



1	Multi-gas and multi-source comparisons of six land use emission datasets
2	and AFOLU estimates in the Fifth Assessment Report
3	
4	Short title: AFOLU dataset comparisons
5	
6	Authors affiliation:
7	Rosa Maria Roman-Cuesta ^{1,2} *, Martin Herold ² , Mariana C. Rufino ¹ , Todd S. Rosenstock ³ ,
8	Richard A. Houghton ⁴ , Simone Rossi ⁵ , Klaus Butterbach-Bahl ^{6,7} , Stephen Ogle ⁸ , Benjamin
9	Poulter ⁹ , Louis Verchot ^{10,11} , Christopher Martius ¹ .
10	
11	¹ Center for International Forestry Research (CIFOR), P.O Box 0113 BOCBD, Bogor 16000,
12	Indonesia.
13	² Laboratory of Geo-Information Science and Remote Sensing - Wageningen University.
14	Droevendaalsesteeg 3, 6708PB. Wageningen. The Netherlands.
15	³ World Agroforestry Centre (ICRAF). PO Box 30677-00100, Nairobi. Kenya.
16	⁴ Woods Hole Reseach Center. 149 Woods Hole Road Falmouth, MA, 02540-1644, US.
17	⁵ Global Environmental Monitoring Unit, Institute for Environment and Sustainability,
18	European Commission, Joint Research Centre, TP. 440 21020 Ispra, Varese 21027, Italy,
19	⁶ International Livestock Research Institute (ILRI) P.O. Box 30709. Nairobi 00100, Kenya
20	⁷ Karlsruhe Institute of Technology, Institute of Meteorology and Climate Research (IMK-
21	IFU), Garmisch-Partenkirchen, Germany
22	⁸ Natural Resource Ecology Laboratory, Campus Delivery 1499, Colorado State University,
23	Fort Collins, Colorado 80523-1499, USA.





- ⁹ Ecosystem Dynamics Laboratory. Montana State University. P.O. Box 172000. Bozeman,
- 25 MT 59717-2000. USA.
- ¹⁰ International Center for Tropical Agriculture, Km17 Recta Cali-Palmira, Apartado Aéreo
- 27 6713, Cali, Colombia.
- ¹¹ Earth Institute Center for Environmental Sustainability, Columbia University, New York,
- 29 USA.
- * Corresponding author. Telephone:+31317485919, Fax: Email: rosa.roman@wur.nl

31

- 32 Keywords: AFOLU, Land use greenhouse gas emissions, Land Use Land Cover Change and
- 33 Forestry, LULUCF, mitigation, Fifth Assessment Report, gross emissions flux.

34

35 ABSTRACT

The Agriculture, Forestry and Other Land Use (AFOLU) sector contributes with ca. 20-25% 36 of global anthropogenic emissions (2010), making it a key component of any climate change 37 mitigation strategy. AFOLU estimates remain, however, highly uncertain, jeopardizing the 38 mitigation effectiveness of this sector. Global comparisons of AFOLU emissions have shown 39 40 divergences of up to 25%, urging for improved understanding on the reasons behind these differences. Here we compare a diversity of AFOLU emission datasets (e.g. FAOSTAT, 41 42 EDGAR, the newly developed AFOLU "Hotspots", "Houghton", "Baccini", and EPA) and estimates given in the Fifth Assessment Report, for the tropics (2000-2005), to identify 43 plausible explanations for the differences in: i) aggregated gross AFOLU emissions, and ii) 44 disaggregated emissions by sources, and by gases (CO₂, CH₄, N₂O). We also aim to iii) 45 46 identify countries with low agreement among AFOLU datasets, to navigate research efforts.





47 Aggregated gross emissions were similar for all databases for the AFOLU: 8.2 (5.5-12.2), 8.4 and 8.0 Pg CO₂e.yr⁻¹ (Hotspots, FAOSTAT and EDGAR respectively), Forests: 6.0 (3.8-10), 48 5.9, 5.9 and 5.4 PgCO₂e.yr⁻¹ (Hotspots, FAOSTAT, EDGAR, and Houghton), and 49 Agricultural sectors: 1.9 (1.5-2.5), 2.0, 2.1, and 2.0 PgCO₂e.yr⁻¹ (Hotspots, FAOSTAT, 50 EDGAR, and EPA). However, this agreement was lost when disaggregating by sources, 51 continents, and gases, particularly for the forest sector (fire leading the differences). 52 Agricultural emissions were more homogeneous, especially livestock, while croplands were 53 the most diverse. CO_2 showed the largest differences among datasets. Cropland soils and 54 enteric fermentation led the smaller N₂O and CH₄ differences. Disagreements are explained 55 by differences in conceptual frameworks (e.g. carbon-only vs multi-gas assessments, 56 definitions, land use versus land cover, etc), in methods (Tiers, scales, compliance with 57 Intergovernmental Panel on Climate Change (IPCC) guidelines, legacies, etc) and in 58 assumptions (e.g. carbon neutrality of certain emissions, instantaneous emissions release, etc) 59 that call for more complete and transparent documentation for all the available datasets. 60 Enhanced dialogue between the carbon (CO₂) and the AFOLU (multi-gas) communities is 61 needed to reduce discrepancies of land use estimates. 62

63

64 1. INTRODUCTION

Modelling studies suggest that to keep the global mean temperature increase to less than 2° C and to remain under 450 ppm of CO₂ by 2100, CO₂ emissions must be cut 41-72% below 2010 levels by 2050 (IPCC, 2014), and global emissions levels must be reduced to zero (a balance between sources and sinks) before 2070 and below zero, through removal processes, after that (Anderson, 2015; UNEP, 2015). To reach these ambitious goals, tremendously rapid improvements in energy efficiency and nearly a quadrupling of the share of zero and low carbon energy supply (e.g. renewables, nuclear energy, and carbon dioxide capture and





storage (CCS), including bioenergy (BECCS)) would be needed by 2050 (IPCC, 2014;
Friedlingstein et al., 2014; Anderson, 2015; UNEP, 2015). Since there is no scientific
evidence on the feasibility of CCS technologies (Anderson, 2015), renewables and the land
use sector are among the most plausible options (Canadell and Schulze, 2014). Optimistic
estimates suggest that the AFOLU sector (here indistinctively also called land use sector)
could contribute from 20 to 60% of the total cumulative abatement to 2030 including
bioenergy (Smith et al., 2014).

79

The AFOLU sector roughly contributes with a quarter (10-12 PgCO₂e.yr⁻¹) of the total 80 anthropogenic GHG emissions (50 PgCO2e.yr⁻¹) (Smith et al., 2014) through a few human 81 activities: deforestation, forest degradation, and agriculture including cropland soils, paddy 82 rice, and livestock (Smith et al., 2014). Despite the acknowledged importance of the 83 emissions from the land use sector in global mitigation strategies, assessing GHG emissions 84 and removals from this sector remains technically and conceptually challenging (Abad-Viñas 85 et al., 2014; Ciais et al., 2014). This challenge relates to an incomplete understanding of the 86 processes that control the emissions from the land use sector (Houghton et al., 2012), 87 especially post-disturbance dynamics (Frank et al., 2015; Poorter et al., 2016) and to various 88 sources of error that range from inconsistent definitions, methods, and technical capacities 89 (Romijn et al., 2012, 2015; Abad-Viñas et al., 2014), to special features of the land use sector 90 91 such as legacy and reversibility/non-permanence effects (Estrada et al., 2014), or to the 92 difficulty to separate anthropogenic from natural emissions (Estrada et al., 2014; Smith et al., 2014). As a result, the AFOLU emissions are the most uncertain of the all the sectors in the 93 global budget, reaching up to 50 percent of the emissions mean (Houghton et al., 2012; Smith 94 et al., 2014; Tubiello et al., 2015). This is important since uncertainties jeopardize the 95 effectiveness of the AFOLU sector to contribute to climate change mitigation. Thus, country 96





compliances to their mitigation targets are likely to be controversial when the uncertainty is
equal to or greater than the pledged emission reductions (Grassi et al., 2008; Pelletier et al.,

99 100 2015).

Currently, data on AFOLU emissions are available through national greenhouse gas 101 inventories, which are submitted to the United Nations Framework Convention on Climate 102 103 Change (UNFCCC), but these national estimates cannot be objectively compared due to differences in definitions, methods, and data completeness (Houghton et al., 2012; Abad-104 Viñas et al., 2014). More comparable AFOLU data are offered in global emission databases 105 such as EDGAR or FAOSTAT (Smith et al., 2014; Tubiello et al., 2015), or more sectorial 106 datasets such as the Houghton's Forestry and other Land Use (FOLU) data (Houghton et al., 107 2012), and the US Environmental Protection Agency non-CO₂ emissions for agriculture -108 including livestock (USEPA, 2013). While national inventories and global databases are 109 currently the best bottom up emissions data we count on, their utility to inform on what the 110 atmosphere receives has been contested. Late research shows disagreements between the 111 trends of reported emissions and atmospheric growth since 1990 for CO₂ (Francey et al., 112 2010, 2013a, 2013b), for CH_4 (Montzka et al., 2011), and for N_2O (Francey et al., 2013b). In 113 the case of CO₂, Francey et al. conclude that the differences between atmospheric and 114 emission trends for CO2 might be more related to under-reported emissions (~9 PgC for the 115 116 period 1994-2005), than to adjustments in the terrestrial sinks (e.g. increased CO_2 removals in 117 oceans and forests). On the other hand, global AFOLU databases suffer from inconsistencies that lead to global CO₂e emissions differences of up to 25% (2000-2009) (Tubiello et al., 118 2015): 12.7 vs 9.9 PgCO₂e.yr⁻¹ for EDGAR and FAOSTAT, respectively. These datasets also 119 disagreed in the contribution of the AFOLU sector to the total anthropogenic budget in 2010 120 (e.g. 21% and 24% for FAOSTAT vs EDGAR), and on the relative share of the emissions 121





- from agriculture versus FOLU since 2010. Thus, while EDGAR implies a relatively equal contribution (IPCC, 2014), FAOSTAT reports agricultural emissions being larger contributors to the total anthropogenic budget $(11.2\pm0.4\%)$ than forestry and other land uses $(10\pm1.2\%)$
- (Tubiello et al., 2015), with a steady growth trend of 1% since 2010.
- 126

Understanding the inconsistencies among AFOLU datasets is an urgent task since they 127 128 preclude our accurate understanding of land-atmosphere interactions, GHG effects on climate forcing and, consequently, the utility of modelling exercises and policies to mitigate climate 129 130 change (Houghton et al., 2012; Grace et al., 2014; Smith et al., 2014; Sitch et al., 2015; Tian et al., 2016). The land use sector plays a prominent role in the Paris Agreement (Art.5), with 131 many countries including it as mitigation targets in their Nationally Determined Contributions 132 (NDCs) (Grassi and Dentener, 2015; Richards et al., 2015; Streck, 2015). It is then urgent to 133 understand how much and why different AFOLU datasets differ in their emissions estimates, 134 so that we can better navigate countries' land-based mitigation efforts, and help to validate 135 their proposed claims under the UNFCCC. 136

137

Here we compare gross AFOLU emissions estimates for the tropics, for 2000-2005, from six
datasets: FAOSTAT, EDGAR, "Houghton", "Baccini", the US Environmental Protection
Agency data (EPA), and a recently produced, spatially explicit AFOLU dataset, that we will
hereon call "Hotspots" (Roman-Cuesta *et al., under review)*.

