Anonymous Referee #1

The authors have done much to address the comments of the reviewers and the revised ms is much stronger as a result. I think that readers will benefit from seeing the exchange among authors and reveiwers.

Although the ms is clearly written, the ms still needs a close editorial check for good grammar throughout.

e.g. line 99, "needful" should be "needed". L108 "Thank for recent advances . . ." There are many examples of grammatical errors that need to be corrected. I appreciate that English is not their first language, so they need the aid of a good editor or colleague to correct mistakes.

The terminology "tidal pools" should be used instead of 'ponds'.

Edward Laws, has carefully checked and thoroughly edited the manuscript in terms of language (changes are in blue in the trackable version).

Minor comments

e.g. line 99, "needful" should be "needed".

Corrected.

L108 "Thank for recent advances . . . "

Follow Reviewer #3, we changed it to "As a result of".

The terminology "tidal pools" should be used instead of 'ponds'.

Changed.

Anonymous Referee #3

The authors have satisfactorily responded to my previous comments, though the editing has muddled the explanation of the model somewhat. I have only minor comments at this point.

Thanks for the reviewer's recognition.

line 108: this should be 'Thanks to' not 'Thanks for'. Probably better to substitute "As a result of"

As suggested, we changed it to "As a result of".

The development of the model was laid out more clearly in the previous version. I am not sure the reviewer comments that brought about this change? At the very least, explicit definition of the source of the atmospheric atm% constant term (0.00366) should be given for unfamiliar readers.

Yes, additional part for model development was added according to previous review. We add the reference of Coplen et al. (1992) for the source of the atmospheric atm% constant term (0.00366).

lines 333-334: N/P ratio below 16 doesn't necessarily mean the system is N limited. 0.5 uM [NO3-] concentration suggests it is not.

We would like to provide more background information originally to show readers P is not limiting during entire incubation. To avoid distraction, we deleted this sentence. Such deletion would not influence the logic flow and story.

lines 392-394: This sentence is unclear.

The sentence "In general, the rates of the first time interval can well predict the following up observations, demonstrating a good predictive performance by using the matrix method instant rate." was changed to "The fact that the rates during the first time interval generally predicted rather well the subsequent observations demonstrated a good predictive performance with the matrix method initial rate."

line 535: These are rate constants, not rates.

Corrected.

line 569: This comment about the author's reply to reviews should not be in the final manuscript.

We decided to add this figure into supplementary information (see Fig. S3).

line 589: I would change this to 'may have been overestimated'

Changed as suggested.

line 608-617: Double check this paragraph for correct usage of RPI, as it there are several typos where it says PRI instead.

We changed to RPI.

The verb tense for many of the sub-headings in Section 4 should be changed, e.g. 4.2.4 should say 'Quantifying' not 'Quantify'

Changed as suggested.

1	Quantification of multiple simultaneously occurring nitrogen
2	flows in the euphotic ocean
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12	

13 Abstract

14 The general features of the N cycle in the sunlit region of the ocean have been recognized are well known, but methodological difficulties have previously 15 16 confounded simultaneous quantitative quantification of information about multiple transformation rates among the many different forms of Nnitrogen pools, ie.eg., 17 18 ammonium (NH₄⁺), nitrite (NO₂⁻), nitrate (NO₃⁻), and particulate/dissolved organic 19 nitrogen (PN/DON), are insufficient due to methodological difficulties. Recent 20 However, recent advances in analytical methods methodology have made it possible to employ a convenient isotope labeling technique for isotopic composition of 21 22 oft measured nitrogen species allowed us to establish a convenient isotope labelling 23 method to quantify *-in situ* dynamic fluxes among oft-measured nitrogen species flows forwithin the euphotic waterzone. By aAddingtion of a single ¹⁵N-labelled NH₄⁺ tracer 24 25 and, we monitoring ofed the changes in in the concentrations and isotopic compositions 26 of the total dissolved nitrogen (TDN), PN, NH4⁺, NO₂⁻, and NO₃⁻ pools <u>allowed us</u> to trace-quantify the ¹⁵N and ¹⁴N flowsfluxes simultaneously. Constraints expressing the 27 balance of ¹⁵N and ¹⁴N fluxes between the different N pools were expressed in the form 28 29 of Based on mass and isotope conservations of every individual pool as well as the 30 whole system, we formulated matrix simultaneous equations, with the unique solution 31 of which via matrix inversion yielded the to simultaneously derive multiple nitrogen transformation rates relevant N fluxes, such as including rates of NH4⁺, NO₂⁻, and NO₃⁻ 32 uptake; ammonia oxidation; nitrite oxidation; DON release, and NH4⁺ uptake by 33

34	bacteria. Theis isotope matrix inversion methodology that we used was designed
35	specifically to analyze the results of incubations for euphotic water column incubation
36	under simulated in situ conditions in the euphotic zone. With By taking into
37	consideration of multisimultaneous -flowsuxes among multiple N pools, we minimized
38	potential <u>biases_artifacts</u> caused by non-targeted processes in traditional
39	source-product methods. The proposed isotope matrix method is-facilitates post-hoc
40	analysis of data convenient in terms of from on-deck incubation experiments and
41	post-hoc data analysis and is feasiblecan be used to probe effects of environmental
42	factors (e.g., pH, temperature, and light) on multiple processes under manipulated
43	controlled conditions.

44 Keywords

45 Ammonium oxidation, isotope, new production, nitrification, regenerated production46

47 **1. Introduction**

48 Nitrogen (N), which is an essential element in-for all organisms² metabolic 49 processes, regulates productivity in the surface waters of many parts of the ocean 50 (Falkowski, 1997; Zehr and Kudela, 2011; Casciotti, 2016). As a limiting nutrient in 51 the euphotic zone, nitrogen rapidly interconverts among five major N compartments: 52 particulate organic nitrogen (PN), dissolved organic nitrogen (DON), ammonium 53 (NH₄⁺), nitrite (NO₂⁻), and nitrate (NO₃⁻) (Fig. 1). Quantitative information on Studies 54 of the rates of transformation of N rates in the marine N-cycle may advance had a 55 major impact on our current understanding of the coupling of autotrophic and 56 heterotrophic processes involving carbon and nitrogen as well as the efficiency of the 57 biological pump (Dugdale and Goering, 1967; Caperon et al., 1979; Harrison et al., 58 1992; Bronk and Glibert, 1994; Dore and Karl, 1996; Laws et al., 2000; Yool et al., 59 2007). Such information would has also facilitated evaluation of ecosystem functions. 60 However, those studies have typically involved the dynamic nature and complexity of 61 the reactions involving nitrogen make it a difficult task to resolve the rates of multiple 62 simultaneous nitrogen transformations. Inventory and isotope tracer methods that have 63 often quantified the rates of only one or a few fluxes been used to measure rate of 64 specific process in previous studies (Ward, 2008, 2011; Lipschultz, 2008 and 65 references therein). The dynamic nature and complexity of the N cycle make 66 simultaneous resolution of the rates of more than a few of the important fluxes a 67 challenging task.

68	The inventory method (monitoring the change of substrate and/or product
69	concentrations over time) was has often been used to determine the uptake rates of
70	ammonium, nitrite, nitrate, and urea (McCarthy and Eppley, 1972; Harvey and Caperon,
71	1976; Harrison and Davis, 1977; Howard et al., 2007) and to examine the occurrence
72	and rate of nitrification (Wada and Hatton, 1971; Pakulski et al., 1995; Ward, 2011).
73	However, unwanted_failure to account for other processes may bias the results. For
74	example, the <u>concentration of substrate (ammonium) pool</u> is controlled simultaneously
75	by consumptions viaremoval via phytoplankton_uptake (PN as the product), nitrifier
76	<u>nitrification (nitrite/nitrate as the product)</u> , and bacterial <u>metabolism</u> (operationally
77	defined DON as product) and by additions via remineralization from heterotrophic
78	bacterial metabolism, zooplankton excretion, and viral lysis. Similarly, the products
79	(NO_{x}) pool of nitrification (NO_{x}) is may be simultaneously consumed
80	contemporarily by phytoplankton during incubation.

The ¹⁵N-labeled tracer technique has been widely used as an <u>direct measureassay</u> 81 82 for specific nitrogen processes since the emergence of isotope ratio mass spectrometry (IRMS). For example, the addition of ¹⁵N-labeled nitrate has been applied to estimate 83 84 new production (Dugdale and Goering, 1967; Chen, 2005; Painter et al., 2014). Likewise, by incubating water to which ¹⁵NH₄⁺ has been added, <u>the</u> nitrification rate 85 (¹⁵NO₃⁻ as product; e.g. Newell et al., 2013; Hsiao et al., 2014; Peng et al., 2016) and 86 ammonium uptake rate (¹⁵N_{PN} as product; e.g. Dugdale and Goering, 1967; Dugdale 87 88 and Wilkerson, 1986; Bronk et al., 1994, 2014) can be measured, respectively, with via

89	incubations in the dark and light-, respectively incubation. However, the interpretation
90	of isotope labelling experiments is confounded by the same encounters similar bias
91	problems in as the inventory method, i.e., multiple processes that occur
92	simultaneously involving eitherimpact the concentrations of source substrates or and
93	products terms in the incubation bottle. In fact, these transformations among
94	pools have significant implications in-for biogeochemical cycles. For instance, Yool et
95	al. (2007) has synthesized available global data and indicated concluded that the
96	fractional contribution of nitrate derived from nitrification to nitrate uptake can be as
97	high as 19–33% in the euphotic zone. However, integration of the relevant rates over a
98	light:dark cycle has been confounded by the fact that nitrate uptake rates were have
99	typically been determined under-during the photoperiodlight conditions, and whereas
100	nitrification <u>rates have been was determined undermeasured under</u> dark conditions (e.g.
101	Grundle et al., 2013),). Nitrate uptake may occur in the dark, but not necessarily at the
102	same rate as in the light (Laws and Wong, 1978), and nitrification is inhibited by light
103	(Dore and Karl, 1996). which are not comparable in terms of their effects on these
104	processes. To overcome this problem integrate rates over the light:dark cycle, 24-h
105	incubations have been used to compensate for the diel cycle of light-sensitive processes
106	(Beman et al., 2012). Yet, interpretation of the results of 24-h incubations may cause
107	calculationbe confounded by artifacts due to the interference from significant transfers
108	of ¹⁵ N and ¹⁴ N among pools. A new method is needful needed to reconcile the
I	

109 above-mentioned biases and the uncomparable parallel incubationsovercome these
110 problems.

