

# The burying and grazing effects of Plateau pika on alpine grassland are small: A pilot study in a semi-arid basin on the Qinghai-Tibetan Plateau

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## Abstract

There is considerable controversy about the role of plateau pika (*Ochotona curzoniae*, hereafter pika) in alpine grassland on the Qinghai-Tibetan Plateau (QTP). It is on one hand considered as a keystone species, on the other hand poisoned. Although significant amount of efforts have been made to study the effects of pika at a quadrat scale ( $\sim\text{m}^2$ ), our knowledge about its distribution and effects at a larger scale is very limited. In this study, we investigated the direct effects, i.e. burying and grazing, of pika by upscaling field sampling at a quadrat scale to a plot scale ( $\sim 1,000\text{ m}^2$ ) by aerial photographing. Altogether, 168 plots were set on 4 different types of alpine grassland in a semi-arid basin on the QTP. Results showed that: 1) the effects of burying by pika piles on the reduction of vegetation cover, biomass and soil carbon/nitrogen were less than 10%, which was much smaller than the effects of bald patches; and 2) pika consumed 8-21% of annual net primary production of grassland. We concluded that the direct burying and grazing effects of pika on alpine grassland were minor in this region. Quadcopter is an efficient and economic tool for long-term repeated monitoring over large regions for further understanding the role of pika.

1

## 2 **1. Introduction**

3 Alpine grassland is important for animal husbandry and occupies about 2/3 of the  
4 total area of the Qinghai-Tibetan Plateau (QTP), but about 1/3 of this resource has  
5 degraded over the last few decades (Li et al., 2011). In addition to overgrazing (Zhang  
6 et al., 2014), climate warming and permafrost degradation (Wang et al., 2008; Yi et al.,  
7 2011), small mammals, especially plateau pika (*Ochotona curzoniae*), are considered  
8 an important cause of grassland degradation.

9 Plateau pika (hereafter pika), a small lagomorph, is believed adversely affecting  
10 alpine grassland by consuming biomass, destroying the sod layer, burying vegetation  
11 with excavated soil and expediting carbon dioxide emission (Qin et al., 2015a). The  
12 bald patches created by pika activity may increase in size over time because of  
13 erosion by wind and/or water (Wei et al., 2007). According to Shang and Long (2007),  
14 16-54% of degraded grassland is severely degraded, the so-called “black soil patch”,  
15 half of which is caused by pika (Li and Sun, 2009). For this reason, local government  
16 considers pika a pest of alpine grassland and has initiated campaigns to eradicate it  
17 since 1958 (Wilson and Smith, 2014). On the other hand, pika is believed to benefit  
18 alpine grassland by increasing infiltration, decreasing runoff (Wilson and Smith, 2014)  
19 and increasing moisture and carbon content (Li and Zhang, 2006) in the top soil (up to  
20 a depth of 10 cm). Pika is also a keystone species on the QTP (Smith and Foggin,  
21 1999; Lai and Smith, 2003). Some authors have suggested that pika is an indicator  
22 rather than a cause of grassland degradation; pika population increases quickly only  
23 after the grassland has already been degraded (Harris, 2010; Wangdwei et al., 2013).

24 Although the role of pika in alpine grassland ecosystem is receiving more and  
25 more attention, there have been few quantitative studies at plot scale (e.g.  $\sim 1000 \text{ m}^2$ ,  
26 Guo et al., 2012; Wangdwei et al., 2013). Typically, studies on pika effects have  
27 compared vegetation and soil characteristics and carbon fluxes at a quadrat scale ( $\sim \text{m}^2$ )  
28 among plots with different number densities of pika burrows (Guo et al., 2012; Li and  
29 Zhang, 2006; Liu et al., 2013; Wei et al., 2007; Wilson and Smith, 2014). For example,  
30 Liu et al. (2013) investigated the role of pika in alpine steppe meadows studying 8

1 plots with pika burrow exit numbers varying from 0 to 76 burrow exits/100m<sup>2</sup> and  
2 found that a higher density of pika burrow exits was associated with lower net  
3 ecosystem exchanges, aboveground biomass and number of species. There are  
4 different levels of heterogeneity on grassland surfaces. For example, Wei et al. (2007)  
5 classified the grassland surface into six types: 1) mound height > 10 cm; 2) mound  
6 height between 0 and 10 cm; 3) erosion pit between 0 and 5 cm; 4) erosion pit  
7 between 5 and 10 cm; 5) erosion pit > 10 cm; and 6) undisturbed. It is critical that  
8 measurements taken at a quadrat scale be converted to a plot scale in order to properly  
9 quantify the role of pika. However, it is hard and inefficient to walk around ground to  
10 count the number of burrow exits or piles of pika *in situ* on large amounts of plots (e.g.  
11 Liu et al., 2013), not to say to quantify their area fractions in each plot. Therefore, few  
12 studies have quantified the effects of pika on alpine grassland at plot scale.

13 Lightweight Unmanned aerial vehicles (UAVs) have developed rapidly due to  
14 miniaturization and low cost of various sensors and embedded computers (Salami et  
15 al., 2014). UAVs have become a popular platform at a low cost for high precision  
16 photography and other applications recently. Photography with cm-level resolution  
17 can be achieved using widely-used camera (Colomina and Molina, 2014). In this  
18 study, we applied a UAV with camera to take aerial photos and aimed to: 1) test  
19 whether pika burrow exits and piles information can be retrieved from aerial  
20 photographs at a plot scale; 2) upscale the measurements of biomass, soil carbon and  
21 nitrogen measured at quadrat scale to plot scale and quantitatively assess the burying  
22 and grazing effects of pika.

## 23 **2. Methodology**

### 24 **2.1 Study area and field work**

25 The study area is located in the source region of the Shule River Basin on Qilian  
26 Mountain at the northeastern edge of the QTP, China (Figure 1 a). The area has an  
27 arid continental climate. The average annual air temperature and precipitation are  
28 about -4.0 °C and 200-400 mm (Chang et al., 2016). There are four typical types of  
29 alpine grassland in the study area: alpine steppe (AS); alpine steppe meadow (AStM);  
30 alpine meadow (AM); and alpine swamp meadow (ASwM) (Figure 1 b-e). The soil

1 moisture ranges from dry in AS grassland to wet in ASwM grassland (Qin et al.,  
2 2014). Accordingly, the dominant species was *Stipa purpurea* in AS grassland and  
3 *Kobresia pygmaea* in ASwM grassland (Table 1). We conducted field studies with  
4 field sampling and aerial photographing. We set up 3 30 m X 30 m plots in each of  
5 four types of grassland in 2012, and we set up 9 50 cm X 50 cm quadrats evenly in  
6 each plot. We took photo on each quadrat from May 19 to August 30, 2012 at a height  
7 of 1.4 m. The protocol of measurement and estimation of fractional vegetation cover  
8 can be found in Section 2.2 and 2.4, respectively. Vegetation cover usually peaks  
9 during the end of July and beginning of August (Figure 2).

