

**Spatial variability of organic matter molecular composition and elemental geochemistry
in surface sediments of a small boreal Swedish lake**

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Abstract.

The composition of sediment organic matter (OM) exerts a strong control on biogeochemical processes in lakes, such as those involved in the fate of carbon, nutrients and trace metals. While between-lake spatial variability of OM quality is increasingly investigated, we explored in this study how the molecular composition of sediment OM varies spatially within a single lake, and related this variability to physical parameters and elemental geochemistry. Surface sediment samples (0-10 cm) from 42 locations in Härsvatten – a small, boreal forest lake with a complex basin morphometry – were analyzed for OM molecular composition using pyrolysis gas chromatography/mass spectrometry, and for the contents of twenty-three major/trace elements and biogenic silica. 162 organic compounds belonging to different biochemical classes of OM (e.g., carbohydrates, lignin and lipids) were identified. Close relationships were found between the spatial patterns of sediment OM molecular composition and elemental geochemistry. Differences in the source types of OM (i.e. terrestrial, aquatic plant and algal OM) were linked to the individual basin morphometries and chemical status of the lake. The variability in OM molecular composition was further driven by the degradation status of these different source pools, which appeared to be related to sedimentary physico-chemical parameters (e.g., redox conditions) and to the molecular structure of the organic compounds. Given the high spatial variation in OM molecular composition within Härsvatten and its close relationship with elemental geochemistry, the potential for large spatial variability across lakes should be considered when studying biogeochemical processes involved in the cycling of carbon, nutrients and trace elements or when assessing lake budgets.

Keywords.

Lake sediment; spatial variability; organic matter; molecular composition; Py-GC/MS; elemental geochemistry

1. Introduction

In lake basins, a wide range of factors are known to influence the transport and fate of sedimentary material, such as the location of inlet streams, catchment topography, land-use patterns, fetch, basin morphometry and sediment focusing. Sediment focusing results from a combination of factors such as wind and wave action, basin slope and the settling velocity of different particle sizes, which all contribute to the redistribution of light, fine-grained material rich in clays, organic matter (OM) and associated trace elements from shallower to deeper waters (Blais and Kalff, 1995; Ostrovsky and Yacobi, 1999). While sediment focusing is important, catchment and lake characteristics can be complex and exert a primary influence on spatial patterns in sediment geochemistry, such as in relation to land use in near-shore areas (Dunn et al., 2008; Vogel et al., 2010; Sarkar et al., 2014), complex lake/basin morphometries (Bindler et al., 2001; Rydberg et al., 2012) or river inflows (Kumke et al., 2005). The presence of macrophytes or wind-induced water currents have also been shown to affect the spatial distribution of e.g., lead (Pb), phosphorus (P) and OM (Benoy and Kalff, 1999; Bindler et al. 2001).

Because trace metals and nutrients are primarily associated with – or are part of – OM, studies focusing on the spatial patterns of metal or nutrient accumulation typically include an analysis of the OM content. The two standard approaches to determine sediment OM content are the analysis of loss-on-ignition (LOI; Ball, 1964; Santisteban et al., 2004) or the analysis of elemental carbon (C). However, either approach inherently treats OM as a homogeneous sediment component. Recent studies interested in the role of lake sediments as a long-term C sink have likewise mainly treated OM and C as a homogeneous component (e.g., Sobek et al., 2003; Tranvik et al., 2009; Heathcote et al., 2015). Even if this approach is rational from a global perspective of calculating C budgets, treating OM as a homogeneous component is overly simplistic from the perspective of developing insights into the biogeochemical behavior of OM and its influence on C, nutrients and

52 trace metals cycling, and does not take full advantage of the information provided by differences in
53 the OM quality.

54 In boreal lakes the sediment composition is often dominated by OM, typically ranging from 20
55 to 60 % on a dry weight basis. Biogenic silica (bSi) may account for as much as 45 % of the
56 sediment dry weight (Meyer-Jacob et al., 2014), and the remaining sediment mainly consists of
57 detrital mineral matter and possibly authigenic minerals. Lake OM is an extremely heterogeneous
58 and complex mixture of molecules that are derived from residues of plants, animals, fungi, algae
59 and micro-organisms, and which are either transported into the lake from the surrounding
60 catchment (allochthonous) or produced within the lake (autochthonous). Furthermore, these organic
61 compounds may undergo transformations within the water column and the sediment through both
62 biotic and abiotic processes. There have been a few studies where the spatial complexity in OM
63 quality within a lake basin has been assessed using infrared spectroscopy, which yields qualitative
64 information on variations in OM quality (Korsman et al., 1999; Rydberg et al., 2012), or
65 quantitative analyses of photopigments and lipids (Ostrovsky and Yacobi, 1999; Trolle et al., 2009;
66 Vogel et al., 2010; Sarkar et al., 2014). However, little work has been done to detail how the
67 molecular composition of the sediment OM matrix varies spatially within a lake, considering a
68 large number of organic biochemical classes and compounds.

69 To characterize OM composition at the molecular level, the most commonly used methods are
70 based on liquid or gas chromatography (LC or GC) coupled to fluorescence or mass spectrometry
71 (MS) detection. These methods provide quantitative data on original organic compounds found in
72 the analyzed samples, including highly specific biomarkers of, e.g., OM sources, and have been
73 successfully employed to study OM composition and reactivity in environmental matrices as well
74 as to reconstruct environmental changes (e.g., changes in vegetation, algal productivity) from peat
75 or sediment cores. However, the associated sample preparation procedures, i.e.
76 extraction/hydrolysis and derivatization, are fastidious and specific to the different biochemical
77 classes of organic compounds such as carbohydrates, proteins/amino acids, lipids, chlorophylls and

lignin (e.g., Wakeham et al., 1997; Dauwe and Middelburg, 1998; Tesi et al., 2012). Moreover, sample masses > 10 mg are required. Hence, studies where different OM biochemical classes are targeted using these wet chemical extraction and GC/LC-MS methods are very scarce. But, efforts in characterizing the whole OM composition at the molecular level can bring important insights because the different biochemical classes of OM do not always include specific biomarkers for the different existing sources of OM (e.g., terrestrial plants, macrophytes, higher plants, mosses, algae, bacteria). For example, lignin oligomers are only specific of higher plants (Meyer and Ishiwatari, 1997) and proteins/amino-acids mainly provide biomarkers for bacteria and planktonic production (Bianchi and Canuel, 2011). Moreover, the different biochemical classes of OM do not present the same reactivity; for example, proteins/amino-acids and neutral carbohydrates have been shown to be among the most reactive organic molecules (e.g., Fichez, 1991; Dauwe and Middelburg, 1998; Amon and Fitznar, 2001; Tesi et al., 2012). Advanced ultrahigh-resolution MS techniques, i.e. Fourier transformed-ion cyclotron resonance-mass spectrometry (FT-ICR-MS) or Linear trap quadrupole-orbitrap-MS, enable the determination of a large number of organic molecular formulas in liquid samples (> 1000; e.g., Hawkes et al., 2016). These method have been successfully used to link variability in the molecular composition of dissolved OM (DOM) with different factors and/or processes of environmental ecosystems, such as climate, hydrology and OM degradation in boreal lakes (Kellerman et al., 2014; Kellerman et al., 2015) or optical properties and DOM photochemical alterations in wetland and seawater (Stubbins and Dittmar, 2015; Wagner et al., 2015). However, in addition to the limited access to these advanced MS techniques due to instrumental costs, extraction/hydrolysis steps are required when studying solid samples, which make these methods also specific to the different biochemical classes of organic compounds.

To study the variability of OM composition in sediments, pyrolysis-gas chromatography/mass spectrometry (Py-GC/MS) is a good compromise between (i) the quantitative LC/GC-MS or the high-resolution MS methods that target specific compounds and, (ii) the qualitative, non-molecular information provided by high-throughput techniques such as infrared spectroscopy or 'RockEval'

104 pyrolysis. Py-GC/MS analysis requires no complex sample preparation, but yields semi-quantitative
105 data on >100 organic compounds that are chemical fingerprints of the different OM biochemical
106 classes, which include specific biomarkers for OM sources and degradation status (Faix et al., 1990;
107 Faix et al., 1991; Peulvé et al., 1996; Nierop and Buurman, 1998; Schulten and Gleixner, 1999;
108 Lehtonen et al., 2000; Nguyen et al., 2003; Page, 2003; Buurman et al., 2005; Fabbri et al., 2005;
109 Kaal et al., 2007; Vancampenhout et al., 2008; Schellekens et al., 2009; Carr et al., 2010; Buurman
110 and Roscoe, 2011; De La Rosa et al., 2011; McClymont et al., 2011; Micić et al., 2011; Stewart,
111 2012).

112 In the present study, we apply our newly optimized Py-GC/MS method to characterize the
113 molecular composition of natural OM in surface sediments (0-10 cm) from 42 locations within the
114 lake basin of Härsvatten. Härsvatten is a small boreal forest lake in southwestern Sweden that was
115 previously studied for the spatial distribution of Pb and OM contents (Bindler et al., 2001). Our
116 objective here was to comprehensively investigate how the molecular composition of sediment OM
117 varies spatially across a lake with several basins. Our specific research questions were: (i) what are
118 the spatial patterns within a single lake for various organic biochemical classes and compounds?;
119 (ii) how does the spatial pattern of the OM molecular composition relate to physical parameters
120 (i.e., bulk density and water depth) and elemental, inorganic geochemistry of the sediment
121 material?; (iii) which factors or processes (e.g., provenance, transport pathway and mineralization)
122 appear to explain the in-lake spatial variability of the OM molecular composition?

123 2. Materials and Methods

124

125 2.1 Study site and samples

126

127 Härsvatten is a boreal forest lake located in southwestern Sweden (58°02' N 12°03' E) in the
128 Svartedalen nature reserve. This culturally acidified, clear-water, oligotrophic and fishless lake has
129 been intensively monitored since the 1980's (national database, Dept. of aquatic sciences and
130 assessment, Swedish university of agricultural sciences, Uppsala, Sweden; www.slu.se), during
131 which time the pH has ranged from 4.2-4.5 in 1983–1987 to 4.7-5.6 in 2010–2014. The lake is
132 dimictic with a thermal stratification between 10 and 15 m depth in the summer. Approximately 80
133 % of the lake bottom is within the epilimnion. The surface areas of the lake and its catchment are
134 0.186 and 2.03 km², respectively. The catchment is characterized by an uninhabited, coniferous-
135 dominated forest (*Picea abies* Karst. and *Pinus sylvestris* L.), which extends to the rocky shoreline.
136 The bedrock consists of slow-weathering granites and gneisses that are covered by thin and poorly
137 developed podsollic soils.

138 The basin of Härsvatten can be divided into four general areas (Bindler et al., 2001): 1) the
139 main south basin, which represents about half of the lake area (sample sites S1–24; maximum
140 depth, 24.3 m) and includes the lake's small outlet stream; 2) a north basin (sample sites N1–11;
141 maximum depth, 12 m), which includes a small inlet stream draining from the headwater lake
142 Måkevatten that enters Härsvatten through a small wetland; 3) an east basin, which has a maximum
143 depth of nearly 10 m (sample sites E1–6) and is separated from the main north–south axis of the
144 lake by a series of islands and shallow sills (<3 m water depth); and 4) a generally shallow (<3 m
145 water depth) central area separating the north/east and south basins (sample sites M1–6).

