

1 Exploring the distance between nitrogen and phosphorus 2 limitation in mesotrophic surface waters using a sensitive 3 bioassay

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13 Abstract

14 The balance in microbial net consumption of nitrogen and phosphorus was investigated
15 in samples collected in two mesotrophic coastal environments: the Baltic Sea
16 (Tvärminne field station) and the North Sea (Espegrend field station). For this, we have
17 refined a bioassay based on the response in alkaline phosphatase activity (APA) over a
18 matrix of combinations in nitrogen and phosphorus additions. This assay not only
19 provides information on which element (N or P) is the primary limiting nutrient, but
20 also gives a quantitative estimate for the excess of the secondary limiting element (P⁺ or
21 N⁺, respectively), as well as the ratio of balanced net consumption of added N and P over
22 short time scales (days). As expected for a Baltic Sea late spring-early summer situation,
23 the Tvärminne assays (n=5) indicated N-limitation with an average P⁺=0.30±0.10 μM-P,
24 when incubated for 4 days. For short incubations (1–2 days), the Espegrend assays
25 indicated P-limitation, but the shape of the response surface changed with incubation
26 time, resulting in a drift in parameter estimates toward N-limitation. Extrapolating back
27 to zero incubation time gave P-limitation with N⁺≈0.9 μM-N. The N:P ratio (molar) of
28 nutrient net consumption varied considerably between investigated locations; from
29 2.3±0.4 in the Tvärminne samples to 13±5 and 32±3 in two samples from Espegrend.
30 Our assays included samples from mesocosm acidification experiments, but statistically
31 significant effects of ocean acidification were not found by this method.

32 Keywords: alkaline phosphatase activity, bioassays, mesotrophic temperate seas,
33 nutrient limitation, phytoplankton

34 1 Introduction

35 N to P balance is a core biogeochemical feature of aquatic systems as highlighted in
36 Redfield's classical question of whether it is the chemistry of seawater that has
37 determined the stoichiometry of the marine organisms, or biology is the cause for the
38 "normal" 16:1 (molar) ratio between N and P in seawater (Redfield et al., 1963). The

1 issue of surface ocean nutrient limitation is as acute as ever (Moore et al., 2013), since it
2 has bearings on phenomena ranging from the global carbon cycle, where it plays a key
3 role in the dynamics of the ocean's biological pump (Ducklow et al., 2001); via basin
4 scale issues such as N deficiency in Arctic water of Pacific origin (Lehmann et al., 2005),
5 P deficiency in Eastern Mediterranean deep waters (Krom et al., 1991) and the North
6 Atlantic gyre (Mather et al., 2008); via regional issues such as the question of P and/or N
7 removal from the Baltic Sea (Elmgren and Larsson, 2001; Granéli et al., 1990; Råike et
8 al., 2003); to local ecosystem characteristics such as P-deficient brackish layer
9 overlaying potentially more N-limited marine waters in the fjords of western Norway
10 (Thingstad et al., 1993). The classical idea of predominantly N-limitation in marine
11 systems (as opposed to predominantly P-limitation in limnic systems) (Hecky and
12 Kilham, 1988) has also become considerably more nuanced, not only due to the cases
13 mentioned above, but also with the identification of the High Nutrients Low Chlorophyll
14 (HNLC) areas as being iron-limited (Franck et al., 2003), phosphorus and iron as co-
15 limiting elements of nitrogen fixation in the tropical North Atlantic (Mills et al., 2004)
16 and N₂ fixers potentially having a competitive advantage in oligotrophic P-starved
17 regions (Landolfi et al., 2015). While some of the mechanisms behind these apparent
18 deviations from Redfield stoichiometry seem to be well understood, there are others
19 which lack generally accepted explanations.

20 In deep waters with most of the bioavailable N and P converted to NO₃ and PO₄, the
21 chemical determination of N or P in excess of the Redfield ratio may be relatively
22 straight forward. In biogeochemistry this excess is calculated on the basis of measured
23 nitrate and phosphate, and is referred to as N* and P*, e.g. $N^* = NO_3 - 16PO_4 + 2.9 \text{ mmol m}^{-3}$
24 (Sarmiento and Gruber, 2006). In productive surface waters this is a more complex
25 issue. A potential solution to the chemically intractable problem of measuring a large
26 suite of presumably bioavailable pools is to use a quantitative bioassay, i.e. to ask the
27 organisms how much of the primary and secondary limiting elements they can "see".

28 In productive waters, both N and P may accumulate over time in pools of DON and DOP
29 with different grades of bioavailability. Microbes have flexible stoichiometry as their
30 content of storage materials, structural carbohydrates, nucleic acids and lipids vary with
31 growth conditions (Bertilsson et al., 2003; Geider et al., 2002; Krauk et al., 2006). There
32 are also differences in the stoichiometry of different functional groups of organisms,
33 where e.g. bacteria (Fagerbakke et al., 1996) tend to have N:P ratio significantly lower
34 than 16. Which element that first becomes limiting, and how much of the secondary
35 limiting element then remains in excess may thus depend not only on the total pools as
36 conceptually expressed by N*, but to vary as a function of the biological structure of the
37 food web and its pre-history. Although conceptually related to N*,P*, the answer to what
38 excess nutrients the organisms "see" may therefore differ even between systems with
39 the same, chemically defined N*. We therefore have chosen to use the symbols N⁺ and P⁺
40 for surplus nitrogen and phosphorus as determined by bioassays, to distinguish these
41 numbers from their chemically defined analogs N* and P*.