## 10 **2.2 Field sampling**

11 For each grassland type, we delineated 4 surface types: vegetation patch; new pika  
12 pile (with loose soil and a burrow exit nearby); old pika pile; and bald patch (Figure 3  
13 d-g). At end of July 2014, we randomly set up 3 quadrats with iron frames measuring  
14 50 cm × 50 cm on each surface type in each type of grassland (Figure 3 a). For new  
15 and old pika pile surface types (Figure 3 f and g), the iron frames were placed so as to  
16 cover vegetation as little as possible. We took one picture of each quadrat with an  
17 ordinary digital camera (Fujifilm (Japan), 1000 megapixels) held vertically at a height  
18 of ~1.4 m (Figure 3 d-g). Five soil cores were collected on each quadrat with a  
19 stainless auger (5 cm in diameter) down to 40 cm (Figure 3 c), and bulked as one  
20 composite sample. Three replicates on each surface type of each grassland type were  
21 sampled.

22 At the beginning of August 2015, we set three round plots with radius of 14 m around  
23 sampling place in each type of grassland (Figure 3 h). Distance between plots was  
24 over 50 m. We covered all burrow exits with soil within each plot. The number of  
25 burrow exits which were opened was counted after 72 hours. Then we put trap on  
26 each of the opened burrow exit, and checked whether pika was caught after 48 hours.  
27 The experiment protocol was approved by Department of Qinghai Prataculture (Due  
28 to the small size of experiment, only oral approval was granted).

## 29 **2.3 Aerial photographing**

30 At beginning of August 2015, we selected 14 locations, among which 4, 4, 4 and 2

1 locations were in AS, AStM, AM and ASwM grasslands respectively. (Figure 1).  
2 There were 3, 2, 0, and 0 locations on the alluvial terrace; and 1, 2, 2, and 2 locations  
3 on river terrace. All locations are generally flat with slope less than  $4^\circ$ . Grasslands of  
4 these locations are used for grazing during migration between settlement and  
5 mountain areas in May-June and September-October. Pikas of these locations are not  
6 poisoned. One location in each type of grassland was over the above-mentioned  
7 sampling plots and quadrats (Figure 3 a). On each location, DJI drone (Phantom 3  
8 Professional, DJI Innovation Company, China) was auto-piloted to 12 preset way  
9 points to take photo at a height of 20 m with camera looking vertically down using  
10 software development kits (Yi, submitted). Altogether 168 aerial photos were taken.  
11 The Phantom 3 Professional is a light-weight (about 1280 g including battery and  
12 propellers) four-wheel drone. It is equipped with an autopilot system with 0.5 m  
13 vertical accuracy and 2.0 m horizontal accuracy. It is integrated with a Sony EXMOR  
14 Sensor (maximum image size:  $4000 \times 3000$ ) and a 3-axis gimbal. Each aerial photo  
15 covers roughly  $35 \text{ m} \times 26 \text{ m}$  (Figure 3 a and b), and each pixel covers roughly  $1 \text{ cm}^2$   
16 ground area, when photo is taken at a height of 20 m.

#### 17 **2.4 Image analysis**

18 For those images taken on ground, we selected the part of the image within the iron  
19 frame and retrieved green fractional vegetation cover (GFVC) using a threshold  
20 method based on excess green index ( $\text{EGI} = 2G - R - B$ ; with R, G, B being red, green  
21 and blue bands, respectively) of each pixel. More specifically, to calculate GFVC we:  
22 1) provided an initial value of EGI threshold and compared it with each pixel; 2) if the  
23 EGI of a pixel was greater than the threshold, the pixel was considered a vegetation  
24 pixel and assigned a green color; otherwise it was considered a non-vegetation pixel  
25 and assigned a yellow color; 3) compared the classified image with the original  
26 picture. Steps 1) to 3) were iterated to adjust the threshold value until the vegetation  
27 shapes in the classified image fit those of the original picture (Figure 4). Finally, we  
28 calculated GFVC by dividing the number of vegetation pixels into the total number of  
29 pixels.

30 For pictures taken from the air (Figure 5), the new and old pika piles were marked

1 manually with rectangles so as to include as little intact vegetation as possible (Figure  
2 5). We plotted the contours of the vegetation and bald patches using OpenCv Library:  
3 1) adjusted the EGI value until its contours fit well with the shape of the vegetation  
4 and bald patches (Figure 5), 2) calculated the area in each contour in units of pixel  
5 using OpenCv Library; and 3) we subtracted the number of vegetation and  
6 non-vegetation pixels of new and old pika piles from the vegetation and bald patch  
7 contours, respectively. To exclude very small patches, we only considered the patches  
8 with area greater than 10 cm<sup>2</sup>. The area fractions of vegetation and bald patches, new  
9 and old pika piles were then calculated by dividing the number of pixels in each  
10 surface type by the total number of pixels (see Figure 3b).

11 It is worth mentioning that no true FVCs of grassland at both quadrat and plot scale  
12 exist. Results from supervised classification have been usually used as “true” FVC  
13 values. For example, supervised classification using WinCAM software was used in  
14 Yi et al. (2011), Ren et al. (2014) and Zhou et al. (2015). Yi et al. (2011) found that  
15 widely used visual estimation had large variations among different estimators, while  
16 WinCAM classification was time-consuming. Ren et al. (2014) found that try & error  
17 threshold-based estimation was comparable to that of WinCAM classification and was  
18 more efficient than WinCAM. In addition to EGI based threshold method, we also  
19 tried green relative vegetation index ( $GRVI=(G-R)/(G+R)$ , Motohka et al., 2010), we  
20 did not try the normalized difference vegetation index ( $NDVI=(NIR-R)/(NIR+R)$ ,  
21 where NIR is near infrared band), due to lack of NIR band in a common camera.

## 22 **2.5 Laboratory analysis**

23 Soil samples were processed in the following steps: 1) air-dried in natural condition  
24 avoiding direct sunshine; 2) the gravel, >2 mm in size, was sieved, separated and  
25 weighted by electronic balance (0.01g); 3) the remaining soil samples with diameter  
26 less than 2 mm were ground to pass through a 0.25 mm sieve and were then sent to  
27 Lanzhou University for analysis of soil organic carbon (SOC) and total nitrogen (TN)  
28 concentration. A detailed description of the analysis methods for SOC and TN can be  
29 found in Qin et al. (2014).

## 30 **2.6 Data analysis**

### 1 **2.6.1 Plot scale biomass, soil organic carbon and total nitrogen**

2 Based on the relationship between GFVC and aboveground biomass (AGB) at a  
3 quadrat scale, established using datasets of the same study area (Qin et al., 2014), we  
4 calculated AGB (kg/ha) = 21.6 × GFVC for each of surface type. For each plot, we  
5 calculated the overall AGB with the following equation:

$$6 \quad \text{AGB}_{\text{plot}} = \text{AGB}_{\text{np}}f_{\text{np}} + \text{AGB}_{\text{op}}f_{\text{op}} + \text{AGB}_{\text{bp}}f_{\text{bp}} + \text{AGB}_{\text{vp}}f_{\text{vp}} \quad (1)$$

7 Where plot, np, op, bp, and vp represent plot, new pika pile, old pika pile, bald  
8 and vegetation patches, respectively; f represents area fraction (%) of each surface  
9 type. The SOC and TN at plot scale were then calculated in a similar way as that of  
10 AGB.

