

Hydrogen dynamics in soil organic matter as determined by ^{13}C and ^2H labeling experiments

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Abstract: Understanding hydrogen dynamics in soil organic matter is important to predict the fate of ^3H in terrestrial environments. One way to resolve hydrogen fate and to point out processes is to examine the isotopic signature of the element in soil. However, non-exchangeable hydrogen isotopic signal in soil is complex and depends on the fate of organic compounds and microbial biosyntheses that incorporate water-derived hydrogen.

5 To decipher this complex system and to understand the close link between hydrogen and carbon cycles, we followed labeled hydrogen and labeled carbon all along natural-like soil incubations. We performed incubation experiments with three labeling conditions: 1- $^{13}\text{C}^2\text{H}$ double-labeled molecules in the presence of $^1\text{H}_2\text{O}$, 2- ^{13}C -labeled molecules in the presence of $^2\text{H}_2\text{O}$, 3- no molecule addition in the presence of $^2\text{H}_2\text{O}$. The preservation of substrate-derived hydrogen after one year of incubation (ca. 5% in most cases) was lower than the preservation of substrate-derived carbon (30% in average). We highlighted that 70% of the C-H bonds are broken during the degradation of the molecule which permits the exchange with water hydrogen. Added molecules are used more for trophic resources. The isotopic composition of the non-exchangeable hydrogen was mainly driven by the incorporation of water hydrogen during microbial biosynthesis. It is linearly correlated with the amount of carbon that is degraded in the soil. The quantitative incorporation of water hydrogen in bulk material and lipids demonstrates that non-exchangeable hydrogen exists in both organic and mineral-bound forms. The proportion of the latter depends on soil type and minerals. This experiment quantified the processes affecting the isotopic composition of non-exchangeable hydrogen, and the results can be used to predict the fate of tritium in the ecosystem or the water deuterium signature in organic matter.

20 1 Introduction

Our knowledge of the nature of soil organic matter (SOM) has made great progress in recent decades: it is now considered to be a continuum of progressively decomposing organic compounds (Lehmann and Kleber, 2015), composed of all the components of living material such as glucides, peptides, lipids, organic acids and phenolic compounds (Kelleher and Simpson, 2006). Small molecules are associated with each other in supramolecular

structures or with mineral particles by weak bonds, including H-bonds (Sutton and Sposito, 2005; Lehmann and Kleber, 2015). Their lifetime in soils is controlled more by sorption or protection than from intrinsic chemical recalcitrance to biodegradation (Schmidt et al., 2011; Basile-Doelsch et al., 2015) with the exception of pyrolytic products. Highly degradable compounds, such as glucides and peptides, contribute to the oldest SOM components, and biomarkers tend to indicate that old SOM was derived more from microbial-derived products than from plant-derived molecules as a result of the mineral protection processes (Rafter and Stout, 1970; Derrien et al., 2007; Bol et al., 2009). Carbon dynamics in this continuum have been widely studied using the natural $^{14}\text{C}/^{12}\text{C}$ and $^{13}\text{C}/^{12}\text{C}$ (Rafter and Stout, 1970; Balesdent et al., 1987) ratios and also through labeling experiments (Jenkinson, 1965; Murayama, 1988; Derrien et al., 2004). The results of these experiments have highlighted the different turnover of soil organic pools. Plant material is rapidly decomposed into microbial biomass, and a small portion of both can be protected from biodegradation for decades to centuries, representing the main part of SOM. One part of this organic matter remains stabilized for millennia, especially in deep soil horizons. It is therefore expected that the non-exchangeable hydrogen (NEH) dynamics, bound to carbon in soil, will be controlled by the same processes: organic carbon inherited from vegetation, biodegradation, microbial biosyntheses and stabilization.

Hydrogen has various molecular positions in soil. It can be organic or inorganic and non-exchangeable or exchangeable with available hydrogen. The abiotic exchange of organic hydrogen depends on the strength of the bond and the energy required for exchange (Belot, 1986; Schimmelmann, 1991; Ciffroy et al., 2006; Sauer et al., 2009). Bound to N, O and S, hydrogen is usually exchangeable with ambient water and water vapor (Schimmelmann, 1991; Wassenaar and Hobson, 2000). However, hydrogen bound to carbon is considered to be stable and non-exchangeable due to the strong covalent bonds (Baumgärtner and Donhaerl, 2004; Diabaté and Strack, 1997; Kim et al., 2013). At the ecosystem scale, H bound to C is not exchangeable (Sessions et al., 2004). Hydrogen can also interfere with clay minerals. Interlayer water exchanges with free water within a few hours and is removed after drying (Savin and Epstein, 1970). However, the structural water and the hydroxyl hydrogen of clay are non-exchangeable at room temperature (Savin and Epstein, 1970).

The natural $^2\text{H}/^1\text{H}$ ratio of plants and sediments has been used as a proxy to reconstruct past climate and paleoenvironmental conditions, such as temperature, water use efficiency (Epstein et al., 1976; Sessions et al., 2004; Zech et al., 2014; Tuthorn et al., 2015) The isotopic composition of the NEH preserves the initial composition of the plant and registers the rain isotopic composition (Sessions et al., 2004; Schimmelmann et al., 2006; Ruppenthal et al., 2010). The $\delta^2\text{H}$ of water and exchangeable hydrogen is not stable. Whereas soil organic carbon and nitrogen cycles have been extensively studied, soil organic hydrogen and its recycling in the environment remain poorly understood due to its complex behavior. The total bulk soil, composed of a mixture of non-exchangeable and exchangeable, organic and inorganic hydrogen, makes the hydrogen isotopic composition hard to determine.

The composition and exchanges between these pools can be of great importance when modeling, for instance, tritium fate in the environment. Tritium is a radioactive isotope of hydrogen which was released in large amounts in the atmosphere by nuclear weapon tests in the 60s. Since then, tritium levels have greatly declined because of its relatively short half-life (12.3 years). However the concentration of organically-bound tritium can often be higher than the concentration of water tritium due its longer residence time (Gontier and Siclet, 2011; Kim et al., 2012; Eyrolle-Boyer et al., 2014; Thompson et al., 2015). Models used to predict the fate and behavior of tritium

in the environment often simplify processes linked to the formation and degradation of organic bound tritium. Since tritium behaves as stable hydrogen in the environment, the assessment of the fate and residence time of organic tritium could be improved by quantifying the preservation of organic hydrogen from vegetation, the accumulation of hydrogen from water in the soil and the processes involved in organic matter decomposition and mineralization.

To decipher and quantify the preservation of the organic material and the microbial biosyntheses incorporating water-derived hydrogen, we designed incubation experiments with labeled compounds by assuming that the non-exchangeable hydrogen dynamics are controlled by the carbon dynamics in the soil organic matter. Three scenarios were addressed:

1- $^{13}\text{C}^2\text{H}$ -double labeled molecules in the presence of $^1\text{H}_2\text{O}$, 2- ^{13}C -labeled molecules in the presence of $^2\text{H}_2\text{O}$, 3- no molecule addition in the presence of $^2\text{H}_2\text{O}$. The ^{13}C and ^2H bulk soil isotopic compositions were analyzed at different times to quantify the processes involved. The isotopic composition of lipids was also analyzed as an indicator of organically bound NEH.

The medium-term ^{13}C and ^2H labeling experiments were conducted on different types of soil (clayey leptosol, cambisol and podzol) from 0 to 1 year to highlight and quantify the processes affecting hydrogen based on the carbon dynamics in the soil organic matter and their dependence on the soil properties.

2 Materials and Methods

2.1 Soil sampling

Three soils with contrasting physical and chemical properties were selected for this study (Table 1):

Cambisol: the 0-25 cm surface layer of a cambisol was sampled from an INRA long-term field experiment in the Parc du Château de Versailles, France in March 2014. This soil is a neutral eutric cambisol with a clay composition of 17 %, 33 % of sand, 50 % of silt and a nil carbonate content (Dignac et al., 2005). The plant cover is wheat. After each harvest, wheat residues were returned to the soil, and the first 25 cm were ploughed each fall.