146 In total, we analyzed 44 surface sediment (0–10 cm) samples that were collected in winter
147 1997–1998 (Fig. 1) for a study of Pb and spheroidal carbonaceous particles (Bindler et al., 2001).
148 These samples were collected as follow: short sediment cores (0-25 cm) were taken with a gravity

149 corer from the ice-covered lake in winter 1997 and 1998, and were sectioned on-site into an upper
150 sample (0-10 cm) and a lower sample (10-25 cm; not studied here). In the laboratory, the samples
151 were weighed, freeze-dried, and reweighed to determine the water content and dry mass of the
152 sediment. The freeze-dried samples have been stored in plastic containers within closed boxes
153 shielded from light and at room temperature since winter 1997-1998. Before further analysis in this
154 study, the samples were finely ground at 30 Hz for 3 min using a stainless steel “Retsch” swing
155 mill.

156

157 2.2 Major and trace elements concentrations

158

159 The concentrations of major (Na, Mg, Al, Si, K, Ca, P, S, Mn, Fe) and trace elements (Sc, Ti,
160 V, As, Br, Y, Zr, Ni, Cu, Zn, Sr, Pb) were determined using a wavelength dispersive X-ray
161 fluorescence spectrometer (WD-XRF; Bruker S8 Tiger) and a measurement method developed for
162 powdered sediment samples (Rydberg, 2014). Accuracy was assessed using sample replicates,
163 which were within ± 10 % for all elements.

164 Total mercury (Hg) concentrations were determined using thermal desorption atomic
165 absorption spectrometry (Milestone DMA80) with the calibration curves based on analyses of
166 different masses of four certified reference materials (CRMs). Analytical quality was controlled
167 using an additional CRM and replicate samples included with about every ten samples. The CRM
168 was within the certified range, and replicate samples were within ± 10 % for Hg concentrations < 30
169 $\mu\text{g kg}^{-1}$ and within ± 5 % for concentrations $\geq 30 \mu\text{g kg}^{-1}$.

170 We also included the OM content (in % dry mass), determined as loss-on-ignition (LOI) after
171 heating dried samples at 550°C for 4 h in the earlier study of Bindler et al. (2001).

172 2.3 Biogenic silica concentrations

173

174 Biogenic silica (bSi) was determined by Fourier transform infrared (FTIR) spectroscopy
175 following the approach described in Meyer-Jacob et al. (2014). In brief, sediment samples were
176 mixed with potassium bromide (0.011 g sample and 0.5 g KBr) prior to analysis with a Bruker
177 Vertex 70 equipped with a HTS-XT accessory unit (multisampler). The recorded FTIR spectral
178 information were used to determine the bSi concentrations employing a PLSR calibration based on
179 analyses of synthetic sediment mixtures with defined bSi content ranging from 0 to 100 %.

180 We calculated the mineral Si fraction ($\text{Si}_{\text{mineral}}$) from the difference between the total Si
181 concentration determined by WD-XRF (Sect. 2.2) and the bSi concentration.

182

183 2.4 Organic matter molecular composition

184

185 The molecular composition of OM was determined by pyrolysis-gas chromatography-mass
186 spectrometry (Py-GC/MS) following the method developed by Tolu et al. (2015). In brief, $200 \pm$
187 $10 \mu\text{g}$ sediment was pyrolyzed in a FrontierLabs PY-2020iD oven (450°C) connected to an Agilent
188 7890A-5975C GC/MS system. Peak integration was done using a data processing pipeline under
189 the 'R' computational environment. Peak identification was made using the software 'NIST MS
190 Search 2' containing the library 'NIST/EPA/NIH 2011' and additional spectra from published
191 studies.

192 In the sediments of Härsvatten, 162 Py-products were identified, and peak areas were
193 normalized by setting the total identified peak area for each sample to 100 %. A detailed list of the
194 162 identified organic compounds is provided in the supplementary information along with
195 information on their molecular mass and structure, references for the theoretical mass spectra and
196 calculated or reference retention index values (Table S1). Although the pyrolysis temperature we
197 employed, i.e. 450°C as used in plant science (e.g., Faix et al., 1990; Faix et al., 1991), is different

198 from the pyrolysis temperature which was most commonly used for analyzing soils, sediments, peat
199 records and algae (i.e. $>600^{\circ}\text{C}$), our list is highly similar to published lists of identified pyrolytic
200 organic compounds both in terms of the organic compounds and of their classification into 13 OM
201 classes (Faix et al., 1990; Faix et al., 1991; Peulvé et al., 1996; Nierop and Buurman, 1998;
202 Schulten and Gleixner, 1999; Lehtonen et al., 2000; Nguyen et al., 2003; Page, 2003; Buurman et
203 al., 2005; Fabbri et al., 2005; Kaal et al., 2007; Vancampenhout et al., 2008; Schellekens et al., 2009
204 ; Carr et al., 2010; Buurman and Roscoe, 2011; De La Rosa et al., 2011 ; McClymont et al., 2011;
205 Micić et al., 2011; Stewart, 2012). Pyrolysis at 450°C is preferred to pyrolysis at 650°C when
206 using very small sample mass (few hundred μg) because it avoids complete degradation of some
207 specific biomarkers of OM sources and enables determination of OM degradation status by
208 identification of levosugars (Py products of polysaccharides and/or cellulose) or syringol lignin
209 oligomers, for instance (Tolu et al., 2015).

210

211 2.5 Statistical analysis

212

213 We performed all statistical analyses using SPSS software package PASW, version 22.0. Two
214 separate principal component analyses (PCA) were performed, one for the elemental geochemistry
215 (i.e., dry bulk density (B.D.) and contents of OM (LOI), major/trace elements and bSi) and the
216 other for the OM molecular composition. Prior to the PCA, all data were converted to Z-scores
217 (average = 0, variance = 1). Principal components (PCs) with eigenvalues > 1 were retained and
218 extracted using a Varimax rotated solution. Factor loadings were calculated as regression
219 coefficients, which is analogous to r in Pearson correlations. For convenience the loadings are
220 reported as percentage of variance explained, i.e., as squared loadings. For all PCs, variables with
221 squared loadings < 0.15 are not discussed with respect to that PC. Others variables, e.g., water depth
222 (W.D.) or ratios between elements, were included passively in the PC-loadings plots by using bi-
223 variate correlation coefficients between these variables and the PC-scores of each PC. Hierarchical

224 agglomerative cluster analysis (CA) was performed for the elemental geochemistry and the OM
225 molecular composition datasets using Wards linkages (Ward, 1963) based on squared Euclidean
226 distances. The PC-scores from the PCAs were used instead of the original data in order to eliminate
227 the effects of autocorrelation in the dataset.

228

229 **3. Results and discussion**

230

231 3.1. Sediment elemental geochemistry

232

233 *3.1.1 General description and trends*

234

235 Summary statistics of the elemental geochemical properties of the surface sediments from
236 Härsvatten are presented in Table 1 and the detailed data are given in Table S2 in the SI. The
237 sediments from sites M4 and S15 are two outliers because they have a B.D., bSi, OM and elemental
238 contents (e.g., Na, Mg, Al, K) that deviated by more than four standard deviations from the average
239 values of all analyzed sediment samples (Table 1). Moreover, these sediment samples are too coarse
240 (predominantly sand) for Py-GC/MS analysis according to our method based on $200 \pm 20 \mu\text{g}$
241 analyzed sample mass. Hence, they are excluded from the statistical analyses and discussion. Even
242 when excluding these two sites, the elemental geochemical parameters vary considerably across the
243 lake basin, with Hg, Fe, Co and Mn contents illustrating the greatest variabilities (i.e., coefficient of
244 variation, CVs >60 %) and Al, Br, K, Ti, V, Ni, Mg and Ca contents showing the lowest
245 variabilities (CVs: 17-25 %; Table 1). For most of geochemical properties, the average to median
246 ratios are approximatively 1.0, indicating no extreme values. Slightly higher values were, however,
247 observed for P, Fe, As and Co contents (1.2-1.3), and Mn content is associated with extremely large
248 values outside the population distribution (average:median = 4.1).

249 The lowest B.D. is observed among the three deepest sampling locations (23.5-24.5 m) in the
250 main south basin, where we also find the lowest bSi content and the highest contents in organically
251 bound elements including S, Br, P and certain trace metals, i.e. Cu, Ni, Hg and Zn. These sediments
252 have high OM content (> 50 %), but the highest [OM] (57-58 %) are observed among isolated sites
253 that are located close to the shoreline (N1-2, E3, S5, S23; 3.1-7.4) and which also include the
254 lowest [Al], [P], [K], [Si_{inorganic}], [V] and [Zr]. The highest B.D. and the lowest [OM], [S], [Br],
255 [Cu], [Ni], [Hg] and [Zn] are observed among the shallow sites (1.8-2.5 m) located between the
256 north and east basins and between the larger north and south basins (i.e., sites N10, M1, M5-6),
257 which also contain the highest [bSi], [Sr], [Al], [Y], [Mn] and [Co]. The sediments located at
258 intermediate water depth (9-20 m) in the main south basin (S4, S9, S11, S13-14, S17, S19, S22) are
259 associated with the highest [Fe], [As], [K], [Mg], [Na], [Ti] and [Zr], while among the shallower
260 sites of the south basin we find the highest [Si_{inorganic}]. The lowest [Fe], [As], [Co] and [Y] are
261 observed among the sediments of the east basin, and the sediments of the north basin include the
262 lowest [Mn], [Ca], [K], [Mg], [Na], [Sr] and [Zr]. To identify more accurately the most significant
263 relationships between the different elemental geochemical properties and to explore more precisely
264 their spatial distribution, the results of PCA and cluster analyses are further presented and
265 discussed.

266

267 3.1.2. Principal components of the elemental geochemistry

268

269 For the elemental geochemistry dataset, five principal components were retained. We present
270 only the first four PCs, which together explain 74 % of the total variance (PC1-4_{geo}; Fig. 2), because
271 no reasonable interpretation could be made for PC5_{geo} (10 % of the total variance; Fig. S1 in the
272 SI). PC1_{geo} captures 25 % of the total variance and separates bSi and B.D. (negative loadings) from
273 OM, S, Cu, Hg, Ni, Zn and, to a lesser extent, As and Pb (positive loadings; Fig.2a). This means
274 that bSi and B.D. are significantly positively correlated, and both are significantly negatively

275 correlated to OM, S, Cu, Hg, Ni, Zn and, to a lesser extent, As and Pb. If those parameters do not
276 have significant loadings on PC2-5, it means they are not significantly correlated with the
277 parameters found on PC2-5, the PCs being orthogonal to each other. The negative loadings on PC1
278 are interpreted as reflecting a bSi-rich fraction, while positive loadings indicate an organic-rich
279 fraction that is enriched in organophilic trace metals (Lidman et al., 2014). For PC2_{geo}, which
280 captures 21 % of the total variance, Si_{inorganic}, K, Na, Mg, Zr and Ti have positive loadings, while no
281 element is significantly negatively correlated to PC2_{geo} (Fig. 2a). High PC2_{geo} scores likely
282 represent samples that are richer in silicate minerals such as quartz and clays (Koinig et al., 2003;
283 Taboada et al., 2006).