11 We defined the effect of each surface type ( $E_{\text{type}}$ ) on AGB reduction of grassland as:

$$12 \quad E_{\text{type,agb}} = \frac{(\text{AGB}_{\text{type}} - \text{AGB}_{\text{vp}})f_{\text{type}}}{\sum[(\text{AGB}_{\text{type}} - \text{AGB}_{\text{vp}})f_{\text{type}}]} \times 100 \quad (2)$$

13 Where  $f_{\text{type}}$  represents the area fraction of a surface type in a plot (%),  $\Sigma$  means  
14 the sum. For the vegetation patch surface type,  $E_{\text{type}}$  equals 0 and has no effect in  
15 AGB reduction. The higher the value of  $E_{\text{type,agb}}$ , the higher the effect of a surface type  
16 on plot-scale AGB reduction. The effects on SOC and TN reduction were calculated  
17 in a similar way. The burying effects from pika piles were calculated as the sum of  $E_{\text{np}}$   
18 and  $E_{\text{op}}$ .

### 19 **2.6.2 Plot scale pika number and grazing effects**

20 Two ratios were used in calculating number of pika from aerial photos at plot scale.  
21 First was the ratio ( $r1$ ) between the number of in-use burrow exits and the total  
22 number of burrow exits, and the ratio ( $r2$ ) between the number of pikas caught and the  
23 number of in-use burrow exits, both of which were developed using field data for  
24 each grassland type (Figure 3 h). We then calculated the number of pikas in a plot  
25 covered by each aerial photo (Figure 3 b) with these two ratios and the total number  
26 of pika piles delineated from each aerial photo (Figure 5; equation 3).

$$27 \quad N_{\text{pika}} = N_{\text{pile}} \times r1 \times r2 \quad (3)$$

28 Where  $N_{\text{pika}}$  and  $N_{\text{pile}}$  are the number of pikas and the number of total pika piles  
29 in a hectare, respectively.

1 Each pika consumes ~8.06 kg of grass dry matter per year (Hou, 1995; equation 4).  
2 Pika consumes above-ground biomass more than root system (Sun et al., 2016). The  
3 annual primary production of grassland roughly equals to peak time aboveground  
4 biomass ( $AGB_{plot}$ ; Scurlock et al., 2002). Finally, we estimated the effects of direct  
5 graze consumption by pika ( $E_{graze}$ , %) in a plot (Equation 5).

$$6 \quad AGB_{pika} = N_{pika} \times 8.06 \quad (4)$$

$$7 \quad E_{graze} = \frac{AGB_{pika}}{AGB_{plot}} \times 100 \quad (5)$$

8  $AGB_{pika}$  is the biomass consumed by pika (kg/ha).

9 The data were presented as mean  $\pm$  standard deviation. Statistical analyses were  
10 performed using the SPSS 17.0 statistical software package (SPSS Inc., Chicago, IL,  
11 USA). One-way analysis of variance (ANOVA) and a multi-comparison of a least  
12 significant difference (LSD) test were used to distinguish between differences at the  
13  $p=0.05$  level.

### 14 **3. Results**

#### 15 **3.1 Quadrat scale characteristics**

16 The GFVCs derived using thresholds of EGI and GRVI were similar, with differences  
17 less than 1% (Figure not shown here). Therefore, in the following part, we only  
18 presented results based on EGI threshold.

19 The GFVCs of the vegetation patches were greater than 60% for both AM and ASwM  
20 grasslands, while those of AS and AS<sub>t</sub>M grasslands were less than 30% (Figure 6a).

21 The GFVC of vegetation patches was significantly greater than that of other surface  
22 types for most of the grasslands ( $p<0.05$ ). Because some vegetation was included in  
23 the  $50 \times 50$  cm iron frame, the GFVC of new pika pile was not zero, but was usually  
24 less than 10%. Vegetation also grew on the piles, so the GFVC of old pika pile was  
25 usually greater than that of new pika pile. Bald patch GFVC was similar to that of  
26 new pika pile.

27 The SOC/TN densities of 40 cm soil column ranged between 3.5/0.45 and 8.0/1.2  
28  $kg/m^2$  (Figure 6b and c). Both SOC and TN densities under vegetation patches were  
29 significantly greater than those under bald patch ( $p<0.05$ ). SOCs under vegetation



1 patches of 3 out of 4 grasslands were significantly greater than those under new and  
2 old pika piles (Figure 6b). TNs under vegetation patches were only significantly  
3 greater than those of new and old pika piles on the ASwM grassland (Figure 6c).  
4 Species in vegetation patches were dominant by palatable species, while forbs with  
5 low-nutrient were common on bald patches and old pika piles on all 4 different  
6 grasslands (Table 1).

### 7 **3.2 Area fractions and numbers of surface types at plot scale**

8 Except for the ASwM grassland, the mean area fractions of vegetation patches were  
9 about 30%, and significantly less than bald patches ( $p < 0.05$ ; Figure 7a). The mean  
10 area fractions of new and old pika piles were all less than 2% for all grasslands  
11 (Figure 7b). The mean number of patches of vegetation (bald) patches ranged from  
12  $\sim 33,000/\text{ha}$  ( $17,000/\text{ha}$ ) in AM grassland to  $\sim 100,000/\text{ha}$  ( $67,000/\text{ha}$ ) in AStM  
13 grassland (Figure 7c). The mean number of new (old) pika piles ranged from  $\sim 130/\text{ha}$   
14 ( $160/\text{ha}$ ) to  $\sim 270/\text{ha}$  ( $400/\text{ha}$ , Figure 7d).

### 15 **3.3 Effects of surface types at plot scale**

16 Due to the large area fractions of bald patches (Figure 7a) and low vegetation cover  
17 (Figure 6a), the effects of bald patches on reduction of above-ground biomass ranged  
18 from 80% on ASwM grassland to 98% on AS and AStM grasslands (Figure 8a). The  
19 effects of pika piles were significantly less than that of bald patches. The soil organic  
20 carbon and total nitrogen had the similar pattern as that of above-ground biomass  
21 (Figure 8 b and c).

### 22 **3.4 Grazing effects of pika at plot scale**

23 The mean ratio between in-use burrow exits and total burrow exits ( $r_1$ ) ranged from  
24 0.22 to 0.42, and there were no significant differences among different grassland types  
25 ( $p \geq 0.05$ ; Figure 9a). The mean ratio between number of pikas and in-use burrow  
26 exits ( $r_2$ ) ranged from 0.18 on ASwM grassland to 0.4 on AM grassland (Figure 9b).  
27 The  $r_2$  ratio of ASwM grassland was significantly less than those of the other  
28 grasslands ( $p < 0.05$ ). The mean number of pikas ranged from  $27 \text{ ha}^{-1}$  to  $60 \text{ ha}^{-1}$ , and  
29 there were no significant differences among different types of grasslands ( $p \geq 0.05$ ;  
30 Figure 9c). The graze effects of pika on aboveground biomass ranged from 8% to

1 21%, with that on AStM significantly greater than those of the other grasslands  
2 ( $p < 0.05$ ; Figure 9d).

### 3 **4. Discussion**

#### 4 **4.1 Burying and grazing effects of pika on grassland**

5 Previous studies indicated that pika adversely affect alpine grassland directly through  
6 1) burying of vegetation with soil while burrowing and 2) consumption of vegetation  
7 in competition with domestic animals for food (Yang and Jiang, 2002). However, our  
8 study showed that both new and old pika piles accounted for only a very small area  
9 fraction ( $< 2\%$ ) of the total plot area (Figure 7b), showing that burying has minimal  
10 effects on aboveground biomass, soil carbon and total nitrogen (Figure 8). The  
11 aboveground biomass at peak growing season is usually used as surrogate of annual  
12 net primary production (Scurlock et al., 2002). Pika only accounted for 21% at  
13 maximum on different types of grassland on two different geomorphologies (Figure  
14 9d).