Podzol: the 0-25 cm surface layer of a podzol was sampled from an INRA field experiment in Pierroton (close to Bordeaux, SW France) in May 2014. The lands cover of les Landes de Gascogne is a mixt forest dominated by *Pinus pinaster*. The sampling plot was converted into maize in 1992. The soil is a sandy hydromorphic podzol (Jolivet et al., 2006) with a clay content of less than 5 % and a sand content higher than 90 %. The first 25 cm are ploughed and crop residues are returned to the soil after each harvest.

Leptosol: the 5-10 cm surface layer of a mollic leptosol was sampled from the long-term Ecosystem Research experiment "Oak Observatory at Observatory of Haute-Provence" (O3HP), France in July 2014. The vegetation is dominated by *Quercus pubescens*. This soil is derived from limestone, compact and iron-rich with clay content (mainly smectite) of 54%.

2.2 Soil incubation

2.2.1 Soil preparation

Soil samples were air-dried at 20°C and sieved to 2 mm. Air-dry soil moisture was determined in parallel by oven-drying an aliquot at 105°C during 24h. The residual soil moisture was determined once before adding substrate and water in order to know the exact amount of water at the beginning of incubation.

110 The residual soil moisture was 5.9 % for the cambisol, 0.8 % for the podzol and 8.5 % for the leptosol. Thirty-five grams of cambisol and podzol and 30 g of mollic leptosol were transferred into 210 mL incubation jars. Each incubator was then moistened with ultra-pure water ($\delta^2\text{H} = -63.8 \pm 0.5 \text{‰}$) at 0.24 g g⁻¹ of dry soil for cambisol (6 ml of ultrapure water added) and leptosol (4 ml of ultrapure water added) and 0.10 g g⁻¹ for podzol (3 ml of ultra-pure water added) before pre-incubation at 28°C in the dark for 10 days to re-establish the biological activity to the basal level and to avoid confusion between rewetting-induced and substrate-induced activity.

2.2.2 Substrate incubations

120 Glucose, palmitic acid, phenylalanine and isoleucine were introduced separately in different incubators. They represent the most common primary compounds of the glucide, lipid and protein families found in either plant or microbial matter and contain different functional groups.

¹³C-labeled and ²H-labeled molecules and ²H₂O were provided by Euriso-Top (Cambridge Isotope Laboratories, Inc., Andover, England).

125 The isotopic abundance of each molecule was adjusted to the desired value by mixing labeled and unlabeled sources. We prepared “¹³C²H” (double-labeling) solutions and “¹³C¹H” (mono-labeling) solutions for all molecules. The incubation characteristics are shown in Table S1 (supplementary material). Mixing was performed gravimetrically.

For palmitic acid, the equivalent amount of unlabeled and labeled compound was added to 200 mg of soil and was melted at 70°C. We finely ground the cooled mixture to obtain a homogenized powder that could be added to the incubators. Two powders were prepared: a ¹³C²H-enriched powder and a ¹³C-enriched powder.

Three distinct labeling experiments were performed:

- 1) Experiment 1: “¹³C²H + H₂O”: Double-labeling molecule introduced to the soil with ultra-pure water.
- 2) Experiment 2: “¹³C¹H + ²H₂O”: Mono-labeling molecule introduced to the soil with deuterated water.
- 135 3) Experiment 3: “no molecule + ²H₂O”: Only deuterated water introduced to the soil.

The final humidity of the soil was 0.30 g g⁻¹, 0.15 g g⁻¹, and 0.31 g g⁻¹ dry weight, respectively, for cambisol, podzol and leptosol.

2.2.3 Incubations

140 The 300 incubators were incubated at 28°C in the dark, and three were frozen at 0, 7, 14, and 28 days and 1 year for the cambisol and at 0 and 7 days and 1 year for the two other soils. Jars were briefly opened (few seconds for each samples) every two days during the first three weeks and then every week until the end of incubation to keep the system under aerobic conditions. The evaporation of water is very limited during this step. However, this leads to the renewal of water vapor in the jar head space by ambient atmosphere. The isotopic composition

of ^2H of the new atmosphere in the jar is depleted compared to the previous one. By taking the value of the saturation vapor pressure at 28°C (28 g/m^3), the amount of water contained in the headspace jar (0.17 dm^3) was 4.8 mg. The proportion of the lost labeled water was estimated at 0.7 % the first months and 2 % at one year. The impact of the atmosphere renewal on the isotopic composition was therefore neglected.

150 Control incubators were prepared for each experiment at each time without any added substrate or deuterated water under the same incubation conditions.

To highlight the link between the NEH and carbon dynamics we initially added three different amounts of labeled glucose to the podzol and we analyzed the results after 7 days of incubation.

155 2.3 Lipid extraction

Lipids were extracted to isolate the organic non-exchangeable hydrogen. Lipids extractions were performed on samples that had received glucose and had been incubated for 1 year. Between 10 and 15 g of soil were sub-sampled and phosphate buffer, chloroform and methanol were added (0.8:1:2; v:v:v). After 2 min of ultrasonic and 30 min of warming (37°C), samples were centrifuged for 8 min a 2600 tr min^{-1} . Supernatant was retrieved and stored at room temperature while chloroform and methanol (1:2; v:v) were added to the remaining soil and centrifuged again. Supernatant was retrieved and added to the previous aliquot. Twenty ml of NaCl were then added to the supernatant to distinguish two phases. The denser part was collected and dried under nitrogen.

165 2.4 Isotopic measurements

Prior to analysis, incubated samples were freeze-dried during 28h. Samples were then ground to a fine powder and kept in closed tubes under laboratory atmosphere.

Twenty to fifty-five milligrams of soil were then introduced into a 10 mm tin capsule.

170 Lipid samples were solubilized in dichloromethane before introducing them to the tin capsules. We let the solvent evaporate before the analysis. The mean isotopic signature of this bulk lipid fraction was measured using the same method as for the soil samples.

The ^{13}C and ^2H contents were analyzed simultaneously with a combustion module-cavity ring-down spectroscopy (CM-CRDS) isotope analyzer (Picarro, B2221-i). The organic standards polyethylene (IAEA CH7; $\delta^{13}\text{C} = -32.15 \pm 0.05 \text{ ‰}$; $\delta^2\text{H} = -100.3 \pm 2.0 \text{ ‰}$) and oil NBS-22 ($\delta^{13}\text{C} = -30.03 \pm 0.05 \text{ ‰}$, $\delta^2\text{H} = -119.6 \pm 0.6 \text{ ‰}$) were used to calibrate the measurements. A homemade standard (olive oil) was also used in each run ($\delta^{13}\text{C} = -29.0 \pm 0.2 \text{ ‰}$, $\delta^2\text{H} = -153 \pm 5 \text{ ‰}$). To validate the measurement of highly enriched samples by CM-CRDS, we compared the measurements at initial conditions (time 0 of the incubations, before the degradation of substrate occurred), to the theoretical isotopic composition at initial conditions obtained by calculation. Both the linearity of the $\delta^2\text{H}$ measurement for enriched samples, and the full recovery of labeled NEH during the drying process were confirmed by the measured vs. theoretical $\delta^2\text{H}$ of initial labeled soil samples (mixtures of soil and labeled source before incubation), which yielded to a slope of 1.02 and $r^2 = 0.99$.

180 To deal with the ^2H memory effect often recorded with CM-CRDS (Koehler and Wassenaar, 2012), 5 repetitions were done for each sample; the last three were used for interpretation when standard deviation was less than 1.5
185 ‰ for natural samples and less than 10 ‰ for enriched samples. Moreover, we analyzed samples from the more

depleted to the more enriched and ashes were removed from the combustion tube each 45 samples to limit contamination.