111 Marchant et al. (2016) have reviewed recent methodological advances in marine 112 N-cycle studies using ¹⁵N-labeling substrates combined with nanoSIMS, FISH, or HISH in marine N-cycle studies. These methods provide qualitative information for 113 114 about N transfers at the cellular and molecular level but do not quantitative-quantify 115 rates at the community level. Elskens et al. (2005) conducted A-a comprehensive 116 review of oft-used models for rate derivation was conducted by Elskens et al. (2005), 117 whoand concluded that oversimplified models would risk may lead to biased when 118 results if their underlying assumptions are violated. <u>neverthelessHowever</u>, overly 119 complex models risk could misinterpreting part of the random noise as relevant 120 processes. To address this concern, De Brauwere et al. (2005) Therefore, proposed a 121 model selection procedure was subsequently proposed (De Brauwere et al., 2005). 122 More recently, Pfister et al. (2016) have applied an isotope tracer technique and mass 123 conservation model onto tidal pools study to explore nitrogen flows among dissolved 124 nitrogen pools (NH4⁺, NO₂⁻-, and NO₃⁻) in tidal ponds-pools and found that benthic 125 macrobiota playeds an important role in regulating remineralization flowrates. They 126 also proved found that the dilution effects significantly biaseded the results obtained by 127 with source-product models. Nevertheless, fFor the euphotic zone, where competing 128 processes co-occur, an innovative and convenient method is needed to determine the

129 <u>rates of multiple N fluxes from the results of for on-decksimulated *in-situ* incubations
130 to measure *in situ* multiple N flows is needed.
</u>

131 In this study, we propose an "isotope matrix method". To avoid perturbations, the 132 concentration of the tracer was limited to < 10% or 20-% of the substrate concentration, 133 as suggested by previous researchers (Raimbault and Garcia, 2008; Middelburg and 134 Nieuwenhuize, 2000; Painter et al., 2014). One single tracer, ¹⁵NH₄⁺, was added into an 135 incubation bottle to trace the ¹⁵N flow among the nitrogen pools under simulated *in situ* 136 conditions. Almost all the most fundamental processes in the N cycle can be quantified 137 with this newly proposed method. To demonstrate the applicability of the method, we 138 conducted incubation experiments for-with low-nutrient water fromin the western 139 North Pacific and for-with high-nutrient coastal water off the southeastern China coast. 140 Thanks for As a result of recent advances in these analytical methods for measuring the 141 concentrations and isotopic compositions of various nitrogen species, we were able to 142 use this isotope matrix method becomes applicable to quantify the in situ- dynamic fluxes of nitrogen flows for N in the euphotic zonewater. 143

144 **2. Isotope matrix method**

145 **2.1 Framework of the inter-connections among nitrogen pools**

146 <u>Figure 1 shows In the oxygenated and well-lit euphotic zone</u>, the transformations 147 of N among NH_4^+ , NO_2^- , NO_3^- , PN, and DON <u>in an aerobic euphotic zoneare shown in</u> 148 Fig. 1. The PN <u>is-was</u> operationally defined as the particulate organic nitrogen trapped 149 on a GF/F filter (> 0.7_µm). Dissolved inorganic nitrogen (DIN) and DON <u>are-were</u> 150 <u>equated to</u> the inorganic and organic nitrogen, respectively, in the dissolved fraction 151 that passe<u>ds</u> through a polycarbonate membrane with a 0.22 μ m pore size.<u>Since</u> 152 <u>Because</u> DON includes the N in numerous dissolved organic N compounds, including 153 unidentified organics, urea, amino acids, amines, and amides, DON represents the 154 "bulk" DON and <u>is-was</u> calculated by subtracting the concentrations of NH₄⁺, NO₂⁻, 155 and NO₃⁻ (DIN) from the total dissolved N (TDN).

We consider-used two different types of schemesmodels in to analyze our method<u>data</u>: a low-nitrogen-nutrient model to represent the open ocean and a highnutrient model to represent estuarine and coastal environmentsitrogen (Fig. 1a and 1b). The low nutrient scheme is for the open ocean. In tThe high-nutrient scheme-model, is for estuary and coastal environments where NH₄⁺, NO₂⁻, and NO₃⁻ were assumed tothree dissolved inorganic nitrogen species co-exist. Below, we describe tThe rationale for the twoof model structures is as follows for the two cases.

163 The consumption of reactive inorganic nitrogen $(NH_4^+, NO_2^-, and NO_3^-)$ is 164 dominated by photosynthetic uptake by phytoplankton (F1 and F4 in Fig. 1a; F1, F3, 165 and F5 in Fig. 1b). Heterotrophic bacteria may also be play an important actors forrole in NH4⁺ assimilation (Laws, 1985), and was confirmed by studies later on (e.g.; 166 167 Middelburg and Nieuwenhuize, 2000; Veuger et al., 2004). We took heterotrophic 168 bacterial assimilation of <u>HNH4</u> into account as well (F6 in Fig. 1a and F8 in Fig. 1b) to 169 explore its importance. Though NO₂⁻ may be released during NO₃⁻ uptake (Lomas and 170 Lipschultz, 2006), little NO₂⁻ production from NO₃⁻ was detected by (Santoro et al.,

171	(2013)., especially in high NH_4^+ estuary and coastal sea, nN itrate assimilation may be
172	inhibited in oxygenated aerobic water water, especially in estuaries and coastal seas
173	where the NH ₄ ⁺ concentration is high, and in the absence of nitrate uptake, there is no
174	release of , subsequently, so is the nitrite release. Thus, the nitrite release was ignored
175	in our model. Due to DIN assimilation by phytoplankton, the PN pool may increase,
176	but DON may be released during assimilation (F5 in Fig. 1a and F7 in Fig. 1b) as
177	indicated by previous studies noted by (Bronk et al., (1994;-), Bronk and Ward,
178	(2000;-), and Varela et al., (2005). On the other hand, rThe size of emineralization may
179	refuelthe NH4 ⁺ pool_is increased by remineralization (F2 in both Fig. 1a and 1b)-
180	Meanwhile, ammonium pool is reducedand decreased by nitrification. The latter
181	process, which consists of two basic steps: the ammoniaum oxidation by
182	archaea/bacteria (AOA/AOB) to nitrite (F4 in Fig. 1b) and the-nitrite oxidation to
183	nitrate by nitrite-oxidizing bacteria (NOB) (F6 in Fig. 1b). Although recent studies have
184	revealed a single microorganism that $\frac{\text{may}}{\text{can}}$ completely oxidize NH ₄ ⁺ to NO ₃ ⁻
185	(comammox comammox) (Daims et al., 2015; van Kessel et al., 2015), its-the
186	importance of comammox in the marine environment remains unclear.

187 Specific mechanisms or processes such as grazing and viral lysis may alter the 188 concentrations of NH₄⁺, nitrite, and DON. However, the scope of this study is to 189 determine the nitrogen flows and exchanges<u>fluxes</u> among the often_-measured and 190 operationally defined nitrogen pools. In this context, grazers and viruses belong to the 191 operationally defined PN and DON pools, respectivelyThe organisms that mediate the

192 relevant fluxes are not specifically included in the model. Thus, the results ance of 193 specific process such as grazing and viral lysis <u>has have</u> been incorporated into the 194 paradigm depicted in Fig.ure 1.

195 **2.2** Analytical methods to determine the amounts of ¹⁵N/¹⁴N in various pools

To trace the ¹⁵N movement among pools, our isotope matrix method couples the 197 ¹⁵N-labellinglabeling and inventory methods through by considering changes of both concentrations and isotopic compositions - changes. Analytical methods to determine the concentrations and isotopic compositions of both high and low levels of inorganic/organic nitrogen are in most cases well established and have been reported elsewhere. We determined all of the mentioned-relevant concentrations and isotopic compositions with the exception of the isotopic composition of NH₄⁺.

203 Concentrations of NH₄⁺ higher than 0.5 μ M were measured manually by using the 204 colorimetric phenol hypochlorite technique (Koroleff, 1983). Nanomolar NH₄⁺ 205 concentrations were measured by using the fluorometric o-phthaldialdehyde (OPA) 206 method (Zhu et al., 2013). Concentrations of NO_2^- and of NO_x^- ($NO_2^- + NO_3^-$) were 207 determined with the chemiluminescence method following the protocol of Braman and 208 Hendrix (1989). The detection limits of NO₂⁻ and NO_x⁻ were both ~ 10 nmol L⁻¹, and 209 the corresponding relative precision was better than 5% within the range of 210 concentrations that we measured. By using persulfate as an oxidizing reagent, we 211 oxidized TDN and PN separately to nitrate (Knapp et al., 2005) and then measured the 212 nitrate by using the analytical method for NO_x^- described above.

213	We determined the $\delta^{15}N$ of NO_2^- with the azide method by following the detailed
214	procedures in McIlvin and Altabet (2005). The $\delta^{15}N$ of NO_x^- was determined by using a
215	distinct strain of bacteria that lacked N2O reductase activity to quantitatively convert
216	NO_x^- to nitrous oxide (N ₂ O), which we then analyzed by IRMS (denitrifier method;
217	(Sigman et al., 2001; Casciotti et al., 2002). The isotopic composition of NO_3^- was
218	determined from isotope mass balance (NO _x ⁻ minus NO ₂ ⁻) or measured by the
219	denitrifier method after eliminating preexisting NO_2^- with sulfamic acid (Granger and
220	Sigman, 2009). To determine the $\delta^{15}N$ of TDN and PN, both species were first
221	converted to NO_3^- with the denitrifier method, and then the $\delta^{15}N$ of the NO_3^- was
222	determined as described above. The detection limit of $\delta^{15}N_{PN}$ can be reduced to the
223	nanomole-nanomolar level (absolute amount of nitrogen), which is significantly lower
224	than that the detection limit by using high temperature combustion with an elemental
225	analyzer connected to IRMS.