15 Sun et al. (2016) classified study sites into four classes: 1) approximately zero pika  
16 density ( $0-15 \text{ ha}^{-1}$ ); 2) low pika density ( $15-110 \text{ ha}^{-1}$ ); 3) medium pika density  
17 ( $110-200 \text{ ha}^{-1}$ ), and 4) high pika density ( $200-300 \text{ ha}^{-1}$ ). Our plots belong to the first  
18 two classes (Figure 9 c). Due to different precipitation and temperature conditions, net  
19 primary production, soil carbon and nitrogen exhibits strong spatial heterogeneity  
20 (Luo et al., 2004). Therefore, to properly evaluate the direct burying and grazing  
21 effects of pika on the QTP, large amounts of plots under different combined  
22 conditions of climate and pika densities should be investigated.

#### 23 **4.2 Effects of pika on bald patches**

24 There were bald patches of various sizes on the grasslands (see Figure 5), which  
25 played a much more important role than pika piles in reducing vegetation cover,  
26 aboveground biomass and soil carbon and nitrogen at the plot scale (Figure 8). We  
27 retrieved gravel contours using the threshold of R+G+B and determined whether each  
28 was in a vegetation or bald patch contour. The number of gravel contours in bald  
29 patches was significantly greater than the number in vegetation patch contours (e.g.  
30 Figure 3 e and 5). For example, there was  $\sim 80/5$  gravel/ $\text{m}^2$  in bald/vegetation patches

1 on the AM grassland (Figure not shown). High amounts of gravel content are not  
2 beneficial for nutrient retention and vegetation growth (Qin et al., 2015b): once the  
3 fine soil has been eroded, vegetation in a bald patch is slow to recover (Gao et al.,  
4 2011).

5 Wei et al. (2007) suggested that a bald patch developed from a new pika pile through  
6 its succession to an old pika pile and further erosion by wind and/or water. Other  
7 studies have suggested that a bald patch originates from the collapse of a burrowing  
8 tunnel, repeated freeze and thaw processes, trampling during grazing or some  
9 combination of these factors (Zhou et al., 2003; Cao et al., 2010). However, none of  
10 these suggestions have been supported by field observations (Wilson and Smith,  
11 2014). It is, therefore, critical to perform long-term repeated monitoring studies to  
12 determine: 1) whether bald patches are developed from pika piles or burrow tunnels?;  
13 2) how quickly does a bald patch expand?; and 3) what are the major factors affecting  
14 bald patch expansion?

#### 15 **4.3 Cons and pros of quadcopter in studying pika's effects**

16 Pika piles or burrow exits and bald patches are too numerous to be quantified easily  
17 on ground by human; they are also too small to be identified by regularly available  
18 satellite remote sensing data (Figure 5 and 7). Quadcopter integrated with a camera  
19 has the following advantages in studying pika's effects: 1) large coverage. It can  
20 easily cover an area of  $\sim 1000 \text{ m}^2$  when it is flied at a height of  $\sim 20 \text{ m}$ , therefore, aerial  
21 photos can be used to better characterize patches of different sizes than photos taken  
22 on ground; 2) high resolution. Each pixel represents area if  $\sim 1 \text{ cm}^2$  when photo is  
23 taken at a height of  $\sim 20 \text{ m}$ , which is good enough for identifying pika piles and bald  
24 patches (Figure 5); 3) high locating accuracy. The distance between the center of an  
25 aerial photo and the corresponding preset way point is  $\sim 1 \text{ m}$ , which makes it feasible  
26 for repeated monitoring over the same plots (Yi, submitted); 4) low cost. Each  
27 Phantom 3 quadcopter costs about 1,000 USD; and 5) high efficiency. In our study, it  
28 took only 2 minutes to fly to 12 preset way points and take photos automatically  
29 (Figure 3a).

30 Chen et al. (2016) found that the fractional vegetation cover derived from aerial

1 photos had better correlations with satellite normalized difference vegetation index,  
2 which is usually used to estimate vegetation biomass (e.g. Gao et al., 2013), than  
3 quadrat-scale photo taken on ground on patchy grassland. It is a non-destructive  
4 method to estimate biomass or soil carbon/nitrogen at plot scale with only few  
5 samples at quadrat scale sampled. Therefore, it is feasible to deploy quadcopter to  
6 monitor large amounts of plots in alpine grassland on the QTP repeatedly over a  
7 long-term range.

8 However, we do acknowledge that there are some shortcomings of quadcopter: 1) we  
9 cannot assess role of pika at species level with quadcopter. For example, selective  
10 grazing behavior of pika can sometimes improve alpine grassland biodiversity (Harris  
11 et al. , 2016 and Zhang et al., 2016), which cannot be upscaled to a plot scale in aerial  
12 photos; 2) Quadcopter with a common camera cannot provide soil moisture  
13 information, while the burrowing activity of pika can improve infiltration and increase  
14 soil water content (Wilson and Smith, 2014). Therefore, both aerial surveying with  
15 quadcopter and ground sampling should be used together to investigate the role of  
16 pika comprehensively.

## 17 **5. Conclusions**

18 We up-scaled the quadrat-scale measurements of vegetation cover, biomass, soil  
19 carbon and nitrogen of 4 different surface types, i.e. vegetation and bald patches, new  
20 and old pika piles, to plot-scale using aerial photography. We then assessed the direct  
21 burying and grazing effects of pika. We concluded that both the direct effects were  
22 minor on different types of grasslands on two different geomorphologies. Bald  
23 patches had great impact on the reduction of biomass, soil carbon and nitrogen, but  
24 cannot be directly associated with pika activity at the current stage, which requires  
25 long-term repeated monitoring the changes of piles and burrow tunnels created by  
26 pika. Our study suggested that it is feasible and efficient to use quad-copter to monitor  
27 large amounts of patchy grassland plots and study the roles of pika.

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- 5

- 1 **Table 1.** The latitude, longitude, elevation of four different types of alpine grassland
- 2 and the dominant species on different surface types of each grassland.

Grassland Type	Latitude, Longitude, Elevation	Vegetation patch	Bald patch	Old pika pile
Alpine steppe (AS)	38°38'05.4" 98°06'41.7" 3768 m	<i>Stipa purpurea</i> , <i>Artemisia minor</i>	<i>Heteropappus hispidus</i> (Thunb.) Less., <i>Saussurea arenaria</i> Maxim.	<i>Potentilla bifurca</i> Linn., <i>Saussurea arenaria</i> Maxim.
Alpine steppe meadow (AStM)	38°28'34.6" 98°19'22.8" 3886 m	<i>Carex moorcroftii</i> , <i>Stipa purpurea</i>	<i>Ajania tenuifolia</i> , <i>Potentilla bifurca</i> Linn.	<i>Potentilla bifurca</i> Linn., <i>Saussurea arenaria</i> Maxim
Alpine meadow (AM)	38°25'15.2" 98°18'30.4" 3897 m	<i>Kobresia capillifolia</i> , <i>Carex moorcroftii</i>	<i>Glaux maritima</i> Linn., <i>Polygonum sibiricum</i> Laxm.	<i>Aster tataricus</i> L. f., <i>Polygonum sibiricum</i> Laxm.
Alpine swamp meadow (ASwM)	38°19'56.2" 98°13'35.1" 4043 m	<i>Kobresia pygmaea</i> , <i>Kobresia humilis</i>	<i>Carex atrofusca</i> Schkuh., <i>Glaux maritima</i> Linn.	<i>Polygonum sibiricum</i> Laxm., <i>Veronica didyma</i> Tenore.

