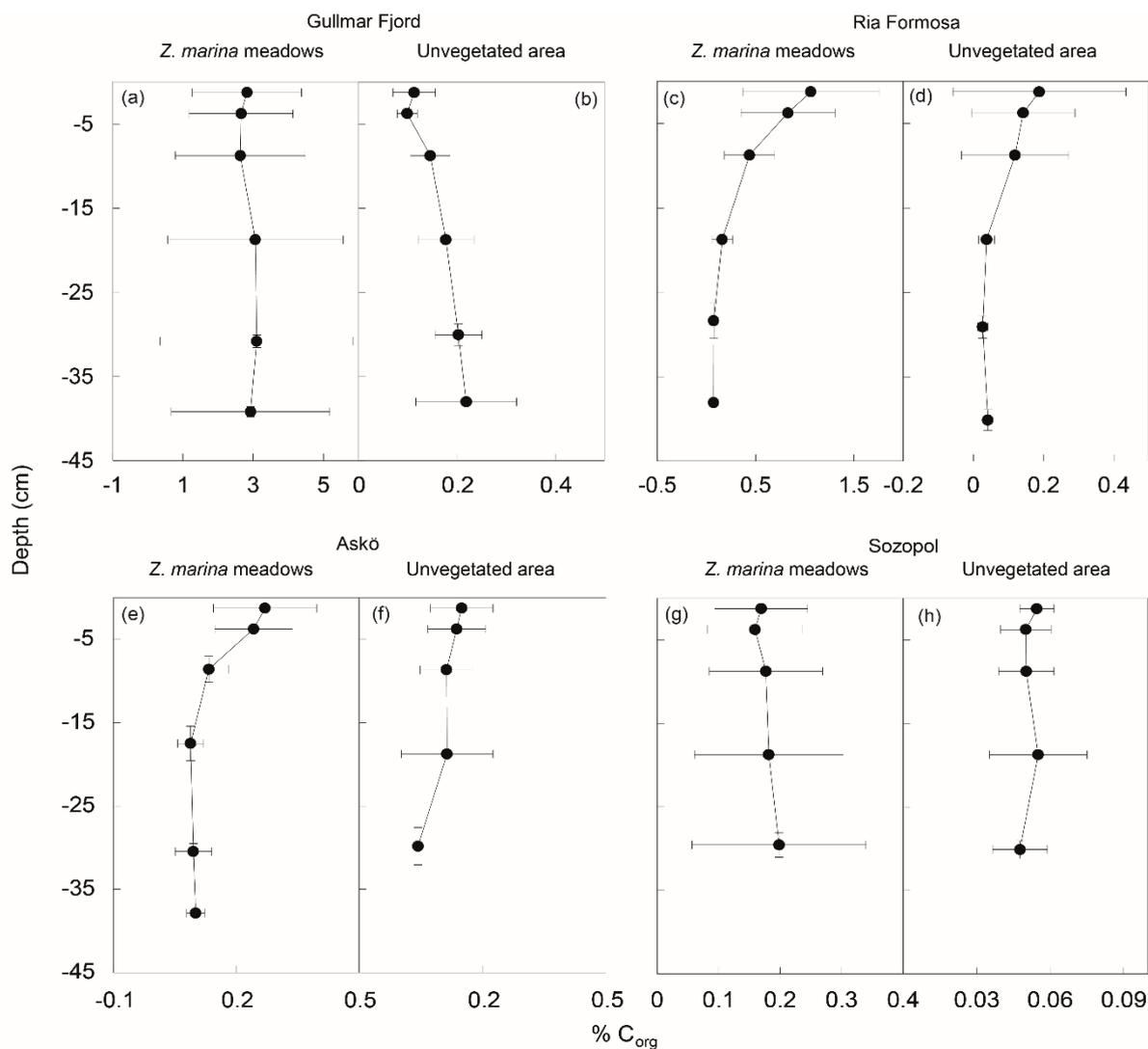


462 Table 2. Summary of Tukey's post hoc tests from significant one-way ANOVA models for comparison of sedimentary
 463 carbon between sites within each area separately. The percent organic carbon (% C_{org}) is presented as a mean of the
 464 content of the top 25 cm sediment, while carbon per unit area (g C cm⁻²) is the total (accumulated) amount of carbon
 465 in the top 25 cm of sediment. The full names of the sites are included in Table 1.