284 Positive loadings on PC3_{geo}, which explains 16 % of the total variance, are found for Al and Fe
285 along with As, P and Y (Fig. 2b). Compared to elements such as Mg, Na and K that are mostly
286 confined to the silicate fraction of sediments, Fe and Al may reflect both detrital material and
287 dissolved or amorphous phases. However, the fact that As and P contents as well as the Fe:Al ratio
288 plot together with Fe and Al contents on the positive side of PC3_{geo} and not with the S content
289 strongly suggest that sediments with high PC3_{geo} scores are associated with higher contents of Fe
290 and Al (oxy)hydroxides, which are known to strongly bind both As and P (Mucci et al., 2000; Plant
291 et al., 2005; Zhu et al., 2013). PC4_{geo} captures 12 % of the total variance and separates Mn, Co, Pb
292 and to a lesser extent Fe (positive loadings) from OM and Br (negative loadings; Fig. 2b). Although
293 Mn, like Fe and Al, is not confined to a specific mineral phase and can reflect both detrital or
294 dissolved and amorphous phases, the positive loadings are interpreted as reflecting Mn
295 (oxy)hydroxides, which bind Pb, especially when they contain cobalt (Co) (Yin et al., 2011). This
296 interpretation is supported by the positive loadings on PC4_{geo} of the ratio Mn:Fe, often used as a
297 paleolimnological proxy for bottom water oxygenation (Naeher et al., 2013). The negative loadings
298 could indicate a terrestrial OM fraction that is rich in Br (Leri and Myneni, 2012).

299 3.1.3 Cluster analysis of the elemental geochemistry

300

301 For the cluster analysis of the elemental geochemistry dataset, we selected a solution of six
302 clusters (cluster_{geo} 1–6; Fig. 1c). The cluster averages and standard deviations of each physical and
303 geochemical variable are given in Table S3 in the SI where they are compared to the averages
304 values of all analyzed sediment samples, hereafter referred as ‘whole-lake average’. Table 3
305 provides the cluster averages for a selection of geochemical parameters.

306 In the south basin, the sediments found at shallower water depth (cluster_{geo} 6; n=10) have a
307 higher B.D., are richer in bSi (negative scores on PC1_{geo}; Fig. 2a) and have lower than whole-lake
308 average trace metal concentrations (Table 1). In contrast, the sediments from the deeper sites
309 (cluster_{geo} 5; n=3) have the lowest B.D. and lowest bSi content (Table 1), and are enriched in OM
310 and trace metals (positive scores on PC1_{geo}; Fig. 2a). The sediments found at intermediate water
311 depths (cluster_{geo} 2; n=8) have positive scores on PC2_{geo} (Fig. 2a), and they have an OM content
312 within 10 % of whole-lake average while trace metal concentrations are above 10 % of whole-lake
313 averages (Table 1). The south basin as a whole has higher P concentrations than the north, east and
314 central areas, and in both intermediate and deeper sites, the sediments are rich in Fe and As
315 (positive scores on PC3_{geo}; Fig. 2b and Table 1).

316 The sediments found at shallow water depth between the north and east basins and in the
317 central area (cluster_{geo} 3; n=4) have the highest B.D. and are the most enriched in both bSi (negative
318 score on PC1_{geo}; Fig. 2a) and Mn and Fe (oxy)hydroxides (positive score on PC4_{geo}; Fig. 2b). A
319 small number of shallow, near-shore sampling locations (cluster_{geo} 4; n=4) have higher OM
320 concentrations than the whole-lake average, and are enriched in S and trace metals (positive scores
321 on PC1_{geo}; Fig. 2a and Table 1).

322 3.2 Sediment organic matter molecular composition

323

324 3.2.1 *General description and trends*

325

326 The Py-products identified in the surface sediments of Härsvatten were classified into 13 OM
327 classes, i.e., carbohydrates, N-compounds, chitin-derived Py-products, phenols, lignin,
328 chlorophylls, *n*-alkenes, *n*-alkanes, alkan-2-ones, steroids, tocopherols, hopanoids, and
329 (poly)aromatics, in agreement with previous studies using Py-GC/MS for different environmental
330 matrices (Faix et al., 1990; Faix et al., 1991; Peulvé et al., 1996; Nierop and Buurman, 1998;
331 Schulten and Gleixner, 1999; Lehtonen et al., 2000; Nguyen et al., 2003; Page, 2003; Buurman et
332 al., 2005; Fabbri et al., 2005; Kaal et al., 2007; Vancampenhout et al., 2008; Schellekens et al., 2009
333 ; Carr et al., 2010; Buurman and Roscoe, 2011; De La Rosa et al., 2011 ; McClymont et al., 2011;
334 Micić et al., 2011; Stewart, 2012). For the sake of making the presentation of the data and the
335 associated discussion more constrained and avoid over-interpreting individual compounds, the 162
336 identified organic compounds were reduced to 41 groups of compounds as described in Table 2.
337 This grouping is based on similarities in the molecular structure within the OM classes (cf. Table
338 S1 in the SI), and preliminary principal components analyses have shown that the compounds
339 within each of our 41 groups are highly positively correlated and thus present the same trends in our
340 study (data not shown). As an example, the 20 identified carbohydrate compounds, previously
341 demonstrated to derive from pyrolysis of polysaccharides and carbohydrates (Faix et al., 1991),
342 have been separated into 6 groups based on the number of C in the heterocycles of these
343 compounds and on their side-chain functional groups. The heterocycle of “furan” and “furanone”
344 compounds contains 4 C and 1 oxygen (O) atoms, and the side-chain is either aliphatic
345 ((alkyl)furans and (alkyl)furanones) or contains an oxygenated functional group (hydroxy- or
346 carboxy-furans and furanones). While the heterocycle of “pyran” compounds has 5 C and 1 O
347 atoms, the one of dianhydrohamnose, levoglucosenone and levosugars consists in 6 C and 1 O. But,

348 the levosugars contain three hydroxyl functional groups whereas dianhydrorhamnose contains 2
349 hydroxyl groups and levoglucosenone have a carbonyl group instead.

350 Summary statistics of these 41 groups of organic compounds are presented in Table 2 and the
351 detailed data are given in Table S4 in the SI. The coefficients of variation for the abundances of the
352 different organic compound groups range from 15 to 106 % with an average of 38 ± 20 %, showing
353 a remarkable in-lake variability of OM molecular composition. For most of the organic compound
354 groups, the average to median ratios are approximatively 1.0, indicating no extreme values.
355 However, slightly higher values (1.2-1.8) are observed for organic compounds derived from higher
356 plants and mosses, i.e. levosugars, lignin oligomers (syringols and guaiacols), *n*-alkanes C25-35,
357 alkan-2-ones C23-31 and tocopherols.

358 Most of the N-compounds, which usually derive more from algae than from higher plants and
359 mosses (Bianchi and Canuel, 2011), have the highest abundances among the three deepest sampling
360 locations (23.5-24.5 m) in the main south basin (S12, S18 and S24). These three deepest sampling
361 locations also present the highest abundances of (i) pyrolytic compounds containing an acetamide
362 functional group previously shown to be a good indicator of the presence of chitin, a component of
363 fungi cell walls and arthropod exoskeletons, in biological and geological samples (Gupta et al.,
364 2007); (ii) phytadienes, i.e., pyrolytic products of chlorophylls (Nguyen et al., 2003); (iii) short-
365 chain alkan-2-ones (2K C13-17); and (iv) steroids. In contrast, most of the carbohydrates, which
366 usually derive mostly from higher plants and mosses (Bianchi and Canuel, 2011), have the highest
367 abundances among the sediments situated close to the shoreline (N1-2, E3, S5, S23) such as for the
368 abundances of phenols, guaiacyl- and syringyl-lignin oligomers, long-chain *n*-alkenes (C27-28:1)
369 and diketodipyrrole (N-compounds), all specific of higher plants and/or mosses OM (Meyers and
370 Ishiwatari, 1993; Schellekens et al., 2009). The highest abundances of long-chain *n*-alkanes (C23-
371 26:0 and C27-35:0) and mid-chain *n*-alkanes (C17-22:0) are, however, observed for the shallower
372 sites (<2 m) situated between the larger north and south basins (sites M5-6).

373 Among the shallow sites (2.5-3.0 m) located between the north and east basin (N10, M1) and
374 the shallow and intermediate water depth (4-20 m) sites of the south basin (S1-4, S6-11, S13-17,
375 S19-22), we find the highest abundances of degradation products of carbohydrates (i.e.,
376 (alkyl)furans & furanones and hydroxyl- or carboxy-furans & furanones), of proteins, amino-acids
377 and/or chlorophylls (i.e., pyridines_O, (alkyl)pyrroles, pyrroles_O, pyrroledione &
378 pyrrolidinedione, pristenes) and of lipids (i.e., short-chain *n*-alkenes and *n*-alkanes – C9-16:1 and
379 C13-16:0) as well as the highest abundances of (poly)aromatic compounds indicative of highly
380 degraded OM (Schellekens et al., 2009; Buurman and Roscoe, 2011). The lowest abundances of the
381 (poly)aromatic and certain aliphatic compounds (i.e. *n*-alkenes C17-22 and C27-28, *n*-alkanes C13-
382 16 and alkan-2-ones C13-17) are observed among the sediments located close to the shoreline (N1-
383 2, E3, S5, S23), while the two shallow sites situated between the larger north and south basins (M5-
384 6) present the lowest abundances for all other organic compounds. To identify more accurately the
385 most significant relationships between the different organic compounds groups and to explore more
386 precisely their spatial distribution, the results of PCA and cluster analyses are further presented and
387 discussed.

388

389 3.2.2 Principal components of OM molecular composition

390

391 For the OM molecular composition dataset, six principal components (PC1-6_{OM}) were retained,
392 which explain 85 % of the total variance (Fig. 3). PC1_{OM}, which captures 30 % of the total variance,
393 separates organic compounds that are produced during OM degradation (positive loadings), from
394 molecules of higher plant or moss origin including those that are readily mineralized (negative
395 loadings; Fig. 3a). Compounds with positive loadings include i) (poly)aromatics (i.e., benzene,
396 acetylbenzene, benzaldehyde, alkylbenzenes C2-9 and polyaromatics); and ii) degradation products
397 of carbohydrates ((alkyl)furans & furanones; Schellekens et al., 2009), proteins, amino acids and/or
398 chlorophylls (aromatic N, (alkyl)pyridines and (alkyl)pyrroles; Jokic et al., 2004; Sinninghe Damsté

et al., 1992) and lipids (short-chain *n*-alkanes – C13-16:0 —, *n*-alkenes – C9-16:1 – and alkan-2-ones – 2K C13-17 –; Schellekens et al., 2009). Therefore, positive PC1_{OM} scores represent samples rich in degraded OM. The molecules of plant origin with negative PC1_{OM} loadings are syringol and guaiacol lignin oligomers that are specific for vascular plants, long-chain *n*-alkenes (C23-26:1 and C27-28:1) deriving from lipids of higher plants and/or mosses (Meyers and Ishiwatari, 1993) and long-chain alkan-2-ones (2K C23-31). Although alkan-2-ones C23-31 may arise with degradative oxidation of *n*-alkanes/*n*-alkenes (Zheng et al., 2011), they are also good biomarkers for mosses such as *Sphagnum* (2K C23-25) and for aquatic higher plants (2K C27-31) (Baas et al., 2000; Hernandez et al., 2001; Nichols and Huang, 2007). Furthermore, negative loadings on PC1_{OM} are found for the anhydrosugars, that are Py-products of fresh, high-molecular weight carbohydrates and cellulose from higher plants and mosses (never reported in Py-chromatograms of algae or arthropods; Marbot, 1997; Nguyen et al., 2003; Valdes et al., 2013) as well as for the ratio anhydrosugars: (alkyl)furans & furanones, a proxy for plant OM freshness (Schellekens et al., 2009), have also (Fig. 3a). Thus, negative PC1_{OM}-loadings likely reflect a fresh pool of OM coming from in-lake vegetation.

