

Revised manuscript has been examined by one of the original reviewers. reviewer still finds two important points that need to be addressed:

- 1) unknown and variable magnitude of inherent non-conservativeness, which can compromise both estimates of internal sources/sinks and budgets.
- 2) nutrient budgets must be re-evaluated and presented in a realistic way, including appropriate revision of Discussion with comparison to other studies.

REPLY: Thanks for your highlighting two points from reviewer. We have addressed all the comments from the Dr. Savchuk Oleg. Please find our revised manuscript.

Further reviewer's comments to revised manuscript by Y. Liu, H.E.M. Meier and K. Eilola
"Nutrient transport in the Baltic Sea - results from a 30-year physical-biogeochemical reanalysis"
submitted to "Biogeosciences"

The manuscript has been substantially improved in many aspects, from clarification of procedures and algorithms to its language. However, besides of possible further stylistic cosmetics, there is still a couple of important obscurities left, which need to be either clarified or entirely removed in order to increase the paper's credibility because retaining clearly questionable issues reduces confidence in more plausible (reliable) results.

1. General comments and suggestions

1.1. As the major objectives of the data assimilation, the revised Introduction lists: a) the reconstruction of the water quality with high resolution, especially in under-sampled periods and areas, b) the estimation of nutrient transports from more realistic nutrient fields, c) the implementation of reconstructed fields as initial and boundary conditions. In addition to these, I would also explicitly stress the capability of dynamical estimating of the water nutrient pools, especially their long-term developments, as indicators of the trophic state (see 2.2 below).

REPLY: We add in the introduction: "A good reanalysis of biogeochemical state variables can dynamically describe indicators of eutrophication such as the long-term development of water nutrient pools."

We add a line to section 6.3 after "The results of the reanalysis can be used to estimate the water quality and ecological state with high spatial and temporal resolution in regions and during periods when no measurements are available."

"This supports improved assessments e.g. of eutrophication status indicators."

We added to section 5.6 (see also reply to 2.2 below): "The 3D nutrient pools constructed by data assimilation methods offer an opportunity to evaluate with improved estimates the changes in Baltic Sea eutrophication. ..."

1.2. However, I still miss the explicit indication of limitations and applicability of this approach already in Introduction and even, perhaps, in Abstract. In your response to me you indicated: "In the introduction section we further clarify the already listed limitations of data assimilation with respect to estimating nutrient budgets and we rewrite the objectives of this study." However, the first mention: "As a reanalysis can never be dynamical consistent and does not preserve mass, momentum and energy..." (Lines 100 – 101) of non-conservativeness seems appeared here out of nowhere. Consider, please,

a) re-formulation as "As a reanalysis does not CONSERVE mass, momentum and energy and, (var. - therefore, consequently, thus) can never be dynamically consistent and ..." and

b) moving to- (or making similar statement) somewhere between lines 68-72.

REPLY: We have changed the corresponding text and move them to the lines 68-72.

1.3. Most important among such limitations is an inherent non-conservativeness of the approach. As you correctly admitted in several places of your responses: “The potential impact from artificial sources or sinks due to data assimilation is of course also included in the reanalysis results. Because of the unknown impact from this “process” it is better to avoid detailed discussions ...especially about the changes in the nutrient pools... about the net sources and sinks.” First of all, why “potential” when it is a real artificial change of simulated masses of nutrients (see also lines 423-425)? Unfortunately, you have also neglected to estimate its possible magnitude as a difference between 3D pools estimated before and after assimilation, e.g. simultaneously with RMSD (lines 218-219). I understand that it might be too laborious to systematically calculate such differences over the entire integration period but then it could be made as a few cases, perhaps, at the very first “act of assimilation” and then a couple of times more, just to show a magnitude. Such estimates are important because I can hardly see a justification of your statement at lines 427-429 in Fig. 3, where RMSDs do not show really significant differences between the beginning and the rest of REAN experiment.

REPLY: We agree that artificial sources or sinks affect the reanalysis results.

The figure below illustrate the magnitude of RMSD difference before and after data assimilation in REANA. The changed magnitude of RMSD by data assimilation is small, which support our text in the discussion --”data assimilation has corrected the mismatch between model state and observation to an “optimal” level during an initial adjustment period. After the adjustment period, the mismatch between model and observation becomes small and the successive adjustment due to data assimilation also becomes small.”

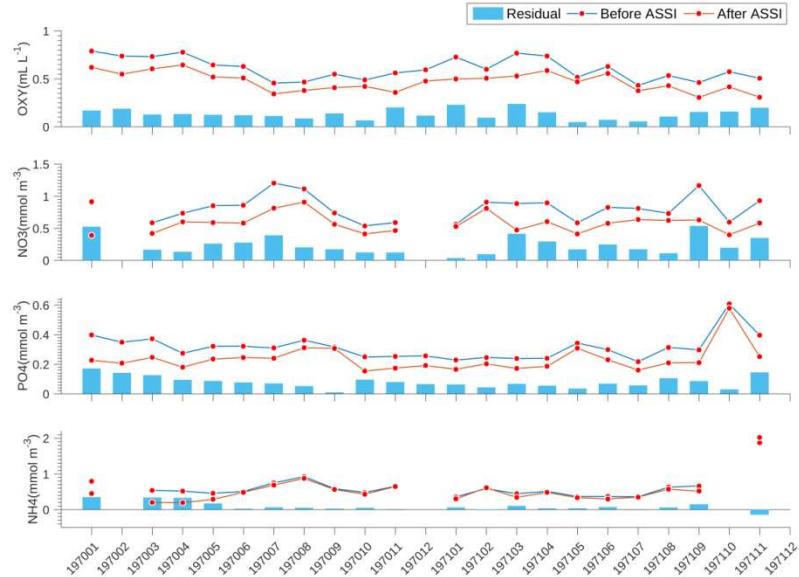


Figure 1. The monthly averaged RMSD before and after the data assimilation and their residual (After-Before) in REANA during the period 1970-1971.

1.4. Furthermore, I now think that instead of “avoiding the discussion” about erroneous sources and sinks, their usage and analyses should be excluded entirely. Consider, please, for instance, your estimates of internal sources and sinks as a difference between integrated inflows and outflows of nutrients into every grid cell, where water transports is calculated as a product of the concentrations and the current velocity vectors in the considered and surrounding cells (Eq. 5). Without data assimilation these concentrations are determined by the explicitly parametrized pelagic and sediment biogeochemical processes and, thus, the resulted differences can justifiably be interpreted as sources and sinks of the model variables. In the case of data assimilation, these concentrations are artificially altered by unknown values, unknowingly variable both in time and space. These dynamical errors are then uncontrollably propagated over the simulation domain. As you correctly noted at lines 458-460: “...usually small artificial sources and sinks from the data assimilation are becoming as important as physically motivated sources and sinks when sums of fluxes are compared.” Therefore, your estimates are not strictly comparable to those in Eilola et al. (2012) and lead to errors if being implemented in budgeting. As you say: “In this study, we focus mainly on nutrient transports derived from the reanalysis” (Line 207) and I advise to stay strictly with that.

REPLY: We agree with your comments and hope that the text of the revised manuscript explains the focus of our study better. However, we keep the revised Figures 10 and 11 in the revised manuscript because we believe that the comparison of transports between sub-basins between our and previous studies is still very interesting. As the calculation of transports across sections follows conservation principles our estimates of nutrient exchange between sub-basins are reliable for the considered time period. The revised calculation of the integrated sources/sinks within the sub-basins includes of course the artificial sources/sinks caused by the data assimilation method and makes an interpretation difficult. However, as our numbers of sources/sinks are similar to those of previous studies we conclude that our overall budgets are reliable.

1.4. Nutrient budgets.

1.4.1. Definitions and algorithms. It is still not clear enough, how the “burial” terms in Figs. 9-10 were calculated. Terminologically, is it a proxy of real sediment burial (i.e. permanent removal from the model domain) or a sediment retention/accumulation, especially at a long-term (1970-1999) budgeting scale? Or even simpler – is it just integral sink or nutrient removal, accounting also for denitrification? Is it actually a difference between nitrogen fixation and denitrification that is accounted here for? Do you consider these “burial” values as “the sediment budgets” (line 99)? If it is the integration of sources/sinks described in Section 5.5 then the critique from 1.3 above apply. If “The sediment sinks (burial) are calculated from the difference between the net deposition of nutrients to the sediments and the release of nutrients from the sediments” (lines 227-228) then it should be better explained a) what does “net deposition” mean – was it explicitly calculated as the difference between total sedimentation of plankton+detritus+ resuspended sediments minus uplifted sediments and b) how the sediment release of nutrients was computed, why it is not presented and analysed, which could be more relevant and interesting than the entire Section 5.5? Please, consider carefully and explain more precisely.

REPLY: we recalculated the nitrogen and phosphorus sink in the Figs10-11 in revised manuscript. We describe this in section 4: “The sinks of the nutrient budgets are calculated from the supplies from land/atmosphere, import/export from other basins and the changes in pelagic nutrient pools during the period (sink=supply+import-export-pool change).”

We removed all discussions about tendencies.

1.4.2. There are also internally contradicting statements and unfulfilled intentions:

Lines 100-102 – “As a reanalysis ... does not preserve mass... the calculated budgets are compared to ... other studies ... as consistency check”. OS: – if only within an order-of-magnitude, because of different approaches and compared periods.

REPLY: We rephrased in section 6.3.” Further, nutrient transports across selected cross-sections or between vertical layers are calculated from the reanalysis with high resolution and improved accuracy. However, one cannot expect that budgets calculated from the summation of internal fluxes from model results with data assimilation are more accurate because usually small artificial sources and sinks from the data assimilation are becoming as important as physically motivated sources and sinks when sums of fluxes are compared. Hence, we calculated (section 5.6) budgets only from inputs and exports and changes in the water pools of nutrients with the aim to compare the reanalysis results with other studies using only observations. It is perhaps not possible to claim that our budgets are more accurate than budgets that are derived from observations only, despite the higher temporal and spatial resolution in model outputs. However, the advantage of the reanalysis is that measurements are extrapolated in space and time based upon physical principles of the model.”

Lines 459-464 – “...one cannot expect that budgets calculated ... with data assimilation are more accurate... Hence, we calculated budgets with the aim ... to estimate the magnitude of artificial sources and sinks by comparing our results with other studies...” OS: Such estimate was not made.

REPLY: See above.

1.4.3. Most likely, the above inconsistencies are the reason why the presented budgets as a whole are highly unrealistic. The budget's terms are presented as average for 30 years including “tendencies” from Table 1. Then the starting and final total stocks (TS) can formally be calculated as follows: $TS_{init} = TS_{average} - dTS * 15$ and $TS_{final} = TS_{average} + dTS * 15$. For the Gulf of Finland such estimates give in kT: $P_{init} = 29.9 - 3.7 * 15 = -25.6$, $P_{final} = 29.9 + 3.7 * 15 = 85.4$, $N_{init} = 127 - 16 * 15 = -113$, $N_{final} = 127 + 16 * 15 + 367$; for the Kattegat: $P_{init} = -29.3$, $P_{final} = 51.7$, $N_{init} = -378$, $N_{final} = 522$, etc.

At the same time, both the total pools and estimated nutrient exports and imports across the basin boundaries looks realistic enough to serve as another solid source of information for other studies and conclusions, if not being compromised by the doubts in “burial” terms estimated with artificial effects of assimilation.

REPLY: we re-estimated the nitrogen and phosphorus sink of the Baltic (please refer to the reply 1.4.1).

1.5. Algorithm of the data assimilation. Now it looks more comprehensive and save the reader from looking into similar papers. Unfortunately, it still too formal and, to my mind, is not narratively explanatory enough for fellow marine biologists and biogeochemists. Perhaps, that is why it leaves (or arises) some additional questions, such as, for instance, about construction of BEC from simulation for 1962-1968, when erroneous effects of the poorly prescribed sediment (especially, P) nutrients should be even larger.

REPLY: The analysis quality of an ensemble data assimilation method is related to the BEC constructed by samples ensemble anomalies. We do not know the “truth” BEC. Typically we use a “good” sample ensemble for constructing BEC. These samples of EnOI are usually obtained from a historical simulation without data assimilation. The ensemble samples are not integrated in the simulation process. As a result, EnOI is a 3D data assimilation method and the model errors are

assumed not to be integrated in simulation process. Actually the analysis increment is a combination of sample ensemble anomalies. In general the variance magnitude of the perturbation of the sample ensemble from historical simulation is not close to the one at the analysis time. Therefore EnOI use a inflate factor (alpha in Equtaion 3, $0 < \alpha \leq 1.0$) to adjust the ensemble samples perturbation's variance. Our reanalysis simulation stated from 1970. We would like to construct the samples from the model snapshots before starting our reanalysis. Further, there are also many observations available and the model results have been validated in the period 1964-1968. Therefore, in this study, we select the samples from free run during the period 1964-1968. In order to keep the "independence" of sample, the snapshots during the period 1964-1968 have been stored every three days. The assimilation scheme in this study is same as that used in Liu et al. (2014). Other sample ensembles could be tested but this is beyond the scope of this study.

When observations become available at the assimilation time, the 'optimal' state variables are produced by Equtaion (1). These 'optimal' state variables are used as 'new' initial condition for the 'next' simulation cycle.

An overall flow diagram below explains the method qualitatively neglected all details.

