Reviews and syntheses: Four Decades of Modeling Methane Cycling in Terrestrial Ecosystems

Xiaofeng Xu^{1, 2, 3}, Fengming Yuan⁴, Paul J. Hanson⁴, Stan D.Wullschleger⁴ Peter E. Thornton⁴, William J. Riley⁵, Xia Song^{1, 3}, David E. Graham⁶, Changchun Song², and Hangin Tian⁷

- 1. Biology Department, San Diego State University, San Diego, CA, USA 5 2. Northeast Institute of Geography and Agro-ecology, Chinese Academy of Sciences, Changchun, Jilin, China 3. Department of Biological Sciences, University of Texas at El Paso, El Paso, TX, USA 4. Climate Change Science Institute and Environmental Sciences Division, Oak Ridge National
- Laboratory, Oak Ridge, TN, USA 10
 - 5. Earth Sciences Division, Lawrence Berkeley National Laboratory, Berkeley, CA, USA
 - 6. Biosciences Division, Oak Ridge National Laboratory, Oak Ridge, TN, USA
 - 7. International Center for Climate and Global Change Research, School of Forestry and Wildlife Sciences, Auburn University, Auburn, AL, USA.

Correspondence to: Xiaofeng Xu (xxu@mail.sdsu.edu) 15

Abstract

Over the past four decades, a number of numerical models have been developed to quantify the magnitude, investigate the spatial and temporal variations, and understand the underlying mechanisms and environmental controls of methane (CH₄) fluxes within terrestrial ecosystems. These CH₄ models are also used for integrating multi-scale CH₄ data, such as laboratory-based incubation and molecular 20 analysis, field observational experiments, remote sensing, and aircraft-based measurements across a variety of terrestrial ecosystems. Here we summarize 40 terrestrial CH₄ models to characterize their strengths and weaknesses and to suggest a roadmap for future model improvement and application. Our key findings are that: (1) the focus of CH₄ models has shifted from theoretical to site- and regional-level applications over the past four decades, (2) large discrepancies exist among models in terms of representing CH₄ processes and their environmental controls, and (3) significant data-model and modelmodel mismatches are partially attributed to different representations of landscape characterization and inundation dynamics. Three areas for future improvements and applications of terrestrial CH₄ models are: (1) CH₄ models should more explicitly represent the mechanisms underlying land-atmosphere CH₄

30 exchange, with an emphasis on improving and validating individual CH₄ processes over depth and horizontal space, (2) models should be developed that are capable of simulating CH₄ emissions across highly heterogeneous spatial and temporal scales, particularly hot moments and hot spots, and (3) efforts should be invested to develop model benchmarking frameworks that can easily be used for model improvement, evaluation, and integration with data from molecular to global scales. These 35 improvements on CH₄ models would be beneficial for the Earth system models and further simulation of climate-carbon cycle feedbacks.

1. Introduction

- Methane (CH₄) is the second most important anthropogenic greenhouse gas, accounting for ~15% of anthropogenic forcing to climate change (Forster et al., 2007; IPCC, 2013; Rodhe, 1990). Therefore, an accurate estimate of CH₄ exchange between land and the atmosphere is fundamental for understanding climate change (Bridgham et al., 2013; Nazaries et al., 2013; Spahni et al., 2011). The ecosystem modeling approach has been one of the most broadly used integrative tools for examining mechanistic processes, quantifying the budget of CH₄ flux across spatial and temporal scales (Arah and Kirk, 2000; Arah and Stephen, 1998; Cao et al., 1995; Curry, 2007; Fung et al., 1991; Huang et al., 1998b; Nouchi et al., 1994; Potter, 1997; Rilev et al., 2011; Walter et al., 1996; Xu et al., 2007; Zhuang
- 1998b; Nouchi et al., 1994; Potter, 1997; Riley et al., 2011; Walter et al., 1996; Xu et al., 2007; Zhuang et al., 2004), and predicting future flux (Anisimov, 2007). Specifically, many CH₄ models have been developed to integrate data, improve process understanding, quantify budgets, and project exchange with the atmosphere under a changing climate (Cao et al., 1995; Grant, 1998; Huang et al., 1998a;
 Potter, 1997; Riley et al., 2011; Tian et al., 2010; Zhuang et al., 2004). In addition, model sensitivity analyses help to design field and laboratory experiments by identifying the most uncertain processes and parameters in the models (Massman et al., 1997; Xu, 2010).

55

Based on the complexity of the CH₄ processes represented, CH₄ models fall into two broad categories: (1) empirical models that are used to estimate and extrapolate measured methanogenesis, methanotrophy, or CH₄ emission at plot, country, or continental scales (Christensen et al., 1996; Eliseev et al., 2008; Mokhov et al., 2007; Wania et al., 2010, 2009); and (2) process-based models that are used

for prognostic understanding of individual CH₄ processes in response to multiple environmental drivers and budget quantification (reviewed below). This separation emphasizes the high-level model structure rather than the specific processes represented, therefore, models with many processes represented with empirical functions are still classified as process-based models if they represent many key processes of CH₄ production, oxidation, and transport. Although this separation is rather arbitrary, it helps understand the characteristics and purpose of models in a systems perspective.

60

Over the past decades, many empirical and process-based models have been developed, for example CASA (Potter, 1997), CH4MOD (Huang et al., 1998b), CLM4Me (Riley et al., 2011), DAYCENT (Del Grosso et al., 2000), DLEM (Tian et al., 2010; Xu and Tian, 2012), DNDC (Li, 65 2000a), ecosys (Grant, 1998), HH (Cresto-Aleina et al., 2015), MEM (Cao et al., 1995), TEM (Zhuang et al., 2004), etc. However, recent analyses and model inter-comparisons have shown that most of these models poorly reproduce regional- to global-scale observations (Bohn and Lettenmaier, 2010; Bohn et al., 2015; Melton et al., 2013; Wania et al., 2013). A comprehensive synthesis and evaluation of the 70 mechanisms incorporated into these models is lacking. This review focuses on primary processes of CH₄ cycling in the terrestrial ecosystems and their representation in the models. The critical CH₄ processes include substrate cycling, methanogenesis, methanotrophy, and transport in the soil profile, and their environmental controls. Emphasis is given to how these mechanisms were simulated in various models and how they were categorized in terms of complexity and ecosystem function. The review focuses on CH₄ models developed for terrestrial ecosystems, which is defined as ecosystems on 75 land and wetlands with less than 2 m standing water. This classification is used to distinguish from pure aquatic ecosystems and considering the important role of wetlands on CH₄ cycling. Therefore, models for understanding reactions in bioreactors (Bhadra et al., 1984; Pareek et al., 1999), mining plots (De Visscher and Van Cleemput, 2003), aquatic ecosystems, and marine systems (Elliott et al., 2011) were 80 excluded. An early pioneering effort of multiplying wetland area by average CH₄ flux to estimate global CH₄ budget was excluded from this review as well (Matthews and Fungi, 1987). This review excludes the CH₄ emission from biomass burning, termites and ruminants, because this paper primarily focuses on soil biogeochemical processes represented in ecosystem models. The model names are determined

by two criteria: (1) if the model has been named in the original publication, it will be used to represent
the model; (2) if the model has not been named, the last name of the first author will be used to name
the model; for example, "Segers model", "Gong model". In this paper we first provide an overview of
the range of processes that have been considered in methane models over the past four decades, and
then further classify existing models as determined by the range of processes considered. We finished
with several suggested research topics, which would be beneficial for better developing and applying
CH₄ model for either understanding CH₄ cycling or quantifying CH₄ budget at various scales.

2. Primary CH₄ Processes

10

Biological methane production in sediments was first noted in the late 18th century (Volta 1777). and the microbial oxidation of methane was proposed at the beginning of the 20th century (Söhngen 1906). Since then, methane cycling processes have been intensively studied and documented 95 (Christensen et al., 1996; Hakemian and Rosenzweig, 2007; Lai, 2009; Melloh and Crill, 1996; Mer and Roger, 2001), and most have been described mathematically and incorporated into ecosystem models (Table 1). Herein, we do not attempt to review all CH₄ processes, as a number of reviews have been published on this topic (Barlett and Harriss, 1993; Blodau, 2002; Bridgham et al., 2013; Cai, 2012; Chen et al., 2012; Conrad, 1995; Conrad, 1996; Hakemian and Rosenzweig, 2007; Higgins et al., 1981; Lai, 2009; Monechi et al., 2007; Segers, 1998; Wahlen, 1993). Rather, we focus on primary CH₄ 00 processes in terrestrial ecosystems, and their environmental controls from a modeling perspective. In this context there exist three major methanogenesis mechanisms, two CH₄ methanotrophy mechanisms. and three aggregated CH₄ transport pathways in plants and soils. We note that most models do not explicitly represent all of these transport pathways, and that the relative importance of these pathways 05 varies substantially in time, space, and with ecosystem types. We also pay attention to several other modeling features including capability for plot- or regional-level simulations, vertical representation of biogeochemical processes, and whether the model is embedded in an Earth System Model (ESM).

The published literature concludes that two processes dominate biological CH_4 production (Conrad, 1999; Krüger et al., 2001): acetoclastic methanogenesis -- CH_4 production from acetate, and hydrogenotrophic methanogenesis - CH_4 production from hydrogen (H₂) and carbon dioxide (CO₂).

Acetoclastic and hydrogenotrophic methanogenesis account for ~50% - 90% and ~10% - 43% of global annual CH₄ produced, respectively (Conrad and Klose, 1999; Kotsyurbenko et al., 2004; Mer and Roger, 2001; Summons et al., 1998). Methylotrophic methanogenesis (producing CH₄ from methanol, methylamines, or dimethylsulfide) is usually considered a minor contributor of CH₄, but may be significant in marine systems (Summons et al., 1998). The proportion of CH₄ produced via any of these pathways varies widely in time, space, and across ecosystem types.

15

Methanotrophy occurs under aerobic (Gerard and Chanton 1993) and anaerobic (Smemo and Yavitt 2011) conditions. These oxidative processes can occur in several locations in soil and plants (Frenzel and Rudolph 1998, Heilman and Carlton 2001, Ström et al. 2005) and using CH₄ either produced in the soil column or transported from the atmosphere (Mau et al. 2013). Large variation in 20 the relative magnitudes of these pathways as a percentage of total methanotrophy has been observed: aerobic oxidation of CH₄ in soil contributes 1% - 90% (King, 1996; Ström et al., 2005), anaerobic oxidation of CH₄ within the soil profile contributes 0.3% - 5% (Blazewicz et al., 2012; Murase and Kimura, 1996), oxidation of CH₄ during transport in plant aerenchyma contributes <1% (Frenzel and Karofeld, 2000; Frenzel and Rudolph, 1998), and oxidation of atmospheric CH₄ contributes $\sim 10 - 100\%$ 25 (ranging from ~10% for wetland to ~100% for upland) (Gulledge and Schimel, 1998a; Gulledge and Schimel, 1998b; Topp and Pattey, 1997) to total methanotrophy in the ecosystem. CH₄ is transported from the soil profile to the atmosphere in typical open-water wetlands by seven pathways which could be aggregated into three: plant-mediated transport accounts for 12~98% (Butterbach-Bahl et al., 1997; 30 Mer and Roger, 2001; Morrissey and Livingston, 1992), diffusion accounts for ~5% for wetlands and > 90% for upland systems (Barber et al., 1988; Mer and Roger, 2001), and ebullition accounts for 10~60% (Chanton et al., 1989; Tokida et al., 2007) of the CH₄ produced in the soil that is emitted to the atmosphere. The plant-mediated transport includes diffusive and advective (associated with gas or liquid flow) transports, soil diffusion includes soil gaseous diffusion and advection and aqueous diffusion and advection.

35

Environmental factors affecting CH₄ processes have many direct and indirect controls. The dominant direct factors controlling methanogenesis and methanotrophy in most ecosystems include

oxygen availability, dissolved organic carbon concentration, soil pH, soil temperature, soil moisture, nitrate and other reducers, ferric iron, microbial community structure, active microbial biomass, wind

speed (Askaer et al. 2011), plant root structure (Nouchi et al. 1990), etc. Indirect factors include soil texture and mineralogy, vegetation, air temperature, soil fauna, nitrogen input, irrigation, agricultural practices, sulfate reduction, and carbon quality, etc. (Banger et al., 2012; Bridgham et al., 2013; Hanson and Hanson, 1996; Higgins et al., 1981; Mer and Roger, 2001). The complicated effects induced by a few key factors on CH₄ processes have been mathematically described and incorporated in many CH₄
models; for example, direct factors such as soil temperature, moisture, oxygen availability, soil pH, and soil redox potential (Grant, 1998; Riley et al., 2011; Tian et al., 2010; Zhuang et al., 2004). The indirect factors such as nitrogen input (Banger et al., 2012), irrigation (Wassmann et al., 2000), and agricultural practices were not reviewed in this study as their impacts are indirect and were modeled through impacts on vegetation and hydrology (Li, 2000a; Ren et al., 2011; Xu et al., 2010).

3. Model Representation of CH₄ Processes

[Insert Figure 1 here] [Insert Figure 2 here]

We reviewed 40 CH₄ models (Fig. 1 & Table 1), which were developed for a variety of purposes. The first CH₄ model was published in 1986 by Lovley & Klug (1986) to simulate *methanogenesis* in freshwater sediments, and since then a number of CH₄ models have been developed and applied at numerous scales (Table 1). For example, Cao et al. developed the Methane Emission Model (MEM) and applied it to quantify the global CH₄ source in rice paddies and the sensitivity of the global CH₄ budget's response to climate change (Cao et al., 1995; Cao et al., 1998). Grant et al (1998) developed the *ecosys* model, which is currently the ecosystem-scale model that most mechanistically
represents the many kinetic processes and microbial mechanisms for methanogenesis, methanotrophy, and CH₄ emission (Grant and Roulet, 2002). Riley et al (2011) developed CLM4Me, a CH₄ module for the Community Land Model, which is incorporated in the Community Earth System Model. The family of LPJ models (LPJ-Bern, LPJ-WHyMe, LPJ-WSL) was developed under the LPJ framework to simulate CH₄ processes, but with different modules for CH₄ cycling; for example, LPJ-Bern and LPJ-

- WHyMe incorporate Walter CH₄ module (Walter and Heimann, 2000; Walter et al., 1996; Wania et al., 65 2009) while LPJ-WSL incorporates the CH₄ module from Christensen et al. (Christensen et al., 1996). The number of CH₄ models has steadily increased since the 1980s (Figs. 1 & 2): 1 in the 1980s, 11 in the 1990s, 14 in the 2000s, and 14 for 2010-2015. This increase in model developments is driven by many factors, including a desire to understand the contribution of CH₄ processes to regional CH₄ budget (Fig. 1). For instance, the Lovley's model was built to understand the CH₄ production and 70 sulfate reduction in freshwater sediment (Lovley and Klug, 1986); while all models published in the 2010s are applicable for CH₄ budget quantification, particularly at regional scale. This rapid increase in CH₄ model development indicates a growing effort to analyze CH₄ cycling and quantify CH₄ budgets across spatial scales. Meanwhile, the key mechanisms represented in the models have increased at a slower pace (Fig. 2). The most important changes are representation of vertically-resolved processes 75 within the soil and regional model simulation. For example, the percentage of the newly developed models with vertically-resolved CH₄ biogeochemistry has increased from 54% before 2000 to ~79% in the recent decade (2010-2015). The proportion of models with regional simulation capability
- 80 \sim 50% before the 2010s to almost 100% afterwards (Fig. 2).