PC2_{OM} captures 14 % of the total variance and positive loadings are associated with (i) mid-chain *n*-alkanes/*n*-alkenes doublets that are known to be released during pyrolysis of resistant biomacromolecules such as cutin, suberin and algaenan (Buurman and Roscoe, 2011); (ii) pristenes, which are resistant degradation products of chlorophylls (Nguyen et al., 2003); and (iii) hopanoids, which are high-molecular weight pentacyclic compounds of prokaryotes, especially bacteria, origin (Meredith et al., 2008; Sessions et al., 2013). No compounds are significantly negatively correlated to PC2_{OM} (Fig. 3a). High PC2_{OM} scores thus represent samples rich in organic molecules that are resistant to degradation.

PC3_{OM} explains 13 % of the total variance and separates carbohydrates and N-compounds that are pyrolytic or degradation products of proteins, amino acids and/or chlorophylls (i.e., pyridines, pyrroledione and pyrrolidinedione) and of chitin on positive side, from aliphatic long-chain *n*-

425 alkanes (C23-26:0 and C27-35:0) coming from lipids of higher plants or mosses on negative side
426 (Fig. 3b).

427 On PC4_{OM}, which explains 13 % of the total variance, positive loadings are found for the
428 diketopiperazines, i.e. specific Py products of proteins or amino acids, the alkylamides and the
429 chlorophyll-derived phytadienes, which altogether indicate fresh algal organic residues (Peulvé et
430 al., 1996; Nguyen et al., 2003; Fabbri et al., 2005; Micić et al., 2010). Py products of chitin (Gupta
431 et al., 2007) and hopanoids, which derived from prokaryotes and mainly bacteria (Meredith et al.,
432 2008; Sessions et al., 2013), also have positive loadings on PC4_{OM}, while no compounds are
433 significantly negatively correlated to PC4_{OM} (Fig. 3b). Therefore, PC4_{OM} reflects OM input from in-
434 lake algae and micro-organisms (e.g., zooplankton, bacteria). Steroids, which have not yet been
435 reported by Py-GC/MS in aquatic matrices, have positive loadings on this PC4_{OM} suggesting that
436 the steroids released by pyrolysis in aquatic samples are mainly of algal origin.

437 For PC5_{OM}, capturing 8 % of the total variance, positive loadings are related to lignin
438 oligomers, which are specific for vascular plants (Meyers and Ishiwatari, 1993), and
439 diketodipyrrole, a N-compound often reported in soil pyrolysates (e.g., Schellekens et al., 2009;
440 Buurman and Roscoe, 2011). No compounds are associated with negative loadings on PC5_{OM} (Fig.
441 3c). Interestingly, the long-chain *n*-alkanes from higher plants or mosses lipids do not have positive
442 loadings on PC5_{OM}. We therefore interpret PC5_{OM} to relate to OM inputs from the forested
443 catchment, which is dominated by coniferous species. Coniferous trees generally have higher lignin
444 contents as compared to other vascular plants (Campbell and Sederof, 1996), while they contain
445 much lower amounts of *n*-alkanes than other plant species (Bush and McInerney, 2013).

446 PC6_{OM} captures 7 % of the total variance and has four compounds with significant positive
447 loadings, i.e., benzene, two benzenes with oxidized side-chain and carboxy- or hydroxy-furans and
448 furanones, i.e. furan and furanone heterocycles with an O atom in the side-chain (Fig. 3c). PC6_{OM}
449 may thus represent an intermediate degradation status of higher plants and/or mosses residues,
450 between the lignin oligomers or anhydrosugars (fresh) and the degraded polyaromatics and

451 benzenes C2-9 or (alkyl)furans and furanones (i.e. furan and furanone heterocycles with an
452 aliphatic side-chain).

453

454 3.2.3 Cluster analysis of OM composition

455

456 As with the elemental geochemistry dataset, a solution of six clusters (cluster_{OM} 1–6) was
457 relevant to represent the data on the spatial heterogeneity of OM molecular composition (Fig. 1d).
458 Each cluster is associated with one or a few of the OM types that were identified by the PC1-6_{OM}
459 (Fig. 3; Sect. 3.2.1). The cluster averages and standard deviations of each organic compound are
460 given and compared to whole-lake averages in Table S5 in the SI. Table 3 provides the cluster
461 averages for ratios indicative of OM source types and their degradation status based on literature
462 data and on the distribution of the organic compounds on PC1-6_{OM}.

463 In the south basin, the majority of sites found at shallower and intermediate water depths group
464 in cluster_{OM} 3 (n=14) and are enriched in degraded and resistant OM (positive scores on PC1_{OM};
465 Fig. 3a). The deep basin sites (cluster_{OM} 2; n=3) are enriched in fresh algal and zooplanktonic OM
466 (positive scores on PC4_{OM}; Fig. 3b). Accordingly, the values for the ratios indicative of higher
467 proportions of fresh, labile algal OM, based on N-compound or chlorophyll composition, are higher
468 in the deeper sites as compared to whole-lake averages, while the values are below or within $\pm 10\%$
469 of whole-lake averages in the sediments found at shallower and intermediate water depths (Table
470 1). In contrast, the ratios indicative of higher plant and moss OM freshness based on carbohydrate
471 or lignin composition have similar values, and lower as compared to whole-lake averages, for all
472 sediments of the south basin. Furthermore, the clusters_{OM} 2 and 3 are characterized by higher values
473 for the ratios specific of algal versus higher plant and moss OM based on the proportions of N-
474 compounds versus carbohydrates or chlorophylls versus lignin and long-chain *n*-alkanes and alkan-
475 2-ones (Table 1). The rest of the south basin sites, fall within cluster_{OM} 1 (n=1), 5 (n=2) or 4 (n=1),
476 which are described below.

477 The majority of sites in the northern half of the lake group within cluster_{OM} 1 (n=15) with
478 isolated shallower sites falling within clusters_{OM} 3 (n=1), 4 (n=2) and 5 (n=2). The sediments of
479 cluster_{OM} 1 are rich in fresh plant (higher plants or mosses) OM coming from in-lake productivity
480 (negative scores on PC1_{OM}; Fig. 3a) and have higher values than whole-lake averages for the ratios
481 specific of in-lake vs terrestrial plant OM and of higher plant OM freshness (Table 1). In contrast,
482 the values for these ratios are below 10 % of whole-lake averages for the south basin sites,
483 indicating that terrestrial input is the main source of plant OM to the sediments of the main basin of
484 Härsvatten.

485 The cluster_{OM} 5 represents some near-shore locations (n=4), which are enriched in OM derived
486 from the coniferous-forested catchment (positive scores on PC5_{OM}; Fig. 3c). The cluster_{OM} 4 (n=4),
487 which groups shallow sites located close to the lake outlet (south basin, S16) and between the north
488 and east basins (N10 and M1), is characterized by high proportions of degraded and resistant OM
489 (positive scores on PC5_{OM}; Fig. 3a). Two shallow sites of the central area (cluster_{OM} 6; n=2) show
490 an enrichment in aliphatic molecules derived from higher plant and moss lipids (negative loadings
491 on PC3_{OM}; Fig. 3b). Both clusters_{OM} 4 and 6 have values for the ratio indicative of in-lake:terrestrial
492 plant OM above 10 % of the whole-lake average, while the values for the ratios specific of algal vs
493 higher plant and moss OM and of OM freshness based on N-compounds and carbohydrates
494 composition are below 10 % of whole-lake averages (Table 1). Cluster_{OM} 6 differs from cluster_{OM} 4
495 by its higher values for the ratios specific of OM freshness based on chlorophyll and lignin
496 composition.

497 3.3 Factors and processes involved in the spatial distribution of OM molecular composition

498

499 The surface sediments used in this study comprise the uppermost 10 cm. Given the inherent
500 variation in sedimentation rates across a lake basin, each bulk sample represents material deposited
501 over different timescales. We know from the developmental work for our Py-GC/MS method using
502 annually laminated sediments that there are transformations in OM composition within the
503 uppermost few cm, i.e., the first few years following deposition (Tolu et al. 2015). Thus, these bulk
504 sediment samples provide initial insights into the spatial variability in molecular OM composition
505 within a lake basin resulting from longer-term sedimentation processes (including those within the
506 sediment) reflecting years to decades.

507 The distribution of both clusters_{geo} and clusters_{OM} within Härsvatten shows a similar general
508 pattern (Fig. 1c and 1d) where a main feature is the separation of most of the sediments located in
509 the north and east basins (cluster_{geo} 1 and cluster_{OM} 1) from those found in the main south basin
510 (clusters_{geo} 2, 5, 6 and clusters_{OM} 2, 3). The other similarities are i) a separation of the sediments
511 within the main, south basin according to water depth, with cluster_{geo} 5 and cluster_{OM} 2 grouping
512 the deeper sites and clusters_{geo} 2, 6 and cluster_{OM} 3 grouping the shallow and intermediate depth
513 sites; and ii) a separation of the shallower sites that are located close to the shore (cluster_{geo} 4 and
514 cluster_{OM} 5) from the ones found between the north and east basins and between the central area
515 and the south basins (cluster_{geo} 3 and clusters_{OM} 4 and 6).

516

517 3.3.1 *Spatial variability in the main south basin*

518

519 As shown previously for OM (as % LOI) and Pb (Bindler et al. 2001), there is a physical and
520 inorganic geochemical gradient from shallower to deeper waters reflecting sediment focusing in the
521 south basin of Härsvatten. B.D. and bSi decrease from shallower (cluster_{geo}6) to intermediate
522 (cluster_{geo}2) and to deeper areas (cluster_{geo}5), whereas there is a progressive enrichment in organic

523 matter and trace elements with increasing water depth (Fig. 1c; Table 1). For example, B.D.
524 decreases from ~ 0.07 to 0.03 g cm^{-3} while OM and Hg increase from ~ 34 to 52% and from ~ 230 to
525 920 ng g^{-1} , respectively, in shallower versus the deepest locations. At intermediate depths
526 ($\text{cluster}_{\text{geo}2}$), OM, B.D., bSi and most trace metals (i.e., Cu, Ni, Hg, Zn) are between those of
527 shallow and deep locations. Sediment focusing is thus an important process for sediment
528 geochemistry in the large, deep basin of Härsvatten, which presents a relatively simple
529 morphometry. According to the model of sediment focusing, the sediments found at shallower (<11
530 m; $\text{cluster}_{\text{geo}6}$), intermediate (11–21 m; $\text{cluster}_{\text{geo}4}$) and deeper water depths (>23 m; $\text{cluster}_{\text{geo}5}$)
531 would correspond to zones of erosion, transportation and accumulation, respectively (Håkanson,
532 1977). The bSi decline, from ~ 15 to 4% , indicates a decrease of diatom production with depth due
533 to increasing light attenuation, and thus suggests that the diatom assemblage is dominated by
534 benthic diatoms as shown for many acidified lakes, such as the surrounding lakes in the Svartedalen
535 nature reserve (e.g., Andersson, 1985; Anderson and Renberg, 1992).