142). We aim to identify differences and plausible explanations behind: i) aggregated AFOLU,

143 FOLU and Agricultural gross emissions, ii) disaggregated contributions of the emission

sources for the different datasets, iii) disaggregated contribution of the different gases (CO₂,

145 CH₄, N₂O), and iv) national scale disagreements among datasets.





147 **2. METHODS**

148 2.1 Study area

Our study area covers the tropics and the subtropics, including the more temperate regions of 149 South America (33° N to 54° S, 161° E to 117° W). Land use change occurs nowhere more 150 rapidly than in this region (Poorter et al., 2016), so its study has global importance. We 151 selected the period 2000-2005 for being the common temporal range for all the datasets. This 152 period is not for the recent past but that does not affect the comparative nature of this 153 research. Our study area focuses at the country level and includes eighty countries, following 154 Harris et al., (2012). We ran the comparisons on gross emissions. Mitigation action can be 155 directed to reducing emissions by the sources, or to increasing the absorptions by the sinks, or 156 to both. While gross and net emissions are equally important, they offer different information 157 (Richter and Houghton, 2011; Houghton et al., 2012). Net land use emissions consider the 158 emissions by the sources and the removals by the sinks in a final emission balance where the 159 removals are discounted from the emissions, closer to what the atmosphere receives. Land use 160 sinks refer to any process that stores GHGs (e.g. forest growth, forest regrowth after 161 disturbances, organic matter stored in soils, etc) (Richter and Houghton, 2011). Gross 162 assessments can consider both the emissions produced by the sources (gross emissions) and 163 the removals absorbed by the sinks (gross removals), but they are not offered in a final 164 165 balance where the sinks are discounted from the sources. They are offered as separate fluxes, 166 instead. They are useful to navigate mitigation implementation since they offer direct 167 information on the sources and sinks that need to be acted upon through policies and measures to enhance and promote mitigation. However, lack of ground data makes the 168 assessment of the sinks much more difficult than the assessment of the sources (Houghton et 169 170 al., 2012; Grace et al., 2014; Brienen et al., 2015) with a particular gap on disturbed standing forests (Poorter et al., 2016). For these reasons, we here focus on gross emissions by the 171





- 172 sources, excluding gross sinks.
- 173

174 2.2 AFOLU datasets

Hotspots: this is a multi-gas (CO₂, CH₄, N₂O) spatially explicit (0.5°) database on gross 175 AFOLU emissions and associated uncertainties for the tropics for the period 2000-2005, at 176 Tier 2 and Tier 3 levels. It identifies Hotspots of AFOLU emissions to help prioritize 177 178 mitigation actions. It combines available published GHG datasets for the key sources of emissions in the AFOLU sector, as identified by the Fifth Assessment Report (AR5) of the 179 Intergovernmental Panel on Climate Change (Smith et al., 2014): deforestation, forest 180 degradation (fire, wood harvesting), crop soils, paddy rice, and livestock (enteric fermentation 181 and manure management). Tier 1 emission estimates of agricultural peatland decomposition 182 are also included. Forest emissions mainly report aboveground biomass (except fire that also 183 reports on soils). More detailed methodological information is available in Roman-Cuesta et 184 al., (under review). 185

186

FAOSTAT: covers agriculture, forestry and other land uses and their associated emissions of 187 CO₂, CH₄ and N₂O, following IPCC, 2006 Guidelines at Tier 1 (Tubiello et al., 2013, 2014). 188 Emissions are estimated for nearly 200 countries, annually, for the reference period of 1961– 189 2012 (agriculture) and 1990-2012 (FOLU), based on national activity data submitted by 190 191 countries and further collated by FAO. Projected emission data are available for 2030 and 192 2050. FAOSTAT includes estimates of emissions from biomass fires, peatland drainage and fires, based on geo-spatial information, as well as on forest carbon stock changes (both 193 emissions and removals) based on national-level FAO Forest Resources Assessment data 194 (FRA 2010). 195





EDGAR: The Emissions Database for Global Atmospheric Research (EDGAR) provides
global GHG emissions from multiple gases (CO₂, CH₄, N₂O, HFCs, PFCs and SF6) at 0.1°
and country levels. The EDGAR database covers all IPCC sectors (energy, industry, waste
management, and AFOLU), mostly applying IPCC 2006 guidelines for emission estimations
(EDGAR 2012). We downloaded the EDGAR's 4.2 Fast Track 2010 (FT 2010). FT 2010
emissions cover the period 2000-2010 in an annual basis, at the country level.

203

"Houghton": Houghton's bookkeeping model calculates the net and gross fluxes of carbon 204 (CO₂ only) between land and atmosphere that result from land management (Houghton, 1999, 205 2012; Houghton and Hackler, 2001; Houghton et al., 2012). The net estimate includes 206 emissions of CO₂ from deforestation, shifting cultivation, wood harvesting, wood debris 207 decay, biomass burning (for deforestation fires only, peatland fires were not included in our 208 version of their data), and soil organic matter from cultivated soils. It also includes sinks of 209 carbon in forests recovering from harvest and agricultural abandonment under shifting 210 cultivation. Unlike the other datasets, all pools are included: live vegetation, soil, slash 211 (woody debris produced during disturbance), and wood products. The model does, however, 212 not include forests that are not logged, cleared or cultivated. Rates of growth and 213 decomposition are ecosystem specific and do not vary in response to changes in climate, CO₂ 214 215 concentrations, or other elements of environmental change. Therefore, forests grow (and 216 wood decays) at the same rates in 1850 and 2015. Unlike other databases all carbon in the 217 ecosystem considered is accounted for: live vegetation, soil, slash (woody debris produced during disturbance), and wood products. We downloaded regional annual emissions from the 218 TRENDS (1850-2005) dataset for the tropics: Central and South (CS) America, tropical 219 Africa and South and South East Asia. Only net emissions were available. No spatially 220 disaggregated data were offered (e.g. countries). Houghton's data are, unlike all the other 221





datasets, net aggregated FOLU estimates, for CO₂-only.

223

"Baccini": These are gross FOLU tropical emissions that derive from Houghton's bookkeeping model and published by Baccini et al., (2012). Data are gross disaggregated emissions estimates for the period 2000-2010: deforestation (4.18 PgCO₂.yr⁻¹), wood harvesting (1.69 PgCO₂.yr⁻¹), biomass burning (2.86 PgCO₂.yr⁻¹), wood debris decay (3.04 PgCO₂.yr⁻¹). Baccini's estimates refer, however, to a tropical area slightly smaller than our study region.

230

The US Environmental Protection Agency (EPA): global non-CO₂ projected emissions for the 231 period 1990-2030 for the Agriculture, Energy, Industrial Processes and Waste sectors, for 232 more than twenty gases. EPA uses future net emissions projections of non-CO₂ GHGs as a 233 basis for understanding how future policy and short-term, cost-effective mitigation options 234 can affect these emissions. EPA follows the Global Emissions Report, which uses a 235 combination of country-prepared, publicly-available reports consistent with IPCC guidelines 236 and guidance (USEPA, 2013). When national emissions estimates were unavailable, EPA 237 produced its own non-CO₂ emissions using IPCC methodologies (e.g., international statistics 238 for activity data, and the default IPCC Tier 1 emission factors). Deviations to this 239 methodology are discussed in each of the source-specific methodology sections of USEPA 240 241 (2012). No FOLU estimates are included in this dataset. We downloaded agricultural emissions offered as 5-year intervals at country level, disaggregated by gas (N₂O and CH₄), 242 and by emission sources. 243

244

IPCC AR5: The AR5 is a synthesis report, not a repository of global data. However, new
AFOLU data are produced by the merging of peer-reviewed data such as Figures 11.2, 11.4,





- 247 11.5 and 11.8 in chapter 11 of the AR5 (Smith et al., 2014). We will contrast our six datasets
- 248 against the data from these newly produced figures.

249

Table 1 shows a summary of key similarities and differences of the assessed AFOLU datasets and the data from the AR5. The exact variables used for each database, are described in Table S1 in the supplementary material (SI). Datasets can be downloaded at the websites described in the reference section.

254

255 2.3 Estimating comparable gross AFOLU emissions for all datasets

256 We focus on human-induced gross emissions only, excluding fluxes from unmanaged land (e.g. natural wetlands). We focus on direct emissions excluding indirect emissions whenever 257 258 possible (e.g. nitrate leaching and surface runoff from croplands). Delayed fluxes (legacies) are important (e.g. underestimations of up to 62% of the total emissions when recent legacy 259 fluxes are excluded) (Houghton et al., 2012) but are frequently omitted in GHG assessments 260 that derive from remote sensing, such as some of the datasets used in this comparison (e.g. 261 deforestation emissions from Harris et al. (2012)). Wood harvesting emissions also excluded 262 legacy fluxes. We assumed instantaneous emissions of all carbon that is lost from the land 263 after human action (Tier 1, IPCC 2006) (e.g. deforested and harvested wood), with no 264 265 transboundary considerations (e.g. the emissions are assigned wherever the disturbance takes 266 place, particularly important for Harvested Wood Products). Life-cycle substitution effects 267 were neither considered for harvested wood (Peters et al., 2012). Some exceptions were allowed when data were already aggregated (e.g. for Houghton's and EPA's datasets we could 268 not exclude indirect emissions linked to forest decay and agriculture, respectively), or because 269 their legacy (past decay) estimates corresponded to an important source (e.g. EDGAR's post 270 burned decay and decomposition emissions represent deforestation) (Tubiello et al., 2015). To 271





- 272 facilitate comparisons, emissions estimates included the exact same emission sources:
- 273 deforestation, wood harvesting, fire, livestock (enteric fermentation + manure management),
- cropland soil emissions, rice emissions, emissions from drained histosols), for CO₂, CH₄, and
- N_2O . See Table S1 in SI to review the exact sources used in each database. Fire emissions do
- 276 not include CO₂ emissions from biomass burning in non-woody vegetation -savannas and
- 277 agriculture as they are assumed in equilibrium with annual regrowth processes (for CO_2
- 278 gases only) (IPCC 2003, 2006).
- 279

280 2.4 Correcting known differences among datasets estimates

Tubiello et al. (2015) identified four main differences that resulted in larger estimates for the 281 EDGAR data than for FAOSTAT, under the AFOLU estimates of the AR5 (Smith et al., 282 2014): 1. The inclusion of energy emissions under the agriculture budget, 2. Inclusion of 283 savannah burning, 3. Higher rice emissions due to the use of the IPCC 1996 guidelines instead 284 of the IPCC 2006 guidance, 4. FOLU's unresolved differences due to unclear metadata on 285 EDGAR's proxy for deforestation (post burned decay and decomposition). We have corrected 286 for the first two in our data comparison. No energy, and no CO₂ for savannah burning have 287 been included in the AFOLU estimates in any of our analyses. 288

289

290 2.5 Country emissions

We estimated the country-scale level of emissions agreement for the three most complete databases: FAOSTAT, EDGAR and Hotspots using the coefficient of variation among data, for AFOLU, forests (deforestation, fire and wood harvesting), crops (cropland soils, paddy rice) and livestock emissions. Percentiles were then used to separate between countries with high level of agreement ($\geq 75^{\text{th}}$ percentile), moderate agreement (50^{th} - 75^{th}), low agreement (25^{th} - 50^{th}), and very low agreement ($\leq 25^{\text{th}}$).





297

298 3. RESULTS AND DISCUSSION

299 3.1 Aggregated AFOLU, FOLU and Agricultural emissions

We found good agreement among datasets for the aggregated tropical scales with AFOLU 300 values of 8.0 (5.5-12.2) (5th-95th percentiles), 8.4 and 8.0 PgCO₂e.yr⁻¹ (for the Hotspots, 301 FAOSTAT and EDGAR, respectively). FOLU (deforestation and forest degradation) 302 contributed with 6.0 (3.8-10), 5.9, 5.9 and 5.4 PgCO₂e.yr⁻¹ for the Hotspots, FAOSTAT, 303 EDGAR, and Houghton datasets respectively. Agriculture (livestock, cropland soils and rice 304 emissions) reached 1.9 (1.5-2.5), 2.5, 2.1, and 2.0 PgCO₂e.yr⁻¹ for the Hotspots, FAOSTAT, 305 EDGAR, and EPA datasets respectively (Figure 1, Table 2). Forest emissions represented 306 \geq 70% of the tropical AFOLU gross mean annual budget for 2000-2005 (our Hotspots 307 database and Houghton showing the highest and the lowest estimates), and agriculture 308 represented the remaining 25-30% AFOLU emissions (FAOSTAT and Hotspots showing the 309 highest and the lowest values). Houghton's FOLU value (5.4 PgCO₂,yr⁻¹) is a net estimate that 310 includes carbon dynamics associated to forest land use changes, and forest removals from 311 areas under logging and shifting cultivation and it is, as expected, lower than the forest gross 312 emissions. Its value for the tropics was, however, higher than the net FOLU value used in the 313 IPCC AR5 (4.03 PgCO₂e.yr⁻¹ for 2000-2009) (Houghton et al. 2012). Since boreal and 314 temperate forest sinks are reported to be quasi-neutral (Houghton et al., 2012), these 315 316 differences are unclear. There is a variety of Houghton' net FOLU estimates in current bibliography (e.g. 4.03 PgCO₂e.yr⁻¹ for 2000-2009 in Smith et al. (2012), 4.9 for 2000 and 4.2 317 for 2010 (Tubiello et al., 2015) that likely correspond to different updates of the same dataset, 318 but create confusion and would call for verified official values that could be consistently used. 319