[Insert Tables 1, 2, and 3 here]

(producing spatial map of CH₄ fluxes with inputs of spatial map of driving forces) has doubled from

The majority of these models were designed to simulate land-surface exchange in saturated ecosystems (primarily natural wetlands and rice paddies) (Huang et al., 1998b; Li, 2000a; Walter et al., 1996) (Table 1). Not all of the models explicitly represented the belowground mechanistic processes for CH₄ production and consumption and the primary carbon biogeochemical processes (Christensen et al., 1996; Ding and Wang, 1996). The land-atmosphere CH₄ exchange is a net balance of many processes including production, oxidation, and transport, which are represented in models with different complexities (Table 2). Some models are quite complicated, while some are relatively simple. The obvious tradeoff in modeling CH₄ cycling is to represent mechanisms as accurately as possible while 90 managing complexity (Evans et al., 2013), and ensuring that additional complexity enhances predictability (Tang and Zhuang, 2008).

85

[Insert Figure 3 here]

[Insert Figure 4 here]

95 Based on a cluster analysis that considers model characteristics including acetoclastic methanogenesis, hydrogentrophic methanogenesis, methanotrophy, different CH_4 transport pathways, multiple soil layer, oxygen availability, current CH₄ models can be classified into three groups (Fig. 3 & 4). The first group of CH₄ models uses a very simple framework for land-surface CH₄ flux, and most were developed before the 2000s (e.g., Christensen's model, CASA, etc.) (Fig 4A). These models 00 treated land-surface CH₄ flux as an empirical function and link it to environmental controls, or soil organic carbon. This group of models ignored the mechanistic processes of methanogenesis, methanotrophy, and CH₄ transport. The second group of CH₄ models considers processes in a relatively simple manner (e.g., one or two primary CH₄ transport pathways, methanogenesis as a function of DOC, oxidation of atmospheric CH₄, etc.); however, the methanogenesis and methanotrophy 05 mechanisms are still not mechanistically represented (Fig. 4B). For example, DLEM simulate CH₄ production with a Michaelis-Menten equation with DOC concentration as substrate (Tian et al., 2010); Walter's model simulates CH_4 production with a simple multiplier between substrate availability and environmental scalars and CH₄ oxidation with a Michaelis-Menten equation (Walter et al., 1996). The third group of CH₄ models explicitly simulates the processes for methanogenesis, methanotrophy, and 10 CH₄ transport as well as their environmental controls, which allows comprehensive investigation of physical, chemical, or biological processes' contribution to land-surface CH₄ flux (Fig. 4C). Of the models in the third group, none of them fully represent all these processes (although some have most of the features described); for example, the ecosys model is one of the few models to represent most of the CH₄ cycling processes shown in Fig. 4C, although it has not been embedded in an Earth System Model.

15 **3.2. Methanogenesis**

Models make use of four types of modeling frameworks (Table 3) to relate methanogenesis to substrate requirements. Similar to Eqs (1) - (4) in Table 3, there are four model algorithms to represent methanogenesis: (1) empirical association between methanogenesis and environmental condition,

including temperature and water table: (2) empirical correlation of methanogenesis with biological 20 variables (particularly heterotrophic respiration and soil organic matter); (3) methanogenesis as a function of concentration of substrate (DOC); and (4) a suite of mechanistic processes simulated for methanogenesis.

Representation of the substrate for methanogenesis may be a key aspect of simulating CH₄ cycling in terrestrial ecosystems (Bellisario et al., 1999); however, more than half of the models 25 examined do not explicitly simulate substrates for methanogenesis. We note, however, that explicit representation of substrates and their effects on methanogenesis requires additional model parameters. and therefore degrees of freedom in the model, which can lead to increased equifinality (Tang and Zhuang, 2008). The optimum complexity level for methanogenesis and consumption models remains to be determined.

30 The first model algorithm correlates methanogenesis with environmental factors and ignores substrate production and its influence on methanogenesis [Eq. (1)] (Table 3). This group includes Christensen's model (Christensen et al., 1996), which simulates the net flux of CH₄ based on fraction of saturated soil column and soil temperature, and the IAP-RAS model (Mokhov et al., 2007), which calculates methanogenesis as an empirical equation of soil temperature. This group has a role in sitespecific interpolation of observations for scaling over time at a given site, but does not explicitly 35 represent carbon or acetate substrate. The second model algorithm directly links methanogenesis with heterotrophic respiration or soil organic matter content, but does not explicitly represent carbon or acetate substrate availability [Eq. (2)]; examples are the LPJ model family (Hodson et al., 2011; Spahni et al., 2011; Wania et al., 2010, 2009) and CLM4Me (Riley et al., 2011). The third model algorithm 40 simulates dissolved organic carbon (DOC) or different pools of soil organic carbon, which are treated as a substrate pool influencing CH₄ production [Eq. (3)]; examples are the MEM model (Cao et al., 1995; Cao et al., 1998) and DLEM (Tian et al., 2010). The fourth model algorithm considers the primary substrates for methanogenesis, that is, acetate and single-carbon compounds [Eq. (4)]; examples are Kettunen's model (Kettunen, 2003), Segers' model (Segers and Kengen, 1998; Segers and Leffelaar,

45 2001a, b; Segers et al., 2001), van Bodegom's model (van Bodegom et al., 2000; van Bodegom et al., 2001), and the ecosys model (Grant, 1998).

Methanogenesis is a fundamental process for CH₄ cycling, and a majority of models simulate methanogenesis in either implicit or explicit ways (Tables 2 & 3). For example, 32 models (i.e. Cartoon model, CASA, CH4MOD, Christensen model, CLM4Me, Ding model, DLEM, DNDC, DOS-TEM, ecosys, Gong model, HH model, IAP-RAS, Kettunen model, Lovley model, LPJ-Brn, LPJ-WHyMe, 50 LPJ-WSL, Martens model, MEM, MERES, ORCHIDEE, SDGVM, Segers model, TCF, TEM, TRIPLEX-GHG, UW-VIC, van Bodegom model, VISIT, Walter model, and Xu model) simulate methanogenesis as one individual process. As a comparison, only three out of 40 CH₄ models reviewed explicitly simulate two methanogenesis pathways (acetoclastic methanogenesis and hydrogenotrophic 55 methanogenesis) (Table 3). As mentioned earlier, it is well-recognized that there are two dominant methanogenesis pathways and their relative combination changes significantly across environmental gradients, for example, along the soil profile (Falz et al., 1999) and across landscape types (McCalley et al., 2014). This lack of representation of two methanogenesis mechanisms might have caused dramatic bias in simulating CH₄ flux temporally and spatially and needs to be addressed in future model improvements.

60

65

Michaelis-Menten-like equations, widely used for simulating CH₄ production and oxidation. consider substrates limiting factors (Segers and Kengen, 1998). A few CH₄ models in the third category of methanogenesis models (linking methanogenesis with a substrate) use the Michaelis-Menten-like equation to compute methanogenesis and methanotrophy rates (Eqs. 3, 5, & 6). For example, DLEM simulates methanogenesis as a function of DOC concentration and other environmental controls, and Michaelis-Menten-like functions were used to compute methanogenesis on the basis of DOC as substrate.

3.3. Methanotrophy

Methanotrophy is another important process for simulating the land-atmosphere exchange of 70 CH₄ (Table 2). Aerobic and anaerobic methanotrophy occurs in different locations in the soil profile, and affect both methanogenesis in the profile and CH₄ diffusing in from the atmosphere. For example,

the oxidation of atmospheric CH₄, rhizosphere and bulk soil oxidation, and oxidation during CH₄ transport from soil to the atmosphere have been measured and modeled (Tables 1 & 2). Anaerobic CH₄ oxidation has been measured (Blazewicz et al., 2012) and has been proposed to be incorporated into ecosystem models (Gauthier et al., 2015).

75

80

It has been confirmed that the aerobic oxidation of CH₄ produced in the soil profile and aerobic oxidation of atmospheric CH₄ play a major role in CH₄ consumption in the system, and that anaerobic oxidation of CH₄ is a minor contributor. Currently, no models explicitly simulate the anaerobic oxidation of CH₄ in soil, although a few recent studies highlighted the importance of this process (Blazewicz et al., 2012; Caldwell et al., 2008; Conrad, 2009; Smemo and Yavitt, 2011; Valentine and Reeburgh, 2000). The key reasons for this omission are that the process has not been mathematically described, the key parameters are uncertain (Gauthier et al., 2015), and the biochemical mechanism is not fully understood.

Methanotrophy has been simulated with dual Monod Michaelis-Menten-like equations with CH₄ 85 and oxygen as limiting factors (Table 3). Recent work has shown that the Michaelis-Menten approach may be inaccurate when representing multi-substrate, multi-consumer networks, and that a new approach (called Equilibrium Chemistry Approximation, ECA) can ameliorate this problem (Tang and Riley 2013, 2015; Zhu et al., 2016). Although the ECA approach has not been applied for simulations of CH₄ emissions, CH₄ dynamics are inherently multi-consumer, including transformations associated with methanogens, heterotrophs, ebullition, advection, diffusion, and aerenchyma transport, even if only 90 one substrate is considered.

3.4. CH₄ within the Soil/Water Profile

95

CH₄ produced in the soil profile or below the water table is not transported immediately into the atmosphere. The time required for CH₄ to migrate from deep soil profile to the atmosphere ranges from minutes to days (depending on temperature, water, soil texture, and emissivity of plant roots), or even a season if the surface is frozen. The majority of current CH₄ models assume that CH₄ transport to the atmosphere occurs immediately after CH₄ is produced, and a portion is oxidized (Tian et al., 2010; Fan

et al., 2013); for models simulating CH₄ flux over minutes to days, the lack of modeled transport may produce unrealistic simulations.

00

Some models do simulate CH₄ dynamics within the soil and water profile (e.g., *ecosys*, CLM4Me), which produces a lag between methanogenesis and emission, allowing for oxidation to be explicitly represented during transport, and is valuable for simulating the seasonality of CH₄ flux (Table 2). For example, the recently observed CH₄ burst in the spring season in some field experiments confirms that the storage of CH₄ produced in winter can produce a strong emission outburst (Song et al., 2012). Without understanding the mechanism of CH₄ storage beneath the soil surface, this phenomenon will be difficult to simulate. In most of the models considering CH₄ storage, the CH₄ is treated as a simple gas pool, under the water table, which will be transported to the atmosphere through several transport pathways.

3.5. CH₄ Transport from Soil to the Atmosphere

10

05

The transport of CH_4 produced and stored in soil column is the bottleneck for CH_4 leaving the system; therefore, this process is an important control on the instantaneous land-surface CH_4 flux. Several important pathways of CH_4 transport to the atmosphere are identified: plant-mediated diffusive and advective transport, aqueous and gaseous diffusion, and ebullition (Beckett et al., 2001; Chanton, 2005; Mer and Roger, 2001; Whiting and Chanton, 1996). Model simulation of these transport pathways uses direct control of simulated land surface CH_4 flux, with CH_4 transport simulation considered in a manner similar to Eq. (7) (Table 3).

15

The majority (83%) of the current models simulate at least one transport pathway. Specifically, 70% of the models simulate CH₄ transport via aerenchyma, 80% simulate gaseous diffusive transport, and 60% simulate ebullition transport (Table 1). More than 50% of models simulated these three transport pathways. Some models simulate explicitly the aqueous and gaseous diffusion of CH₄ (Riley et al., 2011), while most models do not simulate advective transport. Many models simulate diffusion and plant-mediated transport in very simple ways. For model improvement in this area, three issues remain as challenges:

20

(1) Most models treat transport implicitly; for example, the diffusion processes is treated simply as an excessive release of CH_4 when its concentration exceeds a threshold (Tian et al., 2010). This treatment prevents the model from simulating the lag between methanogenesis and its final release to the atmosphere, which has been confirmed to be the key mechanism for hot-moment and hot-spot of CH_4 flux (Song et al., 2012) and for oxidation during transport.

30

35

25

(2) The parameters for plant species capable of transporting gas (i.e., *aerenchyma*) are poorly constrained (Riley et al. 2011), although plant-mediated transport has been identified as the dominant pathway for CH_4 emission in some natural wetlands (Aulakh et al., 2000; Colmer, 2003).

(3) Simultaneously representing aqueous and gaseous phases of CH_4 is one potentially important issue for simulating CH_4 transport from soil to the atmosphere (Tang and Riley, 2014). However, these processes are only explicitly represented in a few extant CH_4 models (Riley et al., 2011; Grant et al., 1998).

3.6. Environmental Controls on CH₄ Processes

Although a suite of environmental factors affects various CH₄ processes, many of these factors are not explicitly simulated in many models. These factors include soil temperature, soil moisture, substrate, soil pH, soil redox potential, and oxygen availability. Many other factors not directly incorporated in the models, could indirectly affect CH₄ cycling. For example, nitrogen fertilizer affects methanogenesis through its stimulating impacts on ecosystem productivity, which in turn affects DOC, soil moisture and soil temperature (Xu et al., 2010). The CLM4Me model simulates permafrost and its effects on CH₄ dynamics, and has a simple relationship for soil pH impacts on methanogenesis (Riley et al., 2011). Wania et al. (2013) reviewed a number of active CH₄ models for their representation of CH₄ production area. In this review, we specifically focus on temperature, moisture, and pH because these factors directly affect CH₄ processes in all environments, and they have been explicitly simulated in the many of the models.

Three types of mathematical functions have been used to simulate the temperature dependence 50 of CH₄ processes: (1) linear functions of air or soil temperature (Eq. 9 in Table 3), (2) Q₁₀ function (Eq. 10 in Table 4), and (3) Arrhenius type function (Eq. 11 in Table 3). Of these three model representations of temperature dependence, the Q_{10} equation is the most common mathematical description. However, the parameters for these empirical functions vary widely across the models (Table 4). Actual temperature responses may diverge significantly from the models at low temperatures, close to the freezing point of water, and high temperatures, close to the denaturation point of enzymes.

[Insert Table 4 here]

Soil moisture is another important factor controlling CH_4 processes, because water limits O_2 diffusion from the air through the soil column and because microbes can become stressed at low matric potential. CH_4 is produced typically under conditions with a low reduction potential, which is normally associated with long-term inundation. Although methanogenesis occurs solely under reducing conditions (methanogenesis within plant biomass under aerobic condition has never been simulated although it has been reported in experiments (Keppler et al., 2006)), methanotrophy occurs under drier, aerobic conditions. A low water content can also limit microbial activity in frozen soils or soils with high osmolarity (Watanabe and Ito, 2008). Therefore, soil moisture has different impacts on different CH_4 processes. Four types of model representation are used to simulate moisture's effects on CH_4 processes (Eqs. 13-16 in Table 3).

- Methanogenesis occurs only in the saturated zone and an exponential function for soil moisture is used to control methanotrophy (e.g., CLM4Me);
- (2) Linear function for moisture impacts (e.g., CLASS use linear function for moisture impact on methanotrophy) (Curry, 2007);
- 70

55

60

- (3) Reciprocal responsive curves for moisture impacts on methanogenesis and methanotrophy (e.g., DLEM) (Tian et al., 2010);
- (4) A bell-shaped curve for methanogenesis (e.g., TEM uses a function similar to Eq. (16) for moisture impacts) (Zhuang et al., 2004).
- Soil pH is another important factor that has been included in a number of CH₄ models (Cao et al., 1995; Zhuang et al., 2004). Methanogens and methanotrophs depend on proton and sodium ion translocation for energy conservation, thus they are directly affected by pH. The pH impacts on CH₄

processes are simulated as a bell-shaped curve although the mathematical functions used to describe pH impacts are different (Eq. 17a, 17b, and 17c). Moreover, even when the same functions were used in

- different models, they were associated with different parameter values, indicating slightly different response functions; for example, the MEM model sets *pH_{min}* (minimum pH value for CH₄ processes being active), *pH_{opt}* (optimal pH value for CH₄ processes being most active), and *pH_{max}* (minimum pH value for CH₄ processes being active) values of 5.5, 7.5, and 9 (Cao et al., 1995). This set of parameter values was adopted in the TEM model (Zhuang et al., 2004), whereas the DLEM model uses values of 4, 7, and 10 (Tian et al., 2010). The CLM4Me model uses a different function while keeping the impact curve at the same shape, but its peak has an optimal pH of 6.2 (Meng et al., 2012). It should be noted that while pH has been confirmed to significantly affect CH₄ production (Xu et al., 2015), the
 - simulation of pH dynamics caused by organic acid in soils remains a key challenge for the incorporation of this phenomenon.
 - 90 For the other environmental factors, model representation is still in its infancy; however, several models consider oxygen availability as an electron acceptor for methanotrophy (e.g., Beckett model, Cartoon model, CLM4Me, *ecosys*, Kettunen model, MERES, Segers model, van Bodegom model, De Visscher model, and Xu model). In addition, only a few models simulate the impacts of the electron acceptor (i.e. nitrate, sulfate, etc.) on CH₄ processes (Table 2). For example, the van Bodegom model all simulates iron biogeochemistry, and the Lovley model, Marten model, and van Bodegom model all simulate sulfate as the electron acceptor and its impacts on methanogenesis and methanotrophy (Lovley and Klug, 1986; Martens et al., 1998; van Bodegom et al., 2001). Explicitly representing these processes enables future coupling of CH₄ cycling to processes that are regionally significant, such as iron reduction on the Alaskan North Slope (Miller et al., 2015). These models have the potential advantage of more accurately simulating biogeochemical processes of carbon and ions, although large uncertainties still exist because of the lack of data for constraining model parameters.