The isotopic composition of ^{13}C and ^2H are expressed by abundance (A) or as δ (‰)

190 $A^{13}\text{C} = {}^{13}\text{C}/({}^{13}\text{C}+{}^{12}\text{C})$ and $A^2\text{H} = {}^2\text{H}/({}^2\text{H}+{}^1\text{H})$

$$\delta\text{‰} = [R_{\text{sample}} / R_{\text{standard}} - 1] * 1000$$

where $R = {}^{13}\text{C}/{}^{12}\text{C}$ or ${}^2\text{H}/{}^1\text{H}$. The international standard was VPDB for carbon and VSMOW for hydrogen.

2.5 Quantification of NEH derived from the labeled source

195 2.5.1. Samples equilibration

We performed equilibration of the labeled and unlabeled samples with the same atmosphere to reach the same isotopic composition of exchangeable hydrogen in the dry soils. To do so, before analysis, unlabeled and labeled samples were equilibrated with the laboratory atmosphere for 2 hours after soil grinding (exchanges also occur during the grinding \approx 20 min and during evaporation by nitrogen flushing in the CM-CRDS introduction line). The differences of $\delta^2\text{H}$ between unlabeled and labeled samples are a mean to eliminate the contribution of labile hydrogen (short-time exchange) in the final isotopic calculation (see mass calculation below). Unlabeled and labeled sample received exactly the same treatment. Hydrogen that did not exchange during the length of equilibration is considered as non-exchangeable in this study.

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2.5.2. Mass balance equations

Table 2 summarizes the different variables used in the mass balance equations and supporting information (S2) is provided for further understanding of calculations.

210 The carbon isotopic composition of the total bulk soil corresponds to the proportion of molecule-derived (labeled) and soil-derived (unlabeled) carbon (Eq. (1)).

$$C_{\text{dfm}} = ({}^{13}\text{A}_{\text{tot}} - {}^{13}\text{A}_{\text{tot}_0}) / ({}^{13}\text{A}_{\text{m}} - {}^{13}\text{A}_{\text{tot}_0}) * C_{\text{tot}} \quad (1)$$

Exchangeable hydrogen has the isotopic composition of the atmosphere when the sample is dry. Non-exchangeable hydrogen from the labeled source was estimated based on the simultaneous measurement of labeled and unlabeled samples equilibrated with the same atmosphere using the equations (2) and (3), which attributes all the excess deuterium (difference between the ${}^2\text{H}$ abundances of the labeled sample and unlabeled control) to the NEH derived from the labeled source atoms (see supplementary material S2 for the calculation).

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In Experiment 1 (labeled molecule):

$$H_{\text{dfm}} = (A_{\text{tot}} - A_{\text{tot}_0}) / (A_{\text{m}} - A_{\text{tot}_0}) * H_{\text{tot}} \quad (2)$$

220 In Experiment 2 (labeled water):

$$H_{\text{dfw}} = (A_{\text{tot}} - A_{\text{tot}_0}) / (A_{\text{w}} - A_{\text{tot}_0}) * H_{\text{tot}} \quad (3)$$

The labeled source is highly enriched compared to natural soil or water: ${}^{13}\text{A}_{\text{m}} = 6.08$ to 16.08 %, $\text{A}_{\text{m}} = 2$ to 3.5 % and $\text{A}_{\text{w}} = 0.26$ % (supplementary material, Table S1).

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The mean ${}^2\text{H}$ abundance of unlabeled soil and water is approximately 0.015 %, and the mean ${}^{13}\text{C}$ abundance is approximately 1.08 %.

Uncertainties in the element and isotope ratio measurements affect the estimate of the amount of labeled-source-derived carbon or hydrogen atoms. To assess the uncertainty in the calculated values H_{dfm} and C_{dfm} , we calculated the statistical error propagation of the uncertainties of the measured isotopic compositions and the element content of the replicated samples (Supplementary material S3).

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3 Results

The raw isotopic composition of the highly enriched samples, ($\delta^{13}C$ and δ^2H measurements) are presented over the incubation period in supplementary material (section S4).

235 In unlabeled samples δ^2H ranged from - 43 to + 27 ‰ for the cambisol, from -58 to -22 ‰ for the podzol and from -74 to - 43 ‰ for the leptosol with less than 3 ‰ variation.

In the first experiment (labeled molecule), measured δ^2H values of labeled samples at initial conditions ranged from 900 ‰ to 2500 ‰ depending on the soil and the added molecule. They still reached 85 to 576 ‰ after one year of incubation.

240 δ^2H values of the second experiment (labeled water) ranged from 0 to 260 ‰ for the cambisol, from -20 to 240 ‰ for the podzol and from 40 to 1000 ‰ for the leptosol over the incubation period.

At each time step, as described in the methodology (see section 2.5.2) the difference between the 2H isotopic composition of the labeled sample and the isotopic value of the unlabeled sample analyzed the same day is calculated. The differences between labeled and unlabeled samples are huge.

245 $\delta^{13}C$ values are between 130 and 110 ‰ at the beginning of the incubation and between 0 and 15 ‰ after one year whereas unlabeled $\delta^{13}C$ values were -26.2 ‰ for the cambisol, -25.4 ‰ for the podzol and -24.7 ‰ for the leptosol with less than 0.3 ‰ variation.

3.1 Comparison of the four substrates mineralization

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The fates of labeled C or H atoms are presented as the mass of C or NEH derived from the labeled source, i.e. molecule or water, C_{dfm} , H_{dfm} , H_{dfw} ; eq. (1), (2), (3) (dfm and dfw stand for derived from the molecule and derived from water, respectively). We first tested the dependence on time (1 week and 1 year), molecule type and soil type on the basis of a three-way ANOVA of each explained variable. Both C_{dfm} and H_{dfw} were dependent on time ($p < 0.001$) and soil ($p < 0.001$) but not on molecule. H_{dfm} was dependent on time ($p < 0.001$), soil type ($p < 0.001$) and molecule ($p < 0.001$). The differences in results for H_{dfm} can be explained by the uncertainty in experiments and measurements. Because we found no significant differences between the molecules for C_{dfm} and H_{dfw} , we considered the different molecule incubations as replicates to simplify the results presentation, with only the mean values shown in graphs. The difference in degradation between molecules is therefore contained in the error bars in fig. 1 and 2.

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3.2 Carbon mineralization

In fig.1, the results of the carbon and hydrogen derived from the molecule are expressed in percent: the amount of H_{dfm} and C_{dfm} calculated in Eq. (6) and (7) relative to the amount of H_m and C_m in the added molecule. The

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degradation of the added molecule was very fast. After 7 days, 42 %, 31 % and 53 % of molecule-derived carbon remained in the cambisol, podzol and leptosol, respectively (fig. 1). This trend is in agreement with previous studies (Murayama, 1988; Derrien et al., 2007) and illustrates the almost complete consumption of the substrate in a few days. Approximately 30 to 50 % of the consumed material was converted into microbial products, and the remaining part was used for heterotrophic respiration. During the following months, the mineralization of organic carbon continued due to the partial consumption of the newly formed microbial carbon by the soil food web.

During the incubation, non-labeled carbon (soil-derived carbon) also decreased by 1.8, 2.4 and 4.5 mg g⁻¹ within one year in the cambisol, podzol and leptosol, respectively.

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3.3 Molecule-derived non-exchangeable hydrogen

After incubation, molecule-derived NEH (H_{dfm}) was considerably lower than molecule-derived carbon for the cambisol and for the podzol (fig. 1): expressed in % of the initial labeled NEH, it was, 12 % and 5 % compared to the 42 % and 31 % recorded for carbon. It is important to note the different fate of the leptosol, where the yield of transfer of NEH reached 55 % after 7 days of incubation (fig. 1). During the following months, H_{dfm} slightly decreased in the three soils (by approximately 6 % \pm 5 after one year).

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3.4 Incorporation of water hydrogen

Experiment 2 (molecule $^{13}C^1H + ^2H_2O$) highlights the incorporation of water hydrogen in the non-exchangeable pool of soil. For the three soils, the incorporation of hydrogen from water increased during the first seven days and continued to slowly increase during the incubation year (fig. 2). Respectively, 0.013 ± 0.001 , 0.008 ± 0.002 and 0.33 ± 0.07 mg g⁻¹ of hydrogen derived from water was found after 7 days of incubation, and 0.06 ± 0.03 , 0.023 ± 0.004 , and 0.845 ± 0.003 mg g⁻¹ was found after one year of incubation for cambisol, podzol and leptosol (fig. 2).