The IPCC AR5 offers a FOLU gross value for the tropics of ca. 8.4 PgCO₂.yr⁻¹ (2000-2007) 321 (Fig 11.8 in AR5, Smith et al., 2014) (Fig S1, SI) which corresponds to Baccini's estimates 322 using Houghton's bookkeeping model. This value is in the upper range of our gross FOLU 323 emissions: 6 (3.8-10) PgCO₂e.yr⁻¹ (2000-2005), and higher than the mean gross FOLU 324 emissions from all the other datasets (approx. 6 PgCO₂e.yr⁻¹) (Table 2). The time periods are 325 not identical and we do not compare the same gases (e.g. the bookkeeping model focuses on 326 CO₂ only, while we run a multi-gas assessment). However, the differences mainly relate to 327 unreported choices behind the inclusion/exclusion of emission sources and the description of 328 their methods, in the AR5. Thus, the 8.4 PgCO₂.yr⁻¹ gross estimate does not include fire, and 329 has larger contributions from shifting cultivation (2.35 PgCO₂.yr⁻¹) and wood-harvesting (2.49 330 PgCO₂.yr⁻¹), than the deforestation and wood-harvesting emissions in our selected datasets 331 (Figure 2). Numbers used in Figure 11.8 also exclude other gross emissions offered in Baccini 332 et al. (2012), which is the citation used in Fig. 11.8. Explicit, complete, and transparent 333 documentation is encouraged for the next AFOLU figures in the IPCC Assessment Reports. 334 Another consideration of AFOLU estimates in the Assessment Reports relates to the use of 335 the bookkeeping model to estimate land use, land use change and forest (LULUCF) 336 emissions. As useful as this model is, its framework does not follow the IPCC AFOLU 337 guidelines (IPCC, 2006), particularly regarding the concept of managed land. Thus, forests 338 that are on managed land but are not suffering from direct human activities are considered 339 340 carbon neutral (Houghton pers. comm.). Partly because of that, the net emission estimates of LULUCF from Houghton et al., (2012) used in the AR5 (4.03 PgCO₂.yr⁻¹, 2010) contrast with 341 the LULUCF estimates produced by country reports submitted to the UNFCCC for the same 342 year, which are close to zero (Grassi and Dentener, 2015). The use of IPCC compliant models 343 for the IPCC Assessment Reports, or/and some documentation that warned about these 344 inconsistencies, would be useful in future assessments. 345





346

347 Emissions in the agricultural sector are mostly net, since sink effects in the soils are small and frequently temporal (USEPA, 2013; Smith et al., 2014). Comparisons against global 348 agricultural emissions show that for the year 2000, global estimates more than doubled our 349 values (e.g. 5 and 5.5 PgCO₂e.yr⁻¹ vs ca. 2 PgCO₂e.yr⁻¹ in all datasets) (Tubiello et al., 2015) 350 (Table 2), suggesting larger contributions of agricultural emissions from non-tropical 351 352 countries. Unexplained methodological differences such as the inclusion or not of indirect emissions and the lack of an exhaustive list of the variables included in the agricultural 353 emissions, difficult further comparisons. 354

355

356 3.2 Disaggregated gross emissions: contributions of the emission sources

While the gross aggregated estimates suggested a good level of agreement among datasets 357 (Figure 1), differences occur when comparing the emissions sources leading the AFOLU 358 budgets (Figure 2). The FOLU sector showed the largest differences, mainly due to the 359 estimates of forest degradation, and particularly fire (FAOSTAT and EDGAR showed the 360 lowest and highest values). The forest sector is the most uncertain term in the AFOLU 361 emissions due to both uncertainties in areas affected by land use changes and other 362 disturbances, and by uncertain forest carbon densities (Houghton et al., 2012; Grace et al., 363 2014; Smith et al., 2014). Agricultural sources were more homogeneous (ca. 2 PgCO₂e.yr⁻¹ 364 365 for all datasets) (Figure 1), with livestock and cropland soil emissions as the most and least 366 similar (Figure 2). The homogeneity in livestock emissions was expected since most datasets use common statistics (FAO) to derive herd numbers per country. 367

368

369 3.2.1 Deforestation

Deforestation emissions were 2.9 (1.0-10.1), 3.7, and 2.5 and 4.2 PgCO₂.yr⁻¹ (Hotspots,





371 FAOSTAT, EDGAR, and Baccini, respectively), with Baccini and EDGAR showing the 372 highest and the lowest values. Their values represent, however, very different scenarios: gross deforestation for the Hotspots and Baccini datasets, net deforestation for FAOSTAT, and 373 forest fire and post-burn decay for EDGAR (Table 3). The Hotspots (Harris et al., 2012) and 374 Baccini et al., (2012) datasets offer gross deforestation estimates that rely on Hansen et al., 375 (2010)'s forest cover loss areas. However, they report different tropical emissions (0.81 and 376 1.14 PgC.yr⁻¹) because they use different carbon density maps: Harris et al.(2012) rely on 377 Saatchi et al.(2011) and Baccini rely on Baccini et al.(2010). EDGAR does not provide a 378 category for deforestation, and their Forest Fire and Decay category (5F) (Table 3, and Table 379 S1 in SI) is used as a proxy for deforestation (Tubiello et al., 2015). Such an approximation 380 leads to underestimations since not all carbon losses from deforestation are necessarily 381 associated with the use of fire (Tubiello et al., 2015). In spite of being net emissions, the 382 deforestation estimates for FAOSTAT were higher than the gross estimates of Hotspots and 383 Baccini. This is partly due to FAOSTAT's inclusion of fire emissions from humid tropical 384 forests (see section 3.2.3), which the other datasets did not. Baccini's larger estimates of gross 385 deforestation included more carbon pools than the other datasets (e.g. soil, CWD, litter). 386 Baccini et al. (2012) reported that their estimated gross and net emissions from tropical 387 deforestation were the same value (4.2 Pg $CO_2 yr^{-1}$). The difference with Houghton's net 388 emissions (5.4 PgCO₂.yr⁻¹) (Figure 2) corresponds, then, to non-offset carbon emissions from 389 390 other land uses and activities included in the bookeeping model: degradation by logging and 391 shifting cultivation, decomposition and decay, and cultivated soils. Houghton's tropical net emissions for 2000-2005 are high, but lower than Houghton's reported net estimates in the 392 80's (7 PgCO₂.yr⁻¹) (Houghton, 1999). 393

394





396 *3.2.2 Forest degradation*

397 Forest degradation can be defined in many ways (Simula, 2009), but no single operational 398 definition has been agreed upon by the international community (Herold et al., 2011a). It typically refers to a sustained human-induced loss of carbon stocks within forest land that 399 remains forest land. In this study, similarly to Federici et al., (2015), we consider degradation 400 any annual removal of carbon stocks that does not account for deforestation, without temporal 401 402 scale considerations (e.g. time needed for disturbance recovery, or time to guarantee a sustained reduction of the biomass). We assessed two major degradation sources: wood 403 harvesting and fire. Soil degradation is poorly captured in many datasets, and mainly focuses 404 on fire in equatorial Asian peatland forests and drained peatlands (Hooijer et al., 2010). 405 Better understanding of the processes and emissions behind forest degradation, would be key 406 for climate mitigation efforts not only because forest degradation is a wide spread 407 phenomenon (e.g. affects much larger areas than deforestation (Herold et al., 2011b)) but also 408 because the lack of knowledge of net carbon effects frequently results in assumptions of 409 carbon neutrality of the affected standing forests, particularly for fire (Houghton et al., 2012; 410 Le Quéré et al., 2014), which is likely leading to an underestimation of forest and AFOLU 411 emissions. 412

413

Gross emissions from forest degradation were larger than deforestation for the Hotspots, EDGAR and Baccini's datasets, with degradation-to-deforestation ratios of 108%, 120%, and 128%, respectively. FAOSTAT had degradation emissions of 60% of the deforestation, partly due to its anomalously low fire contribution (see next section). Houghton et al., (2012) pointed out that global FOLU net fluxes were led by deforestation with a smaller fraction attributable to forest degradation, while the opposite was true for gross emissions (degradation being 267% of deforestation emissions). This large ratio relates to their inclusion of shifting





421 cultivation under degradation. This is a definition issue, which would not fit the definition of
422 degradation chosen in this study, where a complete forest cover loss would represent
423 deforestation and not degradation.

424

425 3.2.3 Fire

Fire led the gross forest degradation emissions in the tropics in 2000-2005 (Figure 2): 2 (1.1-426 2.7), 0.2, 3.4, 2.9 PgCO₂e.yr⁻¹ for the Hotspots, FAOSTAT, EDGAR, and Baccini datasets, 427 respectively) (Figure 2). Our estimates are conservative compared to Van der Werf et al., 428 (2010)'s global emissions of 7.7 PgCO₂e.yr⁻¹ for 2002-2007, due to our removal of CO₂ from 429 deforestation fires (to avoid double counting with deforestation emissions), to the exclusion of 430 fires in grasslands and agricultural residues, and to our smaller study area. FAOSTAT and 431 EDGAR had the lowest and the highest fire values. FAOSTAT lowest values relate to 432 omissions that are currently in the process of being corrected (Rossi pers. comm.): 1. the 433 exclusion of CO₂ from fire in humid tropical forests and other forests (Table 3, Table S1), 434 which FAOSTAT relocated as net forest conversion emissions, partly explaining their larger 435 deforestation values, and 2. The use of default parameters for fuel in peats from the IPCC 436 2006 Guidelines instead of the new IPCC Wetland supplement which offer considerable 437 higher values (Rossi et al., 2016). Moreover, FAOSTAT uses GFED3.0-burned area (Giglio 438 et al., 2010) in their estimates while the other datasets use GFED3.0-emissions (Van der Werf 439 440 et al., 2010). EDGAR fire emissions were the largest most likely because they included decay. Their dataset considers some undefined "forest fires" (5A) and "wetland/peatland fires and 441 decay" (5D) (Table 3; Table S1 in SI). Peatland decay probably explains EDGAR's larger 442 emissions in Asia, while we assume that EDGAR's highest fire emissions for CS America 443 might respond to deforestation fires which were not included in the Hotspots to avoid double 444 counting with deforestation, and relocated in FAOSTAT to deforestation emissions (Figure 3, 445





Table 3). Our Hotspots dataset showed higher gross fire emissions for Africa due to the inclusion of woodland fire, which EDGAR and FAOSTAT probably excluded. Baccini et al., (2012)'s fire emissions: 2.9 PgCO₂e.yr⁻¹ (2000-2010) derive from Houghton's bookkeeping but it is unclear how these emissions were estimated.

