

Quantification of dynamic soil-vegetation feedbacks following an isotopically labelled precipitation pulse

Arndt Piayda^{*,1}, Maren Dubbert^{*,2}, Rolf Siegwolf³, Matthias Cuntz⁴, Christiane Werner²

5 *equal contribution

¹Thünen Institute of Climate-Smart Agriculture, Braunschweig, 38116, Germany

²Ecosystem Physiology, University Freiburg, Freiburg, 79110, Germany

10 ³Lab for Atmospheric Chemistry, Ecosystems and Stable Isotope Research, Paul Scherrer Institut, Villingen PSI, 5232, Switzerland

⁴UMR Ecologie et Ecophysiologie Forestières, UMR1137, INRA-Université de Lorraine, Champenoux-54500 Vandoeuvre Les Nancy, 54280, France

Correspondence to: Arndt Piayda (arndt.piayda@thuenen.de), Maren Dubbert (maren.dubbert@cep.uni-freiburg.de)

15 **Abstract.** The presence of vegetation alters hydrological cycles of ecosystems. Complex plant-soil interactions govern the fate of precipitation input and water transitions through ecosystem compartments. Disentangling these interactions is a major challenge in the field of ecohydrology and pivotal foundation for understanding the carbon cycle of semi-arid ecosystems. Stable water isotopes can be used in this context as tracer to quantify water movement through soil-vegetation-atmosphere interfaces.

20 The aim of this study is to disentangle vegetation effects on soil water infiltration and distribution as well as dynamics of soil evaporation and grassland water use in a Mediterranean cork oak woodland during dry conditions. An irrigation experiment using $\delta^{18}O$ labeled water was carried out in order to quantify distinct effects of tree and herbaceous vegetation on infiltration and distribution of event water in the soil profile. Dynamic responses of soil and herbaceous vegetation fluxes to precipitation regarding event water use, water uptake depth plasticity and contribution to ecosystem soil evaporation and
25 transpiration were quantified.

Total water loss to the atmosphere from bare soil was as high as from vegetated soil, utilizing large amounts of unproductive evaporation for transpiration, but infiltration rates decreased. No adjustments of main root water uptake depth to changes of water availability could be observed during the experiment. This forces understory plants to compete with adjacent trees for water in deeper soil layers at the onset of summer. Thus, understory plants are subjected faster to chronic water deficits,
30 leading to premature senescence at the onset of drought. Despite this water competition, the presence of cork oak trees fosters infiltration and reduces evapotranspirative water losses from understory and soil, both due to altered micro climatic conditions under crown shading. This study highlights complex soil-plant-atmosphere and inter-species interactions controlling rain pulse transitions through a typical Mediterranean savannah ecosystem, disentangled by the use of stable water isotopes.

1 Introduction

Vegetation influences ecosystem water cycling in many ways. Rainfall is intercepted while at the same time infiltration, redistribution and translatory flow might be altered depending on rooting depths and soil structure (Bhark and Small, 2003; Dawson, 1993; Devitt and Smith, 2002; Dubbert et al., 2014c; Schwinning and Ehleringer, 2001; Tromble, 1988). E.g., a dense vegetation layer can strongly reduce soil evaporation (Dubbert et al., 2014c; Wang et al., 2012). In turn, plant transpiration is controlled by soil water availability and distribution and plant species have different abilities to use different soil water pools (i.e. surface vs. deep or ground water). Large parts of ecosystem water losses by transpiration strongly depend on plant functional types, stomatal regulation and leaf area index (*LAI*). Although studies within the last decades emphasized the pivotal role of plant roots for soil water redistribution or the role of plant transpiration on ecosystem water losses (Caldwell, 1987), it remains a major challenge to quantify dynamic soil-vegetation-atmosphere feedbacks within the water cycle.

Stable water isotopes are widely used to trace water transfers in soils, through plants and at the soil-vegetation-atmosphere interface (Werner and Dubbert, 2016; Yakir and Sternberg, 2000). Fractionation between the heavier and lighter isotopes occurs during phase changes (from liquid to gaseous, equilibrium fractionation) and movement (kinetic fractionation). This leads to different stable isotope compositions (δ^2H and $\delta^{18}O$) in various water pools (i.e. rain, groundwater), along soil profiles, in different plant species and between water vapour evaporated from soil compared to water transpired by plants. These differences provide the basis for tracing water through an ecosystem. Utilization of different water pools within the soil by different plant individuals may be possible (Dawson, 1993; Volkmann et al., 2016a). Isotopes can further help to separate transpiration from soil evaporative fluxes (Dubbert et al., 2013; Yepez et al., 2003) or to study infiltration or distribution of precipitation in soils (Garvelmann et al., 2012; Rothfuss et al., 2015). Stable water isotopes have also been used to study water movement at the soil-vegetation interface (Caldwell et al., 1998). The isotopic composition of plant water uptake can be determined by sampling the ‘output’ of the root system, for example the plant xylem, because the water isotopic signatures are usually not altered by plant water uptake (Dawson, 1993). Compared with values observed in the soil water profile, the preferential plant extraction depth or the proportional use of “event water” (i.e. singular precipitation events) can be determined. Although this method has been successfully used to identify processes such as hydraulic lift and soil water redistribution (Caldwell et al., 1998), most datasets were limited in temporal and spatial resolution (Asbjornsen et al., 2008; Kurz-Besson et al., 2006). Over the last decade, the development of field-deployable laser spectroscopy has enabled continuous measurements of water vapour and its isotopic signatures in ecosystem fluxes and atmospheric concentrations. This opens the door for large scale assessment of the soil-vegetation-atmosphere interactions in the water cycle. In particular, these developments have enhanced the spatial and temporal resolution tremendously, furthering the understanding in the fields of plant ecophysiology (Cernusak et al., 2016) and ecosystem physiology (Dubbert et al., 2014a; Dubbert et al., 2014c).

In the present study, we focus on disentangling the vegetation effects on soil water infiltration and distribution as well as dynamics of soil evaporation and grassland water-use in a Mediterranean cork oak woodland. An irrigation experiment with $\delta^{18}O$ labeled water was carried out to quantify the distinct effects of tree and herbaceous vegetation on 1) infiltration and distribution of “event water” (freshly introduced water) in the soil profile and 2) to quantify the dynamic responses of soil and herbaceous vegetation fluxes to precipitation regarding event water-use, plasticity of water uptake depth and contribution to ecosystem *ET*. The following hypotheses were tested:

I. Presence of understory vegetation increases evapotranspirative water loss compared to bare soil, but foster infiltration due to shading.

II. Preferential root water uptake depth of understory plants is unaffected by changes in soil water availability after rain pulses during drought.

III. Tree shading fosters infiltration of event water and reduces evapotranspiration generating favourable soil moisture conditions for understory plants.

2 Material and methods

2.1 Study site and experimental design

Measurements were conducted between May 26 and June 6 2012 in an open cork oak woodland (*Quercus suber* L.) in central Portugal, approximately 100 km north-east of Lisbon (N39°8'17.84'' W8°20'3.76''; Herdade de Machoqueira do Grou). The trees are widely spaced (209 individuals ha⁻¹) with a leaf area index of 1.1 and a gap probability of 0.7 (Piayda et al., 2015).

The herbaceous layer is dominated by native annual forbs and grasses. The site is characterized by Mediterranean climate, with a 30 year long-term mean annual temperature of approximately 15.9 °C and annual precipitation of 680 mm (Instituto de Meteorologia, Lisbon). We established two sites: one directly under the oak crown projected area (tree site, ts) and another one in an adjacent open area (open site, os). Two types of plots (sized 40 × 80 cm) were installed in each site: bare soil plots with total exclusion of above and below-ground biomass (lateral root ingrowth was prevented by vertically inserted trenching meshes around the plots, mesh diameter < 1 μm, Plastok, Birkenhead, UK), and understory plots with herbaceous vegetation (four plots per site and treatment). The sites were kept vegetation free just by regular weeding. We expect no influence of the mesh on infiltration, since the plots were installed one year before the experiment and processes like preferential flow along the mesh is unlikely (For further details see Dubbert et al. (2013)).

After a base line observation, all plots were watered with 20 mm water within one hour using watering cans. The water showed an oxygen isotopic signature of -139.5‰ to trace the influence of different vegetation components on water infiltration. All plots and the surrounding soil were watered equally to avoid lateral gradients and possible differences between trenched and control plots. Thereafter, all measurements were conducted in 7 diurnal cycles over the following 10-12 days. The open and tree sites were watered independently, as the measurement setup did not allow highly resolved

observations of all treatment plots at the same time. Environmental variables (PPFD; soil water content; vpd) were not significantly different between the first and second half of the observation period.