Area	Sites	% C _{org} <i>P</i>	g C cm ⁻² <i>P</i>
Gullmar Fjord			
	F vs. K	<0.001	0.009
	F vs. Rö (r)	<0.001	<0.001
	K vs. Rö (r)	<0.001	<0.001
Ria Formosa			
	I vs. C	ns	ns
	I vs Cr (r)	<0.001	0.005
	I vs. Ir (r)	<0.001	<0.001
	C vs. Cr (r)	0.009	0.035
	C vs. Ir (r)	0.002	0.002
Askö			
	T vs L	ns	<0.001
	T vs. S	0.002	<0.001
	T vs. Tr (r)	ns	ns
	T vs. Gh (r)	<0.001	<0.001
	L vs. S	ns	ns
	L vs. Tr (r)	ns	0.021
	L vs. Gh (r)	<0.001	<0.001
	S vs. Tr (r)	ns	0.026
	S vs. Gh (r)	<0.001	<0.001
Sozopol			
	R vs. G	<0.001	<0.001
	R vs. SB (r)	<0.001	<0.001
	G vs. SB (r)	<0.001	ns

466 r = reference site (unvegetated areas).

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468
 469 Figure 3. Mean sedimentary carbon (% C_{org} ± SD) depth profiles grouped for the different regions showed as mean slice
 470 depth. Note that the scale on the x-axes differs among the different depth profiles.
 471



Table 3. Seagrass meadow variables (mean ± SD) for the different regions.

Region	Shoot density (m ⁻²)	Shoot height (cm)	Seagrass cover (%)	Aboveground biomass				Belowground biomass			
				%C	%N	C:N	g DW m ⁻²	%C	%N	C:N	g DW m ⁻²
Gullmar Fjord	157.9 ± 43.8	81.4 ± 18.2	36.9 ± 14.0	38.8 ± 0.7	2.1 ± 0.3	18.4 ± 2.0	39.4 ± 31.1	34.2 ± 1.4	1.1 ± 0.1	30.2 ± 2.8	253.0 ± 86.0
Ria Formosa	264.9 ± 107.8	32.5 ± 4.4	79.1 ± 10.6	34.4 ± 1.2	1.4 ± 0.1	25.2 ± 1.2	108.3 ± 58.5	30.8 ± 3.2	1.1 ± 0.1	37.8 ± 4.5	494.2 ± 230.2
Askó	338.1 ± 160.3	51.7 ± 12.4	47.6 ± 18.1	37.2 ± 2.0	1.8 ± 0.3	20.9 ± 2.9	255.7 ± 193.4	32.8 ± 2.9	1.1 ± 0.1	31.5 ± 3.2	205.9 ± 88.7
Sozopol	419.6 ± 315.3	63.5 ± 11.2	63.5 ± 11.1	36.0 ± 1.3	1.9 ± 0.3	19.6 ± 3.2	122.0 ± 110.8	30.3 ± 3.0	0.8 ± 0.2	28.6 ± 5.7	86.4 ± 80.6