42 Microorganisms have evolved sophisticated physiological mechanisms to adapt to the
43 different forms of nutrient limitation (Geider et al., 1997; Ivančić et al., 2012; Thingstad
44 et al., 2005; Van Mooy et al., 2009; Lin et al., 2016), including the induction of
45 extracellular enzymes such as alkaline phosphatase (AP) catalyzing the hydrolysis of
46 phosphomonoesters within DOP (Hoppe, 2003). A well-studied model system is the
47 induction of the Pho-regulon in *Escherichia coli*, which leads to expression of a series of
48 P-starvation related genes, including *phoA* coding for AP synthesis (Torriani-Gorini,
49 1994). The induction of AP synthesis seems to be more coupled to a low internal cell

1 quota of P, than directly to low external concentrations of inorganic P (Lin et al., 2016),
2 thus presumably providing a main signal when both external pools and internal storage
3 reserves of P have been depleted below the certain level (Boekel and Veldhuis, 1990;
4 Chróst and Overbeck, 1987). Inducible AP synthesis is a wide-spread feature in
5 microorganisms (Jansson et al., 1988). It is easily measured as AP activity (APA) (Perry,
6 1972, Hoppe, 2003), and thus it has been frequently used as an indicator of P-stress
7 (Jansson et al., 1988; Dyhrman and Ruttenberg, 2006; Lomas et al., 2010).
8 This method was further exploited by Thingstad and Mantoura (2005) in the
9 oligotrophic Eastern Mediterranean, showing that the concentration of added PO_4
10 needed for APA to disappear in a P-limited system, or alternatively the NH_4 needed to
11 induce APA in an N-limited system, could be used as a bioassay to quantitatively
12 estimate N^+ and P^+ , respectively. We here expand this technique by using a matrix-setup
13 including simultaneous gradients in both PO_4 and NH_4 additions. This is applied to
14 samples from the coastal waters of western Norway and the Baltic Sea, confirming that
15 the assay gives informative results also in temperate, mesotrophic environments.

16 **2 Material and Methods**

17 2.1 Study areas and sampling

18 Part of the sampling for this study was performed in mesocosms designed to study
19 acidification effects. In the Baltic, the water was collected as integrated samples (depth
20 0–10 m) in Storfjärden near Tvärminne field station (59° 51.50' N, 23° 15.50' E) on 6
21 August 2012. The collection was performed 45–30 days after the first-last CO_2
22 treatments and 50 days after the mesocosm closure (Paul et al., 2015). Samples were
23 collected from the fjord (417 μatm) and mesocosms M1 (365 μatm), M3 (1007 μatm),
24 M6 (821 μatm) and M7 (497 μatm); where numbers in parentheses are average $f(\text{CO}_2)$
25 over the period Day 1–Day 43. The mesocosms received no nutrient manipulations
26 except the CO_2 treatments. Further details about location and the mesocosm experiment
27 can be found in Paul et al. (2015) and Nausch et al. (2016).

28 The samples from western Norway were collected during a similar mesocosm
29 experiment in Raunefjorden close to Espegrend field station (60° 16.2' N, 5° 11.7' E).
30 From one mesocosm (MR) an integrated (depth 0–20 m) sample (1165 μatm) was
31 collected on 25 May 2015 corresponding to Day 22 after acidification treatment. The
32 fjord sample was collected at nearby landlocked location Kviturspollen (60° 15.8' N, 5°
33 15' E) at the depth of 1 m using a Niskin sampler on 3 June 2015. Samples were pre-
34 filtered through gauze of 112 μm mesh size to minimize the variability due to the
35 occasional large zooplankton.

36 2.2 Matrices of nitrogen and phosphorus additions

37 Samples were distributed in 15 mL Falcon® polypropylene tubes (BD Biosciences®)
38 organized in 10x10 or 8x8 columns x rows (Tvärminne and Espegrend, respectively).
39 PO_4 (KH_2PO_4 10 μM) was added in final concentrations from 0 to 290 nM-P in steps of
40 32.2 nM (Tvärminne) and from 0 to 105 nM-P in steps of 15 nM (Espegrend). Each of the
41 columns received additions of NH_4 (NH_4Cl 200 μM) in final concentrations from 0 to 964
42 nM-N in steps of 107 nM-N (Tvärminne) and from 0 to 2100 nM-N in steps of 300 nM-N
43 (Espegrend). The tubes were incubated in light:dark (16 h:8 h) at 17–18°C (Tvärminne)
44 and in light:dark (12 h:12 h) at 16.5°C (Espegrend), both at irradiance of 78 μmol

1 photons $\text{m}^{-2} \text{s}^{-1}$. Incubation at Tvärminne lasted 4 days for all samples, whilst APA assays
2 for Espegrend were repeated as given in each case.