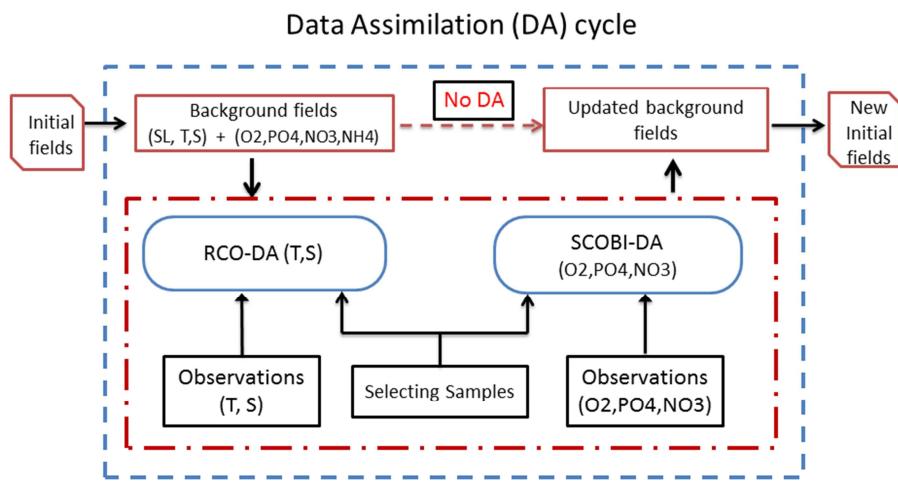


Figure 3. Work flow diagram of the "weakly" coupled data assimilation.

2. Specific comments and suggestions.

2.1. Consider, please, re-estimation of nutrient budgets using only external nutrient inputs and nutrient flows across basin's boundaries. The sinks/sources can then be found as the rest terms assuming either steady state or accounting also for the differences between initial and final pools (see 2.2) below. Then, the analysis of budgets should be revised accordingly both in Results and Discussion.

REPLY: we have recalculated the sink/sources according to your advices in the revised manuscript.

2.2. The potential of the integrated nutrient pools estimation from more realistic (?) 3D nutrient fields reconstructed by data assimilation into 3D coupled physical-biogeochemical could be exploited much more for description of eutrophication. For that as well as for the budgeting, it is not even necessary to calculate annual dynamics but just, say, 5 (10?) year averages in the beginning and end of simulation.

REPLY: We added to section 5.6: "The 3D nutrient pools constructed by data assimilation methods

offer an opportunity to evaluate with improved estimates the changes in Baltic Sea eutrophication. As a example, an investigation of the trophic state from changes in five year average nutrient pools in REANA show that the total nitrogen pool in the Baltic proper increased from 657 kton to 1045 kton from the period 1971-1975 to 1995-1999 while the total phosphorus pool decreased from 469 kton to 448 kton between the same periods. Hence, nitrogen increased by about 59% while phosphorus decreased by about 4%. Similarly the pool of DIN in REANA increased by 80% while DIP decreased by 6%. The corresponding numbers obtained from BED showed an increase of 100% for DIN and an increase of 7% of DIP. The results indicate large increases of nitrogen pools in the Baltic proper during the investigated period but only relatively small changes of phosphorus pools.

Subsequent periodic assessments can be used to reveal future eutrophication changes. While estimating the trophic state, it should be noticed that the change in trophic state depends on the chosen time periods. For example, from the year 1971 to 1999 the total phosphorus (TP) in the Baltic proper increased with 2.7 kton yr^{-1} . This differs from the decrease seen from the five year average change discussed above. The reason is the impact from short term fluctuations of nutrient content (Figure 4) that may be larger than the long-term changes.”

2.3. In addition to running RMSDs, a comparison of such average pools between REAN and FREE simulations (perhaps, also with BED) as the measure of improvement could be interesting too,

REPLY: We added to section 5.1: “The annual averaged pelagic pool of simulated DIN and DIP were also compared to the corresponding pools estimated from the BED data in Baltic proper. The maximal annual differences of DIN and DIP compared to the BED, have been reduced by 57.5% and 72.3%, respectively, from 400 kton and 650 kton in FREE (not shown) to 170 kton and 180 kton in REANA. The remaining differences between REANA and BED may be explained by the methods of integration that differ. BED estimates are based on a limited amount of observations while the model results are based on a large number of grid points with dynamically varying state variables.”

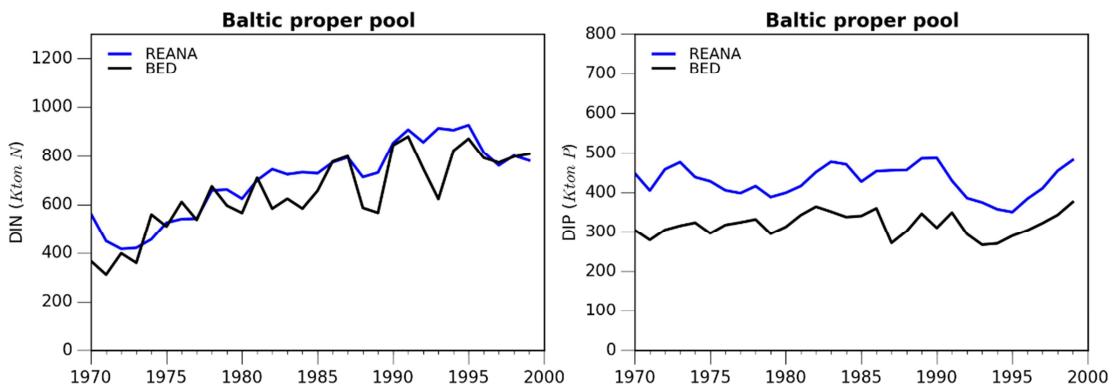


Figure 4. Annual mean integrated pools (in kton) of pelagic dissolved inorganic nitrogen (DIN) and phosphorus (DIP) in the Baltic proper calculated from REANA and from observations with DAS.

2.4. End of Section 3. A few sentences are needed here to explain what does indication “in BED” in Sections 5.2-3 mean and how the long-term average March fields in 5.3 are reconstructed.

REPLY: We have added the text in the Section 3. “The simulated spatial variations of the late winter surface layer nutrient concentrations are compared with the spatial variations reconstructed from BED with the Data Assimilation System (DAS) by Sokolov et al. (1997). Due to insufficient historical data coverage the average March fields were computed for time period (1995–2005) from over 3600 oceanographic stations found in BED. Also nutrient pools of dissolved inorganic nitrogen (DIN) and phosphorus (DIP) calculated with DAS (see Savchuk, 2010) are compared with the

results of this study. See Eilola et al. (2011) for more details about the data handling by DAS.”

And in Section 5.3 we added: “The average March concentrations (1970-1999) of DIP and DIN in the upper layers (0–10m), as well as their ratio (DIN:DIP), were calculated (Fig. 6). Due to insufficient historical data the corresponding BED maps describe the averages for the period 1995–2005 as a basis for model to data comparison.”

2.5. Lines 220-224. As follows from Eq. 5, the product of velocity vector “ V_k ” and concentration “ C_k ”, both taken at the location “ k ”, represents just transports (or outflowing nutrient mass) at this location and not “net transports”, which word better to keep for real differences between integral inflows and outflows, like you use for derivation of (semi-artificial) internal sources and sinks in Section 5.5. as well as for more plausible really net transports between basins and along cross-sections in Section 5.6. Please, check carefully the usage of “net transports” and correct accordingly, for instance, at line 295.

REPLY: Throughout the entire manuscript we changed the definitions. We differentiate now between “vertically integrated transports” and “net transports”. The latter is the difference between export and import of volume or nutrients integrated across the sections in sub-basins.

2.6. Line 273-274 – an improved model description of vertical transports of nutrients in the layers above the halocline.” Since Fig. 4 says nothing about water movements, explain, please, the reason(s) for a statement about improvement of transports – is it due to realistic vertical gradients or something else? Where is this claimed analysis of vertical transports?

REPLY: We have improved the modeling stratification which affects the vertical transports of nutrients, especially the water depths around the thermocline. The vertical distributions of nutrients have been improved above the halocline (Figure 5). For clarify, we changed the text to “an improved model description of vertical DISTRIBUTIONS of nutrients in the layers above the halocline”

2.7. Consider, please, making a separate Section starting from line 372

REPLY: we have done it in revised manuscript.

3. Minor things, technical corrections and language cosmetics

REPLY: Thanks for your comments. We have changed these texts in revised version of manuscript. Title – I feel, “Nutrient transportS...” would sound better

Line 97 - “This information can’t be obtained from neither observations alone or...” The grammar construction should be either “...CAN NOT ... EITHER... OR...” or “CAN... NEITHER... NOR...”, I would recommend the former

Line 100 - “...dynamically consistent.” - it is adverb

Line 101 – replace, please, “preserve” by “conserve” because we are dealing here with the laws of conservation of mass and energy...

Section 5.1 – Please, restore reference to Fig. 3

Line 408 – “...nutrients are not detailed enough known.” Please, rephrase less clumsy: “are poorly

known”, “are insufficiently known”, “are known not detailed enough”, are known not in enough detail”

Line 752 – Ref. to Terruzzi et al (2014) should be to pp. 200-217

Line 817 –“...Northwestern Gotland Basin, Bothnian Sea...” – Gulf of Finland is missing in this listing.

1 Nutrient transports in the Baltic Sea - results from a 30-year 2 physical-biogeochemical reanalysis

3
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8 **Abstract.** Long-term oxygen and nutrient transports in the Baltic Sea are reconstructed using the Swedish
9 Coastal and Ocean Biogeochemical model (SCOBI) coupled to the Rossby Centre Ocean model (RCO). Two
10 simulations with and without data assimilation covering the period 1970–1999 are carried out. Here, the “weakly
11 coupled” scheme with the Ensemble Optimal Interpolation (EnOI) method is adopted to assimilate observed
12 profiles in the reanalysis system. The reanalysis shows considerable improvement in the simulation of both
13 oxygen and nutrient concentrations relative to the free run. Further, the results suggest that the assimilation of
14 biogeochemical observations has a significant effect on the simulation of the oxygen dependent dynamics of
15 biogeochemical cycles. From the reanalysis, nutrient transports between sub-basins, between the coastal zone and
16 the open sea, and across latitudinal and longitudinal cross sections, are calculated. Further, [the spatial
17 distributions of regions with](#) nutrient import or export are examined. Our results emphasize the important role of
18 the Baltic proper for the entire Baltic Sea, with large net [transport \(export minus import\)](#) of nutrients [from the
19 Baltic proper](#) into the surrounding sub-basins (except the [net phosphorus import from](#) the Gulf of Riga and the [net
20 nitrogen import from](#) the Gulf of Riga and Danish Straits). In agreement with previous studies, we found that the
21 Bothnian Sea imports large amounts of phosphorus from the Baltic proper that are [retained](#) in this sub-basin. For
22 the calculation of sub-basin budgets, it is crucial where the lateral borders of the sub-basins are located, because
23 net transports may change sign with the location of the border. Although the overall transport patterns resemble
24 the results of previous studies, our calculated estimates differ in detail considerably.

25 **Keywords:** reanalysis; data assimilation; numerical modelling; Baltic Sea; biogeochemical transports; nutrient
26 budgets

27 **1 Introduction**

28 The water exchange between the Baltic Sea and the North Sea is restricted by the narrows and sills in the Danish
29 transition zone (Fig. 1). The hydrography of the Baltic Sea also depends on freshwater from rivers, which causes
30 large salinity gradients between the surface layer and the saltier bottom layer, and between the northern sub-
31 basins and the entrance area (e.g. Meier and Kauker, 2003). The low-saline outflowing surface water is separated
32 from high-saline inflowing bottom water by a transition layer, the halocline. The bottom water in the deep sub-
33 basins is ventilated mainly by so-called Major Baltic Inflows (MBIs) (Matthäus and Franck, 1992; Fischer and
34 Matthäus, 1996). MBIs can significantly affect biogeochemical processes in the deep basins because of the
35 inflow of large volumes of saline and oxygen-rich water into the Baltic Sea (e.g. Conley et al. 2009; Savchuk,
36 2010). In the Baltic Sea, the density stratification and long water residence time hamper the ventilation of deep
37 waters. As a result, oxygen deficiency is a common feature. Additionally, nutrient loads from agriculture and
38 other human activities of the large population in the catchment area increased nutrient concentrations in the water
39 column. Actually, eutrophication has become a large environmental problem in the Baltic Sea in recent decades
40 (e.g. Boesch et al., 2008; Pawlak et al., 2009; Wulff et al., 2001; Andersen et al., 2015). Therefore, accurate
41 estimates of the ecological state and nutrient and water exchange between sub-basins and between the coastal
42 zone and the open sea are of particular importance in managing the marine environment system.

43

44 On one hand, the estimation of biogeochemical processes, ecological state and nutrient exchange may rely on
45 coupled marine ecosystem-circulation models (e.g. Neumann et al., 2002; Eilola et al., 2009; 2011; Almroth-
46 Rosell et al., 2011; 2015; Maar et al., 2011; Daewel and Schrum, 2013). However, addressing biogeochemical
47 cycles is a challenging task due to the complexity of the system. Obviously, there are large uncertainties in
48 marine ecological simulations (e.g. Eilola et al., 2011). In contrast to the modelling of the physics of the

49 atmosphere or ocean, where a basic description of the motion is provided by conservation equations, there is no
50 basic set of equations that describe the marine ecosystem. Many biogeochemical processes are still poorly known
51 and their uncertainties are difficult to quantify accurately. These potential sources of errors limit the applicability
52 of the models both in forecasting and reanalysis. Further, imperfect initial conditions and model forcing also
53 cause biases in the simulation results.

54
55 On the other hand, estimating nutrient budgets and transports between sub-basins may directly rely on
56 observations and basin integrated budget models (Savchuk, 2005). The estimation accuracy depends on the
57 spatial and temporal coverage of the measurements and the locations of borders between sub-basins. Although
58 the data coverage in the Baltic Sea has gradually increased over time, the lack of observations still makes it
59 difficult to estimate reliable biogeochemical cycles. Today, the availability of satellite sensor data like ocean
60 color data from the OCTS (Ocean Color and Temperature Sensor) and from the SeaWiFS (Sea-Viewing Wide
61 Field-of-View Sensor) has provided the best spatial coverage of measurements. However, these sensors only give
62 an estimate of a few biogeochemical parameters at the surface of the marine ecosystem, and not the state of the
63 entire marine ecosystem in the water column. Continuous observations of the deep ocean are only possible with
64 in situ sensors, which have been deployed at only a limited number of stations (Claustre et al., 2010).