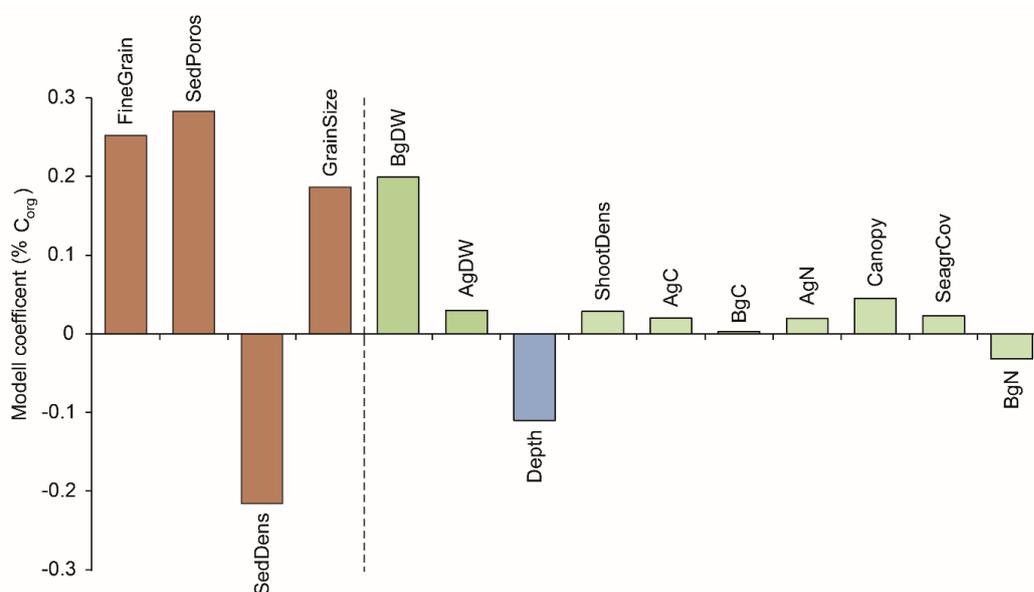


473 Table 4. Seagrass sediment data as mean (\pm SD) for the depth profile (0-25 cm) in the different regions. Mean grain size is
 474 presented with phi (ϕ) units.

Region	Sediment porosity (%)	Sediment density (g DW mL ⁻¹)	% N	C:N	Mean grain size (ϕ)	Sediment particles <0.074 mm (%)
Gullmar Fjord	67.0 \pm 14.1	0.71 \pm 0.33	0.28 \pm 0.16	9.39 \pm 1.26	4.89 \pm 0.93	62.8 \pm 25.6
Ria Formosa	43.0 \pm 5.4	1.13 \pm 0.14	0.08 \pm 0.01	7.15 \pm 0.83	2.34 \pm 0.56	17.9 \pm 5.8
Askö	31.9 \pm 2.4	1.4 \pm 0.15	0.03 \pm 0.01	5.60 \pm 1.27	1.19 \pm 0.79	3.7 \pm 0.6
Sozopol	41.8 \pm 5.2	1.25 \pm 0.04	0.05 \pm 0.04	3.22 \pm 1.25	2.08 \pm 0.27	2.6 \pm 1.9