3 2.3 Alkaline phosphatase activity

4 Measurements of APA were done according to Perry (1972) using 3-o-methyl-
5 fluorescein- PO_4 (final concentration $0.1 \mu\text{M}$) as the substrate. Volumes were modified to
6 the use of fluorescence plate reader by pipetting $200 \mu\text{L}$ subsamples from each Falcon
7 tube into the wells containing the substrate. Results are expressed as increase in relative
8 fluorescence units per hour (RFU h^{-1}). APA in the coastal waters of the western Norway
9 was measured using a PerkinElmer Enspire 2300 plate reader programmed to do 15
10 repeated measurements (time interval 5 min) over a total incubation time of 70
11 minutes. APA was calculated as the slope of the fitted linear regression line. APA in the
12 Baltic Sea was measured by Varian Cary Eclipse fluorometer after 30 minutes incubation
13 with substrate.

14 2.4 Fitting the response surface

15 To interpret the data obtained by this method, an objective algorithm is needed to define
16 the transition between subsamples with high (P-limited) and low (N-limited) post-
17 incubation APA. Thingstad and Mantoura (2005) did this by fitting sigmoidal functions
18 to the observed APA-responses; either a decreasing function parallel to the P-addition
19 axis in the case of a P-limited system, or an increasing function parallel to the N-addition
20 axis in the case of N-limitation. To avoid this pre-fitting choice of function, we here have
21 instead started with the assumption that the P,N -plane is split into a P-limited and an N-
22 limited region by the straight line:

$$23 \quad N = N_0 + rP \text{ Eqn. 1}$$

24 where a negative value of the intercept N_0 corresponds to the excess-N (N^+) present in a
25 P-limited system and $P_0 = \frac{-N_0}{r}$ is the amount of phosphate needed to shift the system to
26 N-limitation. Conversely, a positive value of the intercept N_0 would correspond to the
27 amount of N required to shift an N-limited system into P-limitation, while $P_0 = \frac{N_0}{r}$ then
28 corresponds to the excess-P (P^+) in this N-limited system. The shift from P- to N-
29 limitation, and therefore the expression of APA in a point P,N is assumed to be a function
30 of the distance Z between this point and the line (Fig. 1). The sigmoidal function fitted is:

$$31 \quad \text{APA}_{est} = \frac{A}{1+e^{sZ}} - B \text{ Eqn. 2}$$

32
33 From the geometry of Fig. 1 one can calculate the perpendicular distance Z from the
34 point P,N to the line defined by Eqn.1 as

$$35 \quad Z(P, N) = \frac{1}{\sqrt{1+r^2}} (rP - (N - N_0)). \text{ Eqn. 3}$$

36
37 Here, the exponential function in the denominator of Eqn. 2 replaces the term $\left(\frac{Z}{Z_0}\right)^s$
38 adopted by Thingstad and Mantoura (2005) from standard calculation of lethal
39 concentration (i.e. LC_{50}) in toxicology. This standard expression is undefined for $Z_0=0$
40 and therefore not applicable with our approach where $Z = 0$ along the line defined by
41 Eqn. 1. Visual inspection of residuals in graphs (see Fig. 2A, B) did not suggest
42 systematic deviances between response surfaces fitted with this function and the
43 observed data. Alternative fitting functions have therefore not been explored. With five

1 parameters to fit (r , N_0 , s , A , B), this leaves 95 and 59 degrees of freedom for the
2 Tvärminne and Espegrend set-ups, respectively. The fitted surface APA_{est} has a
3 maximum $A-B$ obtained for co-ordinates combining low P with high N (large negative Z)
4 and $APA_{est}=(A/2)-B$ along the line $N=N_0+rP$ separating the P- and N-limited regions. The
5 parameter s defines the steepness of transition between the two regions perpendicular
6 to this line. B is the background APA_{est} found for high- P , low- N (large positive Z) co-
7 ordinates. The fitting was done using the “fit” function in Matlab® with its default
8 Levenberg-Marquardt algorithm providing the parameter estimates with 95%
9 confidence intervals (c.i.) (code included in SI).

10 3 Results

11 Two examples of the fitted response surface, one from Tvärminne (Fjord) (Fig. 2A) and
12 one from Espegrend (MR) (Fig. 2B) are shown to illustrate the difference in shape of the
13 response in situations apparently N-limited (Tvärminne) and P-limited (Espegrend),
14 with estimated $P^+=0.3 \mu\text{M-P}$ and $N^+=0.4 \mu\text{M-N}$, respectively. All assays are summarized
15 in Table 1.

16 For the two Espegrend samples, the change in shape of the response surface with
17 incubation time was explored (Fig. 3). For both samples, N_0 increased with incubation
18 time ($p \leq 0.05$, Table 2), i.e. the assay results drifted towards increasing N deficiency
19 when using longer incubation times. In the sample MR, r and s decreased significantly
20 over time (Table 2). Using linear regression, the parameter estimates can be
21 extrapolated back to zero incubation time. With this technique the average P^+ for the
22 Tvärminne samples, based on a single incubation time, was $0.3 \mu\text{M-P}$, and the average N^+
23 for the two Espegrend samples, based on backward extrapolation, was $0.9 \mu\text{M-N}$.