2.2 Environmental variables and plant parameters

Photosynthetic photon flux density (PPFD) was measured at both sites at approximately 1.5 m height (PPFD, LI-190SB, LI-COR, Lincoln, USA). Rainfall (ARG100 Rain gauge, Campbell Scientific, Logan, UT, USA), air temperature, and relative humidity (rH, CS-215 Temperature and Relative Humidity Probe, Campbell Scientific, Logan, UT, USA) were measured and 30 min averages were stored by a data logger (CR10x, Campbell Scientific, Logan, UT, USA). Soil temperature (custom built pt-100 elements) was measured at -5 cm depth on vegetation and bare soil plots at both sites and 60 min averages were stored in a data logger (CR1000, Campbell Scientific, Logan, UT, USA; 4 sensors per depth and treatment). Temperature at the soil surface was manually measured on each measurement day in diurnal cycles corresponding with the gas exchange measurements using temperature probes (GMH 2000, Greisinger electronic, Regenstauf, Germany). Volumetric soil water content (θ_v , 10hs, Decagon, Washington, USA) was measured in -5, -15, -30 and -60 cm depth on vegetation and bare soil plots at both sites and 60 min averages were stored in a data logger (CR1000, Campbell Scientific, Logan, UT, USA; 4 sensors per depth and treatment).

Living aboveground biomass of herbaceous plants was determined destructively on five randomly selected, 40 × 40 cm plots at the beginning and end of the experiment in the open and under the trees. All green fresh aboveground plant biomass was collected, divided by species, dried (60 °C, 48 hours) and weighed. Below ground biomass was sampled with soil cores in -5, -15, -30, and -60 cm depth. Oven dried soil was sieved and root biomass was determined gravimetrically. 80 % of root biomass was distributed between -5 to -15 cm depth. Only 5% was distributed above -5 cm and 15% between -20 to -35 cm depth. Total aboveground biomass was relatively low compared to previous years between 42 and 78 g m⁻² (see Fig. A1) with a minimal fraction of dry biomass due to the considerable winter/spring drought in the hydrological year 2012 (Costa e Silva et al., 2015; Dubbert et al., 2014b; Piayda et al., 2014). Dry biomass from the previous season was removed from the plots at the end of summer 2011. While total aboveground biomass was similar between plots, species composition and relative dominance differed with the open sites being dominated by *Tuberaria guttata* and the tree sites by grass and legume species (Dubbert et al., 2014b).

2.3 Cavity Ring-Down Spectrometer based gas-exchange flux and $\delta^{18}\text{O}$ measurements

Water fluxes and isotopic composition were measured with a Wavelength Scanned Cavity Ring-Down Spectrometer (WS-CRDS, Picarro, Santa Clara, USA) in combination with custom built soil chambers (following the design of Pape et al. (2009)) in an open gas exchange system (n=3 per treatment and experimental site). Background and sampling air were measured alternately after stable values were reached. A five minutes interval average was used for the calculation of evapotranspiration (*ET*) and evaporation (*E*). *ET* and *E* were calculated according to von Caemmerer and Farquhar (1981).

Oxygen isotope compositions of soil evaporation (bare soil plots) as well as evapotranspiration of the understory (vegetation plots) were estimated using a mass balance approach (Dubbett et al., 2013; Dubbett et al., 2014c):

$$\delta_E = \frac{u_{out}w_{out}\delta_{out} - u_{in}w_{in}\delta_{in}}{u_{out}w_{out} - u_{in}w_{in}} \quad (1),$$

$$\delta_E = \frac{w_{out}\delta_{out} - w_{in}\delta_{in}}{w_{out} - w_{in}} - \frac{w_{in}w_{out}(\delta_{out} - \delta_{in})}{w_{out} - w_{in}} \quad (2),$$

where u is the flow rate [$\text{mol}(\text{air}) \text{ s}^{-1}$], w is the mole fraction [$\text{mol}(\text{H}_2\text{O}) \text{ mol}(\text{air})^{-1}$] and δ is isotope value of the incoming (in) and outgoing (out) air stream of the chamber. Flow rates are measured with humid air so that conservation of dry air gives $u_{in}(1-w_{in}) = u_{out}(1-w_{out})$, which leads to Eq. (2). The second term in Eq. (2) corrects for the increased gas flow in the chamber due to addition of water by transpiration. In addition to isotopic signatures of soil evaporation and understory evapotranspiration, the oxygen isotope signatures of ambient water vapor (in 9 m height) were measured with the CRDS. All measurements were conducted as diurnal courses with 5-6 measurement points between 7 a.m. and 7 p.m. For more details about the chamber design and measurement setup see Dubbett et al. (2013).

2.4 Sampling and measurement of $\delta^{18}\text{O}$ of soil and leaf water

Soil samples for water extraction and $\delta^{18}\text{O}$ analysis were taken on vegetated and bare soil plots using a soil corer. Samples were collected from the soil surface (0-0.5 cm depth), -2, -5, -10, -15, -20, and -40 cm soil depths ($n=4$ per depth and treatment) usually during midday, but on the day of irrigation directly proceeding the irrigation pulse and additionally at 18:00. Mixed leaf samples of the herbaceous vegetation for water extraction were obtained in daily cycles in 2-hourly steps from 8:00 to 18:00 following a destructive sampling scheme affecting the overall amount of living biomass less than 5%. Thus, effects of destructive sampling on observed ET fluxes during the experiment are negligible. Soil and leaf water samples were extracted on a custom build vacuum line by cryogenic distillation. Water $\delta^{18}\text{O}$ analysis was performed by headspace equilibration on an Isoprime IRMS (Elementar, Hanau, Germany) coupled via open split connection to a μgas autosampler (Elementar, Hanau, Germany). Equilibration with 5% He gas was done for 24 hours at 20 °C. For every batch of 44 samples 3 different laboratory standards were analyzed. Laboratory standards were regularly calibrated against VSMOW, SLAP, and GISP water standards (IAEA, Vienna). Analytical precision was 0.1‰.

2.5 Partitioning of evapotranspiration

Oxygen isotope signatures of soil evaporation were calculated using the Craig and Gordon equation (Craig and Gordon, 1965; Dubbett et al., 2013; Haverd and Cuntz, 2010):

$$R_E = \frac{1}{\alpha_k \alpha^+ (1-h)} (R_e - \alpha^+ h R_a) \quad (3),$$

where R_E is the isotope ratio ($^{18}\text{O}/^{16}\text{O}$) of evaporated water vapor and R_e is the isotope ratio of bulk soil water at the evaporating sites. The evaporating site is the vapor-liquid interface below which liquid transport and above which vapor

transport is dominant (Braud et al., 2005). It has been shown for unsaturated soils that this site is related to a strong enrichment in soil water isotopic composition relative to the rest of the soil column and an exponential depletion in isotopic signature within few cm of the underlying soil due to evaporative enrichment of the remaining liquid water (Dubbert et al., 2013; Haverd and Cuntz, 2010). Thus, for R_e and temperature at the evaporating sites (T_e), temperature and oxygen isotope signatures of bulk soil water were measured along the soil profile and those values along the soil profile were used where the strongest enrichment in bulk soil $\delta^{18}O$ could be detected (residual soil water volumetric content was only 1% and therefore neglected). R_a is the isotope ratio of ambient water vapor, α_k is the kinetic fractionation factor, α^+ is the water vapor equilibrium fractionation factor (α_k and $\alpha^+ > 1$; see Majoube (1971); Merlivat (1978); for the formulation of $\alpha_k = \alpha_{diff}^{nk}$ see Mathieu and Bariac (1996)). h is the relative humidity normalized to T_e . R_E can then be transferred to delta notation as $\delta = R_E - 1 \times 1000$.

Although direct estimates of E and $\delta^{18}O_E$ were available for bare soil plots, vegetation depresses E and also influences $\delta^{18}O_E$, for example due to different isotopic signatures of soil water and also temperature at bare soil and vegetated soil patches (Dubbert et al., 2013). Therefore, bare soil plots only served to validate the Craig and Gordon equation, because on bare soil plots E contributes entirely to the evaporative flux and could be tested against modeling results. Finally, the Craig and Gordon equation was used to calculate $\delta^{18}O_E$ of vegetation plots.

The oxygen isotope signature of transpired water vapor $\delta^{18}O_T$ was calculated based on the isotopic signature of bulk leaf water $\delta^{18}O_L$ using the Craig and Gordon equation (Eq. 3) instead of measuring xylem/source water isotopic signatures and modeling $\delta^{18}O_L$ of leaf water at the evaporating sites due to the lack of suberized/lignified plant parts in the herbaceous vegetation. The isotopic signature on the evaporating site $\delta^{18}O_e$ was thus estimated by:

$$\delta^{18}O_e = \frac{\delta^{18}O_L \wp}{1 - e^{-\wp}} \text{ with the Péclet number } \wp = \frac{TL_{eff}}{CD} \quad (4),$$

where L_{eff} is the effective path length of water movement in the leaf mesophyll, which we assumed to be 0.05 m, C is the molar water concentration ($55.6 \times 10^3 \text{ mol m}^{-3}$) and D is the tracer diffusivity in liquid water ($2.66 \times 10^{-9} \text{ m}^2 \text{ s}^{-1}$). T was estimated iteratively with equation (Eq. 5) using ET as initial value. Convergence was generally reached after five iterations. Small differences in isotopic compositions were found compared to a direct use of $\delta^{18}O_L$ in equation (Eq. 3), which were not significant for results shown throughout this work.