65
66 Given the coverage of observations and model deficiencies, we decided to perform a reanalysis based upon a
67 high-resolution, coupled physical-biogeochemical model to estimate the physical and biogeochemical state of the
68 Baltic Sea. For this purpose, data assimilation continuously updates the model variables at the locations of the
69 observations and in their neighborhood. Integration in time of the prognostic model equations allows the spread
70 of the information from the observations within the model domain.

71
72 The assimilation of data into coupled physical-biogeochemical models is confronted by various theoretical
73 and practical challenges. For example, the response of the three-dimensional biogeochemical model to external
74 forcing caused by the physical model is highly non-linear. Further, it is difficult to use the biological
75 observational information to reduce biases in the simulation of ocean physics which has an impact on modeled
76 biogeochemistry (Beal et al., 2010). Besides, data assimilation as used in this study does not conserve mass,
77 momentum and energy. Therefore, a reanalysis with data assimilation can never be dynamically fully consistent.

78 Nevertheless, the use of data assimilation complementing ecosystem modeling efforts has gained widespread
79 attention (e.g. Hoteit et al., 2003; Allen et al., 2003; Natvik and Evensen, 2003; Hoteit et al., 2005; Triantafyllou
80 et al., 2007; While et al., 2012; Triantafyllou et al., 2013; [Teruzzi et al., 2014](#)). [Data assimilation into ecosystem](#)
81 [models has focused both on parameter optimization and on state and flux estimations \(Gregg et al, 2009\)](#). A
82 comprehensive review of biological data assimilation experiments can be found in Gregg et al. (2009).

83
84 In the Baltic Sea, the biogeochemical data assimilation has started to become a research focus. For example,
85 Liu et al. (2014) used the Ensemble Optimal Interpolation (EnOI) method to improve the multi-annual, high-
86 resolution modelling of biogeochemical dynamics in the Baltic Sea. Fu (2016) analyzed the response of a coupled
87 physical-biogeochemical model to the improved hydrodynamics in the Baltic Sea. Recently, several data
88 assimilation studies have focused on the historical reanalysis of salinity and temperature in the Baltic Sea (e.g. Fu
89 et al., 2012; Liu et al., 2013; 2014). Reanalysis has helped enormously in making the historical record of
90 observed ocean parameters more homogeneous and useful for many purposes. For instance, ocean reanalysis data
91 have been applied in research on ocean climate variability as well as on the variability of [biogeochemistry](#) and
92 ecosystems (e.g. Bengtsson et al., 2004; Carton et al., 2005; Friedrichs et al., 2006). Ocean reanalysis can also be
93 used for the validation of a wide range of model results (e.g. Fontana et al., 2013). For instance, the ocean mean
94 state and circulation can be calculated from reanalysis results to evaluate regional climate ocean models (e.g.
95 Meier et al., 2012). Moreover, reanalysis in the ocean is beneficial to the identification and correction of
96 deficiencies in the observational records, as well as filling the gaps in observations. Regional and local model
97 studies may use reanalysis results as initial and boundary conditions. [A good reanalysis of biogeochemical state](#)
98 [variables can dynamically describe indicators of eutrophication such as the long-term development of water](#)
99 [nutrient pools.](#)

100
101 The present paper focuses on the assimilation of profiles of temperature, salinity, nutrients and oxygen in the
102 Baltic Sea following Liu et al. (2014). We aim to reproducing the ocean biogeochemical state with the help of
103 information from both observations and a coupled physical-biogeochemical model for the period 1970-1999.
104 Since 1970 the data coverage in the Baltic Sea is satisfactory. The results of the reanalysis are supposed to be
105 used to estimate the water quality and ecological state with high spatial and temporal resolution in regions and

107 during periods when no measurements are available. Further, nutrient transports across selected cross-sections or
108 between vertical layers are calculated with high resolution and accuracy taking the complete dynamics of
109 primitive equation models into account. This information can~~not~~ be obtained from either observations alone or
110 from model results without data assimilation because the latter might have large biases in both space and time.
111 We assess the nutrient budgets of the water column as well as of the nutrient exchanges between sub-basins and
112 between the coastal zone and the open sea. The calculated budgets are compared to the results of other studies to
113 evaluate our results. Hereby, we follow studies of other regions applying data assimilation for a biogeochemical
114 reanalysis on long-term scale (Fontana et al., 2013; Ciavetta et al., 2016).

115

116 This paper is organized as follows. The physical and biogeochemical models are described in Section 2. Then
117 the observational data set and the method of the reanalysis are introduced in Section 3 and 4, respectively. The
118 experiment results, including comparisons with observations, are presented in Section 5. Finally, in Section 6 and
119 7, discussion and conclusions finalize the paper.

120 2 Models

121 The RCO (Rossby Centre Ocean) model is a Bryan–Cox–Semtner primitive equation circulation model with a
122 free surface (Killworth et al., 1991). Its open boundary conditions are implemented in the northern Kattegat,
123 based on prescribed sea level elevation at the lateral boundary (Stevens, 1991). An Orlanski radiation condition
124 (Orlanski, 1976) is used to address the case of outflow, and the temperature and salinity variables are nudged
125 toward climatologically annual mean profiles to deal with inflows (Meier et al., 2003). A Hibler-type dynamic–
126 thermodynamic sea ice model (Hibler, 1979) with elastic–viscous–plastic rheology (Hunke and Dukowicz, 1997)
127 and a two-equation turbulence closure scheme of the $k-\varepsilon$ type with flux boundary conditions (Meier, 2001) have
128 been embedded into RCO. The deep-water mixing is assumed inversely proportional to the Brunt–Väisälä
129 frequency, with the proportionality factor based on dissipation measurements in the Eastern Gotland Basin (Lass
130 et al., 2003). In its present version, RCO is used with a horizontal resolution of 2 nautical miles (3.7 km) and 83
131 vertical levels, with layer thicknesses of 3 m. RCO allows direct communication between bottom boxes of the
132 step-like topography (Beckmann and Döscher, 1997). A flux-corrected, monotonicity-preserving transport (FCT)

133 scheme is applied in RCO (Gerdes et al., 1991). RCO has no explicit horizontal diffusion. For further details of
134 the model setup, the reader is referred to Meier et al. (2003) and Meier (2007).

135

136 The biogeochemical model called SCOBI (Swedish Coastal and Ocean Biogeochemical model) has been
137 developed to study the biogeochemical nutrient cycling in the Baltic Sea (Marmefelt et al., 1999; Eilola et al.,
138 2009; Almroth-Rosell et al., 2011; 2015). This model handles biological and ecological processes in the sea as
139 well as sediment nutrient dynamics. SCOBI is coupled to RCO (e.g. Eilola et al., 2012; 2013; 2014). With the
140 help of a simplified wave model, resuspension of organic matter is calculated from the wave and current-induced
141 shear stresses (Almroth-Rosell et al., 2011). SCOBI has a constant carbon (C) to chlorophyll (Chl) ratio C:Chl =
142 50 (mg C (mg Chl)⁻¹), and the production of phytoplankton assimilates carbon (C), nitrogen (N) and phosphorus
143 (P) according to the Redfield molar ratio (C:N:P = 106:16:1) (Eilola et al., 2009). The molar ratio of a complete
144 oxidation of the remineralized nutrients is O₂:C = 138. For further details of the SCOBI model, the reader is
145 referred to Eilola et al. (2009, 2011) and Almroth-Rosell et al. (2011).

146

147 RCO-SCOBI is forced by atmospheric forcing data calculated from regionalized ERA-40 data using the
148 regional Rossby Centre Atmosphere (RCA) model (Samuelsson et al., 2011). The horizontal resolution of RCA is
149 25 km. A bias correction method following Meier et al. (2011) is applied to the wind speed. Monthly mean river
150 runoff observations (Bergström and Carlsson, 1994) are used for the hydrological forcing. Monthly nutrient loads
151 are calculated from historical data (Savchuk et al., 2012).

152 **3 Observations**

153 The Baltic coastal shelf observation systems have been largely improved by the joint efforts of the countries
154 surrounding the Baltic Sea. For example, the International Council for the Exploration of the Sea (ICES)
155 (<http://www.ices.dk>) and the Swedish Oceanographic Data Centre (SHARK) (<http://sharkweb.smhi.se>) are
156 collecting the observations with the aim to monitor the Baltic Sea. Furthermore, the Baltic Sea Operational
157 Oceanographic System (BOOS) (<http://www.boos.org/>) is providing near real-time observations and the publicly
158 available database BED (Baltic Environmental Database, <http://nest.su.se/bed>) of the Baltic Nest Institute (BNI)
159 (<http://www.balticnest.org>) store physical and environmental data from BNI partner institutes (see

160 | http://nest.su.se/bed/hydro_chem.shtml). [Also the data of SHARK have been stored in BED.](#) As a result, a
161 | comprehensive data set is collected for the Baltic Sea region. The assimilated observations in this study comprise
162 | both physical (temperature and salinity) and biogeochemical variables (oxygen, nitrate, phosphate and
163 | ammonium) from the SHARK database. Before assimilation, the data were quality controlled. These controls
164 | include checks of location and duplication, and examination of differences between forecasts and observations. A
165 | profile was eliminated from the assimilation procedure when the station was located on land defined by the RCO
166 | bathymetry. We also removed observations when the difference between model forecasting field and
167 | observations exceeds the given standard maximum deviation (for example 4.0 mL L^{-1} for oxygen concentration).
168 | We used an average of the observations in the same layer when there was more than one observation per layer.
169 | These observations cover almost the whole Baltic Sea including Kattegat and the Danish Straits. Figure 2 shows
170 | the number of biogeochemical observation profiles in different sub-basins, and the temporal distribution of these
171 | biogeochemical observations. The number of observations is inhomogeneous in both temporal and spatial
172 | distribution over the period from 1970 to 1999. There are relatively more observations in the Baltic proper than in
173 | other sub-basins. In the Gulf of Riga, a minimum number of observation profiles (30 for oxygen, 30 for
174 | phosphate, 28 for nitrate and 28 for ammonium) is found. Obviously, the number of observations during the
175 | period of 1988-1994 is higher than that during other periods. Further, there are generally less observations from
176 | 1981-1983 than during other periods. The maximum number of observation profiles occurred in 1991 for oxygen
177 | (1,844), phosphate (1,728) and nitrate (1,758). However, the number of ammonium observation profiles has a
178 | maximum value of 1,222 in 1992. Moreover, [compared to other variables](#) the numbers of oxygen and ammonium
179 | observations [are](#) largest and smallest, respectively. These observations from SHARK and BED are used to
180 | validate the model and assimilation results.

181 |
182 | [The simulated spatial variations of the late winter surface layer nutrient concentrations are compared with the](#)
183 | [spatial variations reconstructed from BED with the Data Assimilation System \(DAS\) by Sokolov et al. \(1997\).](#)
184 | [Due to insufficient historical data coverage the average March fields were computed for time period \(1995–2005\)](#)
185 | [from over 3600 oceanographic stations found in BED. Also nutrient pools of dissolved inorganic nitrogen \(DIN\)](#)
186 | [and phosphorus \(DIP\) calculated with DAS \(see Savchuk, 2010\) are compared with the results of this study. See](#)
187 | [Eilola et al. \(2011\) for more details about the data handling by DAS.](#)

189 Here we briefly describe the configuration of the data assimilation system of this study. We focus on the state
 190 estimation via EnOI. The distribution of stochastic errors are assumed to be Gaussian and non-biased. EnOI
 191 estimates an 'optimal' oceanic state at a given time using observations, the numerical model and assumptions on
 192 their respective bias distribution. The relationship between them can be expressed as following:

$$193 \quad \psi^a = \psi^f + \mathbf{K}(d - H\psi^f) \quad (1),$$

$$194 \quad \mathbf{K} = \mathbf{P}^f H^T (H\mathbf{P}^f H^T + (N-1)\mathbf{R})^{-1} \quad (2).$$

195 Where $d \in \mathfrak{R}^m$ is the vector of observations with m being the number of observations. $\psi \in \mathfrak{R}^n$ is the n
 196 dimensional model state vector which includes the sea level anomaly, temperature, salinity, oxygen, phosphate,
 197 ammonium and nitrate. The superscripts a and f refer to “analysis” and “forecast”, respectively. $\mathbf{K} \in \mathfrak{R}^{n \times m}$ is
 198 the Kalman gain matrix and H is an operator that maps the model state onto the observation spaceoften H is
 199 linear interpolation. $d - H\psi^f \in \mathfrak{R}^m$ is the innovation which is calculated in the observation space. $\mathbf{R} \in \mathfrak{R}^{m \times m}$ is
 200 the observation error covariance. N is the number of the ensemble samples. EnOI computes the Background
 201 Error Covariance (BEC) matrix $\mathbf{P} \in \mathfrak{R}^{n \times n}$, which determines how to spread out information from observations in
 202 space and between variables, by the ensemble perturbation matrix $\mathbf{A}' = \mathbf{A} - \bar{\mathbf{A}}$ as follows:

$$203 \quad \mathbf{P} = \frac{\alpha}{N-1} \mathbf{A}' (\mathbf{A}')^T \quad (3).$$

204 Here $\mathbf{A} = (\psi_1, \psi_2, \dots, \psi_N) \in \mathfrak{R}^{n \times N}$ is the sample ensemble and $\bar{\mathbf{A}} = \frac{1}{N} \sum_{i=1}^N \psi_i$ is the sample ensemble mean. The
 205 subscript T denotes the transpose of a matrix and the scaling factor $\alpha \in (0, 1]$ is introduced to tune the variance
 206 of the sample ensemble perturbations to a realistic level in order to capture the variability of model parameters
 207 like temperature and dissolved oxygen, which is dominated by misplacement of mesoscale features and which
 208 varies in location and intensity seasonally. Therefore, we hypothesize that the background errors are proportional
 209 to the model variability on intra-seasonal time scales. We selected the samples from model results of a hindcast