536 In this main basin of Härsvatten, OM originates from a combination of autochthonous algal
537 production and allochthonous input (Sect. 3.2.2). The dominance of benthic diatoms in acidified
538 lakes and the declining bSi content with depth would indicate that the algal material in deeper areas
539 of the basin should mainly derive from resuspended benthic algal production. However, this benthic
540 algal production is not reflected in the OM molecular composition. The sediments from shallow and
541 intermediate water depths ($\text{cluster}_{\text{OM}3}$) are mainly composed of degraded and resistant OM, while
542 the sediments from deeper sites ($\text{cluster}_{\text{OM}2}$) are enriched in fresh algal and zooplanktonic OM
543 (Fig. 1d; Sect. 3.2.2). Although our results are based on the top 10 cm of sediment and thus account
544 for different sediment ages, we suggest that the higher proportions of decomposed algal material,
545 based on N-compound and chlorophyll composition (Table 1), at shallower and intermediate water
546 depths (<21 m) than at the deepest sites (23.5–24.5 m) reflect higher mineralization rates of OM in
547 shallow/intermediate areas. Higher OM mineralization rates in shallow/intermediate areas are most
548 probably due to more oxic conditions, which are known to prevail in epilimnetic and metalimnetic

549 sediments (Ostovsky and Yacobi, 1999); the epilimnion in Härsvatten has been assessed to extend
550 to 10–15 m water depth. Higher OM preservation in the deeper area may also be favored by higher
551 accumulation rates as compared to shallow/intermediate areas (as consequence of sediment
552 focusing), but the sedimentation rates in the deeper areas of Härsvatten are nonetheless very low,
553 with the uppermost 30 cm being deposited during the last c. 500 year (Bindler et al., 2001).
554 Moreover, the elemental geochemistry indicates that the sites found at intermediate water depths
555 (cluster_{geo} 6; 11–21 m) correspond in the sediment focusing model to transportation zones, which
556 experience recurrent resuspension events that favor gas exchanges and mineralization of OM
557 (Ståhlberg et al., 2006). Occurrence of oxic conditions at intermediate depths in the south basin is
558 supported by the higher concentrations of Fe, Mn, As, Co and P and the high Fe:Al values, this
559 combination of parameters being often indicative of Fe and Mn (oxy)hydroxides (Table 1; Sect.
560 3.1.1). In line with our hypothesis, higher OM mineralization rates in oxic versus anoxic sediments
561 have previously been reported (Bastviken et al., 2004; Isidorova et al., 2016). However, in contrast
562 to the more algal-derived OM, we do not observe significant differences between the sediments of
563 shallower/intermediate water depths and the deepest sites for ratios indicative of higher plant and
564 moss OM freshness (Table 1). Because higher plant and moss OM is mainly of allochthonous
565 origin in this basin, our results indicate that primarily autochthonous algal OM is mineralized in the
566 epilimnetic and metalimnetic sediments of this deeper, steeper-sloped, basin of Härsvatten. This is
567 consistent with the suggestion that allochthonous OM is recalcitrant to sediment mineralization
568 after its degradation in the catchment and within the water column (Gudasz et al., 2012).

569 Overall, our molecular characterization of OM in the south basin suggests an enrichment in
570 algal versus allochthonous OM (e.g., higher N-compounds:carbohydrates) in the deeper areas of a
571 deep, simple lake basin, in line with previously reported sediment C:N ratios along lake-basin
572 transects (Kumke et al., 2005; Dunn et al., 2008; Bruesewitz et al., 2012). Given our data on the
573 degradation status of algal and allochthonous OM, we believe that this trend in OM quality results
574 from preferential degradation of algal versus allochthonous OM in sediments at

shallower/intermediate water depth in addition to the known focusing of living, and residues of, authochthonous OM towards deeper sites (Ostrovsky and Yacobi, 1999).

3.3.2 Spatial variability in the central, north and east basins and near-shore locations

In the northern half of the lake, 11 of 19 locations fall within cluster_{geo}1 (Fig. 1c), which distinguishes itself geochemically only by somewhat lower than average concentrations of elements often associated with (oxy)hydroxides (i.e., Fe, Mn, As, P and Co; Table 1 and Sect. 3.1.2). Sediments from the shallowest locations can potentially fall in one of four different clusters (clusters_{geo}1, 3, 4 or 6). Thus, for the northern half of the lake there is no evidence of sediment focusing. The effect is either limited by the more gentle slopes of the north and east basins (Blais and Kalff, 1995), modified by the water circulation resulting from the prevailing winds towards the north-east (Bindler et al. 2001, Abril et al., 2004), and/or interrupted by aquatic vegetation that acts as a sediment trap (Benoy and Kalff, 1999). Aquatic vegetation represents a major source of OM to the sediments of the north, east and central basins (clusters_{OM} 1, 4 and 6; Fig. 1d; Table 1 and Sect. 3.2.2). The enrichment of aquatic higher plant or moss OM in these sediments is consistent with field observations during the original sediment coring in winter 1997, where mosses and *Isoetes* (a vascular angiosperm plant) were observed in some parts of the lake to a depth of at least 10 m (Bindler et al., 2001). The presence of such submerged vegetation in Härsvatten is favored by its acidic, clear water (i.e., deeper light penetration), as previously observed for other acidified boreal Swedish lakes, such as the nearby lake Gårdsjön (Andersson, 1985; Grahn, 1985). Benthic aquatic vegetation is also favored in the northern half of Härsvatten by the more gentle slopes, comparatively shallow water depth and thus greater availability of light than in the deep, steeper-sloped south basin where allochthonous input appears as the main source of higher plant and moss OM (Sect. 3.2.2; Table 1).

600 The sediments found across the north and east basins and at the deeper sampling site of the
601 central area (clusters_{OM} 1; Fig. 1d) have the highest proportions of fresh, labile higher plant and
602 moss OM, e.g. anhydrosugars (Sect. 3.2.2; Table 1). Also, the proportions of fresh, labile algal OM
603 is as high as in the deeper anoxic sediments of the main south basin and two times higher than in
604 the sediments found at shallow water depth in the south basin and central areas, although these sites
605 span the same water depth range (3–11 m) and have relatively similar bSi contents (Table 1). These
606 results indicate the accumulation of fresh autochthonous, both plant and algal, OM in sediments
607 associated with in-lake vegetation even if they are below or within the epilimnion (i.e., supposed
608 oxic conditions). A possible explanation is that the input of labile, decomposing in-lake higher plant
609 and moss OM consumes oxygen and results in locally anoxic conditions in the sediment, which in
610 turn lower OM mineralization rates (Bastviken et al., 2004; Isidorova et al., 2016). This hypothesis
611 may explain the lower than whole-lake average concentrations of elements or elemental ratios often
612 associated with (oxy)hydroxides (i.e., Fe, Mn, As, Co, P contents and Fe:Al) in these
613 epilimnetic/metalimnetic sediments (cluster_{geo} 1; Table 1). This interpretation is consistent with
614 laboratory experiments, where, for example, Kleeberg, 2013 had shown that inputs of macrophyte
615 residues to sediments results in oxygen depletion and microbially mediated reduction of Fe and Mn
616 oxides. However, we cannot rule out that other factors, such as shallow groundwater discharges that
617 are rich in (oxy)hydroxides or diagenetic processes that lead to Fe enrichment in sediments, can be
618 involved in the higher concentrations of Fe, Mn and other elements known to be associated with Fe
619 and Mn (oxy)hydroxides in the sediments of the south basin as compared to the sediments of the
620 north and south basins.

621 The shallow sites located between the north and east basins and between the central area and
622 the south basin (i.e., cluster_{geo} 3 and clusters_{OM} 4 and 6; Fig. 1c and 1d) have higher than whole-lake
623 averages bSi contents and values for the ratio in-lake:terrestrial higher plant and moss OM,
624 suggesting that these sediments receive plant OM from in-lake vegetation and algal OM from
625 benthic production (Table 1). However, the proportions of fresh, labile plant and algal OM based on

626 N-compound and carbohydrate composition in these central sediments are much lower than in the
627 sediments found across the north and east basins (Table 1). Probably, these central areas are not
628 sites for aquatic vegetation growth, but receive in-lake plant OM produced within the north and east
629 basins that has been degraded during transport and/or is degraded at these shallow central sites due
630 to more oxic conditions. More oxic conditions at these shallow central sites are also suggested by a
631 higher occurrence of Fe and Mn (oxy)hydroxides (Fe, Mn, As, Co, and P contents, Fe:Al and
632 Mn:Fe above 10 % of whole-lake averages; Table 1). Among these shallow central sites, two
633 locations (cluster_{OM} 6) are specifically rich in higher plant and moss lipids (i.e., C23-35:0; Table S3
634 in the SI) and, have high proportions of fresh higher plant OM based on lignin composition while
635 the proportions of fresh carbohydrates (anhydrosugars) versus total carbohydrates is low (Table 1).
636 This suggests preservation of higher plant cell-wall lipids and lignin with respect to carbohydrates
637 at these two shallow sites, in agreement with the known faster assimilation of carbohydrates versus
638 lipid and lignin structures (Bianchi and Canuel, 2011). However, no reasonable hypothesis could be
639 given to explain this difference in OM molecular composition between the sediments at sites M5-6
640 and the ones at sites N10 and M1 given their similar water depth and elemental geochemistry.

641 Among the sediments found in a small number of near-shore locations (cluster_{geo} 4 and
642 cluster_{OM} 5; n=4), three are located in two more-sheltered bays at the northwestern corner and the
643 southern end of the lake that are more protected from wind circulation (Bindler et al. 2001, Abril et
644 al. 2004). The sediments of these three locations predominantly accumulate terrestrial OM as
645 indicated by the abundance in lignin oligomers and the ratio indicative of in-lake:terrestrial plant
646 OM that are respectively above and below 10 % of the whole-lake averages (Table 3).
647 Accumulation of OM coming from the coniferous-forested catchment most probably explained the
648 high OM content (i.e. 52-58 %, which is as high as in the deeper sediments of the main south basin)
649 as well as the high concentrations of S and trace metals (i.e., Hg, Pb and Zn) in these near-shore
650 sediments (Table 1). Boreal forest soils are known to be enriched in S and trace metals because
651 their organic fraction retains atmospheric S and trace metals deposited over the industrial era

652 (Johansson and Tyler, 2001). Also, there is evidence that the transport of terrestrial OM to boreal
653 aquatic ecosystems is associated with significant inputs of trace metals (Grigal, 2002; Rydberg et
654 al., 2008). Alternatively, high S and trace metal contents could be due to accumulation of metal
655 sulfides due to near-shore groundwater gradients and/or anoxic conditions, or to redox cycling
656 related to the important input of terrestrial OM.