226 The most popular way to determine the N isotopic composition of NH₄⁺ is the 227 "diffusion method", which involves conversion of dissolved NH₄⁺ to NH₃ gas by 228 raising the sample pH to above 9 with magnesium oxide (MgO) and subsequently 229 trapping the gas quantitatively as (NH₄)₂SO₄ on a glass fiber (GF) filter; the isotope ratios of the ${}^{15}N/{}^{14}N$ are then measured using an <u>coupled</u> elemental analyzer <u>coupled</u> 230 231 with an IRMS (Holmes et al., 1998; Hannon and Böhlke, 2008). Alternatively, after 232 removing the preexisting NO₂⁻ from the seawater samples using sulfamic acid, NH₄⁺ is 233 first quantitatively oxidized to NO₂⁻ by hypobromite (BrO⁻) at pH ~12 (BrO⁻ oxidation method), and the protocol of McIlvin and Altabet (2005) is then used to reduce the NO₂⁻ to N₂O (Zhang et al., 2007). Unfortunately, neither of these methods has been established in our lab yet. The isotope matrix method requires the isotopic composition of NH_{4^+} as well, but this requirement can be circumvented by making certain assumptions, as illustrated in our case studies below.

We estimated the amount of ¹⁴N and ¹⁵N atoms in every individual pool for which 239 we knew the concentration and $\delta^{15}N$ ($\delta^{15}N \% = [(R_{sample} - R_{atmN2})/R_{atmN2}] \times 1000$). By 240 assuming the ¹⁵N content of standard atmospheric nitrogen to be 0.365% (Coplen et al., 241 1992), we calculated R_{sample} (¹⁵N/¹⁴N). By defining r_{sample} as ¹⁵N/(¹⁴N+¹⁵N), we directly 242 derived the ¹⁵N and ¹⁴N concentrations of all forms of N, with the exception of NH₄⁺ 243 244 and DON. The r value of the NH₄⁺ was assumed to equal either its initial value or an arbitrarily chosen fraction thereof, and the ¹⁵N and ¹⁴N content of the NH₄⁺ was then 245 246 determined.

247 **2.3 Formation of matrix equations**

In this isotope matrix method, we added <u>a</u>_limited amount of ¹⁵NH₄⁺ into incubation bottles at the very beginning and then monitored the changes of ¹⁵N and ¹⁴N in the measured pools every <u>a</u>-few hours. We assumed isotopic mass balance at every time point in the incubation bottle. In other words, the sums of the variations in the total N, ¹⁵N, and ¹⁴N concentrations were zero for any time interval. We assumed no fractionation between ⁴⁵N and ⁴⁴N for all the transfer reactions among the pools. The fluxes of ¹⁵N and ¹⁴N were therefore equal to the total flux multiplied, respectively, by r_{substrate} and $(1 - r_{substrate})$, <u>respectively</u>. Note that w<u>Although we</u>e did not consider isotope fractionation, though it could <u>have been easily be</u> introduced into the equations if <u>necessary</u>, <u>i.e., by</u> dividing the ¹⁴N flux by a (the ratio of <u>the</u> specific rate constants of ¹⁴N to <u>and</u> ¹⁵N) to obtain, and the flux of ¹⁵N is obtained. Below, we illustrated equations for the two model cases.

According to mass balance, the net changes of the ¹⁵N (or ¹⁴N) concentration of 260 261 an individual N pool in <u>a certain</u> time interval are determined by the inflow and outflow of ${}^{15}N$ (or ${}^{14}N$) (see Fig. 1 and Eqs. 1–14 below). In the low-nitrogen case, the changes 262 of the ¹⁵N concentrations of the NH_4^+ , NO_x^- , and PN pools were expressed by Eq. 1, 2, 263 and 3, respectively. Similarly, the temporal dependence of ${}^{14}N-NH_4^+$, ${}^{14}N-NO_x^-$, and 264 265 14 N-PN were expressed by Eq. 4, 5 and 6, respectively. The mean rate of change in-of 266 the nitrogen pool, i.e. the left side of the each equation, was determined from the data at 267 time zero (t0) and the first time point (t1). For example, when the sampling time interval is-was short, $\Delta [^{14}NH_4^+]/\Delta t$ at the first time point was approximately $\{[^{14}NH_4^+]_{t1}\}$ 268 $- [^{14}NH_4^+]_{t0}]/(t1 - t0)$ where the subscripts indicate the times at which the 269 270 concentrations were measured. The r value applied in the each equation for substrate 271 was the average of the r values for the pool at time zero and the first time point after that 272 for measured pool.

273
$$\frac{\Delta \left[{}^{15}NH_4^{+} \right]}{\Delta T} = \overline{F_2} \times 0.00366 - \overline{F_1} \times \overline{F_{NH_4^+}} - \overline{F_3} \times \overline{F_{NH_4^+}} - \overline{F_6} \times \overline{F_{NH_4^+}}$$
(1)

274
$$\frac{\Delta \left[{}^{15}NO_{x}^{-} \right]}{\Delta T} = \overline{F_{3}} \times \overline{r_{NH_{4}^{+}}} - \overline{F_{4}} \times \overline{r_{NO_{x}^{-}}}$$
(2)

17 / 61

275
$$\frac{\Delta \left[{}^{15}PN \right]}{\Delta T} = \overline{F_1} \times \overline{F_{NH_4^+}} + \overline{F_4} \times \overline{F_{NO_x^-}} - \overline{F_5} \times \overline{F_{PN}}$$
(3)

276
$$\frac{\Delta \left[{}^{14}NH_4^{+} \right]}{\Delta T} = \overline{F_2} \times (1 - 0.00366) - \overline{F_1} \times (1 - \overline{F_{NH_4^+}}) - \overline{F_3} \times (1 - \overline{F_{NH_4^+}}) - \overline{F_6} \times (1 - \overline{F_{NH_4^+}}) \quad (4)$$

277
$$\frac{\Delta \left[{}^{14}NO_{x}^{-} \right]}{\Delta T} = \overline{F_{3}} \times (1 - \overline{r_{NH_{4}^{+}}}) - \overline{F_{4}} \times (1 - \overline{r_{NO_{x}^{-}}})$$
(5)

278
$$\frac{\Delta \left[{}^{14}PN \right]}{\Delta T} = \overline{F_1} \times (1 - \overline{F_{NH_4^+}}) + \overline{F_4} \times (1 - \overline{F_{NO_x^-}}) - \overline{F_5} \times (1 - \overline{F_{PN}})$$
(6)

279 In this study, we conducted a The time series monitoring for overin this study 280 lasted for 24 hours, ... however However, we took used only the first two time points for 281 the rate calculations since suchbecause we felt those rates derivations mightwould be 282 be more closest to the instantaneous in situ rates in of the original 283 environments samples. Note: researchers may apply Although this the isotope matrix 284 method onto may be applied to longer time intervals, however, rates may vary as a 285 result of substrate consumption and/or community change,. Relatively shorter-term 286 incubations is are therefore advisable suggested (see below).

287 <u>Since_Because_the total number of equations and unknowns are equal, a unique</u>
288 solution therefore can be obtained via matrix <u>solution_inversion_for the low_</u>-nutrient
289 model.

In high_-nutrient cases, similar<u>analogously</u>, equations (Eqs. 7–14) can be constructed <u>by-to describe the using transformations amongfluxes between</u> NH_4^+ , NO_2^- , NO_3^- -, and PN (Fig. 1b).

$$293 \qquad \frac{\Delta \left[{}^{15}NH_4^{+} \right]}{\Delta T} = \overline{F_2} \times 0.00366 - \overline{F_1} \times \overline{F_{NH_4^+}} - \overline{F_4} \times \overline{F_{NH_4^+}} - \overline{F_8} \times \overline{F_{NH_4^+}} - (7)$$

294
$$\frac{\Delta \left[{}^{15}NO_2^{-} \right]}{\Delta T} = \overline{F_4} \times \overline{F_3} \times \overline{F_3} \times \overline{F_6} \times$$

297
$$\frac{\Delta \left[{}^{14}NH_4^{+} \right]}{\Delta T} = \overline{F_2} \times (1 - 0.00366) - \overline{F_1} \times (1 - \overline{r_{NH_4^+}}) - \overline{F_4} \times (1 - \overline{r_{NH_4^+}}) - \overline{F_8} \times (1 - \overline{r_{NH_4^+}}) \quad (11)$$

$$300 \qquad \frac{\Delta \left[{}^{14}PN \right]}{\Delta T} = \overline{F_1} \times (1 - \overline{r_{NH_4^+}}) + \overline{F_3} \times (1 - \overline{r_{NO_2^-}}) + \overline{F_5} \times (1 - \overline{r_{NO_3^-}}) - \overline{F_7} \times (1 - \overline{r_{PN}})$$
(14)

Again, a<u>A</u> unique solution can <u>again</u> be obtained <u>via matrix inversion since</u>
 <u>because</u> the numbers of equations and unknowns are equal.