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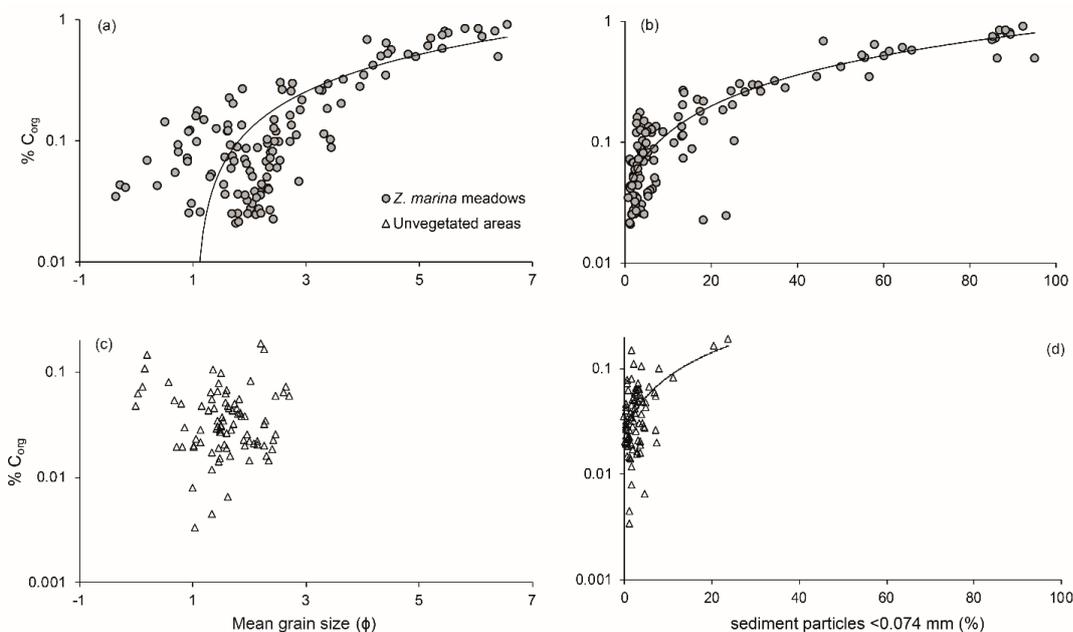
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478 Figure 4. PLS (Partial Least Square) regression model coefficient plot for % C_{org} in sediment (using a mean of the carbon
 479 content for the top 25 cm sediment). The predictor variables are ranked in level of importance (left to right) where the four
 480 variables left of the striped bar having a VIP-value >1 (i.e. FineGrain, SedPoros, SedDens and GrainSize) and hence
 481 significantly influencing % C_{org}. Brown = sediment characteristics, green = seagrass-associated variables and blue = water
 482 depth; Variables included in the model were FineGrain (sediment particles <0.074 mm, %), SedPoros (sediment porosity,
 483 %), SedDens (sediment density, g DW mL⁻¹), GrainSize (mean grain size, φ), Bg and Ag DW (belowground biomass dry
 484 weight, m²), Depth (water depth, m), ShootDens (shoot density, m²), Ag and Bg biomass C and N (biomass carbon and
 485 nitrogen content, %), Canopy (shoot height, cm) and SeagrCov (seagrass cover, %).

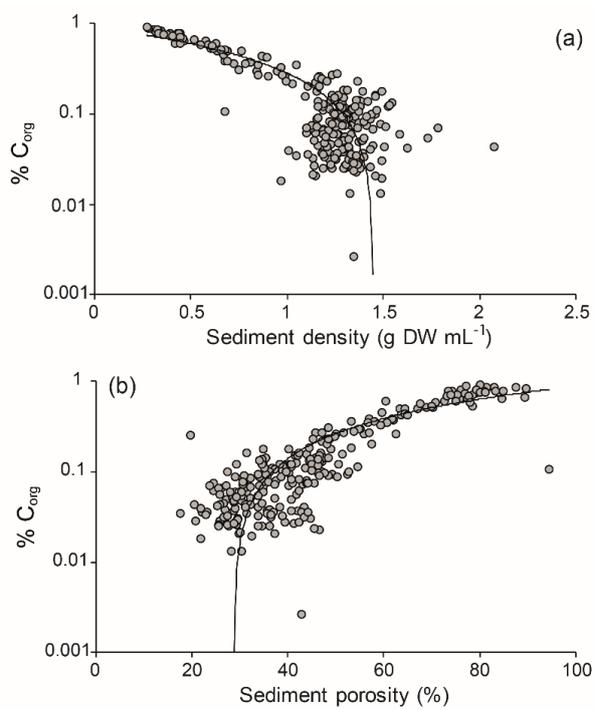
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488 Figure 5. Semi-log plots ($\log_{10}[x+1]$) showing the relationship between % C_{org} and grain size. The % C_{org} is presented with
 489 a log scale as it gave the best fit of the models. Grain size is shown as mean grain size (ϕ) and sediment particles <0.074
 490 mm (%) for *Z. marina* meadows (a and b) and unvegetated areas (c and d). The % C_{org} was positively linked to both sediment
 491 particles <0.074 mm (%) ($R^2 = 0.91$, $P < 0.001$) and mean grain size (ϕ) ($R^2 = 0.74$, $P < 0.001$) for *Z. marina* meadows but
 492 for unvegetated area only sediment particles < 0.074 mm (%) showed this relationship with % C_{org} ($R^2 = 0.42$, $P < 0.001$).