Figure 3 shows the difference in the incorporation of water hydrogen with and without added substrate. The incorporation of water-derived hydrogen was higher when associated with substrate addition. It was twice as high for the podzol after 7 days of incubation. Figure 4 illustrates that this enhancement of incorporation of water hydrogen was linearly dependent on the amount of the substrate added to the soil.

3.5 Isotopic composition of lipids

Carbon and hydrogen isotopic compositions of bulk lipids at 365 days for the control soil are presented in Table 3. The proportions of labeled carbon and hydrogen were calculated as the proportion of the total lipid carbon and hydrogen ($(C_{dfm}/C)_{lipids}$, $(H_{dfm}/H)_{lipids}$, and $(H_{dfw}/H)_{lipids}$) and were compared to the proportion in the bulk soil ($(C_{dfm}/C)_{bulk}$, $(H_{dfm}/H)_{bulk}$, $(H_{dfw}/H)_{bulk}$). The $\delta^{13}C$ and δ^2H of the lipids in the control samples were lower than that of the bulk soil, in agreement with previous work, where the lipid $\delta^{13}C$ was 2-3 ‰ lower than the bulk $\delta^{13}C$ (Chikaraishi and Naraoka, 2001; Hayes, 2001), and the lipid δ^2H was 150‰ lower than the bulk δ^2H (Sessions et

305 al., 1999; Chikaraishi and Naraoka, 2001). The average measured H/C ratio of the lipids of the three soils was 2.1 (molar ratio).

The proportion of molecule-derived carbon in the lipids was 1.0, 0.4 and 0.8 % for the cambisol, podzol and leptosol, respectively, compared to the corresponding values of 0.6, 0.4, and 0.4 % in the bulk organic carbon. The proportion of molecule-derived hydrogen was, respectively 0.10, 0.02 and 0.19 % for the cambisol, podzol
310 and leptosol (Table 3). These values were on the same order of magnitude as the molecule-derived hydrogen in the bulk soil. The proportion of labeled water-derived hydrogen was 1.0, 0.4 and 1.1 % of the total hydrogen content in the lipids and 1.5, 0.9 and 6.8 % in the total bulk soil for the cambisol, podzol and leptosol, respectively.

315 4 Discussion

4.1 Preservation of the organic substrate hydrogen in biosyntheses

The microbial activity is initiated during the first days after the addition of the substrate. The added molecule
320 regardless of its quality is quickly metabolized (fig. 1).

We independently traced the preservation of organic hydrogen (experiment 1) and the incorporation of water-derived hydrogen (experiment 2) during decomposition and biosynthesis. The conservation of organic hydrogen from the initial substrate is very low in both the total and lipid NEH. The carbon-hydrogen bonds are broken during decomposition, and exchange with exchangeable hydrogen can occur. The difference between carbon and
325 hydrogen isotopic fates during the first seven days (fig. 1) reflects the exchange of hydrogen with water during the early stage of degradation. Subsequently, new organic exchangeable hydrogen derived from water can be incorporated into the non-exchangeable pool of organic matter by biological processes. Furthermore, because mineralization of a substrate also results in $^2\text{H}_2\text{O}$ release, one part of the soil organic non-exchangeable ^2H may originate from incorporation of this substrate-derived deuterated water into the non-exchangeable pool. Using the
330 assumption that water is a well-mixed isotopic compartment, this amount is between 3 and 7 % of the residual hydrogen from the organic substrate at 365 days (see supplementary materials for calculation). The isotopic composition of the non-exchangeable organic hydrogen is mainly determined by the water isotopic composition (fig. 2). It is in accordance with the work of Baillif and colleagues that have grown fungi with labeled glucose, water or acetate to trace the incorporation of ^2H during fatty acid biosynthesis. They have demonstrated that
335 water is the main donor of hydrogen atoms in the non-exchangeable pool within the biosynthesis cycle (Baillif et al., 2009). Ruppenthal and colleagues have shown as well that precipitation contributes to 80 % of the isotopic composition of non-exchangeable hydrogen (Ruppenthal et al., 2010). Moreover, the incorporation of water hydrogen is favored by the strength of the C-H bond breakage. It can be weak or strong, depending on the enzyme activated for the degradation of the molecule and the position of the bond (Augusti et al., 2006). When
340 C-H breakdown is favored, the surrounding water imprints its hydrogen isotopic signature on the former bounded-H (Augusti et al., 2006). In the present experiment, we show that more than 70 % of the H-C bonds of the initial molecule are broken in the first biodegradation steps (7 days, fig. 1); therefore, the added molecules are used more for energetic and trophic resources than as building blocks in the biosynthesis.

345 Deuterium can also accumulate in hydration shells, which have stronger hydrogen bridges than the biomolecule (Baumgärtner and Donhaerl, 2004). The accumulation of deuterium from water occurs in the biomatter during biological processes but also during the hydration of molecules (Baumgärtner and Donhaerl, 2004; Turner et al., 2009).

4.2 Carbon-driven acquisition of the non-exchangeable hydrogen isotope signature

350 The rapid mineralization of hydrogen from the molecule is due to biodegradation whereas the rapid incorporation of water during the first seven days of incubation is associated with biosynthesis (figs. 1 and 2). Results obtained on the lipid fraction provide evidence of the formation of organic non-exchangeable hydrogen from H_{dfw} (table 3). Carbon mineralization fosters the formation of non-exchangeable hydrogen from water (figs. 3 and 4). There is also an incorporation of hydrogen from water in the soil in the experiments without substrate. This could be associated to the microbial transformation of carbon already present in soil (fig. 3) but also to the inorganic fraction that could incorporate hydrogen from water. In fig. 2, exchange of hydrogen with water seems to be continuous: H_{dfw} continues increasing, following carbon mineralization. The incorporation of water hydrogen in the organic non-exchangeable pool but also in the inorganic fraction is occurring during the length of the incubation.

360 Results of lipids isotopic compositions show that the amount of newly formed NEH (% of H_{dfw} + % of H_{dfm} in lipids; Table 3) is slightly higher than expected from the theoretical organic C-H bond (% of C_{dfm} in lipids; Table 3). This could be due to complete, stoichiometric labeling of newly biosynthesized lipids, i.e., lipids formed on the labeled organic carbon plus a smaller amount of newly synthesized lipids from unlabeled organic matter. The proportion of molecule-derived carbon is higher in the lipid than in the bulk soil, and the lipids are derived mainly from microbial biosynthesis. However, the proportions of H_{dfm} in the lipids and in the bulk soil are of the same magnitude for the three soils (Table 3): hydrogen is derived evenly from the labeled molecule and from the unlabeled soil during lipids biosynthesis.

370 The proportion of hydrogen derived from the water in lipids is lower than the respective proportion in the bulk soil, which means the proportion of H_{dfw} is not necessarily organic. It could be inorganic hydroxyl, hydrated ions or water in different states. The difference is even higher in the clayey soil (leptosol, Table 3) that contains the highest proportion of hydrogen (table 1).

375 In lipids, hydrogen corresponds to the organic, non-exchangeable hydrogen. The hydrogen in lipids formed from water (H_{dfw}) is only organic, whereas in the bulk soil, the H_{dfw} is organic and inorganic. To estimate the proportion of organic non-exchangeable hydrogen, we assume that H_{dfw}/C_{dfm} in lipids (0.20 on average for the three soils based on mass ratio, Table 3) is approximately the same as the H_{dfw}/C_{dfm} in the organic fraction of the bulk soil. Using the measured C_{dfm} in bulk soil, we can then estimate the total organic H_{dfw} as 0.014, 0.015 and 0.028 $mg\ g^{-1}$, respectively, for the cambisol, podzol and leptosol. The proportion of inorganic H_{dfw} is therefore 0.046, 0.008 and 0.82 $mg\ g^{-1}$ for the three soils. The NEH isotopic composition is mainly controlled by the incorporation of water through biosynthesis, but the inorganic H_{dfw} is not negligible, especially in the clayey soil.