In the above matrix equations, <u>the value of r_{NH4+} </u>, which we did not measure in this study, <u>is-was necessary for theneeded to obtain a</u> solution. <u>Here-To address this issue</u>, we <u>set-assumed</u> various degrees of remineralization to test the effect of isotope dilution (NH₄⁺ addition) <u>in-on</u> our <u>experimental cases_calculated fluxes</u>. We reduced r_{NH4+} values at a constant reduction rate and the total reduction of r_{NH4+} was 0%, 1%, 10%, 20% and

308 50% for the full time span of the 24-h incubation. (The r_{NH4+} of for remineralization (F2) 309 is-was assumed to be constant (0.00366) and equal constant rates that led to total 310 reductions of r_{NH4+} by 0%, 1%, 10%, 20%, or 50% by the end). The value of F2 coupled 311 with given the assumed r_{NH4+} values allowed us to resolve rates under different 312 remineralization conditionsscenarios, and the derived F2 was introduced into a 313 STELLA model for extrapolation purposess (see below). We compared the observed and remineralization-associated simulations to reveal elucidate the effect of 314 remineralization on the calculated rates measure for the time series incubations. 315

316 **2.4 Validation by STELLA**

317 As the aforementioned, tThe "instant rate" initial rates are of particular interest 318 because they are presumably most similar to the *in situ* rates at the time the sample was 319 collected at the original condition is what researchers pursue. Note that the use of The 320 initial rate is here distinguished <u>"instant" here is just to make it distinguishable from</u> 321 rates derived from the longer time-incubations that extended beyond time point t1or 322 more than two time points. To evaluate the applicability of the matrix-derived instant 323 initial rate, here we applied used STELLA 9.1.4 software (Isee systems, Inc.) to 324 construct box models that were consistent with the scenarios depicted in Fig.ure 1. The constructed STELLA model contained two modules (Figs. S1 and S2), one for ¹⁵N and 325 326 the other for ¹⁴N. The connection between these two modules was were connected through the ¹⁵N atom % (rN), which was a measured parameter measured in the 327 incubation experiment. The A model started to run was initialized with these measured 328

329 initial values for of the nitrogen pools at time zero, and the model then to-projected 330 continuous the changes values of corresponding those nitrogen pools as a continuous 331 function of time. Since Because the rates numbers based on the first two time points 332 may might not accurately represent the behavior of the system throughout not guarantee a good performance for the full time course due to system variation, for example, i.e., to 333 334 changes in substrate concentrations and the composition of the microorganism 335 microbial community, we took this model practice (extrapolation using the initial rates) 336 as amounted to a test of the hypothesis that the rates did not changea validation. 337 In this study, wWe assumed the first-order reaction kinetics for in both the 338 low-nutrient and high-nutrient cases., thus, tThe initial rate constant "k" can-could therefore be derived via by dividing the matrix-derived flux F by \overline{C} , (the average 339 340 substrate concentration of during the first two time points). After setting the initial the concentrations of ¹⁵N and ¹⁴N were initialized to-in every pool, the model ran for 24 h 341 342 according tousing the matrix-derived short-term k values. As depicted in Fig.ure 1, all 343 these monitored N pools are-were regulated by F, which is-was assumed to be 344 concentration dependent according to our assumption (Figs. S1 and S2). The output of the model includes included the time courses of the ¹⁵N and ¹⁴N concentrations and the 345 346 ¹⁵N atom % (r_N) of each N species. Through this comparison analysis, we could observe 347 the <u>course temporal</u> evolution of the isotopic composition <u>in of the</u> various N pools.

348 **2.5 Study sites and incubation experiments**

Incubation-The incubation experiments were conducted in two environments with
very different nutrient levels. The low-nutrient <u>case study</u> was conducted on-deck of
the R/V Dongfanghong 2 on a cruise to the Western North Pacific (WNP) (33.3 N,
145.9 E) in the spring of 2015. The site for-of the high-nutrient <u>case study is was in the</u>
Wuyuanwan Bay (WYW) (24.5 N, 118.2 E) in the southern coast of China.

354 The water samples at the WNP station were collected using a 24-bottle rosette 355 sampler. The sampling depth was 25 m, at which the with moderate light intensity was 356 (12% of the surface water irradiance). Two pre-washed 10-L polycarbonate carboys (Nalgene, USA) were used for the incubation. A total of 1.5 mL of 200 µM¹⁵N-labelled 357 358 NH₄Cl tracers containing 98 atom % ¹⁵N (Sigma-Aldrich, USA) was injected into each 359 incubation bottle separately to achieve a final concentration of 30 nM. Incubation-The 360 incubation was carried out immediately with a constant simulated light intensity of (35 361 μ mol <u>E-photons</u> m⁻² s⁻¹) in a thermostatic incubator (GXZ-250A, Ningbo) at the *in situ* 362 temperature.

The WYW station is-was an-located in the inner bay, where the tide was with a regular semidiurnal tide. As a coastal bay, Wuyuanwan, a coastal bay, suffers from the same anthropogenic influences that <u>cause eutrophication result in high nutrient</u> concentrations analogous toin other coastal zones areas in of China. However, the bay water is still-well ventilated and constantly saturated with dissolved oxygen due to tidally induced water exchange. It is an ideal research site to study the dynamic transformations that characterize processes of the coastal nitrogen cycle.

370 The WYW samples were taken on 19 January, 2014 from water depths of 0.3 m 371 and 2.3 m, where the , respectively, with a light intensity intensities of were 80-% and 372 2%, respectively, of the surface water irradiance. Duplicate water samples were 373 collected for from each depth by using submersible pump into to fill pre-washed 10-L 374 polycarbonate bottles (Nalgene, USA). ¹⁵N-labeled NH₄Cl (98 atom % ¹⁵N, 375 Sigma-Aldrich, USA) was added to the incubation bottles with to a final concentration 376 of 1 μ M (~4–% of the ambient concentration). The incubations were carried out 377 immediately in the field. Neutral density_screens that allows allowed 80% and 2% light 378 penetration was applied, respectively were used to simulate the light intensities at 0.3 m 379 and 2.3 m, respectively, for incubation bottles of shallow and deep samples. The 380 temperature was maintained at ~13.7 % by continuously pumped pumping seawater 381 through the flowincubators.

382 The <u>Sample sample of at</u> the first time point (t0) was taken immediately after 383 tracer addition. Subsequent samples were taken at approximately 2–4 h intervals -for 384 DIN and PN analyses. An aliquot of 200 mL was filtered through a 47-mm 385 polycarbonate membrane filter with a 0.22 μ m pore size (Millipore, USA). and tThe 386 filtrates were frozen at -20 °C for chemical analyses in the lab. Particulate matter was 387 collected by filtering 500 ml seawater through pre-combusted (450 °C for 4 h) 25 mm 388 GF/F filters (Whatman, GE Healthcare, USA), under at a pressure of <100 mm Hg. The 389 GF/F filters were freeze-dried and stored in a desiccator prior to for-analysis of PN 390 concentrations and $\frac{15N \text{ atom } \%}{15N \text{ sotopes}}$.

391 3. Results

392 3.1 Ambient conditions and initial concentrations

The water temperature and salinity <u>at a depth of 25 m of in</u> the WNP low nutrient case from 25m waswere 18.4 °C and 34.8, respectively. The dissolved oxygen (DO) was 7.3 mg L⁻¹. The concentrations of NH₄⁺, NO_x⁻⁻, and phosphate was-were 113 ±5 nmol L⁻¹, 521 ± 18 nmol L⁻¹ and 74 ± 2 nmol L⁻¹, respectively. The fact that the N/P ratio was lower less than 16, indicated that ing the system is was N limited.

The water temperature and salinity of thethroughout the water column of the WYW whole water column for high nutrient case wasere 13.5 ±0.1 °C and 29.5 ±0.1, respectively. The DO saturation fell in the ranged 135–140%. The concentrations of nitrogenous species were relatively high, with The inorganic nutrient concentrations of were 30.9 ±0.7 μ mol L⁻¹ for NO₃⁻, 22.3 ±4.3 μ mol L⁻¹ for NH₄⁺, 5.4 ±0.2 μ mol L⁻¹ for NO₂⁻, and 1.5 ±0.1 μ mol L⁻¹ for phosphate. , and The PN concentration was 9.3 ±0.7 μ mol L⁻¹ for PN. Phosphate was 1.5 ±0.1 μ mol L⁻⁴.

405 **3.2 Time-courses of incubations**

406 **3.2.1 Low nutrient case in the WNP**

The observed variation patterns of change for of the bulk NH₄⁺, NO_x⁻, PN, and TDN concentrations and the δ^{15} N of NO_x⁻ and PN during the incubation are shown in Figure 2. Concentrations of NH₄⁺ and NO_x⁻ decreased rapidly from 143 ± 5 to 48 ± 5 nM and 521 ± 18 to 127 ± 11 nM, respectively (Figs. 2a and 2b). In contrast, the PN concentration increased from 437 ± 9 to 667 ± 14 nM (Fig. 2c), and the TDN 412 concentration remained stable, with an average of 6511 ±209 nM (Fig. 2d). Opposite In 413 contrast to to the trend of NO_x^- concentration, $\delta^{15}N-NO_x^-$ increased from 8.9 ±0.2 to 414 171 ±2 ‰ (Fig. 2e). In addition, $\delta^{15}N-PN$ exhibited great changes, increasing from 415 46.8 ±0.2 to 6950 ±314 ‰ (Fig. 2f).

416

3.2.2 High nutrient case in the WYW s

417 The time-series of observational parameters for samples from depths of 80% and 2% 418 sPAR of surface PAR (sPAR) exhibited similar variation trends during the incubation 419 (Fig. 3). During the course of the incubation, NH_4^+ decreased significantly and 420 continuously from 26.6 ± 0.1 (initial concentration) to 17.4 $\pm 0.1 \mu$ mol L⁻¹-. with a The mean reduction rate of-was 0.63 μ mol L⁻¹ h⁻¹ for-for the 80% sPAR sample (Fig. 3a). 421 422 Compared with that of 80% sPAR, The NH4⁺ concentration of 2% sPAR sample 423 decreased slower-more slowly from 24.6 ± 0.1 (initial concentration) to 18.2 $\pm 1.0 \,\mu$ mol L^{-1} with a mean reduction rate of 0.47 μ mol L^{-1} h⁻¹ (Fig. 3a). NO₃⁻ in 80% and 2% 424 425 sPAR samples decreased from 30.1 \pm 0.1 to 28.3 \pm 0.1 μ mol L⁻¹ and from 31.1 \pm 0.1 to 426 29.7 \pm 0.1 µmol L⁻¹, respectively (Fig. 3c). Overall, the nitrate reduction rates were 427 much lower than that of the NH4⁺- reduction rates. Compared to nitrate, NO₂⁻ displayed 428 even slower rates of declining trendse, yet with the rate was significantly higher rate for at 80% sPAR sample relative to that of than at 2% sPAR sample (Fig. 3b). Similar to the 429 430 low nutrient case, PN increased steadily from 8.8 ± 0.1 to 17.7 $\pm 0.9 \,\mu$ mol L⁻¹, with a mean rate of 0.61 μ mol L⁻¹ h⁻¹ for at 80% sPAR sample and from 9.9 \pm 0.1 to 16.0 \pm 2.0 431 μ mol L⁻¹ with a mean rate of 0.44 μ mol L⁻¹ h⁻¹ in-at 2% sPAR (Fig. 3d). The rates of 432

increase <u>rates inof the PN</u> concentration were very close to the <u>rates of decrease rates of</u> NH₄⁺⁻, <u>the indicating indication being that ammonium was the major nitrogen source</u> for growth. The TDN concentration decreased from 78.7 ± 1.6 to 68.4 ± 0.1 μ mol L⁻¹ and <u>form-from 72.8 ± 2.5 to 67.1 ± 0.8 μ mol L⁻¹ <u>at for 80%</u> and 2% sPAR-<u>samples</u>, respectively (Fig. 3e).</u>

438 The δ^{15} N-NO₂⁻ increased from -9.0 ±0.1 to 12.1 ±0.1 ‰ and from ---8.8 ±0.1 to 439 23.3 ± 0.6 % in-at 80% and 2% sPAR-incubation, respectively (Fig. 3g);). Since Because the nitrate pool was relatively large, the values of δ^{15} N-NO₃⁻ ranged from 6.8 440 to 10.1 % with no significant trend over time (Fig. 3h). In addition, δ^{15} N-PN increased 441 442 from 14.8 ± 0.3 to 3078 ± 180 ‰ and from 15.0 ± 0.5 to 2738 ± 66 ‰ for at 80% and 2% 443 sPAR-sample, respectively (Fig. 3i). These significant changes in both the 444 concentration and isotopic compositions of the nitrogen pools over time suggested 445 suggests that there were significant movements of nitrogen transformed significantly among pools and that the labelled ¹⁵N in the NH₄⁺ flowed-moved throughout the 446 447 system, with the exception of nitrate.