210 [simulation without data assimilation from](#) one and a half month before and after the calendar date of the
211 assimilation time [during](#) the period 1964–1968 (Liu et al., 2013). [The snapshots during the period 1964-1968](#)
212 [have been stored every three days.](#) From every year during the selected period 1964–1968 20 snapshots have
213 been selected. [Hence, a total of \$N = 100\$ model samples are adopted to obtain a quasi-stationary BEC matrix.](#)
214 [The analysis by EnOI rely on the sample ensemble because the analysis increment is a linear combination of](#)
215 [sample ensemble anomalies. Other, more “sophisticated” sample ensembles could be tested but this is beyond the](#)
216 [scope of this study.](#) An adaptive scaling factor was calculated to adapt to the instantaneous forecast error variance
217 before each local analysis (Liu et al., 2013; 2014). Further, localization is used to remove unrealistic long-range
218 correlation with a quasi-Gaussian function and a uniform horizontal correlation scale of 70 km. As a result, the
219 quality of fields obtained by data assimilation is determined by the coverage and quality of observations (She et
220 al., 2007). Moreover, the assimilation frequency or window is another factor to affect the assimilation fields.
221 They are directly related to how many observations are entering the assimilation cycling and how often the model
222 initial condition is adjusted by data assimilation (Liu et al., 2013). Here, we select an assimilation window of
223 three days and the assimilation frequency is once every seven days in the reanalysis experiment. It means that all
224 the observations in three days before and after the assimilation time are selected to yield the “new” initial
225 condition for the following simulation during the current assimilation cycle. [When observations become available](#)
226 [at a certain time, the 'optimal' state variables are calculated by Equation 1, which are used as new initial](#)
227 [conditions for the next simulation cycle.](#)

228
229 Based on the above configuration, two experiments from January 1970 to December 1999 have been carried
230 out. One experiment is a simulation without data assimilation (FREE). The other simulation is constrained by
231 observations using the “weakly coupled” assimilation scheme based upon the EnOI method following Liu et al.
232 (2014) which was briefly described [above](#) (REANA). Both simulations, FREE and REANA, are initialized for
233 January 1970. The initial conditions are taken from an earlier run with RCO-SCOBI. The observation error in
234 REANA is defined according to Liu et al. (2014). However, in Liu et al. (2014), only a shorter assimilation
235 experiment for a 10-year period is presented, and so far the reliability of the assimilation scheme in multi-decadal
236 simulations has not been shown. Following Liu et al. (2014), our REANA experiment assimilated both physical
237 and biogeochemical observations. In this study, we focus mainly on nutrient transports derived from the
238 reanalysis.

239

240 To assess the results with (REANA) and without (FREE) data assimilation, the overall monthly mean RMSDs
 241 (root mean square differences) of oxygen, nitrate, phosphate and ammonium were calculated relative to
 242 observations during the whole integration period. The overall monthly mean RMSD is calculated by the
 243 following formula:

244

$$RMSD = \frac{1}{N_j} \sum_{j=1}^{N_j} \sqrt{\frac{1}{N_t} \sum_{i=1}^{N_t} (\varepsilon_t^i)^2} \quad (4),$$

245 where N_t is the number of the observations at assimilation time t and N_j is the number of days of one month
 246 for one field for the entire Baltic Sea. $\varepsilon_t^i = x_{sim}^i(t) - x_{obs}^i(t)$ represents the difference between model result (x_{sim})
 247 and observation (x_{obs}) at time t and at the i^{th} observation location. We calculated ε_t only at the locations of the
 248 observations at the time t , which is calculated by mapping the model field onto the observation space. Here it
 249 should be noted that the RMSDs were calculated before the time of assimilation analysis, and the corresponding
 250 observations were not yet assimilated into RCO-SCOBI (Liu et al., 2014).

251

252 Based on the reanalyzed simulation, the annual mean DIN and DIP transports as well as DIP persistency are
 253 also calculated. These transports (VA_{Trans}) are vertically integrated from the sea floor to the sea surface at every
 254 horizontal position at every time step of the integration according to:

255

$$VA_{Trans} = \sum_{k=1}^N C_k u_k \Delta z_k \quad (5),$$

where $C_k, u_k, \Delta z_k$ and N are the field concentrations of DIN, DIP or organic phosphorus (OrgP), the current
 velocity vector in horizontal direction, vertical dimensions of a grid cell and the number of wet grid cells in the
 water column, respectively. From the net transport vector field both magnitude and streamlines are calculated.

256

257 The total nutrient budgets are calculated from the sum of inorganic and organic bioavailable nutrients. The
 258 combined nutrient supplies from land and from the atmosphere have been taken into account. Nitrogen fixation is
 259 not included in the external supplies. The sinks of the nutrient budgets are calculated from the supplies from
land/atmosphere, import/export from other basins and the changes in pelagic nutrient pools during the period

260 (sink=supply+import-export-pool change). By this definition the nitrogen sink includes nitrogen fixation and
261 denitrification. The nutrient flows for the total budgets are integrated along the selected borders of sub-basins
262 using Equation 5. Annual nutrient flows are averaged for the period 1970-1999. The total amount of nutrients for
263 every sub-basin is calculated from the integral of nutrient concentrations from phytoplankton, zooplankton,
264 detritus and dissolved nutrient times the volume of the sub-basin according to:

265
$$Total = \sum_{i=1}^{N_i} \sum_{j=1}^{N_j} \sum_{k=1}^{N_k} C_{i,j,k} \Delta x_{i,j} \Delta y_{i,j} \Delta z_k \quad (6),$$

where $C_{i,j,k}$, $\Delta x_{i,j}$, $\Delta y_{i,j}$ and Δz_k are the field concentrations (including nutrients from phytoplankton, zooplankton, detritus and dissolved nutrients), the horizontal and vertical dimensions of a grid cell, respectively. N_i , N_j and N_k are the number of grid cells in horizontal and vertical direction for every sub-basin, respectively.

Net transports across sections of sub-basins are calculated from the difference between export and import (relevant for the results in Sections 5.6 and 5.7).

266 5 Results

267 In the following sub-sections, we evaluate the impact of data assimilation on the long-term evolution of biases
268 (Section 5.1), and on vertical (Section 5.2) and horizontal (Section 5.3) distributions of nutrient concentrations.
269 For the evaluation of time series of simulated oxygen, nitrate, phosphate and ammonium concentrations, the
270 reader is referred to Liu et al. (2014, their Figs. 6 and 7). After the evaluation of the assimilation method, we
271 focus on the analysis of nutrient transports in the Baltic Sea based upon our reanalysis data that we consider to be
272 the best available data set for such an analysis. In particular, we analyze the horizontal circulation of nutrients
273 (Section 5.4), the horizontal distribution of nutrient sources and sinks, the nutrient exchange between the coastal
274 zone and the open sea (Section 5.5), and the nutrient budgets of sub-basins (Section 5.6).

275 5.1 Temporal evolution of biases and pools

276 The biases in both FREE and REANA have been calculated relative to observations for dissolved oxygen and
277 inorganic nutrients using Equation 4 (Fig. 3). Data assimilation has significantly reduced bias of the model

278 simulation. Generally, the RMSDs of oxygen and nutrient concentrations in REANA are smaller than that of
279 FREE. However, the degree of improvements differs among the variables. The RMSD of oxygen is mostly
280 greater and smaller than 1.0 mL L^{-1} for FREE and REANA, respectively. The mean RMSD of oxygen during this
281 period has been reduced by 59% (from 1.43 to 0.59 mL L^{-1}). Similar improvements appear in nitrate and
282 phosphate concentrations. The mean RMSDs of nitrate and phosphate in REANA were reduced by 46% (from
283 2.04 to 1.11 mmol m^{-3}) and 78% (from 1.05 to 0.23 mmol m^{-3}) relative to that in FREE, respectively.
284 Furthermore, the variability of RMSD of phosphate in FREE is large during the first 10 years, and decreases
285 afterwards with time. However, the data assimilation cannot always improve the model results. For instance,
286 although the mean RMSD of ammonium is reduced by 45% (from 1.15 to 0.63 mmol m^{-3}), the ammonium
287 concentrations in REANA become worse relative to those in FREE during some months. An example appears in
288 February 1975 when the RMSD of the ammonium concentrations in REANA (3.07 mmol m^{-3}) is greater than that
289 in FREE (2.75 mmol m^{-3}). These results are similar to the findings by Liu et al. (2014). However, here we show
290 that even the 30-year long assimilation is reliable, and that the RMSD of phosphate concentration decreases even
291 further with data assimilation continuing for more than 10 years.

292

293 Further, the annual averaged pelagic pools of simulated DIN and DIP in the Baltic proper are compared to the
294 corresponding pools estimated from the BED data in the Baltic proper (Fig. 4). The maximum annual differences
295 of DIN and DIP pools compared to BED, have been reduced by 57.5% and 72.3%, respectively, from 400 kton
296 and 650 kton in FREE (not shown) to 170 kton and 180 kton in REANA. The remaining differences between
297 REANA and BED may be explained by the methods of integration that differ. BED estimates are based on a
298 limited amount of observations while the model results are based on a large number of grid points with
299 dynamically varying state variables. A similar result for hypoxic area was found by Väli et al. (2013). They
300 showed with the help of model results which were sampled at the same times and locations of the observations
301 that the applied interpolation algorithm underestimated hypoxic area by about 40%.

302 **5.2 Mean seasonal cycle of nutrients**

303 The long-term average seasonal cycles of temperature and inorganic nutrients at monitoring station BY15 at
304 Gotland Deep (for the location, see Fig. 1) give a hint of how data assimilation improves simulated nutrient

305 dynamics in the Baltic proper (Fig. 5). The surface layer temperature and stratification show rapid increase in
306 April to May, with concurrent rapid decrease of nutrient concentrations due to primary production down to 50–60
307 m depths. The cooling and increased vertical mixing in autumn and winter reduce temperatures and bring
308 nutrients from the deeper layers into the surface layers. RCO-SCOBI captures these variations. However,
309 compared to BED, [FREE](#) has obvious biases, such as [overestimated temperature stratification around 30–50 m](#)
310 [depth](#) from late winter to early spring, higher concentration of nutrients at the 50–60m depth, stronger vertical
311 stratification of nutrient concentrations and less decrease of nutrients in the summer, especially below the
312 thermocline, as well as also in the surface layers for phosphate. One reason for the biases is the vertical
313 displacement of the halocline that is too shallow in RCO (e.g. Fig. 4 in Liu et al., 2014). The causes for the model
314 bias in nutrient depletion below the summer thermocline are not known, but possible reasons are discussed by
315 Eilola et al. (2011). These biases [are significantly reduced in the reanalysis which](#) provides an improved
316 description of vertical [distributions](#) of nutrients in the layers above the halocline.

317 5.3 Spatial variations of late winter nutrient concentrations

318 The average March concentrations [\(1970–1999\)](#) of DIP and DIN in the upper layers (0–10m), as well as their
319 ratio (DIN:DIP), were calculated (Fig. 6). [Due to insufficient historical data the corresponding BED maps](#)
320 [describe the averages for the period 1995–2005 as a basis for model to data comparison.](#) In BED [results](#) the
321 highest concentration of DIP occurs in the Gulf of Riga and the Gulf of Finland. Relatively high concentrations
322 of DIP are found in the entire Gotland Basin. The DIP concentrations in the Bothnian Sea and Bothnian Bay are
323 obviously lower than in other regions. Generally, the DIP in [FREE](#) has been largely overestimated in all regions
324 relative to BED, especially in the [Baltic proper](#). Further, in BED high concentrations of DIN occur in coastal
325 waters close to the river mouths of the major rivers in the southern Baltic proper. DIN concentrations in the Gulf
326 of Finland and in the Gulf of Riga are also high, and cover large areas of these gulfs. Unlike the BED data, the
327 DIN in [FREE](#) has high concentrations also in the entire southern and eastern coastal zones of the Baltic proper.
328 As a result, [FREE](#) shows a gradient in DIN concentrations between the coastal zone and the open sea in the entire
329 southern Baltic proper. The DIN and DIP patterns result in high and low DIN:DIP ratios in the Bothnian Bay and
330 Baltic proper, respectively. The highest DIN:DIP ratios are found in the Bothnian Bay in BED and in the Gulf of
331 Riga in [FREE](#). [FREE](#) has captured this large-scale pattern, but there are substantial regional differences. By the

332 constraints of the observation information, REANA has improved the spatial distributions of DIN and DIP
333 significantly [compared to BED](#). In particular, DIP concentrations in REANA are much closer to [the interpolated](#)
334 [BED](#) observations.

335

336 5.4 Mean horizontal circulation of nutrients

337 Nutrient transport directly affects the biogeochemical cycles and the eutrophication of the Baltic Sea. The
338 persistency of [vertically integrated](#) transports (Fig. 7) is defined, for instance, by Eilola et al. (2012). One should
339 note that the results by Eilola et al. (2012) are based upon 30-year averages for the control period 1978-2007 of a
340 downscaled climate scenario from a global circulation model. Similar calculations of transports will therefore be
341 briefly presented in the present study, since the hindcast period is better represented when the model is forced by
342 the assimilated atmospheric (ERA-40) and Baltic Sea data (REANA). DIP has the largest transports in the central
343 parts of the Baltic proper, with high persistency because the volume transports are generally larger in deeper
344 rather than in shallower areas. In the Bornholm Basin and the eastern parts of the central Baltic proper, cyclonic
345 circulation patterns are found. In the western parts of the central Baltic proper, southward transports prevail.
346 Relatively large magnitudes of transports of DIP are also found in the northwestern Gotland Basin, in the
347 southern Bornholm Basin, and through the Slupsk Channel connecting Bornholm Basin and Gotland Basin.
348 Similar transport patterns are also found for DIN, OrgP and OrgN (not shown). [Compared](#) to Eilola et al. (2012),
349 DIN, DIP, OrgP and OrgN transports and their persistency are [larger](#), although the overall patterns are similar.
350 For example, in Eilola et al. (2012, their Fig. 1), large DIN transports appear in the southern Baltic proper and the
351 Bornholm Basin. Similar differences are also found in both DIP and OrgP transports.