657

658 *3.3.3 Implication for in-lake and/or global elemental (e.g., C, nutrients, trace metals) cycling*

659

660 The molecular composition of natural OM has been shown to exert a strong influence on key
661 biogeochemical reactions involved in in-lake and global cycling of C, nutrients and trace metals,
662 such as C mineralization or nutrients/trace metals sorption and transformations into mobile and/or
663 bioavailable species (Drott A et al., 2007; Sobek et al., 2011; Gudas et al., 2012; Tjerngren et al.,
664 2012; Kleeberg, 2013; Bravo et al., 2017). Our work demonstrates that OM molecular composition
665 can vary significantly within a single lake system in relation to basin morphometry, lake chemical
666 and biological status (e.g., presence of macrophytes, which is influenced by, e.g., acidification) and
667 the molecular structure/properties of the different OM compounds (e.g., higher resistant of
668 allochthonous versus autochthonous OM upon degradation). Our results further show that it may be
669 problematic to extrapolate data on OM composition from only a few sites or one basin when scaling
670 up to a whole lake. Thus, investigating sedimentary processes and the resulting fate of C and trace
671 elements using sampling strategies focused on the deepest area of a lake or on single transects from
672 shallower to deeper sites, may not fully capture the variation in either elemental geochemistry or
673 OM composition.

674 Overall, this study underlines that the OM molecular composition and its spatial heterogeneity
675 across a lake are two factors that should be considered to better constrain processes involved in the
676 fate of C, nutrients and trace metals in lake ecosystems, to improve whole-lake budgets for these
677 elements and to better assess pollution risks and the role of lakes in global elemental cycles.

678 **Data availability.**

679 The supporting information includes the raw data for sediment elemental geochemical parameters
680 and for the 41 groups of organic compounds (resulting from the identification of 162 pyrolytic
681 organic compounds) used for the statistical analysis and discussion. Raw data for the 162 pyrolytic
682 organic compounds will be provided upon request from the authors.

683

684 **Author contribution.**

685 J. Tolu and R. Bindler designed research. J. Tolu performed Py-GC/MS analyses with help from L.
686 Gerber and did the data treatment with the data processing pipeline of L. Gerber. J. Tolu and J.
687 Rydberg performed XRF and mercury analyses. J. Tolu and C. Meyer-Jacob performed FTIR
688 measurements and C. Meyer-Jacob determined the inferred bSi. J. Tolu, J. Rydberg, C. Meyer-
689 Jacob and R. Bindler interpreted the data. J. Tolu prepared the manuscript with consistent
690 contributions from J. Rydberg, R. Bindler and C. Meyer-Jacob.

691

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697

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Tables

Table 1. Summary statistics for sediment elemental geochemistry

Whole sample collection except the two outliers							Outliers (M4, S15)			
	Unit	Av. ^a ± sd ^b	CV ^c	Median	A:M ^d	Min ^e -Max ^f	Av. ± sd	CV ^c	Median	Min-Max
W.D.	m	9 ± 7	74	7	1.23	2-25	3.4 ± 0.6	19	3	2.9-3.8
B.D.	g cm ⁻³	0.06 ± 0.02	38	0.06	1.05	0.02-0.13	0.67 ± 0.09	14	0.06	0.61-0.74
bSi	%	13 ± 6	48	12	1.05	4-25	1.9 ± 0.2	0	12	1.7-2.0
LOI	%	38 ± 10	26	37	1.01	10-58	3.6 ± 0.8	20	37	3.0-4.1
[S]	mg kg ⁻¹	11876 ± 5920	50	11305	1.05	4685-29190	2570 ± 552	21	10610	2180-2960
[Br]	mg kg ⁻¹	149 ± 35	23	152	0.99	71-225	16 ± 7	44	148	11-21
[Cu]	mg kg ⁻¹	34 ± 13	37	32	1.07	12-75	9 ± 3	31	31	7-11
[Ni]	mg kg ⁻¹	19 ± 4	24	19	0.99	10-27	12 ± 4	35	19	9-15
[Hg]	µg kg ⁻¹	337 ± 202	60	286	1.18	117-1152	28 ± 9	33	274	21-34
[Pb]	mg kg ⁻¹	192 ± 74	39	184	1.05	58-422	22 ± 16	76	178	10-33
[Zn]	mg kg ⁻¹	219 ± 108	49	207	1.06	43-445	50 ± 16	31	200	39-61
[Al]	%	3 ± 1	17	3	1.06	2-4	5.67 ± 0.01	0.1	3	5.66-5.67
[Y]	mg kg ⁻¹	25 ± 8	32	25	1.01	7-43	20 ± 4	18	25	17-22
[Fe]	%	5 ± 3	65	4	1.26	1-12	3.4 ± 0.1	4	4	3.3-3.5
[As]	mg kg ⁻¹	35 ± 20	56	28	1.26	5-73	<DL		27	0-0
[P]	mg kg ⁻¹	1624 ± 741	46	1401	1.16	655-3769	949 ± 57	6	1389	908-989
[Mn]	mg kg ⁻¹	729 ± 1690	232	180	4.06	94-7981	1060 ± 845	80	184	462-1657
[Co]	mg kg ⁻¹	19 ± 15	77	14	1.39	5-76	17 ± 9	56	14	10-23
[Ca]	mg kg ⁻¹	5261 ± 1306	25	5213	1.01	2860-9300	26540 ± 7566	29	5283	21190-31890
[K]	mg kg ⁻¹	4426 ± 1020	23	4485	0.99	2420-6140	10510 ± 2616	25	4580	8660-12360
[Mg]	mg kg ⁻¹	1488 ± 354	24	1500	0.99	870-2130	7495 ± 3599	48	1515	4950-10040
[Na]	mg kg ⁻¹	1795 ± 659	37	1743	1.03	440-3380	10695 ± 587	5	1783	10280-11110
[Si_{inorganic}]	%	11 ± 4	33	11	1.06	4-21	23 ± 1	3	11	22-23
[Sr]	mg kg ⁻¹	55 ± 16	29	55	1.01	27-116	235 ± 24	10	55	218-252
[Ti]	mg kg ⁻¹	2115 ± 495	23	2200	0.96	997-2870	4357 ± 2348	54	2215	2697-6017
[V]	mg kg ⁻¹	63 ± 15	23	60	1.05	36-101	75 ± 23	31	60	58-91
[Zr]	mg kg ⁻¹	101 ± 31	31	100	1.01	39-160	158 ± 6	4	103	153-162

^a Av.: average; ^b sd: standard deviation; ^c CV: coefficient of variation calculated as relative standard deviation in %; ^d A:M: ratio between average and median; ^e Min.: minimal value; ^f Max.: maximal value

Table 2. Summary statistics for the molecular composition of sediment OM given as relative abundances (expressed in %) of the 41 groups of pyrolytic organic compounds, which belong to 13 classes of OM as indicated by the grey shading (*to be continued*)

Compounds included		Av ^a ± sd	CV	Median	A:M	Min-Max
Carbohydrates						
(Alkyl)-furans & furanones	3-furaldehyde, 2-furaldehyde, 2-acetyl-furan, Methyl-3-furaldehyde, 2(5H)-furanone, Methyl-2-furaldehyde, Dihydro-methyl-furanone, Methyl-2(5H)-furanone, Methyl-2-furaldehyde	15 ± 4	30	14	1.06	8-28
Hydroxy- or carboxy-furans & furanones	2-Furancarboxylic acid, methyl ester; 2,5-Dimethyl-4-hydroxy-3(2H)-furanone; 5-(hydroxymethyl)-2-furaldehyde	4.1 ± 1.2	29	4.0	1.03	0.8-7.5
Pyrans	5,6-dihydro-pyran-2-one, 4-hydroxy-5,6-dihydro-pyran-2-one	3.4 ± 1	30	3.2	1.06	1.2-5.3
Dianhydrorhamnose	Dianhydrorhamnose	1.6 ± 0.5	28	1.7	0.99	0.3-2.7
Levogluconenone	Levogluconenone	2.2 ± 0.4	20	2.2	1.00	1.3-3.1
Anhydrosugars	Anhydrohexose, Levogalactosan, Levomannosan, Levoglucosan	3.7 ± 2.6	71	2.5	1.46	0.8-11
Chitin derived compounds						
Chitin-derived compounds	Acetamide, 3-acetamido-furan, 3-acetamido-4-pyrone, Oxazoline	2.5 ± 1	40	2.6	0.98	0.2-4.2
N-compounds						
(Alkyl)pyridines	Pyridine, 2-methyl-pyridine, 3/4-methyl-pyridine	0.3 ± 0.1	34	0.3	0.95	0.1-0.5
Pyridines_O, <i>i.e. pyridines with side chain containing a "C=O" function</i>	2-acetylpyridine, 3-acetylpyridine, 2-Methyl-5-acetoxypyridine	0.7 ± 0.1	18	0.7	1.00	0.2-0.9
(Alkyl)pyrroles	Pyrrole, Methyl-pyrrole	2.4 ± 0.5	22	2.4	1.01	1.7-3.5
Pyrroles_O, <i>i.e. pyrroles with side chain containing a "C=O" function</i>	2-formyl-pyrrole, 2-acetyl-pyrrole, 2-formyl-1-methylpyrrole	1.0 ± 0.2	25	0.9	1.04	0.5-1.4
Pyrroledione & pyrrolidinedione	2,5-pyrroledione, 2,5-pyrrolidinedione	1.2 ± 0.3	29	1.2	0.98	0.2-1.7
Aromatic N- compounds	Benzeneacetonitrile, Benzenepropanenitrile	0.8 ± 0.3	36	0.8	1.03	0.3-1.4
Indoles	Indole, Methyl-indole	1.5 ± 0.4	24	1.5	1.03	0.5-3.1
Diketodipyrrole	Diketodipyrrole	0.8 ± 0.2	22	0.8	1.01	0.4-1.2
Diketopiperazines	Pro-Ala, Pro-Val, Pro-Val, Cyclo-Leu-Pro, Pro-Pro, Pro-Phe	1.5 ± 0.4	30	1.5	1.02	0.3-2.6
Alkylamides	6 alkylamides	0.6 ± 0.3	51	0.6	1.06	0.1-1.7
Phenols						
Phenols	Phenol, 2-methyl-phenol, 3/4- methyl-phenol, dimethyl-phenol, Ethyl-phenol, Propenyl-phenol	8 ± 1	15	8	1.02	4.4-11.4
Lignin						
Syringols	Syringol, 4-vinyl-syringol, 4-formyl-syringol, 4-allenesyringol, Acetosyringone	0.5 ± 0.4	83	0.4	1.32	0.1-1.9
Guaiacols	Guaiacol, Ethyl-guaiacol, 4-vinyl-guaiacol, 4-propenyl-guaiacol, Vanillin, 4-alleneguaiacol, Acetovanillone, Methyl-ester-vanillic acid, , Guaiacylacetone	3.6 ± 2.3	65	2.9	1.24	1.1-13.5

Table 2. Summary statistics for the molecular composition of sediment OM given as the relative abundances (expressed in %) of the 41 groups of Py organic compounds, which belong to 13 classes of OM that are indicated by the grey shading (*following part*)