448 **3.3 Solutions of the matrix equation and STELLA extrapolation**

449

3.3.1 Low nutrient case

The matrix-derived rate constants (ki) and rates (Fi) are shown in Table 1(A) and 1(B), respectively. Under <u>the</u> no remineralization condition (i.e. r_{NH4+} decreased 0% within 24 hours), <u>the-NO_x⁻</u> uptake (k4 = 0.059 h⁻¹; F4 = 27.2 nmol L⁻¹ h⁻¹) was the highest among all <u>forms of inorganic nitrogen</u> in terms of flux, followed by NH₄⁺

454	uptake (k1 = 0.038 h ⁻¹ ; F1 = 4.9 nmol L ⁻¹ h ⁻¹) and DON release (k5= 0.024 h ⁻¹ ; F5 =
455	11.5 nmol $L^{-1} h^{-1}$). NH ₄ ⁺ uptake by bacteria (k6 = 0.007 h ⁻¹ ; F6 = 1.0 nmol $L^{-1} h^{-1}$) was
456	much lower than that by phytoplankton. The rate constant for nitrification ($k3 = 0.0005$
457	h^{-1}) was the lowest among all fluxes (F3 = 0.07 nmol L ⁻¹ h ⁻¹).
458	By introducing the initial ${}^{15}N$ and ${}^{14}N$ concentrations of NH_4^+ , NO_x^- , PN_x and
459	DON and the calculated rate constants (k1 to k6) into STELLA (Fig. S1), we obtained \underline{a}
460	full time courses for all parameters (Fig. 4). Generally, the model outputs fitted well
461	with the measured values, except for the last time point for $PN_{, the}$ associated ¹⁵ N
462	concentration, δ^{15} N, and r_N (Figs. 4 c, k and o). In general, The fact that the rates of
463	during the first time interval can well generally predicted rather well the following
464	upsubsequent observations, demonstrateding a good predictive performance by
465	using with the matrix method instant-initial rate. Since Because the concentrations of
466	both substrates, e.g., ammonium and NO _x ⁻ were, fitted described well within during the
467	24h_experimentours, the extra non-fitted_PN_that was not well described in
468	observations after the time point of 12 hours likely indicated reflected the participation
469	influence of an additional nitrogen source, i.e., dissolved organic nitrogen that was
470	utilized by phytoplankton (see discussion below) when inorganic nitrogen reached
471	threshold levels (Sunda and Ransom, 2007).
472	In these test runs of regeneration with r_{NH4+} reduction by a total amount of 1-%,

- 473 10-%, 20-% and 50-%, we found that the NH_4^+ consumption rates (k1 and k6) increased
- 474 as the regeneration (k2) increased (Table 1). As indicated in previous studies, such

475 regeneration-induced isotope dilution indeed altered the original results (Table 1 and 476 Fig. 4). More sSpecifically, greater NH_4^+ regeneration resulted in larger differences between these three PN-associated values (¹⁵N-PN, δ^{15} N-PN, and r_{PN}) and the 477 478 STELLA-projected data (Figs. 4 c, k and o). Meanwhile, tThe dilution effect was more 479 significant after 12 hours of incubation. On the other handIn contrast, the effect of r_{NH4+} 480 on NO_x^- -associated parameters associated with NO_x^- was trivial (Figs. 4 b, f, j, n and r). 481 The comparison between the simulation and observation suggested that NH4⁺ 482 regeneration needs to be considered for PN (i.e., uptake) when the remineralization rate 483 is high and the incubation prolongs is longer than 12 hours. Besides remineralization, 484 offsets discrepancies along the time course might possibly be induced caused by by 485 changes in the community change composition of the microbial community as the 486 incubation prolongscontinues.

487

3.3.2 High nutrient cases

488 The results atof 80% sPAR and 2% sPAR under on the assumption of a fixed r_{NH4+} 489 are shown in Table 2(A) and 2(B), respectively. For the high light sample (80 %) sPAR sample), the NH₄⁺ uptake by phytoplankton (F1, 397 nmol L^{-1} h⁻¹) and by 490 bacteria (F8, 282 nmol $L^{-1} h^{-1}$) were much higher than the other rates and were 491 492 followed by the NO₃⁻ uptake rate (F5, 149 nmol $L^{-1} h^{-1}$). The NO₂⁻ uptake (F3) rate was 493 29 nmol L^{-1} h⁻¹, much lower than that of NH₄⁺ and NO₃⁻-uptake. The ammonia oxidation rate (F4) was 0.4 nmol $L^{-1} h^{-1}$, and the nitrite oxidation rate (F6) was zero 494 495 (Table 2A). Since-Because this incubation was conducted in winter with low

496	temperature and <u>under_at_80%</u> sPAR-light conditions, low rates of ammonium and
497	nitrite oxidation were reasonable because both nitrifiers and NOB are sensitive to light
498	(e.g. Olson, 1981a, 1981b; Horrigan et al., 1981; Ward, 2005; Merbt et al., 2012; Smith
499	et al., 2014). The DON release rate by phytoplankton (F7) was zero in this case.
500	In comparison, all the rates in the condition of at 2% sPAR showed a very similar
501	pattern (Table 2B). The only difference was that all the uptake rates were lower for
502	theat 2% sPAR, except for ammonia oxidation, which was higher in the low light.
503	By introducing initial concentrations and calculated rate constants (k1-k8) into the
504	STELLA model (Fig. S2), we obtained successive variations a time series of ¹⁵ N and
505	¹⁴ N concentrations and <u>the</u> r_N <u>values for of NH4</u> ⁺ , NO ₂ ⁻ , NO ₃ ⁻ , PN and DON over time
506	(Fig. 5). In general, the modeled and measured values remained consistent throughout
507	the 15h incubation, demonstrating the capability of the isotope matrix method.
508	Similar to the low_nutrient case, we evaluated the effect of regeneration (see
509	Table 2 and Fig. 5A and 5B). Since Because ammonium uptake was the dominant
510	process, the alterationchanges of the PN pool wereas more significant in comparison
511	with the other pools (Figs. 5 d, n and s). We found again that, as F2 increased, F1 and
512	F8 increased to maintain a constant reduction of the measured $\mathrm{NH_{4^+}}$ concentration
513	(Table 2). Similar to the low-nutrient case, as regeneration increased, the projected
514	course of ¹⁵ N-PN deviated more from observations, and the turning point also appeared
515	earlier, resulting in a larger curvature of r-PN and $\delta^{15}\text{N-PN}$ (Fig. 5d and 5s). This

modeling exercise confirmed the influence of the isotope dilution effect.; <u>Hhowever</u>,
the effect wasis insignificant in the early stage-part of the incubation.

518 **4. Discussion**

519 4.1 Method comparisons

520 **4.1.1 Model structure and rate derivation**

521 The most widespread ¹⁵N model was proposed by Dugdale and Goering (1967), 522 who assumed the isotopic and mass balances in the particulate fraction, the result being 523 in the commonly used formula equation for nitrogenous nutrient uptake. Dugdale and 524 Wilkerson (1986) modified their rate equations further and highlighted the importance 525 of short-term incubations. Although short-term incubation was requested, Collos 526 (1987) demonstrated that anthe formula equation based on the concentration of 527 particles at the end of the experiment, rather than at the beginning, wasis more reliable 528 when more than one N source wasare simultaneously incorporated by the 529 phytoplankton-population. That is, the equation by Collos (1987) corrected for the bias 530 caused by use of unlabeled multiple N utilizationsources.

531Different fromUnlikethe above_-mentioned equations, Blackburn (1979) and532Caperon et al. (1979) proposed ^{15}N isotope dilution models based on the substrate533rather than the product. By measuring the isotope values and concentrations of the534substrate, (e.g. NH4⁺), and then both NH4⁺ consumptions (DON and/or PN as product)535and regeneration rates can be obtained. Glibert et al. (1982) further modified the

536 isotope dilution method and calculated the uptake rate into the PN fraction by 537 substituting the exponential average of r_{NH4+} at the beginning and at the end of an 538 incubation to correct for the isotope dilution existing in the model of Dugdale and 539 Goering (1967). Besides Despite the methodological improvements, imbalance was 540 often observed between the substrate reduction and the increase in the particulate phase 541 in field studies. Laws (1985) introduced a new model that with considered ation of the 542 imbalance and calculated the "net uptake rate" (into PN). Later on, Bronk and Glibert 543 (1991) revised Law²s² model on the basis of the model proposed by Glibert et al. (1982) 544 to calculate the "gross uptake rate" (substrate incorporation into PON particulate 545 organic nitrogen plus DON). Overall speaking, Nnone of the above models considered 546 the mass balance at the whole system scale. Although rates were obtained via 547 analytical solutions, the bias potential due to multiple fluxesows was not completely 548 resolved.