352 5.5 Internal nutrient sources and sinks

353 The horizontal distributions of areas with internal sources and sinks of phosphorus and nitrogen are illustrated in
354 Fig. 8. A net inflow (inflow \geq outflow) of nutrients to [each cell of the horizontal model grid](#) is defined as a sink
355 (import) and counted as positive, while net outflow (inflow \leq outflow) is defined as a source (export) and counted

356 as negative (Eilola et al., 2012). Source areas of DIP generally coincide with sink areas of OrgP, and vice versa.
357 This is also partly true for DIN and OrgN, but the sink for DIN has a large contribution from denitrification that
358 transfers DIN to dissolved N₂. The difference between phosphorus and nitrogen sources and sinks is oxygen
359 dependent, because the removal of N is enhanced at lower oxygen concentrations, while the sediment phosphorus
360 sink is weakened (e.g., Savchuk, 2010). Sediments may even temporarily become a source under anoxic
361 conditions, when older mineral-bound P can be released to the overlying water. Source areas of DIN are mainly
362 found in the Gulf of Riga, and the deeper parts of the Arkona Basin and Bornholm Basin. The largest DIP sources
363 occur in the eastern parts of the Gotland Basin as well as in the deepest parts of the Bornholm Basin and Arkona
364 Basin, whereas the largest sink of OrgP occurs in the central Baltic proper. The main sources of DIP are generally
365 found in regions where water depth is greater than 70 m (in other words below the permanent halocline in the
366 Baltic proper), while the main sources of OrgP (and OrgN) are found in the productive coastal areas shallower
367 than about 30–40 m (see also Fig. 9). Indeed, DIP export is largest in areas with a water depth between 70 and
368 100 m, and decreases towards greater water depths (Fig. 9).

369

370 According to the accumulated import of nutrients (Fig. 9), the magnitude of the DIP export is larger than that
371 of the DIP import. This indicates that not all of the supply of phosphorus from land and atmosphere is retained
372 within the Baltic proper [as will be further discussed from the nutrient budgets in section 5.6](#). For DIN, however,
373 we may notice only a very small net export from the Baltic proper to adjacent sub-basins, while for OrgP and
374 OrgN, imports and exports are almost balanced (Fig. 9). The nitrogen and phosphorus supply from land is
375 implemented in sea areas with a bottom depth usually of 6 m. This is where the river mouths are located in the
376 model.

377

378 There is a large import of DIP to areas with a depth range between 40–70 m (Fig. 9). This import does not
379 show a counter-part in the export of OrgP in Fig. 9. This result might be explained by local processes causing the
380 phytoplankton uptake and sediment deposition of DIP. There is an import of DIN to these areas that together with
381 nitrogen fixation and sediment–water fluxes of DIN may support local production of organic matter. The
382 phosphorus sink may be partly caused by oxygen dependent water–sediment fluxes that bind DIP to ironbound
383 phosphorus in oxic sediments (Almroth-Rosell et al., 2015). This effect is not included in Eilola et al. (2012), but
384 might potentially be accounted for by the adjusted DIP transports in REANA. The results of REANA indicate

385 that there is an additional sink but the relative importance of different processes causing this sink (data
386 assimilation or sediment processes) is, however, not possible to evaluate from the present reanalysis data set.
387

388 A partly opposite exchange profile is found for OrgP (Fig. 9). Coastal areas with a water depth of up to 40 m
389 are exporting organic phosphorus, whereas deeper areas import OrgP. Production in the coastal zone of the Baltic
390 proper and sedimentation in the open sea is almost balanced.
391

392 The largest export of DIN occurs due to rivers in the very shallow coastal zone. The magnitude of DIN imports
393 and exports in areas with greater water depths are much smaller. Obviously, DIN supplied from land is already
394 consumed in the coastal zone (Voss et al., 2005; Almroth-Rosell et al., 2011) and, consequently, only a minor
395 fraction of the nitrogen supplied to the shallow area can continuously reach regions deeper than 100 m (Eilola et
396 al., 2012; Radtke et al., 2012).

397 5.6 Nutrient budgets of sub-basins

398 The Baltic Sea is divided into seven sub-basins according to the selected sections, which form the borders of the
399 sub-basins (Fig. 1). We calculate total nutrient budgets for each of the sub-basins from the reanalysis (Figs. 10
400 and 11). Changes in pools are calculated as differences between 1971 and 1999 because the initial adjustment
401 process due to the assimilation is taking place during the first year (1970) (not shown). The largest annual mean
402 external phosphorus load occurs in the Baltic proper and amounts to 34.2 kton yr⁻¹ (Fig. 10). In addition, in the
403 Baltic proper the largest annual mean phosphorus sink of 25.0 kton yr⁻¹ is also found. Whereas during the period
404 1971–1999 the phosphorus content in the Gulf of Baltic proper increased, we found decreasing phosphorus
405 content in the Gulf of Finland, Bothnian Bay, Bothnian Sea and Danish Straits. Largest export and import of
406 phosphorus between sub-basins are found for the exchange between the Baltic proper and the Gulf of Finland,
407 which amount to 24.3 and 22.5 kton yr⁻¹, respectively. However, the largest net exchange (import minus export)
408 appears between the Baltic proper and Bothnian Sea. It is also found that the Baltic proper exports more
409 phosphorus to neighboring sub-basins than it imports, except for the Gulf of Riga. The annual mean net
410 phosphorus exported from the Baltic proper into the Danish Straits, the Bothnian Sea, the Gulf of Finland and
411 Gulf of Riga during the period 1971–1999 amounts to 1.7, 3.6, 1.8, and -0.6 kton yr⁻¹, respectively. The exchange

412 of phosphorus between the Baltic proper and the Gulf of Riga is smallest relative to the other three neighboring
413 sub-basins. Further, we found that the net transport, import and export of phosphorus into the Bothnian Bay are
414 smallest relative to the other sub-basins.

415

416 Nitrogen transports between Baltic Sea sub-basins are different compared to phosphorus transports (Fig. 11).
417 For example, the Baltic proper has larger nitrogen sinks than external sources. Further, during the period 1971–
418 1999 the nitrogen content decreased in the Gulf of Riga and increased in the Bothnian Bay, respectively. In the
419 Gulf of Finland and Danish Straits, the difference between external supply and internal sink of nitrogen is equal
420 to the net transport into the Gulf of Finland and Danish Straits. The large sink of nitrogen in the Bothnian Bay is
421 noteworthy. We also found relatively large net transports of nitrogen from the Gulf of Riga into the Baltic proper.
422 This is mainly explained by the relatively high nitrate concentrations in the Gulf of Riga relative to other sub-
423 basins.

424

425 The 3D nutrient pools constructed by data assimilation methods offer an opportunity to evaluate with improved
426 estimates the changes in Baltic Sea eutrophication. As an example, an investigation of the trophic state from
427 changes in five year average nutrient pools in REANA shows that the total nitrogen pool in the Baltic proper
428 increased from 657 kton to 1045 kton from the period 1971-1975 to 1995-1999 while the total phosphorus pool
429 decreased from 469 kton to 448 kton between the same periods. Hence, nitrogen increased by about 59% while
430 phosphorus decreased by about 4%. Similarly the pool of DIN in REANA increased by 80% while DIP decreased
431 by 6%. The corresponding numbers obtained from BED showed an increase of 100% for DIN and an increase of
432 7% of DIP. The results indicate large increases of nitrogen pools in the Baltic proper during the investigated
433 period but only relatively small changes of phosphorus pools.

434

435 Subsequent periodic assessments can be used to reveal future eutrophication changes. While estimating the
436 trophic state, it should be noticed that the change in trophic state depends on the chosen time periods. For
437 example, from the year 1971 to 1999 the total phosphorus (TP) in the Baltic proper increased with 2.7 kton yr⁻¹.
438 This result differs from the decrease seen from the five year average change discussed above. The reason is the
439 impact from short term fluctuations of nutrient content (Fig. 4) that may be larger than the long-term changes.

441
442 **5.7 Baltic nutrient flows**
443

444 To further analyze the variability of the budget of the reanalyzed nutrients, Fig. 12 provides the cross sectional,
445 integrated nutrient flows in the different sub-basins. Here the eastward and northward net transports are, by
446 definition, positive. Obviously, the integrated nutrient flows vary significantly in space according to the nutrient
447 loads from land. The inflows and outflows also vary depending on the depth of the water column and nutrient
448 concentrations that influence the vertically integrated mass fluxes. In general, the magnitude of nutrient transports
449 declines along transect A from south to north. For instance, the largest annual northward flow of nitrogen in the
450 Baltic proper reaches 392 kton yr⁻¹, while it is only 133 and 87 kton yr⁻¹ for the Bothnian Sea and Bothnian Bay,
451 respectively.

452
453 In the Baltic proper, inflow and outflow as well as the net northward flow of phosphorus increase from the
454 south until a section along 56.8° N; they remain then about constant until a section along 58.7° N, and thereafter
455 decrease rapidly further to the north. This indicates that major sources are located in the south where the large
456 rivers pour their loads into the Baltic Sea, while the major net sinks are mainly found in the northern parts of the
457 Baltic proper. The behavior of net northward flow of nitrogen is different. Nitrogen transports decrease
458 constantly with increasing latitude because the major sink works differently for nitrogen (i.e. denitrification) than
459 for phosphorus, which is retained mainly by burial in the sediments. The net northward flow decreases at the
460 latitude of the Gulf of Finland where phosphorus (and nitrogen) are transported towards the Gulf, as seen in
461 transect C.

462
463 In the Arkona and Bornholm basins, nitrogen and phosphorus transports increase from the west to the east. Due
464 to the nitrogen load from the Oder River, the inflow of nitrogen increases significantly at the border between the
465 Arkona and Bornholm basins, whereas the outflow does not show any discontinuity. As a result, the net flow of
466 nitrogen shows an accelerated increase. The situation for phosphorus in the Arkona and Bornholm basins is
467 different compared to the nitrogen transports because in- and outflow, as well as the net flow, change direction.
468 The phosphorus loads from the Oder River turn the outflow in the western parts into an inflow of phosphorus in
469 the eastern parts.

470

471 In the Gulf of Finland, in- and outflows generally decline from the west to east. In the entrance of the Gulf of
 472 Finland, the net inflows of nutrients are almost zero. The largest net flow (westward) of nutrients appear at the
 473 inner end of the Gulf of Finland, where the large river Neva enter the Gulf, with a magnitude of 33 kton yr⁻¹ for
 474 nitrogen and 2.6 kton yr⁻¹ for phosphorus, respectively. The net flows of both phosphorus and nitrogen change
 475 their directions in the Gulf of Finland and for nitrogen this change take place closer to the Baltic proper entrance
 476 than for phosphorus. These results indicate that the large supply of nutrients from the Neva River are
 477 accumulated or removed within the Gulf of Finland.

478 **6 Discussion**479 **6.1 Biases of FREE**

480

481 RCO-SCOBI has been widely used for the Baltic Sea and the model was carefully evaluated using various
 482 observational data sets. As any other model, RCO-SCOBI had to be calibrated because many processes including
 483 sources and sinks of nutrients are known not in enough detail. Hence, an “optimal” parameterization of
 484 unresolved processes is one of the requirements for the predictive capacity of the model. Further requirements to
 485 calculate correct transports and transformation processes in addition to optimized model equations are high-
 486 quality atmospheric and riverine forcing data, and high-quality initial and lateral boundary conditions.

487

488 Most of the large biases in FREE are caused by imperfect initial conditions. The reason is that the nutrient
 489 pools in the sediments have not been spun up appropriately. As a consequence, phosphate concentrations in
 490 FREE are higher than observed concentrations at all depths. The biases in surface phosphate concentrations
 491 between model results and observations can influence the seasonal primary production. In REANA, however,
 492 from the beginning of the experiment, the biases are already significantly reduced during the first year and remain
 493 relatively small during the integration compared to FREE. This result indicates a need of new initial conditions of
 494 the sediments.

495

496 **6.2 Non-conservation in REANA**

497

498 In the long-term simulation, the new initial condition for an assimilation cycle differs from the ending ocean state
 499 of the last cycle when at that time observations are available. In this sense, the data assimilation introduces
 500 sources and sinks of the nutrient cycles by interrupting the model simulation and adjusting the initial conditions.
 501 The magnitudes of these artificial sources and sinks are directly related to the biases between model results and
 502 observations. Figure 3 shows that the model has large biases during the beginning of the simulation. However,
 503 data assimilation has corrected the mismatch between model state and observation to an “optimal” level during an
 504 initial adjustment period. After the adjustment period, the mismatch between model and observation becomes
 505 small and the successive adjustment due to data assimilation also becomes small. Further, the adjustment of data
 506 assimilation is related to the spatial-temporal coverage of observations. Here we assimilated only observed
 507 profiles into the model. After every assimilation cycle, the simulation continues with “optimal” initial conditions
 508 based upon conservation principles. As the equations of RCO-SCBI have not been changed, masses of all
 509 constituents of the model are conserved at least during the simulation between two assimilation occasions.

510

511 **6.3 Advantages of data assimilation**

512

513 The advantage of the data assimilation is that model variables at any station are very likely more accurate than
 514 the model output without data assimilation. For instance, time series of profiles or transports across vertical
 515 sections have very likely a smaller bias compared to observations than the corresponding model results without
 516 data assimilation. Compared to available observations the information from the model is higher resolved and
 517 homogeneous in space and time. Of course, it is difficult to evaluate the quality of model results at high
 518 resolution because independent observational data sets are usually missing. An exceptional effort to utilize
 519 independent data was done by Liu et al. (2014) showing that the statement about the added value of data
 520 assimilation is true for [one](#) available, independent cruise data [set](#) at high resolution.