Chlorophylls						
Pristenes	Prist-1-ene, Prist-2-ene	2.7 ± 0.8	28	2.8	0.97	0.4-4.6
Phytadienes	Phytadiene 1, Phytadiene 2	1.9 ± 0.7	35	1.8	1.04	0.2-3.6
n-alkenes						
C9-16:1	n-alkenes C9, C13, C14, C16	3.5 ± 0.8	23	3.6	0.98	1.8-5.1
C17-C22:1	n-alkenes C17, C18, C19, C20, C21, C22	6 ± 1	17	6.2	0.97	3.5-8.9
C23-26_1	n-alkenes C23, C24, C25, C26	2.9 ± 0.9	32	2.7	1.09	0.6-5.4
C27-28:1	n-alkenes C27, C28	0.8 ± 0.4	47	0.7	1.10	0.1-1.4
n-alkanes						
C10-16:0	n-alkanes C10, C11, C12, C13, C14, C15, C16	2.5 ± 0.6	23	2.5	1.03	1.3-4.1
C17-22:0	n-alkanes C17, C18, C19, C20, C21, C22	3.9 ± 0.8	21	4.0	0.98	1.6-5.4
C23-26:0	n-alkanes C23, C24, C25, C26	2.8 ± 1.4	49	2.7	1.07	1.4-8.8
C27-35:0	n-alkanes C27, C28, C29, C30, C31, C32, C33, C35	4.3 ± 3.5	80	3.6	1.20	1.1-21.3
Alkan-2-ones						
2K C13-17	Alkan-2-ones C13, 16, 17	1.3 ± 0.4	33	1.4	0.96	0.6-2.2
2K C19-21	Alkan-2-ones C19, 20, 21	0.3 ± 0.1	45	0.3	0.97	0-0.8
2K C23-31	Alkan-2-ones C23, 14, 25, 26, 27, 28, 29, 31	1.3 ± 0.8	62	1.1	1.24	0.1-3.3
Steroids						
Steroids	Cholest-2-ene, Cholesta-3,5-diene, Stigmasta-5,22-dien-3-ol, acetate, Sitosterol, Cholesta-3,5-dien-7-one, Stigmasta-3,5-dien-7-one	1.2 ± 0.9	70	1.1	1.10	0-4.3
Tocopherols						
Tocopherols	γ-Tocopherol, α-Tocopherol	0.3 ± 0.3	106	0.2	1.75	0-1.5
Hopanoids						
Hopanoids	Trinosphopane, Norhopene, 22,29,30-trisnorhop-17(21)-ene, 22,29,30-trisnorhop-16(17)-ene, Norhopane, 25-norhopene	1.3 ± 0.4	31	1.4	0.94	0.2-1.9
(Poly)aromatics						
Benzene	Benzene	0.9 ± 0.4	43	0.8	1.14	0.4-2.5
Benzaldehyde	Benzaldehyde	0.6 ± 0.3	41	0.6	1.08	0.3-1.5
Acetylbenzene	Acetyl-benzene	1.1 ± 0.4	39	1.0	1.10	0.6-2.3
Alkylbenzenes C3-9	Ethyl-methyl-benzene, Benzene C7, Benzene C9,	1.9 ± 0.5	23	1.8	1.07	1.4-3.5
Polyaromatics	Styrene, Indene, Dihydro-naphthalene, Dihydro-inden-1-one, 1-methyl-napthalene, 2-methyl-napthalene, Biphenyl, Fluorene, Anthracene	1.4 ± 0.4	27	1.3	1.04	0.8-2.1

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Table 3. Whole-lake and clusters average for a selection of elemental geochemical parameters and of ratios indicative of OM source types and their degradation status

SPECIFIC FEATURES IN GEOCHEMISTRY							
	Whole-lake ^a	Near-shore sites	North/East basins	South basin			Shallow central areas
				Shallower	Intermediate depth	Deeper	
		Cluster _{geo} 4	Cluster _{geo} 1	Cluster _{geo} 6	Cluster _{geo} 2	Cluster _{geo} 5	Cluster _{geo} 3
	(n ^b =42)	(n=4)	(n=13)	(n=10)	(n=8)	(n=3)	(n=4)
Water depth (m)	9 ± 7 (78 %) ^c	4 ± 2	5 ± 3	8 ± 3	15 ± 4	24 ± 1	2 ± 1
Bulk density (g cm ⁻³)	0.06 ± 0.02 (33 %)	0.06 ± 0.03	0.07 ± 0.02	0.07 ± 0.02	0.05 ± 0.01	0.026 ± 0.009	0.10 ± 0.02
[bSi] (%)	13 ± 6 (46 %)	12 ± 6	13 ± 3	15 ± 7	7 ± 3	4.2 ± 0.3	21 ± 4
LOI (%)	38 ± 10 (26 %)	50 ± 12	39 ± 5	34 ± 7	37 ± 4	52 ± 2	20 ± 8
[S] (mg kg ⁻¹)	11876 ± 5920 (50 %)	17510 ± 833	11683 ± 3440	7550 ± 1900	12896 ± 3315	26227 ± 4833	4879 ± 148
[Br] (mg kg ⁻¹)	149 ± 35 (23 %)	130 ± 6	153 ± 36	145 ± 35	154 ± 19	204 ± 26	116 ± 32
[Cu] (mg kg ⁻¹)	34 ± 13 (38 %)	36 ± 5	28 ± 6	30 ± 7	42 ± 6	65 ± 10	24 ± 13
[Ni] (mg kg ⁻¹)	19 ± 5 (25 %)	21 ± 1	18 ± 4	17 ± 2	21 ± 4	27 ± 1	12 ± 4
[Hg] (μg kg ⁻¹)	337 ± 202 (60 %)	407 ± 141	251 ± 47	230 ± 69	427 ± 94	917 ± 212	203 ± 87
[Zn] (mg kg ⁻¹)	219 ± 108 (49 %)	279 ± 31	212 ± 68	139 ± 42	305 ± 86	417 ± 33	63 ± 16
[Fe] (%)	5 ± 3 (60 %)	3.1 ± 2.1	2.7 ± 1.7	3.6 ± 1.5	9.1 ± 2.4	4.3 ± 2.2	5.5 ± 1.7
Fe:Al	1.5 ± 0.8 (53 %)	1.0 ± 0.5	1.0 ± 0.6	1.1 ± 0.3	2.5 ± 0.9	1.3 ± 0.6	1.9 ± 0.3
[As] (mg kg ⁻¹)	35 ± 20 (57 %)	27 ± 17	26 ± 16	25 ± 11	64 ± 11	48 ± 14	29 ± 9
[P] (mg kg ⁻¹)	1624 ± 741 (46 %)	927 ± 240	1065 ± 295	2088 ± 730	2074 ± 275	2766 ± 869	1224 ± 216
[Mn] (mg kg ⁻¹)	729 ± 1690 (231 %)	162 ± 53	182 ± 67	184 ± 50	305 ± 93	171 ± 13	5700 ± 1597
Mn:Fe	0.02 ± 0.03 (150 %)	0.007 ± 0.002	0.008 ± 0.003	0.006 ± 0.002	0.004 ± 0.001	0.005 ± 0.002	0.111 ± 0.051
[Co] (mg kg ⁻¹)	19 ± 15 (79 %)	15 ± 8	12 ± 6	13 ± 5	26 ± 11	14 ± 2	49 ± 24
[Pb] (mg kg ⁻¹)	192 ± 90 (47 %)	199 ± 58	132 ± 53	115 ± 42	300 ± 59	315 ± 7	182 ± 96
SPECIFIC FEATURES IN OM COMPOSITION							
	Whole-lake	Near-shore sites	North/East basins	South basin		Shallow central areas	
				Shallower/intermediate depth	Deeper		
		Cluster _{OM} 5	Cluster _{OM} 1	Cluster _{OM} 3	Cluster _{OM} 2	Cluster _{OM} 4	Cluster _{OM} 6
	(n=42)	(n=4)	(n=16)	(n=14)	(n=3)	(n=3)	(n=2)
Water depth (W.D.)	9 ± 7 (78 %)	4 ± 2	7 ± 5	11 ± 5	24.1 ± 0.5	3.2 ± 0.9	1.8 ± 0.1
LOI (%)	38 ± 10 (26 %)	50 ± 12	39 ± 4	36 ± 5	52 ± 2	24 ± 4	14 ± 6
(C23-35:0+2K C23-31): Lignin ^d	In-lake:Terrestrial plant OM	0.8 ± 0.5	3 ± 1	1.7 ± 0.4	1.8 ± 0.6	3 ± 1	19 ± 11
N-compounds : Carbohydrates	Algal:Plant OM	0.37 ± 0.09 (24 %)	0.32 ± 0.08	0.35 ± 0.04	0.6 ± 0.1	0.29 ± 0.02	0.23 ± 0.05
Chlorophylls : Plant lipids+lignin	Algal:Plant OM	0.18 ± 0.09 (50 %)	0.10 ± 0.05	0.13 ± 0.06	0.31 ± 0.07	0.18 ± 0.05	0.03 ± 0.03
Proteins:(alkyl)pyrroles+	Algal OM (N-compounds) freshness	0.39 ± 0.09	0.36 ± 0.05	0.22 ± 0.06	0.42 ± 0.06	0.20 ± 0.08	0.13 ± 0.08
(alkyl)pyridines+Aromatic N							
Phytadienes:pristenes ^c							
Anhydrosugars:(alkyl)furans & furanones	Plant OM (carbohydrates) freshness	0.4 ± 0.1 (25 %)	0.4 ± 0.1	0.37 ± 0.09	0.40 ± 0.06	0.56 ± 0.05	0.42 ± 0.07
Guaiacyl-acid:Guaiacyl-aldehyde ^c	Plant OM (lignin) freshness	0.2 ± 0.2 (100 %)	0.4 ± 0.2	0.3 ± 0.2	0.12 ± 0.11	0.14 ± 0.04	0.08 ± 0.01
Guaiacyl -2C: Guaiacyl -1C ^c	Plant OM (lignin) freshness	0.07 ± 0.03 (43 %)	0.13 ± 0.02	0.07 ± 0.03	0.05 ± 0.02	0.04 ± 0.01	0.04 ± 0.03
Syringyl-2C:Syringyl-1C ^c	Plant OM (lignin) freshness	0.8 ± 0.3 (38 %)	1.23 ± 0.07	1.0 ± 0.2	0.5 ± 0.2	0.6 ± 0.2	0.5 ± 0.1
		1.0 ± 0.8 (80 %)	2.4 ± 0.3	1.1 ± 0.6	0.5 ± 0.2	0.6 ± 0.1	0.3 ± 0.3

^a whole-lake: averages of all analyzed sediment samples excluding the two outlier samples (sites M4, S15; cf. Sect. 3.1.1); ^b n: number of samples; ^c the data are presented as follow: **average** ± standard deviation (*relative standard deviation*); ^d the compounds included in the ratios are given in detail in Table 2 and S1 in the SI.

Light grey background denotes average values below whole-lake average (<10 %); No background denotes values close to whole-lake average (±10 %); Dark grey background are values above whole-lake average (>10 %).