3.7. CH₄ implementation in ESMs

The importance of CH₄ flux in simulating climate dynamics has been well recognized (IPCC 2013; Ringeval et al., 2011); yet few ESMs have implemented a CH₄ module (Ringeval et al., 2011;

Riley et al., 2011; Xu et al., 2014; Hopcroft et al., 2011; Eliseev et al., 2008). While these models have 05 been claimed to be coupled within ESMs, truly fully coupled simulations within ESMs to evaluate CH₄ dynamic impacts on global climate system are rare (Eliseev et al., 2008; Hopcroft et al., 2011). For example, the SDGVM has been coupled within Fast Met Office UK Universities Simulator (FAMOUS), a coupled general circulation model, to study the association between terrestrial CH₄ fluxes with rapid climate fluctuation during the last glacial period (Hopcroft et al., 2011). IAP-RAP 10 model was used to simulate terrestrial CH₄ flux and its contributions to atmospheric CH₄ concentrations and further on climate change. The quasi-coupling between ORCHIDEE WET with an oceanatmosphere general circulation model was used to theoretically evaluate terrestrial CH₄ dynamics on climate system (Ringeval et al., 2011). The CLM application within CESM framework has both CLM4Me and CLM-Microbe module for CH₄ dynamics, but none of them have been applied for a fully 15 coupled simulation to evaluate CH₄-climate feedback. It should be a key research effort for CLM community in next five years to complete this coupling. All previous coupled ESM simulations have concluded that changes in terrestrial CH₄ flux has small impacts on climate change, while they also pointed out that large uncertainties exist. Given the importance of CH₄ as a greenhouse gas and 20 uncertainties in current ESMs in simulating permafrost carbon and CH₄ flux, more efforts should be invested to implement CH₄ module in ESMs and further evaluate the CH₄-climate feedback under different climate scenarios.

3.8. Summary

Through the four decades of modeling CH₄ cycling in terrestrial ecosystems, consensus has been reached on several fronts. First, CH₄ cycling includes a suite of complicated processes, and both the simple and complex models are able to estimate land-surface CH₄ flux to a certain level of confidence, although models of different complexity do provide different results (Tang et al., 2010). Second, although a number of CH₄ models have been developed, several gaps remain that need new model representations (e.g., dynamic linkage between inundation dynamics and the CH₄ module (Melton et al., 2013), anaerobic oxidation of CH₄ (Gauthier et al., 2015)).

Two recent CH₄ model-model inter-comparison projects raised several important points (Bohn et al., 2015; Melton et al., 2013): (1) the distribution of the inundation area is important for accurately simulating global CH₄ emissions, but was poorly represented in CH₄ models; (2) the modeled response of land-surface CH₄ emission to elevated CO₂ is likely biased as a number of global change factors were missing, which indicates the need for modeling with multiple global environmental factors; and (3) the need for comparison with high-frequency observational data is identified as an important task for future model-model inter-comparison. These lessons will be helpful for, and likely addressed during, model improvements and applications of more mechanistic CH₄ models.

Although the primary individual CH₄ processes have been studied and quantified at a certain 40 level of confidence, only a few modeling studies have reported these individual processes as previously discussed. For example three pathways of CH₄ transports were represented in Kettunen, 2003 and Walter et al., 1996, but none of those modeled results have been evaluated against observational results for those individual processes. One reason is that measurements rarely distinguish among individual processes; another reason is that the majority of CH₄ models do not explicitly represent all processes (Table 2). However, a number of studies report significant shifts in the processes contributing to the 45 surface CH₄ flux along environmental gradients or across biomes (Conrad, 2009; Krumholz et al., 1995; McCalley et al., 2014). Projecting CH₄ fluxes into future changing climate conditions requires not only accurate simulations of CH₄ processes, but also shifts among the various processes. In addition, CO₂ flux has been evaluated within the Earth System Modeling framework, but only a few studies have 50 evaluated the CH₄ flux and its contribution to climate dynamics. Given the much higher warming potential and relatively faster rate of increase of atmospheric CH₄, fully coupled simulations are needed to represent the feedbacks between terrestrial CH₄ exchanges and climate. We note that a few recent studies reported a relatively small climate warming-methane feedback from global wetlands and permafrost (Gao et al., 2013; Gedney et al., 2004; Riley et al., 2011). A fully mechanistic CH₄ model that accounts for all the important features is critically needed. In addition, a modeling framework to 55 integrate multiple sources of data, such as microbial community structure and functional activities,

ecosystem-level measurements, and global scale satellite measurements of gas concentration and flux is needed with these mechanistic CH₄ models.

4. Needs for Mechanistic Methane Models

- During the last few years, the scientific community has continued to improve and optimize models to better simulate methanogenesis, methanotrophy, CH₄ transport, and their environmental and biological controls (Xu et al., 2015; Zhu. Q. et al., 2014). A number of emerging tasks have been identified, and progress in these directions is expected. First, linking genomic data with large-scale CH₄ flux measurements will be an important, while challenging, task for the entire community; for example, some work has been carried out in this direction (De Haas et al., 2011; Larsen et al., 2012). An effort has been initialized to develop a new microbial functional group-based CH₄ model, which has the advantages of linking genomic information for each individual process with the four microbial functional groups (Xu et al., 2015). Second, data-data and model-model comparisons are another important effort for model comparison and improvement. One ongoing encouraging feature that all recently developed CH₄ models possess is the capability for regional simulations as well as the possibility to be run at the site level (Riley et al., 2011; Zhu. Q. et al., 2014).
- Third, microbial processes need to be considered for incorporation into ecosystem models for simulating carbon cycling and CH₄ processes (DeLong et al., 2011; Xu et al., 2014). Although a few models explicitly simulate the microbial mechanisms of CH₄ cycling (Arah and Stephen, 1998; Grant, 1998; Li, 2000a; Segers and Kengen, 1998), none of them have been used for regional- or global-scale estimation of microbial contributions to the CH₄ budget. A reasonable experimental design and a well-validated microbial functional group-based CH₄ model should be combined to enhance our capability to apply models to estimate a regional CH₄ budget and to investigate the combination of microbial and environmental contributions to the land surface CH₄ flux (DeLong et al., 2011). Fourth, incorporating
 well-validated CH₄ modules into Earth System Modeling frameworks will allow a fully coupled simulation that provides a holistic understanding of the CH₄ processes, with its connections to many other processes and mechanisms in the atmosphere. Several recently developed models fall in the framework of Earth System Models (Riley et al., 2011; Ringeval et al., 2010), which provide a
 - 18

foundation for this application in a relatively easy way. This effort will likely contribute not only to the

85 CH₄ modeling community, but also to the entire global change science community (Koven et al., 2011). The iron and sulfate biogeochemistry that has been implicitly simulated in a few models (Table 2), but was not included in any of the recently developed models because that effort will likely be achieved over the long term, owing to poor understanding of the mechanisms and the lack of observational data.

[Insert Figure 5 here]

90 Based on the above-mentioned needs and model features as well as the mechanisms for the CH₄ models, the next generation of CH₄ models will likely include several important features (Fig. 5). The models should (1) be embedded in an Earth System Model, (2) consider the vertical distribution of thermal, hydrological, and biogeochemical transport and processes, (3) represent mechanistic processes for microbial CH₄ production, consumption, and transport, and (4) support data assimilation and a model benchmarking system as auxiliary components.

95

5. Challenges for Developing Mechanistic CH₄ Models

Knowledge Gaps - Modeling CH₄ cycling is a dynamic process. As new mechanisms are identified the modeling community should ensure that the mechanisms are well studied and mathematically described, as has occurred over the past decades (Conrad, 1989; McCalley et al., 2014; Schütz et al., 1989; Xu et al., 2015). However, a number of knowledge gaps need to be filled before a 00 full modeling framework of CH₄ processes within terrestrial ecosystems can be achieved. The first gap is either confirmation or rejection of a few recently observed CH₄ mechanisms; these mechanisms need to be fully vetted before being considered for incorporation into a model. The first most well-known mechanism still under debate is aerobic CH₄ production within plant tissue (Beerling et al., 2008; 05 Keppler et al., 2006). Since its first report in 2006 (Keppler et al., 2006), a few studies have confirmed the mechanism in multiple plant species (Wang et al., 2007). While its existence in nature is still under debate (Dueck et al., 2007), this mechanism will likely not be incorporated into an ecosystem model before solid evidence is presented and consensus is reached. The second new mechanism is fungi as a microbial group carrying out CH₄ production (Lenhart et al., 2012). More field- or lab-based

- 10 experiments are needed to investigate this mechanism and its contribution to the global CH₄ budget,
 - 19

probably through a data model integration approach. Third, the aerobic production of methane from the cleavage of methylphosphonate has been demonstrated in marine systems (Karl et al., 2008), but the significance of this process in terrestrial systems is unknown. Forth, the large CH₄ emission from rivers and small ponds are still not fully understood (Holgerson and Raymond, 2016; Martinson et al., 2010), which will likely be a direction for future model improvement.

15

Another knowledge gap is the missing comprehensive understanding of spatial and temporal variations in CH₄ flux; particularly, the "hot spots" and "hot moments" of observed CH₄ flux are still not completely understood (Becker et al., 2008; Mastepanov et al., 2008; Song et al., 2012). The traditional static chamber method of measuring CH₄ emissions could underestimate the CH₄ flux because sparse sampling is unlikely to detect these foci or pulses of unusually high emissions. Better 20 methods are also needed to measure CH₄ cycling during the shoulder seasons in the Arctic and subarctic when fluxes may be most variable (Zona et al. 2016). These knowledge gaps are key hurdles for CH₄ model development efforts. No model has yet been tested for simulating hot spots or hot moments over large spatial or long temporal scales. However, the high range (usually of factor 1-10) of the observed CH₄ flux might cause regional budgets to vary substantially (Song et al., 2012); therefore, mechanistic 25 model representations of these mechanisms are highly needed.

30

Modeling Challenges - Better simulation of CH₄ cycling in terrestrial ecosystems requires improvement in the model structure to represent mechanistic CH₄ processes. First is the challenge to simulate the vertical profile of soil biogeochemical processes and validate such models with observational results. Although some models have a capability for vertical distribution of carbon and nitrogen (Koven et al., 2013; Tang et al. 2013; Mau et al., 2013), a better framework for CH₄ and extension to cover the majority of CH₄ models are needed. This vertical distribution of biogeochemistry is necessary for simulating the vertical distribution of CH₄ processes and CH₄ transport through the soil profile before reaching the atmosphere. A second challenge is incorporating tracer capability. Isotopic tracers (¹³C, ¹⁴C) have been widely used for quantifying the carbon flow and partitioning among 35 individual CH₄ processes (Conrad, 2005; Conrad and Claus, 2005), but for ecosystem models this capability has not been represented even though it is very important to understanding CH₄ processes

and integrating field observational data. A third challenge is to simulate microbial functional groups. Microbial processes are carried out by different functional groups of microbes (Lenhart et al., 2012;

40 McCalley et al., 2014). Therefore, model comparison with individual processes requires representing the microbial population sizes (or active biomass) for specific functional groups (Tveit et al., 2015). This goal has proved more difficult than representing plant functional types or traits in models, because not all microbial taxonomic groups have ecologically coherent functions (Philippot et al., 2010). A fourth challenge is to simulate the lateral transport of dissolved and particulate biogeochemical variables that are necessary to better simulate the storage and transport of CH₄ within heterogeneous landscapes (Weller et al., 1995). A fifth challenge is modeling CH₄ flux across spatial scales. Although a few studies have been used to demonstrate the approach for simulating CH₄ budget at plot scale and eddy covariance domain scale (Zhang et al., 2012), a mechanistic framework to link CH₄ processes at distinct scales is still lacking while highly valuable. Finally, a sixth challenge is accurate simulate and predict, and their impacts on CH₄ processes are challenging (Li et al., 2005).

Data Needs - First, a comprehensive dataset of field measurements of CH_4 fluxes across various landscape types is needed to effectively validate the CH_4 models. Although a number of datasets have been compiled (Aronson and Helliker, 2010; Chen et al., 2012; Liu and Greaver, 2009; Mosier et al., 1997; Yvon-Durocher et al., 2014), some landscape types are still not fully covered. Meanwhile, highfrequency field observational data are also needed, particularly long-term observational data in some less-studied ecosystems; for example Arctic tundra ecosystems have been considered as an important contributor to global CH_4 budget in the changing climate (IPCC, 2013; Koven et al., 2011), however, long-term dataset of CH_4 flux is lacking. It is well-known that inter-annual variation of climate may turn an ecosystem from a CH_4 sink to a CH_4 source (Nauta et al., 2015; Shoemaker et al., 2014); therefore, a long-term observational dataset that covers these temporal shifts in CH_4 flux and its

60

55

21

associated ecosystem information would improve our understanding of the processes and our

representation of them in CH₄ models. Second, microbial community shifts and their role in CH₄

processes are important, although information is incomplete for model representation of this mechanism

- 65 (McCalley et al., 2014; Schimel and Gulledge, 1998). Although a number of studies have reported the microbial community structure and its potential association with changes in CH₄ processes (Monday et al., 2014; Schimel, 1995; Wagner et al., 2005), none of this progress has been documented in a mathematical manner suitable for a modeling representation.
- Third, a comprehensive dataset of all primary CH₄ processes within an individual ecosystem would be valuable for model optimization and validation. Although some datasets exist, no study has investigated all primary individual CH₄ processes within the same plot over the long term. Given the substantial spatial heterogeneity of CH₄ processes, this lack of process representation may cause bias in CH₄ simulations at regional scale. It should be noted that land surface net CH₄ flux is a measurable ecosystem-level process, whereas many individual CH₄ processes are difficult to accurately measure. Therefore, designing field- or lab-based-experiments suitable for measuring these processes is a fundamental need. For example, the anaerobic oxidation of CH₄ has been identified as a critical process for some ecosystem types, but no comprehensive dataset on it is available for model development or improvement.
- Last but not least, high quality spatial data as driving forces and validation data for CH₄ models 80 are critical for model development as well (Melton et al., 2013; Wania et al., 2013). Spatial distribution and dynamics of wetland area probably are the most important data need for CH₄ models (Wania et al., 2013). Spatial distribution of soil temperature, moisture, and texture are fundamental information because they serve as direct or indirectly environmental control on CH₄ processes. Recently launched Soil Moisture Active Passive (SMAP) satellite could be used as an important data source for soil moisture for driving CH₄ model (Entekhabi et al., 2010). It has been identified that soil texture and pH 85 are important for simulating CH₄ processes (Xu et al., 2015). In addition, the atmospheric CH₄ concentration data from satellite could be used as important benchmark for model validation purposes, for example Scanning Imaging Absorption spectrometer for Atmospheric ChartographY (SCIAMACHY) (Frankenberg et al., 2005) and Greenhouse gas Observing SATellite (GOSAT) (Yokota et al., 2009). 90

<u>Data-Model Integration</u> - Model development and data collection are two important, but historically independent scientific approaches; the integration between model development and data collection is much stronger for advancing science (De Kauwe et al., 2014; Luo et al., 2012; Peng et al., 2011). Although data-model integration is recognized as very important for understanding and predicting CH₄ processes and some progress has been made, integrating experiments and models presents multiple challenges, particularly, 1) the methods for integrating data with the models are not well developed for CH₄ cycling; 2) the metrics for evaluating data-model integration are not consistent in the scientific community; and 3) the regular communication between data scientists and modelers on various aspects of CH₄ processes and their model representation is lacking.