521

522

523

524

525

[The results of the reanalysis can be used to estimate the water quality and ecological state with high spatial and temporal resolution in regions and during periods when no measurements are available. This supports improved assessments e.g. of eutrophication status indicators as exemplified in Section 5.6.](#) Regional and local model studies may use the data as initial and boundary conditions. For projections of future climate and for

526 nutrient load abatement scenario simulations the reanalysis has a very high scientific value as reference data set
527 for the historical period of the climate simulations. The evaluation of the regionalized climate (the statistics of
528 mesoscale variability, e.g. the mean state) during the historical period can be done much more accurate based
529 upon the reanalysis data than with sparse observational data. For instance, it is very difficult to calculate the
530 climatological mean state just from observations that are casted only during the ice-free season of the year. Using
531 a reanalysis as reference data for historical climate is a common method in regional climate studies of the
532 atmosphere. Here we provide a corresponding data set for the ocean to evaluate simulated present-day climate.
533

534 Further, nutrient transports across selected cross-sections or between vertical layers are calculated from the
535 reanalysis with high resolution and improved accuracy. However, one cannot expect that budgets calculated from
536 the summation of internal fluxes from model results with data assimilation are more accurate because usually
537 small artificial sources and sinks from the data assimilation are becoming as important as physically and
538 biogeochemically motivated sources and sinks when sums of fluxes are compared. Hence, we calculated in
539 Section 5.6 budgets only from external supply, imports and exports and changes in the water pools of nutrients
540 with the aim to compare the reanalysis results with other studies using only observations. It is perhaps not
541 possible to claim that our budgets are more accurate than budgets that are derived from observations only, despite
542 the higher temporal and spatial resolution in model outputs. However, the advantage of the reanalysis is that
543 measurements are extrapolated in space and time based upon physical principles of the model.
544

545 **6.4 Comparison with other assimilation methods**

546 Fu (2013) estimated the volume and salt transports during the 2003 MBI with a three-dimensional variational
547 data assimilation method (3DVAR) in the Baltic Sea. In the present study, we estimate the impact of the data
548 assimilation based upon the EnOI method on the net volume and nutrient transports as well as calculate budgets
549 for major sub-basins of the Baltic Sea. The volume transports obtained with different assimilation methods may
550 be different. The sea level in Fu (2013) is kept constant in the assimilation process, while sea level in this study is
551 varying accordingly during the assimilation of temperature and salinity based upon the statistical covariances.
552 The variability of sea level may enhance the barotropic flow, which is one of the reasons for the differences in net
553 volume transport between the results by Fu (2013) and REANA. However, transports within the sub-basin are
554 also indirectly affected by the interaction of baroclinicity and topography.

556 | **6.5 Comparison with other studies on nutrient budgets**

558 In contrast to Eilola et al. (2012), in this study areas with DIN export are also found at the southern and eastern
 559 coasts as well as at some small local regions in the inner parts of the Baltic proper (Fig. 8). In REANA, the
 560 magnitudes of DIP imports and exports are larger than in Eilola et al. (2012), and there is pronounced import of
 561 DIP in the western part of the Eastern Gotland Basin below 100 m (Fig. 8) that is not as significant in Eilola et al.
 562 (2012). This, and the larger variability of DIN imports and exports, indicates that there is a higher degree of
 563 small-scale localized transport and production patterns that are not captured by Eilola et al. (2012). Main sinks of
 564 DIN are found in the deeper areas, but significant sinks are also seen in shallow areas and water depths of about
 565 60m. As the assimilation of salinity observations result in a deeper halocline (Liu et al., 2014), the bottom water
 566 in a depth range of 40–70 m contains higher oxygen concentrations than in the simulation without data
 567 assimilation. Hence, in the REANA simulation of this study, more phosphorus is retained by the sediments in the
 568 depth range of 40–70 m than in the simulation by Eilola et al. (2012). The present results show, however, an
 569 export contribution from DIN sources in deeper areas (e.g. 60–90 m depths) that may have been caused by
 570 reduced denitrification efficiency of oxidized sediments in the REANA simulation compared to Eilola et al.
 571 (2012).

573 The in- and outflows of phosphorus between the sub-basins, except the Gulf of Riga and Gulf of Finland,
 574 simulated in REANA are smaller than the results by Wulff and Stigebrandt (1989), Savchuk (2005) and Savchuk
 575 and Wulff (2007). However, the net transports of phosphorus are similar between our results and these earlier
 576 studies in all sub-basins. Moreover, the nitrogen budgets are much lower than the results of earlier studies,
 577 especially in the Baltic proper. It should be kept in mind that the above mentioned studies estimated the nutrient
 578 budgets from mass balance models together with inter-basin transport calculations based upon Knudsen's
 579 formulae to calculate nutrient budgets of the Baltic Sea (see, e.g. Savchuk, 2005). Obviously, there are limitations
 580 in calculations of previous studies. Despite overall uncertainties that also limit the reliability of our results, like
 581 incomplete understanding of selected biogeochemical processes (e.g. nitrogen fixation), lacking information of
 582 sediment parameters, and under-sampled observations in space and time, our approach has the advantage of using
 583 both high-resolution modeling and all available observations made over a 30-year period. Our model results

584 consider the complete set of primitive equations in high-resolution, taking into account not only the volume and
585 salt conservation of sub-basins according to Knudsen's formulae, but also the wind-driven circulation between
586 and within sub-basins. Hence, we have, for the first time, the potential to quantify spatial transport patterns with
587 high confidence even within sub-basins, as in the exchange of nutrients between the coastal zone and the open
588 sea.

589

590 Eutrophication of the Baltic Sea is directly affected by the long-term evolution of external nutrient supply that
591 has three components (waterborne land loads, direct point sources at the coasts, and atmospheric depositions)
592 which are associated with the biogeochemical dynamics of the Baltic Sea. In our study, we used the reconstructed
593 external nutrient input data by Savchuk et al. (2012). Nutrient budgets (Figs. 10 and 11) of sub-basins are time-
594 averaged and represent in our study the overall results of the period 1970–1999. The phosphorus loads vary in
595 different periods, for example, the phosphorus loads in the 1980s are larger relative to the 1990s (see Savchuk et
596 al., 2012). Therefore, the phosphorus supply into the Gulf of Finland is greater in our study compared to Savchuk
597 and Wulff (2007). The greater phosphorus supply changes the phosphorus content and phosphorus concentration
598 in the Gulf of Finland. This is one reason why phosphorus transports between the Gulf of Finland and the Baltic
599 proper in our study are greater than the transports calculated by Savchuk (2005) and Savchuk and Wulff (2007).

600

601 Since our study covers a different time period compared to the studies by Wulff and Stigebrandt (1989),
602 Savchuk (2005) and Savchuk and Wulff (2007) nutrient concentrations and related budgets differ in time and
603 space. Hence, it is not surprising that other studies show deviating results. For example, during the period 1970–
604 1999, HELCOM (2013) showed that the TP concentration generally decreased in the Bothnian Bay and has
605 increased in the Gulf of Riga. However, these changes in TP concentrations were not monotonous. For example,
606 the TP concentration obviously increased during the period 1970–1976 in the Bothnian Bay. While in the
607 Bothnian Sea, TP concentration increased during the period 1970–1983 and decreased during the period 1990–
608 1999. Similarly, changes in total nitrogen concentration differed during different periods.

609

610 Gustafsson et al. (2012) used a process-oriented model that resolves the Baltic Sea spatially in 13 dynamically
611 interconnected and horizontally integrated sub-basins with high vertical resolution to reconstruct the temporal
612 evolution of eutrophication for 1850–2006. Savchuk (2005) and Savchuk and Wulff (2007) applied mass balance

613 models as mentioned above to calculate nutrient budgets of the Baltic Sea. The results of all these models depend
614 on the locations of the sub-basin borders which are chosen as far as possible according to dynamical constraints
615 such as sills or fronts that are parameterized to obtain estimates of the water exchanges. Using a high-resolution
616 circulation model, we showed that nutrient flows within the sub-basins may vary considerably (Fig. 12). For
617 instance, we found east- and westward net transports of nitrogen between the Baltic proper and Gulf of Finland
618 depending on border locations at 23.2° and 24.0° E, respectively. The importance of regional variations of sources
619 and sinks for nutrients on the calculation of transports between sub-basins therefore seems to be significant and
620 needs to be further studied. Given the uncertainty caused by data assimilation in the present study we must
621 however save the detailed studies of these issues to future work where the artificial impact of data assimilation on
622 sources and sinks will be traced and quantified during the run.

623 7 Summary and Conclusion

624 For the first time, a multi-decadal, high-resolution reanalysis of physical (temperature and salinity) and
625 biogeochemical variables (oxygen, nitrate, phosphate and ammonium) for the Baltic Sea was presented. The
626 reanalysis covers the period 1970–1999. A “weakly coupled” assimilation scheme using the EnOI method was
627 used to assimilate all available physical and biogeochemical observations into a high-resolution circulation model
628 of the Baltic Sea.

629 Both assimilated and independent observations collected from different databases were used to evaluate the
630 reanalysis results (REANA). Based on the model–data comparison presented in this study, we found that the
631 model results without data assimilation (FREE) exhibit significant biases in both oxygen and nutrients. The
632 reasons for these biases are not totally understood yet, although it is speculated that the main reasons might be
633 related to the imperfect initial conditions, limitations of model parameterizations, the inaccurate halocline
634 position and correspondingly the hypoxic volume (Liu et al. 2014). Based on the calculation of the overall RMSD
635 of oxygen and nutrient concentrations between model results and not-yet-assimilated observations, the results in
636 REANA are considerably better than those in FREE. The total RMSD of the oxygen, nitrate, phosphate and
637 ammonium is reduced respectively by 0.84 mL L⁻¹, 0.99 mmol m⁻³, 0.88 mmol m⁻³, 0.52 mmol m⁻³. This means

639 that the overall qualities of simulated oxygen, nitrate, phosphate, and ammonium concentrations are improved by
640 59, 46, 78 and 45%, respectively. These results demonstrate the strength of the applied assimilation scheme.
641

642 The observation information entering the model affects the oxygen dependent dynamics of biogeochemical
643 transports significantly due to both improved simulation of physical (e.g. vertical stratification) and
644 biogeochemical parameters (e.g. nutrient concentrations). As examples, we presented improved results of mean
645 seasonal cycles of nutrients, the spatial surface distributions of DIN, DIP and DIN:DIP of the entire Baltic Sea.
646

647 Based on the reanalysis simulation, we analyzed nutrient transports in the Baltic Sea. We found that vertically
648 integrated nutrient transports follow the general horizontal water circulation, and vary spatially to a large extent.
649 In particular, large nutrient transports were found in the Eastern Gotland Basin, in the Bornholm Basin, in the
650 Slupsk Channel and in the north-western Gotland Basin. The persistence of nutrient transports is greater in the
651 eastern and southern than in the northern and western Baltic Sea.
652

653 The horizontal distributions of sources and sinks of inorganic and organic nutrients show large spatial
654 variations and may be partly explained by (1) the external supply of nutrients from land, (2) the topographically
655 controlled horizontal nutrient exchange between sub-basins and between the coastal zone and the open sea, and
656 (3) vertical stratification that determines redox conditions at the sea floor. The latter is important for the
657 sediment-water fluxes of nutrients, and consequently for burial of nutrients in the sediments. The reanalysis
658 results suggest that in the Baltic proper, in most areas with a water depth less than the depth of the permanent
659 halocline at about 70–80 m, DIP is imported and transformed either to OrgP, or buried in the sediments in water
660 depths greater than the wave-induced zone at 40–70 m. Whether the latter is an artefact of the assimilation
661 method or a real sink is unclear. On the other hand, in areas with greater water depth, DIP is exported (e.g.
662 released from the sediments under anoxic conditions). Overall, the Baltic proper exports DIP to neighboring sub-
663 basins.
664

665 Nitrogen transports are very different compared to phosphorus transports. The shallow coastal zone with water
666 depths less than 10 m plays an outstanding role for DIN, because within it, large exports occur due to supplies
667 from land. The high productivity in the shallow areas effectively transfers DIN to OrgN and denitrification

668 decreases the exports of nitrogen from coastal areas to the deeper areas. Most of the exported DIN is removed in
669 shallow waters while at greater depths imports and exports of DIN are much smaller, indicating the important
670 role of the coastal zone for nitrogen removal.

671
672 Detailed nitrogen and phosphorus budgets suggest that nutrient transports in the various sub-basins are
673 controlled by different processes and show different response to external loads and internal sources and sinks. In
674 particular, the Baltic proper is the sub-basin with the largest nutrient exchanges with its surrounding sub-basins.
675 The Baltic proper exports phosphorus to all sub-basins except the Gulf of Riga. Similarly, the Baltic proper also
676 exports nitrogen to all sub-basins except to the Gulf of Riga and Danish Straits. In this sub-basin, also the largest
677 internal sink of all sub-basins was found. Noteworthy is the relatively large net export of phosphorus from the
678 Baltic proper into the Bothnian Sea. This finding is in agreement with previous studies. For the budgets of the
679 sub-basins, it is important where the borders of the sub-basins are located, because net transports may change
680 sign with the location of the border. For instance, in the entrance of the Gulf of Finland, the net phosphorus
681 transport from the Baltic proper is directed eastward, but changes direction at about 26°E. Further to the east, the
682 net phosphorus transport is directed westward.