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1 **Figure Legends**

2 **Figure 1.** a) Source region of Shule River Basin and its location in the Qinghai  
3 Tibetan Plateau; The rectangles indicate the locations of auto-piloted flight (each with  
4 12 way points), 1-4 indicate the location of field sampling on each type of grassland;  
5 b)-e) show aerial photographs of 4 types of alpine grasslands (AS: alpine steppe;  
6 AStM: alpine steppe meadow; AM: alpine meadow; and ASwM: alpine swamp  
7 meadow) investigated in this study. Each photograph covers  $\sim 35 \text{ m} \times 26 \text{ m}$  ground  
8 area.

9 **Figure 2.** Seasonal variations of fractional vegetation cover over May 19-August 30,  
10 2012 on alpine steppe, alpine steppe meadow and alpine meadow grasslands of Shule  
11 River Basin.

12 **Figure 3.** a) Diagram of ground sampling and aerial photographing; b) aerial  
13 photograph on one of 12 way points (solid black rectangles in a), each photo covers  
14  $\sim 35 \text{ m}$  by  $26 \text{ m}$  ground area, and was analyzed to have 4 parts, i.e. VP (vegetation  
15 patch), BP (bald patch), NP (new pika pile) and OP (old pika pile); c) ground  
16 sampling quadrat with  $50 \text{ cm}$  by  $50 \text{ cm}$  for vegetation cover, soil carbon and nitrogen  
17 (open rectangles in a) with red for vegetation patch (d), black for bald patch (e), green  
18 for new pika pile (f), and blue for old pika pile (g); and h) a circular plot with radius  
19 of  $14 \text{ m}$  for counting pika piles and pikas.

20 **Figure 4.** A photo taken on ground (left) and three examples (white rectangles) of  
21 green vegetation (green) classification (1-3 on the right).

22 **Figure 5.** An aerial photo and contours of vegetation patch (red curves, VP), bald  
23 patch (yellow curves, BP), new pika pile (red rectangles, NP), old pika pile (black  
24 rectangles, OP) and enlarged examples on the right for each type. Pink contour  
25 indicates gravel.

26 **Figure 6.** Green fractional vegetation cover (GFVC; %; a) soil organic carbon density  
27 (SOC;  $\text{kg}/\text{m}^2$ ; b) and total soil nitrogen density (TN;  $\text{kg}/\text{m}^2$ ; c) of vegetation patch  
28 (VP), new pika pile (NP), old pika pile (OP) and bald patch (BP) at a quadrat scale of  
29 four types of alpine grasslands (see Figure 1). Error bar indicates  $\pm$  standard deviation,

1 different letters above error bar indicate significant differences among surface types  
2 ( $p < 0.05$ ).

3 **Figure 7.** Area fraction (%) and number ( $\text{ha}^{-1}$ ) of vegetation patch (VP), new pika pile  
4 (NP), old pika pile (OP) and bald patch (BP) at a plot scale of four types of alpine  
5 grasslands (see Figure 1). Error bar indicates  $\pm$  standard deviation, different letters  
6 above error bar indicate significant differences between VP and BP or between NP  
7 and OP ( $p < 0.05$ ).

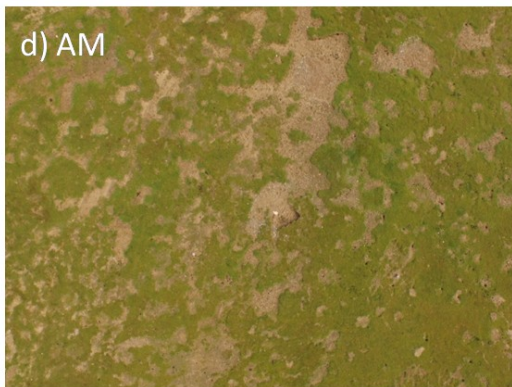
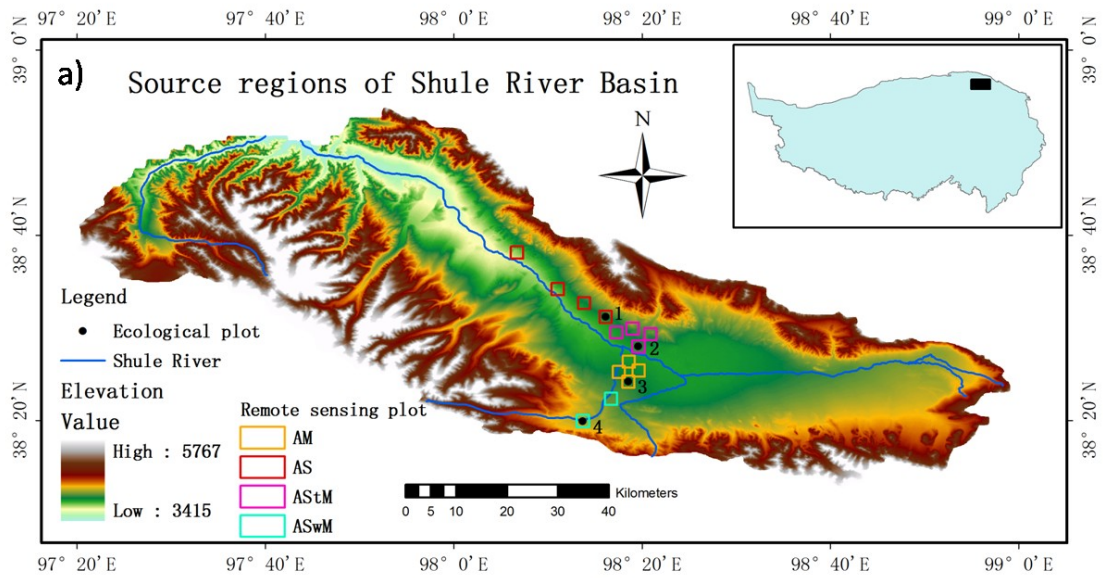
8 **Figure 8.** Effects of new pika pile (NP), old pika pile (OP) and bald patch (BP) on  
9 reduction of fractional vegetation cover a), soil carbon density (SOC); b) and total  
10 nitrogen (TN); c) on four types of alpine grasslands (see Figure 1). Error bar indicates  
11  $\pm$  standard deviation, different letters above error bar indicate significant differences  
12 among different surface types ( $p < 0.05$ ).

13 **Figure 9.** a) ratio between in-use burrow exits and total burrow exits ( $r_1$ ); b) ratio  
14 between number of pika and in-use burrow exits ( $r_2$ ); c) number of pikas ( $\text{ha}^{-1}$ ); and d)  
15 effects of pika grazing on above ground biomass (%) on four types of alpine  
16 grasslands (see Figure 1). Error bar indicates  $\pm$  standard deviation, different letters  
17 above error bar indicate significant differences among different grassland types  
18 ( $p < 0.05$ ).