95

00 Methods for data-model integration have been recently created, for example, Kalman Filter (Gao et al., 2011), Bayesian (Ogle and Barber, 2008; Ricciuto et al., 2008; Schleip et al., 2009; Van Oijen et al., 2005), and Markov Chain Monte Carlo (Casella and Robert, 2005). However, no studies have evaluated these methods for integrating CH₄ data with models. In addition, the metric for evaluating the data-model integration is still not well developed. A very helpful strategy for data-model integration is 05 to solicit timely input from modelers when designing a field experiment. A good example of this is the U.S. Department of Energy-sponsored project Next Generation Ecosystem Experiments - Arctic (ngeearctic.ornl.gov), which was planned with inputs from field scientists, data scientists, and modelers. Another successful example is the U.S. DOE-sponsored project, Spruce and Peatland Responses Under Climatic and Environmental Change (SPRUCE) (mnspruce.ornl.gov), in which the experiment design 10 for data-model integration created an opportunity for modeling needs to be adopted by the field scientists. A modeling framework that focuses on model parameterization and validation ability is under development at Oak Ridge National Laboratory; building model optimization algorithm into an ESM framework will enable more effective parameterization of newly developed CH₄ modules within CLM at site, regional, and global scales (Ricciuto et al, personal communication).

15 6. Concluding Remarks

CH₄ dynamics in terrestrial ecosystems have been intensively studied, and model representation of CH₄ cycling has evolved as new knowledge becomes available. This is inherently a slow process.

Currently, the primary mechanisms for CH₄ processes in terrestrial ecosystems are implicitly represented in many, but not all, terrestrial ecosystem models. Development of CH₄ models began in the

- 20 late 1980s, and the pace of growth has been fast since the 1990s. Model development shifted from theoretical analysis in the 1980s and 1990s to being more applied in the 2000s and 2010s, expressed as being more focused on regional CH₄ budget quantification and integration with multiple sources of observational data. Although some current CH₄ models consider most of the relevant mechanisms, none of them consider all the processes for methanogenesis, methanotrophy, CH₄ transport, and their primary environmental controls. Further, evidence demonstrating that incorporating all of these processes would lead to more accurate prediction is needed. Incorporating sophisticated parameter assimilation, uncertainty quantification, equifinality quantification, and metrics of the benefits associated with increased model complexity would also facilitate scientific discovery.
- The CH₄ models for accurate projection of land-climate feedback in the next few decades 30 should: (1) use mechanistic formulations for primary CH₄ processes, (2) be embedded in Earth System Models for the global evaluation of terrestrial-climate feedback associated with CH₄ fluxes, (3) have the capacity to integrate multiple sources of data, which makes the model not only a prediction tool but also an integrative tool, and (4) be developed in association with model benchmarking frameworks. These four characteristics pave the way for examining CH₄ processes and flux in the context of global change.
- 35 These improvements for CH₄ modeling would be beneficial for ESMs and further simulation of climatecarbon cycle feedbacks.

Acknowledgements:

- The authors are grateful for financial and facility support from the University of Texas at El Paso. The authors are grateful for Dr. Yiqi Luo at University of Oklahoma for his comments on the manuscript. Five anonymous reviewers provided a number of constructive comments that significantly improved this manuscript. This review is part of the CH₄ modeling tasks within the NGEE-Arctic and SPRUCE projects sponsored by the US Department of Energy Office of Science. Contributions by FY, PJH, PET, SDW, and DEG are supported by the U.S. Department of Energy, Office of Science, Office of
- 45 Biological and Environmental Research. Oak Ridge National Laboratory is managed by UT-Battelle, LLC, for the U.S. Department of Energy under contract DE-AC05-00OR22725. CS is supported by the National Natural Science Foundation of China (41125001), HT is supported by NASA Carbon Monitoring System Program (NNX14AO73G) and NASA Interdisciplinary Science Program (NNX14AF93G).

50

References

Anisimov, O. A.: Potential feedback of thawing permafrost to the global climate system through methane emission, Environmental Research Letters, 2, 045016, DOI:10.1088/1748-9326/2/4/045016. 2007.

55 Arah, J. R. M. and Kirk, G. J. D.: Modeling rice plant-mediated methane emission, Nutrient Cycling in Agroecosystems, 58, 221-230, 2000.

Arah, J. R. M. and Stephen, K. D.: A model of the processes leading to methane emission from peatland, Atmospheric Environment, 32, 3257-3264, 1998.

Aronson, E. and Helliker, B.: Methane flux in non-wetland soils in response to nitrogen addition: a meta-analysis, Ecology, 91, 3242-3251, 2010.

Askaer, L., B. Elberling, T. Friborg, C. J. Jørgensen, and B. U. Hansen. 2011. Plant-mediated CH₄ transport and C gas dynamics quantified in-situ in a Phalaris arundinacea-dominant wetland. Plant and Soil **343**:287-301.

Aulakh, M. S., Wassmann, R., Rennenberg, H., and Fink, S.: Pattern and amount of aerenchyma relate to variable methane transport capacity of different rice cultivars, Plant Biology, 2, 182-194, 2000.

Banger, K., Tian, H., and Lu, C.: Do nitrogen fertilizers stimulate or inhibit methane emissions from rice fields?, Global Change Biology, 18, 3259-3267, 2012.

Barber, T. R., Burke, R. A., and Sackett, W. M.: Diffusive flux of methane from warm wetlands, Global Biogeochemical Cycles, 2, 411-425, 1988.

70 Barlett, K. B. and Harriss, R. C.: Review and assessment of methane emissions from wetlands, Chemosphere, 26, 261-320, 1993.

Becker, T., Kutzbach, L., Forbrich, I., Schneider, J., Jager, D., Thees, B., and Wilmking, M.: Do we miss the hot spots? The use of very high resolution aerial photographs to quantify carbon fluxes in peatlands, Biogeosciences, 5, 1387-1393, 2008.

75 Beckett, P. M., Armstrong, W., and Armstrong, J.: Mathematical modelling of methane transport by *Phragmites*: the potential for diffusion within the roots and rhizosphere, Aquatic Botany, 69, 293-312, 2001.

Beerling, D. J., Gardiner, T., Leggett, G., Mcleod, A., and Quick, W. P.: Missing methane emissions from leaves of terrestrial plants, Global Change Biology, 14, 1821-1826, 2008.

80 Bellisario, L., Bubier, J., Moore, T. & Chanton, J.: Controls on CH₄ emissions from a northern peatland. Global Biogeochemical Cycles, 13, 81-91, 1991.

Bhadra, A., Mukhopadhyay, S. N., and Ghose, T. K.: A kinetic model for methanogenesis of acetic acid in a multireactor system, Biotechnology and Bioengineering, XXVI, 257-264, 1984.

Blazewicz, S. J., Petersen, D. G., Waldrop, M. P., and Firestone, M. K.: Anaerobic oxidation of
methane in tropical and boreal soils: Ecological significance in terrestrial methane cycling, Journal of
Geophysical Research: Biogeosciences (2005–2012), 117, G02033, doi:10.1029/2011JG001864. 2012.

Blodau, C.: Carbon cycling in peatlands-A review of processes and controls, Environmental Reviews, 10, 111-134, 2002.

Bohn, T. J. and Lettenmaier, D. P.: Systematic biases in large-scale estimates of wetland methane
 emissions arising from water table formulations, Geophysical Research Letters, 37, 6pp, L22401, doi: 10.1029/2010GL045450. 2010.

Bohn, T. J., Lettenmaier, D. P., Sathulur, K., Bowling, L. C., Podest, E., McDonald, K. C., and Friborg, T.: Methane emissions from western Siberian wetlands: heterogeneity and sensitivity to climate change, Environmental Research Letters, 2, 045015, doi:10.1088/1748-9326/2/4/045015., 2007.

- 95 Bohn, T. J., Melton, J. R., Akihiko, I., Kleinen, T., Spahni, R., Stocker, B., Zhang, B., Zhu, X., Schroeder, R., Glagolev, M. V., Maksyutov, S., Brovkin, V., Chen, G., Denisov, S. N., Eliseev, A. V., Gallego-Sala, A., McDonald, K. C., Rawlins, M. A., Riley, W. J., Subin, Z. M., Tian, H., Zhuang, Q., and Kaplan, J. O.: WETCHIMP-WSL: Intercomparison of wetland methane emissions models over West Siberia, Biogeosciences, 12, 3321-3349, doi:10.5194/bg-12-3321-2015, 2015.
- 00 Bridgham, S. D., Cadillo-Quiroz, H., Keller, J. K., and Zhuang, Q.: Methane emissions from wetlands: biogeochemical, microbial, and modeling perspective from local to global scales, Global Change Biology, 19, 1325-1346, 2013.

Butterbach-Bahl, K., Papen, H., and Rennenberg, H.: Impact of gas transport through rice cultivars on methane emission from rice paddy fields, Plant, Cell & Environment, 20, 1175-1183, 1997.

05 Cai, Z.: Greenhouse gas budget for terrestrial ecosystems in China, Science China - Earth Sciences, 55, 173-182, 2012.

Caldwell, S. L., Laidler, J. R., Brewer, E. A., Eberly, J. O., Sandborgh, S. C., and Colwell, F. S.: Anaerobic oxidation of methane: mechanisms, bioenergetics, and ecology of associated microorganisms, Environmental Science and Technology, 42, 6791-6799, 2008.

10 Cao, M. K., Dent, J. B., and Heal, O. W.: Modeling methane emissions from rice paddies, Global Biogeochemical Cycles, 9, 183-195, 1995.

Cao, M. K., Gregson, K., and Marshall, S.: Global methane emission from wetlands and its sensitivity to climate change, Atmospheric Environment, 32, 3293-3299, 1998.

Casella, G. and Robert, C. (Eds.): Monte Carlo statistical methods, Springer, New York, 2005.

15 Chanton, J. P.: The effect of gas transport on the isotope signature of methane in wetlands, Organic Geochemistry, 36, 753-768, 2005.

Chanton, J. P., Martens, C. S., and Kelley, C. A.: Gas transport from methane-saturated, tidal freshwater and wetland sediments, Limnol. Oceanogr, 34, 807-819, 1989.

Chen, H., Zhu, Q., Peng, C., Wu, N., Wang, Y., Fang, X., Jiang, H., Xiang, W., Chang, J., Deng, X.,
and Yu, G.: Methane emissions from rice paddies natural wetlands, and lakes in China: synthesis and new estimate, Global Change Biology, 19, 19-32, 2012.

Christensen, T. and Cox, P.: Response of methane emission from Arctic tundra to climatic change: results from a model simulation, Tellus B, 47, 301-309, 1995.

Christensen, T. R., Prentice, I. C., Kaplan, J. O., Haxeltine, A., and Sitch, S.: Methane flux from northern wetlands and tundra an ecosystem source modeling approach, Tellus, 48B, 652-661, 1996.

Colmer, T.: Long - distance transport of gases in plants: a perspective on internal aeration and radial oxygen loss from roots, Plant, Cell & Environment, 26, 17-36, 2003.

Conrad, R.: Contribution of hydrogen to methane production and control of hydrogen concentration in methanogenic soils and sediments, FEMS Microbiology Ecology, 28, 193-202, 1999.

30 Conrad, R.: Control of methane production in terrestrial ecosystems. In: Exchange of trace gases between terrestrial ecosystems and the atmosphere, Andrease, M. O. and Schimel, D. S. (Eds.), Springer, New York, pp 39-58. 1989.

Conrad, R.: The global methane cycle: recent advances in understanding the microbial processes involved, Environmental Microbiology Reports, 1, 285-292, 2009.

35 Conrad, R.: Quantification of methanogenic pathways using stable carbon isotopic signatures: a review and a proposal, Organic Geochemistry, 36, 739-752, 2005.

Conrad, R.: Soil microbial processes involved in production and consumption of atmospheric trace gases. In: Advances in microbial ecology, Springer, pp 207-250. 1995.

Conrad, R.: Soil microorganisms as controllers of atmospheric trace gases(H₂, CO, CH₄, OCS, N₂O, and NO), Microbiological Reviews, 60, 609-640, 1996.

Conrad, R. and Claus, P.: Contribution of methanol to the production of methane and its 13C-isotopic signature in anoxic rice field soil, Biogeochemistry, 73, 381-393, 2005.

Conrad, R. and Klose, M.: How specific is the inhibition by methyl fluoride of acetoclastic methanogenesis in anoxic rice field soil?, FEMS microbiology ecology, 30, 47-56, 1999.

45 Cresto-Aleina, F., Runkle, B. R. K., Kleinen, T., Kutzbach, L., Schneider, J., and Brovkin, V.: Modeling micro-topographic controls on boreal peatland hydrology and methane fluxes, Biogeosciences Discussions, 12, 10195-10232, 2015.

Curry, C. L.: The consumption of atmospheric methane by soil in a simulated future climate, Biogeosciences, 6, 2355-2367, 2009.

50 Curry, C. L.: Modeling the soil consumption of atmospheric methane at the global scale, Global Biogeochemical Cycles, 21, GB4012. doi:10.1029/2006GB002818, 2007.

De Haas, Y., Windig, J., Calus, M., Dijkstra, J., De Haan, M., Bannink, A., and Veerkamp, R.: Genetic parameters for predicted methane production and potential for reducing enteric emissions through genomic selection, Journal of dairy science, 94, 6122-6134, 2011.

- 55 De Kauwe, M. G., Medlyn, B. E., Zaehle, S., Walker, A. P., Dietze, M. C., Wang, Y. P., Luo, Y., Jain, A. K., El - Masri, B., and Hickler, T.: Where does the carbon go? A model-data intercomparison of vegetation carbon allocation and turnover processes at two temperate forest free - air CO2 enrichment sites, New Phytologist, 203, 883-899, 2014.
- De Visscher, A. and Van Cleemput, O.: Simulation model for gas diffusion and methane oxidation in landfill cover soils, Waste Management, 23, 581-591, 2003.

Del Grosso, S. J., Ojima, D., Parton, W. J., Mosier, A., Peterson, G., and Schimel, D.: Simulated effects of dryland cropping intensification on soil organic matter and greenhouse gas exchanges using the DAYCENT ecosystem model, Environmental Pollution, 116, S75-S83, 2002.

Del Grosso, S. J., Ojima, D. S., Parton, W. J., Stehfest, E., Heistemann, M., DeAngelo, B. J., and Rose,
 S.: Global scale DAYCENT model analysis of greenhouse gas emissions and mitigation strategies for
 cropped soils, Global and Planetary Change, 67, 44-50, 2009.

Del Grosso, S. J., Parton, W. J., Mosier, A. R., Ojima, D. S., Potter, C. S., Borken, W., Brumme, R., Butterbach-Bahl, K., Crill, P. M., Dobbie, K. E., and Smith, K. A.: General CH₄ oxidation model and

comparisons of CH₄ oxidation in natural and managed systems, Global Biogeochemical Cycles, 14, 999-1019, 2000.

DeLong, E. F., Harwood, C. S., Chisholm, P. W., Karl, D. M., Moran, M. A., Schmidt, T. M., Tiedje, J. M., Treseder, K. K., and Worden, A. Z.: Incorporating microbial processes into climate models, The American Academy of Microbiology, Washington DC, 2011.