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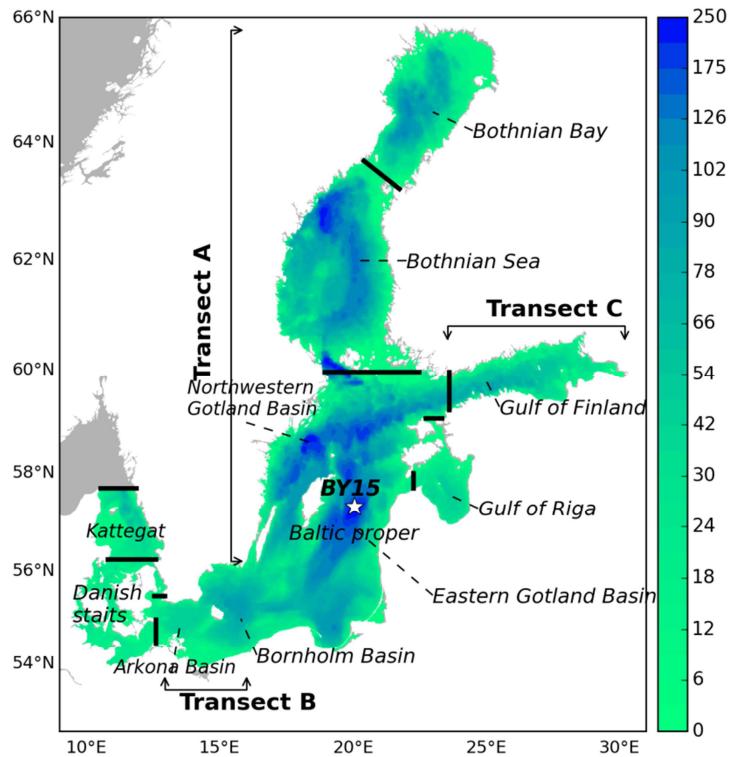
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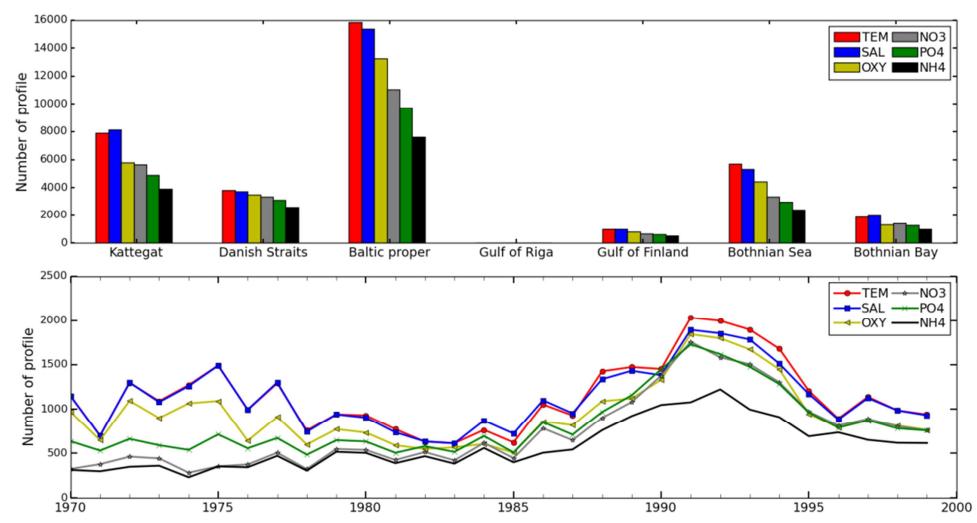
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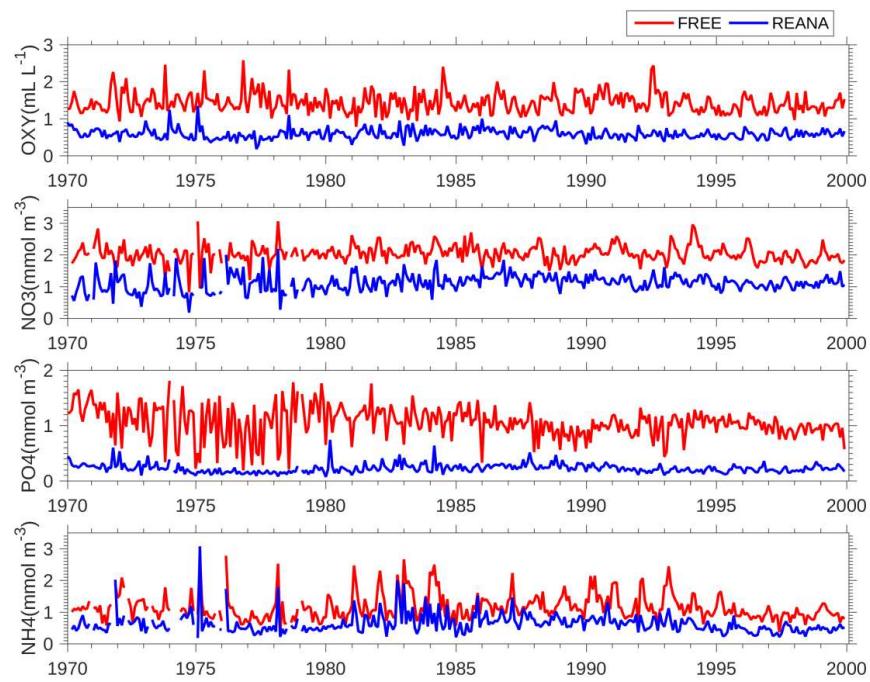
854 Figure 1. The bathymetry of the model (depth in m). The border locations of sub-basins of the Baltic Sea used in
 855 this study are shown by the black lines, and the BY15 station is shown by the white star. [Names of the sub-basins](#)
 856 [are the Kattegat \(KT\), Danish Straits \(DS\), the Baltic proper \(BP\), the Gulf of Riga \(GR\), the Gulf of Finland](#)
 857 [\(GF\), the Bothnian Sea \(BS\), and the Bothnian Bay \(BB\).](#)

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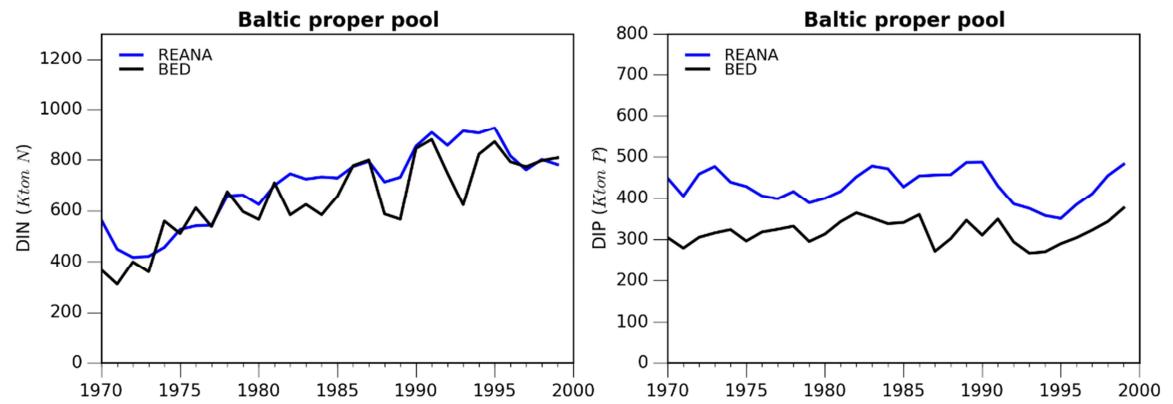


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863 Figure 2. Number of observed profiles in different sub-basins (upper panel) and annual number of profiles from
 864 1970–1999 (bottom panel).



867 Figure 3. Monthly mean root mean square deviation (RMSD) between model results and observations for
 868 oxygen, nitrate, phosphate and ammonium in FREE (red) and REANA (blue).

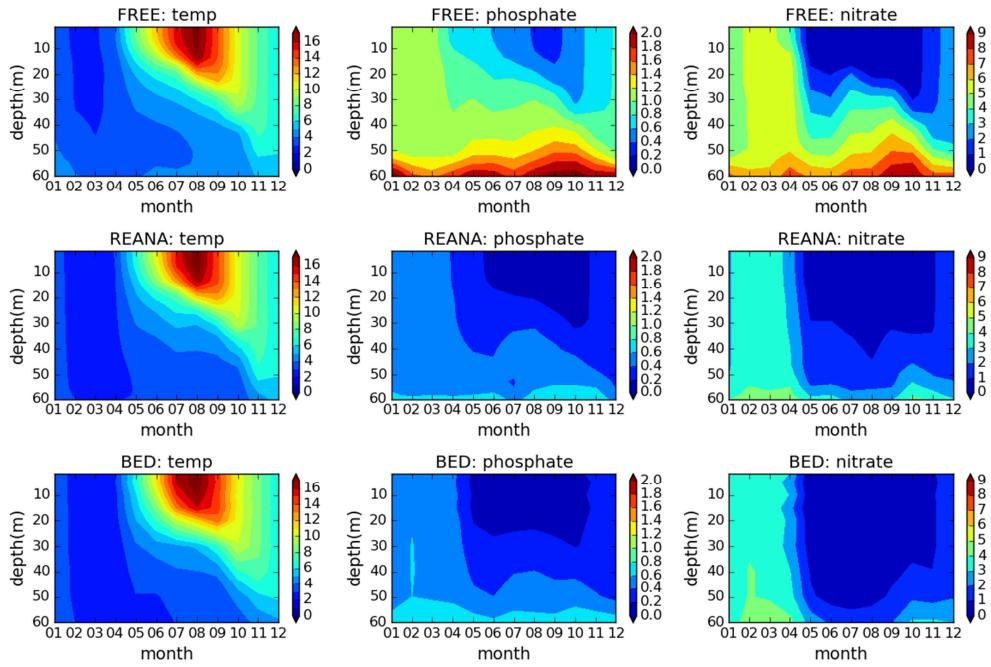


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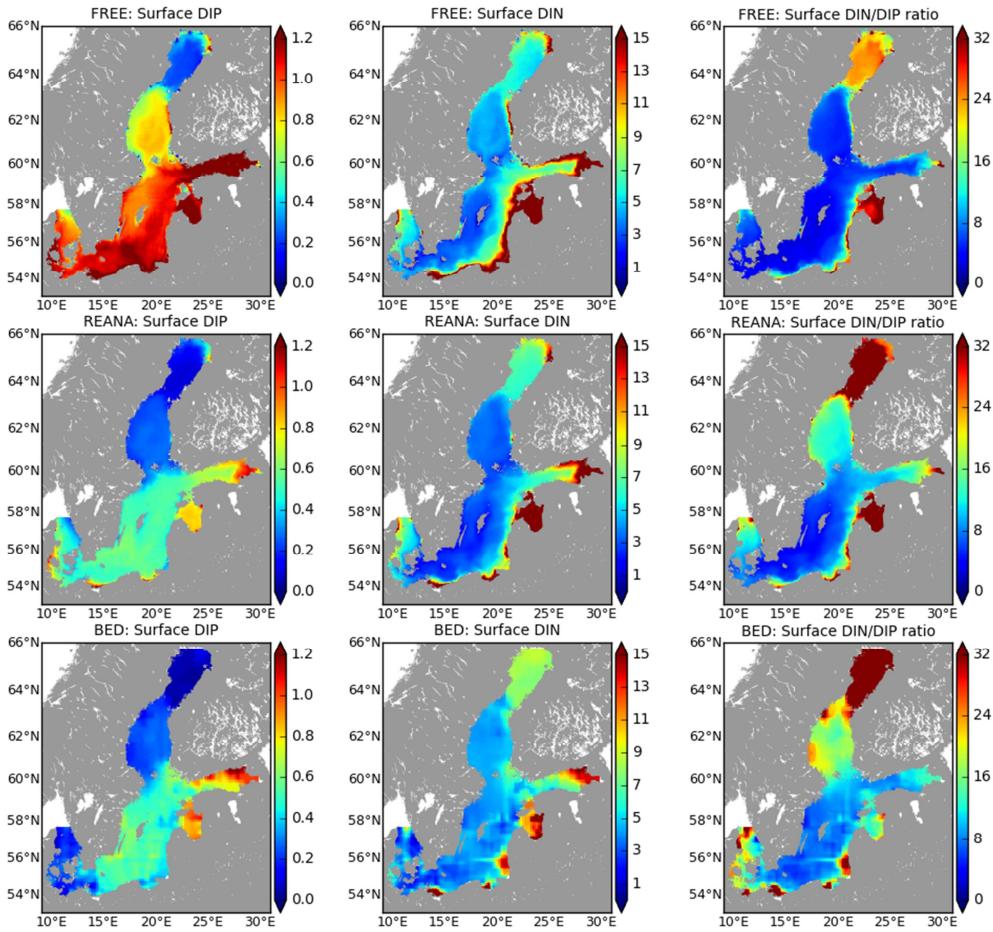
882 Figure 4. Annual mean integrated pools (in kton) of pelagic DIN and DIP in the Baltic proper calculated from

883 REANA and from observations in BED.



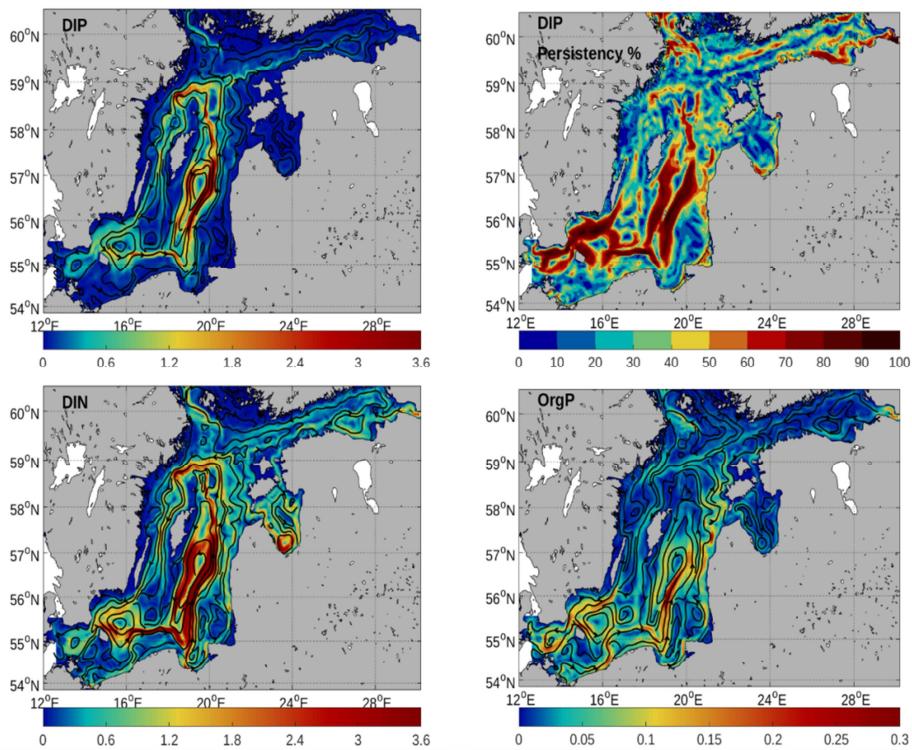
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885 | Figure 5. The seasonal cycle of monthly average (1970–1999) temperature ($^{\circ}\text{C}$), phosphate concentration (mmol m^{-3}),
 886 and nitrate concentration (mmol m^{-3}) at BY15 for FREE (row 1), REANA (row 2), and BED data (row 3),
 887 respectively.