4.3 Hydrogen dynamics in different soil types.

385 The association of organic matter with minerals is known to decrease the decomposition rate of the former (Feng
et al., 2013; Jenkinson and Coleman, 2008; Vogel et al., 2014). This result is observed in our experiment by
comparing the three soils with increasing clay content and is applicable to both H and C in both bulk soil (fig. 1)
and lipids (Table 3). However, the clay content has an important role in the incorporation of water-derived
hydrogen beyond this organic matter stabilization effect. In clayey leptosol, the amount of labeled NEH from the
molecule (H_{dfm} ; fig. 1) is much higher than in the other soils, which may be explained by the preferential use of
390 hydrogen locally near biological reactions. Hydrogen derived from the mineralization of the substrate does not
directly exchange with the total pool of water but with a smaller pool. The resulting local water pool has a less
negative isotopic signature than the remaining water pool. Water incorporation through biosynthesis could then
occur with this smaller pool of 2H -enriched water. Moreover, hydrogen exchange within the whole water pool is
slowed by the presence of clay, which accumulates molecule-derived hydrogen in hydroxyl sites. For this reason,
395 H_{dfw} is also much higher in leptosol than in the other soils. The non-exchangeable hydrogen pool considered in
the present equilibration method contains H in various positions and it may include water molecule acting as
non-exchangeable hydrogen. The inorganic H_{dfw} pool is accounting for less than 1/40 of the amount of water-H
in the leptosol. This pool appears as very slowly exchangeable with “free water”. In situ, on a short-term
dynamics scale, this pool acts as non-exchangeable and is mostly at the hydroxyl position (López-Galindo et al.,
400 2008).

The zonal distribution of organic compounds associated with minerals (Kleber et al., 2007; Vogel et al., 2014)
may control the exchange between soil solution and organic compounds at kinetics that differ according to the
layer within the organo-mineral interaction zone. The non-exchangeable hydrogen dynamics in soil organic
matter are not independent of the mineral structure. The type of clay plays a role in carbon sequestration,
405 depending on the specific surface area of the mineral or aggregate (Vogel et al., 2014; López-Galindo et al.,
2008). In leptosol, clays are mainly smectite and have a high specific surface area. The high content of iron and
hydroxide present in the leptosol also increases the specific surface area of the aggregates, which increases the
organo-mineral association (Baldock and Skjemstad, 2000). Organic carbon cycling itself may be associated with
mineral transformation (Basile-Doelsch et al., 2015), which may involve the newly formed hydroxyl.

410 The short-term dynamics of hydrogen are driven by the incorporation of hydrogen from water by isotopic
exchange and by microbial biosynthesis. However, the increase in the incorporation of water hydrogen with the
soil clay content suggests that part of the hydrogen is bound to clay or organo-mineral complexes. The
production of NEH from water occurs mainly during the first weeks, but slow exchange of water hydrogen
continues during the following year. Lopez-Galindo *et al.* observed the same trend, and they related the
415 accumulation rate to the clay mineral properties (López-Galindo et al., 2008).

4.4 Ecosystem-scale production and fate of non-exchangeable hydrogen

420 In the present experiment, the preservation of non-exchangeable hydrogen from an organic substrate is less than
5 % after one year in soil with a low clay content. Water is the main donor of hydrogen during biosynthesis cycle
favored by the breakage of the C-H bonds of the initial substrate. In this work, we showed that 70 % of the C-H
bonds of the initial substrate were broken during biosynthesis. Concerning the fate of tritium in terrestrial
ecosystems, the isotopic composition of the organic plant material is a minor determinant of the bulk soil organic

matter composition. However, a better preservation of hydrogen from vegetation could arise from higher soil
425 clay content and the subsequent organo-mineral and zonal interactions. Water will be the main donor of
organically bound tritium in the soil and the incorporation will be dependent to the carbon mineralization. Lipid
isotopic composition highlighted that i- water derived hydrogen is involved in biosynthesis and ii- the newly
formed non-exchangeable hydrogen is not necessary organic. Therefore, the incorporation of tritium from water
in NEH pool is dependent on the clay content and on the soil hydrodynamics.

430 In our work, the isotopic composition of the NEH pool is determined by comparing labeled samples with
unlabeled samples equilibrated under the same atmosphere. This method includes inorganic NEH. A proportion
of the inorganic, non-exchangeable hydrogen should be taken into account in the prediction of the dynamics of
hydrogen and tritium.

The main finding of the work is that the long-term fate of hydrogen in terrestrial environments and by extension
435 the fate of tritium will depend on the status of soil carbon dynamics.

The proportion of NEH associated with minerals is itself partially related to the carbon dynamics.

Both the carbon dynamics and the incorporation of inorganic hydrogen in soils should therefore be taken into
account in a conceptual model for the prediction of the long-term fate of hydrogen, and thereafter of tritium, in
soil organic matter. The results of the present study can be used for the parameterization of the carbon-hydrogen
440 coupling in such prediction models.

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450 Supplementary material includes the summary of incubation characteristics (Table S1), mass balance calculation
(S2), propagation error calculation (S3) and $\delta^{13}\text{C}$ and $\delta^2\text{H}$ results of the incubation samples (S4)

References

- Augusti, A., Betson, T. R. and Sclucher, J.: Hydrogen exchange during cellulose synthesis distinguishes climatic and biochemical isotope fractionations in tree rings., *New Phytol.*, 172, 490–499, 2006.
- 455 Baillif, V., Robins, R. J., Le Feunteun, S., Lesot, P. and Billault, I.: Investigation of Fatty Acid Elongation and Desaturation Steps in *Fusarium lateritium* by Quantitative Two-dimensional Deuterium NMR Spectroscopy in Chiral Oriented Media, *J. Biol. Chem.*, 284(16), 10783–10792, doi:10.1074/jbc.M807826200, 2009.
- Baldock, J. A. and Skjemstad, J. O.: Role of the soil matrix and minerals in protecting natural organic materials against biological attack, *Org. Geochem.*, 31(7–8), 697–710, doi:10.1016/S0146-6380(00)00049-8, 2000.
- 460 Balesdent, J., Mariotti, A. and Guillet, B.: Natural ^{13}C abundance as a tracer for studies of soil organic matter dynamics, *Soil Biol. Biochem.*, 19(1), 25–30, 1987.
- Basile-Doelsch, I., Balesdent, J. and Rose, J.: Are Interactions between Organic Compounds and Nanoscale Weathering Minerals the Key Drivers of Carbon Storage in Soils?, *Environ. Sci. Technol.*, 49(7), 3997–3998, doi:10.1021/acs.est.5b00650, 2015.
- 465 Baumgärtner, F. and Donhaerl, W.: Non-exchangeable organically bound tritium (OBT): its real nature, *Anal. Bioanal. Chem.*, 379, 204–209, 2004.
- Belot, Y.: Tritium in plants: A review, *Radiat. Prot. Dosimetry*, 16(1–2), 101–105, 1986.
- Bol, R., Poirier, N., Balesdent, J. and Gleixner, G.: Molecular turnover time of soil organic matter in particle-size fractions of an arable soil, *Rapid Commun. Mass Spectrom.*, 23, 2551–2558, 2009.
- 470 Chikaraishi, Y. and Naraoka, H.: Organic hydrogen-carbon isotope signatures of terrestrial higher plants during biosynthesis for distinctive photosynthetic pathways., *Geochem. J.*, 35(6), 451–458, doi:10.2343/geochemj.35.451, 2001.
- Ciffroy, P., Siclet, F., Damois, C. and Luck, M.: A dynamic model for assessing radiological consequences of tritium routinely released in rivers. Application to the Loire River, *J. Environ. Radioact.*, 90, 110–139, 2006.
- 475 Derrien, D., Marol, C. and Balesdent, J.: The dynamics of neutral sugars in the rhizosphere of wheat. An approach by ^{13}C pulse-labelling and GC/C/IRMS, *Plant Soil*, 267(1–2), 243–253, doi:10.1007/s11104-005-5348-8, 2004.
- Derrien, D., Marol, C. and Balesdent, J.: Microbial biosyntheses of individual neutral sugars among sets substrates and soils, *Geoderma*, 139, 190–198, 2007.
- 480 Diabaté, S. and Strack, S.: Organically Bound Tritium in Wheat after Short-Term Exposure to Atmospheric Tritium under Laboratory Conditions, *J. Environ. Radioact.*, 36(2–3), 157–175, 1997.
- Dignac, M. ., Bahri, H., Rumpel, C., Rasse, D. . and Bardoux, G.: Carbon-13 natural abundance as a tool to study the dynamics of lignin monomers in soil: an appraisal at the Closeaux experimental field (France), *Geoderma*, 128, 3–17, 2005.
- 485 Epstein, S., Yapp, C. J. and Hall, J. H.: The determination of the D/H ratio of non-exchangeable hydrogen in cellulose extracted from aquatic and land plants, *Earth Planet. Sci. Lett.*, 30, 241–251, 1976.
- Eyrolle-Boyer, F., Boyer, P., Claval, D., Charmasson, S. and Cossonnet, C.: Apparent enrichment of organically bound tritium in rivers explained by the heritage of our past, *J. Environ. Radioact.*, 136, 162–168, doi:10.1016/j.jenvrad.2014.05.019, 2014.
- 490 Feng, W., Plante, A. F. and Six, J.: Improving estimates of maximal organic carbon stabilization by fine soil particles, *Biogeochemistry*, 112(1–3), 81–93, doi:10.1007/s10533-011-9679-7, 2013.
- Gontier, G. and Siclet, F.: Le tritium organique dans les écosystèmes d'eau douce : évolution à long terme dans l'environnement des centres nucléaires de production d'électricité français, *Radioprotection*, 46(4), 457–491,