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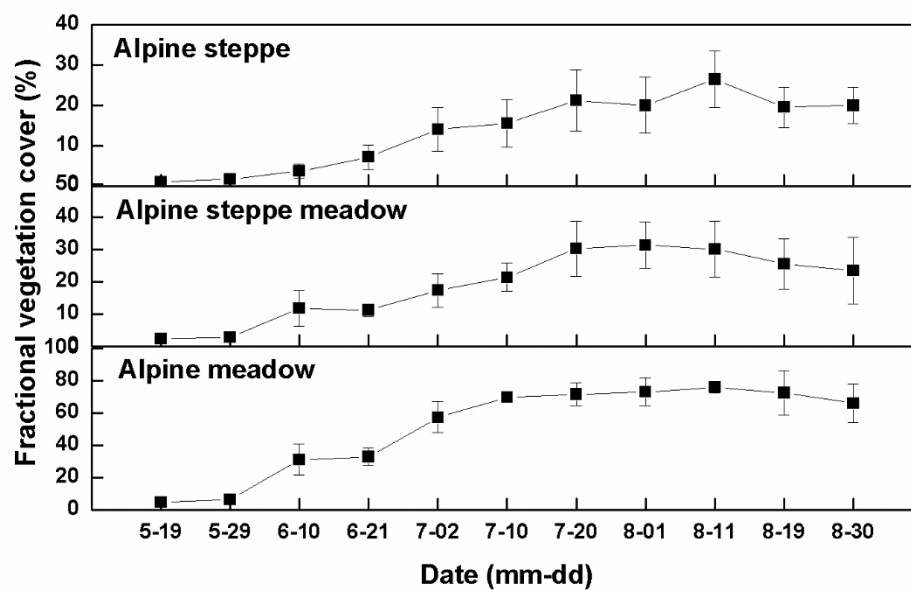
2 **Figure 1.**



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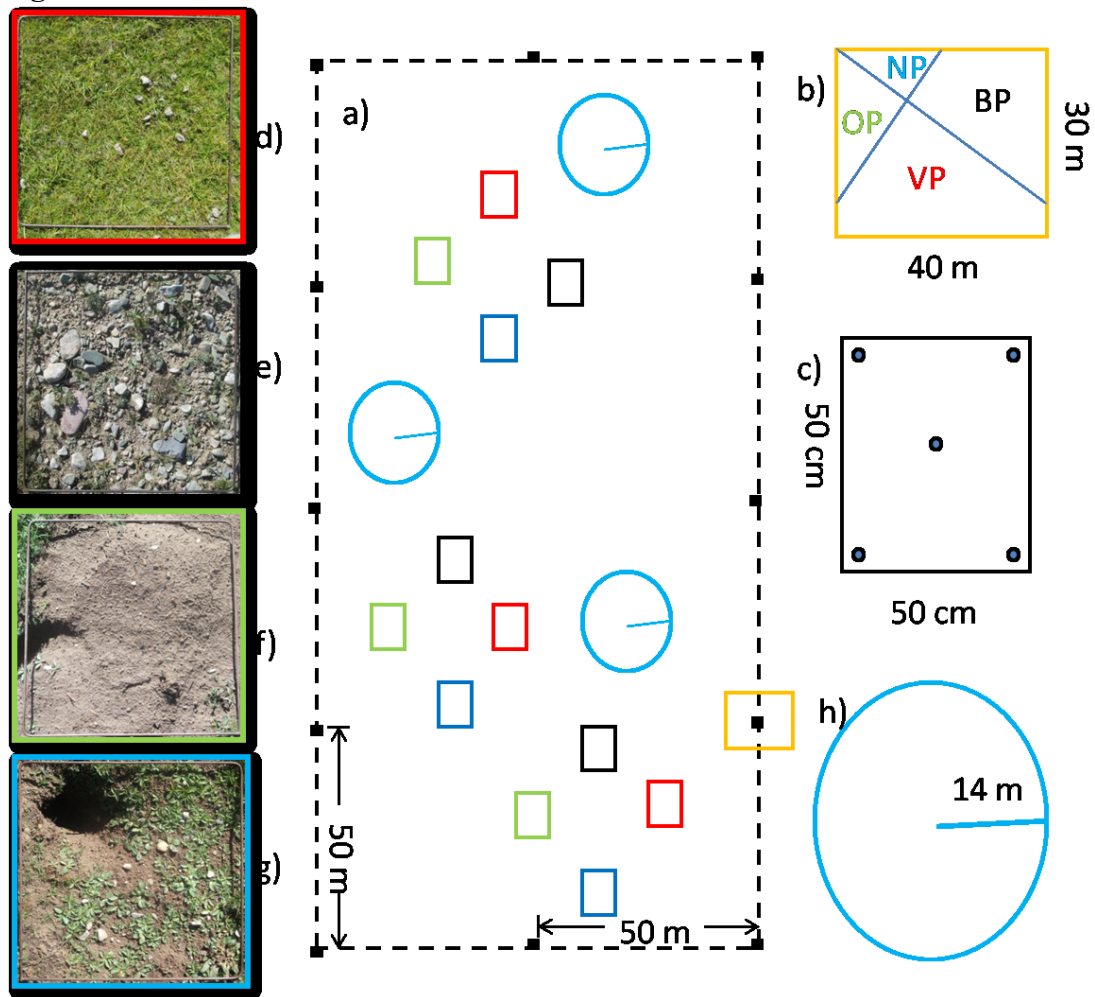
1 **Figure 2.**



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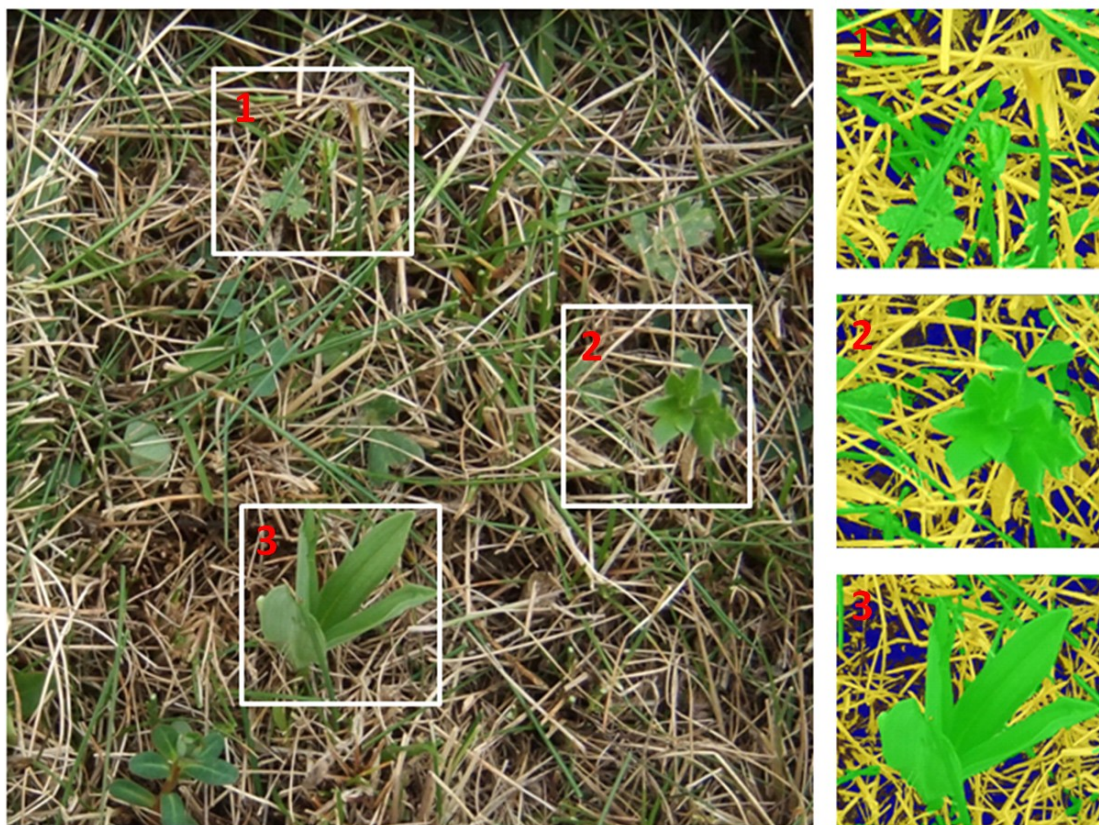
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1 **Figure 3.**