549 To solve multiple co-occurring N processes address this problem, Elskens et al. 550 (2002) formulated a new model that takes into account multiple co-occurring N fluxes 551 in a natural system, The model containsing 3n_+1 equations and an equal number of 552 flux rates, where (n stands for is the number of labelled N substrates) and 3n+1 flux 553 rates, by taking multiple co-occurring N fluxes in natural system into account. 554 Approximate The rates in their model were resolved estimated using by a weighted 555 least squares technique. Additionally, Elskens et al. (2005) subsequently created a 556 process-oriented model (PROM) that accounteding for as many N processes as

557 needed to quantify how specific underlying assumptions affected the estimation 558 behavior of the estimates of all the above-mentioned models. The authors concluded 559 that uncertainties may increase as the incubation is prolongs prolonged and that 560 oversimplified models may risk bias when their underlying assumptions are violated. 561 The most recent attempt to resolve simultaneous N processes was conducted by Pfister et al. (2016), who applied-used parallel incubations (¹⁵N labelled NH₄⁺ and 562 563 NO₃⁻) in tidepools to measure multiple flows among benthic N, ammonium, nitrite, 564 and nitrate pools. In their experiment, six differential equations (with seven unknowns) 565 were constructed basing based on mass and isotope balances and solved by using the 566 ODE function of the R language. Since Because the N content of benthic algae were 567 was not measured due to sampling difficulty difficulties in sampling and spatial 568 heterogeneity in-of biomass, a mass balance at the whole system scale mass balance 569 cannot could not be reachedachieved; . Specifically, thus, the flux rate of DON release 570 cannot could not be be obtained determined.

571 Compared with <u>the</u> methods or models mentioned above, the advantages of <u>the</u> 572 isotope matrix method include (1) the potential biases caused by multi<u>ple</u>-flows were 573 <u>taken into consideration considered undersubject to</u> the <u>circumstance of constraint that</u> 574 <u>there be a mass balance at the system scalelevel</u>; (2) one tracer addition <u>was sufficient</u> 575 <u>to quantify for</u>-multiple *in situ* flows: (parallel incubations, i.e., <u>dark-light</u> and <u>light</u> 576 <u>dark or ¹⁵NH₄⁺ and ¹⁵NO_x⁻, were not needed</u>; (3) <u>simple-post-hoc data processing was</u> 577 <u>simple, and a unique solution can be obtained via matrix derivationinversion</u>; (4) no 578 extra laboratory work <u>is was demanded necessary</u> (see below).

579 **4.1.2 Rate comparisons**

580 Following In accord with Pfister et al. (2016), we estimated all N transformation 581 rates via-using ordinary differential equations (ODEs) for the three cases on the 582 assumption that there is no r_{NH4+}remineralization for comparison was constant (see 583 Table 1–3). In general, the rates values obtained by the matrix inversion and integration 584 of the ODEs were consistent. The rate dDifferences, if anywhen apparent, was were 585 caused by the duration for of the integration., The isotope matrix method was applied to only i.e., shorter time (the first two time points (i.e., time intervals of either 2 or 4 586 587 hours), whereas the ODEs were integrated for calculation) for isotope matrix method 588 and longer time (4 or 5 points for the entire 24-h incubation) for ODE. In Pfister et al. 589 (2016), ODEs were used to analyze data collected at 3 monitoring time points within 590 a 5--h time intervalours were implemented for ODE. Unfortunately, such intensive 591 sampling for on-deck incubations is not practical; <u>howeverHowever</u>, we still 592 strongly recommend the short-term incubations for water column study-studiesas 593 previously suggested. With proper duration, tTwo time points separated by a few hours 594 for integration may be more convenient and realistic for instantaneous rate 595 measureestimates.

Below, we present a comparison with between our results and conventional source-product rate measurements (Collos, 1987) of ammonium oxidation and uptake 598 (Table 3). The matrix-derived NH_4^+ uptake rates for all of the experimental 599 experiments cases were consistent with those the rates (difference < 8%) from the 600 traditional source-product method when the final PN concentration was applied 601 forused in the calculation. The fact that the deviations were larger (13–21%) when the initial PN was applied used, which was supported is consistent by with the conclusions 602 603 of previous studies that estimates involving the final PN concentration are more 604 reliable. Obviously, The deviation could obviously be higher when if the 605 phytoplankton growth rate was higher.

606 On the other handIn contrast, the end-products of ammonium oxidation or 607 nitrification are consumed by phytoplankton continuously, particularly in the euphotic 608 layer zonefull of photosynthetic autotrophs. In many cases, nitrate uptake has been 609 shown to occurs in both the light and dark conditions (e.g. Dugdale and Goering, 1967; 610 Lipschultz, 2002; Mulholland and Lomas, 2008). The significant consumption of 611 end-products (NO_x^- and NO_2^-) -apparently violates the underlying assumption of that 612 underlies the source-product rate calculation. Therefore, the NH_4^+ 613 oxidation/nitrification rate could cannot not be obtained determined with a 614 source-product model,. Although such as all cases in our study since pphytoplankton 615 consumption resulted in a net reduction of NO_x^- in all of our experiments, we were 616 nevertheless able to determine NH₄⁺ oxidation/nitrification rates with the isotope 617 matrix method (Figs. 2b, 3b, and 3c) (see Table 3).

618	In most casesprevious studies, the final isotopic composition rather thanbut not
619	<u>the</u> final concentration of NO_x^- was <u>has been</u> measured; <u>as such As a result</u> ,
620	researchers may not <u>have been</u> aware <u>that the outflow</u> of the greater $^{15}NO_x^-$ outflow
621	was greater than the inflow. For During dark incubations, researchers may also
622	assume insignificant NO_x^- consumption. However, the <u>a</u> "net decrease in end-product"
623	is almost unavoidable when an incubation is conducted under simulated in situ light
624	conditions for to estimate ammonium oxidation. To overcome address this
625	consumption effect induced by the first-order reaction, Santoro et al. (2010, 2013) took
626	NO_x^- removal into account and formulated a new equation, a function of that took
627	<u>account of the nitrification rate (F) and NO_x^- uptake rate (k). Following In accord with</u>
628	Santoro et al. (2010), we calculated the nitrification rate for the low-nutrient case (via a
629	nonlinear least-squares curve-fitting routine in Matlab by using the first three time
630	points of the 15 N-NO _x ⁻ / 14 NNO _x ⁻ measurements <u>.</u>) to be <u>The calculated rate</u> , -0.05 nmol
631	L^{-1} h ⁻¹ (Table 3), which was (~30%) lower than the matrix-derived rate of (0.07 nmol
632	L^{-1} h ⁻¹). By-In contrast, their nitrate uptake rate constant (k = 0.010 h ⁻¹) was only
633	one-sixth of that the rate constant (0.059 h ⁻¹) derived from the matrix method, although
634	a comparable nitrification rate was obtained when the consumption term was taken into
635	account.
636	Surprisingly, when we introduced the values of F and k two parameters by

Surprisingly, when we introduced the <u>values of F and k two parameters by</u>
usingdetermined with the method of Santoro et al. (2010) into STELLA to generate the
time courses of <u>parameters variables</u>, we found <u>that the simulations simulated values</u> of

639 $\delta^{15}NO_x^{-}$ and r_{NOx-} agreed well with that those determined of by the isotope matrix 640 method (Figs. 4j and 4n)., yetHowever, much slower decreasing trends were found for $^{15}NO_x^{-}$, $^{14}NO_x^{-}$, and NO_x^{-} (Figs. 4 b, f and r). Finally, we realized that the formula 641 642 equation produced proposed by Santoro et al. (2010) is-was constrained only by the 643 ratio-changes of the ratios rather than by the changes of the individual concentrations changes in of ${}^{15}NO_x^{-}$ and ${}^{14}NO_x^{-}$. Thus, the nonlinear curve-fitting method by Matlab 644 645 may only provide a correct simulation for only of the ratio change of the ratio. This 646 conclusion implies that the nitrate uptake rate derived from the non-linear curve-fitting 647 method in Matlab should be validated also by the final concentration of nitrate, as was 648 done by Santoro et al. (2013).

In summary, (1) an accurate measurements of concentrations <u>during a</u> time series is vital for all kinds of transformation rate estimates, including the isotope matrix method and (2) the isotope matrix method <u>can overcame overcome</u> various biases that impact estimates made with traditional methods-<u>might encounter</u>.

653 **4.2 Implications for nitrogen biogeochemical processes**

654 <u>Through-Results of use of the isotope matrix method suggest several conclusions</u>
655 <u>with respect to</u>, biogeochemical <u>implications were obtained from various</u>
656 <u>aspectsprocesses</u>.

657 **4.2.1 Remineralization, regeneration, and community succession**

658 The matrix solution fit wellwas consistent with the model runs with variable 659 $r-NH_4^+$ in at time points of no more than 12 hearly stage, the implication being 660 suggesting that dilution effects was were negligible during the early incubation period, at least in our casestudies. Dilution effects could be significant when remineralization 661 662 is intensive and the incubation prolongslonger. Pfister et al. (2016) found that 663 macrofauna (mussels) play an important role in remineralization. While The fact that 664 zooplankton in the water column of our sampled cases waswater samples were not 665 abundant, it might be a reason for the low remineralization rates in our short-term 666 incubations.

667 In the WNP low--nutrient case, after 24 houran incubation of 24 hours, the levels 668 of nitrate and ammonium approached the concentration threshold for phytoplankton 669 utilization (e.g., <30-40 nM NH₄⁺ for *Emiliania huxleyi*; Sunda and Ransom, 2007). In 670 Figure 4, the STELLA projection fitted agreed well with the PN parameters 671 concentrations only for only the first 12 hours. In this case, in fact, we have actually 672 observed phytoplankton succession. Our flow cytometry data (shown in authors reply 673 for Reviewer #2)(Fig. S3)-demonstrated that the cell-number of living eukaryotes 674 eukaryotic cells (4 times higher than Synechococcus) increased in the first 24 hours and 675 started to drop rapidly after 24 hours. On the contraryIn contrast, the growth of 676 Synechococcus continued for throughout 94after 24 hours, even though under 677 constantly low nitrogen nutrient situation concentrations dropped to constantly low 678 level. Such These observations phenomenon suggested that the phytoplankton

679 community was competed competing for nitrogen, source and a major community shift 680 started at around 24 hours. After the time point of at 12 hours, the observed parameters associated with concentrations of ¹⁴N and ¹⁵N in the PN was-were higher than the those 681 682 projection projected by STELLA. The most intriguing phenomenon among PNassociated parameters was the additional ¹⁵N, which should could not have come from 683 684 $^{15}NH_4^+$, in PN. The most likely source of nitrogen source candidate with enriched ^{15}N 685 to support Synechococcus growth was the nitrogen released from dead eukaryotes, which contained freshly consumed ¹⁵N tracer, rather than the ambient DON. More 686 studies are needed to explore nutrient thresholds for different phytoplankton species. 687 688 Nevertheless, our results suggested that short-term-incubations is crucialmust last no 689 more than a few hours for nitrogen uptake studies in the oligotrophic ocean.