24 The assays from Tvärminne mesocosms include an $f(\text{CO}_2)$ gradient. Linear regressions of
25 N_0 ($p=0.55$), r ($p=0.63$) (Fig. 4) and s ($p=0.19$) (not presented) on $f(\text{CO}_2)$ gave no
26 indication of any statistically significant effect of the 45 days exposure of the systems to
27 different CO_2 -levels. Compared to a Redfield N:P value of 16, all the Tvärminne samples
28 gave low r (2.3 ± 0.5 ; mean over samples \pm sd), while the two Espegrend samples gave r of
29 13 ± 2 (Kviturspollen) and 32 ± 3 (MR) (mean \pm sd, both over incubation times).

30 4 Discussion

31 This study extends the demonstrated applicability of this type of assay from its previous
32 use in warm oligotrophic waters (Thingstad and Mantoura, 2005) to mesotrophic
33 temperate environments. We modified the technique so that no *a priori* assumptions are
34 now required as to whether the system investigated is N- or P-deficient. Note that the
35 function used to fit the response (Eqn. 2) was not derived from explicit assumptions on
36 biological mechanisms producing the response, but as a convenient statistical model
37 that fitted the observed responses without obvious systematic patterns in the residuals
38 (Fig. 2). It may, however, be of biological relevance to observe that, with this description,
39 the initially three-dimensional description (P , N , APA) is reduced to two dimensions (Z ,
40 APA): all combinations of P and N that have the same perpendicular distance Z (Eqn. 3)
41 to the line representing N:P balance (Eqn. 1) develop the same APA (Eqn. 2). Contour
42 plot representations of the fitted surfaces in Fig. 2 A and B would thus consist of straight
43 lines parallel to the line described by Eqn. 1.

1 We explored the use of this modified assay in two environments with anticipated
2 differences in ambient N:P stoichiometry. The Tvärminne mesocosm experiment was
3 planned with the expectation of an N-limited spring-summer situation as characteristic
4 in the Baltic Sea (Granéli et al., 1990; Rolff and Elfwing, 2015; Thomas et al., 2003),
5 subsequently transiting from N-limitation towards an N- and P-co-limited situation as
6 the result of “new” N being added through late summer blooming of diazotrophic
7 cyanobacteria (Lignell et al., 2003). This bloom did not occur during the whole
8 Tvärminne experiment and N-limitation at the time of sampling has been confirmed by
9 Nausch et al. (2016) who studied the microbial P-cycle just before our experiment.
10 Nutrient concentrations were not significantly changing throughout the whole
11 acidification experiment (Paul et al., 2015). DIN and DIP equalled $\sim 0.25 \mu\text{mol L}^{-1}$ and
12 $\sim 0.15 \mu\text{mol L}^{-1}$, respectively, giving a ratio of 1.67 (Paul et al., 2015). Compared to a
13 Redfield ratio of 16, these chemical determinations suggests N-deficiency; although not
14 taking other bioavailable forms into account. Our finding of positive N_0 -estimates for all
15 5 samples (Table 1) is in line with this. The Tvärminne assays were performed after the
16 4 days of incubation needed for the APA-responses to emerge. The conclusion of N-
17 limitation is therefore confounded by the potential drift in parameter estimates as was
18 later observed for the Espegrend samples (Fig. 3). The drift obviously complicates the
19 use of this assay since there may be no single incubation time that gives a “correct” set of
20 parameter values. Since the drift seems to be reasonably linear for all parameters (Fig.
21 3), we see it as a promising option to extrapolate the linear regressions back to time 0,
22 assuming this to give values representative for the initial conditions in the water sample.
23 In our case this gives negative N_0 values of -0.8 (-1.4,-0.2) and -1.0 (-2.4,0.4) for the
24 Espegrend samples from MR and Kviturspollen, respectively (intercept with 95% c.i.);
25 suggesting initial P-limitation. This conclusion is in accordance with our expectation
26 since the top layer of the fjords in western Norway has been shown to be P-deficient
27 (Thingstad et al., 1993).

28 The mechanisms behind the drift in parameter estimates have not been studied further
29 here. Three, not mutually exclusive, scenarios may, however, illustrate some of the
30 theoretical possibilities: 1) The microbial food web in the incubated tubes remineralizes
31 P faster than N (Garber, 1984). The assay may then correctly reflect the succession of
32 the limiting nutrient in the sense that the bioavailable pools in the tubes change over
33 time as N becomes immobilized in slowly degradable detritus to a larger extent than P.
34 2) N added in excess of P in the upper P-limited part of the P,N -plane is used by the
35 organisms to produce alkaline phosphatase (rather than biomass). This would lift the
36 response surface for high values of added N which may move the fitted line towards
37 higher N_0 , i.e. towards N-limitation. The use of extra N to produce exo-enzymes for
38 acquisition of P from DOP has recently been argued for, but then with N_2 -fixation as the
39 N source (Landolfi et al., 2015). 3) Successions in the microbial food web move towards
40 organism groups that require more N relative to P, although an increasing dominance of
41 P-rich bacteria (Fagerbakke et al., 1966) would in this scenario produce a drift in the
42 direction opposite to that observed. The r values representing the ratio of N- and P-net
43 consumption are comparable between all the Tvärminne samples (2.3 ± 0.5 , $n=5$ different
44 samples), indicating good reproducibility of the assay for similar water samples. This
45 low value compared to the Redfield value of 16 was, however, strikingly different from
46 the Espegrend samples with one Redfield-like 13 ± 2 (Kviturspollen) and one high 32 ± 3
47 (MR) value, both averaged over incubation times. A similar phenomenon was noted by
48 Thingstad and Mantoura (2005) using this method to study in-out differences in a
49 Lagrangian experiment where orthophosphate was added to the P-deficient surface