Finally, the contribution of T to ET , $ft = T/ET$, can be estimated based on measured understory $\delta^{18}O_{ET}$ and modeled soil $\delta^{18}O_E$ and herbaceous $\delta^{18}O_T$ (Moreira et al., 1997; Yakir and Sternberg, 2000):

$$ft = \frac{\delta^{18}O_{ET} - \delta^{18}O_E}{\delta^{18}O_T - \delta^{18}O_E} \quad (5).$$

This approach is based on the assumption that the isotopic signature of evapotranspiration is a mixing ratio of not more than the two sources (evaporation and transpiration) and that no water vapor is lost other than by the mixing of the two sources with the atmospheric pool (i.e. no condensation).

2.6 Event water partitioning

Event water describes the amount of water in ecosystem pools or fluxes that originates from a certain rain event. To calculate the amount of event water in volumetric soil water content θ that originates from the isotopically labeled watering event, the following linear two source mixing model was used:

$$f_{\theta, eve} = \frac{\delta^{18}O_{\theta} - \delta^{18}O_{\theta, pre}}{\delta^{18}O_{eve} - \delta^{18}O_{\theta, pre}} \quad (6),$$

where $f_{\theta, eve}$ is the fraction of rain event water in θ at a certain time after the event, $\delta^{18}O_{\theta}$ is the stable isotope ratio in θ at a certain time after the event, $\delta^{18}O_{\theta, pre}$ is the stable isotope ratio of soil water before the rain event and $\delta^{18}O_{eve}$ is the stable isotope ratio of the precipitation event water. The model assumes no fractionation of rain event water during infiltration and was solved separately for each depth. Contributions of infiltrated event water to evaporation fluxes from soil and

transpiration fluxes from plant surfaces were calculated analogously:

$$f_{E, eve} = \frac{\delta^{18}O_E - \delta^{18}O_{E, pre}}{\delta^{18}O_{E, eve} - \delta^{18}O_{E, pre}} \quad (7),$$

$$f_{T, eve} = \frac{\delta^{18}O_T - \delta^{18}O_{T, pre}}{\delta^{18}O_{T, eve} - \delta^{18}O_{T, pre}} \quad (8),$$

where $f_{E, eve}$ and $f_{T, eve}$ are the fractions of rain event water in evaporation E and transpiration T . $\delta^{18}O_{E, pre}$ and $\delta^{18}O_{E, eve}$ are the isotopic compositions of evaporation calculated with equation (Eq. 3) assuming that the source water isotopic composition equals either $\delta^{18}O_{\theta, pre}$ at the evaporative site or $\delta^{18}O_{eve}$, respectively. $\delta^{18}O_{T, pre}$ and $\delta^{18}O_{T, eve}$ are the isotopic compositions of transpiration calculated with equation (Eq. 3) and (Eq. 4) assuming that the source water isotopic composition equals either bulk leaf composition before watering $\delta^{18}O_{L, pre}$ or $\delta^{18}O_{eve}$, respectively.

2.7 Root water uptake

The preferential depth of root water uptake by plants along the soil depth was estimated via a linear three source model.

Therefore, the isotopic composition of transpiration $\delta^{18}O_T$ calculated with equation (Eq. 3 and 4) from three independent observations of leaf water compositions $\delta^{18}O_L$ were compared with three independent solutions for isotopic transpiration composition $\delta^{18}O_T$ of equation (Eq. 3), each assuming the current water source for transpiration originating only from an observed depth ($d1 = -5$ cm, $d2 = -15$ cm, $d3 = -30$ cm). Soil depth above and below $d1$ to $d3$ showed negligible root density in the profile and could therefore be excluded from the model. The three possible source fluxes are related to the resulting transpiration flux mixture via the following system of equations (compare e.g. Philips et al. (2005)):

$$\begin{aligned} \delta^{18}O_{T1} &= f_{T, d1} \cdot \delta^{18}O_{T1, d1} + f_{T, d2} \cdot \delta^{18}O_{T1, d2} + f_{T, d3} \cdot \delta^{18}O_{T1, d3} + \varepsilon_1 \\ \delta^{18}O_{T2} &= f_{T, d1} \cdot \delta^{18}O_{T2, d1} + f_{T, d2} \cdot \delta^{18}O_{T2, d2} + f_{T, d3} \cdot \delta^{18}O_{T2, d3} + \varepsilon_2 \\ \delta^{18}O_{T3} &= f_{T, d1} \cdot \delta^{18}O_{T3, d1} + f_{T, d2} \cdot \delta^{18}O_{T3, d2} + f_{T, d3} \cdot \delta^{18}O_{T3, d3} + \varepsilon_3 \end{aligned} \quad (9),$$

$$1 = f_{T,d1} + f_{T,d2} + f_{T,d3}$$

where $f_{T,d}$ denotes the fraction of source water contribution from depths $d1$ to $d3$ to the transpiration flux. The system was solved for $f_{T,d1}$ to $f_{T,d3}$ using a shuffled complex evolution algorithm (Duan et al., 1992) minimizing a multi-objective cost function (Duckstein, 1981) combining the error terms ε_1 to ε_3 for each time step.

2.8 Error propagation

- 5 All results are reported as replicate mean with associated standard error to achieve comparability between different sample sizes. All model calculations were applied to single replica and averaged afterwards. Observed effects were considered statistically different when no overlap of standard errors was observed.

3 Results

3.1 Environmental and soil conditions

- 10 Tree cover significantly influenced diurnal courses of incoming global radiation R_g during daytime on the sites. Strong reductions of R_g between 09:00 and 18:00 o'clock reduced daily sum of energy input $\sum R_g$ by $17.1 \text{ MJ m}^{-2}\text{d}^{-1}$ on the open sites (os) compared to the tree sites (ts) (Fig. 1). However, air temperature and relative humidity was very similar in the open area and below trees with mean values around 66% and 19°C throughout the experiment. Similar to R_g , the amplitude of daily mean soil temperatures T_S in the upper soil layer were smaller on tree sites (bare: 7.4°C , veg: 5.5°C) than in the open area
15 (bare: 14.9° and 11.3°C for bare and vegetated soils, respectively, Fig. 1). In contrast, understory vegetation cover reduced the soil temperature only by $2\text{-}3.6^\circ\text{C}$ on both sites.

- Soil moisture θ prior to the irrigation pulse ranged from 5 to 10% (Fig. 3), which is low compared to the annual average, but typical for the observation period at the end of May and the beginning of the dry season. Systematically, lower soil moisture θ at depths below 20 cm could be observed at the tree sites located close to trees compared to open sites, whereas the upper
20 soil layers showed comparable values for all sites prior to the experiment.

3.2 Oxygen isotope signatures of ecosystem water pools

- Stable oxygen isotope composition of soil water $\delta^{18}\text{O}_S$ for all plots and all depths ranged between -7.3‰ and 10.1‰ before the irrigation. Compared to the very depleted irrigation water signature of -139.5‰ , only small enrichment in $\delta^{18}\text{O}_S$ of on average 0.4‰ in the open sites compared to the tree sites were found and 2.9‰ enrichment of bare soil compared to
25 vegetation plots preliminary to the watering (Fig. 2). Irrigation caused a strong depletion of $\delta^{18}\text{O}_S$ with a peak only 1 h after irrigation in the upper soil layer. Strongest depletion of $\delta^{18}\text{O}_S$ values were found at tree sites on bare soil plots with $\delta^{18}\text{O}_S = -106.06\text{‰}$ and tree sites with vegetation cover with $\delta^{18}\text{O}_S = -85.1\text{‰}$ whereas the open sites showed weaker maximum depletions of $\delta^{18}\text{O}_S = -79.9\text{‰}$ and $\delta^{18}\text{O}_S = -49.4\text{‰}$ on bare soil and vegetation plots, respectively. The nine days following

the irrigation event were characterized by a steady increase of $\delta^{18}O_S$, which was slightly depleted compared to pre-event $\delta^{18}O_S$ nine days after irrigation. In addition to the absolute differences in peak $\delta^{18}O_S$ between sites, the depletion in $\delta^{18}O_S$ was maintained for a longer period at tree sites (Fig. 2).

Oxygen isotope signatures of soil evaporation and leaf water as well as transpired water vapour (Fig. 4) showed an immediate response to the irrigation pulse with peak depletion only 1 hour after labelling for soil evaporation and 3 hours for leaf water and transpired vapour. Subsequently, an exponential rise to pre-event isotope values could be observed in all pools. Depletion in $\delta^{18}O_E$ of soil evaporation was much stronger compared to $\delta^{18}O_T$ of plant transpiration (and leaf water $\delta^{18}O_L$). $\delta^{18}O_E$ of soil evaporation and evapotranspiration $\delta^{18}O_{ET}$ were both more reduced on the tree sites compared to the open sites. A similarly strong vegetation effect could be seen between $\delta^{18}O_E$ on bare soil plots in comparison to understory vegetation plots.