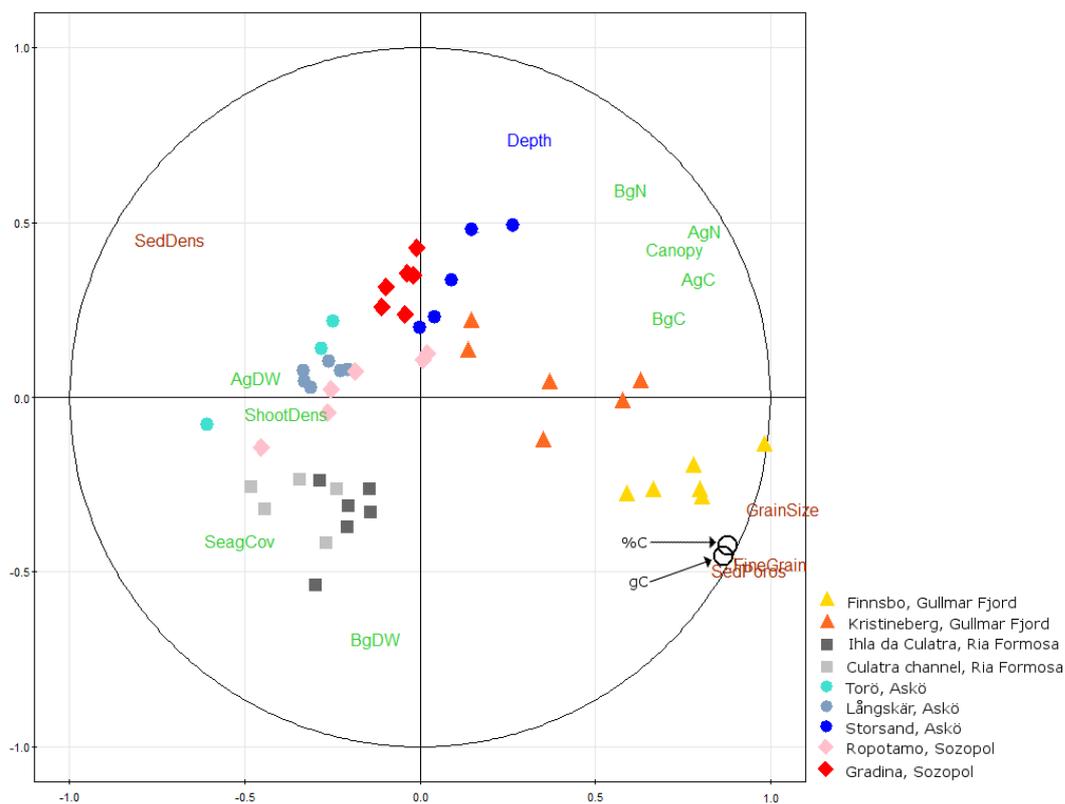
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495 Figure 6. Semi-log plots ($\log_{10}[x+1]$) for sediment density (a) and sediment porosity (b) in relation to % C_{org} for the *Z. marina*
496 sites. The sediment density (g DW mL^{-1}) was negatively influencing the amount of organic carbon ($R^2 = 0.84$, $P < 0.001$),
497 while there was a positive relation between sediment porosity (%) and % C_{org} ($R^2 = 0.80$, $P < 0.001$).

498



499

500 Figure 7. PCA (Principal Component Analysis) showing the nine seagrass sites, the two response variables (sedimentary %
501 C_{org} and $g C cm^{-2}$) and predictor variables (14 in total). The percent organic carbon ($\% C_{org}$) is presented as a mean of the
502 content of the top 25 cm sediment, while carbon per unit area ($g C cm^{-2}$) is the total (accumulated) amount of carbon in the
503 top 25 cm of sediment. The colors of the letters represent different groups of predictor variables; brown = sediment
504 characteristics, green = seagrass-associated variables, blue = water depth. Black circles are the response variables, i.e.
505 organic carbon ($\% C = \% C_{org}$ and $g C = g C cm^{-2}$). For explanations to the abbreviations of predictor variables see Fig. 4.

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507