1 system in the Eastern Mediterranean. While their P-limited out-sample gave an $r=15\pm 2$,
2 the inside system, when driven to N deficiency by the *in situ* phosphate addition, gave a
3 much lower $r=3.0\pm 0.2$. Interestingly, we also here found the lower-than-Redfield r
4 values in the probably N-limited samples from Tvärminne. From the limited number of
5 assays available, the linkage between N deficiency and low r values thus seems
6 consistent. In microorganisms, C:P-ratios are usually more flexible than C:N-ratios
7 (Gismervik et al., 1996; Fagerbakke et al., 1996). P-rich microorganisms in N-deficient
8 environments may thus seem a potential explanation to the observed low r -values in N-
9 limited situations.

10 Considering the difference in sigmoidity (s) for the MR and Kviturspollen samples (Fig.
11 3) it seems that s represents a characteristic of the initial water sample. While s reflects
12 the stoichiometric flexibility in the community response, it would require further
13 investigations to determine whether this flexibility is at cell level and would be seen also
14 in axenic cultures, or is a reflection of differences between species present.

15 **5 Conclusions**

16 We have demonstrated the extension of the APA assay from its previous use in warm
17 oligotrophic, to temperate mesotrophic surface waters. The primary advantage of this
18 technique over traditional nutrient-limitation bioassays is that it indicates which of the
19 elements N or P is the most limiting, while simultaneously providing estimates of the
20 excess in bioavailable forms of the secondary limiting element (N^+ , P^+) along with the
21 ratio between net consumption of the two elements (r). The assay does not require
22 determinations of the large variety of chemical and/or physical forms in which the
23 primary and secondary limiting elements may exist. The assay was found to be
24 complicated by a drift in parameter estimates with incubation time. A backward
25 extrapolation to zero incubation time appears promising. Further work is needed to
26 fully understand the processes creating this drift and also the mechanisms that in some
27 cases generate large deviations in r from the Redfield value of 16. The consortium of
28 ecological processes that create the APA response during incubation are likely to be
29 relevant to processes shaping nutrient limitation in natural aquatic systems,. The
30 experimental setup used in this assay thus seems also to have a potential as a tool for
31 future studies on the ecological stoichiometry of aquatic microbial food webs.
32

33 **Data availability**

34 Original data are given in Supplementary Information (SI) for each assay in the form of a
35 Matlab[®] program that will also fit the response surface as shown in Fig. 2.

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34

1 Table 1: Estimates (with 95% c.i.) of the intercept (N_0) and the slope (r) of the line
 2 $N=N_0+rP$ separating the N- and P-limited regions as illustrated in Fig. 1. s represents the
 3 steepness of transition from N- to P-limitation, perpendicular to the line. R^2 -values are
 4 for the fitted response surfaces.