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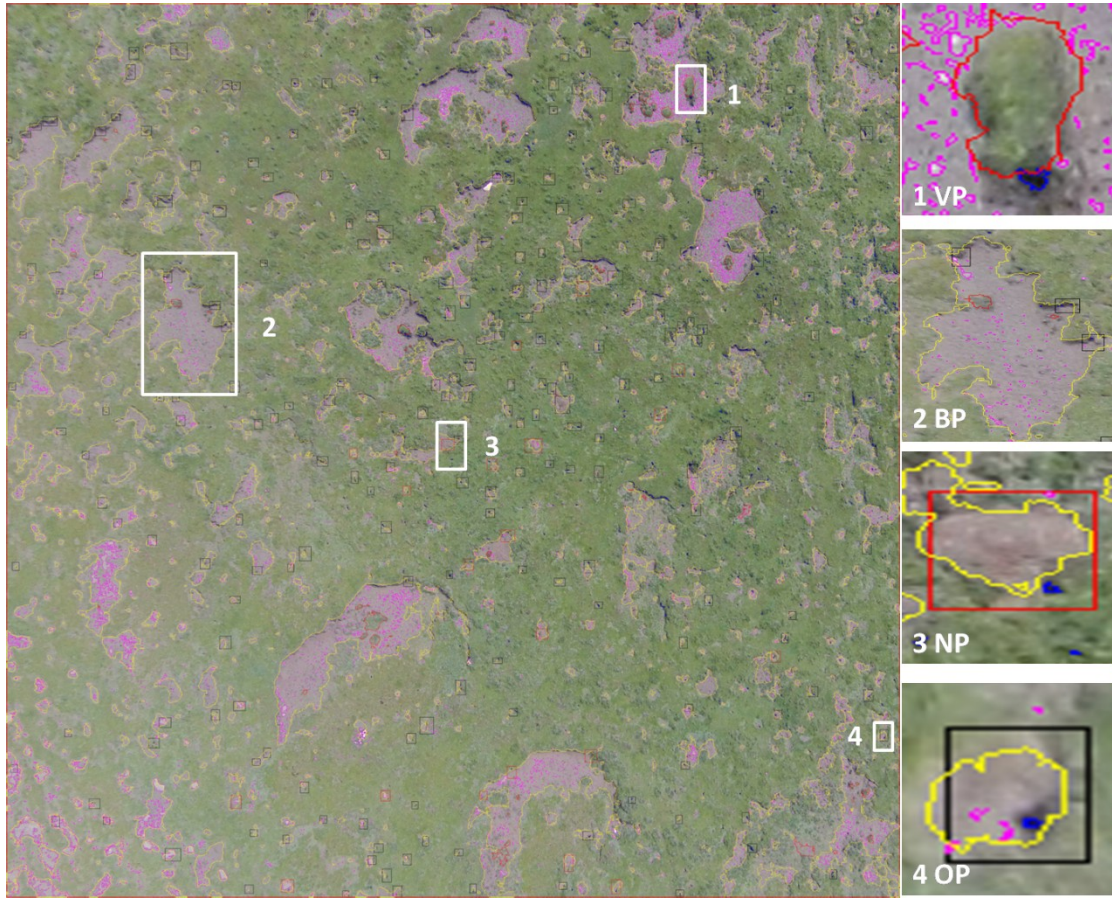
1 **Figure 4.**