450 In spite of the importance of fire as a degradation source, this variable is frequently 451 incompletely included, either through unaccounted gases (e.g. CH₄ and N₂O are excluded in the carbon community but their omission represent 17-34% of the gross CO_2 fire emissions) 452 (Valentini et al., 2014; Roman-Cuesta et al., under review), or to unaccounted components 453 (e.g fires in tropical temperate forests such as conifers or dry forests such as woodlands, are 454 frequently excluded) (Houghton et al., 2012). Unaccounted fire emissions also derive from 455 methodological choices (e.g. only inter-annual fire anomalies being considered) (Le Quéré et 456 al., 2014), from poor satellite observations such as understory fires in humid closed canopy 457 forests) (Alencar et al., 2006; 2012, Morton et al., 2013), or satellite fire omissions in certain 458 regions (e.g. high Andean fires) (Bradley and Millington, 2006; Oliveras et al., 2014). Other 459 omissions relate to the current exclusion of non-Asian peatland fires (e.g American tropical 460 montane cloud forest peatland fires) (Asbjornsen et al., 2005; Roman-Cuesta et al., 2011; 461 Oliveras et al., 2013; Turetsky et al., 2015). 462

463

Fire suffers, moreover, from a series of assumptions that do not apply so easily to other types of degradation: 1. Assuming a non-human nature of the fires (deforestation fire *vs* wildfires), which in tropical areas contrasts with multiple citations referring to the 90% human causality of fires (Cochrane et al., 1999; Roman-Cuesta et al., 2003; Alencar et al., 2006; Van der Werf et al., 2010). 2.Assuming *force-majeure* conditions that lead to non-controllable fires due to extreme climate conditions, which frequently results in incomplete assessment and reporting of emissions. This assumption contrasts with research on how human activities have seriously





471 increased fire risk and spread in the tropics (Uhl and Kauffman, 1990; Laurance and Williamson, 2001; Roman-Cuesta et al., 2003; Hooijer et al., 2010), and clearly expose how 472 most of the fires in the humid tropics would not occur in the absence of human influences 473 over the landscape (; Roman-Cuesta et al., 2003). 3. Assuming carbon neutrality and full 474 biomass recovery after fire in standing forests. This is a generous assumption that contrasts 475 with numerous studies on tropical forest die-back following fire events in non-fire adapted 476 477 humid tropical forests (Cochrane et al., 1999; Barlow et al., 2008; Roman-Cuesta et al., 2011; Brando et al., 2012; Oliveras et al., 2013; Balch et al., 2015). All these phenomena casts 478 doubts on the robustness of these assumptions and call for a much more comprehensive 479 inclusion of fire emissions into forest degradation budgets. 480

481

482 3.2.4 Wood harvesting

There is not a unique way to estimate wood harvesting emissions as exposed in the guidelines 483 for harvested wood products of the IPCC (IPCC 2006). Assumptions regarding the final use 484 of the wood products, decay times, substitution effects, international destination of the 485 products and time needed for forests to recover their lost wood, can fully change the emission 486 budgets. In out study, wood harvesting emissions were 1.2 (0.7-1.6), 2.0, 1.7 PgCO₂.yr⁻¹ for 487 the Hotspots, FAOSTAT and Baccini data, respectively (Tables 3, Table S1 in SI). Harvested 488 wood products derive from FAO's country reports (e.g. FAOSTAT forest products). All 489 490 datasets included fuel wood and industrial roundwood (Tables 3, Table S1), EDGAR 491 excluded fuelwood from the AFOLU budget and placed it instead into the energy budget (EDGAR, 2012), which explains its absence in Figure 2. Wood harvesting emissions were 492 larger in FAOSTAT than in the Hotspot data (Figure 2) partly due to the inclusion of some 493 extra categories of fuels (e.g. charcoal and residues) that were not included in the Hotspot 494 database (Table 3, Table S1 in SI). Charcoal represents 26% of the total wood-harvesting 495





emissions in FAOSTAT. Differences on wood harvesting affected more Asia and CS America
(where our Hotspot data were half of FAOSTAT's), whilst Africa presented almost identical
values (Figure 3), reasons for these continental differences are unclear. Baccini's high
emissions on wood harvesting could partly relate to their inclusion of extra biomass due to
felling damages (e.g. 20-67% of the AGB is damaged, and 20% is left dead in BGB)
(Houghton, 1999).

502

503 *3.2.5 Livestock*

Livestock emissions were the most homogeneous among the emissions sources (Figure 2) 504 with estimates of 1.2 (0.8-1.5), 1.1, 1.2, 1.1 PgCO₂e.yr⁻¹ for the Hotspots, FAOSTAT, 505 EDGAR and EPA respectively, in range with the estimates in the AR5 (Fig 11.5 in Smith et 506 al., 2014). Values were similar in spite of deriving from different Tiers (e.g. Tier 3 for Herrero 507 et al., (2013), Tier 1 for FAOSTAT and EDGAR. EPA used Tier 3 but for incomplete data 508 series, otherwise Tier 1 was applied (USEPA, 2013)). All datasets included enteric 509 fermentation (CH₄) and manure management (N₂O, CH₄). All of them relied on FAO data for 510 livestock heads, although they used different years (e.g. 2000 for Herrero et al., (2013) data in 511 the Hotspots, and 2007-2010 for EDGAR). From a continental perspective, FAOSTAT and 512 EDGAR estimates were the closest while the Hotspots and EPA's were less similar. The 513 Hotspots showed higher emissions for Africa and Asia and lower for CS America, than the 514 515 other three datasets. Divergences likely relate to different Tiers. CS America and Asia showed 516 the highest values, with Africa following closely (Figure 3), similar to what is reported in the AR5 (Smith et al., 2014). Globally, livestock is the largest source of CH₄ emissions, with 517 three-fourth of the emissions coming from developing countries, particularly Asia (USEPA, 518 2013, Tubiello et al., 2014). Three out of the top-5 emitting countries are in the tropics: 519





520 Pakistan, India and Brazil (USEPA, 2013) and while Asia hosts the largest livestock

emissions, the fastest growing trends in 2011 correspond to Africa (Tubiello et al., 2014).

522

523 3.2.6 Cropland emissions

The estimates of cropland emissions reached values of 0.18 (0.16-0.19), 0.56, 0.6 and 0.64 524 PgCO₂e.yr⁻¹ for the Hotspots, FAO, EDGAR and EPA datasets respectively, for N₂O and CO₂ 525 526 from changes in soil organic carbon content. Cropland soil emissions (N₂O and soil organic carbon stocks (CO₂) heavily depend on land management practices (e.g. tillage, fertilization 527 and irrigation practices) and climate (Crowther et al., 2015). We chose exactly the same land 528 practices in all datasets to allow comparisons (Table 3,S1 in SI). For this reason, we excluded 529 N₂O emissions from grassland soils, drainage of organic soils, and restoration of degraded 530 lands (Table 3). This restrictions resulted in lower emissions than those estimated for cropland 531 soils in the AR5 (Fig. 11.5 in Smith et al., 2014). The Hotspots and EPA showed the lowest 532 and the highest estimates (Figures 2,3). With the exception of the Hotspots, the other datasets 533 agreed well at the tropical scale, with FAOSTAT and EDGAR being almost identical, also at 534 continental scales. EPA disagreed more than the other datasets at the continental scales, with 535 underestimations for Asia, probably related to the parameterization of their emission model. 536 All three datasets used FAO's activity data, and for EDGAR and FAOSTAT the same 537 emission factors must have been used. The Hotspot showed anomalously low emissions partly 538 539 because it only included six major crop types (maize, soya, sorghum, wheat, barley, millet) 540 for which the emission model (DAYCENT) counted on reliable parametrization (Ogle pers. comm). Emissions from other important crops in the tropics (e.g. sugar cane, tobacco, tea, etc) 541 were excluded, as well as emissions from croplands in organic soils, due to model constraints. 542

543

544 3.2.7 Peatland drainage for agriculture





545 The disaggregation of cropland soil emissions from drained peatands shows large omissions for drained peatlands in the Hotspots database. Emissions were one order of magnitude lower 546 (28 TgCO2e.yr⁻¹) than FAOSTAT (ca. 500 TgCO2e.yr⁻¹) and than the peatland drainage 547 emissions reported in Asia alone by Hooijer et al. (2010) (355-855 TgCO₂e.yr⁻¹) Our lower 548 values relate to much smaller agricultural areas with histosols (0.4 mill ha) than those reported 549 by FAOSTAT for the same countries (7mill ha). Differences relate to the subset of the final 550 551 areas to only those that respond to the six types of crops selected by Ogle et al. (2013) (maize, wheat, sorghum, soya beans, millet and barley), to the unmatching spatial scales of the 552 overlapping layers (1km for histosols and 50km for croplands) which result in 553 underestimations of the final area, and to the use of an Emission Factor of 20 MgC.ha⁻¹ for the 554 Hotspots data, while FAOSTAT used 14.64 MgC.ha⁻¹. 555

556

557 *3.2.8 Paddy rice*

When paddy fields are flooded, decomposition of organic material gradually depletes the 558 oxygen present in the soil and floodwater, causing anaerobic conditions in the soil that favour 559 methanogenic bacteria that produce CH₄. Some of this CH₄ is dissolved in the floodwater, but 560 the remainder is released to the atmosphere, primarily through the rice plants themselves. Net 561 emission estimates for paddy rice were 0.55 (0.4-0.833), 0.33, 0.37, 0.30 $PgCO_{2}e.yr^{-1}$ for the 562 Hotspots, FAOSTAT, EDGAR and EPA datasets, respectively. The Hotspots showed the 563 564 highest emissions (Figure 2), but only in Asia (Figures 3). Part of the reason behind these differences refers to the final gases estimated in Li et al., (2013)'s which included CH₄, N₂O 565 566 and SOC (CO₂) (Table 3, S1), while the others only focused on CH₄. In Li et al., (2013)'s estimates, N₂O were 48% of the CH₄ emissions, explaining the doubled emissions in our 567 database. SOC was a sink, with -0.076 PgCO₂.yr⁻¹. 568





570 Based on the above, Table 4 offers the least reliable emission sources of each dataset.

571

572 3.3 Differences in the relative contribution of greenhouse gases (CO₂, CH₄, N₂O)

573 GHG emissions (CO₂, CH₄, N₂O) showed good agreement at the sectoral level (FOLU and 574 agriculture) (Figure 5), that disappeared at the disaggregated level (Figure 6). CO₂ showed the 575 largest disagreements among datasets and gases, led by forests emissions and particularly fire. 576 SOC accumulation was reported in the Hotspots data (Li et al., 2013) but it is uncertain if it is 577 included in the other datasets.

578

Non-CO₂ emissions were much more homogeneous, with differences among datasets that were approximately 5 times lower than CO₂ variability (e.g. 0.3 vs 1.5) (Figure 6a). Livestock led CH₄ emissions and showed the largest differences among datasets, with the Hotspot data (Herrero et al., (2013) having the lowest CH₄ emissions, which were compensated with larger N₂O than the other datasets (Figure 6b,c).

At a global level, wetlands dominates natural CH₄ emissions, while agriculture and fossil 584 fuels represent 2/3 of all human emissions, with smaller contributions coming from biomass 585 burning, the oceans, and termites (Montzka et al., 2011). Fire non-CO₂ emissions were quite 586 similar among datasets, confirming that FAOSTAT omissions were CO₂ related. Thus, as 587 exposed in FAOSTAT's metadata, only N2O and CH4 are considered in forest fires, 588 589 excluding CO₂ from aboveground biomass. As expected, N₂O emissions in crops showed 590 large differences, with our Hotspots having the lowest values (3 times lower). Rice N₂O emissions were omitted in all datasets except the Hotspots (Li et al., 2013), which also 591 included SOC. 592





594 The importance of multigas assessments relates to their role in radiative forcing (RF) understood as a measure of the warming strength of different human and natural agents (gases 595 and not gases) in causing global warming $(W.m^{-2})$. CO₂ is the most abundant 379 ppm in 596 2005 (400ppm in 2015), leading to an RF of 1.66±0.17 Wm⁻². Fossil fuels and cement 597 production have contributed about three-quarters of that RF, with the remainder caused by 598 land use changes (AR4). The growth rate of CO_2 in the atmosphere in 1995-2005 (1.9 ppm. 599 yr^{-1}) increased the CO₂ RF by 20%, being the largest change observed or inferred for any 600 decade in the last 200 years (AR4). Non-CO₂ GHG are less abundant in the atmosphere 601 $(1,774 \text{ ppb and } 319 \text{ ppb for } CH_4 \text{ and } N_2O \text{ in } 2005 \text{ respectively})$ but have larger warming 602 potentials (x 28 for CH₄) and (x 265 for N₂O) (0.48±0.05 and 0.16±0.02 Wm⁻² in 2005, 603 respectively) (AR4) and shorter lifetimes than CO₂ (~9 and ~120 years, respectively) offering 604 an additional opportunity to lessen future climate change (Montzka et al., 2011). Growth rates 605 in the atmosphere differ among gases with CO₂ and N₂O showing quasi linear increases while 606 CH_4 shows peculiar patterns that are not fully resolved (Montzka et al., 2011). The sensitivity 607 of CH₄ emissions from wetlands to warmer and wetter climates suggests a positive feedback 608 between emissions and climate change that is visible in ice-core records (Montzka et al., 609 2011). In the case of N_2O , and contrarily to the large contribution of non-human CH_4 610 emissions, anthropogenic emissions currently account for most of them (40%) primarly from 611 612 agricultural activities.