5
6

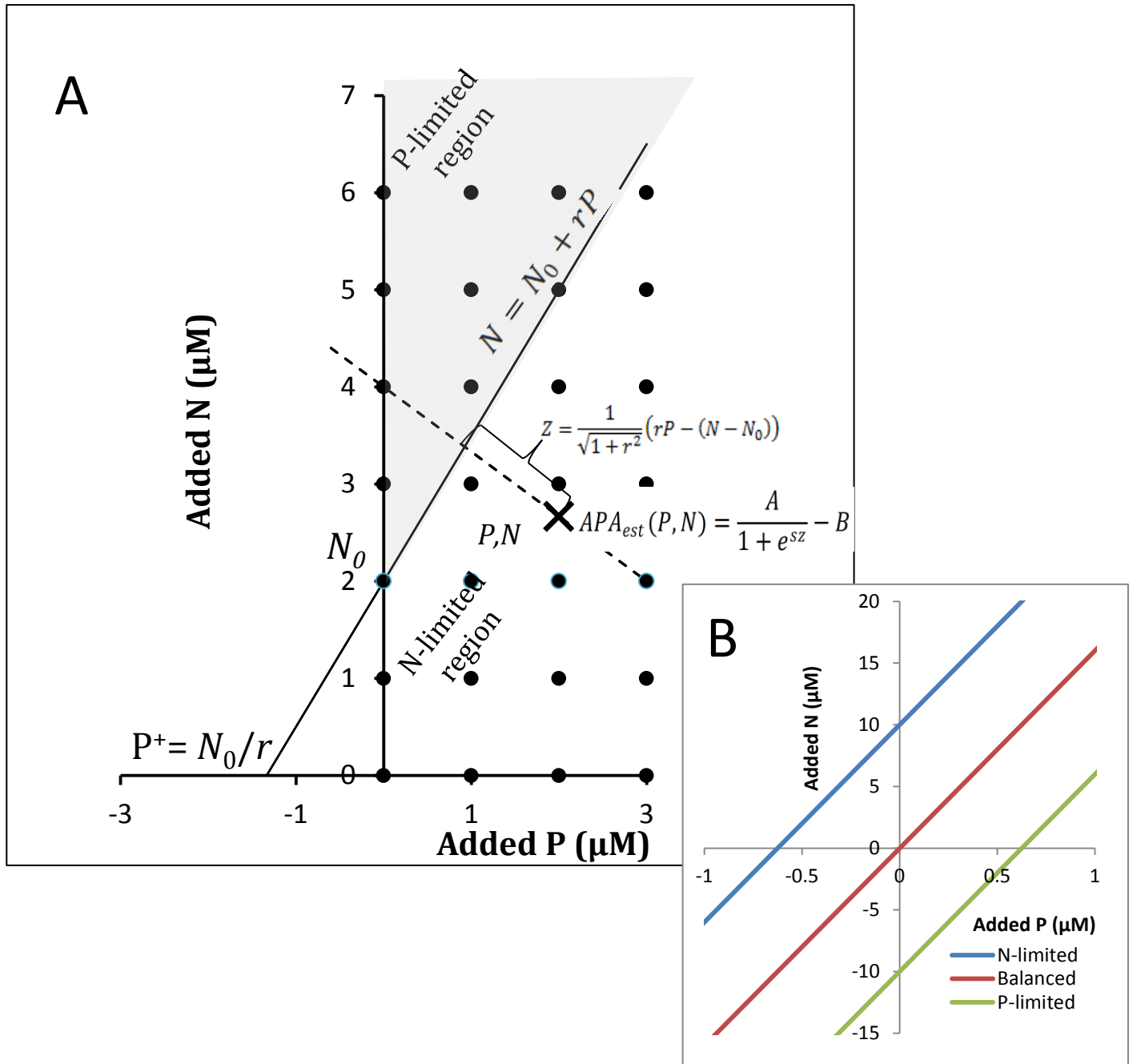
	N_0 ($\mu\text{M-N}$)	r ($\mu\text{M-N}:\mu\text{M-P}$)	s (μM^{-1})	R^2
Tvärminne (Baltic Sea): (incubation time 4 days)				
Fjord	0.79 (0.75, 0.83)	2.5 (2.2, 2.8)	29 (23, 35)	0.674
M1	0.34 (0.30, 0.38)	2.9 (2.7, 3.2)	43 (31, 56)	0.622
M3	0.76 (0.69, 0.83)	2.2 (1.9, 2.5)	19 (15, 24)	0.664
M6	0.64 (0.56, 0.72)	2.2 (1.9, 2.6)	19 (13, 25)	0.569
M7	0.70 (0.66, 0.75)	1.6 (1.4, 1.8)	20 (16, 25)	0.635
Mesocosm Raunefjorden				
Incubation time (days):				
1	-0.41 (-0.60, -0.22)	40 (36, 44)	64 (52, 76)	0.965
2	-0.63 (-1.02, -0.25)	34 (28, 40)	50 (31, 70)	0.764
3	-0.21 (-0.47, 0.06)	31 (27, 36)	48 (33, 63)	0.843
4	-0.13 (-0.34, 0.08)	31 (27, 34)	47 (35, 59)	0.933
4.5	-0.13 (-0.30, 0.04)	30 (27, 33)	51 (40, 63)	0.951
5	0.42 (0.15, 0.69)	26 (23, 29)	36 (25, 47)	0.907
Kviturspollen				
2	-0.09 (-0.26, 0.08)	9.9 (8.5, 11.4)	25 (17, 33)	0.940
3	0.58 (0.46, 0.70)	12 (11, 13)	20 (15, 25)	0.972
4	1.83 (0.72, 2.95)	14 (13, 16)	15 (8, 22)	0.946
5	1.65 (1.32, 1.99)	15 (14, 17)	25 (20, 30)	0.963
7	2.84 (0.10, 5.59)	14 (11, 16)	18 (9, 27)	0.855

7

1 Table 2. Linear regressions of parameter estimates against incubation time for the
 2 Espegrend samples (see Fig. 3). Extrapolation to zero time is given (Day 0).