- doi:10.1051/radiopro/2011139, 2011.
- 495 Hayes, J. M.: Fractionation of Carbon and Hydrogen Isotopes in Biosynthetic Processes, *Rev. Mineral. Geochem.*, 43(1), 225–277, doi:10.2138/gsrng.43.1.225, 2001.
- Jenkinson, D. S.: Studies on the decomposition of plant material in soil. I. Losses of carbon from ¹⁴C labeled ryegrass incubated with soil in the field, *J. Soil Sci.*, 16, 104–115, 1965.
- 500 Jenkinson, D. S. and Coleman, K.: The turnover of organic carbon in subsoils. Part 2. Modelling carbon turnover, *Eur. J. Soil Sci.*, 59, 400–413, 2008.
- Jolivet, C., Angers, D. A., Chantigny, M. H., Andreux, F. and Arrouays, D.: Carbohydrate dynamics in particle-size fractions of sandy spodosols following forest conversion to maize cropping, *Soil Biol. Biochem.*, 38(9), 2834–2842, doi:10.1016/j.soilbio.2006.04.039, 2006.
- 505 Kelleher, B. P. and Simpson, A. J.: Humic Substances in Soils: Are They Really Chemically Distinct?, *Environ. Sci. Technol.*, 40(15), 4605–4611, doi:10.1021/es0608085, 2006.
- Kim, S. B., Bredlaw, M. and Korolevych, V. Y.: HTO and OBT activity concentrations in soil at the historical atmospheric HT release site (Chalk River Laboratories), *J. Environ. Radioact.*, 103(1), 34–40, doi:10.1016/j.jenvrad.2011.08.013, 2012.
- 510 Kim, S. B., Baglan, N. and Davis, P. : Current understanding of organically bound tritium (OBT) in the environment, *J. Environ. Radioact.*, 126, 83–91, 2013.
- Kleber, M., Sollins, P. and Sutton, R.: A conceptual model of organo-mineral interactions in soils: self-assembly of organic molecular fragments into zonal structures on mineral surfaces, *Biogeochemistry*, 85(1), 9–24, doi:10.1007/s10533-007-9103-5, 2007.
- 515 Koehler, G. and Wassenaar, L. I.: Determination of the Hydrogen Isotopic Compositions of Organic Materials and Hydrated Minerals Using Thermal Combustion Laser Spectroscopy, *Anal. Chem.*, 84(8), 3640–3645, doi:10.1021/ac3000489, 2012.
- Lehmann, J. and Kleber, M.: The contentious nature of soil organic matter, *Nature*, 528(7580), 60–68, doi:10.1038/nature16069, 2015.
- 520 López-Galindo, A., Hach-Alí, P. F., Pushkarev, A. V., Lytovchenko, A. S., Baker, J. H. and Pushkarova, R. A.: Tritium redistribution between water and clay minerals, *Appl. Clay Sci.*, 39(3–4), 151–159, doi:10.1016/j.clay.2007.06.005, 2008.
- Murayama, S.: Microbial synthesis of saccharides in soils incubated with ¹³C-labelled glucose, *Soil Biol. Biochem.*, 20(2), 193–199, 1988.
- 525 Rafter, T. A. and Stout, J. D.: Radiocarbon measurements as an index of the rate of turnover of organic matter in forest and grassland ecosystems in New Zealand., in *Radiocarbon Variations and Absolute Chronology*, pp. 401–418, Stockholm., 1970.
- Ruppenthal, M., Oelmann, Y. and Wilcke, W.: Isotope ratios of nonexchangeable hydrogen in soils from different climate zones, *Geoderma*, 155, 231–241, 2010.
- 530 Sauer, P. E., Schimmelmann, A., Sessions, A. L. and Topalov, K.: Simplified batch equilibration for D/H determination of non-exchangeable hydrogen in solid organic material, *Rapid Commun. Mass Spectrom.*, 23(7), 949–956, doi:10.1002/rcm.3954, 2009.
- Savin, S. and Epstein, S.: The oxygen and hydrogen isotope geochemistry of clay minerals, *Geochim. Cosmochim.*, 94, 25–42, 1970.
- 535 Schimmelmann, A.: Determination of the concentration and stable Isotopic composition of nonexchangeable hydrogen in organic matter, *Anal. Chem.*, 63(21), 2456–2459, 1991.