690 **4.2.2** Evaluate Evaluation of the contribution of nitrification to new 691 production

692 Nitrification in the sunlit euphotic zone of the ocean drew not muchlittle attention 693 until recent decades years after molecular evidence led to the discovery of the 694 widespread occurrence of ammonia oxidizing archaea (AOA) discovery in the 695 perspective of molecular evidence (Francis et al., 2005; Santoro et al., 2010, 2013; 696 Smith et al., 2014) and rate measurements based on isotope-isotopic studies (Ward, 697 2011; Santoro et al., 2010; Grundle et al., 2013; Smith et al., 2014). As mentioned in Introduction the Introduction, the conventional "new" production has may have been 698 699 overestimated 19-33% on a global scale due to the "regenerated" nitrate regenerated in

700 the euphotic zone via nitrification process. However, a more realistic evaluation 701 assessment of for the fractional contribution of nitrification to NO₃⁻ uptake can only be 702 achieved when the incubations is are conducted in the same bottle under in situ light 703 conditions instead of parallel incubations in the dark and light. The isotope matrix 704 method is so far the most convenient and suitable method for evaluating the relative 705 importance of co-occurring nitrification and new production in the euphotic oceanzone. 706 In all our experimental casesstudies, the contributions of nitrification to new 707 production were < 1% (Table 4). The This relatively low contribution was probably due 708 to the light inhibition on of nitrifiers for in the WNP case and because of the low water 709 temperature in the sampling season.

710 Nevertheless, light effects in our case-studies is-were significant. Light 711 suppresses nitrification (Ward, 2005; Merbt et al., 2012; Peng et al., 2016). The NH₄⁺ 712 oxidation rate in at 80% sPAR was reduced by 36% relative to that in the rate at 2% 713 sPAR. These Results results agreed are consistent with current knowledge, although 714 some recent evidences has showed shown that some taxa of marine AOA hold-have 715 the genetic expansion capabilities capability to reduce oxidative stress and to repair ultraviolet 716 damage (Luo et al., 2014; Santoro et al., 2015). More study case studiess are needed 717 in the future to explore the vertical distributions of the relative contribution of 718 nitrification to new production in the euphotic zone.

719 **4.2.3 Nutrient preference**

720 Phytoplankton takes differentuse a variety of nitrogenous species as nutrients for 721 growth. McCarthy et al. (1977) introduced a the concept of a relative preference index 722 (PRIRPI) to assess the relative utilization use of a specific different N-forms of N, and 723 when an RPI value >1 indicates a preference for the specific substrate over the other N 724 forms of N. As shown in Table 4, in the NO_3 prevailed low nutrient case NO_3 was preferred, the The fact that the $\frac{PRI}{RPI}$ for (NO_3^-) was very close but slightly higher to 725 726 **PRI**than the RPI for (NH₄⁺), which was probably due to the phytoplankton community 727 structure, as mentioned above. This result was in line is consistent with the result studies in the Sargasso Sea (Fawcett et al., 2011). While-However, in the high-nutrient case 728 NH_4^+ bay, the order of the RPI values was $PRI(NH_4^+) > 1 > PRI(NO_3^-) > PRI(NO_2^-)$, 729 730 the suggesting suggestion being that phytoplankton preferred NH_4^+ over NO_3^- and 731 NO₂⁻, similar to the the results of studies in Chesapeake Bay (McCarthy et al., 1977).

732 **4.2.4 Quantifying various ammonium consumption pathways**

733 In the upper ocean, NH4⁺ cycles rapidly due to the metabolic pathways of the 734 various microorganisms's that metabolic pathways competeing for ammonium. 735 Ammonium may serve as a nitrogen source for phytoplankton assimilation, and as an 736 energy source for ammonia-oxidizing organisms (AOM). Moreover, many studies have shown that bacteria also play a part in NH4⁺ utilization (Middelburg and 737 738 Nieuwenhuize, 2000; Veuger et al., 2004). Our results in the low--nutrient case showed 739 that phytoplankton was-were the main consumers of NH_4^+ consumer (82% of the total 740 NH4⁺ consumption), bacteria Bacteria accounted for another ~17%, and AOM 741 <u>utilized used</u> the <u>rest-remaining</u> 1%. While iIn the <u>high-nutrient studyeutrophic WYW</u> 742 <u>bay</u>, phytoplankton and bacteria each consumed ~50% of the total NH_4^+ (Table 4).

743 **5. Conclusions**

744 Theis isotope matrix method was designed specifically for euphotic water column 745 incubations in the euphotic zone under simulated in situ conditions. By considering 746 multiple -flows among pools on the assumption of and requiring mass balance at the 747 whole--system level, we minimized potential biases caused by non-targeted processes 748 in traditional source-product methods. Given the progress in analytical techniques for 749 measuring concentrations and isotopic compositions of nitrogen species, the isotope 750 matrix method presents is a promising avenue approach for the studying of rates of 751 nitrogen processes fluxes with from a system-wide perspective. Furthermore, the 752 matrix method is also appropriate for probing the effects of environmental factors (e.g., 753 CO₂, pH, temperature, and light intensity) on the interactive N processes in one-a 754 single incubation bottle.

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Table 1. The isotope matrix results for (A) the specific rates and (B) average rates of N processes in the low-nutrient case during the first interval under different r_{NH4+} uariation conditions. And all N transformation rates via ODE following Pfister et al. (2016) on the assumption of no remineralization were estimated for comparison. Note r_{NH4+} variation was manipulated artificially by decreasing r_{NH4+} values at a constant reduction rate and the total reduction of r_{NH4+} was 0%, 1%, 10%, 20% and 50% of the full time span (24 h) of incubation.



	The percentage of r_{NH4+} decrease in 24 h								
Rate constant (k) h^{-1}	0		1%	10%	20%	50%			
	ODE		Ι						
NH4 ⁺ uptake (k1)	0.040	0.038	0.038	0.038	0.038	0.039			
Remineralization (k2)	0	0	0.00001	0.0001	0.0002	0.001			
$\mathrm{NH}_{4^{+}}$ oxidation (k3)	0.0004	0.0005	0.0005	0.0005	0.0005	0.0005			
NO _x ⁻ uptake (k4)	0.060	0.059	0.059	0.059	0.059	0.059			
DON release (k5)	0.017	0.024	0.024	0.024	0.024	0.024			
Bacteria uptake NH ₄ ⁺ (k6)	0.005	0.007	0.008	0.011	0.015	0.028			

1017 **(B)**

	The percentage of r_{NH4+} decrease in 24 h							
Rate $(k \times C)$ nmol L ⁻¹ h ⁻¹ -	()	1%	10%	20%	50%		
	ODE Isotope Matrix							
NH4 ⁺ uptake (F1)	3.8	4.9	4.9	4.9	5.0	5.1		
Remineralization (F2)	0.0	0.0	0.1	0.6	1.2	3.0		
$\mathrm{NH_{4}^{+}}$ oxidation (F3)	0.04	0.07	0.1	0.1	0.1	0.7		
NO _x ⁻ uptake (F4)	19.3	27.2	27.2	27.2	27.2	27.2		
DON release (F5)	9.6	11.5	11.5	11.6	11.6	11.8		
Bacteria uptake NH4 ⁺ (F6)	0.5	1.0	1.0	1.5	2.0	3.7		

Table 2. The isotope matrix results for the rates of N processes in the high-nutrient case at the depth of (A) 80% sPAR and (B) 2% sPAR under different r_{NH4+} variation conditions. And all N transformation rates via ODE following Pfister et al. (2016) on the assumption of no remineralization were estimated for comparison. Note: r_{NH4+} variation was manipulated artificially by decreasing r_{NH4+} values at a constant reduction rate and the total reduction of r_{NH4+} was 0%, 1%, 10%, 20% and 50% of the full time span (15 h) of incubation.

	The percentage of r _{NH4+} decrease in 15 h						
Rate $(k^* \overline{C})$	0		1%	10%	20%	50%	
nmol $L^{-1} h^{-1}$	ODE			Isotope Ma	atrix		
NH ₄ ⁺ uptake (F1)	360	397	397	399	401	408	
Remineralization (F2)	0	0	21	211	424	1043	
NO ₂ ⁻ uptake (F3)	27	29	29	29	29	29	
NH ₄ ⁺ oxidation (F4)	1.1	0.4	0.4	0.4	0.4	0.4	
NO ₃ ⁻ uptake (F5)	190	149	149	149	149	149	
NO ₂ ⁻ oxidation (F6)	1.7	0	0	0	0	0	
DON release (F7)	0	0	0	0	0	0	
Bacteria uptake NH4 ⁺ (F8)	268	282	303	490	701	1314	

1026 **(B)**

	The percentage of r_{NH4+} decrease in 15 h							
Rate $(k^* \overline{C})$	0		1%	10%	20%	50%		
nmol $L^{-1} h^{-1}$	ODE							
NH4 ⁺ uptake (F1)	228	208	208	209	211	216		
Remineralization (F2)	0	0	18.1	179	361	895		
NO_2^- uptake (F3)	7.3	3.1	3.1	3.1	3.1	3.1		
NH ₄ ⁺ oxidation (F4)	1.1	0.7	0.7	0.7	0.7	0.7		
NO ₃ ⁻ uptake (F5)	106	72	72	72	72	72		
NO ₂ ⁻ oxidation (F6)	2.0	0	0	0	0	0		
DON release (F7)	0	0	0	0	0	0		
Bacteria uptake NH4 ⁺ (F8)	202	265	283	442	623	1152		

- 1027 **Table 3.** Comparison of the NH_4^+/NO_x^- uptake and NH_4^+ oxidation/nitrification rates
- 1028 derived from different methods.