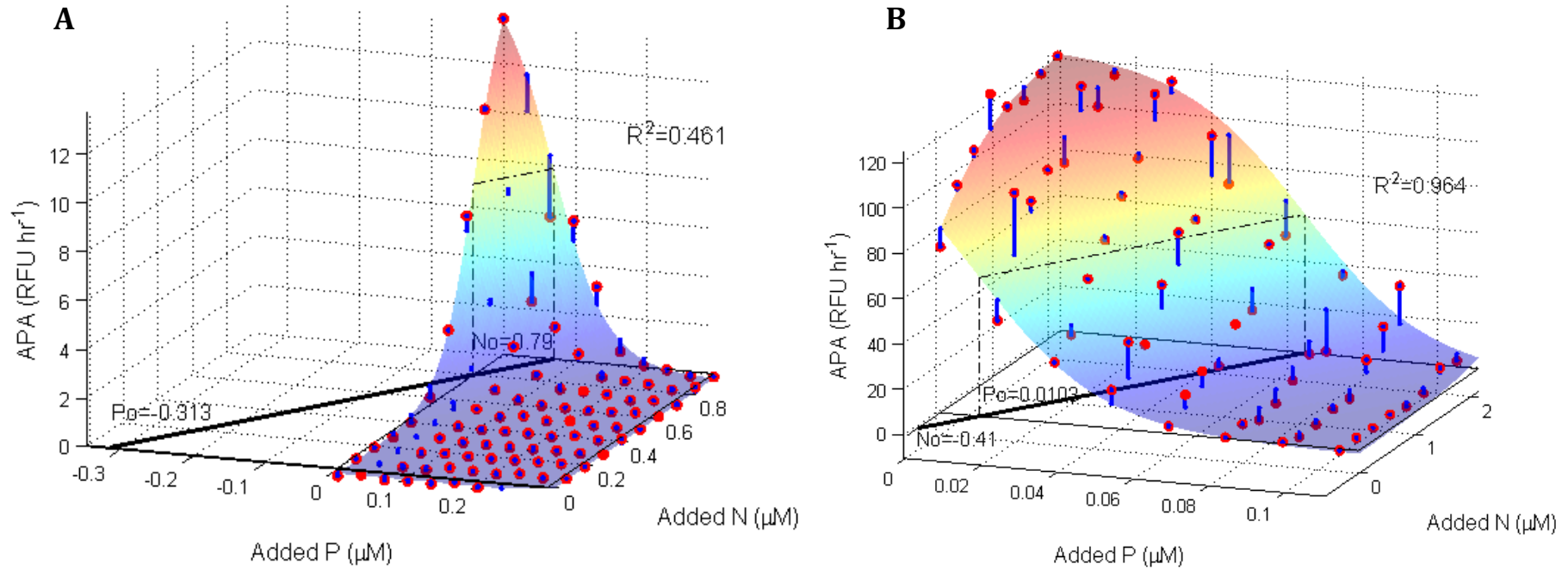
	Intercept (Day 0)	Slope of linear regression	p (H_0 : slope \neq 0)	R ²
MR				
N_0	-0.8 (-1.4,-0.2)	0.19 (0.01,0.37)	0.05	0.674
r	41 (36,46)	-2.8 (-4.1,-1.5)	0.004	0.896
s	65 (48,81)	-4.7 (-9.4,0.0)	0.05	0.656
Kviturspollen				
N_0	-1.0 (-2.4,0.4)	0.6 (0.3,0.9)	0.01	0.917
r	10 (3,16)	0.8 (-0.7,2.3)	0.2	0.479
s	24 (6,42)	-0.8 (-4.7,3.1)	0.6	0.121

3 H_0 null hypothesis



1
2 Figure 1. Illustration of the fitting algorithm used. With APA measured over a 4x7 matrix of
3 combinations in additions of P and N (black dots, Panel A), the objective is to find the line that splits
4 this P, N -plane in an upper P-limited region with high APA (shaded, Panel A) and a lower N-limited
5 region with low APA. This is done by least square fitting of the surface $APA_{est} = \frac{A}{1 + e^{sZ}} - B$ to the
6 APA-values measured in each grid point. APA_{est} is a sigmoidal function of the perpendicular
7 distance $Z = \frac{1}{\sqrt{1+r^2}}(rP + (N - N_0))$ from the point P, N (marked X in Panel A) to the line. The
8 situation illustrated in Panel A represents an N-limited system with the positive N -axis intercept
9 (N_0) and excess-P (P^+) represented by the negative P -axis intercept N_0/r . Panel B illustrates the
10 separating line for three hypothetical situations characterized by N-limitation (blue), N/P balance
11 (red) and P-limitation (green). All three with $r = 16$, but with $P^+ = 0.625 \mu\text{M}$ -P, $P^+ = N^+ = 0$, and $N =$
12 $10 \mu\text{M}$ -N for the N-limited, the balanced, and the P-limited situation, respectively.