Ding, A. and Wang, M.: Model for methane emission from rice paddies and its application in southern 75 China, Advances in Atmospheric Sciences, 13, 159-168, 1996.

Dueck, T. A., De Visser, R., Poorter, H., Persijn, S., Gorissen, A., De Visser, W., Schapendonk, A., Verhagen, J., Snel, J., and Harren, F. J.: No evidence for substantial aerobic methane emission by terrestrial plants: a 13C - labelling approach, New Phytologist, 175, 29-35, 2007.

Eliseev, A. V., Mokhov, I. I., Arzhanov, M. M., Demchenko, P. F., and denisov, S. N.: Interaction of the methane cycle and processes in wetland ecosystems in a climate model of intermediate complexity, Atmospheric and Oceanic Physics, 44, 139-152, 2008.

Elliott, S., Maltrud, M., Reagan, M., Moridis, G., and Cameron - Smith, P.: Marine methane cycle simulations for the period of early global warming, Journal of Geophysical Research: Biogeosciences, 116, G01010, doi:10.1029/2010JG001300, 2011.

85 Entekhabi, D., Njoku, E. G., O'Neill, P. E., Kellogg, K. H., Crow, W. T., Edelstein, W. N., Entin, J. K., Goodman, S. D., Jackson, T. J., and Johnson, J.: The soil moisture active passive (SMAP) mission, Proceedings of the IEEE, 98, 704-716, 2010.

Evans, M. R., Grimm, V., Johst, K., Knuuttila, T., de Langhe, R., Lessells, C. M., Merz, M., O'Malley, M. A., Orzack, S. H., and Weisberg, M.: Do simple models lead to generality in ecology?, Trends in ecology & evolution, 28, 578-583, 2013.

Falz, K. Z., Holliger, C., Grosskopf, R., Liesack, W., Nozhevnikova, A., Müller, B., Wehrli, B., and Hahn, D.: Vertical distribution of methanogens in the anoxic sediment of Rotsee (Switzerland), Applied and Environmental Microbiology, 65, 2402-2408, 1999.

Fan, Z., David McGuire, A., Turetsky, M. R., Harden, J. W., Michael Waddington, J., and Kane, E. S.:
The response of soil organic carbon of a rich fen peatland in interior Alaska to projected climate change, Global change biology, 19, 604-620, 2013.

00

Forster, P., Ramaswamy, V., Artaxo, P., Berntsen, T., Betts, R., Fahey, D. W., Haywood, J., Lean, J., Lowe, D. C., Myhre, G., Nganga, J., Prinn, N. R., Raga, G., Schulz, M., and Dorland, R. V.: Changes in atmospheric constituents and in radiative forcing. In: Climate change 2007: The physical science basis. Contribution of working group I to the fourth assessment report of the intergovernmental panel on

climate change, Solomon, S., Qin, D., Manning, M., and Chen, Z. (Eds.), Cambridge University Press, Cambridge, United Kingdom and New York, USA., pp 133-216. 2007.

Frankenberg, C., Meirink, J. F., Van Weele, M., Platt, U., and Wagner, T.: Assessing methane emissions from global space-borne observations, Science, 308, 1010-1014, 2005.

05 Frenzel, P. and Karofeld, E.: CH₄ emission from a hollow-ridge complex in a raised bog: the role of CH₄ production and oxidation, Biogeochemistry, 51, 91-112, 2000.

Frenzel, P. and Rudolph, J.: Methane emission from a wetland plant: the role of CH4 oxidation in Eriophorum, Plant and Soil, 202, 27-32, 1998.

Fung, I., John, J., Lerner, J., Matthews, E., Prather, M., Steele, L. P., and Fraser, P. J.: Threedimensional model synthesis of the global methane cycle, Journal of Geophysical Research, 96, 1303313065, 1991.

Gao, C., Wang, H., Weng, E., Lakshmivarahan, S., Zhang, Y., and Luo, Y.: Assimilation of multiple data sets with the ensemble Kalman filter to improve forecasts of forest carbon dynamics, Ecological Applications, 21, 1461-1473, 2011.

15 Gao, X., Schlosser, C. A., Sokolov, A., Anthony, K. W., Zhuang, Q., and Kicklighter, D.: Permafrost degradation and methane: low risk of biogeochemical climate-warming feedback, Environmental Research Letters, 8, 035014, doi:10.1088/1748-9326/8/3/035014, 2013.

Gauthier, M., Bradley, R. L., and Šimek, M.: More evidence that anaerobic oxidation of methane is prevalent in soils: Is it time to upgrade our biogeochemical models?, Soil Biology and Biochemistry, 80, 167-174, 2015.

Gedney, N., Cox, P., and Huntingford, C.: Climate feedback from wetland methane emissions, Geophysical Research Letters, 31, L20503, doi:10.1029/2004GL020919, 2004.

20

30

Gerard, G. and Chanton, J.: Quantification of methane oxidation in the rhizosphere of emergent aquatic macrophytes: defining upper limits, Biogeochemistry, 23, 79-97, 1993.

25 Gong, J., Kellomaki, S., Wang, K., Zhang, C., Shurpali, N., and Martikainen, P. J.: Modeling CO₂ and CH₄ flux changes in pristine peatlands of Finland under changing climate conditions Ecological Modelling, 263, 64-80, 2013.

Grant, R. & Roulet, N. Methane efflux from boreal wetlands: Theory and testing of the ecosystem model Ecosys with chamber and tower flux measurements. Global Biogeochemical Cycles, 16, 2-1-2-16. 2002.

Grant, R., Juma, N., and McGill, W.: Simulation of carbon and nitrogen transformations in soil: mineralization, Soil Biology and Biochemistry, 25, 1317-1329, 1993.

Grant, R. F.: A review of the Canadian ecosystem model *ecosys*. In: Modeling Carbon and Nitrogen Dynamics for Soil Management, Shaffer, M. J., Ma, L., and Hansen, S. (Eds.), CRC Press, New York, pp 173-264. 2001.

35

Grant, R. F.: Simulation of methanogenesis in the mathematical model Ecosys, Soil Biology and Biochemistry, 30, 883-896, 1998.

Gulledge, J. and Schimel, J. P.: Low-concentration kinetics of atmospheric CH4 oxidation in soil and mechanism of NH4+ inhibition, Applied and Environmental Microbiology, 64, 4291-4298, 1998a.

40 Gulledge, J. and Schimel, J. P.: Moisture control over atmospheric CH₄ consumption and CO₂ production in diverse Alaskan soils, Soil Biology and Biochemistry, 30, 1127-1132, 1998b.

Hakemian, A. S. and Rosenzweig, A. C.: The biochemistry of methane oxidation, Annual Review of Biochemistry, 76, 223-241, 2007.

Hanson, R. S. and Hanson, T. E.: Methanotrophic bacteria, Microbiology and Molecular Biology
Reviews, 60, 60(2), 439-471., 1996.

Heilman, M. A. and Carlton, R. G.: Methane oxidation associated with submersed vascular macrophytes and its impact on plant diffusive methane flux, Biogeochemistry, 52, 207-224, 2001.

Higgins, I. J., Best, D. J., Hammond, R. C., and Scott, D.: Methane-oxidizing microorganisms, Microbiological Reviews, 45, 556-590, 1981.

50 Hodson, E. L., Poulter, B., Zimmermann, N. E., Prigent, C., and Kaplan, J. O.: The El Nino-Southern Oscillation and wetland methane interannual variability, Geophysical Research Letters, 38, L08810, doi: 10.1029/2011GL046861., 2011.

Holgerson, M. A. and Raymond, P. A.: Large contribution to inland water CO₂ and CH₄ emissions from very small ponds, Nature Geoscience, 9, 222-226, 2016.

55 Hopcroft, P. O., Valdes, P. J., and Beerling, D. J.: Simulating idealized Dansgaard-Oeschger events and their potential impacts on the global methane cycle, Quarternary Science Review, 30, 3258-3268, 2011.

Hosono, T. and Nouchi, I.: The dependence of methane transport in rice plants on the root zone temperature, Plant and Soil, 191, 233-240, 1997.

Huang, Y., Sass, R. L., and Fisher, F. M.: Model estimates of methane emission from irrigated rice
cultivation of China, Global Change Biology, 4, 809–821. doi:10.1046/j.1365-2486.1998.00175.x.1998a.

Huang, Y., Sass, R. L., and Fisher, F. M.: A semi-empirical model of methane emission from flooded rice paddy soils, Global Change Biology, 4, 247-268, 1998b.

Huang, Y., Zhang, W., Zheng, X., Li, J., and Yu, Y.: Modeling methane emission from rice paddies
with various agricultural practices, Journal of Geophysical Research, 109, D08113, doi:10.1029/2003JD004401, 2004.

Inatomi, M., Ito, A., Ishijima, K., and Murayama, S.: Greenhouse gas budget of a cool-temperate deciduous broad-leaved forest in Japan estimated using a process-based model, Ecosystems, 13, 472-483, 2010.

70 IPCC: Summary for policymakers, Cambridge, United Kingdom and New York, NY, USA, 2013.

Ito, A. and Inatomi, M.: Use of a process-based model for assessing the methane budgets of global terrestrial ecosystems and evaluation of uncertainty, Biogeosciences, 9, 759-773, 2012.

Karl, D. M., Beversdorf, L., Björkman, K. M., Church, M. J., Martinez, A., and Delong, E. F.: Aerobic production of methane in the sea, Nature Geoscience, 1, 473-478, 2008.

75 Keppler, F., Hamilton, J. T. G., Brass, M., and Rockmann, T.: Methane emissions from terrestrial plants under aerobic conditions, Nature, 439, 187-191, 2006.

Kettunen, A.: Connecting methane fluxes to vegetation cover and water table fluctuations at microsite level: a modeling study, Global Biogeochemical Cycles, 17, 1051, doi:10.1029/2002GB001958, 2003.

King, G. M.: In Situ Analyses of Methane Oxidation Associated with the Roots and Rhizomes of a Bur
 Reed, Sparganium eurycarpum, in a Maine Wetland, Applied and environmental microbiology, 62, 4548-4555, 1996.

Kotsyurbenko, O. R., Chin, K. J., Glagolev, M. V., Stubner, S., Simankova, M. V., Nozhevnikova, A. N., and Conrad, R.: Acetoclastic and hydrogenotrophic methane production and methanogenic populations in an acidic West - Siberian peat bog, Environmental microbiology, 6, 1159-1173, 2004.

85 Koven, C. D., Riley, W., Subin, Z. M., Tang, J., Torn, M. S., Collins, W. D., Bonan, G. B., Lawrence, D. M., and Swenson, S. C.: The effect of vertically-resolved soil biogeochemistry and alternate soil C and N models on C dynamics of CLM4, Biogeosciences Discussions, 10, 7201-7256, 2013.

Koven, C. D., Ringeval, B., Friedlingstein, P., Ciais, P., Cadule, P., Khvorostyanov, D., Krinner, G., and Tarnocai, C.: Permafrost carbon-climate feedbacks accelerate global warming, Proceedings of the National Academy of Sciences, 108, 14769-14774., 2011.

90

00

Krüger, M., Frenzel, P., and Conrad, R.: Microbial processes influencing methane emission from rice fields, Global Change Biology, 7, 49-63, 2001.

Krumholz, L. R., Hollenback, J. L., Roskes, S. J., and Ringelberg, D. B.: Methanogenesis and methanotrophy within a Sphagnum peatland, FEMS Microbiology Ecology, 18, 215-224, 1995.

Lai, D. Y. F.: Methane dynamics in Northern Peatlands: A Review, Pedosphere, 19, 409-421, 2009.

Larsen, P. E., Gibbons, S. M., and Gilbert, J. A.: Modeling microbial community structure and functional diversity across time and space, FEMS microbiology letters, 332, 91-98, 2012.

Lenhart, K., Bunge, M., Ratering, S., New, T. R., Schuttmann, I., Greule, M., Kammann, C., Schnell, S., Muller, C., Zorn, H., and Keppler, F.: Evidence for methane production by saprotrophic fungi, Nature Communication, 3, 1046, doi:10.1038/ncomms2049, 2012.

Li, C., Frolking, S., Xiao, X., Moore III, B., Boles, S., Qiu, J., Huang, Y., Salas, W., and Sass, R.: Modeling impacts of farming management alternatives on CO₂, CH₄, and N₂O emissions: a case study for water management of rice agriculture of China, Global Biogeochemical Cycles, 19, doi:10.1029/2004GB002341, 2005.

05 Li, C.: Modeling trace gas emissions from agricultural ecosystems, Nutrient Cycling in Agroecosystems, 58, 259-276, 2000a.

Li, C.: Modeling trace gas emissions from agricultural ecosystems, Nutrient Cycling in Agroecosystems, 58, 259-276, 2000b.

Li, T., Huang, Y., Zhang, W., and Yu, Y.: Methane emissions associated with the conversion of
marshland to cropland and climate change on the Sanjiang Plain of Northeast China from 1950 to 2100,
Biogeosciences, 9, 5199-5215, 2012.

Liu, L. and Greaver, T.: A review of nitrogen enrichment effects on three biogenic GHGs: the CO₂ sink may be largely offset by stimulated N₂O and CH₄ emission, Ecology Letters, 12, 1103-1117, 2009.

Lovley, D. P. and Klug, M. J.: Model for distribution of sulfate reduction and methanogenesis in freshwater sediments, Geochimica et Cosmochimica Acta, 50, 11-18, 1986.

Luo, Y., Randerson, J. T., Abramowitz, G., Bacour, C., Blyth, E., Carvalhais, N., Ciais, P., Dalmonech, D., Fisher, J., Fisher, R., Friedlingstein, P., Hibbard, K. A., Hoffman, F., Huntzinger, D. N., Jones, C.

D., Koven, C. D., Lawrence, D. M., Li, D. J., Mahecha, M., Niu, S., Norby, R. J., Piao, S., Qi, X., Peylin, P., Prentice, I. C., Riley, W. J., Reichstein, M., Schwalm, C. R., Wang, Y., Xia, J., Zaehle, S., and Zhou, X.: A framework of benchmarking land models, Biogeosciences, 9, 3857-3874, 2012.

Martens, C. S., Albert, D. B., and Alperin, M. J.: Biogeochemical processes controlling methane in gassy coastal sediments -- Part 1. A model coupling organic matter flux to gas production, oxidation and transport, Continental Shelf Research, 18, 1741-1770, 1998.

Martinson, G. O., Werner, F. A., Scherber, C., Conrad, R., Corre, M. D., Flessa, H., Wolf, K., Klose, 25 M., Gradstein, S. R., and Veldkamp, E.: Methane emissions from tank bromeliads in neotropical forests, Nature Geoscience, 3, 766-769, 2010.

Massman, W., Sommerfeld, R., Mosier, A., Zeller, K., Hehn, T., and Rochelle, S.: A model investigation of turbulence - driven pressure - pumping effects on the rate of diffusion of CO₂, N₂O, and CH₄ through layered snowpacks, Journal of Geophysical Research: Atmospheres (1984–2012), 102, 18851-18863, 1997.

30

20

40

Mastepanov, M., Sigsgaard, C., Dlugokencky, E. J., Houweling, S., Strom, L., Tamstorf, M. P., and Christensen, T. R.: Large tundra methane burst during onset of freezing, Nature, 456, 628-630, 2008.

Matthews, E. & Fung, I.: Methane emissions from natural wetlands: global distribution, area and environmental characteristics of sources, Global Biogeochemical Cycles, 1, 61-86, 1987.

35 Matthews, R. B., Wassmann, R., and Arah, J. R. M.: Using a crop/soil simulation model and GIS techniques to assess methane emissions from rice fields in Asia. I. model development, Nutrient Cycling in Agroecosystems, 58, 141-159, 2000.