3.3 Infiltration and distribution of event water

Daily mean soil moistures θ throughout the experiment were characterized by the ongoing drought at all sites (Fig. 3). Watering the plots with 20 mm increased mean daily soil moisture θ in the upper layers only by 2%_{vol.} to 6%_{vol.} and had no effect on deeper soil layers. However, partitioning event water fractions revealed an extensive replacement of old, pre-event water with new event water up to 4%_{vol.} and even down to depths below -30 cm (Fig. 3), in particular on bare soil plots. Systematically increased infiltration and deepened distribution of event water was observed on tree sites compared to open sites. In the course of the experiment, soil moistures returned to pre-event values and below. The decrease of event water was stronger than that of pre-event water, leaving nearly no trace nine days after the watering.

3.4 Event water use by plant transpiration

While pre-event E on bare soil plots was lower than ET on vegetation plots on both the open and tree sites, E and ET equally peaked with roughly 3.3 mmol m⁻² d⁻¹ on the open sites. However, on the tree sites post-event peak of E at bare soil plots (2.1 ± 0.1 mmol m⁻² d⁻¹) was higher than ET at vegetation plots (1.5 ± 0.2 mmol m⁻² d⁻¹). Moreover, the peak of ET on both sites was shifted by 24 h compared to E and occurred only 2 days after irrigation (Fig. 5). Following peaks in E and ET , evapotranspiration losses declined exponentially to pre-event values 3 days after irrigation on all sites. Partitioning ET on vegetation plots on both sites into soil E and plant transpiration T revealed that the time shift of the response of the ET flux compared to bare soil plots E was caused solely by a slower reaction of T to the irrigation pulse. Throughout the experiment the proportion of T to ET ranged from 9% to 59% on open sites and 17% to 66% on shaded sites. Event water fraction in soil evaporation $f_{E, eve}$ and plant transpiration $f_{T, eve}$ differed considerably with T utilizing only a peak of 12% of the event water while E is fed up to 62% by event water following irrigation (Fig. 6). Nine days after the irrigation pulse event water contribution of T and E converged on average to 10% of the respective flux and differences between $f_{E, eve}$ and $f_{T, eve}$ faded. Event water lost by soil evaporation $f_{E, eve}$ showed no significant differences between open and tree sites nor between bare soil plots and vegetated plots except on the day of watering on the open vegetation plot. Here, $f_{E, eve}$ reached

only about 25%, corresponding to the limited availability of event water in the soil (Fig. 2). No significant differences could be observed between $f_{T, eve}$ on open and vegetation plots.

3.5 Preferential root water uptake depth

Prior to the irrigation pulse we refrained from calculations of preferential root water uptake depth, since the differences in $\delta^{18}O_S$ along soil depth were too small (see above) for a sufficiently accurate prediction power solving equation system (Eq. 9) and derive significant $f_{T, d}$. Following the label pulse, soil water uptake by plants was located solely at soil depths around -30 cm with no change in time or between open and tree sites despite a small uptake of water for transpiration from soil layers around -15 cm on day 0 and 1 after watering (Fig. 7).

4 Discussion

4.1 Infiltration and distribution of event water

Mosaic patterns of vegetation cover by understory plants and trees are characteristic for savannah-type ecosystems (Belsky, 1994; Greig-Smith, 1979). Different vegetation cover is known to alter soil hydrological conditions and micro climate (Scholes and Archer, 1997) which in turn have effects on vegetation cover and ecosystem sustainability in future climate change scenarios (Breman and Kessler, 1999; Pueyo et al., 2012). Infiltration of event water into soil in this ecosystem is strongly altered by understory cover and tree shading. Vegetation cover of understory plants reduced infiltration on average by 24% compared to bare soil (Fig. 3), which clearly contradicts part two of hypothesis I. The reason can be found in interception, subject to instantaneous plant and litter surface evaporation before the first flux observations, which took place one hour after watering. This water uptake limitation could neither be compensated by plant roots, breaking the crust formations which can be observed in the field and are common for Mediterranean soils, limiting hydraulic conductivity of top soils (Eldridge et al., 2010; Goldshleger et al., 2002; Maestre et al., 2002) nor by beneficial shading effects by the above ground biomass, which did not significantly reduce the soil surface temperatures (Fig. 1) and thus, the evaporative demand of boundary layers. The observed infiltration on the day of watering can further be regarded as insignificantly affected by understory root water uptake confirmed by low transpiration fluxes on the day of watering (Fig. 5). This is in contrast to previous studies, which reported beneficial effects of plant cover on daily sums of infiltration during the same period at the onset of drought in 2011 (Dubbert et al., 2014c). However, Dubbert et al. (2014c) only observed precipitation events of light intensity during the period of interest. The present study reports on high intensity precipitation events. Furthermore, aboveground vegetation cover and biomass were reduced by 55 and 30 %, respectively, owing to the additional severe winter/spring drought in 2012. It is thus likely that such a drastic reduction in understory canopy cover eliminates much of the beneficial understory effects on the ecosystem water balance. This unexpected turn in effect direction with increasing precipitation intensity, which depends on vegetation cover and atmospheric evapotranspirative demand, potentially plays a

strong role for the water balance of the ecosystem in the course of ongoing climate change scenarios since the occurrence of extreme precipitation events is expected to increase (IPCC, 2013).

Tree shading had a tremendous impact on the microclimate above understory plant and soil surfaces, but effects on infiltration amount could only be observed on vegetated plots. Reductions of the daily sum of global radiation $\sum R_g$ by 72% and daily peak soil temperatures $T_{S,5cm}$ up to 22% (Fig. 1) generated favorable conditions. Limited instantaneous evaporation from plant surfaces as described above led to 71% higher infiltration amounts (Fig. 3), whereas the high infiltration amounts on bare soil plots were unaffected by tree shading. This confirms part one of hypothesis III on vegetated plots. Previous studies reported similar, positive feedbacks of tree cover for the hydrological cycle in savannah-type ecosystems related to shading effects (Eldridge and Freudenberger, 2005). Effects of altered soil hydraulic properties beneath tree crowns, like the amount of preferential flow fostering infiltration (Bargués Tobella et al., 2014) could not be identified in this study. Supporting findings are given by Bhark and Small (2003); D'Odorico and Porporato (2006). Considering the projected shading by crown cover of the tree layer (minimum of 30% at noon, increasing during the rest of the day (Piayda et al., 2015)), the infiltration enhancement has potentially large benefits on the ecosystem level. A former study of David et al. (2006) under comparable climatic and stand density conditions estimated only minor interception losses of 8% with respect to total canopy throughfall, due to low canopy cover typical for cork oak systems. However, the integral balance of canopy interception losses, increased infiltration and other benefits of tree cover (compare Joffre and Rambal (1993) and Dubbert et al. (2014c)) in this ecosystem could not be analyzed in this study and needs further investigations with regard to tree density and age.

Subsurface distribution of soil water θ was systematically lower at depths below -20 cm at tree sites compared to open sites (Fig. 3). This clearly indicates the enhanced water extraction by tree roots, similar to results of Dubbert et al. (2014b). The observed pattern could not be changed by the event water pulse of 20 mm per hour, equal to a rain event of high intensity on this site. That explains the intense drought stress understory plants suffer during the transition period from moist spring to dry summer, leading to earlier dieback under tree cover (Dubbert et al., 2014b; Moreno, 2008) and contradict part two of hypothesis III. The depth distribution of event water is very similar on bare soil plots that show an overall deeper infiltration of more water than the vegetated plots, caused by the higher infiltration amounts shown before. This negative effect could partially be compensated by higher infiltration amounts below tree shading, but was consumed by tree water uptake from deeper depths within one day. During these dry conditions, pre-event water is located in small pores under high matrix potentials. Infiltrating event water partially displaced pre-event water downwards (Fig. 3) and additionally filled larger pores in the top soil. Thus, event water is more subject to evaporation due to lower matrix potentials in bigger pores than pre-event water. This observation is supported by a rapid decrease of event water content throughout the experiment.

4.2 Dynamic responses of event water-use and plasticity of water uptake depth

Successful biomass production of herbaceous vegetation highly depends on soil water availability in upper soil layers hosting the root system. Occasional precipitation events control the soil water regime (Porporato et al., 2004) which are