613

614 3.4 Country level emissions

Figures 7 and 8 show country level agreement for the AFOLU, forests, cropland and livestock
emission sectors, for the FAOSTAT, EDGAR and Hotspot databases. The use of percentiles
forced each figure to have a similar number of countries per category of agreement (high,
moderate, low and very low), in detriment of sectorial comparisons. Thus, if we contrasted





619

620	agreement. However, we thought it useful to search for within emissions differences, to
621	improve the estimates in each emission sector. No country had high agreement for all the
622	emission sectors, with Brazil, India and Cambodia showing the best results (high agreement in
623	3 out of 4 sectors). CS America (Mexico, Guatemala, Bolivia, Venezuela, Paraguay,
624	Argentina, Uruguay) and Asia (Myanmar, Viet Nam, Thailand, Indonesia, Malaysia) showed
625	the second best agreements (3 out 4 sectors with high or moderate agreement). No country
626	showed very low agreement for all the emission sectors, but African countries (Angola,
627	Botswana, Somalia, Nigeria, Ghana, Cote d'Ivoire), CS American (Chile, French Guiana,
628	Suriname) and Asian (Papua New Guinea, Sri Lanka and Nepal) showed the largest
629	disagreements (3 out of 4 sectors with low or very low agreement). From a sectorial
630	perspective, emissions showed good agreement where they were expected to peak (e.g. forest
631	agreement was high in tropical countries, livestock in Asia, crops in CS America and parts of
632	Asia) (Figures 7,8). From a continental perspective, Africa showed more countries with high
633	levels of disagreement, suggesting the need for further data research.
634	

forests to livestock emissions, the later would have had most countries on the level of high

635 3.5 Some reflections on the datasets

636 *3.5.1 Original goals*

Different datasets were developed for different purposes that have influenced the methods and approaches chosen to estimate their land use GHGs. Thus, EDGAR was created with an air pollution focus making its land emissions weaker. Contrastingly, FAOSTAT carries FAO's focus on land, particularly agriculture, with forest data coming later, through the FRA assessments. The 'Hotspot' database was created to identify the areas with the largest land use emissions in the tropics (emissions hotspots), while Houghton's accent is on historical LULUCF emission trends (since 1850). EPA concentrates on industrial, energy, and





644 agricultural emissions -forests are excluded- with an interest on human health and mitigation. 645 Moreover, several datasets rely on FAOSTAT's long-term agricultural data, which probably explains why the agricultural estimates are more homogeneous (crops, rice, and livestock). 646 FAOSTAT's forest emissions use FRA data, which get updated every 5 years. Different FRA 647 versions strongly influence forest emission and must be considered when comparing 648 estimates (e.g. differences up to 22% between the forest sink estimates using FRA2015 and 649 650 FRA2010 have been reported by Federici et al., 2015). Similarly, different versions of Houghton's bookkeeping TRENDS data, as well as researchers' self-tuned versions of his 651 model, result in emission differences that are difficult to track. 652

653

3.5.2 IPCC guidelines and guidance: Under the UNFCCC, countries are requested to use the 654 655 latest IPCC AFOLU guidelines to estimate their GHG emissions (e.g. IPCC 2006 and 2003 for developed and developing countries, respectively). The use of different guidelines, Tiers, 656 and approaches influences the final emission estimates. Compliance with IPCC has two main 657 consequences: 1. the total area selected to report emissions, and 2. the choice of land use over 658 land cover. In the first case, under IPCC guidance, the total area selected to report emissions 659 would include all the land under human influence (the managed land concept, which includes 660 areas under active and non-active management). Houghton's bookkeeping model (and the 661 carbon modelling community in general) do not comply well with the managed land concept, 662 663 resulting in different net emissions from forest land uses and land use changes (LULUCF) 664 than IPCC compliant country emissions (Grassi and Dentener, 2015; Federici et al., 2016). In the second case, the selection of land uses instead of land covers has partly been behind the 665 recent controversy between FAO and the Global Forest Watch's reported estimates on 666 deforestation trends (REF). Estimates of deforestation that rely on land cover are higher than 667 those using land use, since forest losses under forest land uses -that remain forest land use-668





669 are not considered deforestation (e.g. logged areas will regrow). In our analysis, FAO and Houghton relies on land use for deforestation, while the 'Hotspots' and EDGAR rely on land 670 cover. FAOSTAT and the 'Hotspots' rely on the 2006 IPCC Guidelines for National 671 Greenhouse Gas Inventories (IPCC, 2006). FAOSTAT uses Tier 1 and standard emission 672 factors, while the 'Hotspots' use a combination of Tiers (Tier 3 for all emissions except wood 673 harvesting and cropland emissions over histosols that rely on Tier 1). EDGAR reports the use 674 675 of 2006 IPCC Guidelines for the selection of the emission factors but some of their methodological approaches are not always consistent with IPCC guidelines (e.g. deforestation 676 expressed as the decay of burned forests, wood-harvesting is part of the energy sector, 677 agricultural energy balances are included in the AFOLU budget). EPA methods are reported 678 to be consistent with IPCC guidelines and guidance, with Tier 1 methodologies used to fill in 679 missing or unavailable data (USEPA, 2013). 680

681

682 4. CONCLUSIONS

The Paris Agreement (COP21) counts on the Nationally Determined Contributions (NDCs) as 683 the core of its negotiations to fight climate change. As March 2016, 188 countries had 684 submitted their NDCs under the UNFCC (FAO, 2016) with agriculture (crops, livestock, 685 fishery and aquaculture) and forests as prominent features in meeting the countries' mitigation 686 and adaptation goals (86% percent of the countries include AFOLU measures in their NDCs, 687 688 placing it second after the energy sector) (FAO, 2016). However, there exists large variability 689 in the way countries present their mitigation goals, and quantified sector-specific targets are rare (FAO, 2016). Variability relates not only to the lack of a standardized way to report 690 mitigation commitments under the NDCs, but also to uncertainties and gaps in the AFOLU 691 data. The Paris Agreement relies on a 5-year cycle stock-taking process to enhance mitigation 692 ambition, and to keep close to the 2°C target. To be effective and efficient, stock-taking needs 693





694	robust, transparent and certain numbers (at least with known uncertainties). This is true both
695	for national emission reports and NDCs, but also for the global datasets that can be used to
696	review the feasibility of countries' mitigation claims, and the real space for further mitigation
697	commitments. We have here compared the gross AFOLU emissions of six datasets to search
698	for disagreements, gaps, and uncertainties, focusing on the tropical region. Conclusions
699	depend on the spatial scale. At the tropical scale:
700	• Data aggregation offers much closer emission estimates than disaggregated data (e.g.
701	country level, continental level, gas level, emission source level).
702	• Forest emissions are the most uncertain of the AFOLU sector, with deforestation
703	having the highest uncertainties.
704	• Agricultural emissions, particularly livestock, are the most homogeneous of the
705	AFOLU emissions.
706	• Forest degradation, both fire and wood harvesting, show the largest variabilities
707	among databases.
708	• CO ₂ is the gas with longer-term influence in climate change trends, but remains the
709	most uncertain of the AFOLU gases.
710	• In absolute values, GHG disaggregation shows the largest differences for CO ₂ in fire
711	emissions.
712	• N_2O variability affected all the emission sources, making it the most dissimilar of the
713	non-CO ₂ gases.
714	• Emissions from histosols/peatlands remain incomplete or fully omitted in most
715	datasets.
716	At continental level:
717	• The level of disagreement of the emission sources at continental scale makes it
718	difficult to track the most possible drivers behind the emissions.





- 719 At country level:
- Countries with higher agreement among databases were present in all continents, with
- 721 Africa showing the highest levels of country disagreement.
- 722

723 4.1 Next steps

724 4.1.1 Enhancing dialogue between the carbon and the AFOLU research communities

Research ran by the carbon community is pivotal for AFOLU assessments and while these 725 two research communities overlap, they do not focus on exactly the same topics. The carbon 726 community works with CO₂ emissions-only, fully excluding non-CO₂ gases, particularly N₂O. 727 It moreover rather focuses on forests and associated land use changes, excluding emissions 728 from agriculture. The AFOLU community has, contrarily, a multi-gas approach (CO₂, CH₄, 729 N₂O) and includes emissions from both forests and agriculture. For these reasons, estimates 730 of the carbon community cannot be considered as AFOLU estimates, and certain confusion 731 appears in the IPCC's AR5 with an incorrect AFOLU labelling (Table 11.1, Fig S2 in SI). 732 733 There is great space for these two communities to cooperate but further dialogue is needed to promote closer and more coordinated action. Future steps might include the adoption of the 734 managed land concept by the carbon community; and ways to include legacy emissions by 735 736 the AFOLU community.

737

738 4.2.2 Improving data quality

The quality of the reported AFOLU emissions can be assessed through the UNFCCC principles: completeness, comparability, consistency, accuracy and transparency, which can help navigate the improvements of national monitoring systems. From these principles, the reviewed datasets performed well in *consistency* (they applied similar methods and assumptions over time, with the exception of 'Hotspots' that did not include temporal data).





744	Transparency was excellent for FAOSTAT with well elaborated and publicly available
745	metadata linked to their offered data, while EDGAR performed poorly due to insufficient
746	metadata. Improving transparency is an urgent call for future action. Accuracy and
747	uncertainty are also urgent calls. Thus, in spite of their importance to fully understand the
748	emission trends and dynamics, only Houghton and the 'Hotspots' provided uncertainties.
749	FAO offered uncertainties as a percent value for each emission source. Completeness and
750	omissions are also urgent tasks because all datasets are incomplete (Table 1) (e.g. missing
751	pools, missing gases) and omissions affect all datasets. Complete emission reporting should
752	consider the importance of:
753	• Forest soil CO ₂ and N ₂ O emissions (Werner et al., 2007) (e.g. N ₂ O tropical forest soil
754	emissions of 0.7 $PgCO_2e.yr^{-1}$).
755	- Emissions from CH_4 and $N_2\text{O}$ from drained peatland soils, and from wetlands over
756	managed land (e.g. conservation).
757	• All forest fire types (e.g. temperate conifers and woodlands; understory fires over
758	humid closed canopy forests (Alencar et al., 2006; Morton et al., 2013) (e.g. 85,500
759	km ² , 1999-2010 in southern Brazilian Amazon); fire emissions over peatland soils and
760	peatland forests out of Asia (Román-Cuesta et al., 2011; Oliveras et al., 2014) (e.g. 4-8
761	TgCO ₂ e, 1982-1999, for the tropical high Andes from Venezuela to Bolivia)
762	\bullet CO2 emissions from other components of wood harvesting other than fuel and
763	industrial roundwood (e.g. charcoal, residues).
764	• CO ₂ emissions from tree biomass loss due to fragmentation (Numata et al., 2010; Pütz
765	et al., 2014) (e.g 0.2 Pg C y^{-1})
766	• CO_2 due to decomposition and decay of forests under extreme events: hurricanes
767	(Read and Lawrence, 2003; Negron-Juarez et al., 2010) (e.g the 2005 convective
768	storm, the Amazon basin suffered from an estimated tree mortality of 542±121 million





769	trees); intense droughts (Phillips et al., 2009, 2010; Brienen et al., 2015) (e.g. the 2005
770	Amazonian drought resulted in 1.2-1.6 PgC emissions and the atmosphere has yet to
771	see 13.9 PgCO ₂ (3.8 PgC) of the Amazon necromass carbon produced since 1983);

772

Further suggestions on improving data gaps and knowledge for the AFOLU sector have been reported by Smith *et al.* (2014); Houghton *et al.* (2012); USEPA (2013) and Sist *et al.* (2015), with a focus on soil data and crop production systems, as well as an improved understanding of the mitigation potentials, costs and consequences of land use mitigation options.

777

778 **5. REFERENCES**

- Abad-Viñas, R., Blujdea, V., Federici, S., Hiederer, R., Pilli, R., Grassi, G.: Analysis and 779 proposals for enhancing Monitoring, Reporting, and Verification of greenhouse gases from 780 Land Use, Land Use Change and Forestry in the EU. Technical Report 781 071201/2011/211111/CLIMA.A2. Joint Research Centre, Ispra, Italy, 2015. Available at: 782 http://publications.jrc.ec.europa.eu/repository/bitstream/JRC91414/lb-na-26813-en-n.pdf 783 Achard, F., Beuchle, R., Mayaux, P., Stibig, H. J., Bodart, C., Brink, A., Carboni, S., Desclée, 784 785 B., Donnay, F., Eva, H. D., Lupi, A., Raši, R., Seliger, R. and Simonetti, D.: Determination of tropical deforestation rates and related carbon losses from 1990 to 2010. Global Change 786 Biology, 20, 2540-2554, 2014. 787 788 Alencar, A., Nepstad, D., Vera-Diaz, MC.: Forest Understory Fire in the Brazilian Amazon in
- ENS and non-ENSO years: Area Burned and Committed Carbon Emissions. EarthInteractions, 10, 1-17, 2006.
- Anderson, K.: The inconvenient truth of carbon offsets. Nature News, 484, 7, 2012.
- Anderson, K.: Duality in climate science. Nature Geoscience, 8, 898–900, 2015.