Figures

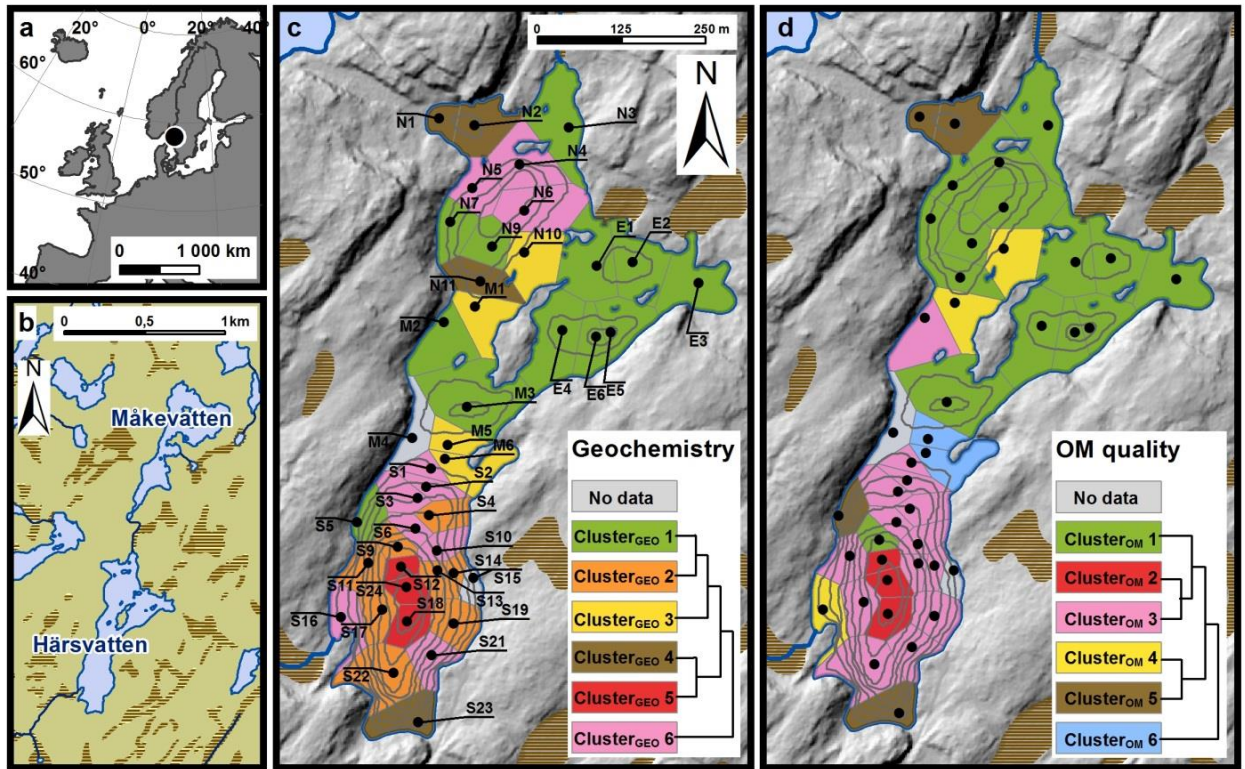


Fig. 1 Maps of Härsvatten showing (a) its location in Europe; (b) its catchment with lakes, mires and larger streams; and (c, d) its bathymetry along with the spatial distribution of the 44 sampling sites and the six selected clusters based on sediment elemental geochemistry (c) and sediment OM molecular composition (d). In the panel c) and d), the dendrogram shows the relationship between the six identified clusters.

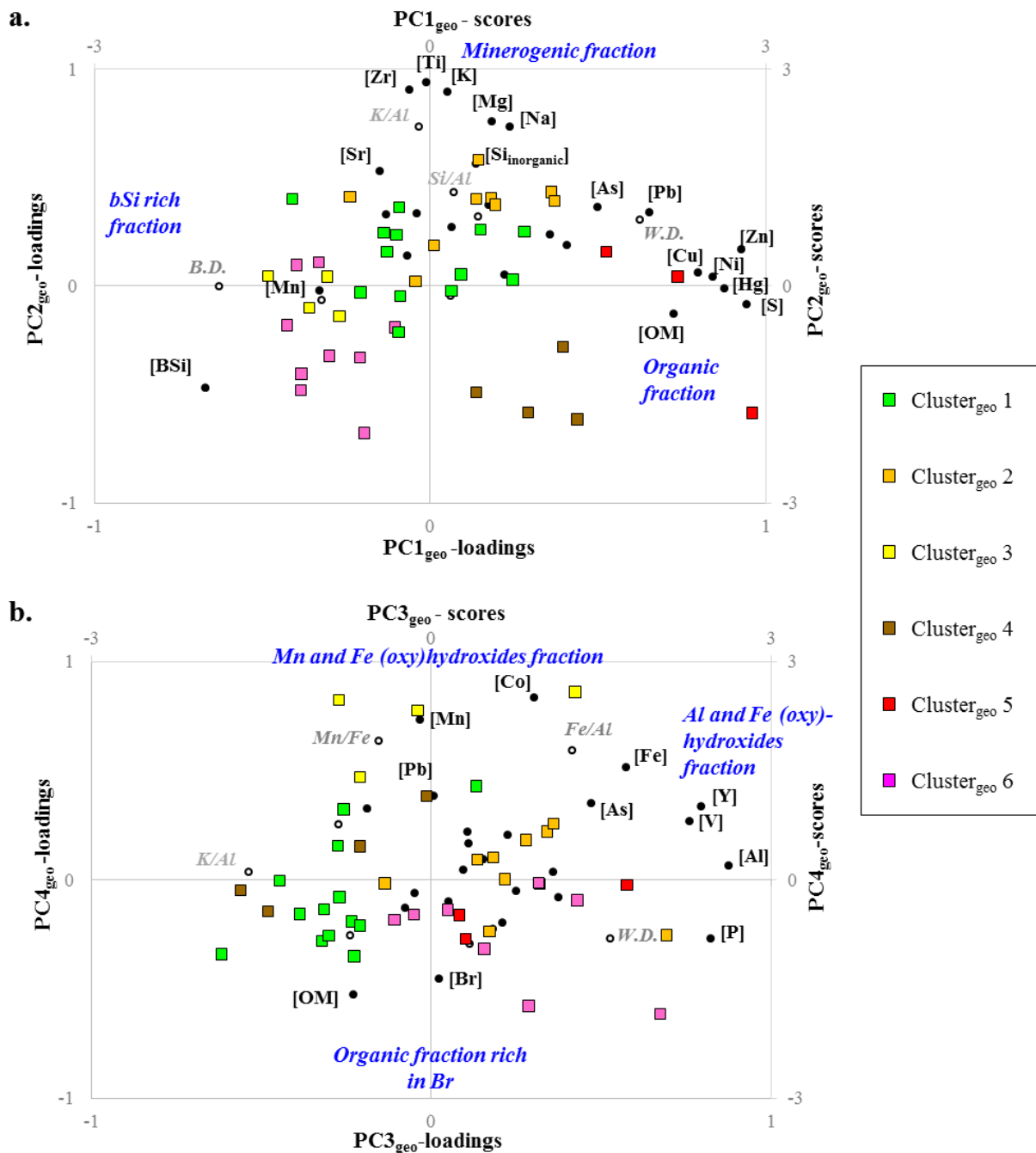


Fig. 2 Combined loading- and score-plots for PCs 1-4 of the elemental geochemistry dataset. For the PC-loadings, filled circles correspond to active variables. Others variables (empty circle and italics letter) were added passively. Sediment samples are colored according to the results of the cluster analysis.

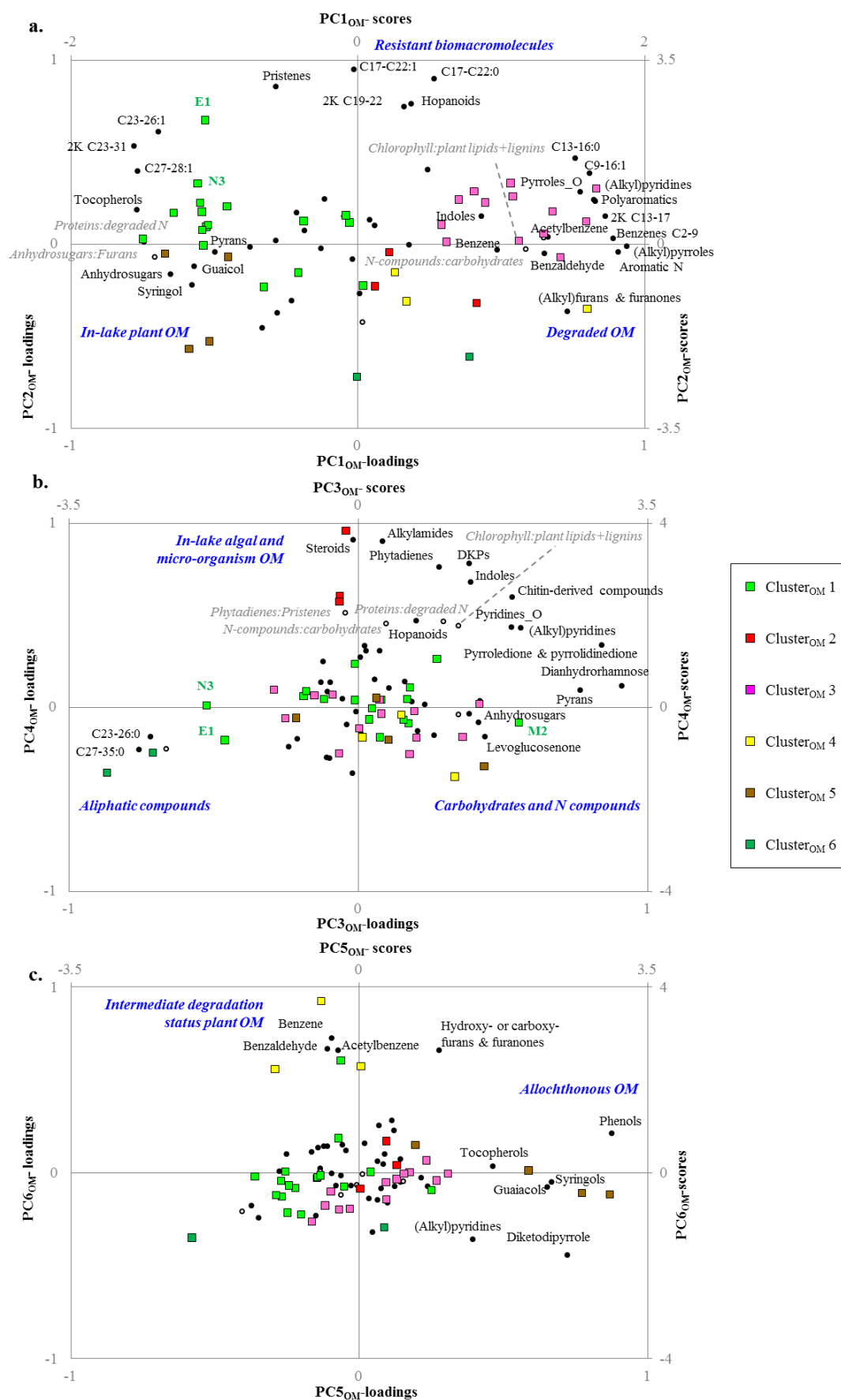


Fig. 3 Combined loading- and score-plots for PCs 1-6 (a, b and c) of the OM molecular composition dataset (i.e. the 41 groups of organic compounds as defined in Table 2). For the PC-loadings, filled circles correspond to active variables. Others variables (empty circle and italics letter) were added passively. Sediment samples are colored according to the results of the cluster analysis.