prone to substantial changes in future climate change scenarios by stronger short term fluctuations of drought events (IPCC, 2013). Thus, a rapid adaptation of preferential root water uptake depth is crucial. This is particularly important for herbaceous vegetation in order to maximize the utilization of different soil water pools for a successful seed production, longevity and inter species competition (Ehleringer and Dawson, 1992; Rodriguez-Iturbe, 2000). It could be clearly shown that understory transpiration T responded slower to an incoming precipitation pulse than soil evaporation E , with a time lag of about 24h. ET on vegetated plots and E on bare soil plots showed equally high peaks and a comparable decline until the end of the experiment, providing no evidence for higher water losses due to the presence of understory and contradicting part one of hypothesis I. During the entire experiment, E was the dominant flux on both, tree and open sites, with a comparable contribution of transpiration T to evapotranspiration ET of 36% and 41% (Fig. 5), respectively. This small loss of transpiration water originates on one hand from the longer time response lag of T , on the other hand from only little event water reaching deeper soil layers, where understory plants have their main root water uptake depth. Event water use of the understory vegetation was overall low, since no shift of root water uptake depth could be observed within the nine days of the experiments (Fig. 7) leading to comparably small isotopic depletion of bulk leaf water and transpiration (Fig. 4), which supports hypothesis II. This is in agreement with previous findings where annual savannah species were not fast enough readjusting their water extraction depth in order to exploit precipitation water more efficiently (Asbjornsen et al., 2008; Kulmatiski and Beard, 2013). More importantly, during that period of the year the dry conditions in the upper soil layers forces understory plants in the direct vicinity of trees to compete for soil water at lower depths where the trees have their roots (i.e. tree sites). This observation clearly opposes the widely discussed two-layer hypothesis, proposing independent ecological niches for root water uptake of trees and understory plants in savannahs in order to avoid competition (Hipondoka et al., 2003; Holdo and Planque, 2013; Kulmatiski et al., 2010; Walter et al., 1971). Moreover, exponential soil profiles of plant available nitrogen causes a coupled water and nutrient competition between herbs and trees in this ecosystem during spring (Dubbart et al., 2014b). Modeling studies of e.g. Nippert et al. (2015) already suggested that understory plants do not exploit all accessible soil layers (including the top layers with high drought risk) in order to maximize water availability. Lower, but more resilient production is achieved instead by limiting root growth and water uptake to deeper depths, which could be confirmed by this study. It has to be additionally considered that the herbaceous vegetation already reached its growth peak when the experiment was conducted and thus maximizing root water uptake might not be a priority for the understory community past the growth peak and during seed production. Dubbart et al. (2014b) showed that the understory community is strongly adapted on a small spatial scale to the presence of oak trees regarding its species composition and overall vegetation period length. This is also observed in this study, with grasses dominating the understory community below the trees and forbs dominating in open areas (Fig. A1). Effectively this leads to an earlier seed production and senescence of less drought tolerant grasses in water competition with trees and a longer vegetation period of drought tolerant native forbs (i.e. *Tuberaria guttata* or *Tolpis barbata*) in open areas. Consequently, while understory species in the open area remained a net sink for carbon during the entire experiment, the understory community below the trees was at the verge of

senescence and turned into a net source for carbon at the last experimental date (Fig. A2), adding explanation to the site-specific differences of transpiration rate in response to event water (Fig. 5).

Recently, Volkmann et al. (2016a) used a similar flux / isotope approach to test the widespread dogma that plant water uptake depth is primarily controlled by root density distribution. While grassland species did not strongly alter their uptake pattern during the measurement campaign their water uptake depth profile was not in accordance with their root density distribution, with 85 % in the upper 10 cm of the soil profile. This clearly indicates that adapting the water uptake to soil water availability plays a role, but probably on longer time scales than what we observed during the 10 day's lasting experiment. The development of membrane-based *in-situ* methods of soil water (Gaj et al., 2015; Rothfuss et al., 2015; Volkmann et al., 2016a), xylem sap sampling (Volkmann et al., 2016b) and transpiration (Dubbert et al., 2014a; Dubbert et al., 2017) will advance the studies of dynamic changes in ecohydrological soil-vegetation feedbacks in the future. Furthermore, the coupling of isotope laser spectrometers to gas-exchange chambers and soil or xylem equilibration probes overcomes the cost and time consuming classical destructive sampling methods. Recent studies (Orlowski et al., 2013) showed significant isotopic deviations between actual soil water that is available for the plants and water that is cryogenically extracted from soil samples depending on soil type. While we did not observe this in sandy soils at our study site, these effects might severely hamper the usefulness of destructive soil sampling techniques in clay or loam soils. The newly developed *in situ* techniques will thus facilitate cost-effective measurements of soil or xylem isotopic signatures with highest resolution, enhancing our capacity to study the dynamics in soil water infiltration, in the uptake of water by plants and in the partitioning of evapotranspiration.

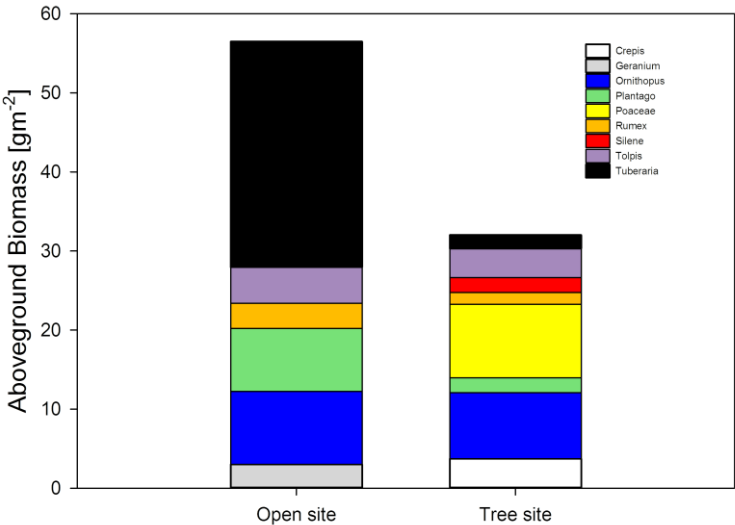
20 **5 Conclusion**

In this study, the various interactions between understory vegetation and trees of a Mediterranean cork oak woodland affecting the ecosystem water flows could be quantified. The immediate on-site determination (with CRDS) of the isotope ratios from different soil and ecosystem compartments in combination with *in situ* sampling methods enhanced the resolution, precision and reliability of our results. This facilitated the tracing of the fate of rain pulse transitions through a typical Mediterranean savannah ecosystem using stable water isotopes.

Regardless of the presence of vegetation, the total evapotranspirative water loss of soil and understory remains unchanged, but infiltration rates decreased by 24% (hypothesis I rejected). Still, the amount of unproductive evaporation is largely reduced, in favour of transpiration. Adjustments of main root water uptake depth to changing soil water availability after rain pulses could not be observed (hypothesis II supported). Consequently the understory plants could not utilize the applied precipitation of 20 mm. Hence, these understory plants were forced into water competition with trees, rooting at deeper soil layers. Crown shading of cork oak trees altered micro climatic conditions, thus fostering infiltration and considerably reducing understory and soil evapotranspiration (hypothesis III, part one supported). Despite these benefits, understory

plants in immediate vicinity of trees suffer from systematically lower soil moistures in deeper layers leading to premature senescence at the onset of drought (hypothesis III, part two rejected).

Appendix A



5 Figure A1: Aboveground biomass on vegetated plots during the experiment time given for each genus. Standard errors are not given for the sake of clarity, but amount on average 30% of displayed genus biomass.

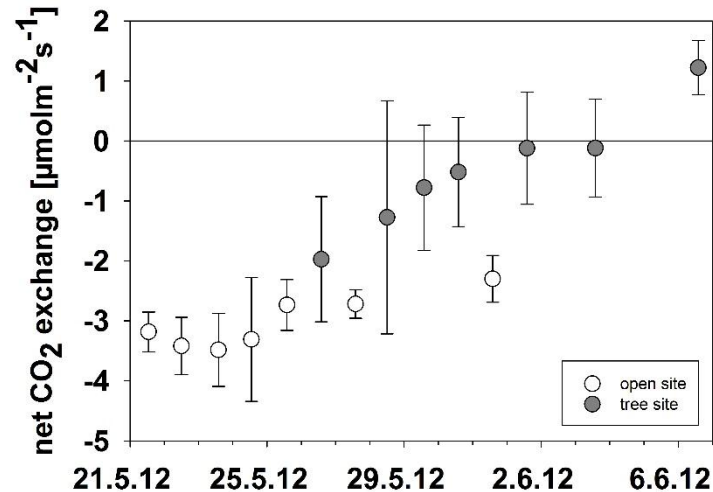


Figure A2: Mean midday net ecosystem exchange (NEE) of the understory vegetation at the open site (white circles) and the tree site (dark grey circles).

Author contribution

Arndt Piayda and Maren Dubbert contributed equally to experimental work, data analysis and writing the manuscript. Rolf
5 Siegwolf proofread the manuscript. Matthias Cuntz contributed to data analysis and proofread the manuscript. Christiane
Werner proofread the manuscript.

Competing interests

The authors declare that they have no conflict of interest.

Acknowledgements

10 We gratefully acknowledge excellent help in the laboratory by Ilse Thaufelder. Funding for this study was provided by the
DFG grants to CW and MC (WATERFLUX Project: # WE 2681/6-1, # CU 173/2-1), as well as MD (# DU 1688/1-1).