- 793 Asbjornsen, H., Gallardo-Hernández, C., Velázquez-Rosas, N., García-Soriano, R.: Deep
- round fires cause massive above- and below-ground biomass losses in tropical montane
- cloud forests in Oaxaca, Mexico. Journal of Tropical Ecology, 21, 427-434, 2005.
- Baccini, A., Goetz, SJ., Walker, WS., Laporte, N. T., Sun, M., Sulla-Menashe, D., Hackler, J.,
- 797 Beck, P. S. A., Dubayah, R., Friedl, M. A., Samanta, S., Houghton, RA.: Estimated carbon
- dioxide emissions from tropical deforestation improved by carbon-density maps. Nature
- 799 Climate Change, 2, 182-185, 2012.
- Balch, J., Brando, P., Nepstad, D. Coe, M., Silverio, D., Massad, T., Davidson, E., Lefebvre,
- 801 P., Oliveira-Santos, C., Rocha, W., Cury, R., Parsons, A., Carvalho, K.: The susceptibility of
- 802 Southeastern Amazon Forests to Fire: Insights from a Large Scale Burn Experiment.
- BioScience, 65, 893-905, 2015.
- 804 Barlow, J., Peres, C.: Fire-mediated dieback and compositional cascade in an Amazonian
- forest. Philosophical Transactions of the Royal Society, 363, 1787-1794, 2008.
- 806 Bellassen, V., Luyssaert, S.: Managing forests in uncertain times. Nature, 506, 153-156, 2014.
- 807 Brando, PM., Nepstad, DC., Balch, JK., Bolker, B., Christman, MC., Coe, M., Putz, F.: Fire-
- induced tree mortality in a neotropical forest: the roles of bark traits, tree size, wood density,
- and fire behaviour. Global Change Biology, 18, 630-641, 2012.
- 810 Brando, PM., Balch, JK., Nepstad, DC., Morton, D., Putz, F., Coe, M., Silverio, D., Macedo,
- 811 M., Davidson, E., Nobrega, C., Alencar, A., Soares-Filho, B.: Abrupt increases in
- 812 Amazonian tree mortality due to drought-fire interactions. Proceedings of the National
- 813 Academy of Sciences, 11, 6347-6352, 2014.
- 814 Brienen, R., Phillips, O., Feldspausch, T., Gloor, E., Lloyd, J., Lopez-Gonzalez, G.,
- 815 Morteagudo-Mendoza, A., Malhi, Y., Lewis, S., Vasquez Martinez, R., Alexiades, M.,
- 816 Alvarez, E., Alvarez-Loayzada, P., Zagt, R.: Long term decline of the Amazon carbon sink.
- 817 Nature, 519, 344-361, 2015.





- 818 Canadell, J., Schulze. D. Global potential of biospheric carbon management for climate
- mitigation. Nature Communications, 5, 5282-5293, 2014.
- 820 Ciais, P., Dolman, J., Bombelli, A., Duren, A., Peregon, A., Rayner, P., Miller, C., Gobron,
- 821 N., et al.: Current systematic carbon-cycle observations and the need for implementing a
- policy-relevant carbon observing system. Biogeosciences, 11, 3547-3602, 2014.
- 823 Cochrane, M., Alencar, A., Schulze, M., Souza, C., Nepstad, D., Lefebvre, P., Davidson, E.:
- Positive feedbacks in the fire dynamics of closed canopy tropical forests. Science, 284,
- 825 1832-1835, 1999.
- 826 Crowther, T., Thomas, S., Maynard, D., Baldrian, P., Covey, K., Frey, S., van Diepen, L.,
- 827 Bradford, M.: Biotic interactions mediate soil microbial feedbacks to climate change.
- Proceedings of the National Academy of Science, 112, 7033-7038, 2015.
- 829 EDGAR. The Emissions Database for Global Atmospheric Research (2012) Part III:
- 830 Greenhouse gas emissions. http://edgar.jrc.ec.europa.eu/docs/IEA_PARTIII.pdf
- 831 Estrada, M., Lee, D., Murray, B., O'Sullivan, R., Penman, J., Streck, C.: Land Use in a Future
- 832 Climate Agreement. # S-LMAQM-13-CA-1128 U.S. Department of State, 2014. Available
- at: http://merid.org/land-use-in-ADP/
- FAO. The Agriculture sectors in the intended nationally determined contributions. Analysis.
- Environmental and natural resource management working paper 61. Rome, Italy. 2016.
- 836 Available at: http://www.fao.org/3/a-i5687e.pdf
- Federici, S., Tubiello, F., Salvatore, M., Jacobs, H., Schmidhuber, J.: New estimates of CO2
- forest emissions and removals: 1990-2015. Forest Ecol.Manag., 3, 89-98, 2015.
- 839 Federici, S., Grassi, G., Harris, N., Lee, D., Neeff, T., Penman, J., Sanz-Sanchez, M., Wolosin
- 840 M.: GHG fluxes from forests: an assessment of national reporting and independent science
- in the context of the Paris Agreement. Working Paper. UCLA, San Francisco, 2016.





- 842 Available at: http://www.climateandlandusealliance.org/wp-content/uploads/2016/06/
- 843 GHG_Fluxes_From_Forests_Working_Paper.pdf
- 844 Friedlingstein, P., Andrew, R., Rogelj, J., Peters G., Canadell J., Knutti, R., Luderer, G.,
- Raupach, M., Schaeffer, M., van Vuuren, D., Le Quéré, C.: Persistent growth of CO₂
- emissions and implications for reaching climate targets. Nature geoscience, 7, 709-715,
- 847 2014.
- 848 Forest Resources Assessment (FRA) (2005) http://www.fao.org/forestry/fra/fra2005/en/
- 849 Forest Resources Assessment (FRA) (2010). http://www.fao.org/forestry/fra/fra2010/en/
- Frank, D., Reichstein, M., Bahn, M., Thonicke, K., Frank. D., Mahecha, M., Smith, P., van
- der Velde, M. et al. Effects of climate extremes on the terrestrial carbon cycle: concepts,
- processes and potential future impacts. Global Change Biology, 21, 2861-2880, 2015.
- 853 Francey, R., Trudinger, C., Van der Schoot, M., Krummel, P., Steele, L., Langenfelds, L.:
- Differences between trends in atmospheric CO2 and the reported trends in anthropogenic
- 855 CO₂ emissions. Tellus, 62B, 316-328, 2010.
- 856 Francey, R., Trudinger, C., Van der Schoot, M., Law, M., Krummel, P., Langenfelds, R.,
- 857 Steele, P., Allison, C., Stavert, A., Andres, R., Rödenbeck, C.: Atmospheric verification of
- anthropogenic CO₂ emission trends. Nature Climate Change, 3, 520-525, 2013a.
- 859 Francey, R., Trudinger, C., Van der Schoot, M., Law, M., Krummel, P., Langenfelds, R.,
- Steele, P., Allison, C., Stavert, A., Andres, R., Rödenbeck, C.: Reply to Anthropogenic CO₂
- emissions. Nature Climate Change, 3, 603-604, 2013b.
- 862 Giglio, L., Randerson, J., van der Werf, G., Kasibhatla, P., Collatz, G., Morton D., DeFries,
- R.: Assessing variability and long-term trends in burned area by merging multiple satellite
 fire products. Biogeosciences, 7, 1171–1186, 2010.
- Grace, J., Mitchard, E., Gloor, E.: Perturbations in the carbon budget of the tropics. Global
- 866 Change Biology, 20, 3238-3255, 2014.





- 867 Grassi, G., Dentener, F.: Quantifying the contribution of the Land Use Sector to the Paris
- 868 Climate Agreement. The LULUCF sector within the Intended Nationally Determined
- 869 Contributions. EUR 27561.JRC Science for Policy Report. Ispra, Italy, 2015. Available at:
- 870 http://publications.jrc.ec.europa.eu/repository/bitstream/JRC98451/jrc%20lulucf-
- indc%20report.pdf
- 872 Grassi, G., Monni, S., Federici, S., Achard, F., Mollicone, D.: Applying the conservativeness
- 873 principle to REDD to deal with uncertainties of the estimates. Environmental Research
- ketters, 3, 035005, 2008.
- Hansen, M., Stehman, S., Potapov, P.: Quantification of global gross forest cover loss.
 Proceedings of the National Academy of Sciences, 107, 8650-8655, 2010.
- 877 Harris, N., Brown, S., Hagen, S., Saatchi, S., Petrova, S., Salas, W., Hansen, M., Potapov, P.,
- Lotsch, A.: Baseline Map of Carbon Emissions from Deforestation in Tropical Regions.
 Science, 336, 1576-1578, 2012.
- 880 Herrero, M., Havlik, P., Valin, H., Notenbaert, A., Rufino, M., Thornton, P., Blümmel, M.,
- Weiss, F., Grace, D., Obesteiner, M.: Biomass use, production, feed efficiencies, and
 greenhouse gas emissions from global livestock systems. Proc. Natl. Acad. Sci., 110, 2088820893, 2013.
- Herold, M., Roman-Cuesta, RM., Heymell, V., Hirata, Y., Van Laake, P., Asner, G., Souza,
- 885 C., Avitabile, V., MacDicken, K.: A review of methods to measure and monitor historical
- carbon emissions from forest degradation. FAO. Unasylva, 238, 16-24, 2011a. Available at:
- 887 http://www.fao.org/docrep/015/i2560e/i2560e04.pdf
- 888 Herold, M., Roman-Cuesta, RM., Mollicone, D., Hirata, Y., Van Laake, P., Asner, G., Souza,
- 889 C., Skutch, M., Avitabile, V., MacDicken, K.: Options for monitoring and estimating hisoric
- so carbon emissions from forest degradation in the context of REDD+. Carbon Balance and
- 891 Management, 6, 13-20, 2011b.





- 892 Hooijer, A., Page, S., Canadell, J., Silvius, M., Kwadijk, J., Wosten, H., Jauhiainen, J.:
- 893 Current and future CO2 emissions from drained peatlands in Southeast Asia.
- Biogeosciences, 7, 1505-1514, 2010.
- Houghton, RA.: The annual net flux of carbon to the atmosphere from changes in land use
- 896 1850-1990. Tellus B, 51, 298-313, 1999.
- Houghton, RA.: How well do we know the flux of CO2 from land-use change? Tellus B, 62,
 337-351, 2010.
- Houghton, RA.: Carbon emissions and the drivers of deforestation and forest degradation in
 the tropics. Current Opinion in Environmental Sustainability, 4, 597-603, 2012.
- 901 Houghton, RA., Hackler, JL.: Carbon Flux to the Atmosphere from Land-Use Changes: 1850
- 902 to 1990. ORNL/CDIAC-131, NDP-050/R1. Carbon Dioxide Information Analysis Center,
- 903 U.S. Department of Energy, Oak Ridge National Laboratory, Oak Ridge, Tennessee, U.S.A,904 2001.
- Houghton, RA., House, JI., Pongratz, J., van der Werf, G., DeFries, R., Hansen, M., Le
 Quere, C., Ramankutty, R.: Carbon emissions from land use and land-cover change.
 Biogeosciences, 9, 5125-5142, 2012.
- IPCC. Intergovernmental Panel on Climate Change.: Good Practice Guidance for Land Use,
 Land Use and Forestry.. IPCC National Greenhouse Gas Inventory Programme. (ed. Penman
 J, Gytarsky M, Hiraishi T, Krug T, Kruger D, Ppatti R, Buendia L, Miwa K, Ngara T,
 Tanabe K, Wagner F) IGES. Kanagawa, Japan, 2003 http://www.ipcc-
- 912 nggip.iges.or.jp/public/gpglulucf/gpglulucf_files/GPG_LULUCF_FULL.pdf
- 913 IPCC. Intergovernmental Panel on Climate Change.: AFOLU Guidelines for National
 914 Greenhouse gas Inventories. Vol. 4: Agriculture, Forestry and Other Land Use (eds
 915 Eggleston S, Buendia L, Miwa K, Ngara T, Tanabe K). IGES, Kanagawa, Japan. 2006.
- 916 http://www.ipcc-nggip.iges.or.jp/public/2006gl/vol4.htm