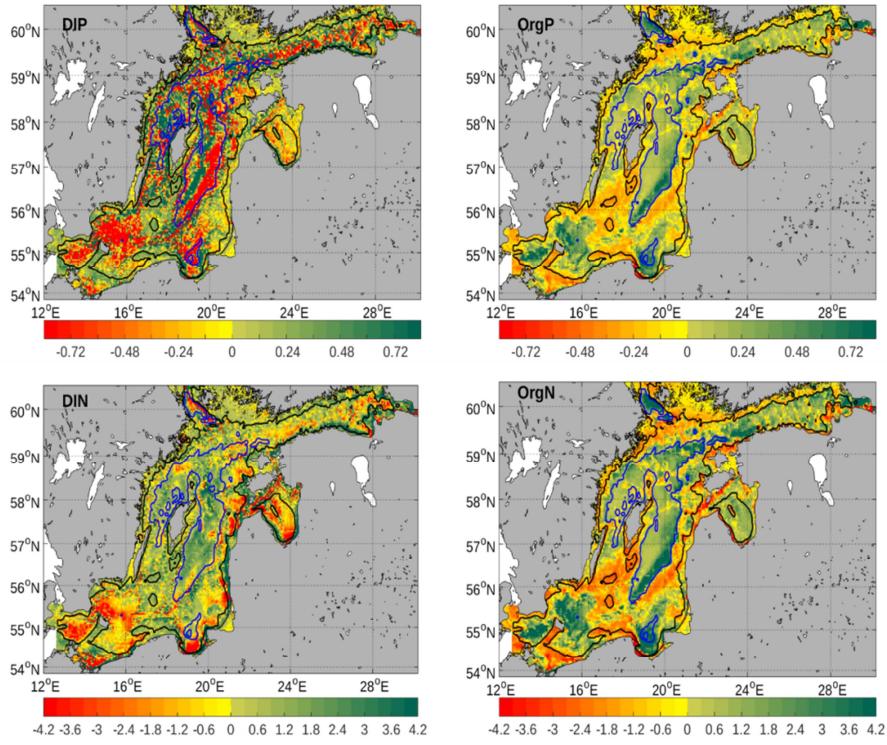


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889 | Figure 6. Simulated monthly (March) mean (1970–1999) surface layer (0–10 m) concentrations of DIP (mmol m⁻³) (left), DIN (mmol m⁻³) (middle), and the corresponding DIN to DIP ratio (right) from FREE and REANA are
 890 | shown in rows 1 and 2, respectively. The corresponding BED maps in row 3 are calculated from observations
 891 | monitored during the period 1995–2005.



895 | Figure 7. Annual mean DIP transports and the corresponding DIP persistency, DIN and OrgP transports for
 896 | REANA averaged for the period 1970–1999. The black solid lines with arrows show the streamlines and
 897 | direction of transports. The magnitude of transports ($\text{kton km}^{-1} \text{ yr}^{-1}$) and the persistency (%) are shown by the
 898 | background color. The corresponding values are shown in the colored bars.



900
901
902 | Figure 8. Spatial distributions of annual mean import of DIP, OrgP, DIN and OrgN averaged for the period 1970–
903 | 1999. The magnitude of import and its corresponding value ($\text{kton km}^{-2} \text{ yr}^{-1}$) are shown by the background color
904 | and color bar, respectively. Green colors denote positive values (import), and yellow to red colors denote
905 | negative values (export). The black and blue lines show 30 and 100 m depth contours of the model, respectively.

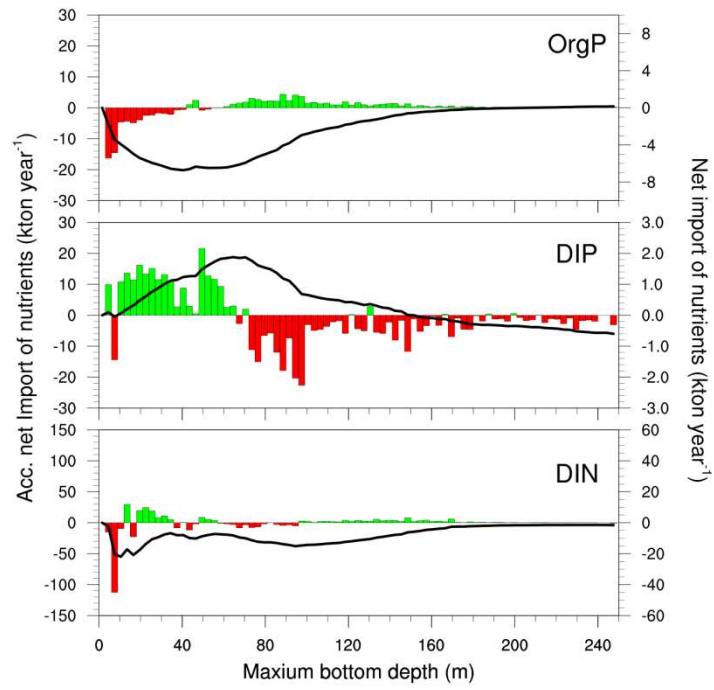
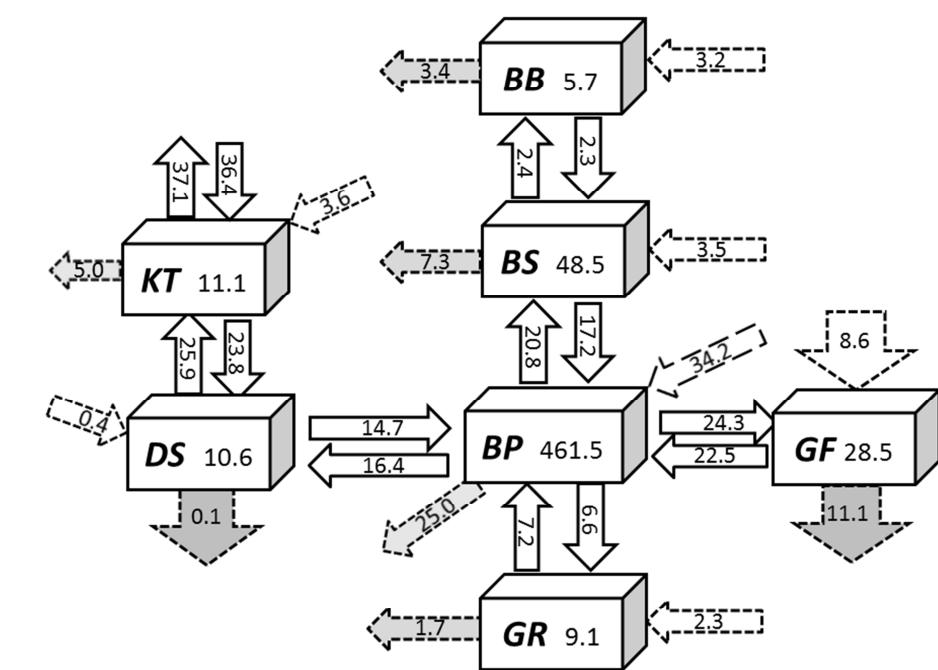
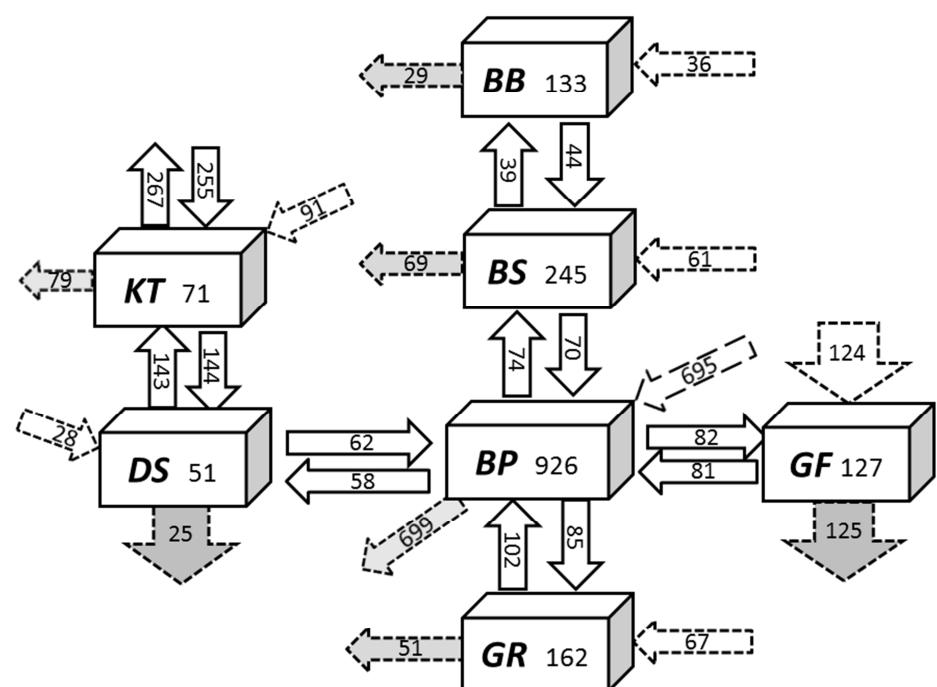


Figure 9. Annual mean, accumulated net imports (black lines) and imports of OrgP, DIP and DIN (color bars) to regions with the same depth in the Baltic proper averaged for the period 1970–1999.



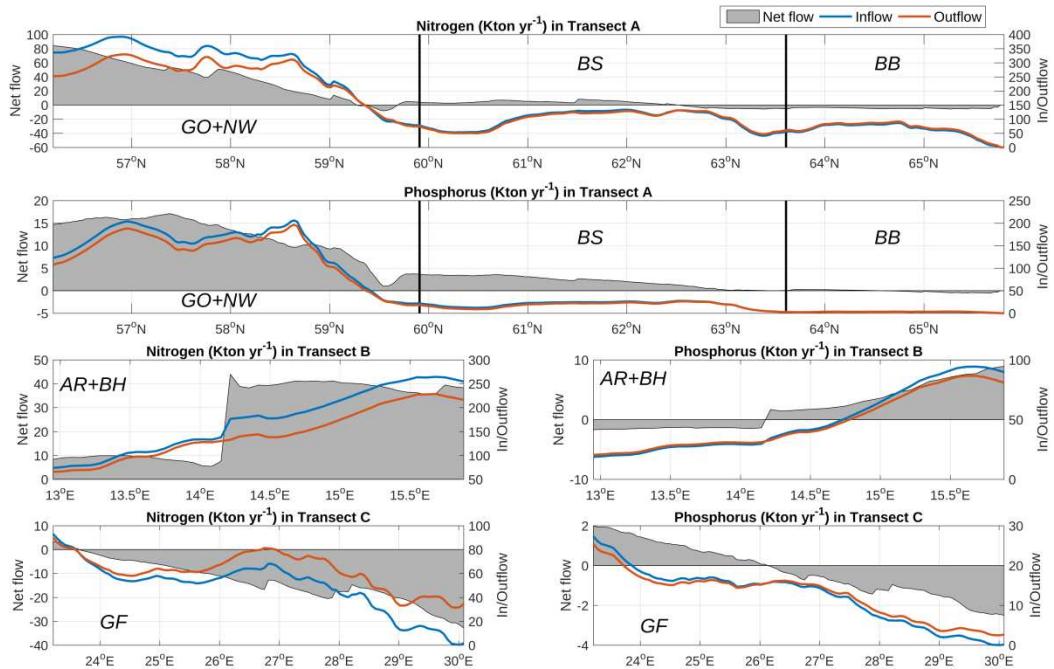
908 | Figure 10. Annual mean total phosphorus budgets of the Baltic Sea averaged for the period 1971–1999. The
 909 | average total amounts are in kton, and transport flows and sink/source fluxes (external nutrient inputs/[sink](#)) are in
 910 | kton yr^{-1} . External nutrient inputs from atmosphere and land are combined.

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912

913 | Figure 11. The same as Figure 9, but for nitrogen.



914

915 | Figure 12. Annual mean fluxes of nitrogen (in kton yr^{-1}) and phosphorus (in kton yr^{-1}) as a function of the cross
 916 | sections along transects following the latitude and longitude in the Baltic sub-basins. Northward and eastward
 917 | fluxes are, by definition, positive and called inflows. Southward and westward flows are called outflows. Net
 918 | flow is the difference between inflow and outflow. Here, AR, BH, GO, NW, GF, BS, and BB represent the Arkona
 919 | Sea, Bornholm Sea, Eastern Gotland Basin, Northwestern Gotland Basin, [Gulf of Finland](#), Bothnian Sea and
 920 | Bothnian Bay, respectively. Transect A summarizes fluxes from the southern Baltic proper to the Bothnian Bay.
 921 | Transect B describes the Baltic Sea entrance area from the Arkona Basin to the Bornholm Basin, and transect C
 922 | summarizes fluxes in the Gulf of Finland (see Fig. 1).