References

- Asbjornsen, H., Shepherd, G., Helmers, M. and Mora, G., 2008. Seasonal patterns in depth of water uptake under contrasting
annual and perennial systems in the Corn Belt Region of the Midwestern US. *Plant and Soil*, 308(1-2): 69-92.
- 15 Bargués Tobella, A., Reese, H., Almaw, A. and Bayala, J., 2014. The effect of trees on preferential flow and soil infiltrability
in an agroforestry parkland in semiarid Burkina Faso *Water Resources Research*, 50.
- Belsky, A.J., 1994. Influences of Trees on Savanna Productivity: Tests of Shade, Nutrients, and Tree-Grass Competition
Ecology, 75.
- Bhark, E.W. and Small, E.E., 2003. Association between plant canopies and the spatial patterns of infiltration in shrubland
20 and grassland of the Chihuahuan Desert, New Mexico. *Ecosystems*, 6(2): 185-196.
- Braud, I., Bariac, T., Gaudet, J.P. and Vauclin, M., 2005. SiSPAT-Isotope, a coupled heat, water and stable isotope (HDO
and (H₂O)-O-18) transport model for bare soil. Part I. Model description and first verifications. *Journal of*
Hydrology, 309: 277-300.
- Breman, H. and Kessler, J.-J., 1999. *Woody Plants in Agro-Ecosystems of Semi-Arid Regions* Advanced Series in
25 *Agricultural Sciences*. Springer Berlin, Heidelberg.
- Caldwell, M.M., 1987. *PLANT ARCHITECTURE AND RESOURCE COMPETITION*, Schulze, E.-D. And H. Zwoelfer.
Ecological Studies, pp. 164-179.
- Caldwell, M.M., Dawson, T.E. and Richards, J.H., 1998. Hydraulic lift: Consequences of water efflux from the roots of
plants. *Oecologia*, 113(2): 151-161.
- 30 Cernusak, L.A. et al., 2016. Stable isotopes in leaf water of terrestrial plants. *Plant, Cell and Environment*.
- Costa e Silva, F. et al., 2015. Effects of extremely dry winter on net ecosystem carbon exchange and tree phenology at a cork
oak woodland. *Agricultural and Forest Meteorology*.
- Craig, H. and Gordon, L.I., 1965. Deuterium and oxygen-18 variations in the ocean and the marine atmosphere. Paper
presented at the Stable Isotopes in Oceanographic Studies and Paleotemperatures, Spoleto, Italy.

- D'Odorico, P. and Porporato, A., 2006. Soil moisture dynamics in water-limited ecosystems. *Dryland Ecohydrology*: 31-46.
- David, T.S. et al., 2006. Rainfall interception by an isolated evergreen oak tree in a Mediterranean savannah. *GHydrological Processes*, 20: 23.
- 5 Dawson, T.E., 1993. HYDRAULIC LIFT AND WATER-USE BY PLANTS - IMPLICATIONS FOR WATER-BALANCE, PERFORMANCE AND PLANT-PLANT INTERACTIONS. *Oecologia*, 95(4): 565-574.
- Devitt, D.A. and Smith, S.D., 2002. Root channel macropores enhance downward movement of water in a Mojave Desert ecosystem. *Journal of Arid Environments*, 50(1): 99-108.
- Duan, Q., Sorooshian, S. and Gupta, V., 1992. Effective and efficient global optimization for conceptual rainfall-runoff models. *Water Resources Research*(28).
- 10 Dubbert, M., Cuntz, M., Piayda, A., Maguas, C. and Werner, C., 2013. Partitioning evapotranspiration - Testing the Craig and Gordon model with field measurements of oxygen isotope ratios of evaporative fluxes, *Journal of Hydrology*, pp. 142-153.
- Dubbert, M., Cuntz, M., Piayda, A. and Werner, C., 2014a. Oxygen isotope signatures of transpired water vapor – the role of isotopic non-steady-state transpiration under natural conditions. *New Phytologist*, 203: 1242-1252.
- 15 Dubbert, M., Kübert, A. and Werner, C., 2017. Impact of Leaf Traits on Temporal Dynamics of Transpired Oxygen Isotope Signatures and Its Impact on Atmospheric Vapor. *Frontiers in plant science*, 8: 12.
- Dubbert, M. et al., 2014b. Influence of tree cover on herbaceous layer development and carbon and water fluxes in a Portuguese cork oak woodland. *Acta Oecologica-International Journal of Ecology*.
- Dubbert, M. et al., 2014c. Stable oxygen isotope and flux partitioning demonstrates understory of an oak savanna contributes up to half of ecosystem carbon and water exchange. *Frontiers in plant science*, 5: 530.
- 20 Duckstein, L., 1981. Multiobjective optimization in structural design: the model choice problem. In: E. Atrek (Editor), *New Directions in Optimum Structural Design*. Wiley.
- Ehleringer, J.R. and Dawson, T.E., 1992. Water uptake by plants: perspectives from stable isotope composition. *Plant, Cell and Environment*, 15.
- 25 Eldridge, D.J. et al., 2010. Interactive Effects of Three Ecosystem Engineers on Infiltration in a Semi-Arid Mediterranean Grassland Ecosystems, 13.
- Eldridge, D.J. and Freudenberger, D., 2005. Ecosystem wicks: Woodland trees enhance water infiltration in a fragmented agricultural landscape in eastern Australia. *Australian Ecology*, 30.
- Gaj, M. et al., 2015. In-situ unsaturated zone stable isotope (2H and 18O) measurements in semi-arid environments using tunable off-axis integrated cavity output spectroscopy. *Hydrological Earth System Sciences Discussions*, 12.
- 30 Garvelmann, J., Kulls, C. and Weiler, M., 2012. A porewater-based stable isotope approach for the investigation of subsurface hydrological processes. *Hydrology and Earth System Sciences*, 16(2): 631-640.
- Goldshleger, N., Ben-Dor, E., Benyamini, Y., Blumberg, D. and Agassi, M., 2002. Spectral properties and hydraulic conductance of soil crusts formed by raindrop impact. *International Journal of Remote Sensing*, 23.
- 35 Greig-Smith, P., 1979. Pattern in Vegetation. *The Journal of Ecology*, 67.
- Haverd, V. and Cuntz, M., 2010. Soil-Litter-Iso: A one-dimensional model for coupled transport of heat, water and stable isotopes in soil with a litter layer and root extraction. *Journal of Hydrology*, 388: 438-455.
- Hipondoka, M., Aranibar, J., chiara, C., Lihavha, M. and Macko, S., 2003. Vertical distribution of grass and tree roots in arid ecosystems of Southern Africa: niche differentiation or competition? . *Journal of Arid Environments*, 54.
- 40 Holdo, R.M. and Planque, R., 2013. Revisiting the Two-Layer Hypothesis: Coexistence of Alternative Functional Rooting Strategies in Savannas. *Plos One*, 8.
- IPCC, 2013. *Climate Change 2013. The Physical Science Basis*.
- Joffre, R. and Rambal, S., 1993. HOW TREE COVER INFLUENCES THE WATER-BALANCE OF MEDITERRANEAN RANGELANDS. *Ecology*, 74(2): 570-582.
- 45 Kulmatiski, A. and Beard, K.H., 2013. Root niche partitioning among grasses, saplings, and trees measured using a tracer technique. *Oecologia*, 171.
- Kulmatiski, A., Beard, K.H., Verqej, R.J.T. and February, E.C., 2010. A depth-controlled tracer technique measures vertical, horizontal and temporal patterns of water use by trees and grasses in a subtropical savanna. *New Phytol*, 188.
- Kurz-Besson, C. et al., 2006. Hydraulic lift in cork oak trees in a savannah-type Mediterranean ecosystem and its contribution to the local water balance. *Plant and Soil*, 282(1-2): 361-378.
- 50