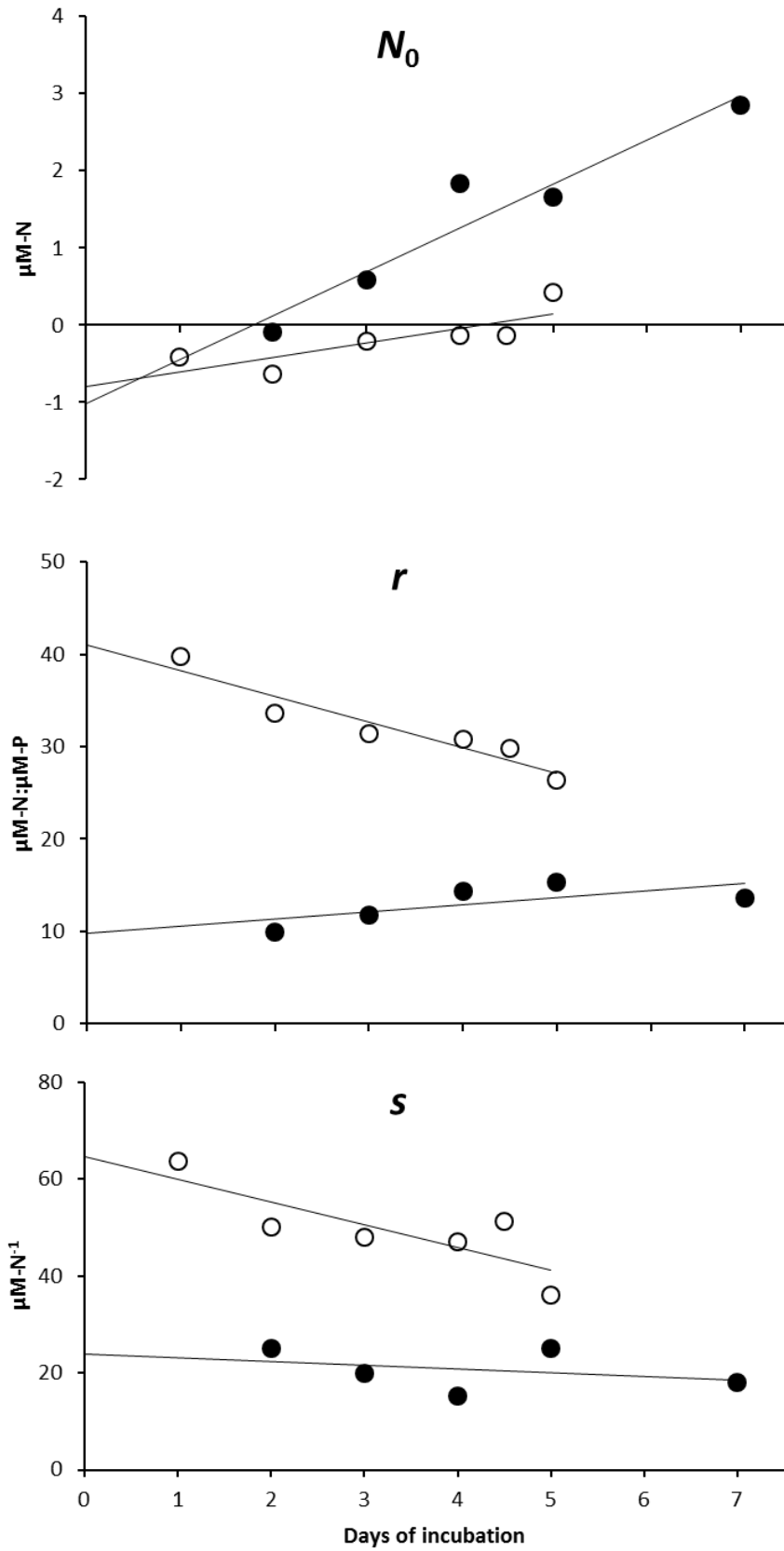
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2

3 Figure 2. Measured APA values (red), fitted response surface (blue – red gradient from N- to P-limited) and residuals (blue) for assays:
4 (A) Fjord from Tvärminne and (B) MR from Espegrend (Day 1); illustrating situations interpreted as N-limited with P⁺=0.3 μM-P and a P-
5 limited with N⁺=0.4 μM-N, respectively.

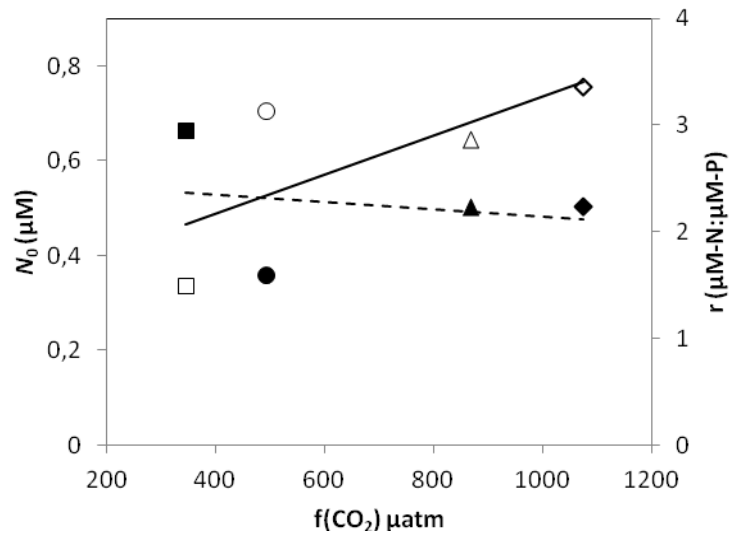
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2

3 Figure 3. Change in parameter estimates with incubation time for the two samples from western
4 Norway. Kviturspollen has filled symbols, mesocosm MR has open symbols.

5



1
 2 Figure 4. Scatterplots between $f(\text{CO}_2)$ and estimates of N_0 (open symbols, solid regression line) and
 3 r (closed symbols, dotted regression line) for Tvärminne mesocosms M1 (squares), M3 (diamonds),
 4 M6 (triangles) and M7 (circles). Regression slopes are not significant ($p=0.27$ and 0.79 ,
 5 respectively).