- Schimmelmann, A., Sessions, A. L. and Mastalerz, M.: Hydrogen Isotopic (d/H) Composition of Organic Matter During Diagenesis and Thermal Maturation, *Annu. Rev. Earth Planet. Sci.*, 34(1), 501–533, doi:10.1146/annurev.earth.34.031405.125011, 2006.
- 540 Schmidt, M. W. I., Torn, M. S., Abiven, S., Dittmar, T., Guggenberger, G., Janssens, I. A., Kleber, M., Kögel-Knabner, I., Lehmann, J., Manning, D. A. C., Nannipieri, P., Rasse, D. P., Weiner, S. and Trumbore, S. E.: Persistence of soil organic matter as an ecosystem property, *Nature*, 478(7367), 49–56, doi:10.1038/nature10386, 2011.
- Sessions, A. L., Burgoyne, T. W., Schimmelmann, A. and Hayes, J. M.: Fractionation of hydrogen isotopes in lipid biosynthesis, *Org. Geochem.*, 30(9), 1193–1200, doi:10.1016/S0146-6380(99)00094-7, 1999.
- 545 Sessions, A. L., Sylva, S. P., Summons, R. E. and Hayes, J. M.: Isotopic exchange of carbon-bound hydrogen over geologic timescales, *Geochim. Cosmochim.*, 68(7), 1545–1559, 2004.
- Sutton, R. and Sposito, G.: Molecular structure in soil humic substances: the new view, *Environ. Sci. Technol.*, 39(23), 9009–9015, 2005.
- 550 Thompson, P. A., Kwamena, N.-O. A., Ilin, M., Wilk, M. and Clark, I. D.: Levels of tritium in soils and vegetation near Canadian nuclear facilities releasing tritium to the atmosphere: implications for environmental models, *J. Environ. Radioact.*, 140, 105–113, doi:10.1016/j.jenvrad.2014.11.009, 2015.
- Turner, A., Millward, G. E. and Stemp, M.: Distribution of tritium in estuarine waters: the role of organic matter, *J. Environ. Radioact.*, 100(10), 890–895, doi:10.1016/j.jenvrad.2009.06.018, 2009.
- 555 Tuthorn, M., Zech, R., Ruppenthal, M., Oelmann, Y., Kahmen, A., del Valle, H. F., Eglinton, T., Rozanski, K. and Zech, M.: Coupling $\delta^{2}\text{H}$ and $\delta^{18}\text{O}$ biomarker results yields information on relative humidity and isotopic composition of precipitation – a climate transect validation study, *Biogeosciences*, 12(12), 3913–3924, doi:10.5194/bg-12-3913-2015, 2015.
- 560 Vogel, C., Mueller, C. W., Höschen, C., Buegger, F., Heister, K., Schulz, S., Schloter, M. and Kögel-Knabner, I.: Submicron structures provide preferential spots for carbon and nitrogen sequestration in soils, *Nat. Commun.*, 5, 2947, doi:10.1038/ncomms3947, 2014.
- Wassenaar, L. I. and Hobson, K. A.: Improved Method for Determining the Stable-Hydrogen Isotopic Composition (dD) of Complex Organic Materials of Environmental Interest, *Environ. Sci. Technol.*, 34, 2354–2360, 2000.
- 565 Zech, M., Zech, R., Rozanski, K., Hemp, A., Gleixner, G. and Zech, W.: Revisiting Mt. Kilimanjaro: Do n-alkane biomarkers in soils reflect the $\delta^{2}\text{H}$ isotopic composition of precipitation?, *Biogeosciences Discuss.*, 11(6), 7823–7852, doi:10.5194/bgd-11-7823-2014, 2014.

570 Tables and Figures

Table 1: Pedologic, geographic information and carbon and hydrogen isotopic composition of the three bulk soils.

Soil	Latitude, longitude	Sampling depth (cm)	plant cover	% clay	% silt	% sand	C mg g ⁻¹	H mg g ⁻¹	N mg g ⁻¹	pH	Mean bulk $\delta^{13}\text{C}$ (‰)	Mean bulk $\delta^2\text{H}$ (‰)
Cambisol	48° 48' 0.2 N; 2° 06' 33 W	0-25	wheat	17	50	33	12	3.2	1.27	6.8	-26 ± 0	-53±3
Podzol	44° 44' 33 N; 0° 47' 37 W	0-25	maize	5	3	92	20	2.3	0.84	5.5	-26 ± 0	-51±3
Leptosol	43° 56' 06 N; 5° 42' 06 W	5-10	oak forest	54	37	9	39	13.7	3.59	7.5	-25 ± 0	-52±2

575 Table 2: definition of the variable used in calculations

	Variables	Definition
Quantity (mg g ⁻¹ of dry soil)	H _{tot}	Total amount of hydrogen in the soil
	C _{tot}	Total amount of carbon in the soil
	H _{dfw}	Amount of non-exchangeable hydrogen derived from water
	H _{dfm}	Amount of non-exchangeable hydrogen derived from molecule
	C _{dfm}	Amount of carbon derived from molecule
	H _m	Initial amount of non-exchangeable hydrogen in the added molecule
	C _m	Initial amount of carbon in the added molecule
Abundance	¹³ A _{tot_0}	¹³ C abundance of the unlabeled experiment (control)
	A _{tot_0}	² H abundance of the unlabeled experiment (control)
	¹³ A _{tot}	¹³ C abundance of the total bulk soil
	A _{tot}	² H abundance of the total bulk soil
	¹³ A _m	Initial ¹³ C abundance of the labeled molecule
	A _m	Initial ² H abundance of the labeled molecule
	A _w	Initial ² H abundance of the labeled water

580 Table 3: $\delta^{13}\text{C}$, $\delta^2\text{H}$ of bulk soil and lipids at 365 days of incubation for cambisol, podzol and leptosol and the proportion of carbon and hydrogen derived from the labeled-source. In brackets the concentration in mg g^{-1} of carbon and hydrogen derived from the labeled-source

	$\delta^{13}\text{C}$ ‰		$\delta^2\text{H}$ ‰		% and concentration (mg g^{-1}) of the labelled source in the bulk soil			% and concentration (mg g^{-1}) of the labelled source in lipids		
	Bulk	lipids	bulk	lipids	% Cdfm (mg g^{-1})	% Hdfm (mg g^{-1})	% Hdfw (mg g^{-1})	% Cdfm (mg g^{-1})	% Hdfm (mg g^{-1})	% Hdfw (mg g^{-1})
Cambisol										
Experiment 1	5.3 ± 0.9	19 ± 3	179 ± 2	133 ± 43	0.6 % (0.054 ± 0.009)	0.082 % (0.0031 ± 0.0003)	-	1 % (3.3 ± 0.6)	0.1 % (0.13 ± 0.01)	-
Experiment 2	5 ± 1	17 ± 11	422 ± 15	36 ± 6	0.6 % (0.055 ± 0.009)	-	1.5 % (0.06 ± 0.03)	1 % (4.2 ± 0.8)	-	1 % (0.68 ± 0.09)
Podzol										
Experiment 1	13.2 ± 0.7	7 ± 5	94 ± 39	-106 ± 42	0.4 % (0.07 ± 0.01)	0.07 % (0.0017 ± 0.0007)	-	0.4 % (1.8 ± 0.9)	0.02 % (0.05 ± 0.02)	-
Experiment 2	7.0 ± 0.7	6 ± 7	259 ± 25	-38 ± 55	0.4 % (0.07 ± 0.01)	-	0.9 % (0.023 ± 0.004)	0.4 % (1.4 ± 0.1)	-	0.4 % (0.4 ± 0.2)
Leptosol										
Experiment 1	22 ± 3	97 ± 2	723 ± 10	458 ± 65	0.4 % (0.15 ± 0.03)	0.24 % (0.029 ± 0.003)	-	0.8 % (5.9 ± 0.2)	0.19 % (0.18±0.01)	-
Experiment 2	25 ± 1	76 ± 20	798 ± 58	23 ± 5	0.4 % (0.14 ± 0.03)	-	6.8 % (0.8 ± 0.1)	0.8 % (4.8 ± 0.8)	-	1.1 % (0.95 ± 0.03)

585

List of figures:

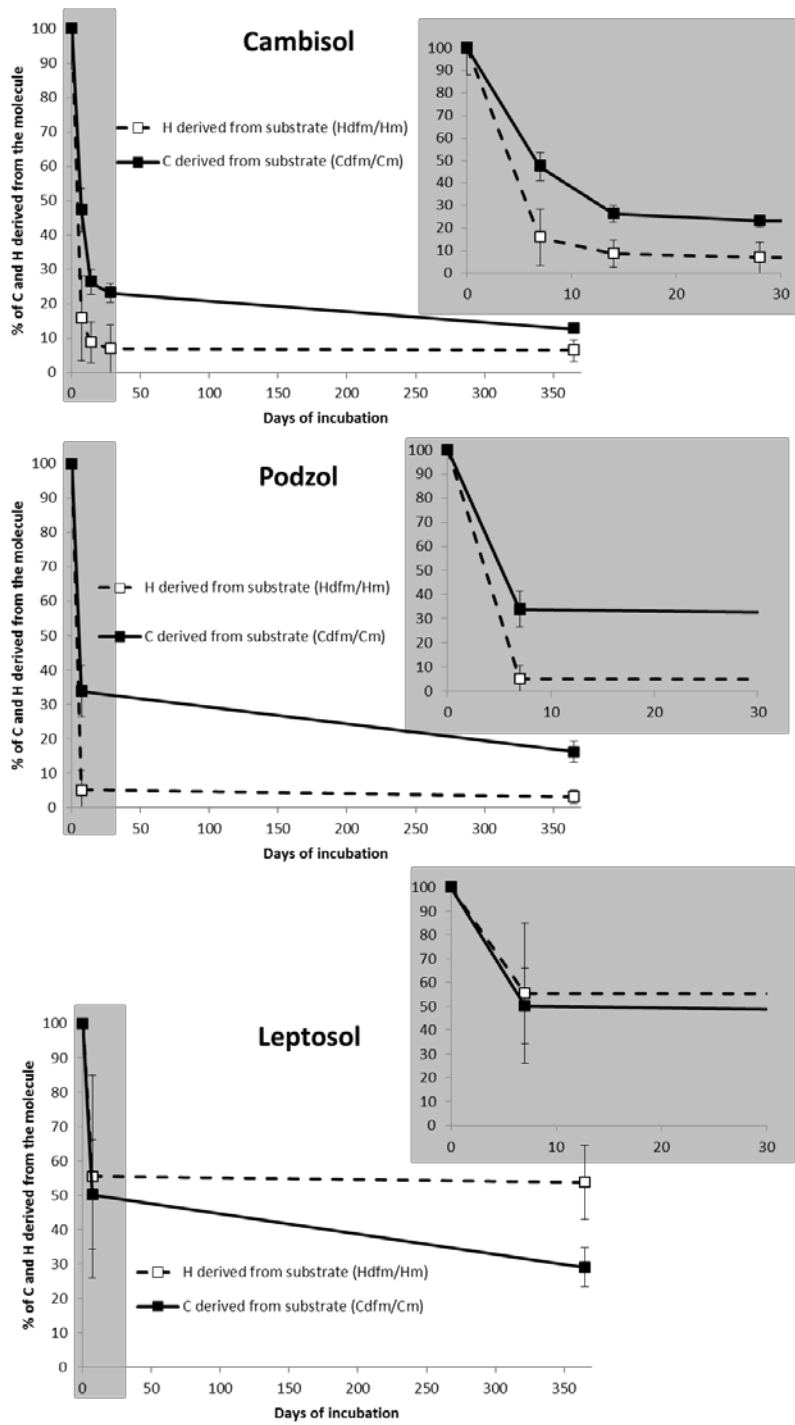
590 Figure 1: Percentage of non-exchangeable hydrogen and carbon derived from the added molecule during one year of incubation for cambisol, podzol and mollic leptosol. The grey part corresponds to the results from 0 to 28 days. The line corresponds to the mean value calculated at each time for all molecules experiments.

595 Figure 2: Concentration of non-exchangeable hydrogen derived from the water H_{dfw} and derived from the molecule H_{dfm} for cambisol, podzol and mollic leptosol from 0 to 365 days. The line corresponds to the mean value calculated at each time for all molecules experiments. Note that the scale of Y axis is different for the three soils.

600 Figure 3: Concentration of non-exchangeable hydrogen derived from water with and without addition of substrate for the cambisol, podzol and leptosol from 0 to 28 days of incubation. Note that the scale of Y axis is different for the three soils.

Figure 4: Amount of substrate-derived H and C at 7 days of incubation versus the concentration of substrate as carbon (0.14, 0.29 and 0.43 mg g^{-1}). The concentration C in experiment 1 is equal to 0.43 mg g^{-1} .

Fig. 1



605

Fig. 2

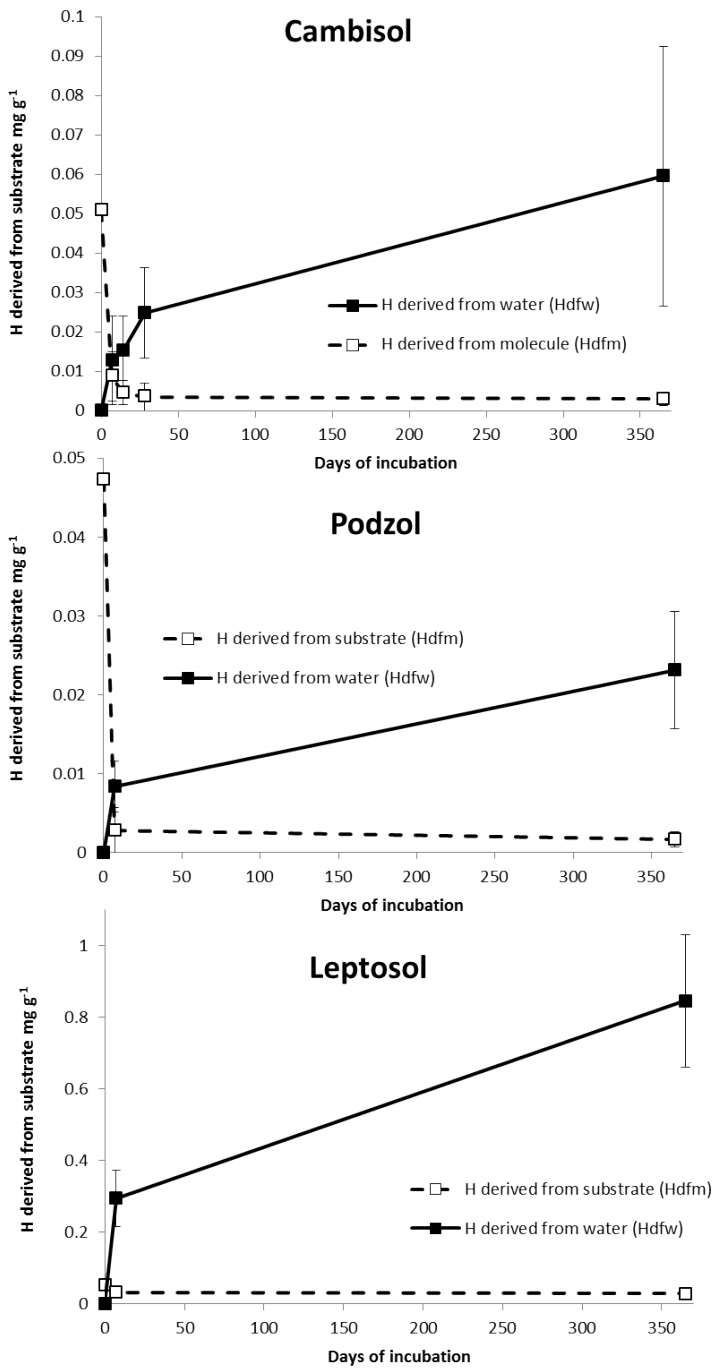


Fig.3

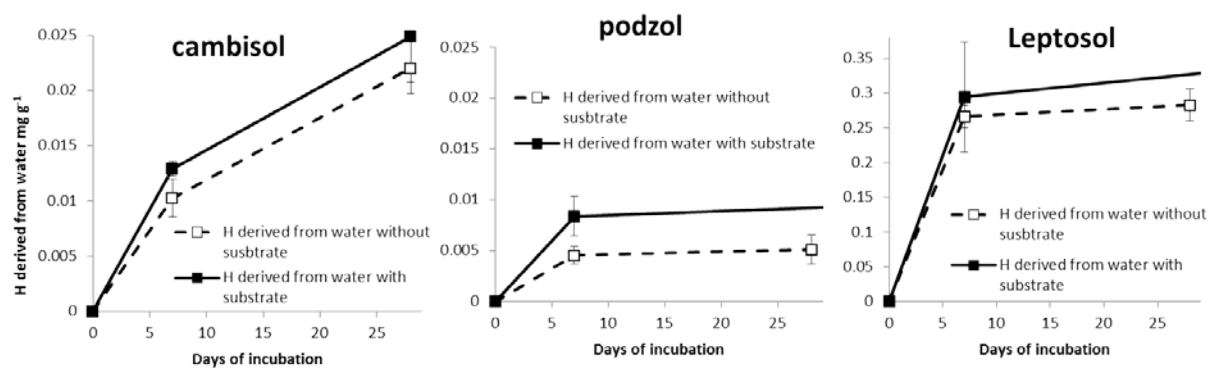


Fig.4