- Maestre, F.T., Huesca, M., Zaady, E., Bautista, S. and Cortina, J., 2002. Infiltration, penetration resistance and microphytic crust composition in contrasted microsites within a Mediterranean semi-arid steppe. *Soil Biology and Biochemistry*, 34: 4.
- 5 Majoube, M., 1971. Fractionation in O-18 Between Ice and Water Vapor. *Journal de Chimie Physique et de Physico-Chimie Biologique*, 68(4): 625-&.
- Mathieu, R. and Bariac, T., 1996. A numerical model for the simulation of stable isotope profiles in drying soils. *Journal of Geophysical REsearch*, 101: 12.685-12.696.
- Merlivat, L., 1978. Molecular Diffusivities of (H₂O)-O-16 and (H₂O)-O-18 in Gases. *Journal of Chemical Physics*, 69(6): 2864-2871.
- 10 Moreira, M.Z. et al., 1997. Contribution of transpiration to forest ambient vapour based on isotopic measurements. *Global Change Biology*, 3(5): 439-450.
- Moreno, G., 2008. Response of understorey forage to multiple tree effects in Iberian dehesas, *Agriculture, Ecosystems and Environment*, pp. 244.
- 15 Nippert, J.B., Holdo, R.M. and Sayer, E., 2015. Challenging the maximum rooting depth paradigm in grasslands and savannas. *Functional Ecology*, 29.
- Orlowski, N., Frede, H.-G., Brüggemann, N. and Breuer, L., 2013. Validation and application of a cryogenic vacuum extraction system for soil and plant water extraction for isotope analysis. *Journal of sensors and Sensor Systems*, 2: 179-193.
- Pape, L. et al., 2009. An automated dynamic chamber system for surface exchange measurement of non-reactive and reactive trace gases of grassland ecosystems. *Biogeosciences*, 6(3): 405-429.
- 20 Philips, D.L., Newsome, S.D. and Gregg, J.W., 2005. Combining sources in stable isotope mixing models: alternative methods *Oecologia*, 144.
- Piayda, A. et al., 2014. Drought impact on carbon and water cycling in a Mediterranean *Quercus suber* L. woodland during the extreme drought event in 2012. *Biogeosciences*.
- 25 Piayda, A. et al., 2015. Influence of woody tissue and leaf clumping on vertically resolved leaf area index and angular gap probability estimates. *For. Ecol. Manage.*
- Porporato, A., Daly, E. and Rodriguez-Iturbe, I., 2004. Soil Water Balance and Ecosystem Response to Climate Change *The American Naturalist*, 164.
- Pueyo, Y., Moret-Fernandez, D., Saiz, H., Bueno, C.G. and Alados, C.L., 2012. Relationships Between Plant Spatial Patterns, Water Infiltration Capacity, and Plant Community Composition in Semi-arid Mediterranean Ecosystems Along Stress Gradients *Ecosystems*, 16.
- 30 Rodriguez-Iturbe, I., 2000. Ecohydrology: A hydrologic perspective of climate-soil-vegetation dynamics. *Water Resources Research*, 36(1): 3-9.
- Rothfuss, Y. et al., 2015. Long-term and high-frequency non-destructive monitoring of water stable isotope profiles in an evaporating soil column. *Hydrol. Earth Syst. Sci.*, 19: 4067-4080.
- 35 Scholes, R.J. and Archer, S.R., 1997. Tree-grass interactions in savannas. In: D.G. Fautin (Editor), *Annual Review of Ecology and Systematics*. *Annual Review of Ecology and Systematics*, pp. 517-544.
- Schwinning, S. and Ehleringer, J.R., 2001. Water use trade-offs and optimal adaptations to pulse-driven arid ecosystems. *Journal of Ecology*, 89(3): 464-480.
- 40 Tromble, J.M., 1988. WATER INTERCEPTION BY 2 ARID LAND SHRUBS. *Journal of Arid Environments*, 15(1): 65-70.
- Volkman, T.H.M., Haberer, K., Gessler, A. and Weiler, M., 2016a. High-resolution isotope measurements resolve rapid ecohydrological dynamics at the soil-plant interface. *New Phytologist*.
- Volkman, T.H.M., Kühnhammer, K., Herbstritt, B., Gessler, A. and Weiler, M., 2016b. A method for in situ monitoring of the isotope composition of tree xylem water using laser spectroscopy. *Plant, Cell and Environment*.
- 45 von Caemmerer, S. and Farquhar, G.D., 1981. SOME RELATIONSHIPS BETWEEN THE BIOCHEMISTRY OF PHOTOSYNTHESIS AND THE GAS-EXCHANGE OF LEAVES. *Planta*, 153(4): 376-387.
- Walter, H., Burnett, J.H. and Mueller-Dombois, D., 1971. *Ecology of tropical and subtropical vegetation*. Oliver and Boyd Edinburgh.

- Wang, L. et al., 2012. Dryland ecohydrology and climate change: critical issues and technical advances. *Hydrology and Earth System Sciences*, 16(8): 2585-2603.
- Werner, C. and Dubbert, M., 2016. Resolving rapid dynamics of soil-plant-atmosphere interactions. *New Phytologist*.
- Yakir, D. and Sternberg, L.D.L., 2000. The use of stable isotopes to study ecosystem gas exchange. *Oecologia*, 123(3): 297-311.
- Yepez, E.A., Williams, D.G., Scott, R.L. and Lin, G.H., 2003. Partitioning overstory and understory evapotranspiration in a semiarid savanna woodland from the isotopic composition of water vapor. *Agricultural and Forest Meteorology*, 119(1-2): 53-68.

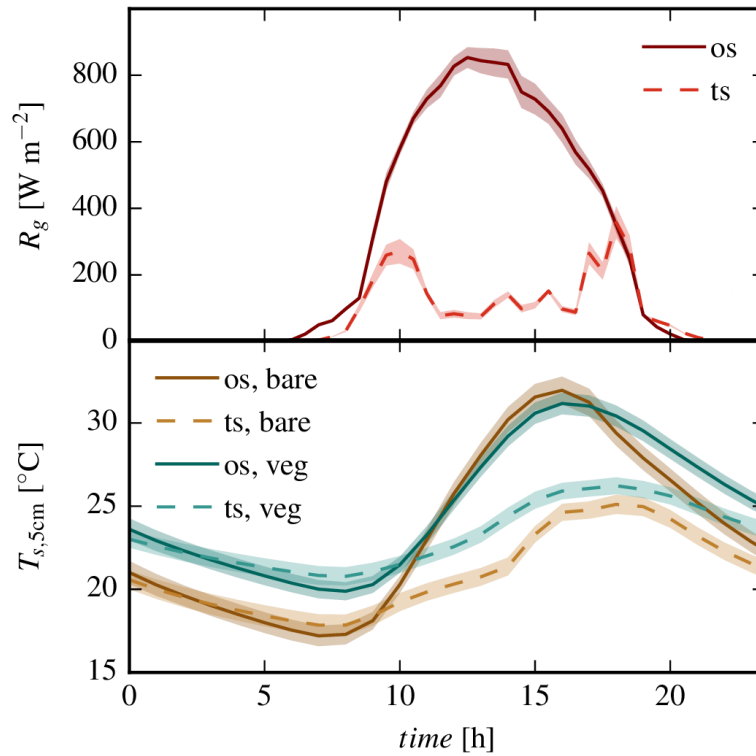


Figure 1: Daily cycles, averaged over the experiment period, of a) global radiation R_g in 1.5 m height and b) soil temperature $T_{s,5cm}$ in -5 cm depth under bare soil (bare) or vegetation cover (veg). Observations at open sites between tree crowns (os) and shaded sites beneath tree crowns (ts) are shown. Uncertainty bands display standard error.

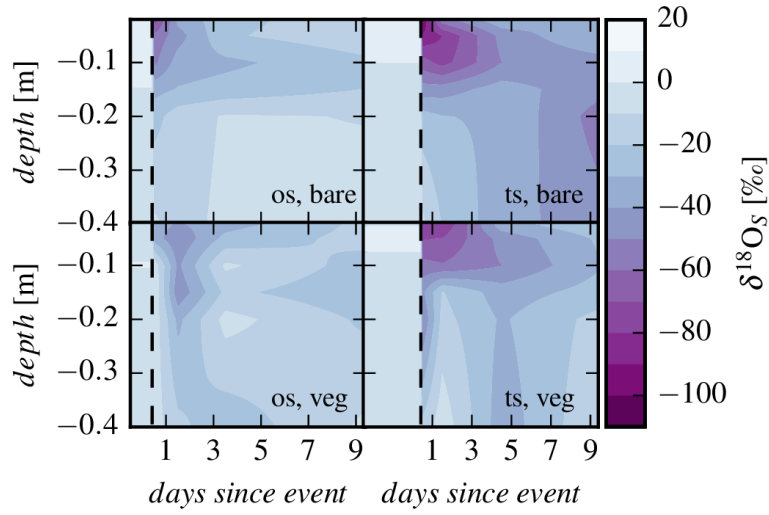


Figure 2: Mean daily isotopic composition of soil water $\delta^{18}O_s$ during experiment under bare soil (bare) or vegetation cover (veg) at open sites between tree crowns (os) and shaded sites beneath tree crowns (ts). Dashed lines mark time of watering event. Interpolation method: linear. The standard error for soil isotopic composition during the experiment amounts on average 1.4‰ in natural abundance.

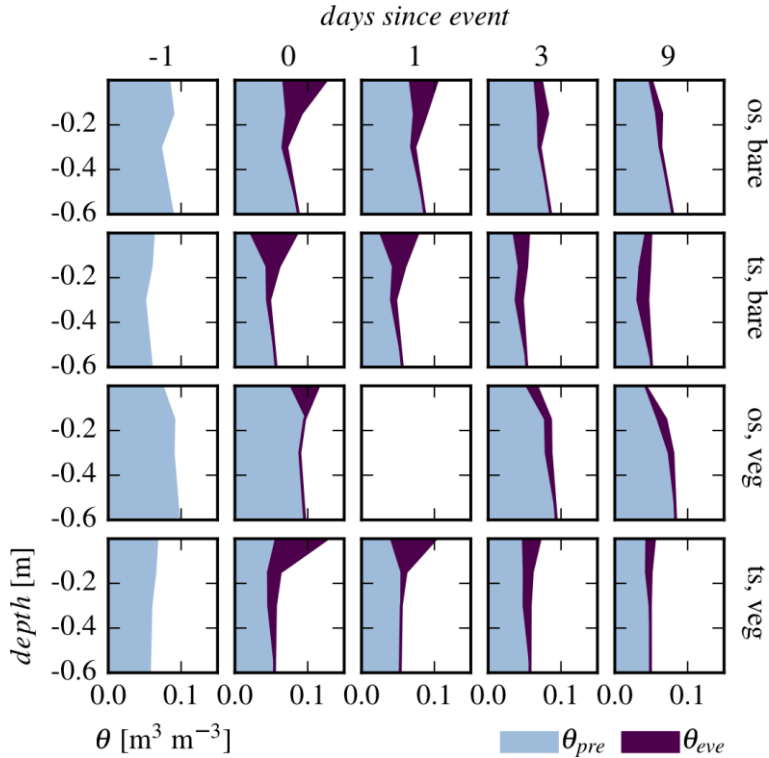


Figure 3: Mean daily soil water content θ along soil depth separated in pre-event soil water content θ_{pre} and infiltrated event soil water content θ_{eve} . Observations are displayed for plots under bare soil (bare) or vegetation cover (veg) at open sites between tree crowns (os) and shaded sites beneath tree crowns (ts). Numbers on top mark days since the watering event. Uncertainties for soil moisture observations during the experiment amount on average 2.3%_{vol.} propagated from the observations. Event water partitioning for day 1 on open, vegetated plots needed to be omitted due to insufficient field data quality.