- 917 IPCC. Intergovernmental Panel on Climate Change.: Summary for Policymakers. In: Climate
- 918 Change 2014: Mitigation of Climate Change.Contribution of Working Group III to the Fifth
- 919 Assessment Report of the Intergovernmental Panel on Climate Change (eds Edenhofer O,
- 920 Pichs-Madruga R, Sokona Y, E, Farahani E, Kadner S, Seyboth K, Adler A, Baum I,
- 921 Brunner S, Eickemeier P, Kriemann B, Savolainen J, Schlomer S, von Stechow C, Zwickel
- 922 T, Minx JC). Cambridge University Press, Cambridge, United Kingdom and New York,
- 923 NY, USA, 2014.
- 924 Laurance, W., Williamson, G.: Positive feedbacks among forest fragmentation, drought and
- climate change in the Amazon. Conservation Biology, **15**, 1529–1535, 2001.
- 926 Le Quéré, C., Peters, G.P., Andres, R. J., Andrew, R. M., Boden, T. A., Ciais, P.,
- 927 Friedlingstein, P., Houghton, R.A., Marland, G., Moriarty, R., Sitch, S., Tans, P., Arneth,
- 928 A., Arvanitis, A., Bakker, D. C.E., Bopp, L., Canadell, J. G., Chini, L. P., Doney, S. C.,
- 929 Harper, A., Harris, I., House, J.I., Jain, A. K., Jones, S. D., Kato, E., Keeling, R. F., Klein
- 930 Goldewijk, K., Körtzinger, A., Koven, C., Lefèvre, N., Maignan, F., Omar, A., Ono, T.,
- 931 Park, G.-H., Pfeil, B., Poulter, B., Raupach, M. R., Regnier, P., Rödenbeck, C., Saito, S.,
- 932 Schwinger, J., Segschneider, J., Stocker, B. D., Takahashi, T., Tilbrook, B., van Heuven, S.,
- Viovy, N., Wanninkhof, R., Wiltshire, A., and Zaehle, S. Global carbon budget 2013, Earth
 System Science Data, 6, 235-263, 2014.
- Li, C., Salas, W., DeAngelo, B., Rose, S.: DNDC9.5 in EPA (2013) Global Mitigation of non-
- 936 CO2 Greenhouse Gases: 2010-2030. EPA Technical Report-430-R-13-011, US, 2013.
- 937 Country data available at: http://www.epa.gov/climatechange/EPAactivities/economics/
 938 nonco2 projections.html
- Montzka, SA., Dlugokencky, EJ., Butler, JH.: Non-CO2 greenhouse gases and climate
 change. Nature, 476, 43-51, 2011.





- 941 Morton, DC., Le Page, Y., DeFries, R., Collatz, GJ., Hurtt, GC.: Understory fire frequency
- and the fate of burned forests in southern Amazonia. Philosophical Transactions of the
- 943 Royal Society B, 368, 20120163, 2013.
- 944 Negron-Juarez, RI., Chambers, J., Guimaraes, G., Zeng, H., Raupp, C., Marra, D., Ribeiro,
- 945 G., Saatchi, S., Nelson, B., Higuchi, N.: Widespread Amazon forest tree mortality from a
- single cross-basin squall line event. Gephysical Research Letters, 37, L16701, 2010.
- Numata, I., Cochrane, M., Roberts, D., Soares, J., Souza, C., Sales, M.: Biomass collapse and
 carbon emissions from forest fragmentation in the Brazilian Amazon. Journal of
 Geophysical Research, 115, G03027, 2010.
- 950 Ogle, S. et al. in EPA (2013) Global Mitigation of non-CO2 Greenhouse Gases: 2010-2030.
- 951 EPA Technical Report-430-R-13-011, 2013.(data available upon request) Country data
- available at:http://www.epa.gov/climatechange/EPAactivities/economics/nonco2projections.
 html
- 954 Oliveras, I., Malhi, Y., Salinas, N., Huaman, V., Urquiaga-Flores, E., Kala-Mamani, J.,
- 955 Quintano-Loaiza, JA., Cuba-Torres, I., Lizarraga-Morales, N., Roman-Cuesta, RM.:
- 956 Changes in forest structure and composition after fire in tropical montane cloud forests near

the Andean treeline. Plant Ecology and Diversity, 7, 329-340, 2013.

- 958 Oliveras, I. Anderson, D., Malhi, Y.: Application of remote sensing to understanding fire
- 959 regimes and biomass burning emissions of the tropical Andes. Global Biogeochemical Cy
- 960 cles, 28, 480-496, 2014.
- Pelletier, J., Busch, J., Potvin, C.: Addressing uncertainty upstream or downstream of
 accounting for emissions reductions from deforestation and forest degradation. Climatic
- 963 Change, 130, 635-648, 2015.
- Peters, G., Davis, J., Andrew, R.: A synthesis of carbon in international trade. Biogeosciences,
 9, 3247-3276, 2012.





- 966 Phillips, O., Aragao, L., Lewis, S., et al.: Drought Sensitivity of the Amazon Rainforest.
- 967 Science, 323, 1344-1347, 2009.
- 968 Phillips, O., van der Heijden, G., Lewis, S., et al.: Drought-mortality relationships for tropical
- 969 forests. New Phitologist, 187, 631-646, 2010.
- 970 Phillips, O., Lewis, S.: Evaluating the tropical forest carbon sink. Global Change Biology, 20,
- 971 2039-2041, 2014.
- 972 Poorter, L., Bongers, F., Aide et al.: Biomass resilience of Neotropical secondary forests.
- 973 Nature, 530, 211-214, 2016.
- 974 Pütz, S., Groeneveld, J., Henle, K., Knogge, C., Martensen, A., Metz, M.: Long-term carbon
- loss in fragmented Neotropical forests. Nature communications, 5, 5037-5045, 2014.
- Read, L., Lawrence, D.: Recovery of biomass following shifting cultivation in dry tropical
 forests of the Yucatan. Ecological Applications, 13, 85–97, 2003.
- 978 Richards, M., Bruun, T., Campbell, BM., Gregersen, L., Huyer, S., Kuntze, V., Madsen, T.,
- 979 Oldvig, M., Vasileiou, I.: How countries plan to address agricultural adaptation and
- 980 mitigation: An analysis of Intended Nationally Determined Contributions. CGIAR Research
- 981 Program on Climate Change, Agriculture and Food Security (CCAFS), Copenhagen, 2015.
- Richter, D., Houghton, RA.: Gross CO2 fluxes from land-use change: implications for
 reducing global emissions and increasing sinks. Carbon management, 2, 41-47, 2011.
- 984 Roman-Cuesta, RM., Gracia, M., Retana, J.: Environmental and human factors influencing
- fire trends in Enso and non-Enso years in tropical Mexico. Ecological Applications, 13,
 1177–1192, 2003.
- Roman-Cuesta, RM., Salinas, N., Asbjornsen, H. et al.: Implications of fires on carbon
 budgets in Andean cloud montane forest: The importance of peat soils and tree resprouting.
- 989 Forest Ecology and Management, 261, 1987–1997, 2011
- 990 Roman-Cuesta, RM., Rufino, M., Herold, M., et al.: Hotspots of tropical land use emissions:





- 991 patterns, uncertainties, and leading emission sources for the period 2000-2005.
- Biogeosciences, 2016, in press.
- 993 Romijn, E., Herold, M., Koistra, L., Murdiyarso, D., Verchot, L.: Assessing capacities of non-
- Annex I countries for national forest monitoring in the context of REDD+. Environmental
- 995 Science Policy, 19–20, 33–48, 2012.
- 996 Romijn, E., Lantican, C., Herold, M., Lindquist, E., Ochieng, R., Wijaya, A., Murdiyarso, D.,
- Verchot, L.: Assessing change in national forest monitoring capacities of 99 tropical
 countries. Forest Ecology and Management, 352, 109–123, 2015.
- 999 Saatchi, S., Harris, N., Brown, S., et al.: Benchmark map of forest carbon stocks in tropical
- regions across three continents. Proceedings of the National Academy of Sciences, 108,9899-9904, 2012.
- 1002 Simula, M.: Towards defining forest degradation: comparative analysis of existing definitions.
- 1003 Forest Resources Assessment. Working Paper 154. FAO. Rome, Italy, 2009. Aavailable at:
- 1004 ftp://ftp.fao.org/docrep/fao/012/k6217e/k6217e00.pdf
- 1005 Sist, P., Rutishauser, E., Peña-Claros, M., et al.: The Tropical Managed Forests Observatory:
- a research network addressing the future of logged forests. Applied Vegetation Science, 18,
- 1007 171-174, 2015.
- Smith, P., Martino, D., Cai, Z., et al.: Greenhouse gas mitigation in agriculture. Philosophical
 Transactions of the Royal Society B: Biological Sciences, 363, 789-813, 2008.
- 1010 Smith, P., Bustamante, M., Ahammad, H., et al.: Agriculture, Forestry and Other Land Use
- 1011 (AFOLU). In: Climate Change 2014: Mitigation of Climate Change. Contribution of
- 1012 Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on
- 1013 Climate Change (eds Edenhofer O, Pichs-Madruga R, Sokona Y, E, Farahani E, Kadner S,
- 1014 Seyboth K, Adler A, Baum I, Brunner S, Eickemeier P, Kriemann B, Savolainen J,





- 1015 Schlomer S, von Stechow C, Zwickel T, Minx JC). Cambridge University Press, Cambridge,
- 1016 United Kingdom and New York, NY, USA, 2014.
- 1017 Streck, C.: Forests and Land Use in the Paris Agreement. Climate Focus. 2015. Available at:
- 1018 http://www.climatefocus.com/sites/default/files/20151223%20Land%20Use%20and%20the
- 1019 %20Paris%20Agreement%20FIN.pdf
- 1020 Tian, H., Lu, C., Ciais, P., Michalak, A., Canadell, J., Saikawa, E., Huntzinger, D., Gurney,
- 1021 K., Sitch, S., Zhang, B., Yang, J., Bousquet, P., Bruhwiler, L., Chen, G., Dlugokencky, E.,
- 1022 Wofsy, S.: The terrestrial biosphere as a net source of greenhouse gases to the atmosphere,
- 1023 Nature, 531, 225–228, 2016.
- Tubiello, F., Salvatore, M., Rossi, S., Ferrara, A., Fitton, N., Smith, P.: The FAOSTAT
 database of greenhouse gas emissions from agriculture. Environmental Research Letters, 8,
 015009-19, 2013.
- 1027 Tubiello, F., Salvatore, M., Cóndor Golec, R., Ferrara, A., Rossi, S., Biancalani, R., Federici,
- 1028 S., Jacobs, H., Flammini, A.: Agriculture, Forestry and Other Land Use Emissions by
- 1029 Sources and Removals by Sinks 1990 2011 Analysis. Working Paper Series ESS/14-
- 1030 02.FAO Statistical Division. Rome, Italy, 2014. Available at:
- 1031 http://www.fao.org/docrep/019/ i3671e/i3671e.pdf
- 1032 Tubiello, F., Salvatore, M., Ferrara, A., House, J., Federici, S., Rossi, S., Biancalani, R.,
- 1033 Condor Golec, R., Jacobs, H., Flammini, A., Prosperi, P., Cardenas-Galindo, P.,
- 1034 Schmidhuber, J., Sanz Sanchez, MJ., Srivastava, N., Smith, P. The contribution of
- 1035 Agriculture, Forestry and other Land Use Activities to Global Warming, 1990-2012. Global
- 1036 Change Biol., 21, 2655–2660, 2015.
- 1037 Turetsky, M., Benscoter, B., Page, S., Rein, G., Van der Werf, G., Watts, A.: Global
- vulnerability of peatlands to fire and carbon loss. Nature Geosciences, 8, 11-14, 2015.