Process	Case	Depth (m)	Isotope Matrix method (this study)	Rates based on Ref A*	Traditional method Ref B*	Rates followed Ref C*		
			$(nmol L^{-1} h^{-1})$					
NH4 ⁺ uptake	Low nutrient	25	4.9	3.8	4.6			
Nitrification	Low nutrient	25	0.07	0.04	_	0.05		
NOx ⁻ uptake	Low nutrient	25	27.2	19.3		4.6		
NH4 ⁺ uptake	High -80%sPAR	0.2	397	360	387			
$\mathrm{NH_{4}^{+}}$ oxidation	High -80%sPAR	0.2	0.4	1	—			
NH4 ⁺ uptake	High -2% sPAR	2.3	208	228	192			
NH ₄ ⁺ oxidation	High -2% sPAR	2.3	0.7	1	_			

1029 Ref A* stands of rates calculation by ODE followed Pfister et al. (2016)

1030 Ref B* stands of rates calculation followed Collos (1987)

1031 Ref C* stands of rates calculation followed Santoro et al. (2010)

- 1033 Table 4. The contribution of nitrification derived NO_x^- to NO_x^- uptake (%), N
- 1034 preference index, and the proportion of NH_{4^+} consumption by phytoplankton, bacteria

Case Depth (m)	nitrification	RPI	RPI	RPI	A/TNH_4^+	B/TNH_4^+	$*C/TNH_4^+$	
	to NO ₃ ⁻	for	for	for	consumption	consumption	consumption	
	(111)	uptake (%)	\mathbf{NH}_{4}^{+}	NO_2^-	NO_3^-	(%)	(%)	(%)
Low nutrient	25	0.3	0.9		1.0	82.1	16.8	1.2
High -80%sPAR	0.2	0.3	1.6	0.6	0.5	58.4	41.5	0.1
High -2% sPAR	2.3	0.9	1.8	0.1	0.5	43.9	56.0	0.1

1035 and nitrifier to total NH_4^+ consumption in low and high nutrient cases.

- 1036 *A, *B, *C stands for NH4⁺ utilized by phytoplankton, bacteria and nitrifier,
- 1037 respectively. TNH_4^+ consumption stands for total NH_4^+ consumption.

1039 Figure Captions

1040 Fig. 1. Model schemes with the most fundamental nitrogen transformation processes in

1041 low- (a) and high- (b) nutrient aquatic environments. Arrows stand for the transfer

1042 flux/rate from the reactant to product pool. The structure and inter-exchanges in the

- 1043 high-nutrient case (Fig. 1b) are the same as in (a), except that NO_x^- is divided into NO_2^-
- 1044 and NO_3^{-} .

1045 **Fig. 2.** The observational data in the low-nutrient case for (a) $[NH_4^+]$, (b) $[NO_x^-]$, (c) 1046 [PN], (d) [TDN], (e) $\delta^{15}N$ -NO_x⁻, (f) $\delta^{15}N$ -PN. The regular and inverse open triangles 1047 stand for the paralled samples and the analytical errors are shown.

Fig. 3. The observational data in the high-nutrient case for (a) $[NH_4^+]$, (b) $[NO_2^-]$, (c) $[NO_3^-]$, (d) [PN], (e) [TDN], (f) [PN+TDN], (g) $\delta^{15}N-NO_2^-$, (h) $\delta^{15}N-NO_3^-$ and (i) $\delta^{15}N-PN$. The light and dark red diamonds stand for the paralled samples in 80% sPAR case and the black regular and inverse open triangles stand for the paralled samples in 2% sPAR case. The analytical errors are shown in figures.

1053 Fig. 4. The observed and STELLA-derived values in the low-nutrient case for (a) $[^{15}NH_4^+]$, (b) $[^{15}NO_x^-]$, (c) $[^{15}N-PN]$, (d) $[^{15}N-DON]$, (e) $[^{14}NH_4^+]$, (f) $[^{14}NO_x^-]$, (g) 1054 $[^{14}$ N-PN], (h) $[^{14}$ N-DON], (i) r_{NH4+} , (j) r_{NOx-} , (k) r_{PN} , (l) r_{DON} , (m) δ^{15} N-NH4⁺, (n) 1055 δ^{15} N-NO_x⁻, (o) δ^{15} N-PN, (p) δ^{15} N-DON, (q) [NH₄⁺], (r) [NO_x⁻], (s) [PN] and (t) 1056 1057 [DON]. The black regular and inverse open triangles represent the paralleled observed 1058 values; the black, green, blue, magenta and pink solid lines stand for the STELLA 1059 model simulations when r_{NH4+} decreases 0%, 1%, 10%, 20% and 50% in 24 h, 1060 respectively. The dashed lines in (b), (f), (j), (n) and (r) were generated from nonlinear 1061 least-squares curve-fitting by Matlab following Santoro et al. (2010).

1062 Fig. 5. The observed and STELLA-derived values in the high-nutrient case of (A) 80%

1063 sPAR depth and (B) 2% sPAR depth for (a) $[^{15}NH_4^+]$, (b) $[^{15}NO_2^-]$, (c) $[^{15}NO_3^-]$, (d)

1064 [15 N-PN], (e) [15 N-DON], (f) [14 NH₄⁺], (g) [14 NO₂⁻], (h) [14 NO₃⁻], (i) [14 N-PN], (j)

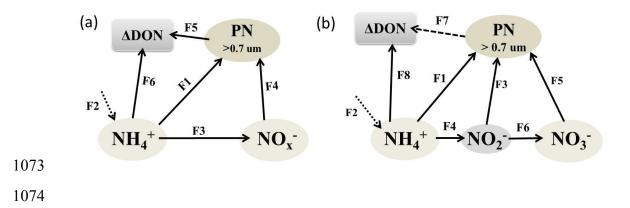
1065 [¹⁴N-DON], (k) r_{NH4+} , (l) r_{NO2-} , (m) r_{NO3-} , (n) r_{PN} , (o) r_{DON} , (p) $\delta^{15}N-NH_4^+$, (q) 1066 $\delta^{15}N-NO_2^-$, (r) $\delta^{15}N-NO_3^-$, (s) $\delta^{15}N-PN$, (t) $\delta^{15}N-DON$, (u) [NH₄⁺], (v) [NO₂⁻], (w) 1067 [NO₃⁻] (x) [PN] and (y) [DON]. The black regular and inverse open triangles

1068 represent the duplicate observational values; the black, green, blue, magenta and pink

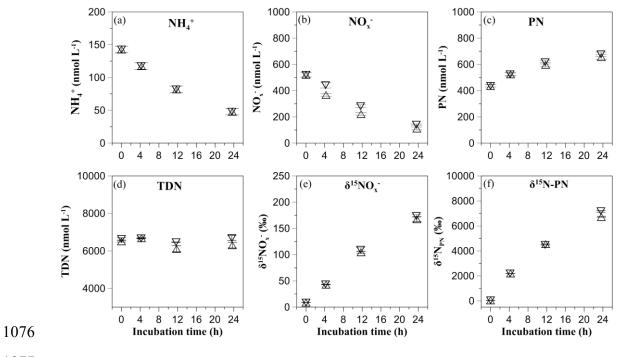
1069 solid lines represent the STELLA model simulations of r_{NH4+} decreases 0%, 1%, 10%,

1070 20% and 50% in 15 h, respectively.

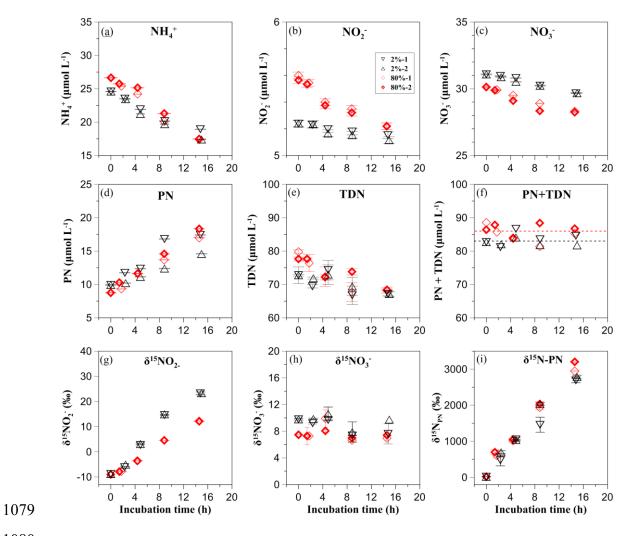
1072 **Fig. 1**





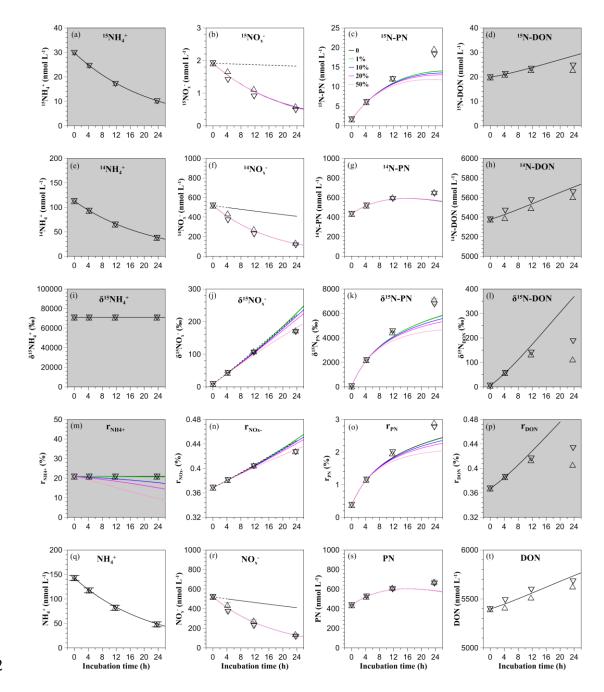






1080

1081 Fig. 4



1082