Mau, S., Blees, J., Helmke, E., Niemann, H., and Damm, E.: Vertical distribution of methane oxidation and methanotrophic response to elevated methane concentrations in stratified waters of the Arctic fjord Storfjorden (Svalbard, Norway), Biogeosciences, 10, 6267-6278, 2013.

McCalley, C. K., Woodcroft, B. J., Hodgkins, S. B., Wehr, R. A., Kim, E.-H., Mondav, R., Crill, P. M., Chanton, J. P., Rich, V. I., Tyson, G. W., and Saleska, S. R.: Methane dynamics regulated by microbial community response to permafrost thaw, Nature, 514, 478-481, 2014.

Melloh, R. A. and Crill, P. M.: Winter methane dynamics in a temperate peatland, Global 45 Biogeochemical Cycles, 10, 247-254, 1996.

Melton, J. R., Wania, R., Hodson, E. L., Poulter, B., Ringeval, B., Spahni, R., Bhon, T., Avis, C. A., Beerling, D. J., Chen, G., Eliseev, A. V., Denisov, S. N., Hopcroft, P. O., Lettenmaier, D. P., Riley, W. J., Singarayer, J. S., Subin, Z. M., Tian, H., Zurcher, S., Brovkin, V., Van Bodegom, P. M., Kleinen, T.,

Yu, Z. C., and Kaplan, J. O.: Present state of global wetland extent and wetland methane modelling:
conclusions from a model inter-comparison project (WETCHIMP), Biogeosciences, 10, 753-788, 2013.

Meng, L., Hess, P., Mahowald, N., Yavitt, J., Riley, W., Subin, Z., Lawrence, D., Swenson, S., Jauhiainen, J., and Fuka, D.: Sensitivity of wetland methane emissions to model assumptions: application and model testing against site observations, Biogeosciences, 9, 2793-2819, 2012.

Mer, J. L. and Roger, P.: Production, oxidation, emission and consumption of methane by soils: a review, European Journal of Soil Biology, 37, 25-50, 2001.

Miller, K. E., Lai, C.-T., Friedman, E. S., Angenent, L. T., and Lipson, D. A.: Methane suppression by iron and humic acids in soils of the Arctic Coastal Plain, Soil Biology and Biochemistry, 83, 176-183, 2015.

Mokhov, I. I., Eliseev, A. V., and denisov, S. N.: Model diagnostics of variations in methane emissions
by wetlands in the second half of the 20th century based on reanalysis data, Doklady Earth Sciences, 417, 1293-1297, 2007.

Monechi, S., Coccioni, R., and Rampino, M. R.: Large ecosystem perturbations : causes and consequences, Geological Society of America, Boulder, Colo., 2007.

Morrissey, L. and Livingston, G.: Methane emissions from Alaska arctic tundra: An assessment of local
 spatial variability, Journal of Geophysical Research: Atmospheres (1984–2012), 97, 16661-16670,
 1992.

Mosier, A., Delgado, J., Cochran, V., Valentine, D., and Parton, W.: Impact of agriculture on soil consumption of atmospheric CH₄ and a comparison of CH₄ and N₂O flux in subarctic, temperate and tropical grasslands, Nutrient Cycling in Agroecosystems, 49, 71-83, 1997.

70 Murase, J. and Kimura, M.: Methane production and its fate in paddy fields: IX. Methane flux distribution and decomposition of methane in the subsoil during the growth period of rice plants, Soil science and plant nutrition, 42, 187-190, 1996.

75

Nauta, A. L., Heijmans, M. M., Blok, D., Limpens, J., Elberling, B., Gallagher, A., Li, B., Petrov, R. E., Maximov, T. C., and van Huissteden, J.: Permafrost collapse after shrub removal shifts tundra ecosystem to a methane source, Nature Climate Change, 5, 67-70, 2015.

Nazaries, L., Murrell, J. C., Millard, P., Baggs, L., and Singh, B. K.: Methane, microbes and models: fundamental understanding of the soil methane cycle for future predictions, Environmental Microbiology, 15(9):2395-417, doi:10.1111/1462-2920.12149., 2013.

Nouchi, I., Hosono, T., Aoki, K., and Minami, K.: Seasonal variation in methane flux from rice paddies
 associated with methane concentration in soil water, rice biomass and temperature, and its modelling,
 Plant and Soil, 161, 195-208, 1994.

Nouchi, I., S. Mariko, and K. Aoki. 1990. Mechanism of methane transport from the rhizosphere to the atmosphere through rice plants. Plant physiology **94**:59-66.

Ogle, K. and Barber, J. J.: Bayesian data—model integration in plant physiological and ecosystem ecology. In: Progress in botany, Springer, Verlag Berlin Heidelberg, pp 281-311, 2008.

Pareek, S., Matsui, S., Kim, S. K., and Shimizu, Y.: Mathematical modeling and simulation of methane gas production in simulated landfill column reactors under sulfidogenic and methanogenic environments, Water science and technology, 39, 235-242, 1999.

Peng, C., Guiot, J., Wu, H., Jiang, H., and Luo, Y.: Integrating models with data in ecology and palaeoecology: advances towards a model–data fusion approach, Ecology letters, 14, 522-536, 2011.

90

Philippot, L., Andersson, S. G., Battin, T. J., Prosser, J. I., Schimel, J. P., Whitman, W. B., and Hallin, S.: The ecological coherence of high bacterial taxonomic ranks, Nature Reviews Microbiology, 8, 523-529, 2010.

Potter, C. S.: An ecosystem simulation model for methane production and emission from wetlands,
Global Biogeochemical Cycles, 11, 495-506, 1997.

Potter, C. S., Davidson, E. A., and Verchot, L. V.: Estimation of global biogeochemical controls and seasonality in soil methane consumption, Chemosphere, 32, 2219-2246, 1996.

Ren, W., Tian, H., Xu, X., Liu, M., Lu, C., Chen, G., Melillo, J., Reilly, J., and Liu, J.: Spatial and temporal patterns of CO₂ and CH₄ fluxes in China's croplands in response to multifactor environmental changes, Tellus Series B-Chemical and Physical Meteorology, 63, 222-240, 2011.

Ricciuto, D. M., Davis, K. J., and Keller, K.: A bayesian calibration of a simple carbon cycle model: the role of observations in estimating and reducing uncertainty, Global Biogeochemical Cycles, 22, GB2030, doi:2010.1029/2006GB002908., 2008.

Ridgwell, A. J., Marshall, S. J., and Gregson, K.: Consumption of atmospheric methane by soils: a process-based model, Global Biogeochemical Cycles, 13, 59-70, 1999.

Riley, W. J., Subin, Z. M., Lawrence, D. M., Swenson, S. C., Torn, M. S., Meng, L., Mahowald, N. M., and Hess, P.: Barriers to predicting changes in global terrestrial methane fluxes: analyses using CLM4Me, a methane biogeochemistry model integrated in CESM, Biogeosciences, 8, 1925-1953, 2011.

10 Ringeval, B., de Noblet-Ducoudre, N., Ciais, P., Bousquet, P., Prigent, C., Para, F., and Rossow, W. B.: An attempt to quantify the impact of changes in wetland extent on methane emissions on the seasonal and interannual time scales, Global Biogeochemical Cycles, 24, GB2003, doi: 10.1029/2008GB003354., 2010.

Ringeval, B., Friedlingstein, P., Koven, C., Ciais, P., de Noblet-Ducoudré, N., Decharme, B., and
 Cadule, P.: Climate-CH₄ feedback from wetlands and its interaction with the climate-CO₂ feedback,
 Biogeosciences, 8, 2137-2157, 2011.

Rodhe, H.: A comparison of the contribution of various gases to the greenhouse effect, Science, 248, 1217-1219, 1990.

Schimel, J.: Ecosystem consequences of microbial diversity and community structure. In: Arctic and
 alpine biodiversity: patterns, causes and ecosystem consequences, Springer, Springer-Verlag Berlin
 Heidelberg, pp 239-254, 1995.

Schimel, J. P. and Gulledge, J.: Microbial community structure and global trace gases, Global Change Biology, 4, 745-758, 1998.

Schleip, C., Rais, A., and Menzel, A.: Bayesian analysis of temperature sensitivity of plant phenology in Germany, Agricultural and Forest Meteorology, 149, 1699-1708, 2009.

Schütz, H., Seiler, W., and Conrad, R.: Processes involved in formation and emission of methane in rice paddies, Biogeochemistry, 7, 33-53, 1989.

Segers, R.: Methane production and methane consumption: a review of processes underlying wetland methane fluxes, Biogeochemistry, 41, 23-51, 1998.

30 Segers, R. and Kengen, S. W. M.: Methane production as a function of anaerobic carbon mineralization: a process model, Soil Biology and Biochemistry, 30, 1107-1117, 1998.

Segers, R. and Leffelaar, P. A.: Modeling methane fluxes in wetlands with gas-transporting plants 1. single-root scale, Journal of Geophysical Research, 106, 3511-3528., 2001a.

Segers, R. and Leffelaar, P. A.: Modeling methane fluxes in wetlands with gas-transporting plants 3.
plot scale, Journal of Geophysical Research, 106, 3541-3558, 2001b.

Segers, R., Rappoldt, C., and Leffelaar, P. A.: Modeling methane fluxes in wetlands with gastransporting plants 2. soil layer scale, Journal of Geophysical Research, 106, 3529-3540., 2001.

Shoemaker, J. K., Keenan, T. F., Hollinger, D. Y., and Richardson, A. D.: Forest ecosystem changes from annual methane source to sink depending on late summer water balance, Geophysical Research Letters, 41, 673-679, 2014.

Smemo, K. A. and Yavitt, J. B.: Anaerobic oxidation of methane: an underappreciated aspect of methane cycling in peatland ecosystems?, Biogeosciences, 8, 779-793, 2011.

40

50

Söhngen, N.: Über bakterien, welche methan als kohlenstoffnahrung und energiequelle gebrauchen. Zentrabl Bakteriol Parasitenk Infektionskr, 15, 513-517, 1906.

45 Song, C., Xu, X., Sun, X., Tian, H., Sun, L., Miao, Y., Wang, X., and Guo, Y.: Large methane emission upon spring thaw from natural wetlands in the northern permafrost region, Environmental Research Letters, 7, 034009, doi:10.1088/1748-9326/7/3/034009. 2012.

Spahni, R., Wania, R., Neef, L., van Weele, M., Pison, I., Bousquet, P., Frankenberg, C., Foster, P. N., Joos, F., Prentice, I. C., and Van Velthoven, P.: Constraining global methane emissions and uptake by ecosystems, Biogeosciences, 8, 1643-1665, 2011.

Ström, L., Mastepanov, M., and Christensen, T. R.: Species-specific effects of vascular plants on carbon turnover and methane emissions from wetlands, Biogeochemistry, 75, 65-82, 2005.

Summons, R. E., Franzmann, P. D., and Nichols, P. D.: Carbon isotopic fractionation associated with methylotrophic methanogenesis, Organic Geochemistry, 28, 465-475, 1998.

55 Tagesson, T., Mastepanov, M., Mölder, M., Tamstorf, M. P., Eklundh, L., Smith, B., Sigsgaard, C., Lund, M., Ekberg, A., and Falk, J. M.: Modelling of growing season methane fluxes in a high-Arctic wet tundra ecosystem 1997-2010 using in situ and high-resolution satellite data, Tellus B, 65, 19722, http://dx.doi.org/10.3402/tellusb.v65i0.19722, 2013.

Tang, J. and Riley, W.: Technical Note: Simple formulations and solutions of the dual-phase diffusive transport for biogeochemical modeling, Biogeosciences, 11, 3721-3728, 2014.

Tang, J. and Riley, W.: A total quasi-steady-state formulation of substrate uptake kinetics in complex networks and an example application to microbial litter decomposition, Biogeosciences, 10, 8329-8351, 2013.

Tang, J. and Zhuang, Q.: Equifinality in parameterization of process - based biogeochemistry models:
A significant uncertainty source to the estimation of regional carbon dynamics, Journal of Geophysical Research: Biogeosciences, 113, G04010, doi:10.1029/2008JG000757, 2008.

Tang, J., Zhuang, Q., Shannon, R., and White, J.: Quantifying wetland methane emissions with processbased models of different complexities, Biogeosciences, 7, 3817-3837, 2010.

Tian, H., Chen, G., Lu, C., Xu, X., Ren, W., Zhang, B., Banger, K., Tao, B., Pan, S., and Liu, M.:
Global methane and nitrous oxide emissions from terrestrial ecosystems due to multiple environmental changes, Ecosystem Health and Sustainability, 1:art4. http://dx.doi.org/10.1890/EHS14-0015.1, 2015.

Tian, H., Xu, X., Liu, M., Ren, W., Zhang, C., Chen, G., and Lu, C.: Spatial and temporal patterns of CH₄ and N₂O fluxes in terrestrial ecosystems of North America during 1979-2008: application of a global biogeochemistry model, Biogeosciences, 7, 2673-2694, 2010.

75 Tian, H., Xu, X., Lu, C., Liu, M., Ren, W., Chen, G., Melillo, J. M., and Liu, J.: Net exchanges of CO₂, CH₄, and N₂O between China's terrestrial ecosystems and the atmosphere and their contributions to global climate warming, Journal of Geophysical Research, 116, G02011, doi: 10.1029/2010JG001393, 2011.

Tokida, T., Mizoguchi, M., Miyazaki, T., Kagemoto, A., Nagata, O., and Hatano, R.: Episodic release of methane bubbles from peatland during spring thaw, Chemosphere, 70, 165-171, 2007.

80

Topp, E. and Pattey, E.: Soils as sources and sinks for atmospheric methane, Canadian journal of soil science, 77, 167-177, 1997.

Tveit, A. T., Urich, T., Frenzel, P., and Svenning, M. M.: Metabolic and trophic interactions modulate methane production by Arctic peat microbiota in response to warming, Proceedings of the National
Academy of Sciences, 112, E2507-E2516, 2015.

Valentine, D. L. and Reeburgh, W. S.: New perspectives on anaerobic methane oxidation, Environmental Microbiology, 2, 477-484, 2000.

van Bodegom, P. M., Leffelaar, P. A., Stams, A. J. M., and Wassmann, R.: Modeling methane emissions from rice fields: variability, uncertainty, and sensitivity analysis of processes involved, Nutrient Cycling in Agroecosystems, 58, 231-248, 2000.

Van Bodegom, P. M., Wassmann, R., and Metra-Corton, T. M.: A process-based model for methane emission predictions from flooded rice paddies, Global Biogeochemical Cycles, 15, 247-263, 2001.

Van Oijen, M., Rougier, J., and Smith, R.: Bayesian calibration of process-based forest models: bridging the gap between models and data, Tree Physiology, 25, 915-927, 2005.

95 Volta, A.: Lettere dell'Illustrissimo Signor Volta Alessandro sull'aria inflammabile native dele paludi. In, Giuseppe Marelli, Milano. 1777

Wagner, D., Lipski, A., Embacher, A., and Gattinger, A.: Methane fluxes in permafrost habitats of the Lena Delta: effects of microbial community structure and organic matter quality, Environmental Microbiology, 7, 1582-1592, 2005.

00 Wahlen, M.: The global methane cycle, Annual Review of Earth and Planetary Sciences, 21, 407-426, 1993.

Walter, B. P. and Heimann, M.: A process-based, climate-sensitive model to derive methane emissions from natural wetlands: Application to five wetland sites, sensitivity to model parameters, and climate, Global Biogeochemical Cycles, 14, 745-765, 2000.

05 Walter, B. P., Heimann, M., Shannon, R. D., and White, J. R.: A process-based model to derive methane emissions from natural wetlands, Geophysical Research Letters, 23, 3731-3734, 1996.

Wang, Z., Han, X., Wang, G. G., Song, Y., and Gulledge, J.: Aerobic methane emission from plants in the Inner Mongolia Steppe, Environmental Science and Technology, 42, 62-68, 2007.