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1 **Figure 5.**

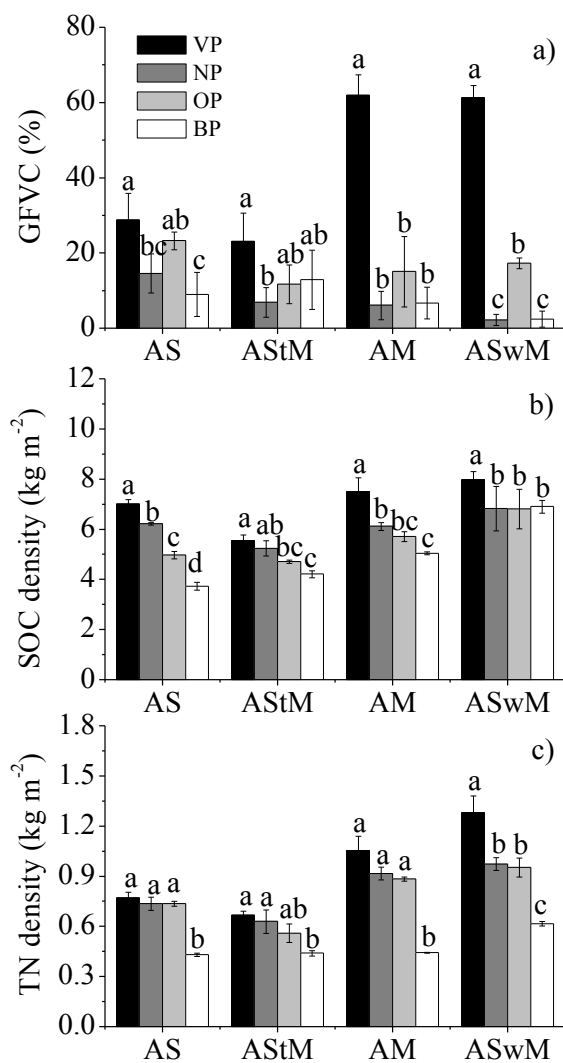


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1 **Figure 6.**



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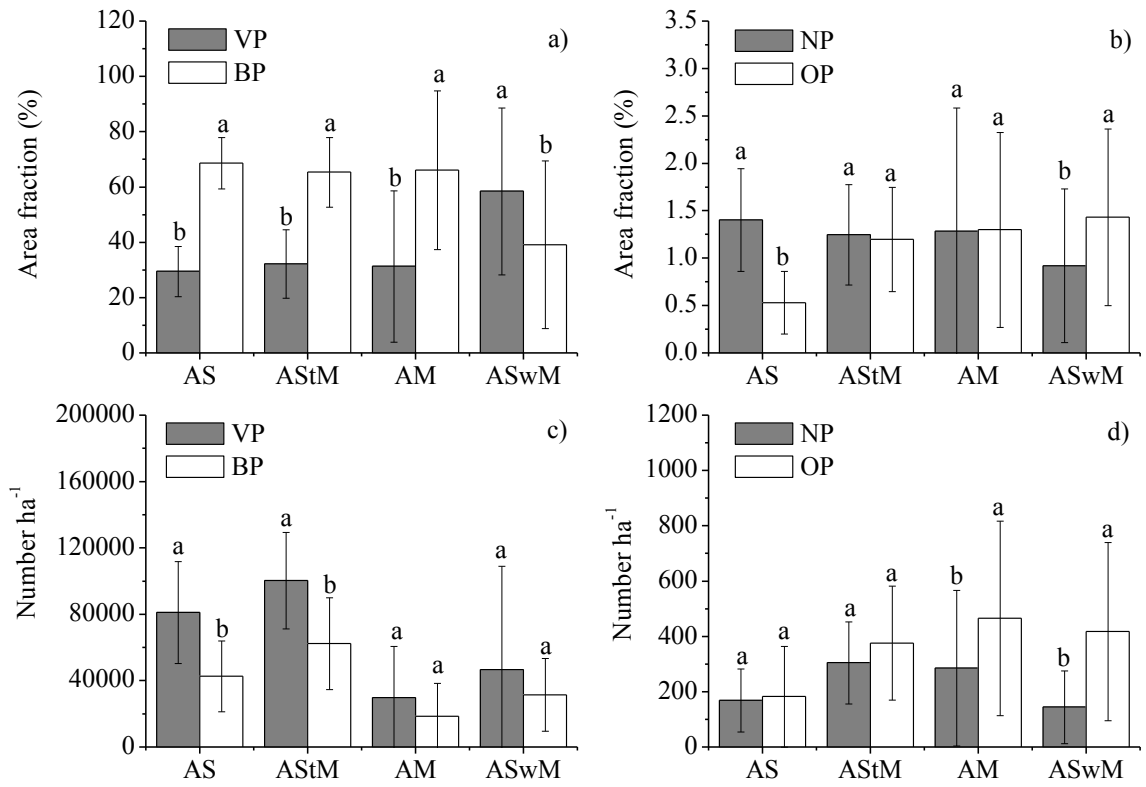
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1 **Figure 7.**

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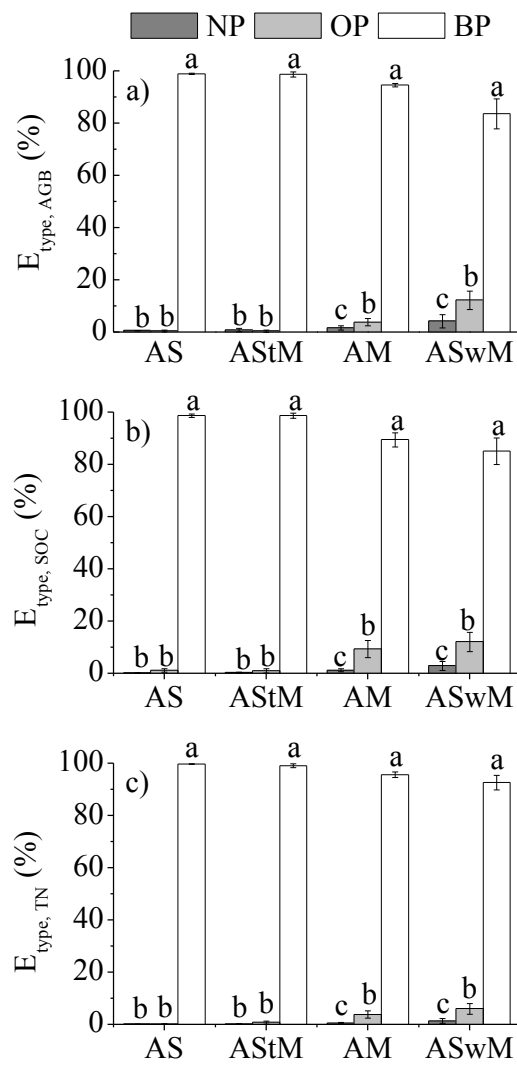
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1 **Figure 8.**



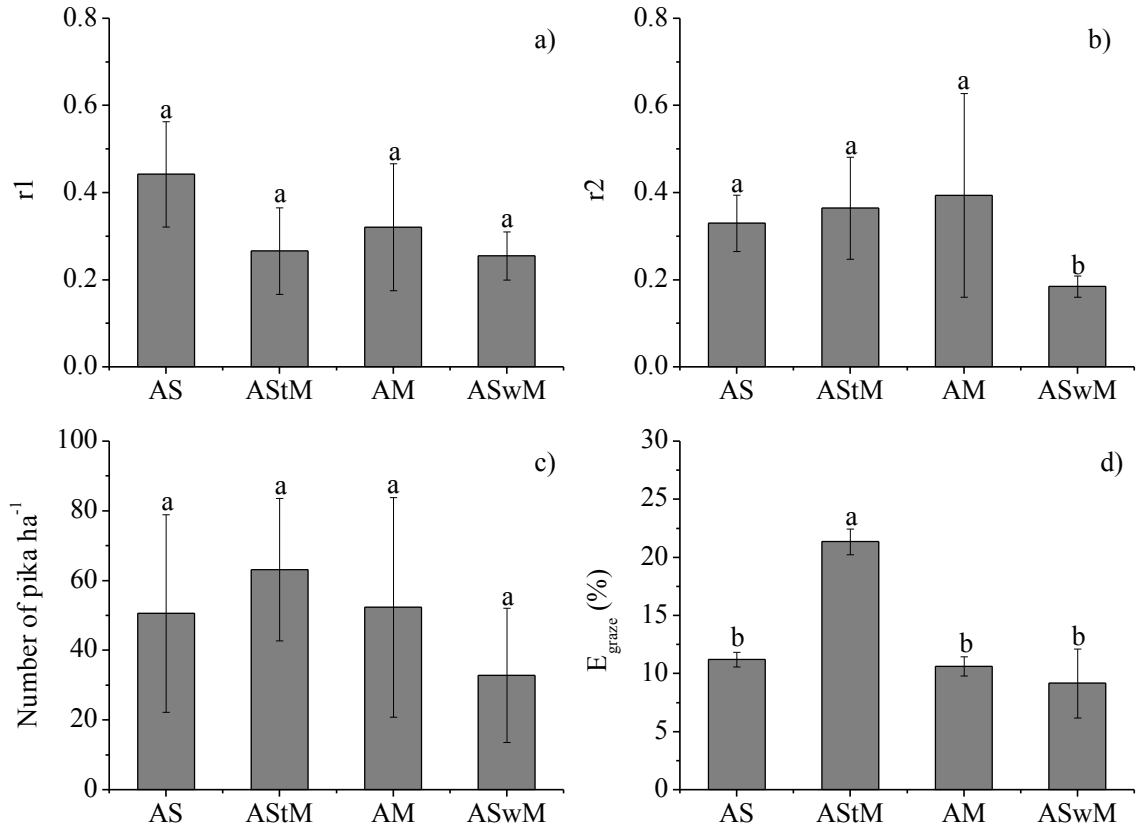
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1 **Figure 9.**

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