- 1039 Uhl, C., Kauffman, J.: Deforestation effects on fire susceptibility and the potential response of
- the tree species to fire in the rainforest of the eastern Amazon. Ecology, 71, 437–449, 1990.
- 1041 USEPA. United States Environmental Protection Agency. Global Anthropogenic Non-CO₂
- 1042 Greenhouse Gas Emissions: 1990-2030. EPA 430-R-12-006. Washington, DC. 2012.
- 1043 Available at: http://www.epa.gov/climatechange/economics/international.html
- 1044 USEPA. United States Environmental Protection Agency. Global Mitigation of non-CO2
- 1045 Greenhouse Gases: 2010-2030. Technical Report-430-R-13-011, 2013. Available at:
- 1046 http://www.epa.gov/climate change/Downloads/EPAactivities/MAC_Report_2013.pdf
- 1047 Valentini, R., Arneth, A., Bombelli, A. et al.: A full greenhouse gases budget of Africa:
- synthesis, uncertainties and vulnerabilities. Biogeosciences, 11, 381-407, 2014.
- 1049 Van der Werf, G., Randerson, J., Giglio, L., Collatz, G., Mu, M., Kasibhatla, P., Morton, D.,
- 1050 DeFries, R., Jin, Y., van Leeuwen, T.: Global fire emissions and the contribution of
- 1051 deforestation, savannah, forest, agricultural, and peat fires (1997–2009). Atmospheric
- 1052 Chemistry and Physics, 10, 11707–11735, 2010.
- 1053 Werner, C., Butterbach-Bahl, K., Haas, E., Hickler, T., Kiese, R.: A global inventory of N₂O
- 1054 emissions from tropical rainforest soils using a detailed biogeochemical model. Global
- 1055 biogeochemical Cycles, 21, GB3010, 2007.
- 1056

1057 6. CONTRIBUTIONS

- 1058 RMRC, MR, MH designed the study. SO, BP provided data and ran quality controls of the
- data. RMRC, MR, MH, KBB, TR, LV, CM, SR, RH, SO, BP discussed the results and
- 1060 contributed to writing.

1061

1062 7. ACKNOWLEDGEMENTS

1063 This research was generously funded by the Standard Assessment of Mitigation Potential and





- 1064 Livelihoods in Smallholder Systems (SAMPLES) Project as part of the CGIAR Research
- 1065 Program Climate Change, Agriculture, and Food Security (CCAFS). Funding also came from
- 1066 two European Union FP7 projects: GEOCarbon (283080) and Independent Monitoring of
- 1067 GHG Emissions-N° CLIMA.A.2/ETU/2014/0008. Partial funds came through CIFOR from
- the governments of Australia (Grant Agreement # 46167) and Norway (Grant Agreement
- 1069 #QZA-10/0468). In the memory of Changsheng Li.





1070 Figures



1072



1076







Figure 2: Tropical gross annual emissions (2000-2005) comparisons, for the leading emission
sources in the AFOLU sector, for the Hotspots, FAOSTAT, EDGAR, Baccini, EPA and
Houghton datasets, in this order. Houghton's data are net land use emissions rather than
deforestation and are offered for visual comparisons against Baccini's gross deforestation
estimate.











- 1085 Figure 3: Continental disaggregated emissions for the individual emission sources in
- 1086 PgCO₂e.yr⁻¹. Bars indicate uncertainty estimates (1σ from mean). No uncertainty estimates
- 1087 are available for the other datasets.
- 1088







1090

Figure 4: Disaggregation of cropland soil emissions from drained peatands for the datasets where data were available in a disaggregated manner (FAOSTAT and Hotspots). Organic

1093 soils were excluded in EPA's cropland emissions.







1094

Figure 5: Contribution of the different AFOLU GHGs (CO2, CH4 and N2O) for the different
 datasets. Bars indicate uncertainty estimates (1σ from mean). No uncertainty estimates are

available for the other datasets.











- Figure 6: GHG emission contribution (CO₂, CH₄ and N₂O) of the leading AFOLU emission 1099 sources. Bars indicate uncertainty estimates (1o from mean). No uncertainty estimates are 1100 1101 available for the other datasets. **AFOLU** emissions **Forests emissions** Levels of agreement 1 High Moderate Low Very Low 1102 Figures 7: Country level agreement for AFOLU and forest emissions for the FAOSTAT, 1103 EDGAR and 'Hotspots' databases. The categories of agreement are percentiles of the 1104 coefficient of variation of the emission data (e.g. high agreement \geq 75th percentile, Moderate: 1105
- 1106 50^{th} -75th percentiles, Low: 50^{th} -25th percentiles, Very Low $\leq 25^{\text{th}}$ percentile)







1108	Figures 8: Country level agreement for croplands (cropland soils including histosols and rice)
1109	and livestock emissions, for the FAOSTAT, EDGAR and 'Hotspots' databases. The
1110	categories of agreement are percentiles of the coefficient of variation of the emission data
1111	(e.g. high agreement $\geq 75^{\text{th}}$ percentile, Moderate: 50^{th} - 75^{th} percentiles, Low: 50^{th} - 25^{th}
1112	percentiles, Very Low≤25 th percentile)



		Hotspots	FAOSTAT	EDGAR	Houghton	Baccini	EPA	AR5
	Gross/Net	Gross	Gross	Gross	Net	Gross	Gross	Net
	emissions							
	Uncertainty ^a	\checkmark	No	N_{O}	No	No	No	\checkmark
	Transparency	High	High	Low^{b}	Low	Low	Intermediate	Low
	IPCC compliant	\checkmark	$^{\wedge}$	\wedge	Not fully ^c	Not fully ^d	\checkmark	Not fully ^e
	Forest carbon	AGB + BGB	AGB +	AGB	AGB+BGB+Soil	AGB+BGB+Soil	Soil	AGB+BGB+Soil
	Pools		BGB		+CWD+Litter	+CWD+Litter		+CWD+Litter
	Gases	CO_2, CH_4 ,	$CO_2, CH_4,$	$CO_2, CH_4,$	CO_2	CO_2	$CO_2, CH_4,$	CO_2 for forests.
		N_2O	N_2O	N_2O			N_2O	CO ₂ ,CH ₄ , N ₂ O for
								agriculture and peatlands.
	Tier 1	$^{\wedge}$	$^{}$	\wedge			$^{\wedge}$	
	Tier 2, 3	$^{\wedge}$			$^{\wedge}$	$^{\wedge}$	$^{\wedge}$	
	Spatial	Pixel (0.5°)	Country	Country ^g	Region	Region	Country	Region
	Disaggregation ^f							
	Peatlands	\checkmark	\checkmark	\checkmark	No	No	No	\checkmark
1115	Table 1: difference	es and similaritie	es of the asses	sed AFOLU d	atasets.			

a Uncertainty at the level of disaggregation at which data are available to download. 1116 b Low means there is no metadata available, or metadata does not properly document the processes followed to estimate the emissions. 1117

c EDGAR data on deforestation emissions does not follow IPCC guidelines. 1118 d The bookkeeping approach does not follow the concept of managed land, and does not include the sink of forests remaining forests in managed 1119

land other than logged forests and those regrowing after shifting cultivation. 1120



1114 1113

Tables





- 1121 e Based on Houghton et al., (2012).
- 1122 f Available disaggregated data.
- g We selected data at the country scale to favour comparability with other datasets (e.g. FAOSTAT) even though data are available at pixel level 1123
- 1124 $(0.1^{\circ}).$





1125

Table 2: Summary of (a) tropical gross emissions estimates for agriculture, FOLU and AFOLU for all the datasets (Hotspots, FAOSTAT, EDGAR, EPA, Houghton) (2000-2005) and published data (Baccini et al., 2012, AR5 (Smith et al., 2014)) (2000-2007), and of (b) net global 1126 1127

estimates as reported by Tubiello et al., (2015). Houghton and EPA offer FOLU and agricultural data only, respectively, and therefore estimates 1128

for AFOLU are not complete. *Data exposed in Figure 11.2 in Chapter 11 Smith et al. (2014). 1129

¹¹³⁰ *net FOLU flux estimate.

¹¹³¹ ** Baccini et al., (2012) reported gross estimates for the FOLU components.

*** Baccini et al., (2012) estimates selected for the AR5 FOLU values in Figure 11.8, Chapter 11, WG-III. 1132

1133

1134

1135





		Deforestation	Wood	Fire	Enteric	Manure	Agricultural soils	Cropland over	Rice	Others
			Harvesting		Fermentation	management		histosols		
	CO ₂	$1^1, 2^2, 5^3 6^1$	$1^4, 2^5, 5^4, 6^4$	$1^{6}, 2^{7}, 3^{8}, 5^{9}, 6^{9}$				1 ¹⁰ ,2 ¹¹		3 ¹²
	CH_4			$1^{13}, 2^{14}, 3^{15}$	1,2,3,4	1,2,3,4			1,2,3,4	
	N_2O			$1^{13}, 2^{14}, 3^{15}$		1,2,3,4	$\frac{1^{16.17}, 2^{16.18}, 3^{16.17,19}}{4^{16.19}}$	1,2	1	
	dSOC						1		1	
1136	Table	3: Contribution (of different data	sets to the diff	erent emission sou	urces, disaggregate	ed by GHG gases. 1: Ho	otspots, 2: FAOS	TAT, 3:	
1137	EDGA	AR, 4: EPA (only	non-CO2 agricu	ılture emissior	is including livesto	ock), 5: Houghton	(only CO ₂ FOLU emis	sions. No disaggr	egated da	ita
1138	offerec	d), 6: Baccini et a	ıl., 2012 (only C	02 FOLU em	issions, based on F	Houghton bookkee	ping model). FAOSTA	T are estimated th	hrough T	ier 1
1139	approa	aches.								
1140	¹ Gross	s deforestation.								
1141	² Net d	leforestation								
1142	³ Houg	ghton net CO ₂ -on	ly estimates are	not deforestat	on emissions, but	land use and land	use change fluxes inclu	Iding deforestatio	on, forest	
1143	degrad	lation, and cropla	ind, abandoned	land, and agric	ultural soil organi	c carbon (SOC).				
1144	⁴ Natic	onally reported fu	el wood and inc	dustrial roundv	vood.					
1145	⁵ Natio	onally reported fu	el wood, charco	al, fuel residu	es and industrial π	.poompunc				
						57				



⁶ Long-cycle CO₂ emissions only (e.g savannas and agricultural CO₂ emissions are excluded). CO₂ emissions from peat, forests and woodland

1147 fires (as defined by Van der Werf et al., 2010).

1146

- 7 CO₂ from the combustion of organic soils.
- 8 CO₂ Forest fires + wetland/peatland fires and decay (5A, and 5D classes).
- ¹¹⁵⁰ ⁹ Humid forest deforestation fires, and peatland fires + decay.
- ¹⁰ CO₂ emissions from organic soils. Tier 1 approach. EF=20 tC.ha⁻¹.yr⁻¹ (IPCC 2006). Only for the six crop types reported by the agricultural 1151
- ¹¹⁵² soils (maize, soya, sorghum, wheat, barley, millet). N₂O emissions not included.
- ¹¹ CO₂ emissions from organic soils. Tier 1 approach. EF=20 tC.ha⁻¹.yr⁻¹ (IPCC 2006). N₂O emissions not included. 1153
- 12 CO₂ for fuelwood is part of the energy balance.
- ¹³¹³ CH4 and N2O emissions for peat, forests and woodland, savannahs and agriculture fires.
- ¹⁴ CH4, N2O emissions from fire in humid tropical forests and other forests, as well as CH4, N2O from the combustion of organic soils. 1156
- ¹¹⁵⁷ ¹⁵ CH4, N2O for forest fires + wetland/peatland fires and decay (5A, and 5D classes).
- 1158 ¹⁶ Direct agricultural emissions only
- ¹⁷ ¹⁷ Fertilizers, manure, crop residues
- ¹⁸ Synthetic fertilizers + Manure applied to soils + Crop residues + Manure applied to pastures. 1160
- 1161 ¹⁹ Indirect emissions



Published: 28 June 2016

Biogeosciences Discuss., doi:10.5194/bg-2016-244, 2016

Manuscript under review for journal Biogeosciences







	Hotspots	TATERAT	EDGAR	Houghton	Baccini	EPA	AR5
Deforestation							
Fire							
Wood-							
harvesting							
Livestock							
Cropland							
Paddy Rice							
Peatland							
Other				Forest sinks			Forest
							sinks
Table 4: summa	rv of the lea	st reliable emi	ssion sourc	es (dark orev) f	or the analysed	datasets in thi	s study

1163

1162