Wania, R.: Modelling northern peatland land surface processes, vegetation dynamics and methaneemissions, Doktorarbeit, University of Bristol, pp 1-140.. 2007.

Wania, R., Melton, J. R., Hodson, E. L., Poulter, B., Ringeval, B., Spahni, R., Bohn, T. J., Avis, C. A.,
Beerling, D. J., Chen, G., Eliseev, A. V., Denisov, S. N., Hopcroft, P. O., Lettenmaier, D., Riley, W. J.,
Singarayer, J. S., Subin, Z. M., Tian, H., Zuercher, S., van Bodegom, P. M., Kleinen, T., Yu, Z., and
Kaplan, J. O.: Present state of global wetland extent and wetland methane modelling: methodology of a
model inter-comparison project (WETCHIMP), Geoscientific Model Development, 6, 617-641, 2013.

Wania, R., Ross, I., and Prentice, I. C.: Implementation and evaluation of a new methane model within a dynamic global vegetation model: LPJ-WHyMe v1.3.1, Geoscientific Model Development, 3, 565-584, 2010.

Wania, R., Ross, I., and Prentice, I. C.: Integrating peatlands and permafrost into a dynamic global
vegetation model: 1. Evaluation and sensitivity of physical land surface processes, Global
Biogeochemical Cycles, 23, GB3014, doi:10.1029/2008GB003412. 2009.

15

30

Wassmann, R., Neue, H., Lantin, R., Makarim, K., Chareonsilp, N., Buendia, L., and Rennenberg, H.: Characterization of methane emissions from rice fields in Asia. II. Differences among irrigated, rainfed, and deepwater rice, Nutrient Cycling in Agroecosystems, 58, 13-22, 2000.

25 Watts, J., Kimball, J.S., Parmentier, F., Sachs, T., Rinne, J., Zona, D., Oechel, W., Tagesson, T., Jackowicz-Korczynski, M. and Aurela, M.: A satellite data driven biophysical modeling approach for estimating northern peatland and tundra CO₂ and CH₄ fluxes. Biogeosciences, 11, 1961-1980. 2014.

Weller, G., Chapin, F. S., Everett, K. R., Hobbie, J. E., Kane, D., Oechel, W. C., Ping, C. L., Reeburgh, W. S., Walker, D., and Walsh, J.: The arctic flux study: A regional view of trace gas release, Journal of Biogeography, 22, 365-374, 1995.

Whiting, G. J. and Chanton, J. P.: Control of the diurnal pattern of methane emission from emergent aquatic macrophytes by gas transport mechanisms, Aquatic Botany, 54, 237-253, 1996.

Xu, S., Jaffe, P. R., and Mauzerall, D. L.: A process-based model for methane emission from flooded rice paddy systems, Ecological Modelling, 205, 475-491, 2007.

35 Xu X.F., Hahn M., Kumar J., Yuan F.M., Tang G.P., Thornton P., Torn M., Wullschleger S. Upscaling plot-scale methane flux to an eddy covariance tower domain in Barrow, AK: integrating in-situ data with a microbial functional group-based model. AGU Annual Fall meeting. San Francisco, 2014.

40

55

Xu, X.: Modeling methane and nitrous oxide exchanges between the atmosphere and terrestrial ecosystems over North America in the context of multifactor global change, Ph.D. Dissertation, School of Forestry and Wildlife Sciences, Auburn University, Auburn, 199 pp., 2010.

Xu, X., Elias, D. A., Graham, D. E., Phelps, T. J., Carrol, S. L., Wullschleger, S. D., and Thornton, P. E.: A microbial functional group based module for simulating methane production and consumption: application to an incubation permafrost soil, Journal of Geophysical Research-Biogeosciences, 120, 1315-1333, 2015.

45 Xu, X., Schimel, J. P., Thornton, P. E., Song, X., Yuan, F., and Goswami, S.: Substrate and environmental controls on microbial assimilation of soil organic carbon: a framework for Earth system models, Ecology Letters, 17, 547-555., 2014.

Xu, X. and Tian, H.: Methane exchange between marshland and the atmosphere over China during 1949-2008, Global Biogeochemical Cycles, 26, GB2006, doi:10.1029/2010GB003946, 2012.

50 Xu, X. F., Tian, H. Q., Zhang, C., Liu, M. L., Ren, W., Chen, G. S., Lu, C. Q., and Bruhwiler, L.: Attribution of spatial and temporal variations in terrestrial methane flux over North America, Biogeosciences, 7, 3637-3655, 2010.

Yokota, T., Yoshida, Y., Eguchi, N., Ota, Y., Tanaka, T., Watanabe, H., and Maksyutov, S.: Global concentrations of CO_2 and CH_4 retrieved from GOSAT: First preliminary results, Sola, 5, 160-163, 2009.

Yvon-Durocher, G., Allen, A. P., Bastviken, D., Conrad, R., Gudasz, C., St-Pierre, A., Thanh-Duc, N., and Del Giorgio, P. A.: Methane fluxes show consistent temperature dependence across microbial to ecosystem scales, Nature, 507, 488-491, 2014.

 Zhang, Y., Sachs, T., Li, C., and Boike, J.: Upscaling methane fluxes from closed chambers to eddy
 covariance based on a permafrost biogeochemistry integrated model, Global Change Biology, 18, 1428-1440, 2012.

Zhu, Q., Riley, W. J., Tang, J., and Koven, C. D.: Multiple soil nutrient competition between plants, microbes, and mineral surfaces: model development, parameterization, and example applications in several tropical forests, Biogeosciences, 13, 341-363, 2016.

65 Zhu, Q., Liu, J., Peng, C., Chen, H., Fang, X., Jiang, H., Yang, G., Zhu, D., Wang, W., and Zhou, X.: Modeling methane emissions from natural wetlands by development and application of the TRIPLEX-GHG model, Geoscientific Model Development, 7, 981-999, doi:10.5194/gmd-7-981-2014, 2014.

Zhu, X., Zhuang, Q., Chen, M., Sirin, A., Melillo, J., Kicklighter, D., Sokolov, A., and Song, L.: Rising methane emissions in response to climate change in Northern Eurasia during the 21st century, Environmental Research Letters, 6, 045211, doi:10.1088/1748-9326/6/4/045211., 2011.

Zhu, X., Zhuang, Q., Gao, X., Sokolov, A., and Schlosser, C. A.: Pan-Arctic land–atmospheric fluxes of methane and carbon dioxide in response to climate change over the 21st century, Environmental Research Letters, 8, 045003, doi:10.1088/1748-9326/8/4/045003, 2013a.

Zhu, X., Zhuang, Q., Lu, X., and Song, L.: Spatial scale-dependent land-atmospheric methane
 exchanges in the northern high latitudes from 1993 to 2004, Biogeosciences, 11, 1693-1704,
 doi:10.5194/bg-11-1693-2014, 2014.

Zhu, X., Zhuang, Q., Qin, Z., Glagolev, M., and Song, L.: Estimating wetland methane emissions from the northern high latitudes from 1990 to 2009 using artificial neural networks, Global Biogeochemical Cycles, 27, 592-604, 2013b.

- 80 Zhuang, Q., Melillo, J. M., Kicklighter, D. W., Prinn, R. G., McGuire, A. D., Steudler, P. A., Felzer, B. S., and Hu, S.: Methane fluxes between terrestrial ecosystems and the atmosphere at northern high latitudes during the past century: A retrospective analysis with a process-based biogeochemistry model, Global Biogeochemical Cycles, 18, GB3010, doi:3010.1029/2004GB002239., 2004.
- Zhuang, Q., Melillo, J. M., Sarofim, M. C., Kicklighter, D. W., McGuire, A. D., Felzer, B. S., Sokolov,
 A., Prinn, R. G., Steudler, P. A., and Hu, S.: CO₂ and CH₄ exchanges between land ecosystems and the atmosphere in northern high latitudes over the 21st century, Geophysical Research Letters, 33, L17403, doi:10.1029/2006GL026972, 2006.

Zona, D., Gioli, B., Commane, R., Lindaas, J., Wofsy, S.C., Miller, C.E., Dinardo, S.J., Dengei, S.,
Sweeney, C., Karion, A., Chang, R.Y.-W., Henderson, J.M., Murphy, P.C., Goodrich, J.P., Moreaux,
V., Liljedahi, A., Watts, J.D., Kimball, J.S., Lipson, D.A. & Oechel, W.C. (2016) Cold season
emissions dominate the Arctic tundra methane budget. Proceedings of the National Academy of

90

Sciences, 113, 40-45.

70

Table 1. Terrestrial ecosystem models for CH₄ cycling and the model representation of three pathways
 of CH₄ transport (models are in alphabetical order; author's last name is used if the model name is not available)

Model	Aerenchynma	Diffusion	Ebullition	References
Beckett model	Yes	Yes	No	(Beckett et al., 2001)
Cartoon model	Yes	Yes	Yes	(Arah and Stephen, 1998; Arah and Kirk, 2000)
CASA	Yes	Yes	Yes	(Potter, 1997; Potter et al., 1996)
CH4MOD	Yes	Yes	Yes	(Huang et al., 1998b; Huang et al., 2004; Li et al., 2012)
Christensen model	No	No	No	(Christensen et al., 1996)
CLASS	No	Yes	No	(Curry, 2009; Curry, 2007)
CLM4Me	Yes	Yes	Yes	(Riley et al., 2011)
CLM-Microbe	Yes	Yes	Yes	(Xu et al., 2015; Xu et al., 2014)
DAYCENT	No	Yes	No	(Del Grosso et al., 2002; Del Grosso et al., 2009; Del Grosso et al., 2000)
Ding model	Yes	No	No	(Ding and Wang, 1996)
DLEM	Yes	Yes	Yes	(Tian et al., 2010; Xu and Tian, 2012)
DNDC	Yes	Yes	Yes	(Li, 2000b)
DOS-TEM	Yes	Yes	Yes	(Fan et al., 2013)
ecosys	No	Yes	Yes	(Grant, 2001, 1998)
Gong model	Yes	Yes	Yes	(Gong et al., 2013)
HH model	Yes	Yes	Yes	(Cresto-Aleina et al., 2015)
IAP-RAS	No	No	No	(Eliseev et al., 2008; Mokhov et al., 2007)

Kettunen model	Yes	Yes	Yes	(Kettunen, 2003)
Lovley model	No	No	No	(Lovley and Klug, 1986)
LPJ-Bern	Yes	Yes	Yes	(Spahni et al., 2011)
LPJ-WHyMe	Yes	Yes	Yes	(Wania et al., 2010, 2009)
LPJ-WSL	No	No	No	(Hodson et al., 2011)
Martens model	Yes	Yes	Yes	(Martens et al., 1998)
MEM	No	No	No	(Cao et al., 1995; Cao et al., 1998)
MERES	Yes	Yes	No	(Matthews et al., 2000)
Nouchi model	Yes	Yes	No	(Hosono and Nouchi, 1997; Nouchi et al., 1994)
ORCHIDEE	Yes	Yes	Yes	(Ringeval et al., 2010; Ringeval et al., 2011)
Ridgwell model	No	Yes	No	(Ridgwell et al., 1999)
SDGVM	No	No	No	(Hopcroft et al., 2011)
Segers model	Yes	Yes	Yes	(Segers and Kengen, 1998; Segers and Leffelaar, 2001a, b; Segers et al., 2001)
Tagesson model	No	No	No	(Tagesson et al., 2013)
TCF	Yes	Yes	Yes	(Watts et al., 2014)
TEM	Yes	Yes	Yes	(Zhuang et al., 2004)
TRIPLEX-GHG	Yes	Yes	Yes	(Zhu Q. et al., 2014)
UW-VIC	Yes	Yes	Yes	(Bohn and Lettenmaier, 2010; Bohn et al., 2007)
van Bodegom model	Yes	Yes	Yes	(van Bodegom et al., 2000; Van Bodegom et al., 2001)
VISIT	Yes	Yes	Yes	(Inatomi et al., 2010; Ito and Inatomi, 2012)

De Visscher model	No	Yes	No	(De Visscher and Van Cleemput, 2003)
Walter model	Yes	Yes	Yes	(Walter and Heimann, 2000; Walter et al., 1996)
Xu model	Yes	Yes	Yes	(Xu et al., 2007)

Table 2. Key mechanisms/features of CH₄ processes and their representations in CH₄ models

Key mechanisms	Models
Methanogenesis	Cartoon model, CASA, CH4MOD, Christensen model, CLM4Me, CLM- Microbe, Ding model, DLEM, DNDC, DOS-TEM, <i>ecosys</i> , Gong model, IAP- RAS, Kettunen model, Lovley model, LPJ-Brn, LPJ-WHyMe, LPJ-WSL, Martens model, MEM, MERES, ORCHIDEE, SDGVM, Segers model, TCF, TEM, TRIPLEX-GHG, UW-VIC, van Bodegom's model, VISIT, Walter's model, Xu's model
Methanotrophy	Cartoon model, CASA, CLASS, CLM4Me, CLM- Microbe, DAYCENT, DLEM, DNDC, DOS-TEM, <i>ecosys</i> , Gong model, Kettunen model, LPJ-Bern, LPJ-WHyMe, Martens model, MEM, MERES, ORCHIDEE, Ridgwells model, SDGVM, Segers model, TCF, TEM, TRIPLEX-GHG, UW-VIC, van Bodegom's model, VISIT, De Visscher model, Walter model, Xu model
Anaerobic oxidation of CH ₄	CLM-Microbe, Martens model
Substrate (Acetate/DOC)	CH4MOD, CLM-Microbe, DLEM, DNDC, <i>ecosys</i> , Gong model, Kettunen model, Lovley model, Martens model, MEM, MERES, SDGVM, Segers model, TCF, van Bodegom model, Xu model
Microbial functional groups	CLM-Microbe, , ecosys, Segers model
CH ₄ storage in soil profile	Beckett model, Cartoon model, CLM4Me, CLM-Microbe, <i>ecosys</i> , Kettunen model, Martens model, MERES, Nouchi model, ORCHIDEE, Segers model, UW-VIC, van Bodegom model, VISIT, De Visscher model, Walter model
O ₂ availability for CH ₄ oxidation	Beckett model, Cartoon model, CLM4Me, CLM-Microbe, <i>ecosys</i> , Kettunen model, MERES, Segers model, van Bodegom model, De Visscher model, Xu model
Iron biogeochemistry	van Bodegom model
Sulfate biogeochemistry	Lovley model, Martens model, van Bodegom model
Frozen trapped CH ₄	None
Embedded in Earth System Model	CLASS, CLM4Me, CLM-Microbe, IAP-RAS, ORCHIDEE, SDGVM
Vertical resolved	Beckett model, Cartoon model, CLASS, CLM4Me, CLM-Microbe, DNDC,

⁴⁷

biogeochemistry	DOS-TEM, <i>ecosys</i> , Gong model, HH model, IAP-RAS, Kettunen model, Lovley model, LPJ-Bern, LPJ-WHyMe, LPJ-WSL, Martens model, MERES, ORCHIDEE, Ridgwell model, SDGVM, Seger model, TRIPLEX-GHG, UW- VIC, VISIT, De Visscher model, Walter model, Xu model
Regional-scale, capacity for up- scaling	CASA, CH4MOD, Christensen model, CLASS, CLM4Me, CLM-Microbe, DAYCENT, DLEM, <i>ecosys</i> , Gong model, HH model, IAP-RAS, LPJ-Bern, LPJ-WHyMe, LPJ-WSL, Martens model, MEM, MERES, ORCHIDEE, Ridgwell model, SDGVM, Tagesson model, TCF, TEM, TRIPLEX-GHG, UW-VIC, VISIT, Walter model

Table 3. The mathematical equations used to described the CH₄ processes used in representative models (P_{CH4} is the CH₄ production rate; $O_{xid_{CH4}}$ is the CH₄ oxidation rate; T_{CH4} is the CH₄ transport rate; D_{CH4} is the CH₄ diffusion rate; some parameter may have been changed from original publication to keep relatively consistent in this table)