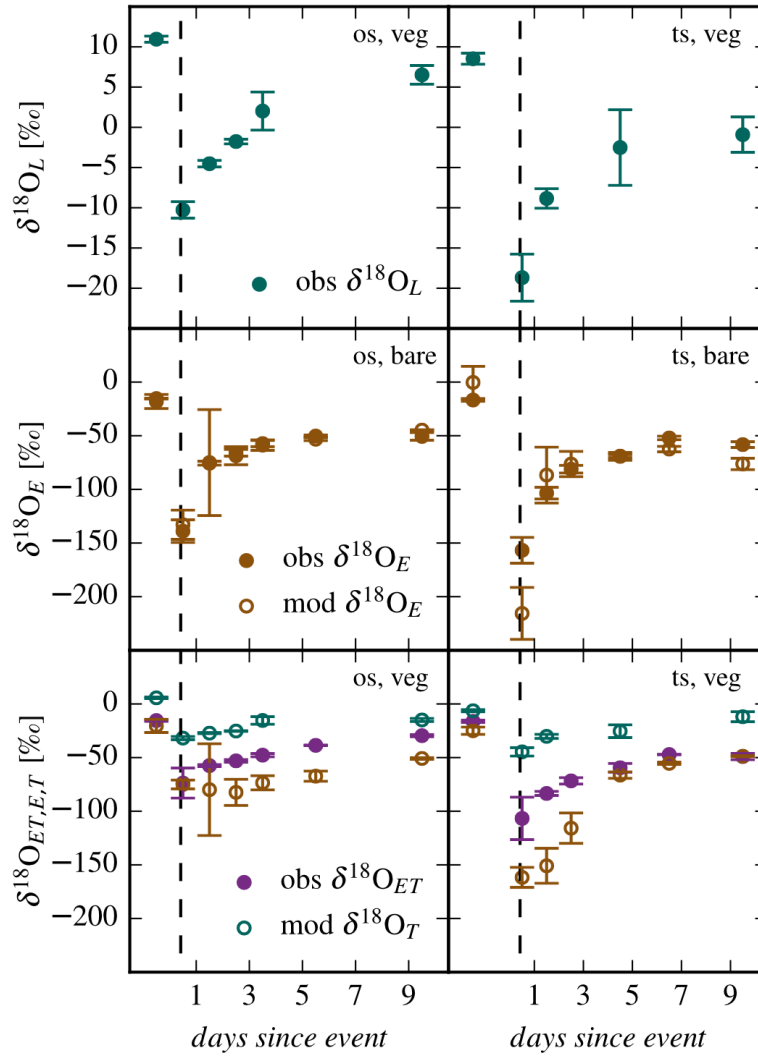


Figure 4: Mean daily isotopic composition of bulk leaf water $\delta^{18}O_L$, soil evaporation $\delta^{18}O_E$, plant transpiration $\delta^{18}O_T$ and combined evapotranspiration $\delta^{18}O_{ET}$ from bare soil (bare) or vegetation plots (veg) at open sites between tree crowns (os) and shaded sites beneath tree crowns (ts). Full dots represent observed values (obs), hollow dots represent modelled values (mod). Dashed lines mark time of watering event. Uncertainty bars display standard error.

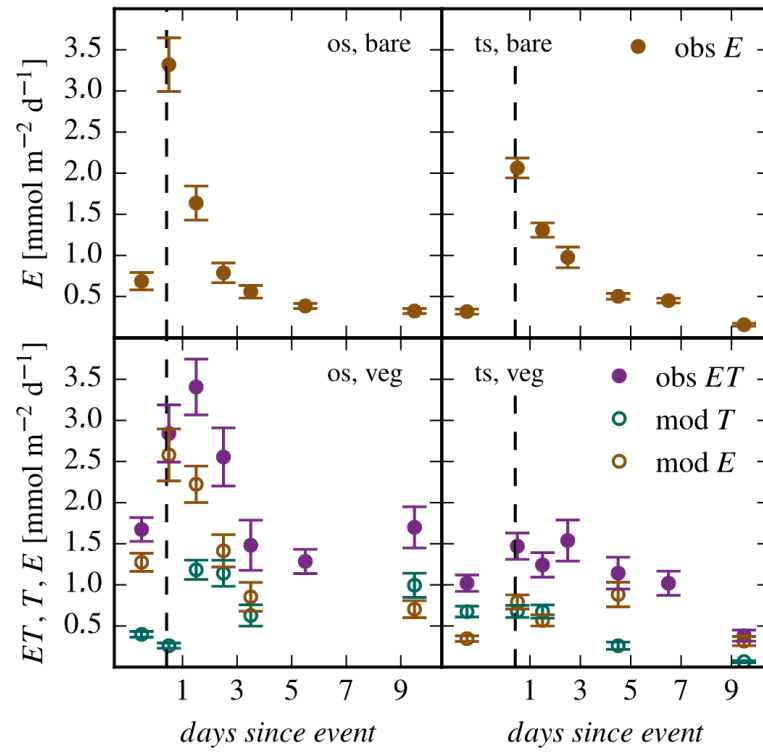


Figure 5: Mean daily flux rates of soil evaporation E , plant transpiration T and combined evapotranspiration ET from bare soil (bare) or vegetation plots (veg) at open sites between tree crowns (os) and shaded sites beneath tree crowns (ts). Full dots represent observed values (obs), hollow dots represent modelled values (mod). Dashed lines mark time of watering event. Uncertainty bars display standard error.

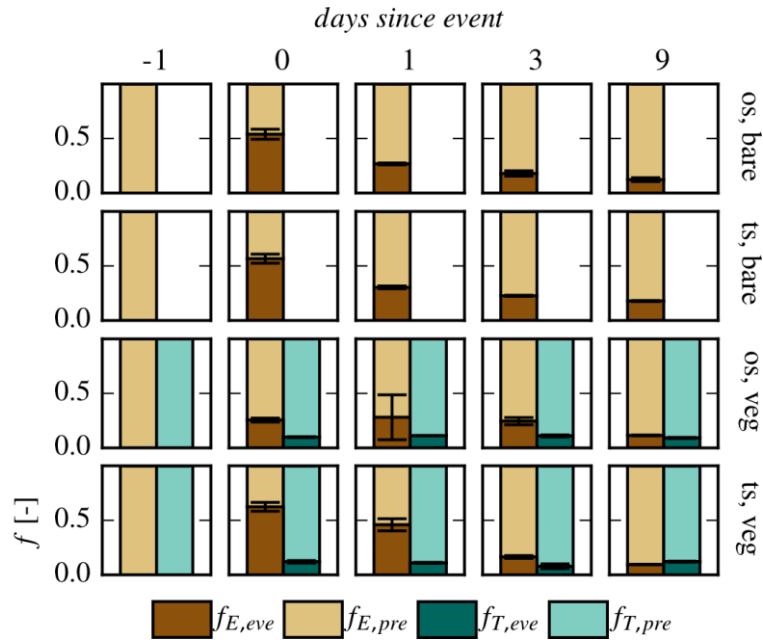


Figure 6: Mean daily fractions f of event water (*eve*) and pre-event water (*pre*) in soil evaporation E and plant transpiration T from bare soil (bare) or vegetation plots (veg) at open sites between tree crowns (os) and shaded sites beneath tree crowns (ts). Numbers on top mark days after watering event. Uncertainty bars display standard error.

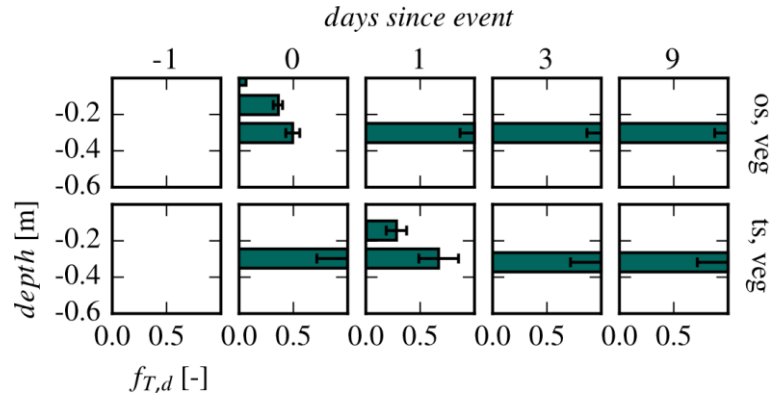


Figure 7: Mean daily fractions of root water uptake $f_{T,d}$ of understory plants for modelled soil depths. Numbers on top mark days after watering event. Uncertainty bars display standard error.