CH ₄ processes		Equations	Ecological description	Model examples
CH ₄ substrate and CH ₄ production	1	$P_{CH_4} = f(T, W)$	A function of temperature (T) and moisture (W)	Christensen model, IAP-RAS, DAYCENT
	2a	$P_{CH_4} = r \times HR \times f(T, W)$	A portion of heterotrophic respiration, affected by temperature (T) and moisture (W)	LPJ family, CLM4Me, Ding model, MERES, TRIPLEX-GHG
	2b	$P_{CH_4} = r \times SOM \times f(T, W)$	A portion of soil organic matter (SOM), affected by temperature (T) and moisture (W); Walter's model use indirect association with NPP	CH4MOD, DOS- Tem, Gong model, HH model, Walter model
	3	$P_{CH_4} = V \times \frac{[DOC]}{K_{DOC} + [DOC]} \times f(T, W)$	A portion of dissolved organic carbon (DOC), affected by temperature (T) and moisture (W)	MEM, DLEM
	4	$P_{CH_4} = f(DOC, Acetate, CO_2) \\ \times f(T, W)$	Mechanistic processes for CH ₄ production are considered, affected by temperature (T) and moisture (W)	Kettunen model, Segers model, van Bodegoms model, and <i>ecosys</i>
CH ₄ oxidation	5	$Oxid_{CH_4} = V \times \left(\frac{[CH_4]}{K_{CH_4} + [CH_4]}\right) \times f(T, W)$	Oxidation as a function of CH ₄ concentration and temperature and moisture	DLEM, TRIPLEX-GHG, VISIT

	(Quid	Oxidation as a function of	Conto on model
	6	Oxid _{CH4}		Cartoon model,
		= V	CH_4 and O_2 concentration,	CLM4Me, CLM-
		$\times \left(\frac{[CH_4]}{K_{CH_4} + [CH_4]}\right) \left(\frac{[O_2]}{K_{O_2} + [O_2]}\right)$	temperature and moisture	Microbe,
		$\left(K_{CH_4} + [CH_4] \right) \left(K_{O_2} + [O_2] \right)$		Kettunen model
		$\times f(T,W)$		
	7			DI FM DNDC
CH ₄ transport	7	$T_{CH_4} = V * ([CH_4] - \overline{[CH_4]})$	V is the parameter for	DLEM, DNDC,
			distance, diffusion	Walter model
			coefficient, etc.; [CH ₄] is	
			the concentration of CH ₄ in	
			the soil/water profile	
			(dissolvability for DLEM, 0	
			for DNDC); and $\overline{[CH_4]}$ is	
			the threshold of CH ₄	
			concentration above which	
			CH ₄ will be transported to	
			the atmosphere via either of	
			the three transport	
			pathways	
	8a	C(z) = C	Aerenchyma transport	CLM4Me
	oa	$A = \frac{C(z) - C_a}{r_L z / r_a} p T \rho_r$	Aerenchyma transport	CLIVITIVIC
		$r_L z / p + r_a$		
	8b		Ebullition	CLM4Me
	00	Moves to first unsaturated layer and	2000000	02111110
		then released to gaseous phase		
	8c	$\Delta[CH_{A}]$	Diffusion of CH ₄ was	CLM4Me, CLM-
		$D_{CH_4} = D \times \frac{\Delta [CH_4]}{\Lambda z}$	simulated following Fick's	Microbe, <i>ecosys</i> ,
			law; CLM4Me separate	
			aqueous and gaseous	Ridgwell model,
			diffusion	TRIPLEX-GHG;
				Sergers model
Temperature	9	$f(T) = a \times T + b$	Linear regression on	DAYCENT,
effects	,		temperature or degree days;	DNDC, IAP-
0110005		$f(T) = a \times T^2 + b \times T + c$	DNDC simulate	RAS, LPJ family
		$f(T) = a \times T^{2} + b \times T + c$ $f(T) = b \times e^{0.2424 \times T}$	temperature impact on	1210, 110 mining
		$f(I) = b \times e^{-1}$	production not on oxidation	
		(77, 77, ->	•	
	10	$f(T) = Q_{10} \frac{(T - T_{ref})}{10}$	Q_{10} equations; T_{ref} is the	CH4MOD,
			reference temperature	CLM-Microbe,
		l		<u> </u>

				CLM4Me, DLEM, VISIT, Kettunen model
	11a	$V_T = V^0 \times \exp\left(\frac{\Delta E}{R} \left[\frac{1}{T^0} - \frac{1}{T}\right]\right)$	Arrhenius equation	Cartoon model, Ding model
	11 b	$f_{T} = \frac{T_{s} \times \exp\left(A - \frac{H_{a}}{R \times T_{s}}\right)}{\left[1 + \exp\left(\frac{H_{al} - S \times T_{s}}{R \times T_{s}}\right) + \exp\left(\frac{S \times T_{s}}{R \times T_{s}}\right)\right]}$	Modified Arrhenius equation; T_s is soil temperature at K ; A is the parameter for $f_T = 1.0$ at T_s = 303.16 K; H_a is the energy of activation (J mol ⁻¹); R is universal gas constant (J mol ⁻¹ K ⁻¹); H_{dl} and H_{dh} are energy of low and high temperature deactivation (J mol ⁻¹)	ecosys
Moisture effects on methanogene sis and methanotroph	12	No moisture effect is simulated, rather inundation area is simulated	No equation, while a temporal and spatial variation of inundation and saturation impacts	CASA
y	13	$F_{\vartheta} = e^{-P/P_c}$	Water stress for oxidation, where P is soil moisture and Pe= -2.4×10^5 mm	CLM4Me
	14	$f(SM) = \begin{cases} 1, \\ \left[1 - \frac{\log_{10}\varphi - \log_{10}(0.2)}{\log_{10}(100) - \log_{10}(0.2)}\right]^{\beta} \\ 0, \end{cases}$	β is an arbitrary constant, Φ is the soil water potential	CLASS
	15	$f_{prod}(SM) = \left(\frac{SM - SM_{fc}}{SM_{sat} - SM_{fc}}\right)^{2}$ $\times 0.368$ $\times e^{\left(\frac{SM - SM_{fc}}{SM_{sat} - SM_{fc}}\right)}$ $f_{oxid}(SM) = 1 - f_{prod}(SM)$	Different impacts on CH_4 production and consumption; SM: soil moisture; SM_{fc} : field capacity; SM_{sat} : saturation soil moisture	DLEM

	16	$f(SM) = \frac{(M_V - M_{min}) \times (M - M_n)}{(M_V - M_{min}) \times (M_V - M_{max}) - (M_V - M_{max})}$	Bell-shape curve	TEM
pH effects	17a	$f(pH) = \frac{(pH - pH_{min}) \times (pH - pH_{min})}{(pH - pH_{min}) \times (pH - pH_{max})} -$	Bell-shape curve	CLM-Microbe, MEM, TEM,
	17 b	$f(pH) = 10^{-0.2335 \times pH^2 + 2.7727 \times pH - 8.6}$	Bell-shape curve	CLM4Me
	17c	$f(pH) = \begin{cases} 0 & pH \le 4 \\ \frac{1.02}{1 + 100000 \times e^{(-2.5 \times pH)}} \\ \frac{1.02}{1 + 100000 \times e^{(-2.5 \times (14 - pH))}} \end{cases}$	Bell-shape curve	DLEM

Table 4. Temperature dependence of CH ₄ processes in various models (blank indicates the Q ₁₀ function	ion
is not used; all temperatures are expressed as °C, 273.15 was used for unit conversion)	

Model	Q ₁₀	Reference temperature (°C)	Note	Sources
CASA			Based on a linear equation with temperature	(Potter, 1997)
DAYCENT			Linear equation $y =$ 0.209 * T + 0.845	(Del Grosso et al., 2000)
LPJ family			Linear function was used	(Hodson et al.,
LPJ-Bern			for temperature impacts on diffusion	2011; Spahni et al., 2011;
LPJ-WHyMe			on unrusion	Wania, 2007)
LPJ-WSL				
Christensen's model	2	2	For temperature > 0, the temperature impact is set to zero when < 0	(Christensen and Cox, 1995)
CH4MOD	3	30	T=30 for $30 < T \le 40$	(Huang et al., 1998b)
CLM4Me	2	2	Parameters for baseline simulation	(Riley et al., 2011)
CLM-Microbe	1.5	13.5		(Xu et al., 2015)
DLEM	2.5	30	For a temperature range of [-5, 30]; temperature impact is set to zero when < -5 or > 30	(Tian et al., 2010)
Kettunenn's model	4.0 for production, 2.0 for oxidation	10	Standard Q ₁₀ function	(Kettunen, 2003)
ORCHIDEE	Abisko site, 2.6; Michigan site, 3.2; Panama site, 1.2	Mean annual temperature	Q ₁₀ function with different parameters across biomes	(Ringeval et al., 2010)

TEM	Alpine tundra: wetland, 3.5; upland, 0.8. Wet tundra: wetland, 2.2; upland, 1.1. Boreal forest: wetland, 1.9; upland, 1.5	Alpine tundra: wetland, -3.0; upland, 8.0. Wet tundra: wetland, - 5.5; upland, 8.0. Boreal forest: wetland, 1.0; upland, 7.0	Q ₁₀ function with different parameters across biomes	(Zhuang et al., 2004)
TRIPLEX- GHG	1.7-16 for production, 1.4- 2.4 for oxidation	25 for optimal, 45 for highest temperature	Modified Q_{10} equation	(Zhu et al., 2014a)
VISIT		Mean annual temperature		(Ito and Inatomi, 2012)
Walter's model	2	Ombrotrophic bog, 12; poor fen, 6.5; oligotrophic pine fen, 3.5; Arctic tundra, 0; swamp, 27	Q ₁₀ function with different parameters across biomes	(Walter and Heimann, 2000)
Cartoon model model		10	Arrhenius equation	(Arah and Stephen, 1998)
ecosys		30	Modified Arrhenius equation	(Grant et al., 1993)

Figure legend

10 Figure 1. The published CH₄ models and modeling trends in terms of applicability and mechanistic representation of CH₄ cycling processes at decadal-scale and the envisioned CH₄ model capability

Figure 2. Percentage of CH₄ models with consideration of some key CH₄ mechanisms. The percentage was calculated as the number of models considering each mechanisms divided by the total number of published models in each time period

15 Figure 3. Cluster analysis showing thee groups of CH₄ models based on model characteristics (lines with same color indicate CH₄ models in same group; green lines represent relatively simple model structure, red lines represent relatively mechanistic models, blue lines represent mechanistic models)

Figure 4. Three types of models with key mechanisms for CH₄ production and oxidation (*SOM*: Soil organic matter; *NPP*: net primary production; *DOC*: dissolved organic carbon; O_{atm} : oxidation of atmospheric CH₄; *P*: plant-mediated transport; *D*: diffusion transport; *E*: ebullition transport; O_{xid} : oxidation; O_{trans} : oxidation of CH₄ during transport)

Figure 5. Key features of future mechanistic CH_4 models with a full representation of primary CH_4 processes in the terrestrial ecosystems. The data assimilation system and model benchmarking system are also shown as auxiliary components to the future CH_4 models

25



Lovely's model in 1986	Nouchi in 1994; MEM in 1995; Christensen, 1996; Ding in 1996; Walter's model in 1996; CASA in 1996, 1997; <i>ecosys</i> in 1998; Martens' model in 1998; CH4MOD in 1998; Segers in 1998; Ridgwell in 1999	Cartoon in 2000; CLASS in 2000; DAYCENT in 2000; DNDC in 2000; MERES in 2000; Beckett in 2011; De Visscher in 2003; IAP-RAS in 2007; TEM in 2004; Kettunen in 2003; LPJ- WHyMe in 2009; ORCHIDEE in 2008; PEATLAND-VU in 2006; UW-VIC in 2007; van Bogedom in 2001; LPJ- WHyMe in 2007; Xu in 2007	CLM-Microbe in 2015; TCF in 2014; DLEM in 2010; CLM4Me in 2011; DOS-TEM in 2013; Gong's model in 2013; HH model in 2015; Tagesson in 2013; TRIPLEX-GHG in 2014; ORCHIDEE-2010; VISIT in 2010; LPJ- Bern in 2011; LPJ-WSL in 2011	
1980s Mechanistic models for understanding CH ₄ processes	1990s Mechanistic models for understanding CH ₄ processes; plot- and regional simulations for quantifying CH ₄ budget	2000s Plot-level model development for CH4 cycling; and regional model for quantifying CH4 budget; mechanistic models for understanding CH4 processes;	2010s Regional model for quantifying CH_4 budget; mechanistic models for understanding CH_4 processes; plot-level model development for CH_4 cycling	Integrative models capable to fuse multiple sources data; mechanistic model with primary CH ₄ cycling including production, oxidation, transport, and environmental controls

Theoretical analysis; mechanistic understanding

Applicable on budget estimation at plot- and regional-scales; integration tool

Fig. 1.

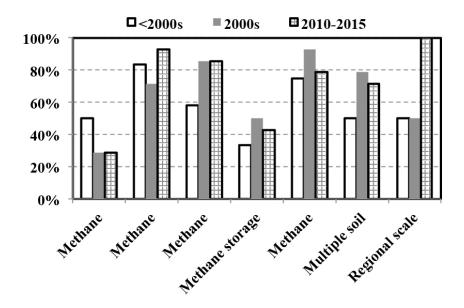


Fig. 2.

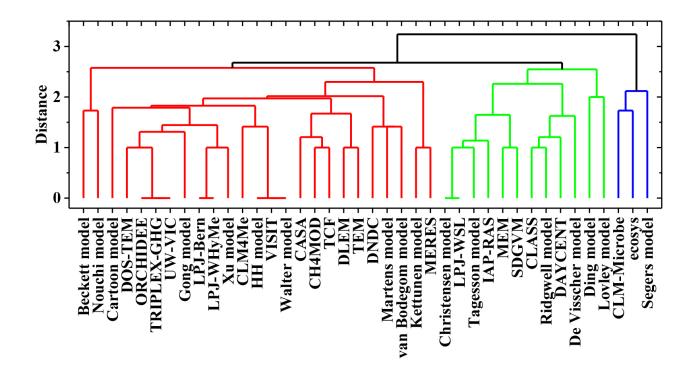


Fig. 3

58

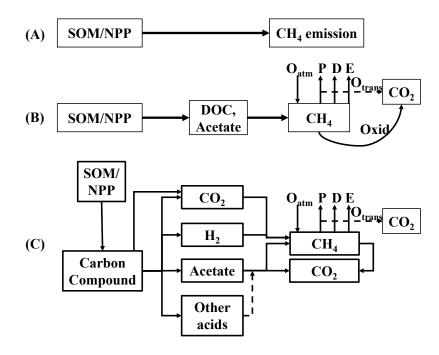


Fig. 4.

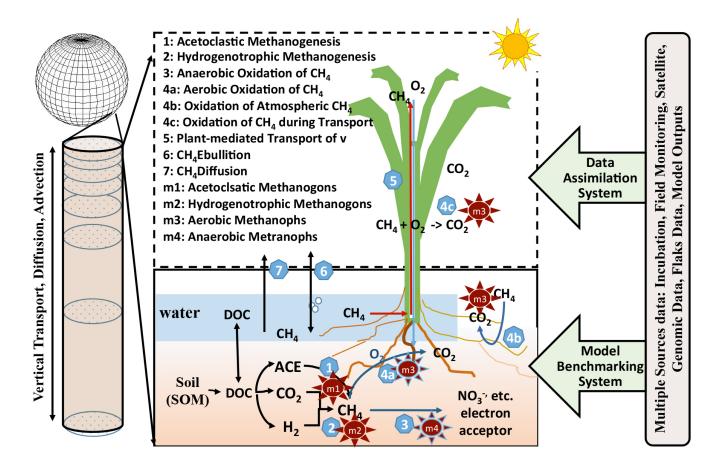


Fig. 5